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Spatiotemporal diversity, structure and trophic guilds of insect assemblages in a semi-arid Sabkha ecosystem

Haroun Chenchouni, Taha Menasria, Souad Neffar, Smail Chafaa, Liès Bradai, Mohamed Nacer Mekahlia, Rachid Chaibi, Djamel Bendjoudi, Abdelkrim Si Bachir

The current study highlights some knowledge on the diversity and structure of insect communities and trophic groups living in Sabkha Djendli (semi-arid area of Northeastern Algeria). The entomofauna was monthly sampled from March to November 2006 using pitfall traps at eight sites located at the vicinity of the Sabkha. Structural and diversity parameters (species richness, Shannon index, evenness) were measured for both insect orders and trophic guilds. The canonical correspondence analysis (CCA) was applied to determine how vegetation parameters (species richness and cover) influence spatial and seasonal fluctuations of insect assemblages. The catches totalled 434 insect individuals classified into 75 species, 62 genera, 31 families and 7 orders, where Coleoptera and Hymenoptera were the most abundant and constant over seasons and study stations. Spring and autumn presented the highest values of diversity parameters. Based on catch abundance, the structure of functional trophic groups was predator (37.3%), saprophagous (26.7%), phytophagous (20.5%), polyphagous (10.8%), coprophagous (4.6%); whereas in terms of numbers of species, they can be classified as phytophagous (40%), predators (25.3%), polyphagous (13.3%), saprophagous (12%), coprophagous (9.3%). The CCA demonstrated that phytophagous and saprophagous as well as Coleoptera and Orthoptera were positively correlated with the two parameters of vegetation, especially in spring and summer. While the abundance of coprophagous was positively correlated with species richness of plants, polyphagous density was positively associated with vegetation cover. The insect community showed high taxonomic and functional diversity that is closely related to diversity and vegetation cover in different site stations and seasons.

1 **Spatiotemporal diversity, structure and trophic guilds of insect**
2 **assemblages in a semi-arid Sabkha ecosystem**

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19 Running title: **Insect Assemblages of a Sabkha Ecosystem**

20 **Abstract**

21 The current study highlights some knowledge on the diversity and structure of insect
22 communities and trophic groups living in Sabkha Djendli (semi-arid area of
23 Northeastern Algeria). The entomofauna was monthly sampled from March to
24 November 2006 using pitfall traps at eight sites located at the vicinity of the
25 Sabkha. Structural and diversity parameters (species richness, Shannon index,
26 evenness) were measured for both insect orders and trophic guilds. The canonical
27 correspondence analysis (CCA) was applied to determine how vegetation
28 parameters (species richness and cover) influence spatial and seasonal fluctuations
29 of insect assemblages. The catches totalled 434 insect individuals classified into 75
30 species, 62 genera, 31 families and 7 orders, where Coleoptera and Hymenoptera
31 were the most abundant and constant over seasons and study stations. Spring and
32 autumn presented the highest values of diversity parameters. Based on catch
33 abundance, the structure of functional trophic groups was predator (37.3%),
34 saprophagous (26.7%), phytophagous (20.5%), polyphagous (10.8%), coprophagous
35 (4.6%); whereas in terms of numbers of species, they can be classified as
36 phytophagous (40%), predators (25.3%), polyphagous (13.3%), saprophagous
37 (12%), coprophagous (9.3%). The CCA demonstrated that phytophagous and
38 saprophagous as well as Coleoptera and Orthoptera were positively correlated with
39 the two parameters of vegetation, especially in spring and summer. While the
40 abundance of coprophagous was positively correlated with species richness of
41 plants, polyphagous density was positively associated with vegetation cover. The

42 insect community showed high taxonomic and functional diversity that is closely
43 related to diversity and vegetation cover in different site stations and seasons.

44 **Keywords:** Sabkha Djendli; entomological biodiversity; insect community ecology;
45 pitfall trapping; Algeria; ecological niche; functional groups; wetlands; conservation
46 biology; semi-arid lands.

47 **1. Introduction**

48 Wetlands are recognized as important ecosystems in terms of biodiversity and
49 functional role. These ecosystems include a remarkable range of habitats that are
50 ecologically considered among the most productive ecosystems worldwide, with large
51 socio-economic importance and high heritage values for humanity. They play crucial
52 and major ecological functions, including trapping, absorbing and eliminating of
53 potential toxic chemicals and pollutants, storage of natural carbon, recycling of
54 nutrients, as well as they contribute to groundwater recharge in arid and semi-arid
55 regions. Unfortunately, wetlands are experiencing rapid degradation due to severe
56 transformations related to intensive human activities (Bobbink et al., 2006; Mitsch et
57 al., 2009).

58 More than 2000 wetlands are listed in Algeria, including 50 sites classified on
59 the Ramsar list of wetlands of international importance (Balla 2012). Most of large
60 inland saline depressions and backwaters “Sabkhas, Chotts, and Oases” are located in
61 arid and semi-arid regions, with a unique agglomeration of this type of sites in
62 northeastern of the country (Chenchouni & Si Bachir, 2010). The most characteristic
63 type of the Algerian wetlands is seasonal/intermittent endorheic type that consists of
64 Sabkha ecosystems “saline lakes” with typical alternation of drought phase in
65 summer and flooding in winter (Khaznadar et al., 2009; Balla, 2012).

66 Large-scale conservation programs focused on wetlands because these habitats
67 support both terrestrial and aquatic biota where biodiversity therein is remarkably

68 high (De Roeck et al., 2007). This biodiversity is the key factor maintaining the
69 structure, stability, and functioning of these ecosystems (Ivask et al., 2008). What
70 makes its conservation at different organizational levels "individual, population,
71 community, ecosystem" has become an issue that deserves national and
72 international attention (Bobbink et al., 2006; Montagna et al., 2012). Moreover,
73 regional contributions have also proven their impact in improving the knowledge
74 and conservation of these habitats (Piñero et al., 2011; Chaibi et al., 2012; Guezoul
75 et al., 2013).

76 As a biological model, invertebrates embrace a large species richness ranging
77 over several taxa with large magnitude of sizes. They colonise various microhabitats
78 and perform an extraordinary diverse functional roles, constituting thus key
79 organisms at different trophic levels inside food webs of wetland ecosystems
80 (Koricheva et al., 2000; Finke & Denno, 2002; Haddad et al., 2009; Piñero et al.,
81 2011). Although their relevant importance in ecosystem functioning of wetlands,
82 invertebrates were slightly used as criteria in conservation programs of wetlands
83 compared to specific criteria based on waterbirds and fishes, since only recently,
84 these organisms as well as other taxa were included in the ninth criterion used by
85 Ramsar Convention for considering wetlands internationally important (Mitsch et al.,
86 2009).

87 Furthermore, it is well known that biodiversity and structure of invertebrates,
88 particularly insects, in saline inland temporary wetlands are governed by two main
89 abiotic factors: hydroperiod "water regimes" and salinity (Bilton et al., 2001; Brock
90 et al., 2005; Gascon et al., 2005; Waterkeyn et al., 2008). Whereas the involved
91 biotic factors are dealing with vegetation traits and various biotic interactions of
92 food webs (Koricheva et al., 2000; Carver et al., 2009; Haddad et al., 2009).
93 However, although species diversity is a good parameter for valuing ecosystems and

94 defining conservation strategies, scarcity of species should also be taken into
95 account (Nijboer & Verdonschot, 2004).

96 Notwithstanding the multi-scale ecological surveys that investigated animal
97 biodiversity of the Algerian wetlands, specifically at the northeast of the country;
98 they focused on waterbirds (e.g. Samraoui and Samraoui, 2008), fishes (e.g. Chaibi
99 et al., 2012), and some other taxa like dragonflies (e.g. Samraoui et al., 2011),
100 whereas the ecology of terrestrial arthropods of Sabkha ecosystems remain very
101 little studied in these saline environments (Hogarth & Tigar, 2002).

