

1 **Using life history trade-offs to understand core-transient structuring of a small mammal**  
2 **community**

3  
4 Sarah R. Supp<sup>1,3\*</sup>, David N. Koons<sup>2</sup>, and S. K. Morgan Ernest<sup>1</sup>

5  
6  
7 <sup>1</sup> Department of Biology and the Ecology Center, Utah State University, Logan, UT, 84322 USA

8 <sup>2</sup> Department of Wildland Resources and the Ecology Center, Utah State University, Logan, UT,  
9 84322 USA

10  
11  
12  
13  
14  
15  
16  
17  
18  
19  
20  
21 <sup>3</sup> Current Address: School of Biology and Ecology, University of Maine, Orono, ME, 04469  
22 USA

23 \*Corresponding author: [sarah@weecology.org](mailto:sarah@weecology.org)

**24 Abstract**

25 An emerging conceptual framework suggests that communities are composed of two  
26 main groups of species through time: core species that are temporally persistent, and transient  
27 species that are temporally intermittent. Core and transient species have been shown to differ in  
28 spatiotemporal turnover, diversity patterns, and importantly, survival strategies targeted at local  
29 vs. regional habitat use. While the core-transient framework has typically been a site-specific  
30 designation for species, we suggest that if core and transient species have local vs. regional  
31 survival strategies across sites, and consistently differ in population-level spatial structure and  
32 gene flow, they may also exhibit different life-history strategies. Specifically, core species  
33 should display relatively low movement rates, low reproductive effort, high ecological  
34 specialization and high survival rates compared to transient species, which may display a wider  
35 range of traits given that transience may result from source-sink dynamics or from the ability to  
36 emigrate readily in a nomadic fashion. We present results from 21 years of capture-mark-  
37 recapture data in a diverse rodent community, evaluating the linkages between temporal  
38 persistence, local abundance, and trade-offs among life-history traits. Core species at our site  
39 conservatively supported our hypotheses, differing in ecological specialization, survival and  
40 movement probabilities, and reproductive effort relative to transient species. Transient species  
41 exhibited a wider range of characteristics, which likely stems from the multiple processes  
42 generating transience in local communities, such as source-sink dynamics at larger regional  
43 scales or nomadic life history strategies. We suggest that trait associations among core-transient  
44 species may be similar in other systems and warrants further study.

45

46 **Keywords:** Core, dispersal, life-history traits, movement, small mammal, source-sink, temporal  
47 persistence, trade-offs, transient

48

## 49 **Introduction**

50 A pervasive characteristic of ecological communities is that they tend to be composed of a  
51 few common species and many rare ones. Traditionally, ecologists have assumed that particular  
52 structuring processes are equally important for explaining patterns of abundance among all  
53 species in a community. Alternately, species abundances have been modeled as a product of  
54 spatially linked metapopulations or metacommunities (e.g., Hanski 1982; Gaston and Lawton  
55 1989; Gotelli 1991; Urban and Skelly 2006). An emerging view posits that species occurring at a  
56 site can be sorted into two different groups that should be modeled differently based on temporal  
57 patterns of occupancy (Magurran and Henderson 2003): core species that display high temporal  
58 persistence (i.e. present most years) and transient species that exhibit low temporal persistence  
59 (i.e. present only occasionally). Core species tend to account for most of the abundance or  
60 biomass within a particular community. The richness of core species is better predicted by local  
61 environmental conditions, presumably reflecting the need for strong local adaptation to  
62 outcompete other species and maintain highly abundant, persistent populations in a specific  
63 habitat (Ulrich and Ollik 2004; Belmaker 2009; Coyle et al 2013). In contrast, transient species  
64 tend to be rare, and their richness in a local community is better predicted by regional factors (e.g.  
65 spatial heterogeneity, regional species pool), presumably because dispersal is a critical process  
66 allowing those species to immigrate into particular habitats (Costello & Myers 1996; Milstead et  
67 al. 2007; Henderson and Magurran 2014). Because core species at a site require a strong match  
68 between their niche requirements and local environmental conditions to maintain populations

69 whereas immigration is the primary process governing the presence of transient species, the core-  
70 transient framework suggests that the spatial and temporal scale of processes influencing the two  
71 groups should differ.

72         Although transient species move into a local site from other habitat patches on the  
73 landscape, and thus suggests the importance of habitat connectivity and linked populations within  
74 a given time frame, the core-transient framework is fundamentally distinct from its spatial analog,  
75 the core-satellite framework (Hanski 1982). The core-transient framework assigns species based  
76 on temporal persistence at a single site (e.g., the fraction of years where present; Magurran and  
77 Henderson 2003; Ullrich and Ollik 2004, Vergnon et al. 2009; Coyle et al. 2013), whereas the  
78 core-satellite framework assigns species based on spatial patterns of occurrence across a regional  
79 metapopulation (e.g., the fraction of sites in the region where present; Hanski 1982). For example,  
80 a species could be identified as regionally “core” because it inhabits a high fraction of habitat  
81 patches, but be identified as temporally “transient” at a single site that it inhabits infrequently  
82 across a time-series. Here, we use “core” and “transient” to refer to species’ occupancy within a  
83 time-series of a single site.

84         Core-transient research has primarily focused on differences between temporally defined  
85 groups at a site in how they respond to the same environment. However, if core and transient  
86 species differ in the importance of local and regional processes in maintaining populations at a  
87 site, they may also differ in the ecological and evolutionary drivers ultimately determining their  
88 presence, abundance and diversity (Magurran and Henderson 2003; Dolan *et al.* 2009, Coyle *et al.*  
89 *al.*2013, Connolly et al. 2014). While little research has focused on this question *per se*, a rich  
90 literature on metapopulations, metacommunities, and the evolution of dispersal suggest that core  
91 and transient species could indeed experience different pressures that would select for different

92 ecologies or life history strategies. Core species must successfully compete in, and adapt to, their  
93 local biotic and abiotic environment. As such, core species that are strongly governed by local  
94 ecological processes experience strong local co-evolutionary pressures with their biotic and  
95 abiotic environment (McCauley 2007). Species that are highly adapted to a particular  
96 environment can evolve reduced dispersal tendencies in part because of the cost of moving into a  
97 maladaptive environment (Kisdi 2002). Limited gene flow through reduced dispersal among  
98 populations enhances the role of local natural selection and adaptation for core species (Hanski  
99 1982, McPeck and Holt 1992; Kisdi 2002; Urban *et al.* 2008). By focusing on using local habitat  
100 and reducing the tendency to disperse, core species must also be able to persist through acclimate  
101 environmental conditions. This likely involves specialized behaviors or traits to allow persistence  
102 under poor conditions (Hanski 1982, Clutton-Brock 1991, Ghalambor & Martin 2000, Kisdi  
103 2002).

104         The ecological and evolutionary pressures on transient species are more complicated.  
105 Although there is only one way to be core, there are two primary processes that may generate  
106 transient species within a local community, each with different resulting expectations of the eco-  
107 evolutionary pressures imposed. One process that can generate transients in a local community is  
108 source-sink dynamics (e.g., Hanski 1982). Source-sink transients are species that are core at other  
109 sites on the landscape through time, but not in the focal site. Because they are well-adapted  
110 elsewhere and immigration from the adapted source population should swamp local adaptation to  
111 the new environment (Holt and Gaines 1992; Lenormand 2002; Kawecki and Holt 2002; Kawecki  
112 2008), we would expect source-sink transients to retain many of the same general life history  
113 characteristics as core species – except with low temporal occupancy.

114 In contrast to source-sink transients, some transients may form short-term reproducing  
115 populations before disappearing and potentially reappearing later via immigration. Processes that  
116 could generate this type of transient include opportunistic tracking of shifting resources across a  
117 region (e.g. Weins 2001; Roshier and Reid 2003; Milstead et al. 2007; Jonzen et al. 2011) or  
118 competition-colonization tradeoffs (e.g., Diamond 1974; Kneitel and Chase 2004). Both cases  
119 create what we will call nomadic transients – individuals that form reproducing populations for  
120 short periods of time in a local habitat before moving elsewhere on the landscape. Nomadism  
121 results from irregular, non-sedentary behavior, and differs from migration in that movements may  
122 be made at any time of year following any path (e.g., no fixed breeding grounds, low home range  
123 stability), and may occupy a large portion of the animal's life (Baker 1978; Roshier and Reid  
124 2003). Nomadic transients are strongly governed by regional ecological processes with high  
125 levels of movement across the landscape that reflects demographic processes or spatial  
126 rearrangement in response to changing conditions though time (O'Donnell 2001; Roshier and  
127 Reid 2003).

128 Increased individual movement may impact the long-term evolutionary dynamics of  
129 nomadic transient species if: 1) high gene flow homogenizes gene pools and inhibits local  
130 adaptation (Urban *et al.* 2008) or 2) intermediate gene flow increases the capacity for local  
131 adaptation in unstable habitats via novel subsidies from the regional gene pool (Roshier and Reid  
132 2003; Urban and Skelly 2006; Loeuille and Leibold 2008). Both the high gene flow and  
133 intermediate gene flow scenarios suggest that on average transient species should be less adapted  
134 to local biotic and abiotic conditions than core species. Because they are less well adapted to local  
135 conditions, they are expected to be at a competitive disadvantage, except for their ability to  
136 exploit novel conditions in unstable environments. If nomadic transient species depend on a

137 regional life history strategy that requires the ability to track suitable environmental conditions  
138 and use heterogeneous landscapes, then they should be strongly associated with traits that enable  
139 them to traverse non-ideal habitat patches and to colonize new suitable habitat patches as they  
140 arise (McCauley 2007). Increased dispersal ability and movement rates come with costs that may  
141 include increased mortality risk and increased time and energetic expenditure (Murray 1967;  
142 Hanski 1982; Waser 1985; Rousset and Gandon 2002). To maximize fitness amidst such costs, it  
143 may be optimal for nomadic transient species to trade-off survival investment toward increased  
144 reproductive allocation (*e.g.*, Clutton-Brock 1991; Stearns 1992) and they may have thus evolved  
145 associations with life-history traits such as low survival probability, high fecundity, early age of  
146 primiparity, and resource or habitat generalism (Diamond 1974; Charlesworth 1980).

147         The core-transient framework can thus provide a key connection of the slow-fast theory of  
148 life history evolution (Ricklefs and Wikelski 2002; Sibly and Brown 2007) with movement and  
149 habitat use strategies. From theoretical and empirical studies on dispersal, metapopulations, and  
150 metacommunities, we might expect a life history strategy for core species focused on local-scale  
151 coexistence to include low movement rates, low mortality rates, high immune function, limited  
152 gene flow among populations, high specialization for local conditions, and decreased fecundity.  
153 Source-sink transients should show similar trait correlations as core species, except that they  
154 should exhibit lower abundances and temporal persistence and have little genetic differentiation  
155 between source and sink populations (Gaggiotti 1996). For nomadic transients, we expect the  
156 opposite set of characteristics from core species. Nomadic species at a site should generally be  
157 expected to have higher movement rates, be able to reproduce in a wider range of habitats, exhibit  
158 increased fecundity, higher gene flow among populations, but lower survival (Mueller and Fagan  
159 2008; Baguette et al. 2013). While this possible link between the core-transient framework and

160 life-history traits is intriguing, there has been no empirical evaluation, especially examining  
161 correlations in traits across species within the same community.

162 We use 21-years of data from a diverse desert rodent community containing information  
163 on movement, mark-recapture rates, and reproduction to test aspects of the hypothesis that core  
164 species have fundamentally different life-history strategies than transient species as expected  
165 from differences in local vs. regional habitat use. We predict that core species will generally be  
166 associated with relatively low movement rates, high survival rates, and low fecundity. We predict  
167 that transient species will display a mixture of traits, depending on whether they are source-sink  
168 or nomadic transients, but generally have more incidence of high movement rates, low survival  
169 rates, high fecundity, and resource or habitat generalism (Figure 1).

