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## The role of macrobiota in structuring microbial communities along rocky shores

Rocky shore microbial diversity presents an excellent system to test for microbial habitat specificity or generality, enabling us to decipher how common macrobiota shape microbial community structure. At two coastal locations in the northeast Pacific Ocean, we show that microbial composition was significantly different between inert surfaces, the biogenic surfaces that included rocky shore animals and an alga, and the water column plankton. While all sampled entities had a core of common OTUs, rare OTUs drove differences among biotic and abiotic substrates. For the mussel *Mytilus californianus*, the shell surface harbored greater alpha diversity compared to internal tissues of the gill and siphon. Strikingly, a 7-year experimental removal of this mussel from tidepools did not significantly alter the microbial community structure of microbes associated with inert surfaces when compared with unmanipulated tidepools. However, bacterial taxa associated with nitrate reduction had greater relative abundance with mussels present, suggesting an impact of increased animal-derived nitrogen on a subset of microbial metabolism. Because the presence of mussels did not affect the structure and diversity of the microbial community on adjacent inert substrates, microbes in this rocky shore environment may be predominantly affected through direct physical association with macrobiota.

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3 The role of macrobiota in structuring microbial communities along rocky shores

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## INTRODUCTION

25  
26 The dynamics and interactions of the macroscopic species on rocky shores of the northeast  
27 Pacific Ocean have been well-characterized and thus have contributed significantly to our  
28 understanding of coastal ecological processes (eg. Paine 1966; Wootton 1994; Estes and  
29 Duggins, 1995). Although some specialized symbiotic associations have been described in rocky  
30 shore species (Secord and Augustine, 2005; Bergschneider and Muller-Parker, 2008), we know  
31 little about multi-taxa microbial associations. There is increasing evidence that many marine  
32 macrobiota have surface biofilms (Grossart *et al.*, 2005; Kvennefors *et al.*, 2012) or  
33 endosymbionts (Zurel *et al.*, 2011; Wegner *et al.*, 2012) or both (e.g. Qian *et al.*, 2006; Taylor *et*  
34 *al.*, 2007). However, our understanding of the specificity of these associations and their  
35 functional significance remain nascent with some notable exceptions (Webster and Taylor, 2012;  
36 Fan *et al.*, 2013; Heisterkamp *et al.*, 2013). The shelf waters of the California Current Large  
37 Marine Ecosystem (CCLME) maintain diverse and unique microbial communities across  
38 upwelling areas (Bertagnolli *et al.*, 2011). The relatively high productivity of this system has  
39 been attributed to the seasonal upwelling of nitrate (Barber and Smith 1981), which can lead to  
40 significant levels of carbon fixation. The diversity of potential plant and animal ‘host’ species in  
41 the CCLME, their relatively large geographic range, their longevity and the provision of key  
42 resources suggest they provide unique microniches capable of increasing the diversity and  
43 function of microbial communities.

44 If macrobiota directly provide habitat for intertidal microbes (e.g. mussels, Pfister *et al.*,  
45 2010; algae, Miranda *et al.*, 2013) or indirectly provide resources such as nitrogen, in the form of  
46 animal-regenerated ammonium, then our understanding of coastal biogeochemistry is incomplete  
47 without considering the contribution of macrobiota. In the northeast Pacific, Tatoosh Island, WA

48 shows persistent shoreward peaks of ammonium (Pfister *et al.*, 2007), while the tidepools at  
49 Second Beach, WA have animal-regenerated ammonium from the California mussel with locally  
50 enhanced algal productivity, and microbial nitrification (Pfister 2007). We employed 16S rRNA  
51 V4 amplicon sequencing to test whether microbes associated with macrobiota differ from those  
52 in the water column and on inert surfaces. We further compared the identified microbial taxa  
53 with those found when using shotgun metagenomics in a subset of samples for mussel shell  
54 biofilms. Finally, we used a manipulative field experiment of mussel presence or absence in  
55 tidepools to ask whether the increased rates of nitrogen remineralization and uptake  
56 demonstrated with mussels by tracer ammonium addition (Pather *et al.*, 2014) resulted in  
57 changed bacterial communities on inert surfaces.

## 58 MATERIALS AND METHODS

59 We asked whether microbial diversity and abundance were distributed differentially among  
60 intertidal microhabitats by 1) sampling distinct microhabitat types including artificial and natural  
61 surfaces at Tatoosh Island (48.32°N, 124.74°W), and 2). Sampling artificial and natural  
62 substrates in the context of an animal removal experiment at nearby Second Beach (48°, 23'N,  
63 124, 40'W). Tatoosh Island is 0.7 km off the northwestern tip of Washington State, USA and has  
64 been well studied ecologically. Previous metagenomic analysis of biofilms associated with the  
65 shells of mussels at Tatoosh Island demonstrated a rich microbial assemblage, with the genetic  
66 capacity for nitrogen cycling (Pfister *et al.*, 2010). On 6 Aug 2009, we sampled both inert and  
67 biogenic substrates in situ, as well as artificial substrates. Biogenic substrates sampled included  
68 the surface of the red alga *Prionitis sternbergii* (n=2), the anemone *Anthopleura elegantissima*  
69 (n=2), and gill tissue (n=3) and siphon tissue (n=1) of the California mussel *Mytilus*  
70 *californianus*. These biogenic hosts were chosen because they are persistent members of the

71 community and have relatively long-lived and sessile tissues that might provide a predictable  
72 substrate. In addition to the biogenic surfaces, rocks were chipped off of bench adjacent to  
73 mussel beds. Artificial substrates (glass crucible covers, 3 cm diam, [www.leco.com](http://www.leco.com)) were  
74 attached with epoxy on rock adjacent to mussel beds (n=4) and in tidepools (n=5) on 10 Jun  
75 2009 and left *in situ* for 2 months. We chose glass crucible covers because they provided  
76 heterogeneous surface, yet an inert substance. Immediately after a late morning collection at low  
77 tide, samples were frozen and sent to -80°C storage at the University of Chicago prior to  
78 extraction. Environmental characteristics of the seawater, including nutrient concentrations,  
79 were recorded as part of a regular sampling program (Wootton and Pfister, 2012) and are  
80 reported in Table S1. The Upwelling Index for this latitude (48°N) was, on average, positive  
81 ([www.pfeg.noaa.gov](http://www.pfeg.noaa.gov)).

82 We tested whether the presence of mussels had entrained different microbial assemblages by  
83 sampling natural and artificial substrates in tidepools where mussels had been removed or  
84 unmanipulated since 2002 at Second Beach (Pfister, 2007), a north-facing complex of rocks 2  
85 km east of Neah Bay, WA, USA within the Makah Tribal Reservation. For seven years prior to  
86 our microbial sample collection, mussels have been excluded from 6 tidepools that previously  
87 contained mussels by pulling them out by hand at an approximately monthly interval during the  
88 spring and summer months. Other tidepools that naturally had mussels served as controls. The  
89 biogeochemistry of these experimental and control tidepools have been characterized multiple  
90 times, allowing us to test for mussel effects on tidepool nutrient concentrations (nitrate, nitrite,  
91 ammonium, and phosphorus), and seawater pH and oxygen and how it related to microbial  
92 community structure. These parameters were measured 9 times in each tidepool during Aug 2009  
93 and 12 times in Jul and Aug of 2010. We also estimated ammonium remineralization and

94 removal rates with a tracer experiment with enriched  $^{15}\text{NH}_4^+$  in 2010 (Pather *et al.*, 2014). A ~  
95  $2\text{cm}^2$  piece of rock was chiseled from 5 pools with and 5 without mussels during morning hours  
96 on 24 Aug 2009. We also sampled from glass cover slips that had incubated for 3 months in  
97 these pools (n=7 controls, n=6 mussel removal pools) with a copper paint barrier to exclude  
98 molluscan grazers. To compare this benthic microbial assemblage to microbes in the water  
99 column, we collected 3 plankton samples on 25mm GF/F filters by pumping 300 mL of seawater  
100 on an incoming tide (1100-1130h). Due to the 0.7  $\mu\text{m}$  pore size, we likely excluded many free-  
101 living microbes, though high species richness still resulted (see below). All samples were frozen  
102 and sent to storage at  $-80^\circ\text{C}$  at the University of Chicago prior to extraction.

103 For DNA extraction, we used the Power Soil DNA Extraction Kit (MoBio). The rock,  
104 crucible cover substrates, and mussel shells were both swabbed with sterile cotton applicators  
105 and brushed with sterile spiral dental brushes that were placed in the extraction solution.  
106 Coverslips, filters, excised mussel gill or mussel siphon, anemones, and algae were placed into  
107 the beadbeater vial and pulverized. Thus, our sampling of biogenic substrata included the  
108 potential that microbes were part of the host tissue. The PCR amplification protocol followed  
109 Caporaso *et al.* (2011) for multiplexing 16S rRNA samples. The PCR products were cleaned  
110 with MoBio™ UltraClean htp PCR Clean-up kit. We amplified the V4 variable region of the  
111 16S rRNA gene from community DNA using bar-coded primers according to Earth Microbiome  
112 Project standard protocols ([www.earthmicrobiome.org](http://www.earthmicrobiome.org)).