102 Located in high plains of Northeast Algeria, the Sabkha Djendli is a seasonal salt
103 lake whose flora was thoroughly surveyed throughout the waterbody vicinity (Neffar
104 et al., 2014). However, the faunal communities, including insects, inhabiting the
105 Sabkha and its environs were very little investigated in connection with their
106 biotope, except some ornithological surveys of wintering waterbirds (e.g. Samraoui
107 and Samraoui, 2008; Bensizerara et al., 2013).

108 The study of relationships between spatiotemporal variation of invertebrate
109 communities and ecological parameters provides valuable information for conservation
110 assessment and restoration planning and may efficiently guide the implementation of
111 future management program (Comin & Comin, 1992; Fischer & Lindenmayer, 2007;
112 Montagna et al., 2012). Furthermore, the assessment of functional trophic groups is
113 crucial to outline the structure of food webs and accordingly identify any perturbation in
114 the ecosystem functioning (Chesson & Huntly, 1997; Gascon, 2005), particularly under
115 changing environmental conditions. Indeed, some insect groups such as dragonflies,
116 hoverflies and some ground beetles (particularly Carabidae) represent good indicators
117 of biodiversity assessment and monitoring in wetlands and mesic environments (Rainio
118 & Niemelä, 2003; Sánchez-Fernández et al., 2006; Hepp & Melo, 2013). In fact the core
119 aim of the current study is placed within the perspective of insect biodiversity
120 assessment for conservation purpose as outlined here above.

121 Thereby, the objectives of this pioneer study are dealing with the framework of
122 understanding entomofauna composition of Sabkha Djendli. This treatise aims to (i)
123 provide accurate information on the spatiotemporal variation of the composition,
124 structure and diversity indices of the insect community inhabiting the vicinity of the
125 Sabkha; (ii) evaluate ecological status and diversity of the functional trophic groups in
126 relation to seasons and site orientations of the salt lake; (iii) understand the structural
127 and functional similarities of entomofauna communities living around the Sabkha; (iv)
128 assess the effect of seasons and site orientations on the spatiotemporal abundance
129 variations of both insect orders and functional trophic groups; and (v) determine how
130 vegetation parameters influence spatial and seasonal fluctuations of insect
131 assemblages.

132 **2. Materials and methods**

133 **2.1. Study area**

134 Sabkha Djendli (35°42'56"N, 6°31'46"E) is one part of the wetland complex at
135 the High Plains region including Batna in eastern Algeria (Figure 1). The site is a
136 temporary lake with brackish-salt water that highly depends on rainfall amounts and
137 water regime. Sabkha Djendli covers about 3,700 ha with an average altitude of 833
138 m in an area where inhabitants are mainly involved in agricultural activities like
139 cereal and fruit cultivation and livestock of sheep and cattle.

140 Based on meteorological data provided by the meteorological station of Batna
141 (WMO Id: 60468) of the period 1974–2013, the climate of the study area is typically
142 semi-arid Mediterranean, characterised by cold-wet winters and hot-dry summers.
143 The dry period extends over four months from June to September. Precipitations are
144 erratic and experience large temporal variations. The coldest month is January with
145 an average temperature of 5.3°C and the hottest month is July with an average
146 temperature of 25°C. The relative humidity of the air fluctuates between 40% to

147 75% and the winds are generally low in dominance west to south-west, with the
148 passage of Sirocco in summer during July–August.

149 The natural vegetation is represented by halophytes such as *Atriplex halimus*,
150 *Suaeda fruticosa*, *Suaeda vermiculata*, and *Sarcocornia fruticosa*, but also other
151 spontaneous vegetation like *Tamarix gallica*, *Artemisia herba-alba* and *Juncus*
152 *maritimus* (Neffar et al., 2014). The current entomological survey was carried on the
153 belt of halophytic vegetation surrounding the Sabkha (Figure 1).

154 **2.2. Sampling design**

155 At eight cardinal and inter-cardinal points of the site border of Sabkha Djendli,
156 the insect fauna was monthly sampled during the period March to November 2006.
157 Halophytic vegetation dominated in the entire sampled area. At each sampling
158 points, insects were trapped using nine pitfall traps (Spence & Niemela, 1994),
159 which were set up inside a square plot of 400 m² (20 m × 20 m). These uncovered
160 traps are aligned 3–3 along three rows and spaced from each other with 5 m
161 (Figure 1). Each trap was filled to 3/4 of water containing a wetting agent, and its
162 catches were monthly recovered after one week trapping since first setting day. The
163 caught specimens were identified to the genus and species. The nomenclature and
164 taxonomy of species were based on up-to dated references (Bouchard et al., 2011;
165 de Jong 2013; Löbl and Smetana, 2013; Anichtchenko et al., 2014; AntWeb, 2014;
166 Eades et al., 2014).

167 **2.3. Data mining and statistical analysis**

168 Data of insect catches from the nine uncovered traps were pooled to form one
169 sample per sampling station per month. Data were presented by taxonomic orders and
170 trophic groups and were expressed for orientation points and seasons to facilitate
171 spatiotemporal comparisons for all the following parameters. The relative abundance

172 (RA) was determined as the ratio of number of individuals rounded to the total number
173 of individuals recorded (N_i). Occurrence frequency (Occ) was calculated for each
174 species by the number of stations wherein the species was found / the total number of
175 sampled stations (Magurran, 2004). Four species groups are distinguished by Bigot and
176 Bodot (1973), according to their frequencies of occurrence: Very accidental species
177 (Vac): an occurrence of less than 10%; Accidental species (Acc) occurrence varies
178 between 10 and 24%; Common species (Cmt) are present in 25–49%; Constant species
179 (Cst) are present in 50% or more of the samples.

180 Biodiversity of insects was assessed by species richness "SR", which
181 corresponds to the total number of identified insect species at each station or
182 season. In addition, Shannon's index ($H' = -\sum p_i \times \log_2 p_i$) and evenness
183 (Evenness = $H' / \log_2 SR$) were applied for measuring insect diversity in each sampled
184 station and season period based on the relative density p_i of the i^{th} species
185 (Magurran 2004).

186 Jaccard similarity index (C_j) was used to compare insect species richness
187 between stations taken in pairs. Given two stations A and B, C_j was computed as:
188 $C_j = c / (a + b - c)$. Where a and b = the total number of species present in station A
189 and B, respectively; c = the number of species found in both stations (Magurran,
190 2004). Agglomerative hierarchical clustering (AHC) was applied to cluster sampled
191 stations according to their species richness based on Jaccard similarity index (C_j).
192 The method used in agglomeration was unweighting pair-group average.

193 Two-way ANOVAs were applied including the effect of 'Orientation' and 'Season'
194 to test spatiotemporal variations of abundances of both taxonomic orders and
195 trophic groups. Moreover, Pearson's Chi-squared test (χ^2) was applied to look for
196 dependencies between the distributions of structural traits values (N_i , SR , N_i/SR , H' ,
197 Evenness) of the functional trophic groups vis-à-vis both study stations and seasons.

198 The spatiotemporal gradients of insect assemblages were analyzed in relation
199 with vegetation traits using a canonical correspondence analysis (CCA). The data
200 used were the abundances of both taxonomic orders and trophic groups on the
201 study seasons and orientations where they were counted. For the spontaneous
202 vegetation, two parameters were assessed at each orientation and season: the
203 vegetation cover (%) and total species richness (number of plant species). These
204 data were generated from Neffar et al. (2014). Since the CCA has the ability to
205 combine ordination and gradient analysis functions in a readily
206 interpretable manner, it was applied to relate spatiotemporal insect abundances to
207 vegetation variables in order to highlight relationships between spatiotemporal
208 variations of insects and vegetation traits as explanatory variables (Jongman et al.,
209 1995). At the end of overcoming the disadvantage effect of scale differences in data,
210 insect densities as well as vegetation variables were normalized using normal
211 distribution transformation based on the average and standard deviation of each
212 input. Finally, Pearson's correlation was used to test the significance of relationships
213 between densities of insect assemblages (of both taxonomic orders and trophic
214 groups) and vegetation parameters (vegetation cover and species richness).

