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Piecing together the biogeographic history of *Chenopodium vulvaria* L. using botanical literature and collections

This study demonstrates the value of legacy literature and historic collections as a source of data on environmental history. *Chenopodium vulvaria* L. has declined in Northern Europe and is of conservation concern in several countries, whereas in other countries it has naturalised and is considered an alien weed. It is hypothesised that much of its former distribution was the result of repeated introductions from its native range in southern Europe and that its decline in northern Europe is the result of habitat change and a reduction in number of propagules imported to the north. An historical analysis of its ecology and distribution was conducted by mining legacy literature and historic botanical collections. Text analysis of habitat descriptions written on specimens and published in botanical literature covering a period of more than 200 years indicate that the habitat and introduction pathways of *C. vulvaria* have changed with time. Using the naturalised alien range in a climate niche model it is possible to project the range in Europe. By comparing this predicted model with a similar model created from all observations it is clear that there is a large discrepancy between the realised and predicted distributions. It is concluded that if *C. vulvaria* was native to northern Europe, then it was only ever a rare species, however it was more common in the 18th and 19th centuries due to a combination of repeated introductions and the creation of suitable habitats by people.

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4 Introduction

5 Legacy biodiversity literature is a potential mine of useful information on the past distributions of
6 organisms. While these texts have always been available in academic libraries, their accessibility
7 and discoverability has been significantly enhanced by projects such as the Biodiversity Heritage
8 Library (www.biodiversitylibrary.org) and other online digital libraries. The ability to search a
9 whole corpus of historical literature for a Latin name of an organism dramatically increases the
10 accessibility of this scientific information and makes literature searches possible that once would
11 have been unfeasible. In parallel, the widespread digital imaging of herbarium specimens and
12 transcription of their labels has also made these data considerably more accessible, which,
13 combined with historic literature, has created a large pool of information from which the
14 phytogeographic historian can draw evidence (Vellend et al., 2013).

15 *Chenopodium vulvaria* L., a small, inconspicuous species that grows largely in places disturbed
16 by mankind. It is not remarkable morphologically, but it is nonetheless distinctive due to its foul
17 smell, which is described as similar to that of rotten fish. Its distinctiveness makes it particularly
18 suited to a study using historic literature, because there is less concern that published accounts
19 refer to other species as a result of misidentification.

20 *C. vulvaria* is currently a red-listed species in several countries including Sweden
21 (www.artfakta.se), the United Kingdom (Cheffings *et al.*, 2005), Belgium (Kestemont, 2010),
22 Luxembourg (Colling, 2005), Czech Republic (Grulich, 2012) and some regions of France
23 (Ferrez, 2005). In contrast, it has naturalised in California (Calflora, 2014), Argentina
24 (Planchuelo, 1975; Giusti, 1997), Chile (Boelcke et al., 1985) and Australia (Atlas of Living
25 Australia, 2014). *C. vulvaria* is widespread in countries bordering the Mediterranean and
26 eastward to Afghanistan and Mongolia (Jalas & Suominen 1980; Meusel, Jäger & Weinert, 1992).
27 Yet it is clear from historical literature that it was common in parts of northern Europe during the
28 18th and 19th centuries. Turner (1548) wrote “*It groweth muche aboute the walles in Bon in*
29 *Germany*”; Bucher (1806) wrote in the Flora of Dresden “*An den strassen der vorstadt und sonst*
30 *gemein*” translated as “By the streets of suburbs and usually common”; Curtis (1777) stated “*This*
31 *species is very common in the neighbourhood of London...*” and Hooker (1821), in his flora of
32 Scotland, describes it as “*frequent*”.

33 The native distribution of *C. vulvaria* is unknown and its long association with man-made

34 disturbance makes this impossible to determine. Floras in Northern and Central Europe variously
35 describe it either as a native or an archaeophyte, though the evidence for categorizing it in either
36 category is slim and is probably based on the anthropogenic habitats that *C. vulvaria* often
37 inhabits.

38 Many other members of the Amaranthaceae live in disturbed, nutrient rich habitats and may be
39 halophytic. *C. vulvaria* itself is often found in disturbed, eutrophic and coastal habitats. In
40 general, species of such habitats are increasing and spreading in northern Europe (Wróbel,
41 Tomaszewicz & Chudecka, 2006; Van Landuyt et al., 2008; Smart et al., 2003; Šerá, 2011;
42 Groom, 2013). So at face value, *C. vulvaria* appears well adapted to modern habitats in Europe
43 and yet it has declined.

44 One possible explanation for its apparent decline in northern and central Europe may be a
45 misunderstanding of its former occurrence, its presence in the north being the result of propagule
46 pressure from its heartland in southern Europe, constantly reinforcing the introduced populations
47 in the north. One or many introduction pathways may have existed that delivered *C. vulvaria* seed
48 outside of its normal range and these pathways have since reduced in importance, causing a
49 collapse in the population. Another possible explanation is change to its former habitat, though
50 the details of its ecology are too poorly known to understand what these changes may have been.

51 For non-woody plants there are few sources of data to examine recent biogeographic change.
52 Palynology and the study of archaeological remains can be useful, but many species do not have
53 a sufficiently distinctive anatomy to identify them from their remains. In these cases, historical
54 literature and collections may be the only source of data on their former habitats and locations.
55 Given the shortage of data an alternative approach, widely used to model the potential
56 distribution of organisms, is bioclimatic modelling. Many studies have used observations from
57 the known native range of a species to extrapolate its potential invasive range (e.g. Macfadyen &
58 Kriticos, 2012). In ecological theory the potential bioclimatic range is generally considered to be
59 larger than the realised distribution as a consequence of additional non-climatic limitations to
60 distribution, such as edaphic factors (Araujo & Peterson, 2012). However, in the case of *C.*
61 *vulvaria* the native range is not known and frequent non-persistent introductions mean that the
62 realised distribution predicted from observations may be larger than its true bioclimatic range.
63 For *C. vulvaria* the location of naturalisation in Australia, North America and South America
64 might be a clearer indication of its bioclimatic range than within Europe, where it is hard to
65 distinguish established from casual occurrences. Assuming that this species is well established

66 and stable in its alien range, we can use the known naturalised range to model the climate
67 envelope and extrapolate this to Europe to identify the areas where the climate is suitable for *C.*
68 *vulvaria*. In this manner we can indicate those places where this species has been observed but is
69 unlikely to be persistent.

70 My hypothesis is that *C. vulvaria* was formally more abundant in northern Europe and its current
71 decline is the result of changes in the introduction pathways and loss of habitat. In this study I
72 draw on botanical literature and specimens to identify habitat change and historic introduction
73 pathways. I use text analysis of habitat descriptions to demonstrate how its habitat has changed
74 over the past 200 years and I use bioclimatic niche modelling to contrast the realised range within
75 Europe with the projected range based upon naturalised occurrences outside Europe.

