

1 **Using Encyclopedia of Life's TraitBank to identify plant traits associated with vulnerability**

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12

13 **Abstract**

14 Certain biological traits seem to predispose some species to greater extinction risk than others
15 and, when vulnerability information is limited, could be used as proxies to identify understudied
16 species likely in need of protection. In the past, identifying broadly applicable traits associated
17 with extinction risk has been hampered by the difficulty of collecting information for a broad
18 range of species (both geographically and taxonomically), with most comparative analyses
19 focusing on regional and/or taxon specific patterns. However, efforts to collect and compile
20 existing trait information from regional and taxon specific datasets into a single repository are
21 making it possible to analyze patterns between traits and vulnerability on ever broader scales.
22 We compared trait information from one such repository, the Encyclopedia of Life (EOL)
23 TraitBank, to information on threat status from the International Union for the Conservation of
24 Nature (IUCN) Red List to determine whether such consolidated data can help either clarify

25 previously identified associations or identify new associations between plant traits and threat risk.
26 Using generalized linear mixed effects models (GLMM's) we found five plant traits that could be
27 used to predict whether a species is threatened or not: 'plant growth form', 'life cycle habit', 'low
28 temperature tolerance', 'soil depth' and 'foliage porosity in winter'. Threatened plant species
29 tend to be trees rather than shrubs or grasses, live for more than one season (i.e. perennials), be
30 less tolerant of cold temperatures, require deep soil for good growth, and have less dense foliage
31 in the winter. Both 'plant growth form' and 'life cycle habit' have been identified as correlates of
32 vulnerability in past studies and both are relatively easy to use as diagnostic characters, which is
33 ideal for identifying understudied species of potential conservation concern. However, how these
34 two traits relate to vulnerability seems context dependent, with opposite relationships in past
35 regional or taxon specific studies. Unfortunately we could not identify further traits that could
36 add explanatory power to those relationships with the existing data on EOL's TraitBank but the
37 repository is still growing and the inclusion of additional traits and/or species from other datasets
38 could clarify these relationships further.

39

40 **Introduction**

41 Human activities are resulting in the loss of alarming numbers of species both regionally
42 (extirpations) and globally (extinctions), with some scientists referring to the current era as the
43 'sixth mass extinction event' in history (Pimm et al., 1995; Stork, 2010). The fossil record
44 indicates that there have been fewer mass extinction events for plants than animals in the past
45 (Willis & Bennett, 1995). However, since the arrival and settlement of humans, a combination of
46 direct exploitation, habitat change (loss, degradation and/or fragmentation), invasive species
47 introductions, and climate change has led to plant species declines, extirpations and extinctions
48 worldwide (Thomas et al., 2004; Hahs et al., 2009; Stork, 2010). Given limitations in money and

49 time for conservation, it would be impossible to implement conservation measures for the
50 298,000 plant species estimated worldwide (Mora et al., 2011), only a small proportion of which
51 have been studied (Vié et al., 2009). To best conserve the plant species remaining, strategies are
52 needed to identify and focus efforts towards those species that are likely to disappear fastest and
53 therefore need help most urgently.

54
55 Certain life history, ecological, and morphological traits appear to make some organisms more
56 vulnerable to human activities than others and, where direct vulnerability information is limited,
57 could be used as proxies to identify understudied plant species in need of conservation action.
58 Although relationships between plant traits and vulnerability have not been studied as well as in
59 other taxa, at least 80 of 118 plant traits assessed seem to have some relationship with at least one
60 metric of vulnerability (plant species abundance, distribution, threatened status, or rarity)
61 (reviewed in Murray et al., 2002; Walker & Preston, 2006; Stork et al., 2009; Neel & Che-
62 Castaldo, 2013; Musters et al., 2013; Bradshaw et al., 2008; McIntyre et al., 2014). Many of these
63 identified traits were only examined in a single study and, among those examined in multiple
64 studies (e.g. plant height, seed size, growth form), their relationships with vulnerability seem to
65 depend on the scale and context of the study (Murray et al., 2002). For example, shorter plant
66 species were more likely to go extinct than taller species after European settlement in New
67 Zealand (Duncan & Young, 2000), but taller species are more likely to have a threatened status
68 among legumes (family Fabaceae) on a global scale (Bradshaw et al., 2008). Establishing general
69 relationships between plant traits and vulnerability has been difficult partly because studies tend
70 to be limited in scope either geographically (e.g. Duncan & Young, 2000; Walker & Preston,
71 2006; Neel & Che-Castaldo, 2013; Musters et al., 2013), taxonomically (e.g. Bradshaw et al.,
72 2008), or both (e.g. Sodhi et al., 2008). A further difficulty is that most studies have only

73 analysed relationships with a single trait in isolation (reviewed in Murray et al., 2002), with some
74 more recent notable exceptions (e.g. Duncan & Young, 2000; Walker & Preston, 2006; Neel &
75 Che-Castaldo, 2013; Musters et al., 2013) and, therefore, may have missed important
76 relationships among traits. The limited scope of past studies likely stems from the difficulty of
77 collecting sufficient amounts of information about multiple traits for species from a broad
78 taxonomic range and on a global scale.