215 **3. Results**

216 ***3.1. Insect community and taxonomic composition***

217 Pitfall sampling of entomofauna at Sabkha Djendli revealed an insect community
218 composed of 75 species from 434 individuals caught. This entomofauna can be
219 classified into 7 orders, 31 families and 62 genera (Table 1). Coleoptera was the best
220 represented with 238 (54.8%) individuals caught belonging to 39 species and 15
221 families, followed by Hymenoptera with 149 (34.3%) individuals of 18 species and 8
222 families, then came Orthoptera with 22 individuals (10 species and 2 families). The
223 orders Dermaptera, Heteroptera, Homoptera and Diptera were poorly represented by

224 either species or catch abundance. Furthermore, the identified entomofauna included
225 five functional trophic groups: the phytophagous with 30 species, predators with 19
226 species, polyphagous with 10 insects, saprophagous with 9 species and coprophagous
227 with 7 species.

228 **3.2. Relative abundance and occurrence**

229 The main species with high relative abundance (RA) of catch were *Calathus*
230 *circumseptus* (21.9%), *Cataglyphis biskrense* (15.4%), *Tetramorium biskrensis* (6%),
231 *Zabrus* sp. (3.9%), *Anomala dubia* (3.9%), *Scarites laevigatus* (3.7%) and *Carabus* sp.
232 (3%), respectively. Furthermore, families that dominated in terms of catches belonged
233 to Coleoptera and Hymenoptera, including Formicidae with a total of 111 individuals
234 (25.6%), Carabidae with 95 individuals (21.9%), Carabidae with 53 individuals
235 (12.2%) and Scarabeidae with 40 individuals (9.2%), and Apidae with 29 individuals
236 the equivalent of 6.0% of total catches (Table 1).

237 Regarding spatial occurrence of insect species at the eight sampled stations,
238 almost all species (66 species) were accidental and very accidental, nevertheless
239 three species were constant ($\text{Occ} \geq 50\%$) during the study period: *Chlaenius*
240 *circumseptus* (Callistidae), *Cataglyphis bicolor* (Formicidae) and *Tetramorium*
241 *biskrensis* (Formicidae). Common species ($\text{Occ} = 25\text{--}50\%$) were characterized by six
242 species: *Scolia* sp. (Scoliidae), *Apis mellifera* (Apidae) *Zabrus* sp. (Carabidae)
243 *Carabus* sp. (Carabidae) *Scarites laevigatus* (Carabidae) *Forficula auricularia*
244 (Forficulidae) (Table 1).

245 **3.3. Spatiotemporal composition and diversity**

246 The sampled station located southern Sabkha Djendli possessed the highest
247 values of catch seize (93 individuals, $RA=21.4\%$), species richness (27 species) and
248 the ratio Ni/SR (3.4). Whereas the highest values of Shannon index and evenness
249 were respectively recorded at station of West, Southeast, South, and East. Although

250 this later station (East) had the lowest values of insect composition ($N_i=43$,
251 $RA=9.9\%$, $SR=19$).

252 As for seasons, values of diversity parameters of insect assemblages were
253 higher during spring and autumn, with a slight leaning to spring values. However,
254 the summer scored the lowest values. Overall, sampling insects using pitfall traps at
255 Sabkha Djendli revealed a diversity equals to 4.7 bits according to Shannon index
256 with an evenness of 76% (Table 2).

257 Despite the differences in insect abundances between seasons and study plot
258 orientations, analysis of variance revealed no significant differences for the various
259 taxonomic orders of identified insects based on two factors 'Orientation' and
260 'Season' (Table 3).

261 **3.4. Spatial similarities of the entomofauna**

262 The assessment of similarities of insect assemblage compositions between the
263 sampled stations based on species richness revealed low similarities ranging between
264 12.8 and 35.3%. The highest values of Jaccard index ($C_j=35.3\%$) were observed
265 between North and West stations (Table 4).

266 According to values of Jaccard's index, the eight sampled stations were clustered
267 into four different groups: (i) the first group gathered all stations located at South,
268 East and West of the Sabkha including E, S, W, SE and NE, (ii) the SW station was
269 distinguished alone and (iii and iv) the third and the fourth class represented by NW
270 and N, respectively (Figure 2).

271 **3.5. Structure and diversity of functional trophic groups**

272 Predators and saprophagous held the highest catch rates with 37.3% and 26.7%
273 of the total, respectively. Predators were more pronounced in south stations (42 ind.)
274 especially in autumn (74 ind.) and spring (62 ind.), while saprophagous are
275 concentrated in southwest (24 ind.) and south (22 ind.) stations during the summer
276 (46 ind.).

277 In terms of numbers of species, phytophagous were the most abundant with 30
278 species distributed almost equally along seasons and sampled stations. As for *Ni/SR*
279 ratio, it varied between 1 and 11 with an average of 3.2 in study stations and
280 seasons, i.e. that each species of a given trophic group comprises an average of 3.2
281 individuals. This ratio is higher in saprophagous with 12.9, chiefly in stations of
282 south (11), northwest (8.5), southwest (8), during the summer (9.2) and spring (9).
283 Predators came in second place with 8.5 individuals per species.

284 The Shannon's index showed high diversity among phytophagous ($H'=4.3$) in
285 both seasons and sampled stations. The values of this index were lower among
286 predators, less important in polyphagous. Regarding saprophagous, the values
287 recorded the lowest rates. Similarly to evenness, where coprophagous (91%),
288 phytophagous (87%) and polyphagous (83%) showed higher values compared to
289 values of predators (68%) and saprophagous (36%). It is noteworthy that apart from
290 evenness, the coprophagous indicated the lowest values of ecological indices
291 calculated for different trophic groups of insects.

292 The Chi-square test revealed a significant dependence for the distribution of the
293 number of individuals of trophic groups along the orientations ($\chi^2=80.62$, $P<0.001$)
294 and seasons ($\chi^2=24.57$, $P=0.002$) (Table 5). However, no significant dependence
295 was observed for the rest of the features (Species richness, *Ni/SR* ratio, Shannon
296 index, evenness) of trophic groups according to stations orientations and seasons.

297 Although the Chi-square test revealed a significant dependence for abundances
298 of trophic groups on site directions and seasons (Table 5). However, the ANOVA
299 showed that the abundance of each trophic group did not vary significantly
300 according to seasons nor station orientations, except for the variation of
301 polyphagous numbers between site orientations ($P=0.015$) (Table 6).

302 **3.6. Relationship between insect communities and vegetation**

303 The Eigenvalues of CCA applied for insect assemblages and vegetation
304 parameters in canonical axis 1 and 2 were high and explaining 65.55% and 34.45% of
305 constrained inertia, respectively. According to CCA, the density of polyphagous was
306 positively associated with vegetation cover, but this parameter had a negative
307 influence on the number of individuals of predators, Hymenoptera and Dermaptera,
308 especially in autumn at northeast, southeast, north and east stations. In addition,
309 coprophagous abundance was positively related with species richness of plants,
310 however, Diptera and Homoptera were located on the negative side of the axis
311 representing species richness of plants; and this in northwest, west, southwest and
312 west stations. The phytophagous and saprophagous as well as Coleoptera and
313 Orthoptera were also positively correlated with the both axes of the parameters of
314 vegetation, particularly in spring and summer seasons. Conversely, the two
315 parameters of vegetation negatively influenced on Heteroptera densities in north and
316 east stations (Figure 3).