76 **Methods**

77 Observation and specimen details were collected in a Common Data Model (CDM) database
78 which is the central component of the EDIT Platform for Cybertaxonomy (Ciardelli et al., 2009;
79 Berendsohn et al., 2011). Two methods were used to extract observations from literature, either
80 XML markup or direct data entry. Digitised treatments were marked up with XML using the
81 GoldenGate editor (<http://plazi.org/?q=GoldenGATE>, Sautter, Böhm & Agosti, 2007); uploaded
82 to the PLAZI taxonomic treatment repository (plazi.org) and imported to the CDM database.
83 Alternatively the observation details were copied from the treatment and entered manually into
84 the CDM database using the EDIT Taxonomic Editor (Ciardelli et al., 2009). Observations where
85 gathered from the biodiversity literature by reading the BHL corpus systematically after
86 searching for *Chenopodium vulvaria* L. and its synonym *Chenopodium olidum* Curt. Other
87 published observations were gathered from publications in the Library of the Botanic Garden,
88 Meise. A list of the sources of observations of *C. vulvaria* is available in supplementary file S3. A
89 complete survey of non-digitised literature is unfeasible, but there was an effort to check multiple
90 floras of every European country and any other country with a temperate climate suitable for *C.*
91 *vulvaria*.

92 Digitised observation data was also gathered from databases, primarily from GBIF (data.gbif.org,
93 accessed 08 Nov 2013; see appendix), but also from the Atlas of Living Australia (2014); the
94 Botanical Society of Britain and Ireland (2013) and Herbaria@home (2013). Scientific articles
95 and websites containing observations were also discovered using search engines

96 (scholar.google.be; google.be). Data from databases were standardised and imported directly into
97 the CDM database.

98 Specimen data were gathered from herbaria by transcription of label information. Specimens
99 from the following herbaria are included in the study, their names and abbreviations follow the
100 Index Herbariorum (<http://sciweb.nybg.org/science2/IndexHerbariorum.asp>). Botanical Garden
101 Meise (BR); Botanical Museum Berlin-Dahlem (B); Botanische Staatssammlung München (M);
102 Bulgarian Academy of Sciences (SOM); Bulgarian Academy of Sciences (SOMF); Charles
103 University in Prague (PRC); Herbar J.H. Fabre (FABR); Institut Botànic de Barcelona (BC);
104 Institute of Biodiversity and Ecosystem Research, Nationaal Herbarium Nederland (L); Moscow
105 State University (MW); Museu Nacional de História Natural e da Ciência (LISU); Museum
106 National d'Histoire Naturelle (P); National Academy of Science, Kyrgyzstan (FRU); Natural
107 History Museum, London (BM); Natural History Museum of Denmark (C); New York State
108 Museum (NYS); Reading University (RNG); Royal Botanical Gardens, Kew (K); Sapienza
109 University of Rome (HFLA), Sofia University (SO); South London Botanical Institute (SLBI);
110 Universidad Nacional del Sur Herbario (BBB); Universität Wien (WU); Universidad de
111 Concepción, Chile (CONC); University of Alaska Herbarium (ALA); University of Birmingham
112 (BIRM); University of California (UC); University of British Columbia (UBC); University of
113 Manchester (MANCH); University of Wales (ABS); Wageningen University (WAG) and others
114 contributing data to Global Biodiversity Information Facility (GBIF) (supplementary files S2).
115 Many other herbaria and herbarium catalogues were searched without finding specimens and
116 several herbaria were contacted and either contained no specimens or did not respond.
117 Undoubtedly there are more specimens and observations of *C. vulvaria* to be discovered, but I
118 believe these to be a representative sample and a large proportion of those that exist. Undated
119 specimens were not used in the study; however, it is usual for published observations to be
120 undated. Therefore the publication date was used for undated observations in literature. Studying
121 biographical information of collectors it is clear that most undated observations in old floras are
122 within 35 years of the publication date and authors tend to provide dates when they are not recent.
123 In total 2456 observations were collected from specimens and literature. These data span 465
124 years from 1548 to 2013, though there are only two observations from the 16th century, two from
125 the 17th century and nineteen from the 18th century.

126 Text analysis of habitats

127 The text describing the habitat of *C. vulvaria* was collected from 104 floras, 33 scientific articles,
128 119 specimens and 5 websites, covering the years 1787 to 2014. The texts were written in 12
129 languages, English (35%) German (20%), French (17%), Latin (12%), Dutch (4%), Italian (3%),
130 Portuguese (3%), Spanish (3%), Hungarian (1%), Danish (1%), Catalan (1%) and Czech (<1%).
131 Each description was broken down into tokens consisting of either single words or short phrases
132 describing a single aspect of the habitat. Thus the description “*In Straßen, an Häusern,*
133 *Stallungen, Düngerstätten*” was broken down into the tokens “*Straßen*” (roads), “*an Häusern*”
134 (near houses), “*Stallungen*” (stables) and “*Düngerstätten*” (mature heaps). This process created
135 475 habitat tokens. These tokens were then translated to English using native speakers of English,
136 German, French and Dutch and for other languages a combination of Google Translate
137 (translate.google.be) and the multilingual collaborative dictionary Wiktionary (wiktionary.org).
138 To conduct the analysis it was necessary to reduce the number of habitat terms, which was done
139 in two stages. The anglicized tokens were first simplified to closely related terms. Thus the terms
140 “*by foot of the city walls*”, “*along walls*”, “*under walls*”, “*mud walls*”, “*foot of walls*”, “*under*
141 *walls*”, “*under a wall*” and “*foot of the church yard wall*” were all replaced by “*by walls*”. This
142 process reduced the number of habitat words to fifty. These fifty words were then arranged into
143 logically related categories. Thus “*by walls*”, “*by fences*” and “*by hedges*” were grouped together
144 under the term “*boundaries*”. This reduced the number of habitat categories to fifteen (animal
145 waste, boundaries (including walls), coastal, disturbed and grazed land, dry & bare soil,
146 habitation, hills, horticulture, industry, rail, roads, sand and rock, shipping, waste, wetland. A full
147 list of the tokens contributing to each category is provided in the appendix (Table S1).
148 Throughout the process the tokens were kept associated with the date; either the year the
149 specimen was collected, observed or the year of publication. To analyse the use of habitat words
150 in the collected corpus the simplified habitat terms were pooled into 20 year periods from 1780
151 onwards. The proportional use of each habitat term was then calculated for each period.
152 Statistical analysis was conducted in R (version 2.15.2) using generalized linear modelling with
153 binomial errors, weighted with the number of tokens contributing to each pool. All models were
154 checked for overdispersion using the ratio of the residual deviance and the degrees of freedom,
155 but none were found to be overdispersed.

156 Analysis of distribution

157 Except for the rare occasions when coordinates were available with the specimen or

158 observation, georeferencing was carried out manually according to best practise (Chapman &
159 Wieczorek, 2006). Error radii for coordinates were not available for most records in databases,
160 but they were estimated for the coordinates georeferenced in this study, however, they were not
161 used to select data for the analysis. The average error radius was 11 km and the mode and median
162 were both 10 km. *C. vulvaria* is a largely lowland species and errors in georeferencing of these
163 magnitudes are insignificant for bioclimatic modelling at a global scale compared to the other
164 inherent biases in these data.

165 Species distribution modelling was conducted using the BioVel Ecological niche modelling
166 workflow and services (www.biovel.eu). The ecological niche modelling workflows were run on
167 6th Aug 2014. BioVeL is funded by the EU's Seventh Framework Program, grant no. 283359. The
168 workflow uses the Maxent method based upon Phillips, Dudik & Schapire (2004) and using the
169 openModeller web service (de Souza Muñoz et al., 2011). Models were created using the default
170 parameter and all 19 layers of the WorldClim global climate layers 10 arc minutes, version 1.4,
171 release 3 (Hijmans, 2005).