113

#### 114 *Sequence analysis*

115 We used QIIME (v. 1.7.0, Quantitative Insights into Microbial Ecology; [www.qiime.org](http://www.qiime.org)) to  
116 filter reads and determine OTUs as described in Caporaso *et al.* (2010a, 2011). Briefly, we used

117 the open reference OTU picking workflow, where sequences were first clustered with the  
118 Greengenes (Dec 2012) reference database (McDonald *et al.*, 2012); we then allowed OTUs that  
119 did not cluster with known taxa (at 97% identity) in the database to cluster *de novo*. Singleton  
120 and chloroplast-derived sequences were removed prior to downstream analyses. Representative  
121 sequences for each OTU were aligned using PyNast, with a minimum alignment overlap of 75 bp  
122 (Caporaso *et al.*, 2010b). Alignments were used to build a phylogenetic tree (FastTree; Price, *et*  
123 *al.*, 2009). We computed alpha diversity metrics among substrates using the `alpha_diversity.py`  
124 script in QIIME (chao1, phylogenetic diversity and equitability), using the same sequence depth  
125 for all samples (50000 sequences per sample). We used the `beta_diversity_through_plots.py`  
126 script to compute beta diversity distances between samples (weighted UniFrac), and to construct  
127 principal component (PCoA) plots, thus accounting for both the phylogenetic composition  
128 (Lozupone *et al.*, 2011) and the relative abundance of taxa. To test for significant sample  
129 groupings based on these distance metrics, we employed PERMANOVA and PERMDISP using  
130 the `compare_categories.py` script in QIIME. We tested whether the abundance of particular  
131 OTUs differed significantly among different substrates using ANOVA analyses (Bonferonni  
132 corrected) with the `otu_category_significance.py` script. OTU networks were constructed using  
133 the `make_otu_network.py` script in QIIME. We further visualized the extent to which OTUs  
134 were shared or unique among samples using Cytoscape network layouts ([www.cytoscape.org](http://www.cytoscape.org)).  
135 Finally, we tested for patterns in species co-occurrence as a function of mussels with a  
136 checkerboard score (c-score) analysis using the `oecosimu` function (Vegan package) with the  
137 ‘quasiswap’ method (99 simulations) for null model construction (Barberán *et al.*, 2012).

138 Because nitrogen metabolism in association with animals was demonstrated in these locales  
139 (Pather *et al.* 2014, Pfister *et al.* in press), we tested whether taxa known to be involved in



140 nitrogen metabolic pathways were present using 3 methodologies. We first examined the taxa  
141 identified (down to the level of genus) with the Greengenes database and 16S data in each  
142 sample. From literature reports, we assigned taxa to one of 4 transformations: nitrification  
143 (either ammonia oxidation or nitrite oxidation), anammox, or nitrate reduction via DNRA or  
144 denitrification based on genera associated with each metabolism (Table 1), comparing categories  
145 in R (version 2.15, [www.R-project.org](http://www.R-project.org)).

146 Our second analysis of potential nitrogen cycling used PICRUSt to predict the percent of  
147 sequences associated with nitrogen metabolism in our 16s data (Langille *et al.*, 2013). Briefly, all  
148 OTUs not assigned to the Greengenes database were removed from the OTU table (closed  
149 reference), abundance was normalized by 16s rRNA read number, and PICRUSt metagenome  
150 predictions were calculated (Greengenes May 2013 release). By inferring the gene families  
151 present in our 16s data, we compared whether the different living and inert substrates on Tatoosh  
152 Island differed in the amount of OTUs associated with nitrogen metabolism or whether inert  
153 substrates in tidepools with or without mussels differed in the abundance of taxa related to  
154 nitrogen metabolism.

155 The discovery of nitrogen metabolizing taxa with 16S rRNA data require that those taxa are  
156 described and identical or analogous sequences are available. In contrast, shotgun metagenomics  
157 directly identifies the sequence associated with metabolic function. Two of our samples of *M.*  
158 *californianus* shells were also shotgun pyrosequenced (Pfister *et al.*, 2010), which allowed us to  
159 compare nitrogen metabolisms detected with shotgun metagenomics with 16S OTU reads and  
160 PICRUSt predictions.. We sequenced the 16S rRNA V4 amplicons using the Illumina platform  
161 described above from DNA archived from a previous extraction from Tatoosh Island, where  
162 microbial community biofilms associated with mussel shell surfaces were extracted and

163 metagenomes sequenced using 454 GS-flx pyrosequencing (Pfister *et al.*, 2010). Briefly, 6  
164 mussels were collected on 10 April 2008 from 6 tidepools and 6 additional mussels  
165 approximately 5 m apart on an adjacent exposed rocky bench. All shells were immediately  
166 cleaned of all soft tissue. Mean shell length was 4.47 cm and 4.42 cm for tidepool and bench  
167 mussels, respectively. We thus compared the 16S rRNA V4 amplicon Illumina sequences to the  
168 shotgun metagenomic pyrosequencing to determine overlaps in key taxa. The two metagenomes  
169 from Pfister *et al.*, (2010) were reanalyzed using MG-RAST for all taxonomic matches and using  
170 the SEED Subsystems database for nitrogen metabolism with maximum e-value of  $e < 10^{-5}$   
171 (<http://metagenomics.anl.gov/>).

## 173 RESULTS

174 *Some distinction exists among microbial assemblages associated with different substrates at*  
175 *Tatoosh Island*

176 Between 54,490 and 250,432 sequences per sample were generated for 26 samples from a  
177 range of materials including inert surfaces (rock and glass crucible lids) as well as mussel shells  
178 and tissues, algal fronds, sea anemones, and the filtered plankton. All samples were rarified to  
179 50,000 sequences per sample. OTU richness (total diversity) estimates were greatest for inert  
180 substrates and the water column, while the lowest richness was associated with mussel gill and  
181 siphon tissue (ANOVA,  $F_{7,16}=4.968$ ,  $p=0.004$ ). Species richness was highly correlated with other  
182 metrics of diversity including chao1 and phylogenetic diversity ( $r=0.98$  to  $0.99$ ,  $p < 0.001$ ), as was  
183 equitability (or evenness,  $r=0.780$ ,  $p < 0.001$ ). Alphaproteobacteria, including Rhodobacteraceae  
184 and Hyphomonadaceae, dominated the algal *Prionitis* tissue and the inert substrates, while  
185 Gammaproteobacteria, especially Vibrionaceae, dominated mussel gill and siphon tissue (Fig 1).

186 The communities associated with mussel shells, anemones, and filtered plankton samples were  
187 similarly dominated by Gammaproteobacteria. Mussel shells and anemones similarly had many  
188 Vibrionaceae OTUs, with shells also harboring Moritellaceae. OTUs in the Psychromonadaceae  
189 were prominent in the plankton. Given our use of a 0.7  $\mu$ M filter, the OTU richness may be  
190 underestimated if the smallest bacteria were not retained.

191 In addition to differences in alpha diversity, the microbial community composition and  
192 structure on different biotic and inert substrates showed differences in beta diversity. First, the  
193 same substrates clustered in a PCoA analysis based on weighted UniFrac distances (Fig 2), e.g.  
194 rock substrates clustered with the glass crucibles, while mussel gill and siphon tissue clustered  
195 together. The filtered plankton samples were highly similar to each other, while the anemone and  
196 *Prionitis* tissues suggest greater differences among individual hosts. Substrate differences were  
197 significant with a permuted ANOVA ( $F_{5,18}=6.570$ ,  $p<0.001$ ) when we grouped substrates into 6  
198 categories (anemone, red alga, plankton, mussel shell, mussel internal tissue, and inert  
199 substrates). Analysis of the differences among those 6 categories showed that each differed  
200 significantly from one or several others, except for filtered POM, which did not statistically  
201 differ from any other group.

202 A second test indicating beta diversity differences among substrates was revealed in an  
203 ANOVA on OTU abundance. There were 10 OTUs that differed significantly in abundance  
204 (Bonferroni corrected ANOVA,  $p<0.05$ , Fig 3) and these distinctions came primarily from their  
205 abundance in association with macrobiota. For example, *Moritella* and *Aliivibrio* were found on  
206 mussel shell and gill tissue, while Cyanobacteria were primarily associated with algal fronds.