170

## 171 **Materials and Methods**

### 172 *Study site and data*

173 We evaluated the relationship between life history traits and core-transient status at our  
174 site using 21 years (1989-2009) of capture-mark-recapture (CMR) data for rodents from a long-  
175 term experimental site in the Chihuahuan desert in southeastern Arizona (the Portal Project field  
176 site; Ernest *et al.* 2009). The study site consists of 24, 0.25 ha fenced plots (50 m X 50 m). Each  
177 month, year-round, plots are trapped on a grid consisting of 49 evenly spaced permanent stakes  
178 to survey the rodent community and to maintain experimental treatments. Four gates cut into  
179 each side of the fenced plots allow free passage of rodents in and out of plots. Large-bodied and  
180 behaviorally dominant kangaroo rats (*Dipodomys spp.*) have enlarged auditory bullae that make  
181 it possible to selectively exclude them from plots that have a smaller gate size (n=8). Total  
182 rodent removal plots have no gates (n=6), while control plots (n=10) have relatively large gates



183 that allow all species access (Brown 1998). Rodents are known to follow barriers until a path  
184 opens (Singleton et al. 1998), ensuring that plot gates are discovered, and there are not  
185 significant differences in species colonization of the plots compared to the natural landscape  
186 (Brown and Maurer 1985). Upon capture, each individual was marked by toe clipping, ear tags  
187 or with a permanent, subcutaneous passive integrated transponder (PIT) tag that allowed it to be  
188 uniquely identified upon capture. For each captured individual, we recorded species, sex,  
189 reproductive status, hind foot length, weight, and individual PIT tag. For our analysis, we right-  
190 censored data from individuals after the point that they were captured on total rodent removal  
191 plots, or from kangaroo rat individuals captured on kangaroo rat removal plots because these  
192 individuals were subsequently removed from the study site.

193 Since the small mammal community includes diverse species (n=21) representing a suite  
194 of different feeding guilds and life history strategies (Table 1), our site is ideal for evaluating  
195 certain traits associated with core and transient species in 3 main feeding guilds: granivores  
196 (n=15), folivores (n=4) and carnivores (n=2). At our site, species across the three guilds have a  
197 wide range of body sizes (4-280 g) and divergent evolutionary histories (Bininda-Edmonds et al.  
198 2007), leading to differing levels of adaptation to the arid environment, which results in them  
199 being differently suited to local and regional habitat use strategies. During 1989-1999,  
200 individuals were marked using ear and toe tags, and during 2000-2009, individuals were mainly  
201 marked with PIT tags. We conducted extensive data cleaning and error checking to ensure that  
202 potential problems in the data (*e.g.*, duplicate tags, uncertainty in sex or species) were resolved.  
203 In cases where the data with identical tags could be clearly partitioned into unique individuals,  
204 we assigned new unique tag numbers to each individual. In cases where data could not be clearly

205 partitioned into individuals, or where species identity was questionable, the data were excluded  
206 from analysis.

207 Core and transient species status was assigned based on temporal persistence, as defined  
208 by the proportion of years that each species was present (1989-2009) on control plots at our site.  
209 Species that were present in at least 2/3 of the years ( $\geq 0.66$ ) were considered core (*sensu* Coyle  
210 *et al.* 2013). Species that were present less than 1/3 of the years ( $\leq 0.33$ ) were considered  
211 transient. All other species were considered intermediate temporal status.

### 212 *Movement, Survival, and Fecundity*

213 Using individual-level recapture data, we assessed movement characteristics for each  
214 species. Here, we define individual movement at each time-step relative to the individual's  
215 previous capture location, and not in a geographic context defined by the individual's birthplace,  
216 burrow, or location relative to other individuals. We were unable to measure geographic  
217 dispersal (e.g., relative to natal location) using our dataset, but we assume that patterns of  
218 individual movement (including maintenance, breeding, foraging, and exploratory movements)  
219 are correlated with and encompassed by dispersal, which is an outcome of movement (Roshier  
220 and Reid 2003). Locations of the permanently marked trap stakes were recorded in 2010 using  
221 ProMark3 GPS Units with error  $< 2$  cm. We recorded the distance in meters between trap stakes  
222 among chronologically ordered capture histories for each individual. For each species, we binned  
223 the individual movement data by 6-meter increments that roughly represent the distance between  
224 stakes (with bin 1 representing distance 0-3 meters, or recapture at the same stake), and plotted  
225 the data in histograms. For each species, we calculated the modal distance and the mean plus one  
226 standard deviation of the log (Y+1) transformed data to determine a species-level benchmark at  
227 which each movement distribution transitions into long-distance movements. We chose this

228 transformation to meet the assumptions of normality and because there are many 0 m movements  
229 (Sokal and Rohlf 2012). For each species, these histograms provide insight into the frequency at  
230 which individuals move short versus long distances. Using the combined individual movement  
231 distances of the core species within each feeding guild (granivore, folivore, and carnivore), we  
232 set the mean plus one standard deviation of  $\log(Y+1)$  transformed data as our guild-level  
233 benchmark defining a short versus a long distance movement to compare across all species. We  
234 used this method because home range size likely differs based on trophic group (Mace and  
235 Harvey 1983).

236 To more thoroughly evaluate life-history relationships between movement and apparent  
237 survival, conditional on recapture probability, we used a multistate capture-mark-recapture  
238 (CMR) modeling approach in Program Mark version 7.0 (White and Burnham 1999, White and  
239 Cooch 2012) through the R programming environment 3.0.2 (R Core Development Team 2012)  
240 and package RMark (Laake and Rexstad 2008; Laake *et al.* 2012). To address our questions of  
241 whether core and transient species differ in apparent survival ( $\Phi$ ), recapture ( $p$ ), and movement  
242 probabilities ( $\psi$ ), we used a two-state model where all species were first recorded in state 1, and  
243 were switched to state 2 (or from state 2 back to state 1) conditional on the distance between  
244 trapping stakes upon recapture being greater than the guild-level benchmark defining short  
245 distance movements. In this two-state CMR model, transitioning between states indicates long  
246 distance movement, and staying in the same state indicates short distance movement, conditional  
247 on apparent survival and recapture probabilities. We defined apparent survival probability as the  
248 probability that an individual alive in trapping period  $i$  survived and did not emigrate from the  
249 entire study area by trapping period  $i+1$ . We defined recapture probability at trapping period  $i+1$   
250 as the probability that a live individual anywhere on the study area was recaptured in a trap. All

251 probabilities were measured over a time scale of approximately one month, the time between  
252 trapping events. To address inconsistencies in the data, we controlled for omitted trap periods  
253 (when trapping did not occur or the site was only partially trapped) by fixing recapture  
254 probability to zero for those occasions. It should be noted that we could not differentiate between  
255 permanent emigration and death, which may affect interpretation of our survival estimates. Thus,  
256 low apparent survival probabilities may indicate low actual survival, high permanent emigration  
257 from the entire study area, or both. We evaluated each species separately in RMark to estimate  
258 apparent survival, recapture, and transition probabilities (White and Cooch 2012), except for  
259 transient granivores, which we grouped together because there were not enough captures to  
260 analyze species separately. Pooling data for all species into a single dataset, and designating  
261 species or strategies with factors, led to an extremely large CMR dataset that prohibited  
262 computational analysis using MARK and RMark. We thus used post hoc analyses to compare the  
263 estimates for core versus intermediate and transient species. For further details on our RMark  
264 analysis, please refer to our code, which is maintained online in a public GitHub repository along  
265 with the data (<https://github.com/weecology/portal-rodent-dispersal>) and is available in the  
266 online supplement.

267 To assess reproductive effort for each species, we tracked the reproductive history for  
268 captured individual females within each calendar year. We considered females with enlarged  
269 and/or red nipples or who were pregnant (researcher could feel embryos) to be actively  
270 reproducing. If a female was marked in reproductive condition during consecutive trapping  
271 periods, we assumed it to be a single reproductive event. Reproductive condition recorded across  
272 non-consecutive trapping periods was considered as multiple reproductive events. We used data  
273 from females because males display reproductive signals for a larger portion of the year, and

274 male reproductive status does not necessarily indicate recent copulation or reproductive success.  
275 For each species, we also recorded litter size and number of litters per year from the literature  
276 (Hoffmeister 1986).

277 To compare the life-history traits among groups, we standardized the data and results to  
278 account for different units (  $[(x - \text{mean}(x))/\text{sd}(x)]$  ) and controlled for the influence of  
279 phylogenetic relatedness on our results (Felsenstein 1985). We used a generalized least squares  
280 model to test for the correlation of traits and trait trade-offs with phylogenetic relatedness (APE,  
281 Paradis et al. 2004; Geiger, Harmon et al. 2008; picante, Kembel et al. 2010) based on a  
282 published mammalian phylogeny (Bininda-Emonds et al. 2007; PhyloOrchard, O'Meara et al.  
283 2013) and using the assumption of linear decrease in trait covariance (bm gls; Brownian motion  
284 model). We also compared species using a linear regression and PCA biplots to determine the  
285 relative influences of temporal status, feeding guild, and phylogenetic relatedness on movement  
286 patterns and life history traits.

287

## 288 **Results**

### 289 *Core-transient species designation*

290 During the 21-year study period, we captured 12,651 individuals from the 21 species  
291 included in the analysis (Table 1). Based on the proportion of years that each species was  
292 present, we placed species into three temporal persistence categories (Figure 2) consisting of 10  
293 core (*Dipodomys ordii*, *D. merriami*, *Chaetodipus baileyi*, *C. penicillatus*, *Perognathus flavus*,  
294 *Peromyscus eremicus*, *Reithrodontomys megalotis*, *Neotoma albigula*, *Onychomys torridus*, and  
295 *O. leucogaster*), 7 transient species (*C. hispidus*, *C. intermedius*, *Peromyscus leucopus*, *R.*  
296 *montanus*, *R. fulvescens*, *Baiomys taylori*, and *Sigmodon ochrognathus*), and 4 intermediate

297 species (*D. spectabilis*, *Peromyscus maniculatus*, *S. fulviventer*, and *S. hispidus*). Granivores and  
298 folivores were represented across all three persistence categories, but both carnivorous species  
299 were designated “core” (Table 1). We removed *Chaetodipus hispidus* from our main  
300 comparisons because there were only 10 individuals with recaptures and we could not run  
301 multistate CMR models on such a small sample. Since transient species were rarely captured at  
302 the site and thus do not have enough data points for robust statistical analysis, we focus our  
303 results on the comparison between core and non-core species, where non-core refers to  
304 intermediate and transient species grouped together, generally.

### 305 *Movement, Survival, and Fecundity*

306 For all the species, movement distances between recaptures were strongly unimodal and  
307 left-skewed. Transient and intermediate species generally had a longer tail on their movement  
308 distributions than core species and distributions for many non-core species had a secondary  
309 mode, suggesting more long-distance movements, larger home ranges, and possible emigration  
310 off-site (Table 1, Figure 3; Figure A1). Transient species that had few long-distance movements  
311 may be attributed to high mortality, low detectability on the site, low recapture due to rapid  
312 movement off the site, or a combination of these. Core species tended to move shorter distances  
313 (median species-level movement benchmark; core=42.91 meters [29.96-152.45]; non-core=65.24  
314 [24.70-365.01]). Among granivores, core species had a lower frequency of moving a relatively far  
315 distance away from the previous trap location (median benchmark=41.50 m) than intermediate  
316 (median benchmark=72.42 m) or transient granivore species (median benchmark=63.42 m;  
317 Table 1). Among folivores, the core species *N. albigula* generally moved shorter distances  
318 (benchmark=39.76 m) than the intermediate (benchmark=53.31 m) and transient species (mean  
319 benchmark=91.85 m; Table 1). Core guild-level movement benchmarks were used to

320 parameterize the CMR model and differed across trophic groups ( $\text{benchmark}_{\text{granivore}}=36.70$ ;  
 321  $\text{benchmark}_{\text{folivore}}=39.76$ ;  $\text{benchmark}_{\text{carnivore}}=88.07$ ).