317 All the obtained significant-correlations were positive with vegetation parameters.
318 These concerned the abundances of Dermaptera in connection with species richness
319 of plants; and Orthoptera and Coleoptera with vegetation cover and richness of
320 plants. As for the trophic groups of the entomofauna, the correlation test was
321 significant for the numbers of polyphagous, phytophagous and saprophagous vs.
322 vegetation cover on the one hand, and coprophagous, phytophagous and
323 Saprophagous vs. plant species richness on the other (Table 7).

324 **4. Discussion and conclusion**

325 Salt lakes offer exceptional conditions for ecological studies of aquatic
326 ecosystems, due to the frequency and intensity of changes in the biological
327 communities compared to freshwater ecosystems (Comin & Comin, 1992). This

328 feature is most notable in arid regions, so that these habitats are home to many
329 original and well-adapted life forms (Chenchouni, 2012a).

330 Out of all the conducted samples, the Sabkha of Djendli houses 75 insect species
331 related to 31 families and 7 orders. In terms of individual numbers caught, the
332 orders of Coleoptera and Hymenoptera dominate other insect orders, while
333 Dermaptera, Heteroptera, Homoptera and Diptera are slightly present with very
334 similar densities in different study stations. This distribution of the composition could
335 be attributed to the low dispersal ability of these insects, as well as the scarcity of
336 these categories (Cobos, 1987), but mostly to the ineffectiveness of pitfall traps to
337 capture flying insects since this type of trap is specifically designed for ground
338 arthropods (Spence & Niemelä, 1994).

339 Since saline environments in hot arid regions are characterized by large spatial
340 and temporal fluctuation of water level and salinity, community of inland insects can
341 be modelled either by the synergistic effect of several factors (abiotic and biotic)
342 that are related to these two parameters; or by the predominance of one factor over
343 others (e.g. vegetation parameters) (Vidal-Abarca et al., 2004; Velasco et al., 2006).
344 Moreover, the state of the composition of insect assemblage in inland saline
345 environments can be explained by the morphological and physiological adaptations
346 necessary to cope with the extreme and unpredictable conditions of these habitats
347 on one hand, and their life cycle, phenological adaptations and behaviour on the
348 other hand (Cloudsley-Thompson, 1975; Louw and Seely, 1982).

349 The study of variations in the frequency of abundance and occurrence of
350 different insect orders shows that the beetles represent the most abundant order
351 that appears regularly in different sampling stations during the study period. This
352 frequency is reflected by the presence of three constant species (Coleoptera and
353 Hymenoptera), six common species (Coleoptera, Hymenoptera and Dermaptera)
354 and 66 accidental species. This finding is in contrast to the observation made by

355 Boix et al. (2008) where it has been found that beetles are the most affected group
356 within insects of saline environments, while our results are similar to those of Vidal-
357 Abarca et al. (2004) who argue that in the salt wetlands of arid and semi-arid areas,
358 Coleoptera and Diptera were the most abundant groups because of their large
359 adaptation to critical and extreme conditions. It is well known that the beetles are
360 the most abundant and occurring insect group in nature (Bouchard et al., 2011). In
361 addition to their dominance in the animal kingdom, they are an important food
362 resource for consumers at different levels in the food web; thereby their number of
363 species represents a good biological indicator of habitat quality (Rainio & Niemelä,
364 2003; Sánchez-Fernández et al., 2006). Moreover, because of their sensitivity to
365 environmental modifications, they constitute a model of choice for assessing the
366 diversity of habitats (Haddad et al., 2009).

367 Regarding insect species richness, the highest value is recorded in the west and
368 south stations with 27 species. According to Neffar et al. (2014), these stations are
369 characterized by certain homogeneity in their floristic composition. These areas are
370 grazed and fertilized by dung they receive and therefore stimulate the development
371 of certain flowering herbaceous and thus attract more pollinators. While cattle dung
372 favor the abundance of coprophagous, mostly Scarabaeidae in our case. These
373 observations were confirmed by the CCA where we found that coprophagous density
374 was positively correlated with plant diversity, which was negatively associated with
375 west and south stations. The vegetation significantly affects the different trophic
376 groups (herbivores, parasitoids and predators) of the insect fauna living at the
377 herbaceous layer, through its floristic composition and functional diversity
378 (Koricheva et al. 2000), but also through the density of vegetation cover that creates
379 a microclimate for soil-dwelling species (Siemann, 1998). According to Haddad et al.
380 (2009), species richness of predators and herbivores is positively related to species
381 richness and plant biomass, without being affected by its composition. However, in

382 lentic ecosystems, high electrical conductivity causes a significant decline in the
383 abundance and taxonomic richness of macroinvertebrate fauna (Waterkeyn et al.,
384 2008; Carver et al., 2009).

385 Based on the values of the Shannon index and evenness, insect diversification is
386 well marked in the different stations and seasons, indicating a balance between the
387 number of sampled invertebrate populations, although it may be that the
388 constituent species of assemblages are generalists, adapting to most environmental
389 conditions, as suggested by Rainio & Niemelä (2003) and Montagna et al. (2012).

390 Furthermore, the dominance of accidental species (66/75) may be connected to
391 the sparse structure of vegetation of the Sabkha. Because the presence of dense
392 vegetation reduces predation against herbivores that therein also find abundant
393 food, but also reduces the antagonistic effect between predators (Finke & Denno,
394 2002); this is not the case with the open vegetation of Sabkha Djendli which is
395 characterized by a medium to low coverage (Neffar et al. 2014). Otherwise the same
396 type of structure and composition of vegetation cover are almost noted in arid and
397 semi-arid wetlands of Algeria and North Africa (Khaznadar et al., 2009; Chenchouni,
398 2012b). This particular pattern of species occurrences may also be explained by the
399 unpredictable environmental changes inciting species to the coexistence, and
400 consequently the increase of diversity (Chesson & Huntly, 1997; Piñero et al., 2011).
401 But generally, seasonality remains the primary determinant factor of invertebrate
402 diversity in any ecosystem (Wolda, 1988). Because the metabolism of poikilotherms
403 requires low investment in energy, making of these invertebrates highly effective
404 organisms for the survival in extreme environments (Heatwole, 1996). This explains
405 the significant variation in predator numbers between the studied seasons.

406 The study of trophic status of insect species reveals their affiliation to different
407 ranks of consumers and thus these species virtually occupy different levels in the
408 food web. Species richness decreases in the following order herbivores> predators>

409 polyphagous> saprophages> coprophagous with 40%, 25.3%, 13.3%, 12.0% and
410 9.3%, respectively. According Piñero et al. (2011), seasonal variations have profound
411 effects not only on the number of species, abundance and biomass of invertebrates
412 during different times of the year, but also on the trophic and functional structures
413 of communities. For his part, Siemann (1998) suggested that the diversity, quality
414 and/or composition of plant species can in their turn influence the diversity of higher
415 trophic levels, not only by changing the diversity of herbivores, parasites and
416 predators but also by affecting the quality of the food of herbivores and the ease
417 with which they can be captured. Therefore, the spatiotemporal variation in traits of
418 vegetation (composition and cover) between the eight stations and seasons (Neffar
419 et al., 2014) is the cause of the significantly uneven spatiotemporal distribution
420 (according to Chi-square test) of insect group densities. Indeed, the CCA has allowed
421 the characterization of insect assemblage responses to vegetation parameters.

422 The comparison of specific composition between different stations using the
423 Jaccard index shows low similarity values, commonly not exceeding 35%. This
424 similarity would find its explanation in the heterogeneity of ecological conditions for
425 this fauna, in particular the composition and structure of the sparse vegetation which
426 is based of halophytes including *Suaeda* spp. *Atriplex* spp. and *Salicornia* spp. (Neffar
427 et al., 2014), reflecting thus the degraded conditions prevailing on the
428 physicochemical properties of soil in which they grow (Khaznadar et al., 2009).
429 According to Baguette (1992), the inter and intra-specific competitions, predation and
430 parasitism regulate the spatial and temporal distribution of species and structure
431 communities. Also, the distribution of a given species is a dynamic phenomenon that
432 involves a set of extinction and recolonisation stages of local populations following
433 changes in environmental conditions. Even more so, several studies have shown that
434 changes in communities across habitats are influenced by environmental variables, in

435 particular the type of substrate (Ligeiro et al., 2010) and even the coarse organic
436 matter (Hepp & Melo, 2013).