172 Non-European observations used for modelling were only those locations where it was clear,
173 either from the notes on the specimens or from floras, that the species forms persistent population
174 at these sites. If there was any doubt to the status modern floras were consulted to ascertain the
175 persistence of the species in the area. The locations with non-native populations outside Europe
176 were Southern Argentina; California; Chile; South Australia; Tasmania; Tierra del Fuego
177 (Argentina) and Victoria (Australia). *C. vulvaria* is also recorded from South Africa and New
178 Zealand, but its status there is not clear. It is also believed to occur natively in Mongolia but only
179 one observation was found. A total of 42 observations from the naturalised range were used to
180 model the range. However, weeding of duplicates during the workflow reduced the number to 32.
181 The dates of these records were from 1863 to 2012, though 86% dated from 1950 onward. For
182 modelling the realised range, all global observations were used which resulted in 1894
183 observations after weeding of duplicates.

184 **Results**

185 **Text analysis**

186 Four habitat categories were notably more frequent than the others (Fig. 1). These categories are

187 firstly waste, including rubbish piles, rubble, ruins and waste places of all kinds; secondly,
188 boundaries, mainly at the base of walls; thirdly, roads and roadsides, including streets and farm
189 tracks; and fourthly horticulture, such as gardens and other cultivated places. The habitat
190 categories in Fig. 1 are not mutually exclusive, but often describe different aspects of the same
191 habitat such as the proximity to landscape features, soil type, nutrient status and moisture.
192 In summary, the habitat analysis underscores several aspects. *C. vulvaria* is strongly associated
193 with mankind, natural habitats such as coastal and wetlands are mentioned infrequently. It is
194 intolerant of competition; none of the habitats are defined by other vegetation, such as meadows,
195 woodland or heaths. It is frequently associated with transport routes and it is usually associated
196 with some form of soil disturbance.
197 When the use of these terms was compared over time, no significant change was found for the
198 use of terms relating to animal waste, coastal, dry & bare soil, habitation, hills, horticulture,
199 industry, rail, roads, shipping and waste. Figure 2 shows the changes of eight of these categories,
200 including the only four where there were significant changes. The significant changes were
201 increases in the proportion of the terms related to wetland ($p < 0.01$, $DF=11$), sand and rock ($p <$
202 0.05 , $DF=11$) and disturbed and grazed land ($p < 0.05$, $DF=11$), whereas there has been a
203 significant decrease in the proportion of terms related to boundaries ($p < 0.001$, $DF=11$). Of these
204 significant changes only terms relating to boundaries were also highly frequent in the corpus (Fig.
205 1).

206 Introduction vectors, pathways and origins

207 Clear expressions of the introduction vector were rare on specimens and in publications. Where
208 introduction vectors are evident they are summarised in Table 1. Ballast soil at ports was the
209 earliest vector mentioned in the corpus and it was also most frequently mentioned. However, it
210 stops being mentioned in the early 20th century. Several specimens and observations implicate the
211 transport of ore. *C. vulvaria* was reported on Chromite in Baltimore, USA between 1953–1958
212 (Reed, 1964); in Norway in 1954 (Uotila, 2001); on manganese ore in Norway between 1931–
213 1935 and near an ore crushing plant in Kyrgyzstan in 1961 (Lazkov, Sennikov & Naumenko,
214 2014). Various agricultural products are mentioned as vectors such as grain, but no mention of its
215 introduction as an herbal medicine or other produces commonly imported from the
216 Mediterranean such as tobacco, even though *C. vulvaria* is frequently associated with waste.

217 Evidence for the pathways of introductions is scant, but shipping and railways are mentioned.
218 Although roads are the most frequently mentioned transport system (Fig. 1), it is unclear if the
219 presence of this species on roads relates to the introduction pathway or whether roads just provide
220 suitable habitat.

221 Evidence for the origin of introductions is also slim, though where the origin is mentioned it is
222 always from a country in the Mediterranean region (Uotila, 2001). There is no evidence of return
223 introductions from naturalised populations outside Europe.

224 **Comparing actual climatic niche and realised distribution**

225 The observations of *C. vulvaria* within Europe are from an inseparable mixture of stable
226 populations and casual occurrences. It is therefore impossible to validate a model for the true
227 climatic niche of *C. vulvaria*. For this reason I have not attempted to refine the output of the
228 models by adjusting their default parameters or eliminating climate layers. It is nevertheless
229 informative to contrast models created from the known naturalised range outside Europe with the
230 realised range within Europe (Fig. 3, Fig. 4). The actual climatic niche, predicted from
231 observations from the naturalised range outside Europe predicts the presence of *C. vulvaria* in
232 southern and western Europe, North Africa and the Middle-east, notably, Spain, western France
233 and Turkey (Fig. 3). The actual observations and the climate niche model created from them
234 show a much wider distribution, which extends much further north and eastward than the niche
235 model created from the naturalised range (Fig. 3, Fig. 4). Figure 5 shows the locations of actual
236 observations and the dates they were made. It demonstrates that there has been a general decline
237 in the number of observations from northern Europe, but it also suggests unevenness in surveying
238 effort between different countries and different time periods.

239 **Discussion**

240 This study tracks the distribution and habitat changes of *C. vulvaria* over more than 200 years.
241 Over this period botanical literature becomes more common and sufficiently abundant for
242 analysis. Simultaneously botanical specimens became more frequently collected and better
243 documented, further adding to the analysable corpus of historical documents.

244 Over the past two centuries many social, economic and technological changes have occurred that
245 may have influenced the abundance and distribution of *C. vulvaria*. Some key events in this
246 period are the expansion of the railway network in the 19th century, the adoption of motorised
247 road transport in the early 20th century, the decline in the uses of horses for transport and

248 agriculture in the 20th century; the transition from sail to steam powered ships at the turn of the
249 20th century; the discovery of herbicides in the mid-20th century and the Green Revolution in the
250 latter half of the 20th century. *C. vulvaria* is an anthropophilic species and to some extent benefits
251 from this association, however, for the same reason it will be more acutely affected by changes in
252 human culture than many other species.

253 Text analysis was able to identify key habitat features of *C. vulvaria*. This species has been, and
254 still is, strongly associated with mankind, both as a weed of cultivation and as a ruderal plant.
255 The analysis identifies habitat traits such as its avoidance of competition and the association with
256 waste. The genus *Chenopodium* is considered to be nitrophilous, indeed *C. vulvaria*, is sometimes
257 associated with habitats linked to animal dung, however, it is much more commonly associated
258 with other types of waste or cultivated place (Fig. 1).

259 The temporal analysis of habitat change indicates that *C. vulvaria* is still associated with many of
260 the same habitats it was in the past, such as agriculture, transport and waste (Fig. 2). However, in
261 the 20th century habitat descriptions have included proportionally more words related to natural or
262 semi-natural habitats, such as grazing, sand and wetland.

263 The reference to wetland amongst the habitats needs further explanation, because *C. vulvaria* is
264 not a typical wetland plant. It does not grow in water, but colonises bare soil exposed in the
265 summer at the margins of rivers, ditches and lakes. Thus its association with wetland is of an
266 opportunistic colonizer of habitats free from competition, rather than a true wetland plant.