322 When using the two-state CMR model in Mark to compare apparent survival, recapture,  
 323 and movement probabilities among core and transient species, differences were most pronounced  
 324 among core granivores versus transient and intermediate granivores (Table 1; Table S1). On  
 325 average, core species had a lower probability of moving a long distance (mean  $\psi_{\text{core}}=0.14$  [range  
 326 0.03-0.37]), but higher recapture ( $p_{\text{core}}=0.39$  [range 0.21-0.63]) and apparent survival  
 327 probabilities (mean  $\Phi_{\text{core}}=0.76$  [range 0.63-0.84]) than non-core species (mean  $\psi_{\text{non-core}}=0.25$   
 328 [range 0.02-0.86], mean  $p_{\text{non-core}}=0.35$  [range 0.07-0.66], mean  $\Phi_{\text{non-core}}=0.67$  [range 0.45-0.83]).

329 Reproductive results from the field data were best explained by phylogeny. For all  
 330 species in *Heteromyidae* (5 core and 3 non-core), the majority of captured females were never  
 331 recorded in reproductive condition (Table 1; Figure A2). However, despite generally much lower  
 332 abundance, species in *Cricetidae* were observed in reproductive condition more often. For  
 333 example, nearly 50% of *Peromyscus eremicus* (core) and *P. maniculatus* (non-core) were  
 334 recorded in reproductive condition at least once per year (Table 1) and *N. albigula* (core) females  
 335 were often found in reproductive condition. However, *Sigmodon* (non-core) females were almost  
 336 never recorded as reproductive (Table 1). The lack of observed reproduction may suggest that  
 337 *Sigmodon* rarely reproduce at the site or that sampling error associated with the small number of  
 338 captures affected our results. *Onychomys* (core) females were rarely captured when reproductive,  
 339 but other data suggest that *O. torridus* may reproduce multiple times per year (Table 1, Table 2).

#### 340 *Phylogeny and Trade-offs*

341 PCA results suggested that species can be grouped in multivariate space by their traits  
 342 and core-transient status (Figure 4), and that traits appear to be strongly conserved within family

343 (Figure A3). Phylogeny (family) was a significant predictor of  $\psi$  (linear model;  $\text{lm}(\psi \sim \text{family})$ ,  
344  $p=0.016$ ,  $r^2=0.28$ ),  $\Phi$  ( $\text{lm}(\Phi \sim \text{family})$ ,  $p=0.006$ ,  $r^2=0.35$ ), and mean abundance ( $\text{lm}$   
345 ( $\text{abundance} \sim \text{family}$ ),  $p=0.004$ ,  $r^2=0.37$ ), but not for the proportion of years a species was present.  
346 Mean abundance was positively related to the proportion of years a species was present in the  
347 study area when phylogeny was controlled for (bm gls,  $p=0.005$ ). Body size was not a significant  
348 predictor for the proportion of years present, mean abundance, survival, reproduction, or  
349 movement (bm gls,  $p \gg 0.05$ ). We did not detect strong movement-survival trade-offs or  
350 movement-reproduction trade-offs in this community using linear regression or phylogenetic  
351 methods (Figures A4, A5). There was a relationship between observed modal movement distance  
352 and temporal persistence ( $\text{lm}(\text{distance mode} \sim \text{proportion years present})$ ,  $p=0.03$ ,  $r^2=0.23$ ; bm  
353 gls,  $p=0.05$ ) where species that persisted longer moved shorter distances.

## 355 Discussion

356 Our study provides the first test, to our knowledge, of whether life-history traits are  
357 associated with the temporal persistence patterns of species in a community. Life-history traits  
358 were generally conserved within evolutionary lineages, which in our system are also related to  
359 higher or lower degrees of adaptation to desert environments. Our results provide some  
360 indication that life history traits are also linked to the core-transient structure among species at  
361 our site. Analysis of the movement, survival and reproduction data indicated that core species at  
362 our site moved relatively short distances and had higher apparent survival rates than intermediate  
363 and transient species, consistent with our predictions. Core species were generally more  
364 abundant than transient or intermediate species, which may suggest greater competitive  
365 dominance or ecological specialization. Transient species at our site exhibited lower apparent



366 survival and moved longer distances, on average, than core species, but the range of estimates  
367 within this group varied widely. We interpret the high variation in survival and movement  
368 estimates across species within the intermediate and transient groups as support for the idea that  
369 this heterogeneous group includes both source-sink and nomadic transients. Among traits, we did  
370 not detect strong movement-survival trade-offs, and too few reproductive events were observed  
371 to test for survival-reproduction trade-offs.

372 Other research at our site also supports the idea that our core and transient species have  
373 different effects on community structure and resource use. Core species that exhibit high self-  
374 investment, strong competitive ability, and strong local adaptation are also expected to  
375 successfully exploit most of the available resources at a local site (Kneitel and Chase 2004). The  
376 most abundant core species at our site are indeed responsible for the majority of resource  
377 consumption (Thibault et al. 2004), are behaviorally dominant (*Dipodomys* spp.; Reichmann and  
378 Price 1993), and have cascading impacts on rodent and plant communities (Brown and Heske  
379 1990; Heske *et al.* 1994; Valone and Schutzenhofer 2007). Pocket mice (*Chaetodipus* spp.),  
380 another group of core species, become dominant in the absence of kangaroo rats (Ernest and  
381 Brown 2001). Therefore, it is generally true in this system that species with lower movement,  
382 higher survivorship, and potentially lower reproductive rates are also the competitively dominant  
383 species responsible for the majority of resource consumption (Thibault et al. 2004).

384 The life-history estimates for transient and intermediate species (n=11) generally  
385 supported our hypotheses. We expected that results for this group of species would exhibit broad  
386 interspecific variation because of the presence of both nomadic and source-sink transients.  
387 Unfortunately, it is not possible to definitively distinguish between these groups using our data,  
388 but using a combination of our results and the well-studied natural history of these species, we

389 can make some informed predictions. Species that exhibited trait correlations more similar to  
390 core species could be source-sink transients (Lenormand 2002; Kawecki 2008), or could be  
391 undergoing local extinction (Gibson et al. 1999). Because they are core somewhere else, source-  
392 sink transients may generally be inferior competitors that temporarily colonize in response to  
393 resource pulses and density dependent dispersal at other locations, both of which could create a  
394 source-sink dynamic over time (Heske *et al.* 1994; Thibault *et al.* 2004; Milstead et al. 2007).  
395 *Dipodomys* and *Chaetodipus* species that were not core at our site are either undergoing local  
396 extinction due to directional habitat change (*D. spectabilis*; Valone et al. 2002, Thibault et al.  
397 2004) or are not adapted to the habitat at our site (*C. intermedius* and *C. hispidus*; Hoffmeister  
398 1986; Paulson 1988, Williams et al. 1993) and their occasional presence can likely be attributed  
399 to temporary dispersal from nearby source populations (Wilson and Ruff 1999). At our site,  
400 *Sigmodon* and *Reithrodontomys* are prairie-adapted species (Webster and Jones 1982;  
401 Hoffmeister 1986) that usually arrive during years and seasons where climatic conditions lead to  
402 higher than normal grass cover (Thibault *et al.* 2004). During our study period these species had  
403 relatively low abundance and were rarely recorded in reproductive condition – strong evidence  
404 that our site represents a habitat sink for these populations. Species that exhibited trait  
405 correlations different from core species could be nomadic transients. *Peromyscus maniculatus*  
406 and *leucopus* at our site were generally rare, moved longer distances more often, potentially  
407 reproduce multiple times per year, and use a wider resource base than the core species  
408 (Hoffmeister 1986; Wilson and Ruff 1999), evidence that these species could readily colonize  
409 new locations. In addition, *Baiomys taylori* seem to be dependent on high ground cover, and thus  
410 requiring populations to nomadically track habitats receiving more than the usual amount of  
411 water (Wilson and Ruff 1999). Thus, through these two different processes, source-sink and

412 nomadic transient species maintain presence in the regional metacommunity and comprise a  
413 dynamic component of the local community.

414         The core-transient framework, integrated with an eco-evolutionary viewpoint, suggests  
415 that core and transient species may be responsible for maintaining different aspects of ecosystem  
416 function at a local site. Because of their higher local adaptation and stronger ability to exploit  
417 resources under local conditions, core species may be particularly important for maintaining  
418 ecosystem function (Grime 1998; Henderson and Magurran 2014). In contrast, because of their  
419 higher movement rates, transient species may be particularly important for maintaining local  
420 diversity through time and in response to disturbance (Dornelas et al 2014, Henderson and  
421 Magurran 2014; Supp and Ernest 2014). Our site exhibits high annual composition turnover  
422 while generally maintaining species richness (Brown et al. 2001, Goheen et al. 2005), which  
423 indicates a role for nomadic transients in the regional metacommunity that can opportunistically  
424 fill empty niches and thus maintain community-level species richness. Our results suggest that  
425 species with locally adapted traits well suited to maximizing ecosystem functions related to  
426 resource use may also have lower movement rates. If so, this relationship between traits and  
427 movement patterns may be particularly important to explore in a metacommunity context.

428         The core-transient and core-satellite frameworks are merely two different ways of  
429 considering commonness and rarity in ecological systems. Temporal patterns of species  
430 persistence at a site depend on environmental filtering and competitive processes (Magurran and  
431 Henderson 2003; Coyle et al. 2013), but also partially depend on spatial patterns of habitat  
432 heterogeneity and movement between linked populations (Hanski 1982; Grime 1998; Gibson et  
433 al. 1999). A difficulty in linking the two ideas is that they define commonness and rarity in two  
434 different ways – species are categorized by occupancy of a regionally defined area (core-

435 satellite) or species are categorized by occupancy of a single site through time (core-transient).  
436 Combining the two frameworks would require modeling patch occupancy in the region through  
437 time, while allowing for core-satellite switching (e.g., Gaston and Lawton 1989, Gotelli 1991),  
438 and assigning core-transient species for each site across the time-series. Presumably, such a  
439 model would lead to several new categories of species: (a) regionally core species that remain  
440 core throughout the time-series and are temporally core at sites in which they occur, (b)  
441 regionally core species that switch to satellite species in poor years (e.g., experience local  
442 population crashes) or that display enhanced occupancy in good years (e.g., experience density  
443 dependent dispersal), and are mixed across sites as temporally core or transient, (c) satellite  
444 species that are temporally core in sites in which they occur, (d) satellite species that switch to  
445 regionally core in high resource years (e.g. irruptive dynamics), and are mixed across sites as  
446 temporally core or transient, (e) satellite species that switch the sites they occupy through time,  
447 and thus are temporally transient across the sites in which they occur (e.g., nomadism).  
448 Synthetically understanding the characteristics of species that display different or overlapping  
449 syndromes of commonness (e.g., locally abundant, high regional occupancy, and/or high  
450 temporal persistence) or rarity (locally rare, low regional occupancy, and/or low temporal  
451 persistence) remains a critical area for future research in community ecology.

452         Assessing whether there are life history implications for species using local versus  
453 regional scale processes for population maintenance is challenging. Long-term mark-recapture  
454 data that provides information on movement, survival, and reproduction rates across an entire  
455 community, including rare species, is very difficult data to obtain. Despite the shortcomings of  
456 our data, our results suggest that core and transient species may differ not only in their temporal  
457 persistence, but also in their use of landscapes, traits for resource use, and structure of their life

458 histories. If other studies support our findings, it suggests that core-transient species traits should  
459 be modeled separately to understand how these groups might respond differently to  
460 environmental change. Core and source-sink transient species generally associated with low  
461 levels of gene flow or that are unable to track shifts in the location of suitable habitat across the  
462 landscape, would be predicted to undergo catastrophic decline under environmental changes that  
463 alter local conditions long-term (*e.g.* temperature, resource availability). Alternately, nomadic  
464 transient species that emigrate readily and exhibit high levels of gene flow among populations  
465 would be predicted to exhibit relatively small change in response to the same scenario. In the  
466 drive to better understand the response of biodiversity to perturbations, a temporal perspective of  
467 species demographics and persistence represents a critical link in identifying the linkages  
468 between local and regional richness patterns and predicting community response to change.