79
80 Efforts such as the Encyclopedia of Life's (EOL) TraitBank are making it possible to compare a
81 broader range of species (both geographically and taxonomically) and traits than was possible in
82 the past, allowing for more thorough analyses of plant traits associated with vulnerability. The
83 EOL TraitBank is "a searchable, comprehensive, open digital repository for organism traits,
84 measurements, interactions and other facts for all taxa across the tree of life"
85 (<http://eol.org/info/516>; Parr et al., 2014). The depository currently includes trait information for
86 thousands of plant species, compiled from published studies (e.g. Chave et al., 2009), regional
87 databases (e.g. the U.S. based USDA plants database (<http://plants.usda.gov>) and the
88 Mediterranean based BROT Database (<http://www.uv.es/jgpausas/brot.htm>)), as well as
89 databases specific to particular taxa (e.g. Shaw Nature Reserve Lily Database -
90 <http://www.missouribotanicalgarden.org/visit/family-of-attractions/shaw-nature-reserve.aspx>) or
91 biomes (National Wetland Plant List - <http://rsgisias.crrel.usace.army.mil/NWPL/>). Our goal in
92 this paper is to assess whether the consolidated data in EOL's TraitBank can help clarify
93 previously identified associations between plant species traits and their vulnerabilities and/or
94 identify relationships with previously untested traits.

96 **Methods**

97 *Compiling plant trait and vulnerability data*

98 For plant species trait data we relied on EOL's TraitBank (accessed on 15 July 2014),
99 transforming and/or combining some of the traits to make analyses easier. We manually searched
100 the TraitBank repository for all life history, morphological, or ecological traits that included
101 information for any plant species (Kingdom Plantae). Three traits contained information about
102 seed size ("seed mass", "seeds per pound", and "seeds per unit") so we converted each into
103 weight per seed (in grams) and combined them into a single "seed weight" trait, using averages
104 for any species with more than one value. We also combined two traits that indicate whether a
105 plant species has some commercial use ("commercial availability" and "uses") into a single
106 "commercially used" trait, assigning each species to a category of either "used" or "not used"
107 based on their original categories. Lastly, we converted the categorical "active growth period"
108 trait (indicating the seasons in which plant growth occurs for a given species), into a numeric
109 trait: "active growth duration", that indicates instead the number of seasons in the year (1-4) with
110 active growth.

111
112 For our metric of plant species vulnerability, we categorized each plant species as either
113 "threatened" or "lower risk", using the IUCN Red List extinction risk categories (IUCN, 2014;
114 accessed on 01 August 2014). The IUCN Red List assigns each evaluated species to one of nine
115 ordered risk categories, using the best data available and standard, rigorous criteria: Data
116 Deficient (DD), Least Concern (LC), Lower Risk (LR), Near Threatened (NT), Vulnerable (VU),
117 Endangered (EN), Critically Endangered (CR), Extinct in the Wild (EW), or Extinct (EX). We
118 searched the Red List for all assessed plant species and compiled information on their threat
119 status and taxonomy (kingdom, phylum, class, order, family, genus, species), removing any
120 species assessed as Data Deficient, since extinction risk is uncertain for those species. Since all

121 traits did not have representatives from each of the eight remaining extinction categories, we re-
122 assigned each species to one of two categories: “threatened” (VU, EN, CR, EW, EX), and “not
123 threatened” (LC, LR, and NT). The IUCN Categories and Criteria manual (IUCN, 2012),
124 indicates that three categories are “threatened” (VU, EN, CR). Although there were few species
125 within the two extinct categories (EW and EX), we also assigned those to our “threatened”
126 category, assigning all others to a “not threatened” category (LC, LR, NT).

127

128 *Analyzing relationships between plant traits and extinction risk categories*

129 To identify relationships between species plant traits and the probability of being in a threatened
130 Red List category we used generalized linear mixed effects models (GLMM’s) and a stepwise
131 model building approach. GLMM’s have been used to test relationships between plant traits and
132 vulnerability in several previous studies (e.g. Bradshaw et al., 2008; Sodhi et al., 2008). One of
133 its advantages is that GLMM’s can be used to test which combination of traits explain the most
134 variance in extinction risk among plant species while controlling for the effect of phylogeny.
135 Phylogenetically paired contrasts are another popular regression based method for comparing
136 traits and vulnerability (Purvis, 2008), but are less appropriate when categorical variables are
137 included in the analysis, as is the case for many plant traits in TraitBank.

138

139 We fitted GLMM’s to the plant trait and vulnerability data using the ‘lmer’ function from the
140 lme4 package in R (Bates et al., 2014). For each GLMM, we coded “threat” as a binomial
141 response variable and each plant trait as a fixed effect, with a logit link function (Bates, 2010). To
142 control for the fact that closely related species can have similar traits through shared evolutionary
143 histories, we controlled for pseudoreplication stemming from phylogenetic influences by

144 including taxonomy as a nested random factor in each GLMM (phylum/class/order/family). We
145 did not have sufficient replication within genera to include genus in the nested random factor.