207 Further, the beta diversity differences we detected among substrates were a function of the  
208 abundance of OTUs among samples. By comparing OTUs shared between substrates in a

209 network plot, it was evident that the abundant OTUs that were detected at least 5000 times were  
210 generally common to all substrates (Fig 4a). In contrast, when we examined shared diversity  
211 among rare OTUs (those detected only 5 to 10 times across the dataset, Fig 4b), there was strong  
212 differentiation among substrates, indicating that the rare OTUs were responsible for the majority  
213 of the compositional differences between the microbial communities associated with different  
214 substrates.

215  
216 *The presence of mussels has little impact on microbial assemblages in tide pools at Second*  
217 *Beach*

218 In experiments performed at Second Beach, the biogeochemical parameters of tidepools were  
219 affected by the presence or absence of mussels (Supplementary Table S1). A principal  
220 components analysis that included the ammonium regeneration and removal rates (Pather *et al.*,  
221 2014), the maximum seawater pH and dissolved oxygen, ammonium, nitrate, nitrite and  
222 phosphorus measured in the tidepools over both daytime and nighttime low tides indicated that  
223 the first principal component explained 81.7% of the variance and differed among mussel versus  
224 no mussel tidepools ( $p=0.049$ , Supplementary Figure S1). Rock had more than twice the  
225 microbial diversity of coverslips (Rock=3,727 OTUs, Coverslips=1,750 OTUs;  $F_{1,18}=140.59$ ,  
226  $p<0.001$ ), perhaps reflecting greater time in the environment. Both types of substrata maintained  
227 more diverse and equitable community profiles in tidepools where mussels were removed, than  
228 in those with mussels present (Rock=3,282 OTUs, Coverslips=1,343 OTUs; Fig 1b, Two-way  
229 ANOVA,  $F_{1,18}=12.759$ ,  $p=0.002$ ). However, there was no interaction between substrate and  
230 mussel presence ( $F_{1,18}=0.013$ ,  $p=0.909$ ), indicating that the distinction between microbial  
231 communities associated with natural rock and glass coverslip communities did not depend upon

232 the presence of mussels. The equitability with which diversity was distributed was strongly  
233 correlated with total diversity ( $r=0.920$ ,  $p<0.001$ ), indicating that when mussels are removed, the  
234 equitability of taxa also increases. Further, the microbial communities associated with the rock  
235 substrate in tidepools at Second Beach were similar in OTU composition with rock substrate at  
236 Tatoosh Island (Fig 1a vs 1b).

237 A high degree of OTU sharing and community structure similarity were observed between  
238 microbial communities associated with rock surface or coverslip samples regardless of whether  
239 mussels were present (Figs 5 and 6); this was supported by the absence of OTUs with  
240 significantly different relative abundances between tidepools with or without mussels based on  
241 Bonferroni-corrected ANOVAs. However, there were fewer shared OTUs among rock samples  
242 when OTUs that were rare were considered against OTUs that were common (26.7 % versus  
243 96.9% shared). There was no relationship in the degree of OTU sharing with mussel presence or  
244 absence. Indeed, in contrast to our comparison of Tatoosh Island substrates, both rock and  
245 coverslip samples showed no differentiation between common or rare OTUs as a function of  
246 mussels, and the clustering of nodes was highly similar. Further, the presence of mussels was not  
247 associated with any changes in the relative weight of deterministic and stochastic forces  
248 governing community assembly. OTU co-occurrence was significantly non-random on rock  
249 surfaces (c-score analysis;  $p<0.01$ ), and this pattern did not change with mussels.

250

251 *Patterns in the distribution of a subset of taxa involved in nitrogen metabolism do show*  
252 *responses to the presence of mussels in tide pools on Second Beach.*

253 Nitrifying taxa were at low incidence throughout our samples, while taxa related to nitrate  
254 reduction were found in almost every sample (Table 2). Filtered plankton had the highest

255 incidence of nitrifying OTUs as a result of the genus *Paracoccus*. The relatively high incidence  
256 of denitrifying taxa in the anemone *Anthopleura* was primarily driven by matches in the  
257 Campylobacteraceae, the group including *Campylobacter*, a taxon that harbors *nrf* genes  
258 (Pittman and Kelly, 2005). Although most samples also had genera in the Planctomycetes, only  
259 some are known to perform anammox (Fuerst and Sagulenko, 2011), and these genera were  
260 either absent from our samples or characterized as “other Planctomycetes”. We tallied these  
261 OTUs for Table 2, but did not perform statistical analyses and interpret these with caution.  
262 OTUs associated with nitrogen metabolism, as inferred from PICRUSt, were found across all  
263 Tatoosh substrates (Table 2).

264 The percentage of taxa associated with nitrogen transformations and residing on mussel  
265 shells was compared between the 16s rRNA amplicon data and existing shotgun metagenomic  
266 data (Pfister *et al.*, 2010). The metagenomic data revealed a greater proportion of nitrogen  
267 metabolizing taxa, including taxa that were not observed in the amplicon data (Table 2). An  
268 analysis of SEED Subsystem functions for the two mussel shell metagenomes yielded estimates  
269 of 1.4 % and 1.3% of the 68,676 and 63,950 proteins with functions known to be related to  
270 nitrogen metabolism that were discovered in each sample. The PICRUSt analysis, which  
271 inferred functional gene presence, indicated that 0.83% of the OTUs discovered were related to  
272 nitrogen function (Table 2), a value closer to the metagenome discovery rate than our discovery  
273 analyzing only the taxa in Table 1 with 16S rRNA data, and likely larger due to the inclusion of  
274 many nitrogen-metabolizing taxa in addition to those in Table 1.

275 When we compared the presence of OTUs with taxa associated with certain nitrogen  
276 metabolisms (e.g. Table 1) in our experimental tidepools, we found that the presence of mussels  
277 increased the incidence of putative nitrate reducing taxa on rock substrate, but mussels had no

278 effect on putative nitrifiers (Table 3). The nitrogen metabolisms on inert substrates that were  
279 inferred through PICRUST also did not differ between tidepools with or without mussels (Table  
280 3). The maximum dissolved inorganic nitrogen in each tidepool did not correlate with the  
281 observed diversity ( $r=-0.318$ ,  $p=0.371$ ). Over all 46 samples from Tatoosh and Second Beach  
282 that we analyzed, the discovery rate of OTUs with known nitrogen transformations (Table 1) was  
283 unrelated to the observed diversity ( $r=-0.127$ ,  $p=0.399$ ,  $n=46$ ).

## 284 285 DISCUSSION

286 All sampled substrates had microbial assemblages, which showed variation in important  
287 taxonomic and functional properties. Differences in diversity and OTU composition may reflect  
288 differences in colonization preference, temporal dynamics, the duration in the environment to  
289 accrue microbes, and host-specific interactions on the longer-lived macrobiota. The red alga, the  
290 anemone and the mussels are all long-lived and could have been present for years prior to our  
291 sampling and had ample time to accumulate a microbial assemblage. The clonal nature of the  
292 anemone *Anthopleura elegantissima* might even provide nearly immortal tissue for microbial  
293 proliferation. Relatively high diversity on rocks and crucible lids may indicate a number of  
294 micro-niches, perhaps as a result of surface irregularities on these substrates. The biotic  
295 substrates we sampled are especially likely to have a number of microbial niches, including  
296 microtopography and strong oxygen gradients (Heisterkamp et al. 2013).

297 Similarities between the filtered plankton and the animal tissue may have resulted from  
298 mussel and anemone tissue harboring planktonic microbiota, due to their feeding activities.  
299 Although increased sample size will be needed to quantify the extent of within-host  
300 heterogeneity, the animals we examined were dominated by Gammaproteobacteria, including the

301 symbiont-bearing anemone. In contrast, coral reef invertebrates with eukaryotic photosynthetic  
302 symbionts are have been shown to be dominated by Alphaproteobacteria (e.g. Bourne *et al.*,  
303 2013), suggesting differential colonization drivers in different symbiont-bearing invertebrates.  
304 Nevertheless, many specialized symbionts of marine bivalves fall within the  
305 Gammaproteobacteria (Stewart and Cavanaugh, 2006; Newton *et al.*, 2008).