#### 470 **Acknowledgements**

471 We would like to thank P. Adler, M. Pfrender, E. P. White, L. Aubry, K. Shoemaker, B. Ross,  
472 M. Pendergast, S. Mortara, A. Hurlbert, J. Coyle, M.A. Leibold, The Weecology Lab at Utah  
473 State University (USU), and the Graham Lab at Stony Brook University for their valuable  
474 comments and discussion on the development of the ideas and analyses presented in the  
475 manuscript. One anonymous reviewer and Ilkka Hanski provided comments on the manuscript  
476 during the review process. We are grateful for the many scientists and volunteers that helped  
477 collect data for the long-term Portal Project. SRS was supported by NSF grant DEB-1100664  
478 and fellowships from the USU Ecology Center and the School of Graduate Studies.

479

480 **References**

- 481 Baguette, M., S. Blanchet, D. Legrand, V.M. Stevens, and C. Turlure. 2013. Individual dispersal,  
482 landscape connectivity and ecological networks. *Biological Reviews* 88: 310-326.
- 483 Baker, R.R. 1978. *The evolutionary ecology of animal migration*. Hodder and Stoughton,  
484 London.
- 485 Belmaker, J., Y. Ziv, N. Shashar, and S.R. Connolly. 2008. Regional variation in the hierarchical  
486 partitioning of diversity in coral-dwelling fishes. *Ecology* 89: 2829-2840.
- 487 Belmaker, J. 2009. Species richness of resident and transient coral-dwelling fish responds  
488 differently to regional diversity. *Global Ecology and Biogeography*. 18: 426-436.
- 489 Bininda-Emonds, O.R.P., M. Cardillo, K.E. Jones, R.D.E. MacPhee, R.M.D. Beck, R. Grenyer,  
490 S.A. Price, R.A. Vos, J.L. Gittleman, and A. Purvis. 2007. The delayed rise of present-day  
491 mammals. *Nature* 446: 507-512.
- 492 Brown, J.H. 1998. The Granivory Experiments at Portal, in *Experimental Ecology: Issue and*  
493 *Perspectives*. W.J. Reseraris, and J. Bernardo, editors. Oxford University Press, New York, New  
494 York, USA.
- 495 Brown, J.H., and E.J. Heske. 1990. Control of a desert-grassland transition by a keystone rodent  
496 guild. *Science* 250: 1705-1707.
- 497 Brown, J.H., S.K.M. Ernest, J.M. Parody, and J.P. Haskell. 2001. Regulation of diversity:  
498 maintenance of species richness in changing environments. *Oecologia* 126: 321-332.
- 499 Charlesworth, B. 1980. *Evolution in Age-structured Populations*. Cambridge University Press.  
500 Cambridge, UK.
- 501 Connolly, S.E., M.A. MacNeil, M.J. Caley, N. Knowlton, E. Cripps, M. Hisano, L.M. Thibaut,  
502 B.D. Bhattacharya, L. Benedetti-Cecchi, R.E. Brainard, A. Brandt, F. Bulleri, K.E. Ellingsen, S.

- 503 Kaiser, I. Kroncke, K. Linse, E. Maggi, T.D. O'Hara, L. Plaisance, G.C.B. Poore, S.K. Sarkar,  
504 K.K. Satpathy, U. Schuckel, A. Williams, R.S. Wilson. 2014. Commonness and rarity in the  
505 marine biosphere. *Proceedings of the National Academy of Sciences of the United States* 111:  
506 8524-8529.
- 507 Costello, M.J. and A.A. Myers. 1996. Turnover of transient species as contributor to the richness  
508 of a stable amphipod (Crustaceae) fauna in a sea inlet. *Journal of Experimental Marine Biology*  
509 *and Ecology* 202: 49-62.
- 510 Cottingham.K.L., B.L. Brown and J.T. Lennon. 2001. Biodiversity may regulate the temporal  
511 variability of ecological systems. *Ecology Letters* 4: 72-85.
- 512 Coyle, J.R., A.H. Hurlbert, and E.P. White. 2013. Opposing mechanisms drive richness patterns  
513 of core and transient bird species. *American Naturalist*. 181: E83-90
- 514 Diamond, J.M. 1974. Colonization of exploded volcanic islands by birds: the supertramp  
515 strategy. *Science* 184: 803-806.
- 516 Dolan, J.R., M.E. Ritche, A. Tunin-Ley, and M.D. Pizay. 2009. Dynamics of core and transient  
517 species in the marine plankton: tintinnid ciliates in the north-west Mediterranean Sea. *Journal of*  
518 *Biogeography* 36: 887-895.
- 519 Ernest, S.K.M., and J.H. Brown. 2001. Delayed compensation for missing keystone species by  
520 colonization. *Science* 292: 101-104.
- 521 Ernest, S.K.M., T.J. Valone, and J.H. Brown. 2009. Long-term monitoring and experimental  
522 manipulation of a Chihuahuan Desert ecosystem near Portal, Arizona, USA. *Ecology* 90:1708.
- 523 Felsenstein, J. 1985. Phylogenies and the comparative method. *American Naturalist* 125: 1-15.
- 524 Gaggiotti, O.E. 1996. Population genetic models of source-sink metapopulations. *Theoretical*  
525 *Population Biology* 50: 178-208.

- 526 Gaston, K.J. and J.H. Lawton. 1989. Insect herbivores on bracken do not support the core-  
527 satellite hypothesis. *The American Naturalist* 134: 761-777.
- 528 Ghalambor, C.K. and T.E. Martin. 2000. Parental investment strategies in two species of  
529 nuthatch vary with state-specific predation risk and reproductive effort. *Animal Behavior* 60:  
530 263-267.
- 531 Gibson, D.J., J.S. Ely, and S.L. Collins. 1999. The core-satellite species hypothesis provides a  
532 theoretical basis for Grime's classification of dominant, subordinate, and transient species.  
533 *Journal of Ecology* 87: 1064-1067.
- 534 Goheen, J.R., E.P. White, S.K.M. Ernest, and J.H. Brown. 2005. Intra-guild compensation  
535 regulates species richness in desert rodents. *Ecology* 86: 567-573.
- 536 Gotelli, N.J. 1991. Metapopulation models: the rescue effect, the propagule rain, and the core-  
537 satellite hypothesis. *The American Naturalist* 138: 768-776.
- 538 Grime, J.P. 1998. Benefits of plant diversity to ecosystems: immediate, filter, and founder  
539 effects. *Journal of Ecology* 86: 902-910.
- 540 Hanski, I. 1982. Dynamics of regional distribution: the core and satellite species hypothesis.  
541 *Oikos* 38: 210-221.
- 542 Harmon L.J., J.T. Weir, C.D. Brock, R.E. Glor, and W. Challenger. 2008. GEIGER:  
543 investigating evolutionary radiations. *Bioinformatics* 24:129-131.
- 544 Henderson, P.A. and A.E. Magurran. 2014. Direct evidence that density-dependent regulation  
545 underpins the temporal stability of abundant species in a diverse animal community. *Proceedings*  
546 *of the Royal Society B: Biological Sciences*, 281: 20141336.
- 547 Heske, E.J., J.H. Brown, and S. Mistry. 1994. Long-term experimental study of a Chihuahuan  
548 desert rodent community: 13 years of competition. *Ecology* 75: 438-445.



- 549 Hoffmeister, D.F. 1986. *Mammals of Arizona*. University of Arizona Press and Arizona Game  
550 and Fish Department. Tucson, Arizona, USA.
- 551 Holt R.D. and M.S. Gaines. 1992. Analysis of adaptation in heterogeneous landscapes:  
552 implications for the evolution of fundamental niches. *Evol. Ecol.* 6:433–47
- 553 Jonzen, N., E. Knudsen, R.D. Holt, and B.E. Saether. 2011. Uncertainty and predictability: the  
554 niches of migrants and nomads. *in Animal Migration: A Synthesis* pp 91-109.
- 555 Kawecki, T.J. 2008. Adaptation to marginal habitats. *Annual Review of Ecology, Evolution, and*  
556 *Systematics* 39: 321-342.
- 557 Kawecki T.J. and R.D. Holt. 2002. Evolutionary consequences of asymmetric dispersal rates.  
558 *American Naturalist* 160:333–47
- 559 Kembel, S.W., P.D. Cowan, M.R. Helmus, W.K. Cornwell, H. Morlon, D.D. Ackerly, S.P.  
560 Blomberg, and C.O. Webb. 2010. Picante: R tools for integrating phylogenies and ecology.  
561 *Bioinformatics* 26:1463-1464.
- 562 Kisdi, E. 2002. Dispersal: risk spreading versus local adaptation. *American Naturalist.* 159: 579-  
563 596.
- 564 Kneitel, J.M. and J.M. Chase. 2004. Trade-offs in community ecology: linking spatial  
565 scales and species coexistence. *Ecology Letters* 7: 69-80.
- 566 Laake, J., E. Rakimberdiev, B. Collier, J. Rotella, and A. Paul. 2012. RMark: R Code for MARK  
567 Analysis. R package version. 2.1.0. <http://CRAN.R-project.org/package=RMark>.
- 568 Laake, J. E., and E. Rexstad. 2008. RMark – an alternative approach to building linear models in  
569 MARK. pages C1-C115 in E. Cooch and G. White, editors. *Program MARK: A gentle*  
570 *introduction, 9<sup>th</sup> Ed.*

- 571 Lenormand, T. 2002. Gene flow and the limits to natural selection. *Trends in Ecology and*  
572 *Evolution* 17:183–89.
- 573 Mace, G.M. and P.H. Harvey. 1983. Energetic constraints on home-range size. *American*  
574 *Naturalist* 121: 120-132.
- 575 Magurran, A.E., S. Khachonpisitsak, and A.B. Ahmad. 2011. Biological diversity of fish  
576 communities: pattern and process. *Journal of Fish Biology* 79: 1393-1412.
- 577 Magurran, A.E., and P.A. Henderson. 2012. How selection structures species abundance  
578 distributions. *Proceedings of the Royal Society B* 279: 3722-3726.
- 579 Magurran, A.E., and P.A. Henderson. 2010. Temporal turnover and the maintenance of diversity  
580 in ecological assemblages. *Philosophical Transactions of the Royal Society B* 365: 3611-3620.
- 581 Magurran, A.E., and P.A. Henderson. 2003. Explaining the excess of rare species in natural  
582 species abundance distributions. *Nature* 422: 714-716.
- 583 McCauley, S.J. 2007. The role of local and regional processes in structuring dragonfly  
584 distributions across habitat gradients. *Oikos* 116: 121-133.
- 585 McPeck, M.A., and R.D. Holt. 1992. The evolution of dispersal in spatially and temporally  
586 varying environments. *American Naturalist* 140: 1010-1027.
- 587 Milstead, W.B., P.L. Meserve, A. Campanella, M.A. Previtali, D.A. Kelt, and J.R. Gutierrez.  
588 2007. Spatial ecology of small mammals in north-central Chile: Role of precipitation and refuges.  
589 *Journal of Mammalogy* 88: 1532-1538.
- 590 Mueller, T. and W.F. Fagan. 2008. Search and navigation in dynamic environments – from  
591 individual behaviors to population distributions. *Oikos* 117: 654-664.
- 592 Murray, B.G. 1967. Dispersal in vertebrates. *Ecology* 48: 975-978.