146
147 To identify individual plant traits associated with threat risk that should be considered in more
148 complex models, we compared models with and without each trait. In each case, we compared
149 full models (including the trait with taxonomy as a random factor) with reduced models
150 (including only taxonomy as a random factor) using the ‘anova’ R function (R Core Team, 2013).
151 We calculated the percent deviance explained by inclusion of each trait in the model and retained
152 traits if their addition significantly reduced model deviance according to a Chi-squared test and
153 false discovery rate corrected significance levels to account for multiple comparisons (Zuur et al.,
154 2009). To further assess the relative strength of each full model, we calculated conditional and
155 marginal R^2 values according to Nakagawa & Schielzeth (2013). Other studies have used
156 Akaike’s Information Criteria (AIC) or Bayesian information criteria (BIC) to compare the
157 strengths of evidence for competing models (e.g. Bradshaw et al., 2008; Sodhi et al., 2008).
158 However, those criteria are less suitable for comparison among models built using different
159 datasets and many of the EOL TraitBank traits had information for different subsets of species
160 (Nakagawa & Cuthill, 2007).

161
162 To test whether combined traits explain more of the deviance than any individual trait, we built
163 GLMM’s with combinations of traits retained from the individual analyses and compared those
164 with GLMM’s with individual traits. Starting with pairs of traits with the highest significance (i.e.
165 lowest P-values), we used the ‘anova’ R function (R Core Team, 2013) to test for differences in
166 deviance between models with and without each trait in combination (Zuur et al., 2009; Bates,

167 2010). Variables were only retained if they were not correlated with a trait already in the model
168 and their inclusion significantly reduced model deviance according to a Chi-square test.

169
170 To determine whether there were any relationships among traits that could confound comparisons
171 with threat risk, we tested for correlations/independence using Spearman rank tests or Chi-square
172 tests, depending on the trait. For traits with continuous and/or ordered categorical values we
173 tested for covariance using Spearman correlations in the 'rcorr' function from the Hmisc package
174 in R (Harrell, 2014). To test for associations among categorical traits (ordered or unordered), we
175 used Chi-square tests of independence ('chisq.test' function in R; R Core Team, 2013).

176

177 **Results**

178 We found 92 life history, ecological, morphological, and usage traits on EOL's TraitBank that
179 include information for plant species, 47 of which were useful for our comparisons with threat
180 risk (i.e. had a plausible relationship with threat risk and sufficient sample sizes). Several traits
181 were management-based with no clear hypothesis for how they could relate to species
182 vulnerability (e.g., garden persistent range, germination treatment). The number of records
183 returned by our EOL TraitBank search ranged from 1 to 48973 for each of the plant traits we
184 identified, with an average (mean) of 2917 records. Our search of the IUCN Red List returned
185 19381 species, 1526 of which were assessed as Data Deficient, leaving 17856 species assessed in
186 one of the categories informative for our analyses (i.e. could be assigned to a "threatened" or "not
187 threatened category"). However, far fewer species (2340) had both threat risk information from
188 the IUCN Red List and information for at least one of the 47 plant traits from EOL useful for our
189 analysis, with each trait having information for a subset of those species (range = 1-1307 spp.;
190 mean = 316 spp.). Consequently, several traits had information for too few species for analysis

191 (e.g. 'submergence tolerance'). It was possible to fit GLMM's using traits with information for as
192 few as 14 species (i.e. 'seed ripening date'). However, it was not possible to combine traits with
193 such small sample sizes in more complex GLMM's.

194
195 According to GLMM analyses with individual traits, only five of the 47 traits we analyzed were
196 significantly associated with variation in threat risk among plant species (Table 1). In

197 comparisons of GLMM's with and without each trait, 'plant growth habit', 'low temperature
198 tolerance', 'life cycle habit', 'foliage porosity in winter', and 'soil depth' reduced between 2.7
199 and 16.8% of the deviance in whether a plant species was threatened or not, significant even after
200 correcting for multiple comparisons using the false discovery rate ($P < 0.01$). 'Plant growth habit'
201 had the most significant influence on reduction in deviance (i.e. lowest P-value) and resulted in
202 the GLMM with the highest relative strength in predicting threat risk (marginal and conditional
203 R^2 values > 0.98), indicating that trees have higher threat risk than shrubs and shrubs have higher
204 threat risk than forbs/herbs or graminoids (Fig. 1a). However, low temperature tolerance
205 explained more of the deviance in threat risk (17%) compared to its null model, indicating that
206 plants that can withstand colder temperatures are less likely to be threatened (Fig 2a).

207 Furthermore, plants that survive multiple seasons (perennials) were more often categorized as
208 threatened than annual species (Fig. 1b) and those that have denser foliage in the winter were at
209 greater risk than those with sparse foliage at that time of year (Fig. 1c). Finally, plants that need
210 deeper soil for good growth had higher probabilities of being in a threatened IUCN Red List
211 category (Fig 2b).

212
213 None of the GLMM models with multiple traits performed significantly better than individual
214 models, except for those containing correlated traits. The combination of foliage porosity in

215 winter and soil depth explained significantly more deviance than without either of those traits (p
216 < 0.02). However, those two traits are linked. Plants that require deeper soil tend to have less
217 foliage in the winter (Spearman rank test; p < 0.0001). Likewise, a combined model including life
218 cycle habit and low temperature tolerance explained significantly more deviance in threat risk
219 than without either trait (p < 0.005), but those two traits are negatively correlated: perennial
220 species tend to have lower temperature tolerances than annual or biennial species (Spearman rank
221 test; p < 0.0001).