306 The great similarity in microbial community structure on inert substrates in tidepools  
307 regardless of the presence or absence of mussels (Figure 1b, 3, 5, 6), suggests that the enhanced  
308 nitrogen regeneration and uptake that has been demonstrated with mussels (Pather *et al.*, 2014)  
309 could be due to the microbes on the mussels themselves. The only important functional  
310 difference that we found was a greater incidence of nitrate reducing taxa on rocks in the presence  
311 of mussels (Table 3), a result likely explained by the ability of mussels to temporarily reduce  
312 oxygen levels to the point where reducing processes are favored. OTUs associated with  
313 nitrification did not differ. Although mussel presence was associated with a decreased alpha  
314 diversity on inert substrates (Fig 1), this did not lead to significant differences in the network of  
315 OTUs that were shared across tidepools with and without mussels and thus did not affect beta  
316 diversity (Fig 6). Hence, the macrobiota did not drive any shifts in community structure on  
317 nearby inert substrates, even though tissue-associated microbial communities were significantly  
318 differentiated. Both 16S rRNA and shotgun metagenomic analyses (Table 2) suggest that  
319 macrobiota host OTUs that are important for nitrogen cycling. This corroborates work by Welsh  
320 and Castadelli (2004) that showed nitrogen metabolism was hosted within a related mussel  
321 species. Similarly, deep-sea mussels are known to host nitrogen-utilizing symbionts (Lee and  
322 Childress, 1995). In contrast with the shell surface microbes, those with mussel gill and siphon  
323 were less diverse (Fig 1). Thus, the elevated nitrite concentrations in tidepools with mussels,



324 suggesting increased nitrification (Pfister, 2007), may be the result of a direct effect of habitat  
325 provisioning for microbes by the macrobiota, rather than simply a microbial community shift on  
326 other substrates due to nutrient provisioning. Alternatively, the nitrogen increase due to mussels  
327 may not be enough to drive microbial community differences. Indeed, neither microbial  
328 community composition nor the expression of several functional genes in coastal sediments  
329 showed major changes in response to nutrient perturbations (Bowen *et al.*, 2011). Thus, further  
330 exploration of macrobiota as the repository of microbial function, not just as providers of  
331 nutrient resources, is warranted.

332 We note that nitrogen-transforming taxa were found with a higher incidence in the  
333 metagenomic data than the 16S rRNA V4 amplicon sequencing of mussel shells. Although there  
334 are technical differences between the two sequencing methodologies that could lead to detection  
335 differences, such as GC bias (Ross *et al.*, 2013), metagenomic data directly identify genes for  
336 nitrogen transformations that are relatively conserved and identifiable (independent of host  
337 phylogeny), even if the taxa hosting these genes are uncharacterized. Many of the taxa identified  
338 in the amplicon survey are not closely related to cultured isolates, and their N-cycling status is  
339 unknown. Although this limits our ability to infer function from phylogeny, it nevertheless is an  
340 analysis that may become increasingly insightful as our knowledge of sequence-based diversity  
341 increases. While we recognize that OTU analysis using 16S rRNA data predictions may not yet  
342 be the strongest lens to detect function, the increased characterization of taxa involved in  
343 nitrogen metabolisms (Ward *et al.* 2011, Munn 2011), and the analyses of 16S data with  
344 PICRUST (Table 2) suggests that functional inference is possible.

345 Marine microbial community structure has been shown to be composed of a multitude of rare  
346 taxa that have deep phylogenetic differences; the extent to which this 'rare biosphere' (Sogin *et*

347 *al.*, 2006) drives community function is unknown, though deep sequencing efforts have revealed  
348 that there could be the equivalent of a ‘seed bank’ of rare taxa that are persistent with only  
349 relative abundance changing through time and space (Lennon and Jones, 2011; Caporaso *et al.*,  
350 2012; Gibbons *et al.*, 2013). In some ecosystems, rare taxa have also been shown to be as  
351 metabolically active as common species (Hamasaki *et al.*, 2007), suggesting that rarity does not  
352 preclude functional importance. Although seawater and rock had the highest OTU diversity, the  
353 differences among filtered plankton and inert substrates versus biogenic substrates (mussels,  
354 seaweed, anemone) sampled *in situ* demonstrated that macrobiota enhance beta diversity by  
355 hosting unique OTUs (Fig 3, Fig 4b), presumably via the provisioning of unique habitats or  
356 resources. It is possible that particular microbial taxa end up in association only with intertidal  
357 macrobiota, though the selectivity of these associations requires further temporal and spatial  
358 sampling. Our results with these benthic macrobiota are, however, in direct contrast to analyses  
359 of seawater where the patterns of beta diversity did not differ among abundant and rare taxa  
360 (Amaral-Zettler *et al.*, 2010).

361 The demonstration that macrobiota host a unique microbial community compared to the  
362 water around them is supported by this study and other recent work with tadpoles (McKenzie *et*  
363 *al.*, 2012) and marine algae (Michelou *et al.*, 2013). Although it has been recognized for several  
364 decades that benthic invertebrates in deep sea environments host unique taxa, it may be that  
365 benthic macrobiota common to large parts of the ocean are also repositories for unique  
366 microorganisms (Grossart *et al.*, 2005; Lee *et al.*, 2011; Bengtsson *et al.*, 2012; Jackson *et al.*,  
367 2012) and loci for important biogeochemistry (Martinez-Garcia *et al.*, 2008; Heisterkamp *et al.*,  
368 2013). The rocky intertidal flora and fauna, though relatively well-understood in terms of the  
369 interactions among macrobiota, likely also interact and mediate productivity via a rich microbial

370 community that we are just beginning to describe. These intertidal macrobiota may also harbor a  
371 unique set of taxa adapted to host-associated niches, thereby promoting microbial community  
372 diversity in the coastal ocean.

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#### REFERENCES

381 Amaral-Zettler L, Artigas LP, Baross J, Loka Bharathi PA, Boetius A, Chandramohan D, Herndl  
382 G, Kogure K, Neal P, Pedros-Alio C, Ramette A, Schouten S, Stal L, Thessen A, de Leeuw  
383 J, Sogin M. 2010. A Global Census of Marine Microbes. In: McIntyre, A (ed). *Life in the*  
384 *World's Oceans: Diversity, Distribution, and Abundance*. Blackwell Publishing: London. pp  
385 223-245.

386 Barber RT, Smith RL. 1981. Coastal upwelling ecosystems. In: A. Longhurst (Ed.), *Analysis of*  
387 *Marine Ecosystems*, Academic Press: New York, pp. 31–68.

388 Barberán A, Bates ST, Casamayor EO, Fierer N. 2012. Using network analysis to explore co-  
389 occurrence patterns in soil microbial communities. *International Society for Microbial*  
390 *Ecology Journal* **6**: 343–351

- 391 Bengtsson MM, Sjøtun K, Lanzén A, Øvreås L. 2012. Bacterial diversity in relation to secondary  
392 production and succession on surfaces of the kelp *Laminaria hyperborea*. *International*  
393 *Society for Microbial Ecology Journal* **7**: 1452–1458.
- 394 Bergschneider, H. and Muller-Parker, G. 2008. Nutritional role of two algal symbionts in the  
395 temperate sea anemone *Anthopleura elegantissima* brandt. *Biological Bulletin* **215**: 73–88.
- 396 Bertagnolli A.D., Treusch A.H., Mason O.U., Stingl U., Vergin K.L., Chan F. 2011. Bacterial  
397 diversity in the bottom boundary layer of the inner continental shelf of Oregon, USA.  
398 *Aquatic Microbial Ecology* **64**: 15-25.
- 399 Bourne D.G., Dennis P.G., Uthicke S, Soo R.M., Tyson G.W., Webster N. 2013. Coral reef  
400 invertebrate microbiomes correlate with the presence of photosymbionts. *International*  
401 *Society for Microbial Ecology Journal* **7**: 1452–1458.
- 402 Bowen, J.L., Ward B.B., Morrison H.G., Hobbie J.E., Valiela I., Deegan L.A., Sogin M.L. 2011.  
403 Microbial community composition in sediments resists perturbation by nutrient enrichment.  
404 *International Society for Microbial Ecology Journal* **5**: 1540-1548
- 405 Caporaso J.G., Kuczynski J., Stombaugh J., Bittinger K., Bushman F.D., Costello E.K., Fierer N,  
406 Gonzalez Peña A, Goodrich JK, Gordon JI, Huttley GA, Kelley ST, Knights D, Koenig JE,  
407 Ley RE, Lozupone CA, McDonald D, Muegge BD, Pirrung M, Reeder M, Reeder J,  
408 Sevinsky JR, Turnbaugh PJ, Walters WA, Widmann J, Yatsunenko T, Zaneveld J, Knight R.  
409 2010. QIIME allows analysis of high-throughput community sequencing data. *Nature*  
410 *Methods* **7**:335-336.