437 The spatial variability of the insect fauna of Sabkha Djendli is related to the
438 combination of several factors, among others, the climate is critical to the
439 distribution of arid arthropods (Langlands et al., 2006), the reproductive potential
440 and dispersal capabilities (Thompson and Townsend, 2006), and environmental
441 heterogeneity may be a contributing factor to their low dispersion.

442 The halophytic belt of Sabkha Djendli have a high richness of insects especially
443 in spring and autumn, coinciding in part with their breeding period. As the recorded
444 species are mostly phytophagous, their number naturally increases in the spring
445 with the increase of plant diversity and vegetation cover, whereas the predators
446 generally depend on the availability of prey (Koricheva et al., 2000; Haddad et al.,
447 2009). This statement is supported by findings of the CCA where the abundant
448 insect groups (Coleoptera, phytophagous) were found linked to vegetation
449 parameters mainly in spring and summer.

450 Following this study, the use of pitfall traps in Sabkha Djendli revealed some
451 knowledge about the entomofauna. The insect community shows high taxonomic
452 richness and diversity in different stations and seasons. The composition of
453 functional trophic groups are closely related to diversity and vegetation cover. The
454 conservation of this biological heritage so rich but little known, non-invested and
455 generally underestimated by managers, can only be possible by improving and
456 deepening our knowledge about biodiversity including the functional communities in
457 relation with threatening factors and disturbances that affect the proper conduct of
458 their vital activities.

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Figure 1

Location of the sampled station and sites at Sabkha Djendli (Batna, Northeast Algeria) and sampling design of pitfall traps.

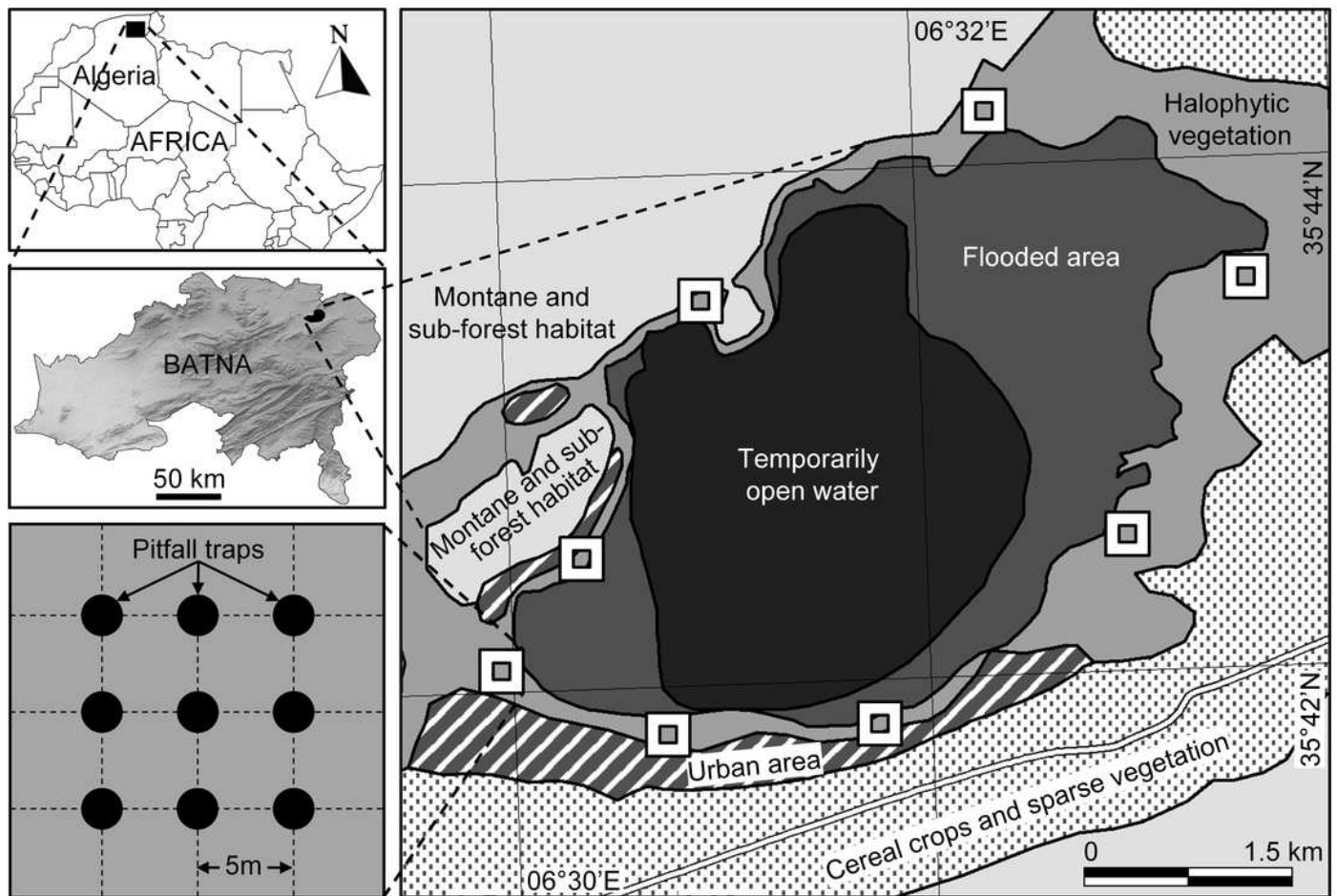


Figure 2

Dendrogram of agglomerative hierarchical clustering (AHC) illustrating species richness similarity (Jaccard coefficient) among insects captured from eight stations around Sabkha Djendli (linkage rule: unweighted pair-group average).