222

223 **Discussion**

224 By comparing EOL TraitBank plant traits and IUCN Red List threat status information, we found
225 evidence supporting associations between five traits and threat risk, two of which have been
226 identified as correlates of vulnerability in the past: ‘plant growth habit’ and ‘life cycle habit’.
227 Plant growth habit (or plant growth form, as it is sometimes termed), has been found to be
228 associated with several metrics of vulnerability including distribution/rarity (reviewed in Murray
229 et al., 2002), extinction proneness (Sodhi et al., 2008), and threatened status (Bradshaw et al.,
230 2008), so it is perhaps not surprising that we found such a strong relationship, with a conditional
231 R^2 value of 0.99. However, the direction of the relationship between plant growth form and
232 vulnerability has differed somewhat among past studies. Our results are consistent with patterns
233 reported for plants in central Australia and Sydney, where shrubs and trees were found to be rarer
234 than herbs (Oakwood et al., 1993) and results among legume species, within which trees had
235 higher probability of being threatened than shrubs or herbs and shrubs had higher probabilities of
236 being threatened than herbs (Bradshaw et al., 2008). In contrast, our results differ from those
237 reported for British plants, where trees are less likely to be rare (Kelly & Woodward, 1996), and
238 for Australian *Eucalyptus* species in which trees are more common than shrubs (Murray et al.,

239 2002). Our results for 'life cycle habit' (sometimes referred to as 'longevity') also differ from
240 results reported for rarity in grassland plants, among which annuals and biennials were less
241 abundant with narrower distributions than perennials (Eriksson & Jakobsson, 1998). Our hope
242 was to clarify previously identified relationships between plant species traits and their
243 vulnerabilities but it appears that these two traits may be context dependent. Since we were
244 unable to find any associations with the other traits in our analysis that might account for some of
245 this context, further analyses may be needed to explore additional factors.

246
247 In addition to the two previously identified traits, we found associations between threat risk and
248 three apparently untested traits: 'low temperature tolerance', 'soil depth' and 'foliage porosity in
249 winter'. Although we were unable to find any comparative analyses that examine relationships
250 between low temperature tolerance and vulnerability, bioclimatic models predict that plant
251 species with narrower tolerance ranges (including temperature tolerances) will be more
252 vulnerable to anthropogenic climate change (Midgley et al., 2002). High temperature tolerance
253 may be more important as warming continues, but low temperature tolerance may also impart
254 some resilience to the increase in stochastic events predicted to occur with climate changes.
255 Unfortunately we did not have access to high temperature tolerance data but it would be
256 interesting to compare the two to determine their relative and combined influence on
257 vulnerability. 'Soil depth' is essentially another measure of abiotic tolerance; plants requiring
258 deeper soil should be more vulnerable as it could limit the possible range of environments they
259 can inhabit, thus limiting their ability to spread into new areas when habitats are disturbed or
260 altered through climate change. Possible mechanisms linking foliage porosity in winter and threat
261 risk are less clear. In general, plants that can go through periods of dormancy can also withstand
262 dramatic changes (Stork et al., 2009). High porosity in winter could be linked to a plant's ability

263 to shed leaves during periods less suitable for growth. However, if that were the case we would
264 expect to find a relationship between ‘shedability’ (deciduous or evergreen) and threat risk, but
265 we did not. Since foliage porosity is correlated with soil depth, it could also be that it is this
266 indirect connection that is responsible for the relationship with threat risk.

267
268 Interestingly, we did not find significant relationships among other traits previously hypothesized
269 to be predictive of plant vulnerability. In particular, plant height has been linked to vulnerability
270 in at least three separate studies, with some indicating short plants are more vulnerable (Mitchley,
271 1988; Duncan & Young, 2000; Walker & Preston, 2006), and another indicating tall species are
272 at greater risk (Bradshaw et al., 2008). However, these studies were either regionally based
273 (Mitchley, 1988; Duncan & Young, 2000; Walker & Preston, 2006) or focused on specific plant
274 taxa (Bradshaw et al., 2008). If the direction of the relationships between plant height and
275 vulnerability depends on the context, as suggested by these past studies, we may not have found a
276 relationship with this trait because it is not generalizable on broader geographic and taxonomic
277 scales. Seed production is another trait that has been linked to vulnerability in several studies but,
278 unlike relationships with plant height, results have been consistent with more productive species
279 being less vulnerable (reviewed in Murray et al., 2002). We did not find a significant relationship
280 between the associated EOL trait ‘fruit/seed abundance’ and threat risk. However, our analysis
281 was hampered by the limited number of species from a threatened category with this information
282 (only 15 spp.).