- 411 Caporaso J.G., Bittinger K., Bushman F.D., DeSantis T.Z., Andersen G.L., Knight R. 2010.  
412 PyNAST: a flexible tool for aligning sequences to a template alignment. *Bioinformatics* **26**:  
413 266-7.
- 414 Caporaso J.G., Paszkiewicz K., Field D., Knight R., Gilbert J.A. 2012. The Western English  
415 Channel contains a persistent microbial seed bank. *International Society for Microbial  
416 Ecology Journal* **6**:1089–1093.
- 417 DeChaine E.G., Cavanaugh C.M. 2006 Symbioses of Methanotrophs and Deep-Sea Mussels  
418 (Mytilidae: Bathymodiolinae). In: Overmann, J (ed). *Molecular Basis of Symbiosis. Progress  
419 in Molecular and Subcellular Biology*. Volume **41**: 227-249. Springer.
- 420 Estes, J.A., Duggins, D.O. 1995. Sea otters and kelp forests in Alaska: generality and variation in  
421 a community ecological paradigm. *Ecological Monographs* **65**: 75-100.
- 422 Fan, L., Liu, M., Simister, R., Webster, N.S., and Thomas, T. 2013. Marine microbial symbiosis  
423 heats up: the phylogenetic and functional response of a sponge holobiont to thermal stress.  
424 *International Society for Microbial Ecology Journal* **7**: 991–1002.
- 425 Fuerst, J.A., Sagulenko, E. 2011. Beyond the bacterium: Planctomycetes challenge our concepts  
426 of microbial structure and function. *Nature Reviews Microbiology* **9**: 403-413.
- 427 Gibbons, S.M., Caporaso, J.G., Pirrung, M., Field, D., Knight, R., Gilbert J.A. 2013. Evidence  
428 for a persistent microbial seed bank throughout the global ocean. *Proceedings of the National  
429 Academy of Sciences, USA* **110**:4651-4655.

- 430 Gilbert, J.A., Steele, J.A., Caporaso, J.G., Steinbrück, L., Reeder, J., Temperton, B., Huse S,  
431 McHardy AC, Knight R, Joint I, Somerfield P, Fuhrman JA, Field D. 2012. Defining  
432 seasonal marine microbial community dynamics. *International Society for Microbial Ecology*  
433 *Journal* **6**: 298–308.
- 434 Grossart, H.-P., Levold, F., Allgaier, M., Simon, M., and Brinkhoff, T. 2005. Marine diatom  
435 species harbour distinct bacterial communities. *Environmental Microbiology* **7**: 860–873.
- 436 Hamasaki, K, Taniguchi, A, Tada Y, Long, R. A., Azam, F. 2007. Actively growing bacteria in  
437 the inland Sea of Japan, identified by combined bromodeoxyuridine immunocapture and  
438 denaturing gradient gel electrophoresis. *Applied Environmental Microbiology* **73**: 2787–  
439 2798.
- 440 Heisterkamp, I.M., Schramm, A., Larsen, L.H., Svenningsen, N.B., Lavik, G., de Beer, D., Stief,  
441 P. 2013. Shell biofilm-associated nitrous oxide production in marine molluscs: processes,  
442 precursors and relative importance. *Environmental Microbiology* **15**:1943-55.
- 443 Jackson, S.A., Kennedy, J., Morrissey, J.P., O'Gara, F., Dobson, A.D. 2012. Pyrosequencing  
444 reveals diverse and distinct sponge-specific microbial communities in sponges from a single  
445 geographical location in Irish waters. *Microbial Ecology* **64**: 105-16.
- 446 Kvennefors, E.C.E., Sampayo, E., Kerr, C., Vieira, G., Roff, G., Barnes, A.C. 2012. Regulation  
447 of bacterial communities through antimicrobial activity by the coral holobiont. *Microbial*  
448 *Ecology*. **63**: 605–618.
- 449 Langille, M.G.I., Zaneveld, J., Caporaso, J.G., McDonald, D., Knights, D., Reyes JA, Clemente  
450 JC, Burkepille DE, Vega Thurber RL, Knight R, Beiko RG, Huttenhower C. 2013. Predictive

451 functional profiling of microbial communities using 16S rRNA marker gene sequences.  
452 *Nature Biotechnology* **31**: 814-821.

453 Lee, O.O., Wang, Y., Yang, J., Lafi, F.F., Al-Suwailem, A., Qian, P.Y. 2011. Pyrosequencing  
454 reveals highly diverse and species-specific microbial communities in sponges from the Red  
455 Sea. *International Society for Microbial Ecology Journal* **5**: 650-64.

456 Lee, R.W., Childress, J.J. 1995. Assimilation of inorganic nitrogen by seep mytilid 1a, an  
457 undescribed deep-sea mussel containing methanotrophic endosymbionts: fate of assimilated  
458 nitrogen and the relation between methane and nitrogen assimilation. *Marine Ecology*  
459 *Progress Series* **123**: 137-148.

460 Lennon, J.T., Jones, S.E. 2011. Microbial seed banks: the ecological and evolutionary  
461 implications of dormancy. *Nature Reviews Microbiology* **9**: 119-130.

462 Lozupone, C., Lladser, M.E., Knights, D., Stombaugh, J., Knight, R. 2011. UniFrac: an effective  
463 distance metric for microbial community comparison. *International Society for Microbial*  
464 *Ecology Journal* **5**: 169–172.

465 Martínez-García, M., Stief, P., Díaz-Valdés, M., Wanner, G., Ramos-Esplá, A., Dubilier, N.,  
466 Anton J. 2008. Ammonia-oxidizing Crenarchaeota and nitrification inside the tissue of a  
467 colonial ascidian. *Environmental Microbiology* **10**: 2991-3001.

468 McDonald, D., Price, M.N., Goodrich, J., Nawrocki, E.P., DeSantis, T.Z., Probst, A., Anderson  
469 GL, Knight RL, Hugenholtz P. 2012. An improved Greengenes taxonomy with explicit ranks  
470 for ecological and evolutionary analyses of bacteria and archaea. *International Society for*  
471 *Microbial Ecology Journal* **6**: 610–618.

- 472 McKenzie, V.J., Bowers R., Fierer N., Knight R., Lauber C.L. 2012. Co-habiting amphibian  
473 species harbor unique skin bacterial communities in wild populations. *International Society*  
474 *for Microbial Ecology Journal* **6**: 588-596. doi: 10.1038/ismej.2011.129.
- 475 Meyer, F., Paarmann, D., D'Souza, M., Olson, R., Glass, E.M., Kubal, M., Paczian T, Rodriguez  
476 A, Stevens R, Wilke A, Wilkening J, Edwards RA. 2008. The metagenomics RAST server -  
477 a public resource for the automatic phylogenetic and functional analysis of metagenomes.  
478 *BMC Bioinformatics* **9**: 386.
- 479 Michelou, V.K., Caporaso, J.G., Knight, R., Palumbi, S.R. 2013. The ecology of microbial  
480 communities associated with *Macrocystis pyrifera*. *PLoS One* 2013 Jun 19;8(6):e67480
- 481 Newton, I.L., Girguis, P.R., Cavanaugh, C.M. 2008. Comparative genomics of vesicomid clam  
482 (Bivalvia: Mollusca) chemosynthetic symbionts. *BMC Genomics* **9**: 585. doi: 10.1186/1471-  
483 2164-9-585.
- 484 Miranda, L.N., Hutchison, K., Grossman, A.R., and Brawley, S.H. 2013. Diversity and  
485 abundance of the bacterial community of the red macroalga *Porphyra umbilicalis*: Did  
486 bacterial farmers produce macroalgae? *PLoS ONE* **8**: e58269.
- 487 Munn, C. 2011. *Marine Microbiology: Ecology and Applications*, 2<sup>nd</sup> Edition. Garland Science,  
488 New York, NY, USA.
- 489 Paine, R.T. 1966. Food web complexity and species diversity. *American Naturalist* **100**: 65–75.
- 490 Pather, S., Pfister, C.A., Post, D.A., Altabet, M.A. 2014. Ammonium cycling in the rocky  
491 intertidal: remineralization, removal and retention. *Limnology Oceanography* **59**: 361-372.
- 492 Pfister, C.A. 2007. Tidepool mussels locally increase nutrients and algal growth. *Ecology* **88**:  
493 1647–1653.



- 494 Pfister, C.A., Wootton, J.T., Neufeld, C. 2007. The relative roles of coastal and oceanic  
495 processes in determining physical and chemical characteristics of an intensively sampled  
496 nearshore system. *Limnology Oceanography* **52**: 1767-1775.
- 497 Pfister, C.A., Meyer, F., Antonopoulos, D.A. 2010 Metagenomic profiling of a microbial  
498 assemblage associated with the California mussel, *Mytilus californianus*: a node in networks  
499 of carbon and nitrogen cycling. *PLoS ONE* 5(5): e10518. doi:10.1371/journal.pone.0010518.
- 500 Pittman, M.S., Kelly D.J. 2005. Electron transport through nitrate and nitrite reductases in  
501 *Campylobacter jejuni*. *Biochemical Society Transactions* **33**: 190-2.
- 502 Price, M.N., Dehal P.S., Arkin A.P. 2009. FastTree: Computing Large Minimum-Evolution  
503 Trees with Profiles instead of a Distance Matrix. *Molecular Biology and Evolution* **26**: 1641-  
504 1650.
- 505 Qian, P.Y., Dobretsov, S., Dahms, H.U., and Pawlik, J. 2006. Antifouling activity and microbial  
506 diversity of two congeneric sponges *Callyspongia* spp. from Hong Kong and the Bahamas.  
507 *Marine Ecology Progress Series* **324**: 151–165.
- 508 Ramette, A. 2007. Multivariate analyses in microbial ecology. *FEMS Microbiol Ecology* **62**:  
509 142–160.
- 510 Ross, M.G., Russ C., Costello M., Hollinger A., Lennon N.J., Hegarty R., Nusbaum C, Jaffe DB.  
511 2013. Characterizing and measuring bias in sequence data. *Genome Biology* **14**:R51.  
512 doi:10.1186/gb-2013-14-5-r51
- 513 Secord, D. and Augustine, L. 2000 Biogeography and microhabitat variation in temperate algal-  
514 invertebrate symbioses: zooxanthellae and zoochlorellae in two Pacific intertidal sea  
515 anemones, *Anthopleura elegantissima* and *A. xanthogrammica*. *Invertebrate Biology* **119**:

516 139–146.