- 593 O'Donnell, C.F.J. 2001. Home range and use of space by *Chalinolobus tuberculatus*, a temperate  
594 rainforest bat from New Zealand. *Journal of Zoology* 253: 253-264.
- 595 O'Meara, B.C., L. Harmon and J. Eastman. 2013. PhyloOrchard: Important and/or useful  
596 phylogenetic datasets. R package version 0.3.
- 597 Paulson, D.D. 1988. *Chaetodipus hispidus*. *Mammalian Species*. No. 320. The American Society  
598 of Mammalogists. pp 1-4.
- 599 Paradis E., J. Claude, and K. Strimmer. 2004. APE: analyses of phylogenetics and evolution in R  
600 language. *Bioinformatics* 20: 289-290.
- 601 R Development Core Team. 2013. R: A language and environment for statistical computing. R  
602 Foundation for Statistical Computing. Vienna, Austria. ISBN 3-900051-07-0. [http://www.R-](http://www.R-project.org/)  
603 [project.org/](http://www.R-project.org/).
- 604 Reichmann, O.J., and M.V. Price. 1993. Ecological Aspects of Heteromyid foraging. Pages 539-  
605 574 in H.H. Genoway and J. H. Brown, editors. *Biology of the Heteromyidae*
- 606 Ricklefs, R.E. and M. Wikelski. 2002. The physiology/life-history nexus. *Trends in Ecology and*  
607 *Evolution* 17: 462-468.
- 608 Roshier, D. and J. Reid. 2003. On animal distributions in dynamic landscapes. *Ecography* 26:  
609 539-544.
- 610 Rousset, F. and S. Gandon. 2002. Evolution of the distribution of dispersal distance under  
611 distance-dependent cost of dispersal. *Journal of Evolutionary Biology* 15: 515-523.
- 612 Sibly, R.M. and J.H. Brown. 2007. Effects of body size and lifestyle on evolution of mammal  
613 life histories. *Proceedings of the National Academy of Sciences* 104: 17707-17712.
- 614 Singleton, G.R., Sudarmaji, S. Suriapermana. 1998. An experimental field study to evaluate a  
615 trap-barrier system and fumigation for controlling the rice field rat, *Rattus argentiventer*, in rice

- 616 crops in West Java. *Crop Protection* 17: 55-64.
- 617 Sokal, R.R. and F.J. Rohlf. 1995. *Biometry*, 3<sup>rd</sup> Ed. W. H. Freeman and Company, New York,  
618 New York, USA.
- 619 Stearns, S. 1992. *The Evolution of Life Histories*. Oxford University Press. Oxford, UK.
- 620 Thibault, K.M., E.P. White, and S.K.M. Ernest. 2004. Temporal dynamics in the structure and  
621 composition of a desert rodent community. *Ecology* 85: 2649-2655.
- 622 Ulrich, W. and M. Ollik. 2004. Frequent and occasional species and the shape of relative-  
623 abundance distributions. *Diversity and Distributions* 10: 263-269.
- 624 Urban, M.C. and D.K. Skelly. 2006. Evolving metacommunities: Toward an evolutionary  
625 perspective on metacommunities. *Ecology* 87: 1616-1626.
- 626 Valone, T.J. and M.R. Schutzenhofer. 2007. Reduced rodent biodiversity destabilizes plant  
627 populations. *Ecology* 88: 26-31.
- 628 Valone, T.J., M. Meyer, J.H. Brown, and R.M. Chew. 2002. Timescale of perennial grass  
629 recovery in desert arid grasslands following livestock removal. *Conservation Biology* 16: 995-  
630 1002.
- 631 Vergnon, R., N.K. Dulvy, R.P. Freckleton. 2009. Niches versus neutrality: uncovering the  
632 drivers of diversity in a species-rich community. *Ecology Letters* 12: 1079-1090.
- 633 Waser, P.M. 1985. Does competition drive dispersal? *Ecology* 66: 1170-1175.
- 634 Webster, W.D., and J.K. Jones, Jr. 1982. *Reithrodontomys megalotis*. *Mammalian Species*. No.  
635 167. The American Society of Mammalogists. pp 1-5.
- 636 Wiens, J. 2001. The landscape context of dispersal. *In* *Dispersal*. Eds J. Clobert. Oxford  
637 University Press, Oxford.

638 Williams, D.F., H.H. Genoways, and J.K. Braun. 1993. Taxonomy and systematics. Pp. 38-196  
639 in *Biology of the Heteromyidae*. Eds H.H. Genoways and J.H. Brown. Special Publication, The  
640 American Society of Mammalogists 10: 1-719

641 White, G. C., and K.P. Burnham. 1999. Program MARK: survival estimation from populations  
642 of marked animals. *Bird Study* 46 Supplement:120-138.

643 White, G.C., and E. Cooch. 2012. *Program MARK: A Gentle Introduction*, 11<sup>th</sup> Ed.

644 <http://www.phidot.org/software/mark/docs/book/>

645 Wilson, D.E. and S. Ruff. 1999. *The Smithsonian Book of North American Mammals*.  
646 Smithsonian Institution Press *in association with the* American Society of Mammalogists.  
647 Washington, D.C. pp. 494-598.

648

649

650

651

652

653

654

655

656

657

658

658

659 **Tables**660 **Table 1.**

661

Species	Guild	Status	N	Mean mass (g)	Mean num. reproduction events/year	Movement benchmark (m)	$\Phi$ (std. error)	$p$ (std. error)	$\Psi$ (std. error)
<i>Dipodomys merriami</i>	Gran.	Core	1972	43.64	0.52	33.13	0.80 (0.00)	0.60 (0.01)	0.06 (0.00)
<i>D. ordii</i>	Gran.	Core	1030	48.32	0.47	30.35	0.75 (0.01)	0.63 (0.01)	0.03 (0.00)
<i>D. spectabilis</i>	Gran.	Intermed.	41	115.01	0.21	24.70	0.83 (0.02)	0.66 (0.04)	0.02 (0.01)
<i>Chaetodipus baileyi</i>	Gran.	Core	2063	31.77	0.40	29.96	0.84 (0.00)	0.56 (0.01)	0.04 (0.00)
<i>C. penicillatus</i>	Gran.	Core	2818	17.01	0.40	44.33	0.83 (0.00)	0.21 (0.00)	0.06 (0.02)
<i>C. hispidus</i>	Gran.	Transient	10	31.74	0.43	16.49	--	--	--
<i>C. intermedius</i>	Gran.	Transient	20	18.46	0.14	33.96	0.79 (0.05)	0.25 (0.07)	0.06 (0.04)
<i>Perognathus flavus</i>	Gran.	Core	648	8.12	0.54	41.50	0.78 (0.01)	0.23 (0.01)	0.06 (0.01)
<i>Peromyscus eremicus</i>	Gran.	Core	479	21.47	0.53	107.18	0.63 (0.02)	0.32 (0.02)	0.37 (0.04)
<i>Pe. maniculatus</i>	Gran.	Intermed.	398	21.41	0.46	120.13	0.62 (0.02)	0.28 (0.03)	0.53 (0.05)
<i>Pe. leucopus</i>	Gran.	Transient	77	22.41	0.24	56.05	0.46 (0.06)	0.44 (0.11)	0.12 (0.06)
<i>Reithrodontomys megalotis</i>	Gran.	Core	1180	10.71	0.23	95.17	0.63 (0.01)	0.27 (0.01)	0.33 (0.20)
<i>R. montanus</i>	Gran.	Transient	18	9.01	0.67	365.01	0.72 (0.14)	0.07 (0.06)	0.86 (0.15)
<i>R. fulvescens</i>	Gran.	Transient	11	13.90	1.50	63.42	0.81 (0.05)	0.63 (0.08)	0.22 (0.08)
<i>Baiomys taylori</i>	Gran.	Transient	106	8.76	0.26	70.65	0.63 (0.05)	0.22 (0.05)	0.12 (0.05)
<i>Sigmodon hispidus</i>	Foliv.	Intermed.	268	88.84	0.12	39.56	0.45 (0.04)	0.38 (0.06)	0.08 (0.03)
<i>S. fulviventer</i>	Foliv.	Intermed.	151	66.01	0.07	67.06	0.66 (0.03)	0.35 (0.04)	0.14 (0.03)
<i>S. ochrognathus</i>	Foliv.	Transient	27	55.55	0.37	91.85	0.70 (0.07)	0.21 (0.07)	0.31 (0.14)
<i>Neotoma albigula</i>	Foliv.	Core	255	167.05	0.66	39.76	0.79 (0.01)	0.33 (0.02)	0.10 (0.02)
<i>Onychomys torridus</i>	Carn.	Core	952	23.81	0.49	80.87	0.76 (0.01)	0.42 (0.01)	0.09 (0.01)
<i>O. leucogaster</i>	Carn.	Core	127	30.16	0.54	152.45	0.77 (0.02)	0.37 (0.03)	0.27 (0.04)

662

663 **Table 1. Summary of species traits from the field data.** Species-level trait details summarizing  
664 feeding guild, core-transient status, total number of individuals tracked through the study (N),  
665 mean body mass across all recorded weights, mean yearly reproductive effort, and species-level  
666 benchmarks defining where each movement distribution transitions into long-distance  
667 movements. Yearly reproductive effort was estimated by taking weighted average of the number  
668 of individual females marked as reproductive 0-4 times per year. -- indicates that we could not  
669 generate multistate CMR estimates due to small sample size ( $N \leq 10$ ).

670

671

672

673

674

675

676

677

678

679

680

681

682

683

684

685

686

687

688

689

690

691

691  
692  
693**Table 2.**

Species	Status	Litter size	Mean litter size	Number of litters per year	Typical breeding months
<i>Dipodomys ordii</i> *	Core	2-3	2.37	1-2	February-July
<i>D. merriami</i> *	Core	2-3	2	1-2	March-October
<i>Chaetodipus baileyi</i> *	Core	1-6	3.6	--	April-August
<i>C. penicillatus</i> *	Core	2-8	4.72	1	April-August
<i>Perognathus flavus</i> *	Core	1-6	4	1	April-August
<i>Peromyscus eremicus</i> *	Core	1-4	2.53	1-4 <sup>†</sup>	Year-round
<i>Reithrodontomys megalotis</i> *	Core	3-7 <sup>†</sup>	3.6	1-10	Year-round
<i>Neotoma albigula</i>	Core	1-4	1.95	≥ 1	Year-round
<i>Onychomys torridus</i>	Core	2-5	3.45	--	March-October
<i>O. leucogaster</i>	Core	3-5	4	--	March-September
<i>D. spectabilis</i> *	Intermed.	1-3	2.1	1-2	January-August
<i>P. maniculatus</i>	Intermed.	1-6	4.29	2-4 <sup>†</sup>	Year-round
<i>Sigmodon hispidus</i>	Intermed.	2-10	5.6	1-9	Year-round
<i>S. fulviventor</i>	Intermed.	4-6 <sup>†</sup>	--	--	Year-round
<i>C. hispidus</i> *	Transient	4-7 <sup>†</sup>	--	1-2 <sup>†</sup>	--
<i>C. intermedius</i> *	Transient	1-7	3.94	--	March-July
<i>P. leucopus</i> *	Transient	2-5	3.61	>1 <sup>†</sup>	February-October
<i>R. montanus</i> *	Transient	1-9 <sup>†</sup>	4 <sup>†</sup>	--	Year-round <sup>†</sup>
<i>R. fulvescens</i> *	Transient	2-4	--	--	Year-round
<i>Baiomys taylori</i> *	Transient	1-5 <sup>†</sup>	2.49	1-9	Year-round
<i>S. ochrognathus</i>	Transient	2-6	--	--	Year-round <sup>†</sup>

708

709 **Table 2. Summary of reproductive life-history traits.** Data from Hoffmeister (1986), with  
 710 missing data filled from Wilson and Ruff (1999)<sup>†</sup>. Dashes (--) indicate no data from either  
 711 reference. Species marked with an asterisk (\*) are directly competing for resources in the  
 712 granivore feeding guild.