283
284 There are several possible reasons why we may not have found more significant relationships
285 between extinction risk and the plant traits we used, including (1) choosing inappropriate traits,
286 (2) biases or inconsistencies in either the predictive variables (traits) or the response variable

287 (threat risk), or (3) insufficient sample sizes for each trait and their combinations. We included a
288 large number of potential traits, many of which were hypothesized to have an influence on
289 vulnerability in previous studies. However, it is possible that we missed some other important
290 traits or other related variables in our analysis. One of the drawbacks of our analysis (and many
291 other similar analyses) is that we only had one trait value for each species. It could be that the
292 variability in traits within a species is as important or more than the mean value for a species (e.g.
293 phenotypic plasticity). However, trait databases rarely include intra-specific variability. In
294 addition to intrinsic traits such as those that we used, other comparative analyses are including
295 extrinsic factors (e.g. rainfall, human population density, invasive species) to identify correlates
296 of vulnerability and extinction risk in other taxa (Murray et al., 2011; Tingley et al., 2013). While
297 these extrinsic factors have proven useful for regional analyses, it could be difficult to summarize
298 such spatially explicit factors for species with large ranges. In all broad comparative analyses of
299 extinction risk, such as this one, there are also concerns about whether the data represent an
300 unbiased sample (González-Suárez et al., 2012). In our data, for example, there could be more
301 information for species that are more conspicuous or have commercial uses. However, as part of
302 our analysis we examined the spread of the data among groups and found no obvious skew in the
303 data. In addition to biases, there could be important differences in the way variables are
304 measured. For instance, other comparative analyses have included more categories in their 'plant
305 growth form' trait than used in the EOL TraitBank (e.g. epiphytes: Sodhi et al., 2008; climbers:
306 Bradshaw et al., 2008). Lastly, our analysis could have been limited by small sample sizes for
307 each trait. It is interesting that the traits we found to be significant in our analysis tended to have
308 information for more species than traits which were not significant (means = 631 spp. vs. 211
309 spp.), but the difference was not significant (t-test, $P > 0.05$).

310

311 The EOL TraitBank has information for a large variety of traits, covering a broad range of
312 organisms (both taxonomically and geographically) but, unfortunately, that was not always the
313 case for individual traits and when focusing on species with threat risk information from IUCN.
314 The 47 traits we were able to use for our analysis is greater than almost all other comparative
315 analyses of extinction risk for plant species (with the notable exception of Musters et al., 2013).
316 Furthermore, even though the numbers of species with both trait and threat risk information was
317 far fewer than the number for each, the mean number of species for each trait in our analyses
318 (316 spp.) was still greater than used in many similar past studies (reviewed in Murray et al.,
319 2002). If we had included regional endangered species in our analyses we might have been able
320 to increase the sample size. However, including regional categorizations could have been
321 problematic because some species can be endangered in one region and not in another. We
322 believed that using the IUCN Red List was the most appropriate method to assess whether a
323 species is assessed throughout its range. Since there were more plant species with threat risk data
324 from IUCN than for most traits, it would also make more sense to look for additional trait
325 information rather than additional threat status information, especially since one of the ultimate
326 goal of this research was to identify relationships that could be used to guide further conservation
327 assessments. There are other plant trait databases that contain information for more species than
328 used in our analyses (e.g. TRY Plant Trait Database: <http://www.try-db.org/TryWeb/Home.php>;
329 GLOPNET Global Plant Trait Network: <http://bio.mq.edu.au/~iwright/glopian.htm>). Fortunately,
330 the EOL TraitBank is still adding new datasets such as these to their repository, with the potential
331 for even stronger analyses in the future. Perhaps the greatest potential for using the EOL
332 TraitBank repository, though, is in even broader scale analyses than in this study, such as those
333 searching for general trends even among kingdoms (e.g. Stork et al., 2009; Musters et al., 2013).
334

335 **Conclusions**

336 Since only a small proportion of known plant species have been studied enough to know their
337 risk of extinction, and there are insufficient resources available to protect them all, strategies are
338 needed to focus efforts towards those species are most in need of conservation. This is
339 particularly important given that human activities represent some of the greatest risks to these
340 species. Our study adds to the growing body of evidence showing that certain life history,
341 ecological, and morphological traits predispose species to higher vulnerabilities. By comparing
342 EOL TraitBank plant traits with IUCN threat risk categories, we were able to identify five such
343 traits: ‘plant growth habit’, ‘life cycle habit’, ‘low temperature tolerance’, ‘soil depth’ required
344 for good growth, and ‘foliage porosity in winter’. Threatened plant species tend to be trees rather
345 than herbs or grasses, live for more than one season (perennials), are able to withstand colder
346 temperatures, can live in shallower soil depths, and have less foliage during the winter. In the
347 absence of direct vulnerability information, these traits, among others found in past studies, could
348 be used to guide further conservation assessments towards those most likely at risk. Ideally, such
349 traits should be easy to measure and generalizable. Fortunately, two of the traits we identified
350 (plant growth habitat, and life cycle habit) seem to satisfy both of these requirements, although
351 their specific relationships may be context dependent. Further analyses using other plant trait
352 databases could clarify these relationships further.

353

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362

363 **References**

364 Bates D. 2010. lme4: Mixed-effects modeling with R. URL <http://lme4.r-forge.r-project.org/book>.

366 Bielby J, Cardillo M, Cooper N, Purvis A. 2010. Modelling extinction risk in multispecies data
367 sets: phylogenetically independent contrasts versus decision trees. *Biodiversity and*
368 *Conservation* 19:113–127.

369 Bradshaw C, Giam X, Tan H. 2008. Threat or invasive status in legumes is related to opposite
370 extremes of the same ecological and life history attributes. *Journal of Ecology* 96:869–883.

371 Chave J, Coomes D, Jansen S, Lewis SL, Swenson NG, Zanne AE. 2009. Towards a worldwide
372 wood economics spectrum. *Ecology letters* 12:351–66.

