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Landscapes of fear: from trophic cascades to applied management and population ecology

Sonny S Bleicher ^{Corresp.} ¹

¹ Ecology and Evolutionary Biology, University of Arizona, Tucson, Arizona, United States

Corresponding Author: Sonny S Bleicher
Email address: bleicher.s.s@gmail.com

Predator-Prey dynamics, and their trophic impacts, have functioned as a focal point in both community and population biology for five decades. The work-group focusing on these dynamics has however largely changed the focus of their work from trophic effects to the study of non-consumptive effects of predation-- the “ecology of fear”. An increasing number of studies chose to spatially chart wildlife populations’ risk assessment and of those the majority use optimal patch-use (giving-up densities) as a continuous measure of fear. These charts, “landscapes-of-fear” (LOFs) originated in conservation literature and the reintroduction of wolves to Yellowstone. Today, they are used to study population habitat selection and venture into the evolutionary context with studies examining the mechanisms by which species coexist in the same physical space. This review predicts increase in, and encourages the use of, LOFs: as a conservation tool to assess species land-use; as a bridge between ecology and neurology with stress hormones as indicators fear; and as a tool to compare species’ evolutionary dynamics within a community context.

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2 ecology

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4 Sonny S Bleicher

5 Department of Ecology and Evolutionary Biology, University of Arizona, Tucson, Arizona,

6 United States

7

8 **Introduction**

9 The study of community ecology has developed from a study of how species affect each
10 other in terms of resource competition to the study of how that competition over evolutionary
11 history; i.e. how the ability of species to extract resources impacted the interactions between
12 species historically, and to the community structure we observe at the present. This historical
13 shift can be traced back to the models that formed the study of trophic cascades, and with that the
14 birth of the predator-prey dynamics research group. Naturally this group continued to focus and
15 narrow the research, eventually leading to the study of non-consumptive effects of predators on
16 entire communities colloquially named “the ecology of fear”.

17 While the ecology of fear continued to focus on means by which the community structure
18 impacts specific species behaviors, some of the researchers involved chose to broaden the study
19 onto a system level. These researchers began to assess ecosystem health using the trophic
20 cascades as the basis for a new theory of behavioral cascades reverberating down the food chain
21 and affecting habitat selection of species along the chain.

22 John Laundré (2001) called this effect “Landscapes of fear” (LOF), and it has gained
23 dominance in studies ever since (Figure 1). This review is directed to analyze how the LOF has
24 changed from a theory based in trophic cascades and became a useful tool for conservation and a
25 measurable attribute of population behavior. This review is intended to build on the previous
26 review (Laundre *et al.*, 2010) and will guide the reader through three major sections: (1) the
27 historical background leading to the development of the ecology of fear research group and
28 segue to the early applications of the LOF. (2) It will discuss the methods by which researchers
29 worldwide have measured the LOF and how they applied the theory. And lastly, (3) this review
30 will make predictions for the directions in which the study of the LOF is developing and identify
31 possible applications for future use.

32 **Historical base: Trophic cascades**

33 The world is green because predators manage the populations of grazing species
34 concluded Hairston (Hairston, Smith & Slobodkin, 1960) in what later became known as the
35 “Green World Hypothesis”. Since this ground breaking paper a large focus in community
36 ecology was directed at the study of trophic cascades, i.e. the idea that through direct predation
37 the size of a population impacts multiple trophic levels below it and the size of the population of
38 prey dictates the availability of resources that sustain the predators above them.

39 The textbook example used to teach these interactions is population fluxes in Lotka-
40 Volterra of lynx-hare (and mastings) predator-prey cycles in Canadian boreal forests (Hewitt,
41 1921; Fox & Bryant, 1984; Krebs *et al.*, 1995; Lima, 1998). This example highlights the trophic
42 cascades from a bottom-up perspective, i.e. how the availability of resources influences the
43 populations of predators. Resource availability dictates the availability of niches for species to
44 occupy in the community; however, the top-down interactions greatly dictate the traits the