267 The habitat where *C. vulvaria* has declined is along boundaries, particularly along walls, which
268 contributed 80% of the boundary terms. *C. vulvaria* does not grow on walls, but beside them,
269 which appears at first sight to be a rather non-specific habitat description. However, the margins
270 of walls have changed considerable in the past 200 years. Walls were once built using lime
271 mortar, rather than cement, and were frequently painted with whitewash, a mixture of calcium
272 hydroxide and [chalk](#). Whitewash gave the traditional white or pink colour to houses throughout
273 Europe. Consequently, the soil in the immediate vicinity of walls would have been alkaline. *C.*
274 *vulvaria* is not known as an alkaliphile, however it is clearly tolerant of high pH as it has been
275 found on the ultrabasic rock chromite (Reed, 1964). Furthermore, because horses were used for
276 transport and farm animals were driven along roads, the base of walls would have been strewn
277 with animal waste. Such fertile alkaline habits do not occur by walls in modern towns and we can
278 speculate that the technical changes in building practises and changes to transportation have
279 contributed to the decline of *C. vulvaria*.

280 Text analysis is clearly a useful tool for environmental historians, nevertheless, it is susceptible to
281 the fallibility and biases of authors, who may uncritically follow their forbears or write from
282 hearsay rather than experience. Also, botanical activity is spatially and temporarily biased. For
283 example, British and German botanical literature has, and continues to be, more abundant than
284 for other countries in Europe.

285 Compared to the analysis of habitat, evidence for introduction vectors, pathways and origins was
286 limited. The results show that there were multiple vectors introducing *C. vulvaria* to northern
287 Europe, but particularly as a grain contaminant and in ship's ballast. The frequent occurrence of
288 *C. vulvaria* in waste perhaps indicates that its seeds were contaminants of many crops. Indeed,
289 different specimens mentioned *C. vulvaria* in crops of lentils and potatoes. Unfortunately, the
290 source of a casual introduction is rarely obvious by the time the plant is mature. Weed species
291 that are dispersed as seed contaminants have declined throughout Europe in the 20th century; this
292 is, in part, a consequence of improved seed cleaning methods (Hilbig, 1987; Sutcliffe & Kay,
293 2000; Lososová, 2003). Most of these species are considered archaeophytes to northern Europe.

294 Soil was used as ballast on sailing ships during the 18th and 19th centuries to provide stability to
295 cargo ships when not carrying heavy loads. In ports, where heavy materials were loaded, ballast
296 was removed and replaced by cargo. Large hills of ballast soil were a common feature of busy
297 ports, particularly in areas of mining and heavy industry, such as in northern Europe. These
298 ballast hills were a large reservoir of propagules for many species (Carlton, 2011). The large
299 number of specimens and observations reflects the importance of this invasion pathway, but
300 might be somewhat over-represented because botanists were attracted to ballast heaps as a source
301 of novel species and because the vector of the propagules is clear in this case.

302 Ore is also mentioned as an introduction vector to the USA and Norway. Chromium processing
303 began in Baltimore, USA in 1822, at which time only local chromium ore deposits were
304 processed (Newcomb, 1994). However, by the end of the 19th century local chromium deposits
305 were exhausted and processing continued with imported ore until the end of the 20th century.
306 Similarly, Norway is also a large processor of imported chromium ore, for example in 1992 the
307 country imported 187,965 tonnes of chromite ore from Turkey (Plachy, 1992). Indeed, it is likely
308 that some of the chromite imported into Baltimore was also from Turkey where chromite was
309 first mined in 19th century (Zengin, 1957). Therefore, it seems that exports of chromite from
310 Turkey could have been a pathway for dispersal of *C. vulvaria* during the 20th century.

311 Animal dung is often mentioned as a growing medium for *C. vulvaria*, which is indirect evidence
312 for endozoochory. Certainly, other *Chenopodium* species are dispersed in this manner and *C.*
313 *vulvaria* is eaten by ruminants despite its smell (Withering 1776; Haarmeyer et al., 2010). In the
314 21st century yet another vector of *C. vulvaria* introduction has been created, that of imported
315 Olive trees (Hoste et al., 2009). These mature trees are extracted from olive groves with a large
316 amount of soil and are sold in northern Europe as horticultural novelties.

317 Though dispersal vectors are rarely mentioned in the corpus, it is clear that *C. vulvaria* has been
318 dispersed by a wide variety of vectors and through a number of pathways (Table 1). There are
319 historic periods associated with each vector and if this analysis was extended to more species,
320 one would be able to further refine the time frames during which these pathways were operating.
321 From the diversity of distribution vectors it is clear that *C. vulvaria* has been widely introduced
322 outside its natural climatic range and it grows often temporarily. However, with the exception of
323 horticultural imports, introduction pathways of *C. vulvaria* ended midway through the 20th
324 century.

325 The sporadic occurrence of *C. vulvaria* presents a problem for the selection of occurrences for
326 distribution modelling. Unless all casual occurrences are eliminated from the data before fitting,
327 the model would indicate a much broader climatic range. Separating permanent populations from
328 casual occurrences is impossible for Europe where anthropogenic disturbance and trade have
329 confused the quasi-natural distribution. However, in the naturalised range the situation is much
330 clearer. Most, if not all, modern observations of *C. vulvaria* in California, Australia and South
331 America appear to be from naturalised populations, that is to say, the associated meta-data
332 indicates the presence of a population and there no indication of a recent introduction. Therefore,
333 the naturalised distribution outside Europe should reflect the true climatic niche of the species, as
334 long as the distribution is at equilibrium. This assumption seems reasonable since old casual
335 records of *C. vulvaria* occur throughout the world, but naturalised populations persist in only a
336 few of those places. Clearly, introduction events were occurring all over the world for several
337 hundred years of international trade, but *C. vulvaria* only naturalised in a few of those places
338 where the habitat and climate suited it.

339 Projecting the bioclimatic range in Europe from naturalised alien populations elsewhere predict a
340 much more southern and western distribution of *C. vulvaria*, than the modelling using all
341 occurrences (Fig. 3, Fig. 4). Yet, these rather crude models indicate that the naturalised
342 distribution of *C. vulvaria* has a climate much closer to that of southern Europe and North Africa

343 than to northern and central Europe. The distribution models are consistent with my hypothesis
344 that historically *C. vulvaria* was only present in parts of Northern Europe because of repeated
345 introductions, and that, in these places, the climate is unsuitable for lasting populations to exist.
346 Discrepancies between the projected model and the realised distribution could be the result of
347 several factors, either an incorrect model; lack of suitable habitat; spatial variations in surveying
348 effort, or plants growing outside their actual climatic niche due to local factors. The model
349 projecting distribution from the naturalised range is based on relatively few observations and
350 could be improved by more data. Nevertheless, any distribution model of this species has to
351 address the problem of casual occurrences. The shortage of observations from countries such as
352 Turkey and Morocco, in apparent contrast to the models, are at least in part due to lack of
353 collecting in these regions, but also due to the inaccessibility of the data from these countries.
354 These results are a good reminder to those who would extrapolate native ranges onto potentially
355 invasive ranges. Clearly, it is not always possible to predict the naturalised distribution from the
356 native range due to the lack of data and indistinct range boundaries.

357 **Conclusions**

358 Text analysis is a useful technique to study recent ecological and distributional change. Despite
359 its limitations it provided information, which would be difficult, if not impossible to obtain from
360 other sources. As a larger volume of semantically enhanced biodiversity literature becomes
361 available it will allow much more sophisticated habitat analysis covering many more species. The
362 ability to contrast data from different species will strengthen results and allow correction for
363 some of the biases. Furthermore, the development of environmental ontologies and [thesauri](#) will
364 simplify the method and improve repeatability (Buttigieg et al., 2013). This will allow over-
365 representation analysis of ontological terms from one species in comparison to these terms in the
366 whole corpus.