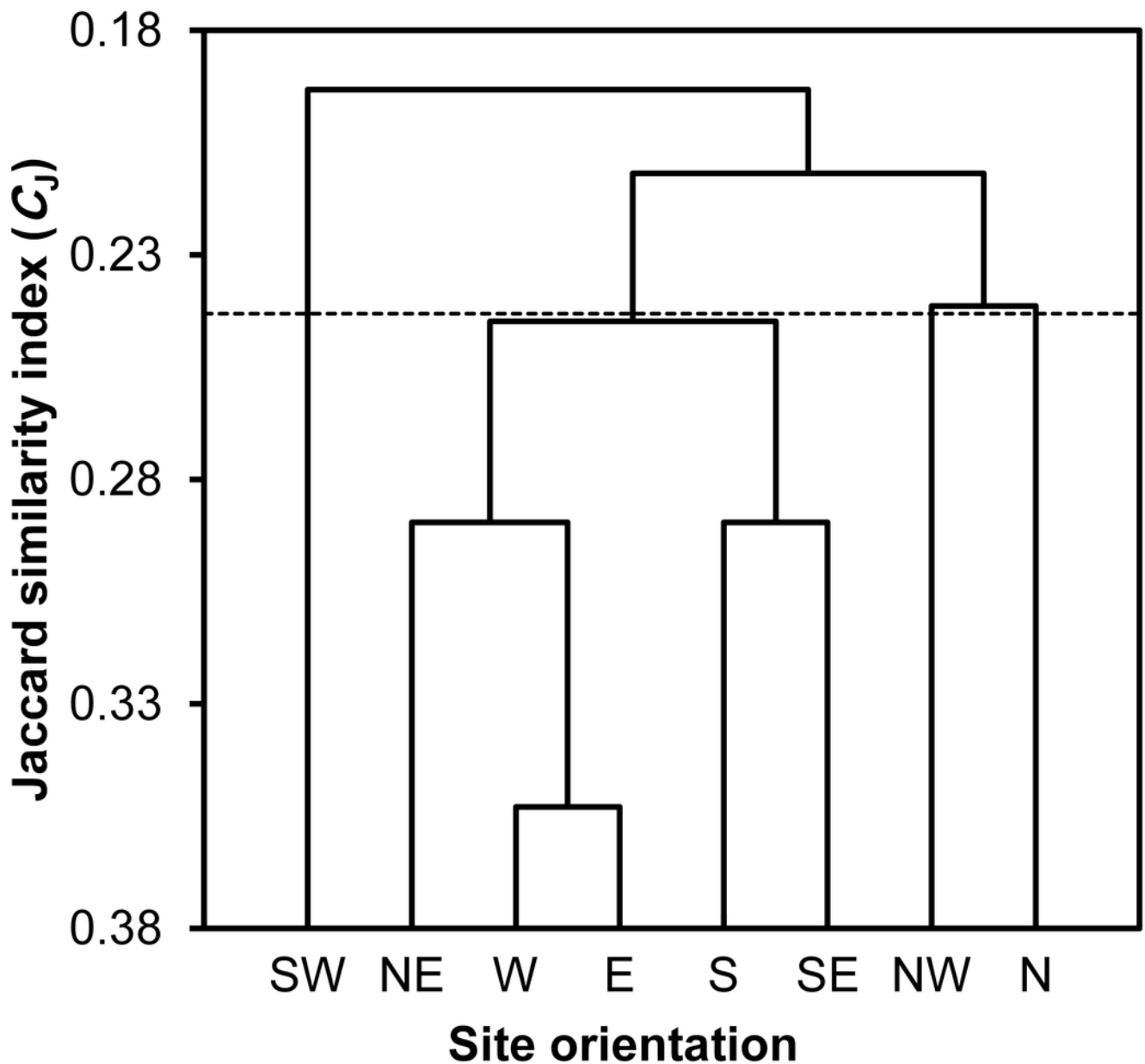


Figure 3

Diagram of the canonical correspondence analysis (CCA) relating spatial and seasonal densities of insect assemblages of both taxonomic orders and trophic groups with vegetation cover and species richness. (Cop: coprophagous, Phy: phytophagous, Pol: polyphagous, Pre: predator, Sap: saprophagous).

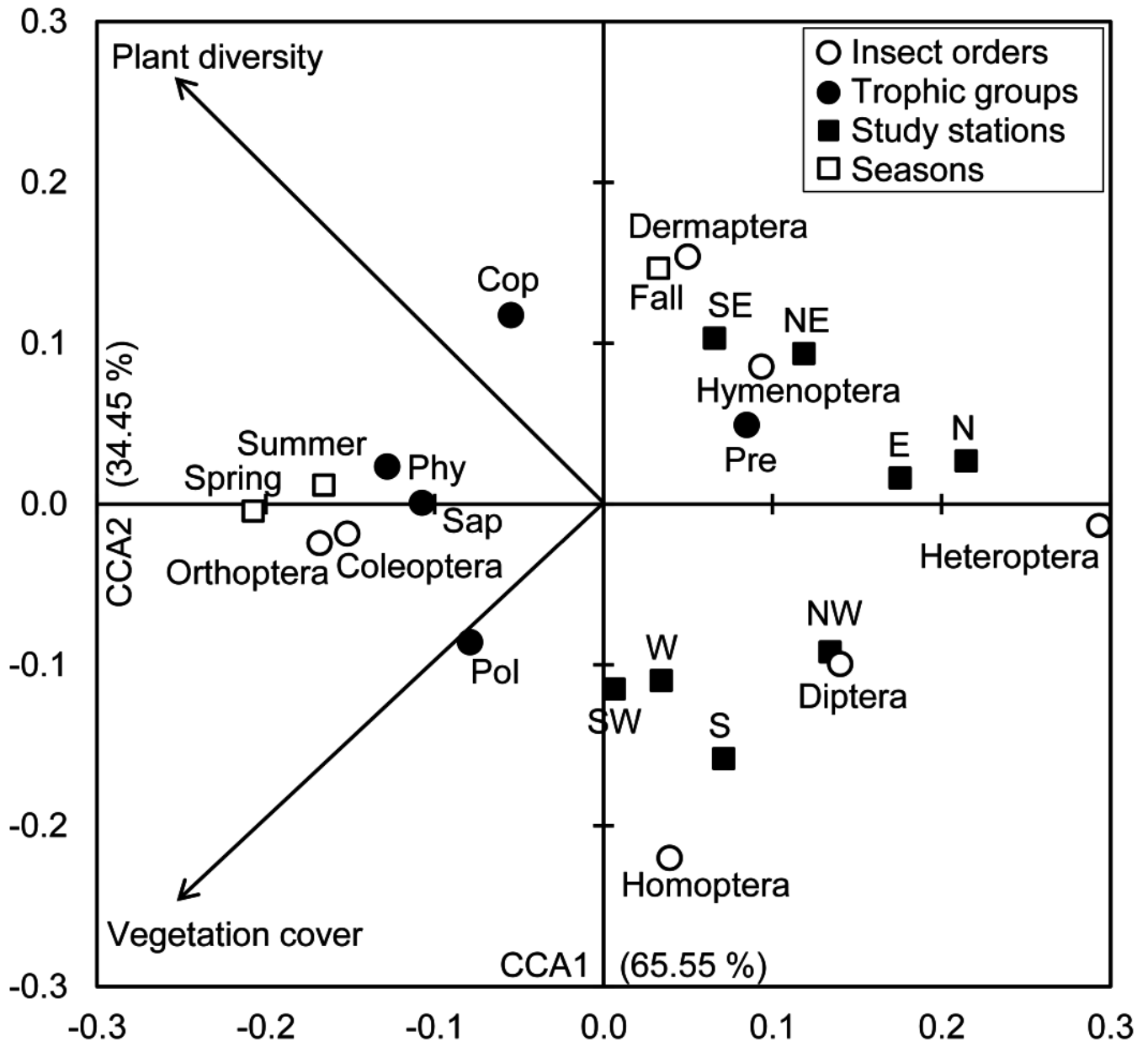


Table 1 (on next page)

Table 1

Systematic list, trophic status, abundances and occurrences of insect species captured using pitfall traps at edges of Sabkha Djendli, Northeast Algeria. (RA: relative abundance (%), FTG: functional trophic groups, Ni: total number of caught individuals, Occ: occurrence frequency, Cop: coprophagous, Phy: phytophagous, Pol: polyphagous, Pre: predator, Sap: Saprophagous, Vac: very accidental species, Acc: accidental species, Cmt: common species, Cst: constant species)