45 species must have to survive within those niches. Robert Paine pioneered the study of the
46 predation implications on prey when he described how a diverse predator community results in
47 diverse prey community (Paine, 1963). In this example, the predation pressure from multiple
48 intertidal zone predators (sea stars) removed the competitive ability of the dominant barnacles
49 and allowed for higher mollusk diversity. More recently, Schmitz and Price (2011) showed a
50 strong positive correlation between the biomass of arachnid predators on the vegetation biomass
51 in an agricultural system. In this system, spiders feed on grasshoppers that damage vegetation.
52 The biomass of spiders thus positively correlates with the health of the vegetation crop. These
53 examples show that the predators negatively affect the prey populations and thus indirectly have
54 positive effects on the vegetation. However a few questions remained unanswered: (1) Do
55 predators manage the populations of prey solely by consuming them? And (2) would the
56 evolution of prey species to manage the risk of predation not overcome the negative impacts
57 these predators have on their prey populations? Thus a study of non-consumptive predator
58 effects was created, i.e. the “ecology of fear”.

59 **Non- Consumptive Effects of Predators: an “Ecology of Fear”**

60 One does not have to study wildlife behavior to understand the effects predation risk has on
61 animal behavior. All one has to do, is think of our own body’s response to a risky situation. Our
62 bodies instinctively respond to the risk in the environment by producing stress hormones. This
63 production results in increased blood pressure, pulse and sensory sensitivity. In essence, our
64 bodies tell us to get out of harm’s way as fast as we can. In this example of our own day-to-day
65 life we find the evidence for the millions of years evolutionary race between predators and prey
66 to maintain the energetic needs of both groups. Prey evolve mechanisms to avoid the risk, and
67 predators find ways to out-gun these defense mechanisms.

68 In non-human examples we expect that predators would have to find a balance between over-
69 hunting a naïve food resource into extinction and starvation from aiming to hunt over vigilant
70 prey (Brown & Vincent, 1992; Lima, 1998, 2002; Wolf & Mangel, 2007; Brown, 2010). The
71 predators must manage the fear in the prey into an optimal state of vigilance (Embar, Mukherjee
72 & Kotler, 2014) and the prey counteracts the predator management through a variety of
73 behavioural strategies, or choices. I will examine a few case studies of such strategies.

74 Many species choose their habitat based on the risk which this habitat poses to them. For
75 example, heteromyid rodents choose habitat based on the density of vegetation. Species
76 (kangaroo rats) that can hop out of harm's way prefer *the* un-encumbered open, and species that
77 are limited in predator evasion strategies (pocket mice) find shelter under thick vegetation
78 (Rosenzweig, 1973; Bleicher, 2014). In kangaroo rats specifically the presence of vipers was
79 shown to be a driver of the choice of the open habitat (Bouskila, 1995). In another example,
80 Gerbilline rodents responded to owls with clear preference for sheltered microhabitats (Kotler,
81 Blaustein & Brown, 1992; Abramsky et al., 1996; Rosenzweig, Abramsky & Subach, 1997;
82 Kotler et al., 2002; St. Juliana et al., 2011; Embar et al., 2014). Habitat fragmentation, or edge
83 effect, has been shown to affect the habitat use (predominantly in the form of avoidance) by song
84 birds (Storch, Woitke & Krieger, 2005; Fischer & Lindenmayer, 2007). The final example is of
85 larger animals (ungulates and primates) selecting habitat where the sightlines allow good
86 visibility of approaching predators (Tadesse, 2012; Abu Baker & Brown, 2013; Sandford, 2013;
87 Coleman & Hill, 2014).

88 Other strategic choices of prey driven by the predator-prey dynamics can include dietary
89 selection and movement patterns. For example (again in the Heteromyid rodents), foragers better
90 equipped for risk management (kangaroo rats) have a more diverse diet than the foragers less

91 well equipped for predator management (pocket mice) who forage what they can (Davidson,
92 Brown & Inouye, 1980). In another example, in the Simpson Desert, dasyurid marsupials avoid
93 risk by covering large distances to search for refuge. These small mammals (20 grams on
94 average) inhabit burrows in the swale of sand dunes located away from the resource dense
95 habitats at the dune crests (Haythornthwaite, 2005; Haythornthwaite & Dickman, 2006). All of
96 these adaptations clearly suggest that a trade-off between resources and predation risk occurs
97 within a spatial dynamic, and as such these predator-prey games can and should be studied using
98 a spatial analysis, a “landscape of fear”.