373 Duncan R, Young J. 2000. Determinants of plant extinction and rarity 145 years after European
374 settlement of Auckland, New Zealand. *Ecology* 81:3048–3061.

375 Eriksson O, Jakobsson A. 1998. Abundance, distribution and life histories of grassland plants: a
376 comparative study of 81 species. *Journal of Ecology* 86:922–933.

377 González-Suárez M, Lucas PM, Revilla E. 2012. Biases in comparative analyses of extinction
378 risk: mind the gap. *The Journal of animal ecology* 81:1211–22.

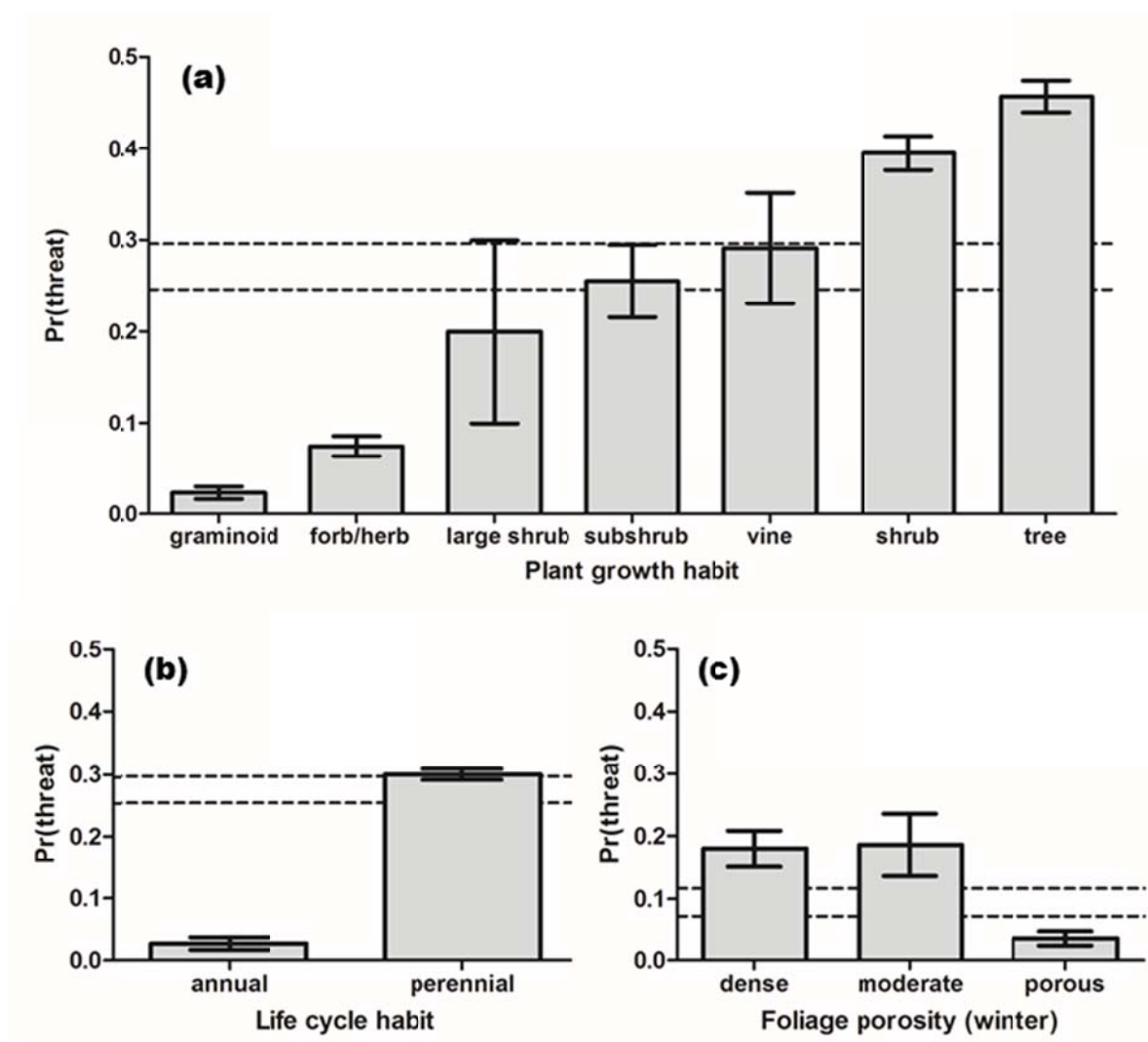
379 Hahs AK, McDonnell MJ, McCarthy M a, Vesk P a, Corlett RT, Norton B a, Clemants SE,
380 Duncan RP, Thompson K, Schwartz MW et al. 2009. A global synthesis of plant extinction
381 rates in urban areas. *Ecology letters* 12:1165–73.

382 Kelly C, Woodward F. 1996. Ecological correlates of plant range size: taxonomies and
383 phylogenies in the study of plant commonness and rarity in Great Britain. *Philosophical*
384 *Transactions of The Royal Society of London. Series B, Biological Sciences* 351:1261–1269.