713

714

715

716

717

718

719

720



720 **Figure Legends**

721

722 **Figure 1. Hypothesized relationships between core-transient status and life history trade-**  
 723 **offs.** Temporal persistence is predicted to be correlated with life-history traits including survival,  
 724 reproduction, movement patterns, and resource use.

725

726 **Figure 2. The proportion of years and the mean proportion of months that a species was**  
 727 **present in the study period (1989-2009).** We designated core ( $\geq 0.66$ ), intermediate ( $< 0.66$  and  
 728  $> 0.33$ ) and transient ( $\leq 0.33$ ) status based on the proportion of years only. Two-letter  
 729 abbreviations refer to the species names (*B. taylori*, BA; *C. baileyi*, PB; *C. hispidus*, PH; *C.*  
 730 *intermedius*, PI; *C. penicillatus*, PP; *D. merriami*, DM; *D. ordii*, DO; *D. spectabilis*, DS; *N.*  
 731 *albigula*, NA; *O. leucogaster*, OL; *O. torridus*, OT; *P. flavus*, PF; *Pe. eremicus*, PE; *Pe.*  
 732 *leucopus*, PL; *Pe. maniculatus*, PM; *R. flavescens*, RF; *R. megalotis*, RM; *R. montanus*, RO; *S.*  
 733 *fulviventer*, SF; *S. hispidus*, SH; *S. ochrognathus*, SO).

734

735 **Figure 3. Histograms grouped by temporal persistence for the distances at which**  
 736 **individuals were recaptured each month.** Note that the y-axes are the percent of observations  
 737 for each bin, but that there is a large difference in total number of individuals captured in each  
 738 group, that is not represented in the histograms (but see Table 1).

739

740 **Figure 4. PCA biplot with confidence ellipses for temporal persistence.** The relationship of  
 741 movement ( $\Psi$  and benchmark), mean number of observed reproductive events for individuals of  
 742 each species per year (fecundity), apparent survival ( $\Phi$ ), recapture rates ( $p$ ), mean abundance,  
 743 and the proportion of years a species was present (persistence). Two letter abbreviations refer to

744 species names as presented in Figure 2. PCA axes 1 and 2 explained a total 64.6% of the  
745 variation.

746

747 **Appendix A.** Additional figures supporting the analyses and results in the manuscript.

748 **Supplementary Material.** Zip file containing the code and data for replicating the analyses.

749

750

## Figure Legends for Appendix A and the Supporting Information

**Figure A1. Panels showing the species-level movement distances calculated from the recapture data.** Two-letter species name abbreviations are as follows from Figure 2.

**Figure A2. Observed yearly reproductive effort for individuals of each species from the field study.** The y-axis represents the proportion females that we tracked that were recorded in reproductive condition 0-4 times per calendar year (no individuals were recorded as having > 4 reproductive events within a calendar year). Two-letter species name abbreviations are as follows from Figure 2.

**Figure A3. PCA biplot with confidence ellipses for phylogenetic relatedness at the family level.** The relationship of movement ( $\Psi$ , benchmark), mean number of reproductive events individuals of a species had per year (fecundity), apparent survival ( $\Phi$ ), recapture rates ( $p$ ), mean abundance, and the proportion of years a species was present (persistence). Two letter abbreviations refer to species names as presented in Figure 2. PCA axes 1 and 2 explained a total of 64.6% of the variation.

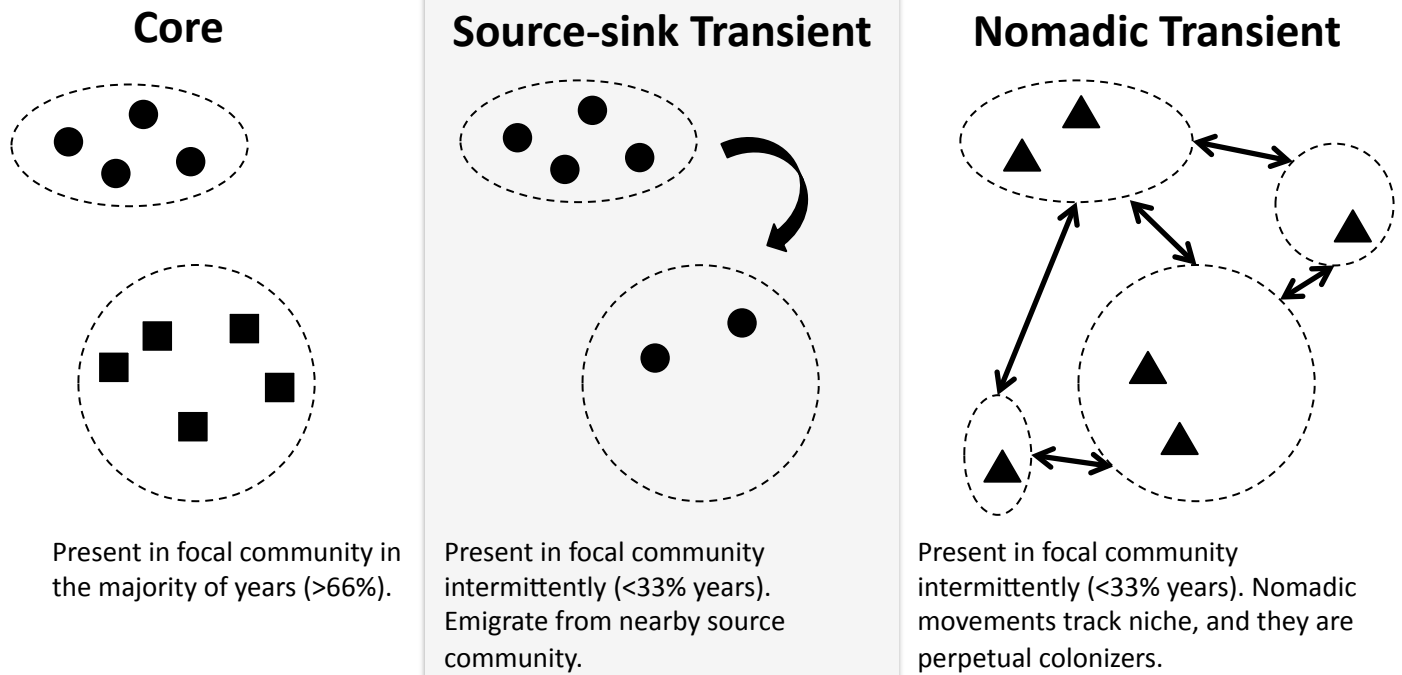
**Figure A4. Apparent survival estimates ( $\Phi$ ) shown against the  $\log(Y+1)$  distance (benchmark) moved for each species and probability of moving a relatively long distance ( $\psi$ ) as estimated using RMark.** The plotted variables have been standardized. Points represent core (pink circles), intermediate (green triangles) and transient (blue squares) species.

**Figure A5. Movement estimates (benchmark,  $\psi$ ) shown against fecundity for each species.** Evidence for a reproduction-movement trade-off is not statistically significant. The plotted variables have been standardized. Points represent core (pink circles), intermediate (green triangles) and transient (blue squares) species.

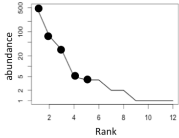
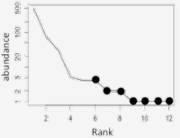
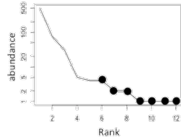
**Figure A6. Relationship between the CMR estimates for apparent survival ( $\Phi$ ), movement ( $\Psi$ ) and recapture ( $p$ ) probability from the CMR analysis for each species, plotted with error bars representing standard error for each estimate.** Points represent core (pink), intermediate (green) and transient (blue) species. Note that the transient species have the largest error bars.

**File S1. Zip file containing the code and data for replicating the analyses.** Code and data are also found in a public GitHub repository (<https://github.com/weecology/portal-rodent-dispersal>).

Figure 1

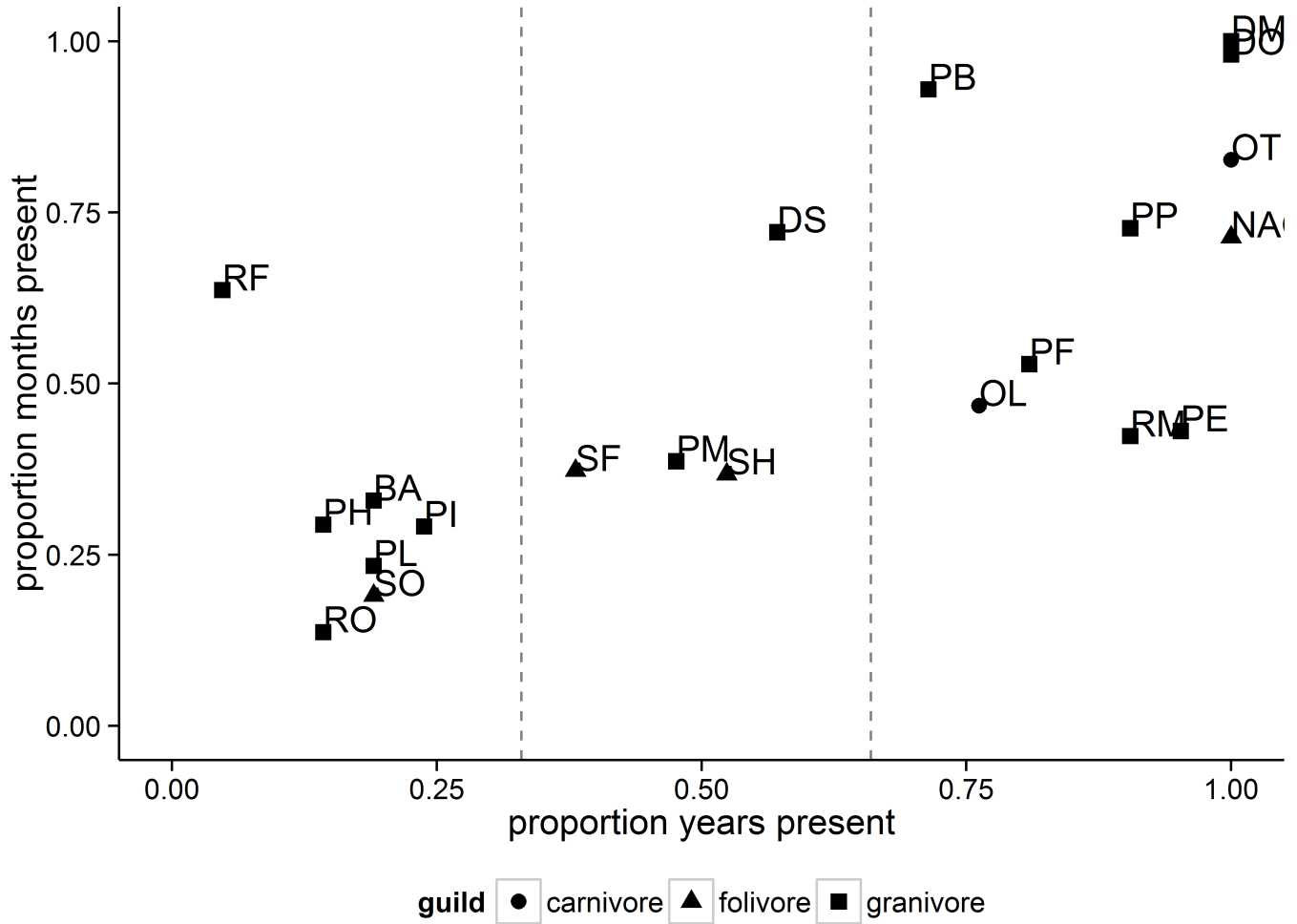


Species fall along a continuum of traits and temporal persistence

Highly adapted to habitat/resources in focal community	Adapted to habitat/resources in source community, but not necessarily to focal community	Habitat/resource generalist that opportunistically colonizes in response to available niches
Generally dominant and abundant in focal community 	May become abundant for short time periods, but are generally rare 	May become abundant for short time periods, but are generally rare 
Competitively dominant	Competitively inferior	Competitively inferior
Self investment Strategy High survival Low fecundity	High survival in source community Lower survival in focal community Low fecundity	Offspring investment Strategy Low survival High fecundity
Strong fidelity to "home"	Colonize population sinks in response to density dependent dynamics in source community	Strong colonizers
Movements tied strongly to defending home resources/territory. Reduced movement.	Movements tied strongly to obtaining and defending home territory. Reduced movement.	Movements tied to obtaining resources and exploring available territory/niches. Increased movement tendencies.
Relatively low gene flow among populations	Relatively low gene flow among populations	Intermediate to high gene flow among populations

Circular habitat is "focal" community. Dashed lines represent that communities are open. Core-transient designation is a concept defined within a focal community, but we hypothesize that temporal persistence should be linked to traits.