367 Analysis of these descriptions indicates that the habitat of *C. vulvaria* has changed over the past
368 two centuries, particularly next to walls. Multiple vectors and pathways have been involved in the
369 human mediated dispersal of *C. vulvaria*, but different vectors and pathways were active in
370 different periods. In the past *C. vulvaria* would have been dispersed to many places outside of its
371 climatic niche. It is reasonable to believe that many of the observations of *C. vulvaria* in northern
372 Europe were the result of introductions and that a reduction in the propagule pressure in recent
373 years has consequently lead to a decline in observations of this species. It is concluded that
374 mankind spread *C. vulvaria* to northern Europe and created habitat for it to grow and then

375 inadvertently removed the habitat and the introduction pathways causing a decline.

376 **References**

377 Araujo MB, Peterson AT. 2012. Uses and misuses of bioclimatic envelope modeling. *Ecology*
378 93:1527–1539.

379 Atlas of Living Australia. 2014. *Available at*

380 <http://bie.ala.org.au/species/Chenopodium+vulvaria> (accessed 25 February 2013).

381 Berendsohn WG, Güntsch A, Hoffmann N, Kohlbecker A, Luther K, Müller A. 2011.

382 Biodiversity information platforms: From standards to interoperability. *ZooKeys* 150:71–87.

383 Boelcke O, Correa NM, Moore DM, Roig FA. 1985. Catálogo de las Plantas Vasculares. In:

384 Boelcke O, Moore DM, Roig FA, eds., *Transecta Botánica de la Patagonia Austral. Proyecto*

385 *Internacional*. Buenos Aires: Consejo Nacional de Investigaciones Científicas y Técnicas, 129–

386 255.

387 Botanical Society of Britain and Ireland. 2013. Distributions Database. *Available at*

388 <http://bsbidb.org.uk>. (accessed 21 February 2013).

389 Bucher, C.T. (1806) *Florae Dresdensis nomenclátor*. Dresden: Walther.

390 Burk I. 1877. List of plants recently collected on ships' ballast in the neighborhood of

391 Philadelphia. *Proceedings of the Academy of Natural Sciences of Philadelphia* 29:105–109.

392 Buttigieg PL, Morrison N, Smith B, Mungall CJ, Lewis SE, and the ENVO Consortium. 2013.

393 The environment ontology: contextualising biological and biomedical entities. *Journal of*

394 *Biomedical Semantics* 4:43.

395 Calflora. 2014. Information on California plants for education, research and conservation, based

396 on data contributed by dozens of public and private institutions and individuals, including the

397 Consortium of Calif. Herbaria. Berkeley (USA). *Available at* www.calflora.org (accessed 3

398 March 2014).

399 Ciardelli P, Kelbert P, Kohlbecker A, Hoffmann N, Güntsch A, Berendsohn WG. 2009. The EDIT

400 Platform for Cybertaxonomy and the Taxonomic Workflow: Selected Components. In: Fischer S,

401 Maehle E, Reischuk R, eds. *INFORMATIK 2009, Im Focus das Leben, Beiträge der 39.*

402 *Jahrestagung der Gesellschaft für Informatik e.V. (GI), 28.9. - 2.10. in Lübeck*. Lecture Notes in

403 Informatics (LNI) 154 - S. 28: 625–638.

- 404 Carlton JT. 2011. Ballast. In: Simberloff D, Rejmanek M, eds. *Encyclopedia of Biological*
405 *Invasions*. Berkeley and Los Angeles: University of California Press, 43–49.
- 406 Chapman AD, Wieczorek J. 2006. Guide to best practices for georeferencing. Copenhagen:
407 Global Biodiversity Information Facility, 1–77.
- 408 Cheffings CM, Farrell L, Dines TD, Jones RA, Leach SJ, McKean DR, Pearman DA, Preston
409 CD, Rumsey FJ, Taylor I. 2005. The vascular plant red data list for Great Britain. *Species Status*
410 *7*:1–116.
- 411 Colling G. 2005. Red list of the vascular plants of Luxembourg. *Ferrantia* 42:80.
- 412 Curtis W. 1777. *Flora Londinensis vol. 5*. London: William Curtis.
- 413 de Souza Muñoz M, De Giovanni R, de Siqueira M, Sutton T, Brewer P, Pereira R, Canhos DAL,
414 Canhos VP. 2011. openModeller: A generic approach to species' potential distribution modelling.
415 *GeoInformatica* 15:111-135.
- 416 Ferrez Y. 2005. Liste rouge de la flore vasculaire menacée ou rare de Franche-Comté Proposition.
417 *Les nouvelles archives de la flore jurassienne* 3:217–229.
- 418 Groom QJ. 2013. Some poleward movement of British native vascular plants is occurring, but the
419 fingerprint of climate change is not evident. *PeerJ* 1:e77.
- 420 Grulich V. 2012. Red list of vascular plants of the Czech Republic: 3rd edition. *Preslia* 84:631–
421 645.
- 422 Giusti L. 1997. Chenopodiaceae. In: Hunziker AT, ed. *Flora fanerogámica argentina*. Córdoba:
423 Proflora-Conicet: 40: 1–53.
- 424 Haarmeyer DH, Bösing BM, Schmiedel U, Dengler J. 2010. The role of domestic herbivores in
425 endozoochorous plant dispersal in the arid Knersvlakte, South Africa. *South African Journal of*
426 *Botany* 76:359–364.
- 427 Herbaria@home. 2013. Available at <http://herbariaunited.org/atHome/> (accessed 21 February
428 2013).
- 429 Hilbig W. 1987. Changes in segetal vegetation under conditions of industrialized agriculture.
430 *Archives of Nature Conservation & Landscape Research* 27:229–249.
- 431 Hijmans RJ, Cameron SE, Parra JL, Jones PG, Jarvis A. 2005. Very high resolution interpolated

- 432 climate surfaces for global land areas. *International Journal of Climatology* 25:1965–1978.
- 433 Hjelt H. 1906. Caryophyllaceae-Resedaceae. In: *Conspectus Florae Fennicae in Acta Societatis*
434 *pro Fauna et Flora Fennica* 30. Finland: Helsingforsiae.
- 435 Hooker WJ. 1821. *Flora Scotica*. London: Archibald Constable and Co. & Hurst, Robinson, and
436 Co.
- 437 Hoste I, Verloove F, Nagels C, Andriessen L, Lambinon J. 2009. De adventievenflora van in
438 België ingevoerde mediterrane containerplanten. *Dumortiera* 97:1–16.
- 439 Jalas J, Suominen J. 1980. *Atlas Florae Europaeae, Vol. 5*. Helsinki: The Committee for Mapping
440 the Flora of Europe and Societas Biologica Fennica Vanamo.
- 441 Kestemont B. 2010. A red list of Belgian threatened species. Brussels: Statistics Belgium.
- 442 Lazkov G, Sennikov A, Naumenko A. 2014. New records in vascular plants alien to Kyrgyzstan.
443 *Biodiversity Data Journal* 2:e1018.
- 444 Lososová Z. 2003. Estimating past distribution of vanishing weed vegetation in South Moravia.
445 *Preslia* 75:71–79.
- 446 Macfadyen S, Kriticos DJ. 2012. Modelling the Geographical Range of a Species with Variable
447 Life-History. *PLoS ONE* 7: e40313.
- 448 Meusel H, Jäger E, Weinert E. 1992. *Vergleichende Chorologie der Zentraleuropäischen Flora,*
449 *Vol. 1*. Jena: Gustav Fischer.
- 450 Mohr C. 1901. Plant Life of Alabama. In: *Contributions [from the United States National](#)*
451 *[Herbarium](#)*. Washington: US Department of Agriculture.
- 452 Newcomb S. 1994. A History of Chromite and Copper in Maryland: The Tyson Years. *Matrix*
453 3:84–92.
- 454 Phillips SJ, Dudík M, Schapire RE. 2004. A maximum entropy approach to species distribution
455 modeling. In: *Proceedings of the Twenty-First International Conference on Machine Learning.*
456 *ACM (New York)*. 655–662.
- 457 Plachy J. 1992. *The Mineral Industry of Norway in Minerals yearbook mineral industries of*
458 *Europe and central Eurasia*. U.S. Bureau of Mines.
- 459 Planchuelo AM. 1975. Study of the fruits and seeds of the genus *Chenopodium* in -Argentina.