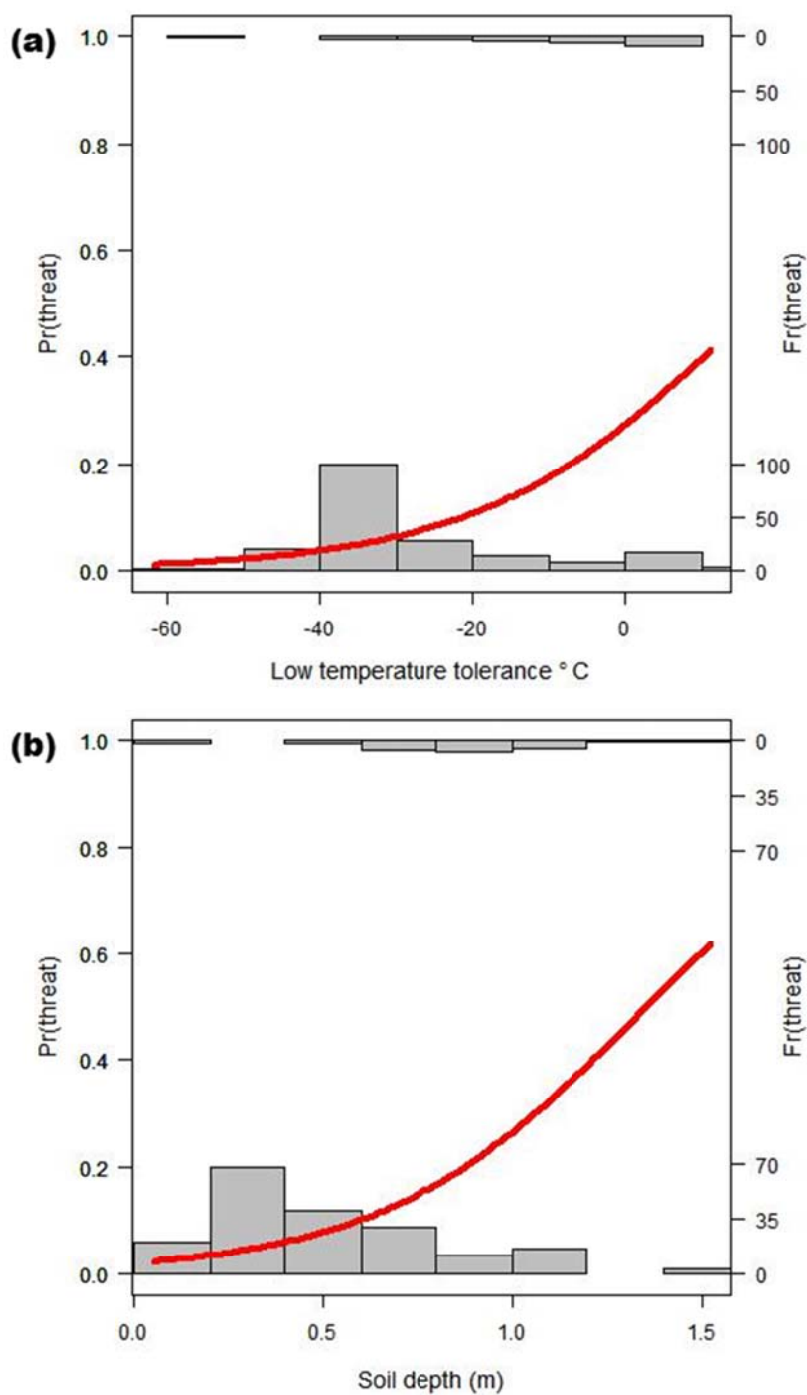
385 McIntyre S, Lavorelt S, Tremont RM. 2014. Plant life-history attributes : their relationship to in
386 herbaceous disturbance response vegetation. 83:31–44.

- 387 Midgley G, Hannah L, Millar D, Rutherford M, Powrie L. 2002. Assessing the vulnerability of
388 species richness to anthropogenic climate change in a biodiversity hotspot. *Global Ecology*
389 *and Biogeography* 11:445–451.
- 390 Mitchley J. 1988. Control of relative abundance of perennials in chalk grassland in southern
391 England. II. Vertical canopy structure. *Journal of Ecology* 76:341–350.
- 392 Mora C, Tittensor DP, Adl S, Simpson AGB, Worm B. 2011. How many species are there on
393 Earth and in the ocean? *PLoS Biology* 9:e1001127.
- 394 Murray K a, Rosauer D, McCallum H, Skerratt LF. 2011. Integrating species traits with extrinsic
395 threats: closing the gap between predicting and preventing species declines. *Proceedings.*
396 *Biological sciences / The Royal Society* 278:1515–23.
- 397 Murray BR, Thrall PH, Gill a. M, Nicotra AB. 2002. How plant life-history and ecological traits
398 relate to species rarity and commonness at varying spatial scales. *Austral Ecology* 27:291–
399 310.
- 400 Musters CJM, Kalkman V, van Strien A. 2013. Predicting rarity and decline in animals, plants,
401 and mushrooms based on species attributes and indicator groups. *Ecology and evolution*
402 3:3401–14.
- 403 Nakagawa S, Cuthill IC. 2007. Effect size, confidence interval and statistical significance: a
404 practical guide for biologists. *Biological Reviews* 82:591–605.
- 405 Nakagawa S, Schielzeth H. 2013. A general and simple method for obtaining R² from
406 generalized linear mixed-effects models. *Methods in Ecology and Evolution* 4:133–142.
- 407 Neel MC, Che-Castaldo JP. 2013. Predicting recovery criteria for threatened and endangered
408 plant species on the basis of past abundances and biological traits. *Conservation biology :*
409 *the journal of the Society for Conservation Biology* 27:385–97.
- 410 Oakwood M, Jurado E, Leishman M, Westoby M. 1993. Geographic ranges of plant species in
411 relation to dispersal morphology, growth form and diaspore weight. *Journal of*
412 *Biogeography* 20:563–571.
- 413 Parr CS, Wilson N, Schulza KS, Leary P, Hammock J, Rice J, R.J. C. 2014. TraitBank: Practical
414 semantics for organism attribute data. *semantic-web-journal.net*.
- 415 Pimm SL, Russell GJ, Gittleman JL, Brooks TM. 1995. The future of biodiversity. *Science (New*
416 *York, N.Y.)* 269:347–50.
- 417 Purvis A. 2008. Phylogenetic approaches to the study of extinction. *Annual Review of Ecology,*
418 *Evolution, and Systematics* 39:301–319.