517 Sogin, M.L., Morrison, H.G., Huber, J.A., Mark Welsh, D., Huse, S.M., Neal, P.R., Arrieta JM,  
518 Herndl GJ. 2006 Microbial diversity in the deep sea and the underexplored “rare biosphere”.  
519 *Proceedings of the National Academy of Science, USA* **103**: 12115–12120.

520 Stahl, D.A., Lane, D.J., Olsen, G.J., Pace, N.R. 1984. Analysis of hydrothermal vent-associated  
521 symbionts by ribosomal RNA sequences. *Science* **224**: 409–411.

522 Stewart, F.J., Cavanaugh C.M. 2006 Symbiosis of thioautotrophic bacteria with *Riftia*  
523 *pachyptila*. *Progress in Molecular and Subcellular Biology*. **41**: 197-225.

524 Taylor, M.W., Radax, R., Steger, D., and Wagner, M. 2007. Sponge-Associated  
525 microorganisms: evolution, ecology, and biotechnological Potential. *Microbiology and*  
526 *Molecular Biology Reviews* **71**: 295–347.

527 Telford, R.J., Vandvik, V., Birks, H.J. 2006. Dispersal limitations matter for microbial  
528 morphospecies. *Science* **312**: 1015.

529 Ward, B.B., D. J. Arp, M. J. Klotz, eds. 2011. *Nitrification*. John Wiley and Sons, Hoboken, NJ,  
530 USA.

531 Webster, N.S. and Taylor, M.W. 2012. Marine sponges and their microbial symbionts: love and  
532 other relationships. *Environmental Microbiology* **14**: 335–346.

533 Wegner, K.M., Volkenborn, N., Peter, H., and Eiler, A. 2013. Disturbance induced decoupling  
534 between host genetics and composition of the associated microbiome. *BMC Microbiology* **13**:  
535 252.

- 536 Welsh, D.T, Castaldelli, G. 2004. Bacterial nitrification activity directly associated with isolated  
537 benthic marine animals. *Marine Biology* **144**: 1029-1037.
- 538 Wootton, J.T. 1994. Predicting direct and indirect effects: An integrated approach using  
539 experiments and path analysis. *Ecology* **75**: 151-165.
- 540 Zurel, D., Benayahu, Y., Or, A., Kovacs, A., and Gophna, U. 2011. Composition and dynamics  
541 of the gill microbiota of an invasive Indo-Pacific oyster in the eastern Mediterranean Sea.  
542 *Environmental Microbiology* **13**: 1467–1476.

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## FIGURE LEGENDS

547 Figure 1. The proportional representation of OTUs among the major microbial groups (colored  
548 bars), with the overall mean observed OTU richness (+SE) among all substrate types at a.  
549 Tatoosh Island, and b. in tidepools where natural rock substrate and coverslips were sampled in  
550 the context of an experimental removal of mussels. The substrates in Fig 1a showed significant  
551 differences in observed richness (ANOVA,  $F_{7,16}=4.968$ ,  $p=0.004$ ) with rocks ( $n=3$ ), crucible lids  
552 ( $n=9$ ) and filtered plankton ( $n=3$ ) showing the greatest richness while the lowest observed  
553 richness was associated with mussel gill ( $n=3$ ) and siphon ( $n=1$ ) tissue. OTU richness of mussel  
554 shell ( $n=2$ ), anemone ( $n=2$ ), and red algae ( $n=2$ ) was intermediate to the others. In b. Tidepools  
555 with mussels removed had greater OTU richness than those with mussels (Two-Way ANOVA,  
556  $F_{1,18}=12.759$ ,  $p=0.002$ ) while rock had over twice the OTU richness of coverslips ( $F_{1,18}=140.59$ ,  
557  $p<0.001$ ); there was no interaction between substrate and mussel presence.

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559 Figure 2. A PCoA of the OTU beta diversity of substrates on Tatoosh Island, demonstrating the  
560 clustering among the different microbial assemblages associated with each substrate. The  
561 weighted UniFrac metric was used to incorporate relative abundance; the first axis explained  
562 40.2% of the variance, while the second explained 14.8%. Differences among substrates were  
563 significant (PERMANOVA,  $F_{5,18}=6.570$ ,  $p<0.001$ ), and groupings that included anemone,  
564 *Prionitis*, mussel shell, mussel tissue, and inert substrates were differentiated while plankton  
565 were indistinguishable from all.

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567 Figure 3. The relative abundance of the 10 OTUs that differed among the Tatoosh Island  
568 substrates (Boniferroni-corrected ANOVA,  $p<0.05$ ).

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Figure 4. Shared OTU diversity among microbes sampled from the substrate groupings as in Fig 1, 2 and 3 at Tatoosh Island and portrayed as a spring-embedded layout, where OTUs that are in common bring nodes or samples together and OTUs that are distinct repel nodes. In a., only common OTUs detected more than 5,000 times are included, while b. shows only rare OTUs that were present 5 to 10 times across the entire dataset.

Figure 5. A PCoA of the OTU diversity of tidepool rock (n=10) versus coverslip (n=12) substrates at Second Beach, demonstrating strong clustering among the microbial assemblages from the two substrates, while the presence of mussels (filled symbols) versus removal of mussels (open symbols) were not a factor for explaining beta diversity. Using weighted UniFrac, the first axis explained 46.5% of the variance, while the second explained 20.3%.

Figure 6. Shared OTU diversity among a. microbes sampled from tidepool rock versus coverslip (lighter green) substrates and b. samples distinguished by whether mussels were present (blue) or absent (red) from tidepools at Second Beach. The spring-embedded layout shows OTUs that are in common bring nodes or samples together and OTUs that are distinct repel nodes. Only common OTUs greater than >5000 are included. Analyses of relatively rare OTUs did not change the network pattern.

## **Table 1** (on next page)

The genera associated with different microbial nitrogen transformations.

Table 1. The genera associated with different microbial nitrogen transformations that were searched in all samples via the Greengenes database. For nitrification, any taxa associated with ammonium or nitrite oxidation are included, while nitrate reduction includes taxa for denitrification and DNRA. For anammox, we searched for taxa with “Candidatus” status.

§Due to the abundance and functional diversity of the *Vibrio* genus and the possibility that many *Vibrio* species are not involved in nitrate reduction, we did not include *Vibrio* OTUs in Tables 2 and 3.

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<b>nitrification</b>	<b>Nitrate reduction</b>	<b>anammox</b>
<i>Alcaligenes</i>	<i>Azospirillum</i>	“Candidatus”
<i>Nitrobacter</i>	<i>Campylobacter</i>	
<i>Nitrococcus</i>	<i>Desulforvibrio</i>	
<i>Nitrolancetus</i>	<i>Nitratifractor</i>	
<i>Nitrosococcus</i>	<i>Nitratiruptor</i>	
<i>Nitrosolobos</i>	<i>Paracoccus</i>	
<i>Nitrosomonas</i>	<i>Rhodobacter</i>	
<i>Nitrosopumilis</i>	<i>Sulfospirillum</i>	
<i>Nitrosovibrio</i>	<i>Wolinella</i>	
<i>Nitrospina</i>	<i>Vibrio</i> §	
<i>Nitrospira</i>		
<i>Nitrotoga</i>		
<i>Paracoccus</i>		

## **Table 2** (on next page)

The OTUs discovered in each substrate type at Tatoosh Island associated with 3 broad nitrogen transformations.

Table 2. The mean percent of OTUs discovered in each substrate type that is associated with each of the 3 broad nitrogen transformations (taxa listed in Table 1) or overall nitrogen metabolism (PICRUSt) at Tatoosh Island. No “Candidatus” were found in the 16S; the anammox category contains Planctomycetes as an estimate of anammox potential only. Mussel shell samples were analyzed with both the V4 region of the 16S rRNA as well as through shotgun metagenomics.