- 460 *Darwiniana* 19:528–565.
- 461 Reed CF. 1964. A flora of the chrome and manganese ore piles at Canton, in the Port of
462 Baltimore, Maryland and at Newport News, Virginia, with descriptions of genera and species new
463 to the flora of eastern United States. *Phytologia* 10:324–406.
- 464 Sautter G, Böhm K, Agosti D. 2007. Semi-Automated XML Markup of Biosystematics Legacy
465 Literature with the GoldenGATE Editor. In: *Proceedings of Pacific Symposium on Biocomputing*
466 *2007, Wailea (USA), January 2007*. International Society for Computational Biology: 391–402.
- 467 Šerá B. 2011. Stress tolerant plant species spread in the road-net. *Ecological Questions* 14:45–46.
- 468 Smart SM, Robertson JC, Shield EJ, van de Poll HAM. 2003. Locating eutrophication effects
469 across British vegetation between 1990 and 1998. *Global Change Biology* 9:1763–1774.
- 470 Sutcliffe OL, Kay QON. 2000. Changes in the arable flora of central southern England since the
471 1960s. *Biological Conservation* 93:1–8.
- 472 Turner W. 1548. The names of herbes. (with an introduction, an index of English names, and an
473 identification of the plants enumerated by Turner) by James Britten. Published for the English
474 Dialect Society by N. Trübner 1881, London.
- 475 Uotila P. 2001. *Chenopodium* L. In: Jonsell B, ed. *Flora Nordica*. Stockholm: Bergius
476 Foundation, 4–31.
- 477 Van Landuyt W, Vanhecke L, Hoste I, Hendrickx F, Bauwens D. 2008. Changes in the
478 distribution area of vascular plants in Flanders (northern Belgium): eutrophication as a major
479 driving force. *Biodiversity and Conservation* 17:3045–3060.
- 480 Vellend M, Brown CD, Kharouba HM, McCune JL, Myers-Smith IH. 2013. Historical ecology:
481 using unconventional data sources to test for effects of global environmental change. *American*
482 *Journal of Botany* 100:1294–1305.
- 483 Withering W. 1776. *A botanical arrangement of all the vegetables naturally growing in Great-*
484 *Britain Vol. 1*. London: Cadel and Elmsley.
- 485 Wróbel M, Tomaszewicz T, Chudecka J. 2006. Floristic Diversity and Spatial Distribution of
486 Roadside Halophytes along Forest and Field Roads in Szczecin Lowland (West Poland).
487 *Polish Journal of Ecology* 54:303–309.
- 488 Zengin Y. 1957. The Mode of Distribution of Chrome-Ores in Peridotites in Turkey. *Bulletin of*

489 *the Mineral Research and Exploration Institute of Turkey* 49:84–92.

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Table 1 (on next page)

Introduction vectors gleaned from historical sources

The vectors stated or implied from specimens and publications, including the range of dates that vectors were mentioned either on specimens or in publications.

Table 1. The introduction vectors stated or implied from specimens and publications, including the range of dates that vectors were mentioned on specimens and in publications.

Vector	Dates	number	Example references and specimens
Ballast	1870 –1912	13	Publications: Burk, 1877; Mohr, 1901; Hjelt, 1906 Specimens: BIRM 032912; MANCH.94943.Kk803; S-H-2810
Grain	1936 –1964	3	Uotila, 2001; Unaccessioned specimens from Nationaal Herbarium Nederland (L);
Wool	1909	1	Observation by I.M. Hayward, Selkirkshire in database of the Botanical Society of Britain and Ireland (2013)
Ore	1931 –1961	4	Uotila, 2001; Reed, 1964; Lazkov, Sennikov & Naumenko, 2014; Specimen S-H-2141
Cork	1956 –1966	1	Uotila, 2001
Horticultural imports	2008	1	Hoste <i>et al.</i> , 2009

Figure 1

The use frequency of words in the collected corpus of *Chenopodium vulvaria* habitat descriptions

The frequency of each habitat category in the corpus of habitat descriptions from literature and specimens. The word and phrase tokens contributing to each category are presented in Table S1.