Classification (RA in %)	Species	FTG	Ni	RA	Occ	Scale
O: DERMAPTERA (3.2)						
F: Forficulidae (3.2)	<i>Anisolabis mauritanicus</i>	Pol	4	0.92	16.7	Acc
	<i>Forficula auricularia</i>	Pol	10	2.30	25.0	Cmt
O: ORTHOPTERA (5.1)						
F: Gryllidae (2.1)	<i>Acheta domesticus</i>	Phy	3	0.69	12.5	Acc
	<i>Gryllus bimaculatus</i>	Phy	2	0.46	8.3	Vac
	<i>Gryllus campestris</i>	Phy	2	0.46	8.3	Vac
	<i>Gryllus</i> sp.	Phy	2	0.46	8.3	Vac
F: Acrididae (3.0)	<i>Acrotylus patruelis</i>	Phy	3	0.69	12.5	Acc
	<i>Calliptamus barbarus</i>	Phy	4	0.92	12.5	Acc
	<i>Ephippiger</i> sp.	Phy	1	0.23	4.2	Vac
	<i>Oedipoda fuscocincta</i>	Phy	3	0.69	12.5	Acc
	<i>Sphingonotus rubescens</i>	Phy	1	0.23	4.2	Vac
	<i>Sphingonotus</i> sp.	Phy	1	0.23	4.2	Vac
O: HETEROPTERA (0.2)						
F: Lygaeidae (0.2)	<i>Lygaeus sexatilis</i>	Phy	1	0.23	4.2	Vac
O: HOMOPTERA (0.2)						
F: Cicadellidae (0.2)	<i>Cicadela variabilis</i>	Phy	1	0.23	4.2	Vac
O: COLEOPTERA (54.8)						
F: Cicindelidae (0.7)	<i>Calomera littoralis</i>	Pre	1	0.23	4.2	Vac
	<i>Cassolaia maura</i>	Pre	2	0.46	8.3	Vac
F: Callistidae (21.9)	<i>Calathus circumseptus</i>	Sap	95	21.89	87.5	Cst
F: Carabidae (12.2)	<i>Calathus</i> sp.	Pre	1	0.23	4.2	Vac
	<i>Macrothorax morbillosus</i>	Pre	1	0.23	4.2	Vac
	<i>Carabus</i> sp.	Pre	13	3.00	37.5	Cmt
	<i>Scarites laevigatus</i>	Pre	16	3.69	41.7	Cmt
	<i>Scarites</i> sp.	Pre	5	1.15	16.7	Acc
	<i>Zabrus</i> sp.	Phy	17	3.92	33.3	Cmt
F: Geotrupidae (0.9)	<i>Geotrupes</i> sp.	Sap	4	0.92	16.7	Acc
F: Scarabaeidae (9.2)	<i>Geotrogus</i> sp.	Sap	7	1.61	16.7	Acc
	<i>Anomala dubia</i>	Pol	17	3.92	20.8	Acc
	<i>Bubas bison</i>	Cop	4	0.92	16.7	Acc
	<i>Gymnopleurus flagellatus</i>	Cop	4	0.92	16.7	Acc
	<i>Onthophagus taurus</i>	Cop	5	1.15	20.8	Acc
	<i>Oxythyrea funesta</i>	Phy	1	0.23	4.2	Vac
	<i>Scarabaeus sacer</i>	Cop	1	0.23	4.2	Vac
	<i>Scarabaeus</i> sp.	Cop	1	0.23	4.2	Vac
F: Silphidae (0.2)	<i>Silpha opaca</i>	Pre	1	0.23	4.2	Vac
F: Staphylinidae (0.7)	<i>Staphylinus olens</i>	Pol	3	0.69	12.5	Acc
F: Cetoniidae (0.5)	<i>Cetonia ablonga</i>	Phy	1	0.23	4.2	Vac
	<i>Cetonia funeraria</i>	Phy	1	0.23	4.2	Vac
F: Cantharidae (0.2)	<i>Cantharis</i> sp.	Phy	1	0.23	4.2	Vac
F: Meloidae (1.4)	<i>Mylabris crocata</i>	Phy	1	0.23	4.2	Vac
	<i>Mylabris quadripunctata</i>	Phy	2	0.46	8.3	Vac
	<i>Mylabris variabilis</i>	Phy	3	0.69	12.5	Acc
F: Tenebrionidae (2.3)	<i>Adesmia microcephala</i>	Sap	1	0.23	4.2	Vac
	<i>Blaps mortisaga</i>	Sap	1	0.23	4.2	Vac
	<i>Blaps nitens</i>	Sap	1	0.23	4.2	Vac
	<i>Opatrum</i> sp.	Sap	2	0.46	8.3	Vac
	<i>Tentyria bipunctata</i>	Sap	1	0.23	4.2	Vac
	<i>Tentyria</i> sp.	Sap	4	0.92	12.5	Acc
F: Dermestidae (1.2)	<i>Dermestes</i> sp.	Cop	1	0.23	4.2	Vac
	<i>Trogoderma</i> sp.	Cop	4	0.92	16.7	Acc

F: Cucujidae (0.7)	<i>Canthartus</i> sp.	Pol	3	0.69	8.3	Vac
F: Curculionidae (0.5)	<i>Coniocleonus excoriatus</i>	Phy	1	0.23	4.2	Vac
	<i>Lixus punctiventris</i>	Phy	1	0.23	4.2	Vac
F: Chrysomelidae (2.3)	<i>Chrysomela</i> sp.	Phy	9	2.07	16.7	Acc
	<i>Entomoscelis</i> sp.	Phy	1	0.23	4.2	Vac
<hr/>						
O: HYMENOPTERA (34.3)						
F: Formicidae (25.6)	<i>Camponotus</i> sp.	Pre	1	0.23	4.2	Vac
	<i>Cataglyphis bicolor</i>	Pre	67	15.44	58.3	Cst
	<i>Messor barbarous</i>	Pre	7	1.61	12.5	Acc
	<i>Pheidole pallidula</i>	Pol	1	0.23	4.2	Vac
	<i>Tapinoma nigerrimum</i>	Pre	9	2.07	20.8	Acc
	<i>Tetramorium biskrense</i>	Pre	26	5.99	50.0	Cst
F: Vespidae (0.2)	<i>Polistes gallicus</i>	Pre	1	0.23	4.2	Vac
F: Apidae (4.9)	<i>Apis mellifera</i>	Phy	10	2.30	29.2	Cmt
	<i>Apis</i> sp.	Phy	5	1.15	16.7	Acc
	<i>Bombus pascuorum</i>	Phy	2	0.46	4.2	Vac
Megachilidae (1.2)	<i>Megachile</i> sp.	Phy	5	1.15	16.7	Acc
Halictidae (0.9)	<i>Sphecodes</i> sp.	Phy	1	0.23	4.2	Vac
	<i>Halictus</i> sp.	Phy	3	0.69	12.5	Acc
F: Scoliidae (1.6)	<i>Scolia</i> sp.	Pre	7	1.61	25.0	Cmt
F: Sphecidae (0.7)	<i>Ammophila hirsute</i>	Pre	1	0.23	4.2	Vac
	<i>Ammophila sabulosa</i>	Pre	1	0.23	4.2	Vac
	<i>Sphex funerarius</i>	Pre	1	0.23	4.2	Vac
F: Mutillidae (0.2)	<i>Mutilla</i> sp.	Pre	1	0.23	4.2	Vac
<hr/>						
O: DIPTERA (2.1)						
F: Tabanidae (0.5)	<i>Tabanus</i> sp.	Pol	2	0.46	8.3	Vac
F: Muscidae (0.7)	<i>Musca domestica</i>	Pol	1	0.23	4.2	Vac
	<i>Musca</i> sp.	Pol	2	0.46	8.3	Vac
F: Sarcophagidae (0.9)	<i>Sarcophaga</i> sp.	Pol	4	0.92	16.7	Acc

Table 2 (on next page)

Table 2

Spatial and seasonal variation of the diversity parameters of insect assemblages in Sabkha Djendli, Northeast Algeria

Parameter	Orientation								Season		Total	
	S	SW	W	NW	N	NE	E	SE	Spring	Summer	Fall	
Abundances (<i>Ni</i>)	93	48	60	44	52	50	43	45	163	116	155	434
RA (%)	21.	11.	13.	10.	11.	11.	9.	10.			35.	
Species richness	4	0	8	1	8	5	9	3	37.5	26.7	6	100
(<i>SR</i>)	27	24	27	19	17	22	19	22	46	38	51	75
Ratio <i>Ni</i> / <i>SR</i>	3.4	2.0	2.2	2.3	3.1	2.3	3	2.0	3.5	3.1	3.0	5.8
Shannon index (<i>H'</i>)	3.9	3.6	4.1	3.5	3.1	3.8	7	3.9	4.5	4.0	4.6	4.7
Evenness (%)	81	79	86	83	77	84	86	87	82	76	82	76

Table 3 (on next page)