Figure 2



**Figure 3**

core

intermediate

transient

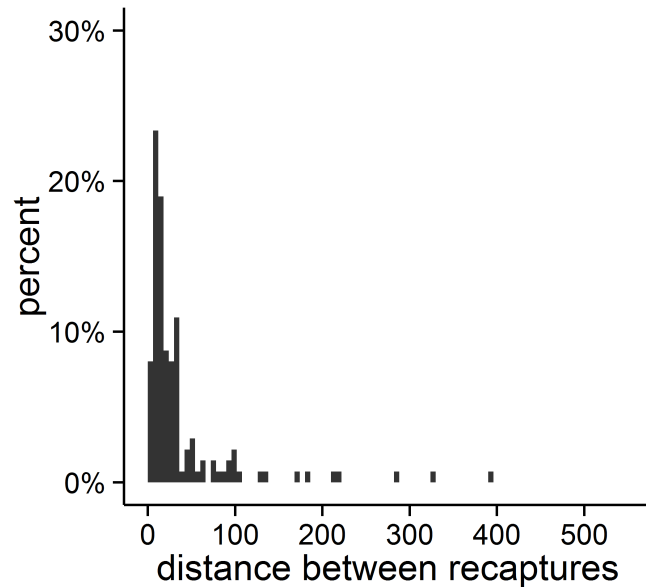
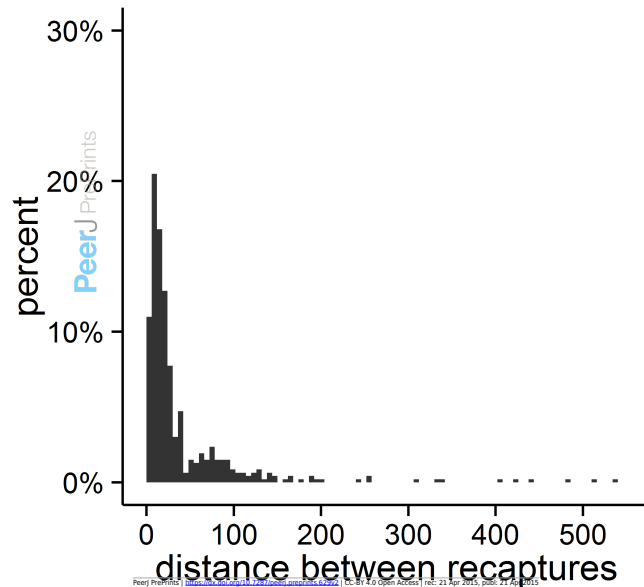
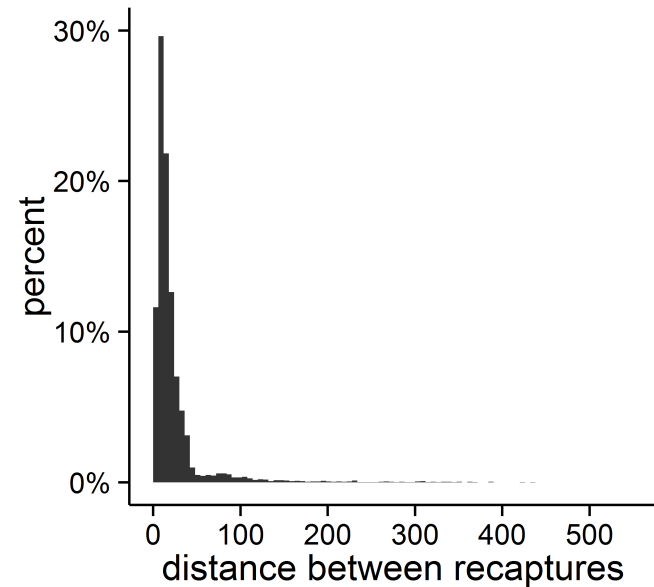