- 419 Sodhi NS, Koh LP, Peh KS-H, Tan HTW, Chazdon RL, Corlett RT, Lee TM, Colwell RK, Brook
420 BW, Sekercioglu CH et al. 2008. Correlates of extinction proneness in tropical angiosperms.
421 *Diversity and Distributions* 14:1–10.
- 422 Stork NE. 2010. Re-assessing current extinction rates. *Biodiversity and Conservation* 19:357–
423 371.
- 424 Stork NE, Coddington J a, Colwell RK, Chazdon RL, Dick CW, Peres C a, Sloan S, Willis K.
425 2009. Vulnerability and resilience of tropical forest species to land-use change.
426 *Conservation biology : the journal of the Society for Conservation Biology* 23:1438–47.
- 427 Thomas CD, Cameron A, Green RE, Bakkenes M, Beaumont LJ, Collingham YC, Erasmus BFN.
428 2004. Extinction risk from climate change. *Nature*:145–148.
- 429 Tingley R, Hitchmough R a., Chapple DG. 2013. Life-history traits and extrinsic threats
430 determine extinction risk in New Zealand lizards. *Biological Conservation* 165:62–68.
- 431 Vié JC, Hilton-Taylor C, Stuart SN. 2009. *Wildlife in a Changing World: An Analysis of the 2008*
432 *IUCN Red List of Threatened Species*. Gland, Switzerland: International Union for the
433 Conservation of Nature.
- 434 Walker KJ, Preston CD. 2006. Ecological Predictors of Extinction Risk in the Flora of Lowland
435 England, UK. *Biodiversity & Conservation* 15:1913–1942.
- 436 Willis K, Bennett K. 1995. Mass extinction, punctuated equilibrium and the fossil plant record.
437 *Trends in ecology & evolution* 10:308–309.
- 438 Zuur AF, Ieno EN, Walker N, Saveliev AA, Smith GM. 2009. *Mixed Effects Models and*
439 *Extensions in Ecology with R*. Springer.
- 440



441 Figure 1: Probabilities of plant species being classified as 'threatened' (Pr(threat)) for each level
 442 of the three categorical traits found to significantly reduce deviance when added to GLMM's that
 443 only consider taxonomy as a nested random factor. Error bars represent standard deviations of
 444 bootstrapped probabilities (10 000 iterations using 70% of the data). Dotted horizontal lines
 445 represent the range in fitted values from the associated GLMM with each trait.



446 Figure 2: Binned frequencies showing how often plant species are classified as either 'threatened'
 447 (=1) or 'not threatened' (=0) (Fr(threat); gray bars), and modelled probabilities of a plant species
 448 being 'threatened' (Pr(threat); red line) for each of the two continuous traits found to significantly
 449 reduce deviance when added to GLMM's that only consider taxonomy as a nested random factor.

450 Table 1: Generalized linear mixed effects model (GLMM) results for the five plant traits that significantly reduced deviance in threat
 451 risk (i.e. whether a species is listed in a threatened IUCN Red List category or not) when added to null models that only include the
 452 effect of taxonomy. Results shown include numbers of species used in each model (*Spp*), maximum log-likelihoods (*LL*), marginal R^2
 453 values ($R^2_{GLMM(m)}$), conditional R^2 values ($R^2_{GLMM(c)}$), percent deviance explained (*%DE*), and results from chi-square tests comparing
 454 models with and without the trait (*Chi-square* and *P*).

Trait	<i>Spp</i>	<i>LL</i>	$R^2_{GLMM(m)}$	$R^2_{GLMM(c)}$	<i>%DE</i>	<i>Chi-square</i>	<i>P</i>
Plant growth habit	1192	-396.442	0.982	0.996	9.468	82.917	1.26×10^{-14}
Life cycle habit	1307	-479.457	0.154	0.894	2.740	27.016	1.36×10^{-06}
Low temperature tolerance	222	-57.651	0.273	0.878	16.803	23.287	1.40×10^{-06}
Soil depth	213	-63.162	0.215	0.717	8.057	11.070	0.0008777
Foliage porosity (winter)	219	-60.911	0.279	0.856	9.232	12.391	0.002039

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