Table 2. The mean percent of OTUs discovered in each substrate type that is associated with each of the 3 broad nitrogen transformations (taxa listed in Table 1) or overall nitrogen metabolism (PICRUSt) at Tatoosh Island. No “Candidatus” were found in the 16S; the anammox category contains Planctomycetes as an estimate of anammox potential only. Mussel shell samples were analyzed with both the V4 region of the 16S rRNA as well as through shotgun metagenomics.

<u>Substrate</u>	Substrate type	<b>Nitrification</b>		<b>Nitrate reduction</b>		<b>Anammox</b>		<b>N metabolism</b>
		<u>16s</u>	<u>Meta-genome</u>	<u>16s</u>	<u>Meta-genome</u>	<u>16s</u>	<u>Meta-genome</u>	<u>PICRUSt</u>
Mussel siphon	biogenic	0.0000		0.0326		0.0008		0.8811
Mussel gill	biogenic	0.0000		0.4159		0.0009		0.9295
Mussel shell	biogenic	0.0000	0.4604	0.0049	0.9882	0.0043	0.1711	0.8344
filtered plankton	biogenic	0.0102		0.1851		0.0391		0.7636
<i>A. elegantissima</i> (anemone)	biogenic	0.0000		1.8961		0.0804		0.8658
<i>Prionitis</i> (red alga)	biogenic	0.0000		0.4278		0.0399		0.6638
rock	inert	0.0048		0.0551		0.0390		0.7119
crucible lid	artificial	0.0094		0.2120		0.0218		0.6325

### **Table 3**(on next page)

OTUs discovered on inert substrates in experimental Second Beach tidepools associated with 3 broad nitrogen transformations.

Table 3. The mean percent of OTUs discovered on inert substrates in experimental Second Beach tidepools that were associated with each of the 3 broad nitrogen transformations (taxa listed in Table 1) or overall nitrogen metabolism (PICRUSt). P-values are listed for t-tests for a significant difference on each substrate as a function of mussel presence. The only significant contrast was the greater incidence of OTUs associated with nitrate reduction on natural rock substrate in tidepools with mussels.

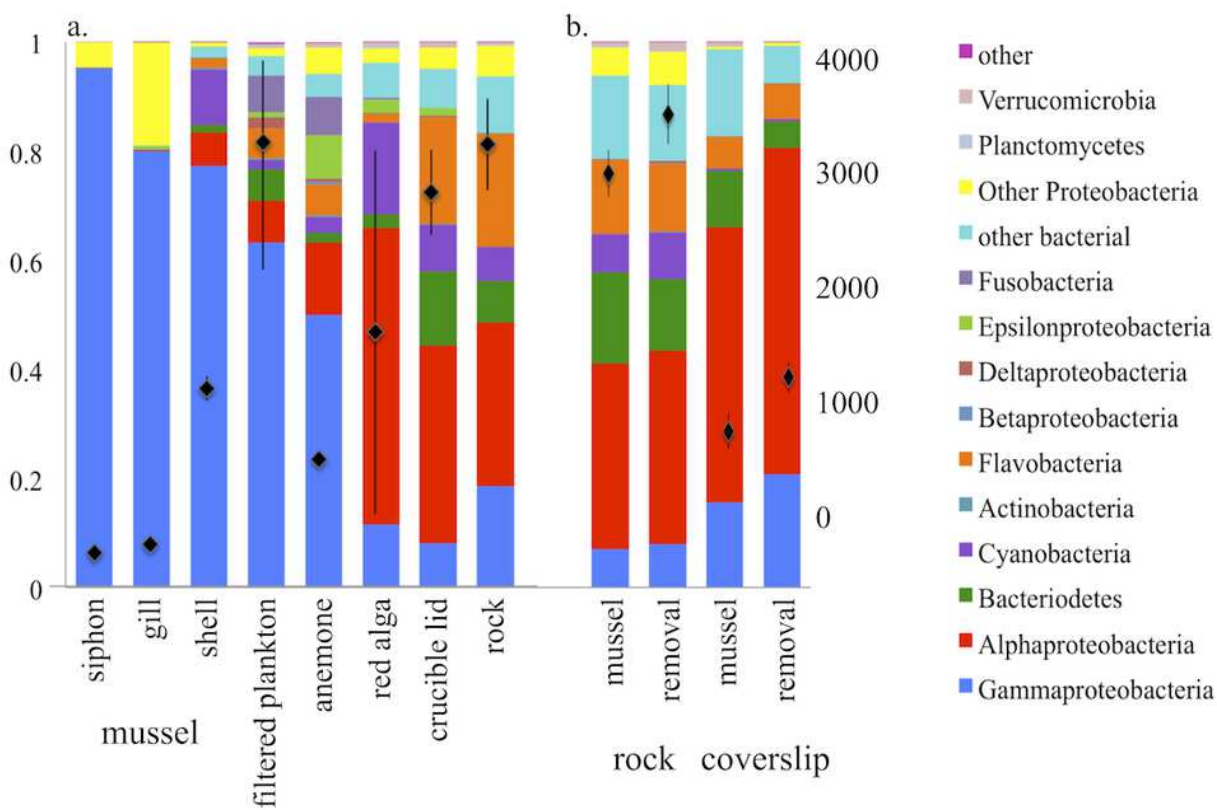
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Tidepool substrates	Nitrification		Nitrate reduction		N metabolism	
	16s	T-test	16s	T-test	PICRUSt	T-test
Natural rock substrate with mussels (n=5)	0.0001	p=0.284	0.1890	p=0.043*	0.7433	p=0.952
Natural rock substrate without mussels (n=5)	0.0011		0.0151		0.7422	
Coverslip with mussels (n=6)	0.0004	p=0.456	0.0737	p=0.471	0.7421	p=0.218
Coverslip without mussels (n=6)	0.0019		0.1785		0.7084	

# Figure 1

The proportional representation of OTUs and the mean observed OTU richness among substrates sampled.

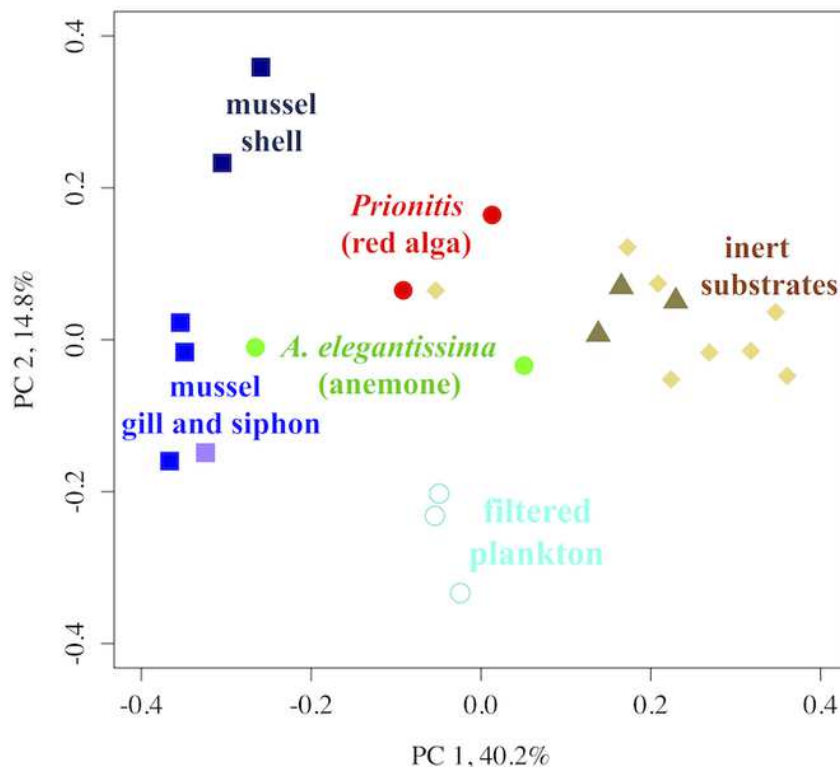
Figure 1. The proportional representation of OTUs among the major microbial groups (colored bars), with the overall mean observed OTU richness (+SE) among all substrate types at a. Tatoosh Island, and b. in tidepools where natural rock substrate and coverslips were sampled in the context of an experimental removal of mussels. The substrates in Fig 1a showed significant differences in observed richness (ANOVA,  $F_{7,16}=4.968$ ,  $p=0.004$ ) with rocks ( $n=3$ ), crucible lids ( $n=9$ ) and filtered plankton ( $n=3$ ) showing the greatest richness while the lowest observed richness was associated with mussel gill ( $n=3$ ) and siphon ( $n=1$ ) tissue. OTU richness of mussel shell ( $n=2$ ), anemone ( $n=2$ ), and red algae ( $n=2$ ) was intermediate to the others. In b. Tidepools with mussels removed had greater OTU richness than those with mussels (Two-Way ANOVA,  $F_{1,18}=12.759$ ,  $p=0.002$ ) while rock had over twice the OTU richness of coverslips ( $F_{1,18}=140.59$ ,  $p<0.001$ ); there was no interaction between substrate and mussel presence.