Table 3

ANOVAs testing the variances in the abundance of insect orders according to site orientations and study seasons in Sabkha Djendli, Northeast Algeria

Orders	Variations	<i>Df</i>	<i>MS</i>	<i>F</i>	<i>P</i>
Dermaptera	Season	2	0.17	0.17	0.845
	Orientation	7	1.12	1.15	0.390
	Residuals	14	0.98		
Orthoptera	Season	2	1.04	0.92	0.423
	Orientation	7	0.55	0.48	0.832
	Residuals	14	1.14		
Heteroptera	Season	2	0.04	1.00	0.393
	Orientation	7	0.04	1.00	0.471
	Residuals	14	0.04		
Homoptera	Season	2	0.04	1.00	0.393
	Orientation	7	0.04	1.00	0.471
	Residuals	14	0.04		
Coleoptera	Season	2	15.04	0.60	0.561
	Orientation	7	29.21	1.17	0.378
	Residuals	14	24.95		
Hymenoptera	Season	2	54.29	1.56	0.246
	Orientation	7	47.23	1.35	0.298
	Residuals	14	34.91		
Diptera	Season	2	0.50	1.11	0.358
	Orientation	7	0.33	0.72	0.655
	Residuals	14	0.45		

Table 4(on next page)

Table 4

Similarity matrix of insect assemblages in the studied sites at Sabkha Djendli, Northeast Algeria. The values are referred to Jaccard similarity index " C_j " (under the diagonal) and the number of shared species (above the diagonal). Station locations "orientations" of the first row are associated in the first column with species richness (SR) values

Orientations	S	SW	W	NW	N	NE	E	SE
SE (22)	11	7	9	6	8	8	6	
E (19)	9	7	12	7	6	10		17.1
NE (22)	11	8	10	8	6		22.2	32.3
N (17)	7	5	5	7		25.8	20.0	18.2
NW (19)	10	8	9		17.1	22.6	24.2	24.1
W (27)	13	8		22.5	35.3	25.6	12.8	24.3
SW (24)	9		17.9	19.4	21.1	13.9	22.9	18.6
S (27)		28.9	24.3	28.9	18.9	27.8	31.7	21.4

Table 5 (on next page)

Table 5

Spatial and seasonal variations of insect trophic guilds living in Sabkha Djendli, Northeast Algeria

Parameter	Orientations								Seasons			Total
	S	SW	W	NW	N	NE	E	SE	Sprin g	Summe r	Fall	
Individual numbers	$(\chi^2_{28}=80.62, P<0.001)$								$(\chi^2_8=24.57, P=0.002)$			
Coprophagous	4	1	4	1	0	3	3	4	8	5	7	20 (4.6%)
Phytophagous	20	10	13	7	8	6	11	14	35	26	28	89 (20.5%)
Polyphagous	4	3	19	7	5	4	2	3	22	13	12	47 (10.8%)
Predator	42	10	21	12	28	21	15	13	62	26	74	162 (37.3%)
Saprophagous	22	24	3	17	11	16	12	11	36	46	34	116 (26.7%)
Species richness	$(\chi^2_{28}=15.65, P=0.971)$								$(\chi^2_8=4.81, P=0.777)$			
Coprophagous	4	1	3	1	0	2	3	3	4	4	5	7 (9.3%)
Phytophagous	8	9	8	7	6	6	6	9	19	14	15	30 (40%)
Polyphagous	4	3	5	4	3	2	2	2	9	5	8	10 (13.3%)
Predator	8	8	8	5	5	8	4	6	10	9	15	19 (25.3%)
Saprophagous	2	3	3	2	3	4	4	2	4	5	8	9 (12%)
Ratio N_i/SR	$(\chi^2_{28}=22.09, P=0.779)$								$(\chi^2_8=3.35, P=0.911)$			
Coprophagous	1.0	1.0	1.3	1.0	0.0	1.5	1.0	1.3	2.0	1.3	1.4	2.9
Phytophagous	2.5	1.1	1.6	1.0	1.3	1.0	1.8	1.6	1.8	1.9	1.9	3.0
Polyphagous	1.0	1.0	3.8	1.8	1.7	2.0	1.0	1.5	2.4	2.6	1.5	4.7
Predator	5.3	1.3	2.6	2.4	5.6	2.6	3.8	2.2	6.2	2.9	4.9	8.5
Saprophagous	11.0	8.0	1.0	8.5	3.7	4.0	3.0	5.5	9.0	9.2	4.3	12.9
Shannon's index	$(\chi^2_{28}=13.18, P=0.992)$								$(\chi^2_8=0.58, P=0.999)$			
Coprophagous	2.0	0.0	1.5	0.0	0.0	0.9	1.6	1.5	1.9	1.9	2.2	2.5
Phytophagous	2.5	3.1	2.7	2.8	2.4	2.6	2.2	2.8	3.8	3.5	3.7	4.3
Polyphagous	2.0	1.6	1.2	1.8	1.4	0.8	1.0	0.9	2.6	2.0	2.9	2.7
Predator	2.5	2.9	2.8	2.1	1.3	2.4	1.7	2.0	2.5	2.4	2.9	2.9
Saprophagous	0.3	0.9	1.6	0.5	1.1	1.2	1.2	0.4	0.9	0.7	1.6	1.2
Evenness (%)	$(\chi^2_{28}=17.71, P=0.999)$								$(\chi^2_8=3.85,$			

P=0.999)

Coprophagous	100	0	95	0	0	92	¹⁰ ₀	95	95	96	96	91
Phytophagous	85	98	88	100	93	¹⁰ ₀	86	89	90	92	93	87
Polyphagous	100	¹⁰ ₀	50	92	86	81	¹⁰ ₀	92	80	88	95	83
Predator	83	97	92	92	56	80	84	79	76	75	75	68
Saprophagous	27	56	¹⁰ ₀	52	69	59	60	44	47	30	52	36

Table 6 (on next page)

Table 6

Two-way ANOVAs testing the variation of abundance of insect trophic guilds between seasons and station orientations in Sabkha Djendli, Northeast Algeria.

Trophic groups	Variations	<i>Df</i>	<i>MS</i>	<i>F</i>	<i>P</i>
Coprophagous	Season	2	0.29	0.32	0.731
	Orientation	7	0.86	0.94	0.507
	Residuals	14	0.91		
Phytophagous	Season	2	2.79	0.64	0.542
	Orientation	7	6.90	1.58	0.220
	Residuals	14	4.36		
Polyphagous	Season	2	3.79	1.46	0.266
	Orientation	7	10.14	3.90	0.015
	Residuals	14	2.60		
Predator	Season	2	78.00	3.07	0.079
	Orientation	7	37.50	1.47	0.254
	Residuals	14	25.43		
Saprophagous	Season	2	5.17	0.51	0.614
	Orientation	7	15.14	1.48	0.251
	Residuals	14	10.21		

Table 7 (on next page)

Table 7

Pearson correlation tests between abundance of orders and trophic guilds of insects and spontaneous vegetation characteristics (cover and species richness) of Sabkha Djendli, Northeast Algeria

Variables	Vegetation cover		Plant species richness	
	<i>r</i>	<i>P</i>	<i>r</i>	<i>P</i>
Taxonomic orders				
Dermaptera	0.214	0.527	0.613	0.045
Orthoptera	0.717	0.013	0.796	0.003
Heteroptera	-0.030	0.930	0.109	0.750
Homoptera	0.591	0.055	0.295	0.378
Coleoptera	0.656	0.028	0.742	0.009
Hymenoptera	0.215	0.526	0.505	0.113
Diptera	0.308	0.357	0.257	0.445
Trophic groups				
Coprophagous	0.429	0.188	0.786	0.004
Phytophagous	0.586	0.048	0.748	0.008
Polyphagous	0.604	0.049	0.571	0.066
Predator	0.262	0.437	0.488	0.128
Saprophagous	0.591	0.046	0.717	0.013