Figure 4

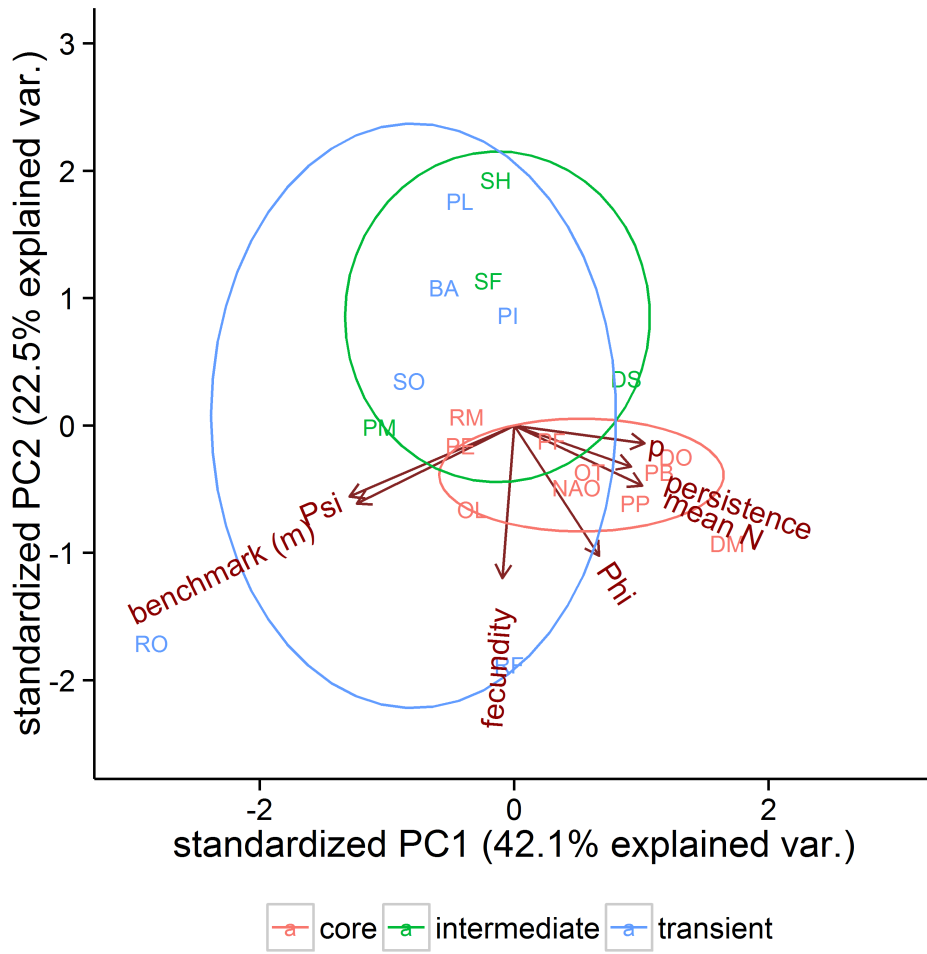




Figure A1

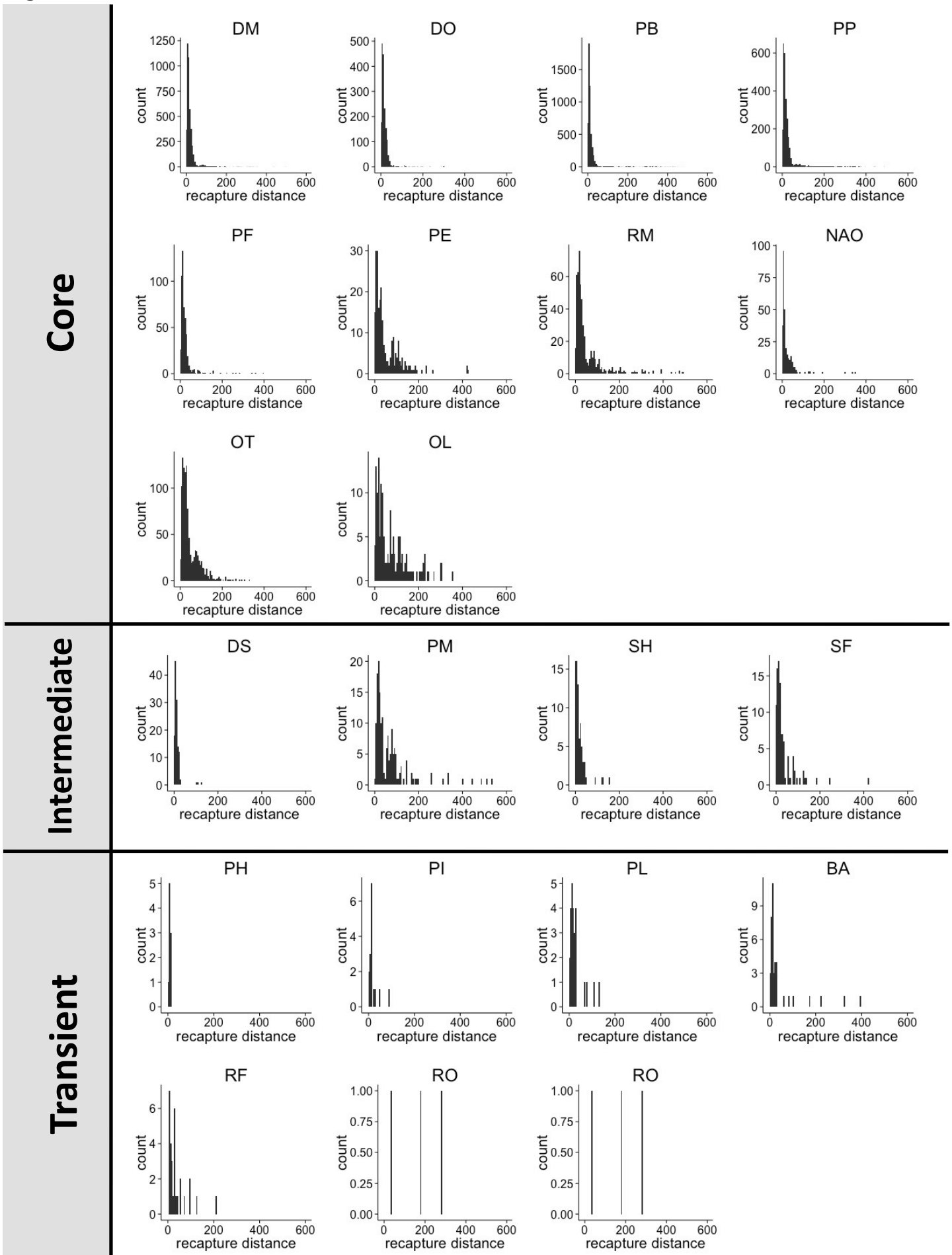


Figure A2

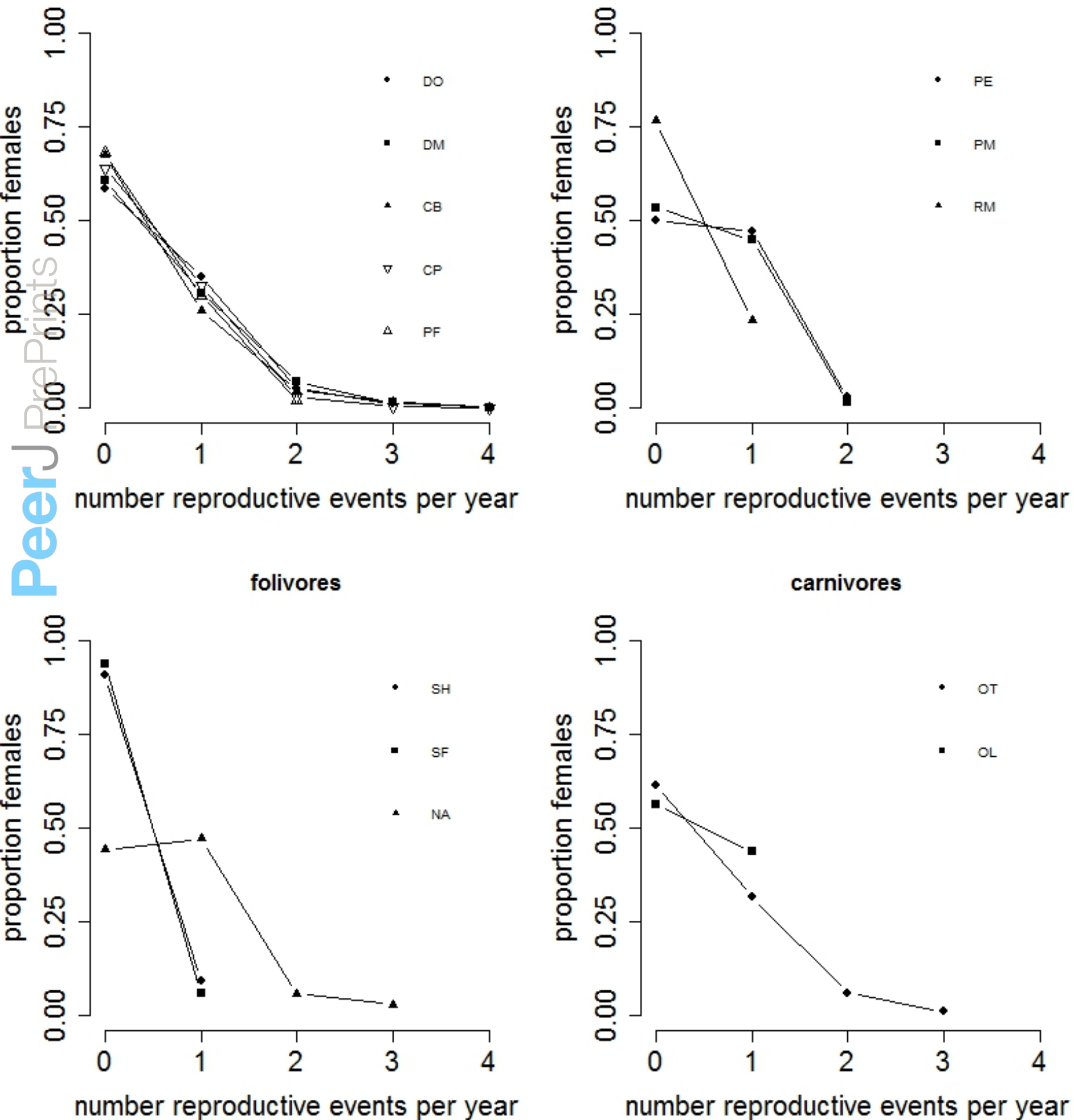


Figure A3

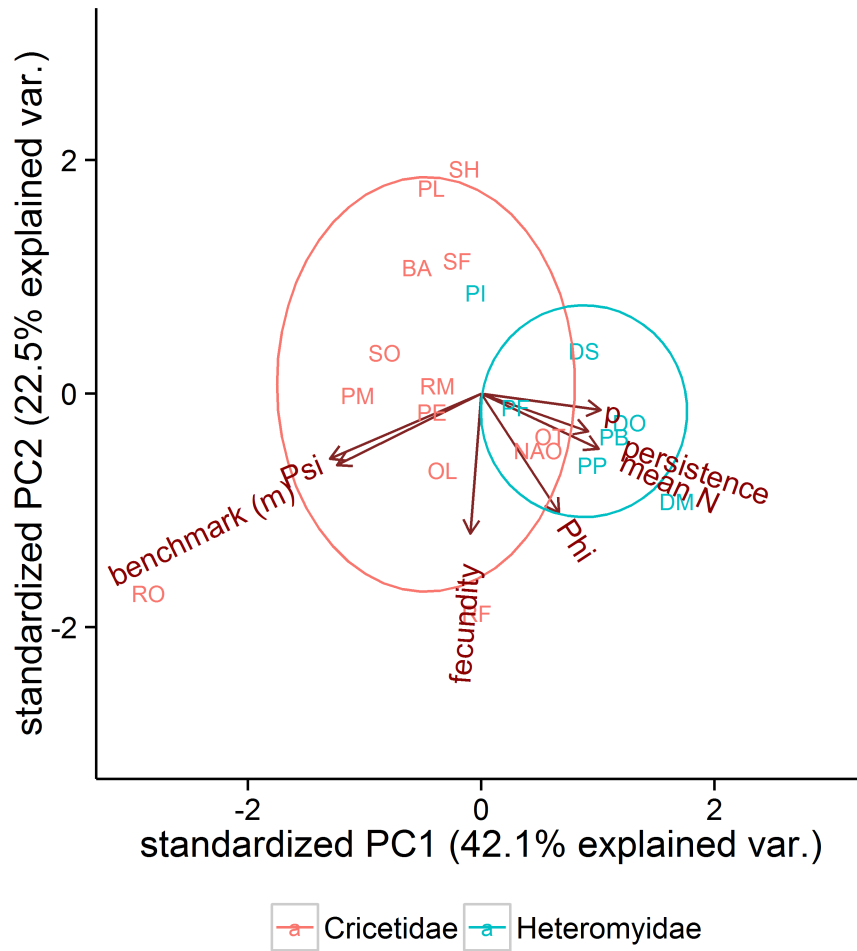


Figure A4

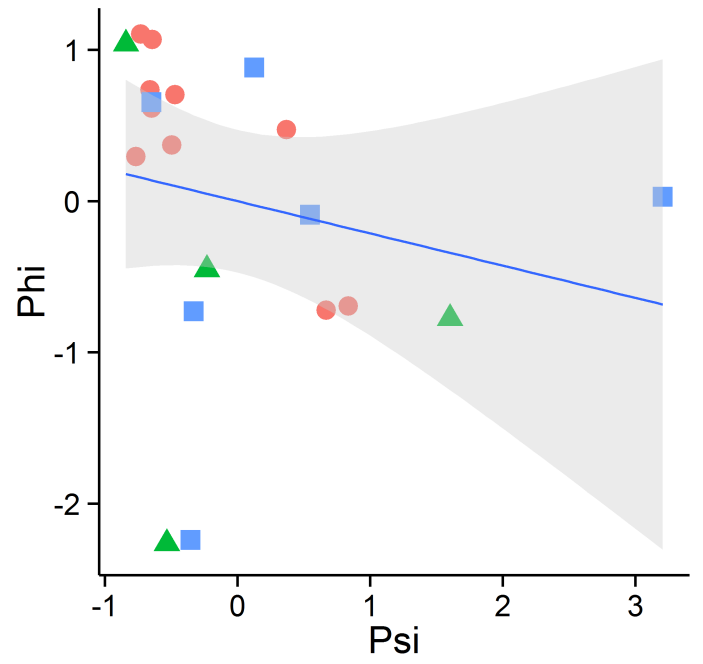
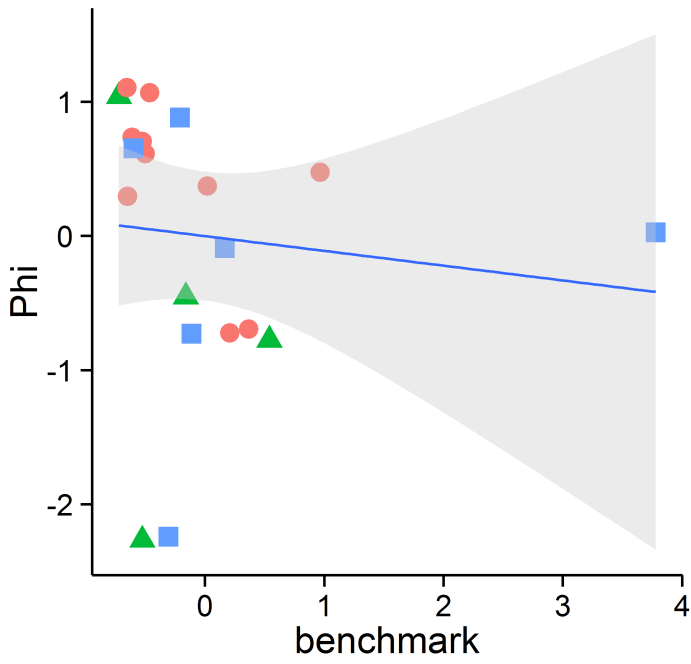


Figure A5

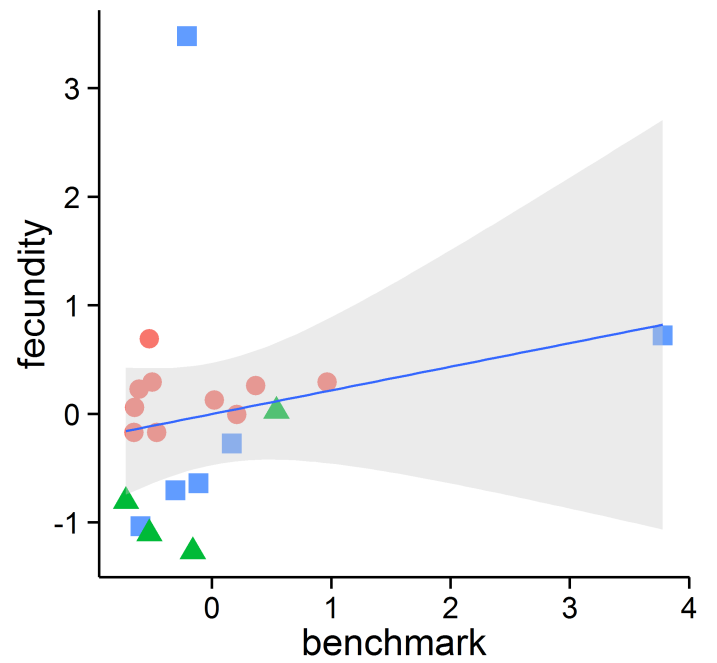
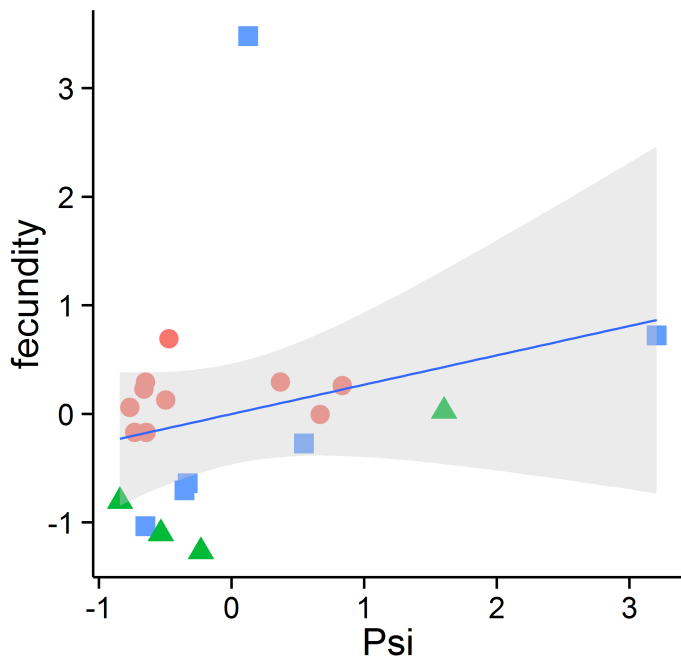


Figure A6