# Figure 2

A PCoA of the OTU beta diversity of substrates on Tatoosh Island.

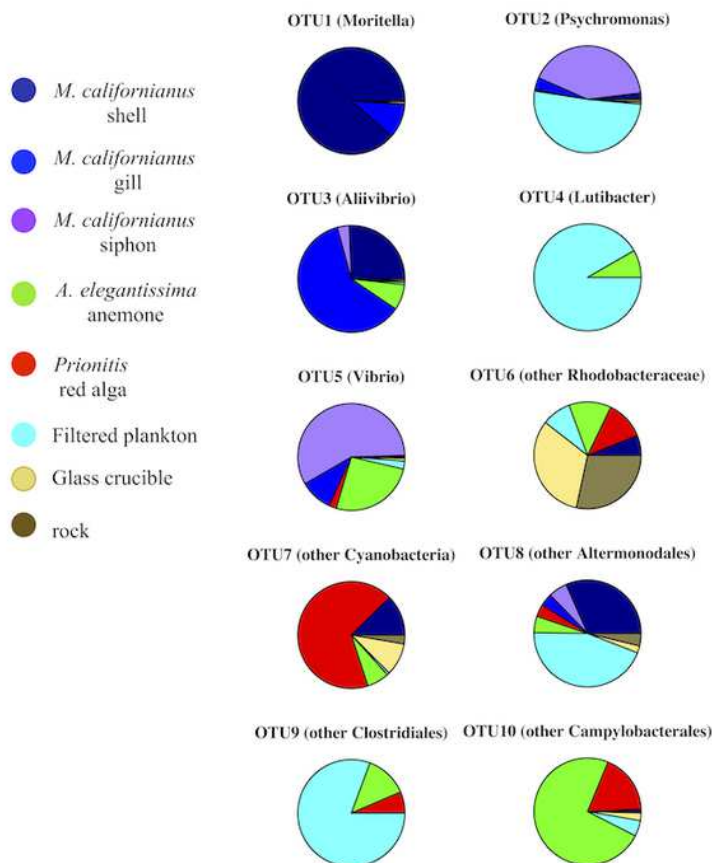
Figure 2. A PCoA of the OTU beta diversity of substrates on Tatoosh Island, demonstrating the clustering among the different microbial assemblages associated with each substrate. The weighted UniFrac metric was used to incorporate relative abundance; the first axis explained 40.2% of the variance, while the second explained 14.8%. Differences among substrates were significant (PERMANOVA,  $F_{5,18}=6.570$ ,  $p<0.001$ ), and groupings that included anemone, *Prionitis*, mussel shell, mussel tissue, and inert substrates were differentiated while plankton were indistinguishable from all.



# Figure 3

The relative abundance of the 10 OTUs that differed among the Tatoosh Island substrates.

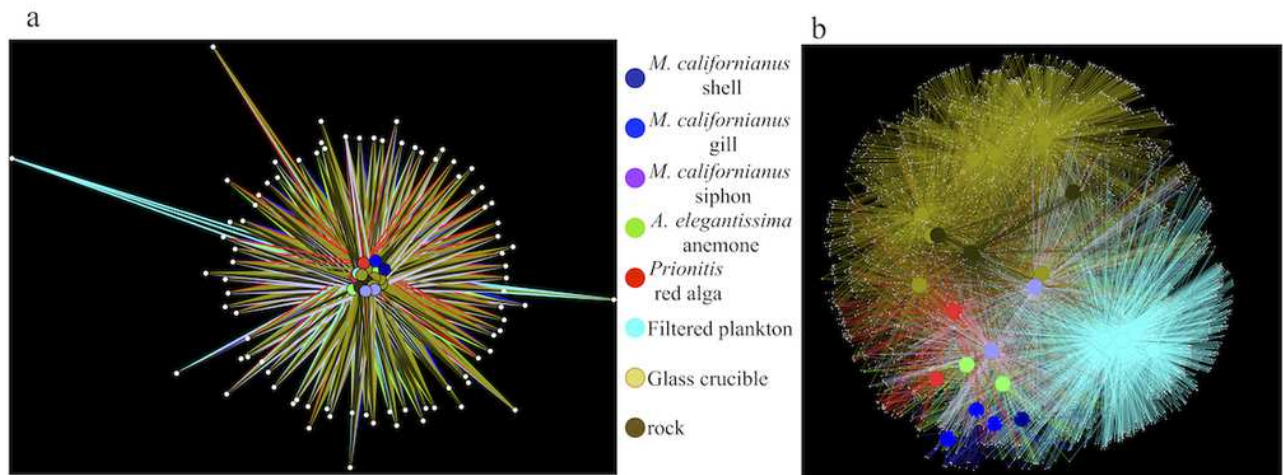
Figure 3. The relative abundance of the 10 OTUs that differed among the Tatoosh Island substrates (Boniferroni-corrected ANOVA,  $p < 0.05$ ).



# Figure 4

Shared OTU diversity among microbes sampled from the substrate groupings at Tatoosh Island and portrayed as a spring-embedded layout.

Figure 4. Shared OTU diversity among microbes sampled from the substrate groupings as in Fig 1, 2 and 3 at Tatoosh Island and portrayed as a spring-embedded layout, where OTUs that are in common bring nodes or samples together and OTUs that are distinct repel nodes. In a., only common OTUs detected more than 5,000 times are included, while b. shows only rare OTUs that were present 5 to 10 times across the entire dataset.

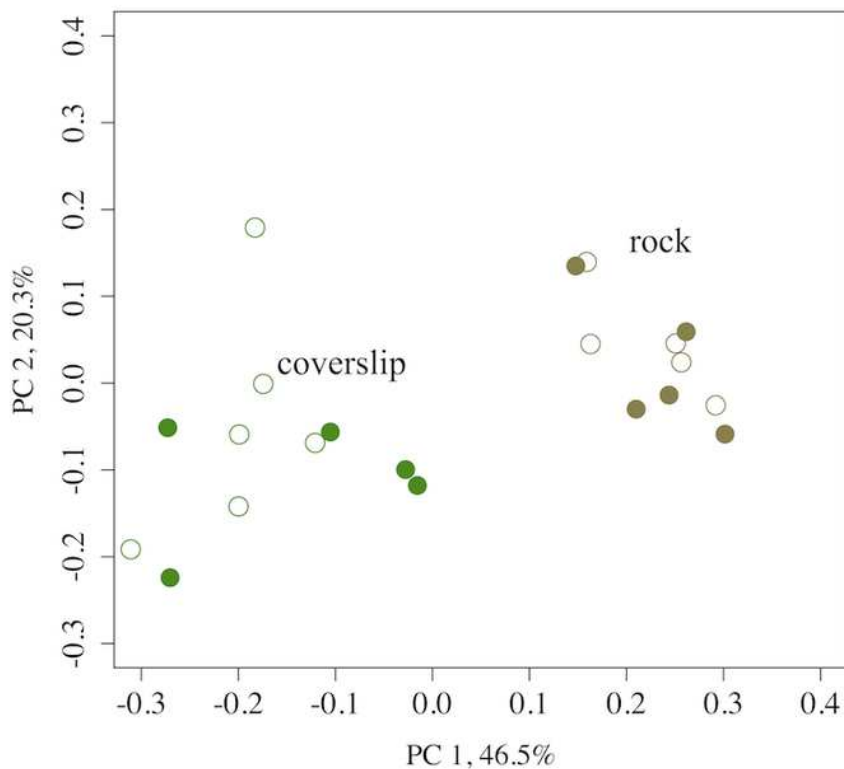




# Figure 5

A PCoA of the OTU diversity of tidepool rock versus coverslip substrates at Second Beach, WA.

Figure 5. A PCoA of the OTU diversity of tidepool rock (n=10) versus coverslip (n=12) substrates at Second Beach, demonstrating strong clustering among the microbial assemblages from the two substrates, while the presence of mussels (filled symbols) versus removal of mussels (open symbols) were not a factor for explaining beta diversity. Using weighted UniFrac, the first axis explained 46.5% of the variance, while the second explained 20.3%.



# Figure 6

Shared OTU diversity among microbes sampled from tidepool rock versus coverslip substrates in tidepools at Second Beach, WA.

Figure 6. Shared OTU diversity among a. microbes sampled from tidepool rock versus coverslip (lighter green) substrates and b. samples distinguished by whether mussels were present (blue) or absent (red) from tidepools at Second Beach. The spring-embedded layout shows OTUs that are in common bring nodes or samples together and OTUs that are distinct repel nodes. Only common OTUs greater than >5000 are included. Analyses of relatively rare OTUs did not change the network pattern.