99 **Spatial Analysis of Risk and Resources: A “landscape of fear”**

100 **A. Theory**

101 Two major research labs (Brown and Laundré), cooperatively pioneered the development of
102 theory of the LOF (Laundre, Hernandez & Ripple, 2010). Initially, the landscapes of fear were
103 proposed as metaphor for the implications of the predators on prey behavior on a landscape level.
104 It was used to predict the implications of reintroduction of predators such as wolves on the
105 populations of elk and bison (Laundré, Hernández & Altendorf, 2001). Only a relatively low
106 proportion of the literature actually develops the theory (Figure 2) and further information on its
107 development may be found in Laundré (2010). The theory was expanded to study the effects of
108 landscape heterogeneity (Brown & Kotler, 2004) culminating in measured maps combining
109 vegetation, refuge, resource availability and risk. In an example with cape ground squirrels
110 (*Xerus inauris*), the LOF was interpreted as a cost benefit analysis of energetic values over
111 change in the landscape (Joules/ meter) (Van Der Merwe & Brown, 2008). This conversion
112 allowed researchers to weigh the study of energetic gain from food patches and compare those
113 gains to the cost of risk avoidance (*cf.* Brown 1988)

114 It is important to draw attention to a common misinterpretation on the LOF theory. The LOF
115 is not the study of how the features in the landscape cause the organisms to assess risk
116 differently, but rather how a population of organisms sense the risk in the environment as a
117 continuous variable of the landscape itself. This virtual, spatially dependent, variable is
118 influenced by multiple factors: (1) the intensity of predation the population senses in the
119 environment (predator community structure and activity); (2) the population size (safety in
120 numbers); (3) inter-species competition (competition may embolden populations when resources
121 are scarce) and (4) the energetic state of the population (hungry animals will take more risk).
122 Some features of physical environment influence the LOF (e.g. blocked sightlines). However
123 given the dependency on all the above, one should observe these separately, or combine these
124 features with a spatial analysis.

125 Descriptive attributes of the LOF began taking increased priority in the literature as the
126 variety of applications increased (see section on applications below). The various attributes of
127 the LOF drew inspiration from other ecological and evolutionary theories to describe the zones
128 of different risk characteristics. In a study of striped mice, the features of risky habitat was
129 described as “islands” of fear, a reference to the island biogeography theory and the SLOSS
130 debate, emphasizing the impacts of both borders and edges and the distribution of safety zones in
131 the environment (Abu Baker & Brown, 2010). In the previous review Laundré et al. (2010)
132 prefer to describe the landscape features as valleys (safe) and peaks (risky) in an aim to show
133 that risk assessment is a quantitative attribute and not a binomial characteristic (two distinct
134 outcomes of risk or safety). Lastly in an aim to quantify this rate of change, my own work has
135 developed an approach to measuring the rate of change of perceived risk in the LOF.

136 This measure can be described as the rugosity of the landscape (Bleicher, Kotler & Brown,
137 2012; Bleicher, 2014). A highly rugose landscape (highly variable with steep changes between
138 points) implies that the population perceives the risk as localized. In comparison, flat landscapes
139 can be interpreted as the result of one of two behavioral assessments (or strategies). (1) A flat
140 LOF may be the result of a very “fearful” population whereas the majority of the entire
141 environment “plateaus” on a high risk contour. In such a LOF, the major focus of the behavior
142 remains in contact with the locations of refuge in the landscape and the risk lessens gradually as
143 one moves near the refuge. Alternately, (2) a population that is “secure” in its management
144 ability of predation risk from the predators in the environment will have a very flat landscape. In
145 this scenario, the zones of risk are less focused and tangible and thus the change between
146 “riskier” and “safer” zones is gradual and not very distinct.

147 **B. Measuring a Landscape of Fear**

148 Fear has many definitions based on the field of biology in which it is studied. In behavioral
149 ecology, the most common definition is the assessment of risk based on a set of criteria
150 dependent on learned and inherited information that drives strategic decision- making (Vincent
151 & Brown, 2005). In most studies that includes information on the lethality of the predators a prey
152 animal may encounter. This variable combines (1) the likelihood a predator will be encountered,
153 (2) environmental factors (vegetation, illumination, resources, etc.) contributing to the likelihood
154 of falling prey and (3) the amount of energy that must be expended to avoid the predation risk
155 (Brown, 1988). Two major theories were derived from that definition (optimal patch use and
156 optimal vigilance), and from those theories tools were developed to measure the perceived risk in
157 the environment (Brown, 1999).