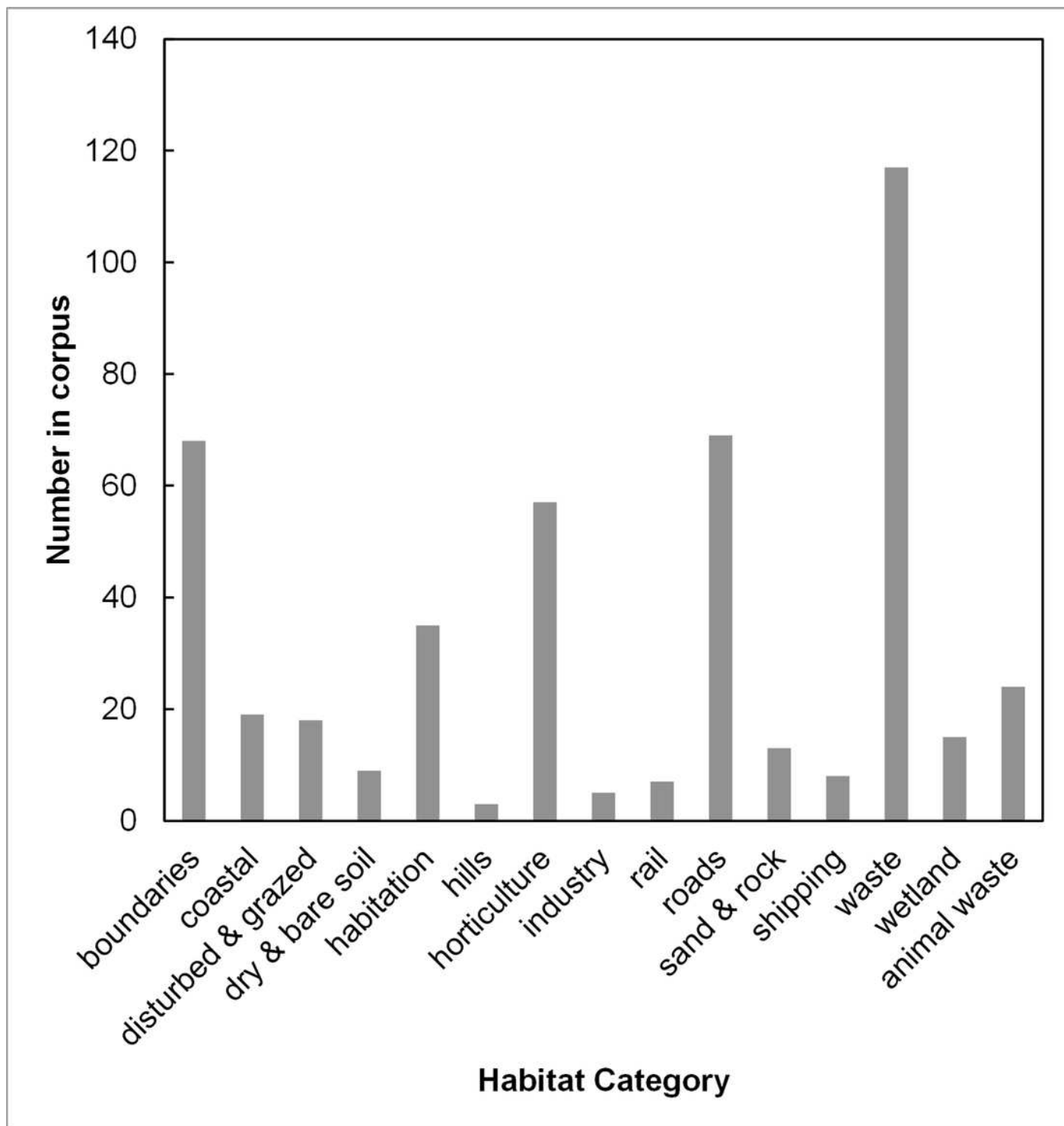


Figure 2

The change with time of habitat categories from the collected corpus of *Chenopodium vulvaria* habitat descriptions

A temporal analysis of the corpus of habitat descriptions from publications and specimens of *Chenopodium vulvaria*. The graphs show the proportion of token usage related to each habitat category for periods of 20 years. The words contributing to each habitat category are listed in Table S1. The best fit lines are from generalised linear models of the data weighted with the number of tokens contributing to each proportion. The categories wetland ($P < 0.01$, $DF=11$), sand and rock ($P < 0.05$, $DF=11$), and disturbed and grazed ($P < 0.05$, $DF=11$) all significantly increased with time. Only the term boundaries decreased with time ($P < 0.001$, $DF=11$). All other categories shown in Fig. 1 did not show significant variations with time.

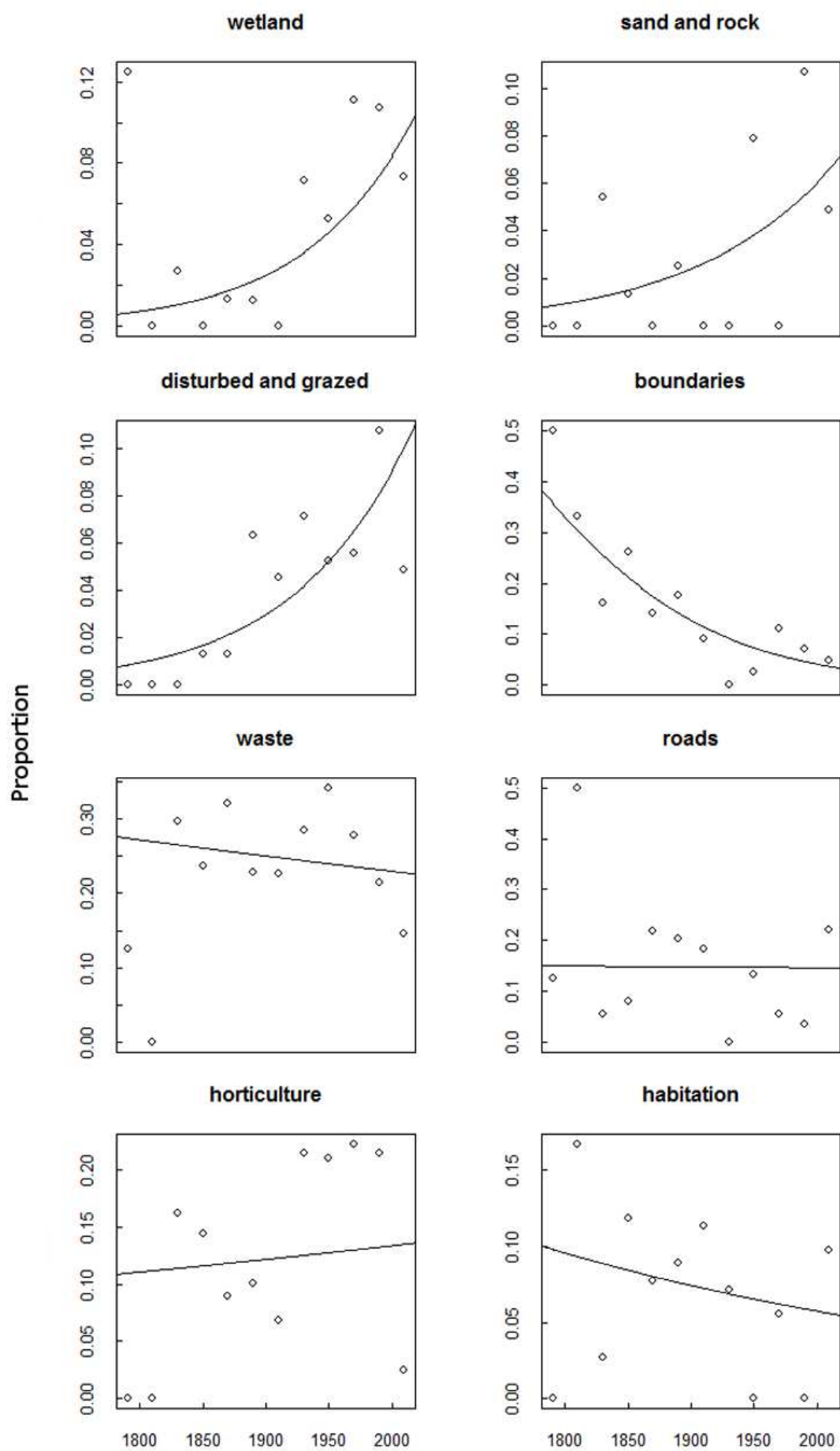


Figure 3

A distribution model created from the naturalised range of *Chenopodium vulvaria* outside Europe and extrapolated back to Europe.

A distribution model of *Chenopodium vulvaria* in Europe, North Africa and the Middle-east projected from its naturalised range in California, South America and Australia. This model aims to predict where, according to the naturalised range, the climate is suitable for persistent populations in Europe as opposed to casual occurrences. The map uses a Mollweide equal area projection.

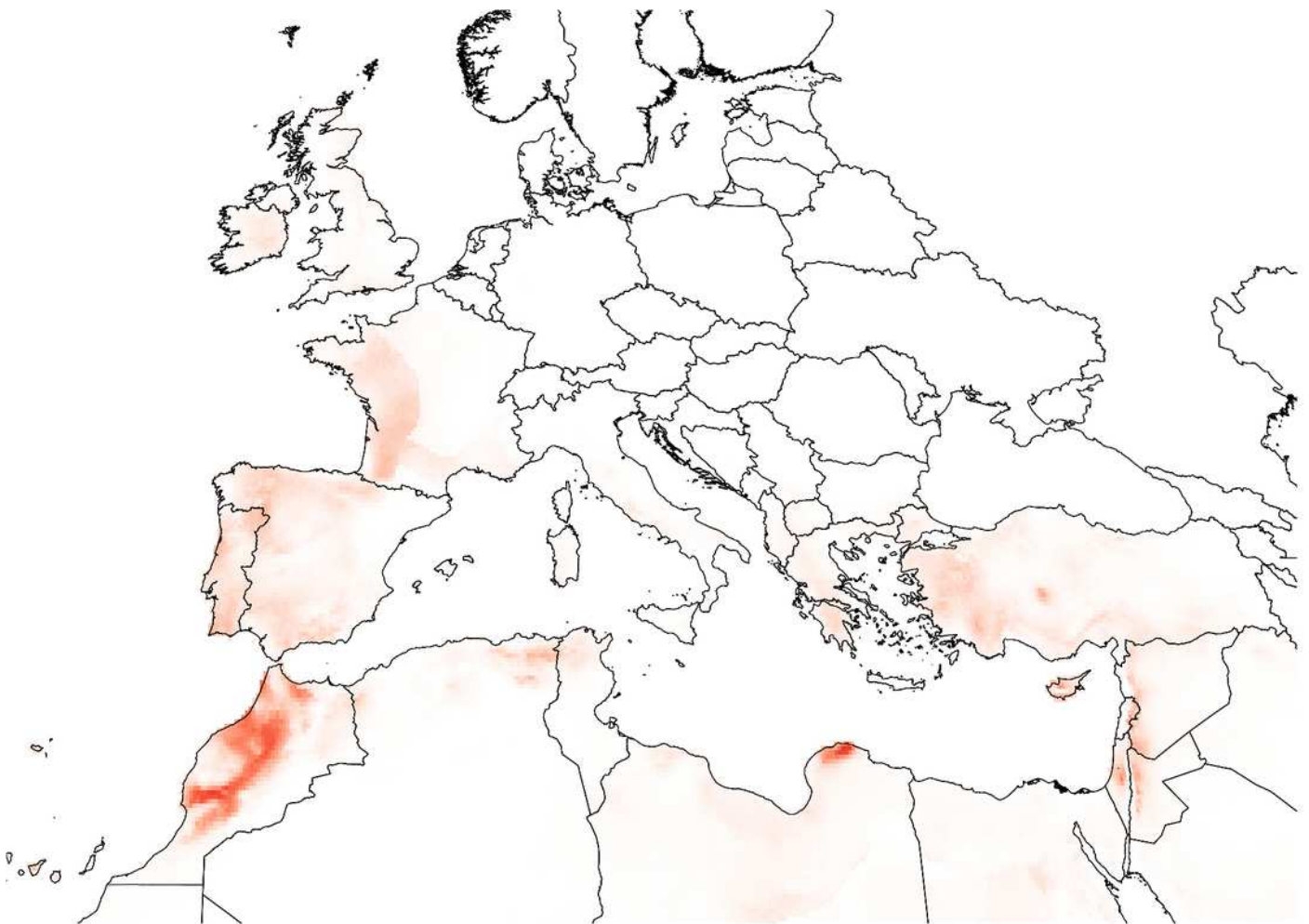


Figure 4

A distribution model of *Chenopodium vulvaria* created from all known locations

A distribution model of *Chenopodium vulvaria* in Europe, North Africa and the Middle-east created from all observations globally. This model aims to delimit the area where the climate is suitable for both stable populations and casual occurrences to occur. The map uses a Mollweide equal area projection.

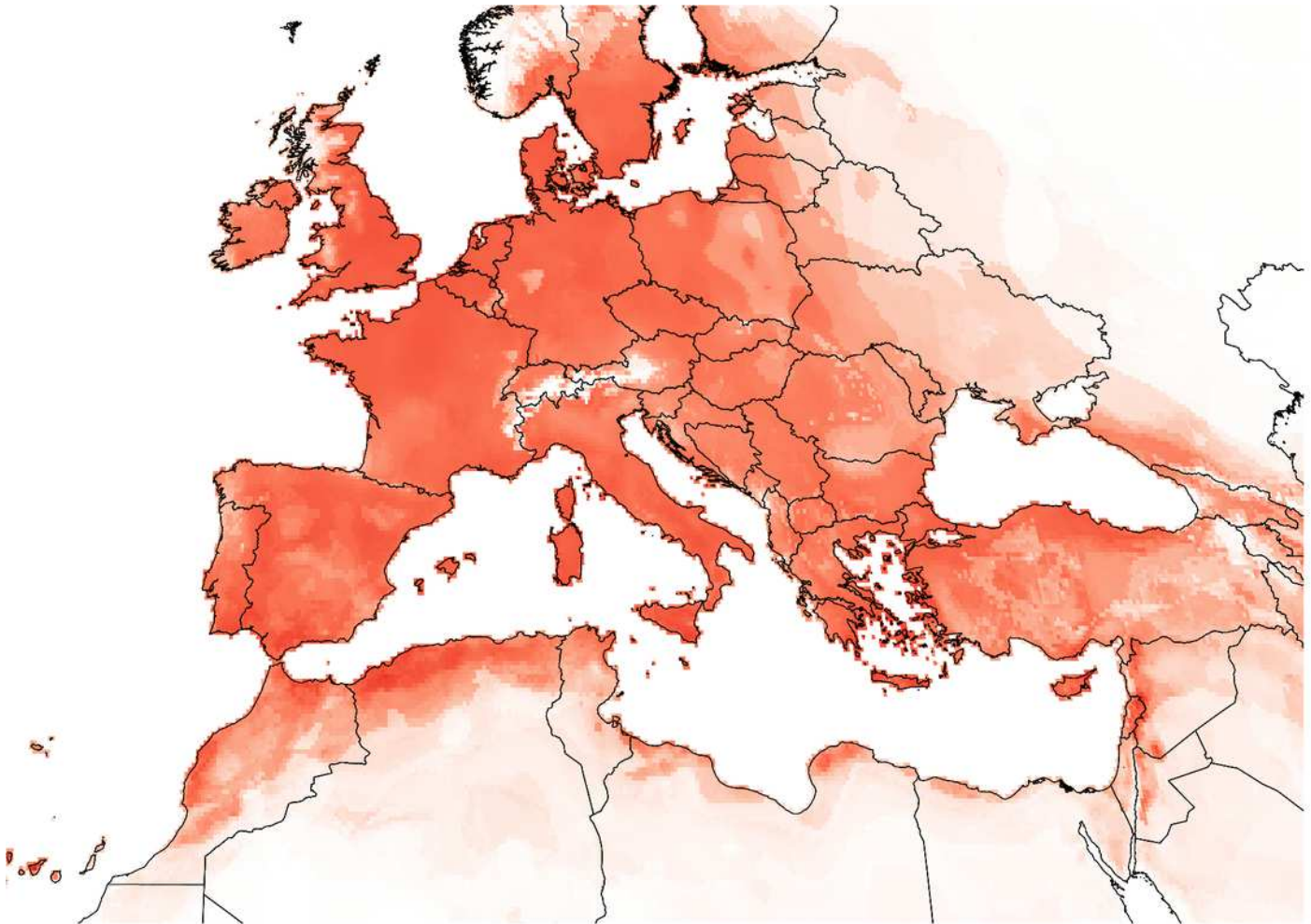


Figure 5

A dated distribution map of *Chenopodium vulvaria* observations from Europe, North Africa and the Middle-east

A distribution map of *Chenopodium vulvaria* in Europe, North Africa and the Middle-east.

Circles represent georeferenced observations either from specimens or from the literature.

The colour of the points denotes the date of observation, yet to emphasise the scarcer old records the date ranges are not equal, but the data is divided into equal-sized subsets. The map uses a Mollweide equal area projection.