158 Risk assessment in wildlife traditionally has been quantified using measures of activity,
159 such as presence or absence of individuals of the species in a habitat. These activity measures
160 present significant difficulties in this type on interpretation. The two general methods that have
161 been used are mark-recapture surveys (Ritchie & Johnson, 2009) and diversity surveys (Rösner
162 et al., 2014) to estimate population density in different habitats. Additionally, in birds, counting
163 the number of active nest sites in different habitat types (Zanette & Jenkins, 2000) served a
164 similar function equivalent to that of mark-recapture surveys. These methods alone are likely not
165 the strongest measure of risk as they are confounded by multiple external factors. For instance,
166 nesting areas and foraging areas likely do not overlap to any significant degree. This can be seen
167 in the above mentioned examples of dunnarts in the Simpson Desert that exhibit a large daily
168 migration pattern. Studies found that these animals transverse a large distance between the
169 burrows they occupy during the day and the locations in which they forage (Dickman, Predavec
170 & Downey, 1995; Haythornthwaite & Dickman, 2006). As a result, the animals may be caught in
171 the path they transverse nightly. Similarly, if burrows (or nests) were counted in the dune crests
172 where the dunnarts forage, (and likely encounter the majority of predators,) we would conclude
173 that this habitat is avoided by the foragers when in fact it is the lack of burrows dug by other
174 species that drives the nightly migration. Thus caution is suggested in the use of surveys alone to
175 quantify a LOF. In addition to using these methods, a stronger (and more reliable) way to
176 measure perceived risk is to measure foraging for small animals and vigilance for larger species.
177 Both foraging behavior and vigilance can be used in ways that are independent of other spatially
178 dependent variables. Examples of such spatially dependent variables are the chance of being
179 trapped for mark-recapture surveys and nest site availability for nests or borrows.

180 Aldo Leopold in observation on the Kaibab Plateau in Arizona proclaims that one can tell the
181 stress (hunger state) on mule deer population by the height of the browse lines on the white pines
182 (Leopold 1949, pg. 56). From this, and many other observations, ecologists have derived that
183 foraging behavior can explain much of the state of the population that left the marks behind;
184 Leopold states: “one need no doubt the unseen” (pg. 57), meaning that one does not need to
185 observe the animal in action to infer what its state was. Brown (1988) suggested the Optimal
186 patch use theory which deduces that the foraging activity of an animal within a resource patch is
187 dictated by the metabolic costs involved in foraging in that patch, the predation risk within that
188 patch and a missed opportunity cost (the value of all other patches in the environment in
189 comparison to the patch the forager is in). From this theory a tool for measuring the perceived
190 risk in the environment was developed, the giving up density (GUD). This tool has been used in
191 hundreds of research projects (Bedoya-Perez et al., 2013), and is the dominant tool that has been
192 used to quantify the LOF (Figure 3A) (Jacob & Brown 2000; Altendorf *et al.* 2001; Van Der
193 Merwe & Brown 2008; Druce *et al.* 2009; Abu Baker & Brown 2010; Baker & Brown 2011;
194 Emerson *et al.* 2011, and more). When the perceived energetic gain is diminished in comparison
195 to the costs (mentioned above), the animal ceases foraging in a patch and the density of food left
196 behind (GUD) becomes a quantifiable measure of these costs. The GUD as a measure for the
197 LOF has proved itself useful for study in small mammals (Appendix I). Examples of studies
198 include the cape ground squirrels mentioned above (Van Der Merwe & Brown, 2008), two
199 competing lemming species in the Canadian tundra (Dupuch et al., 2014), striped mice and
200 elephant voles in African grasslands (Abu Baker & Brown, 2010; Baker & Brown, 2011) and
201 small carnivorous marsupials in the Australian Desert (Bleicher and Dickman, in preparation).
202 GUDs have successfully been applied to measuring risk perception in birds (*e.g.* Brown *et al.*

203 1997; Kotler *et al.* 1998; Oyugi *et al.* 2012) and it is a matter of time before LOF are measured
204 using this tool in fish as well.

205 Foraging behavior is best used to measure the response of populations of small animals,
206 whereas in populations of species of greater biomass vigilance is best measured instead. The
207 theory of optimal vigilance states that an animal will spend a greater amount of time (and
208 energy) in vigilance behavior when the perceived risk of predation is higher in the habitat
209 occupied by that prey individual (Brown, 1999). This theory has been tested greatly in ungulates
210 (Ale & Brown, 2009; Tadesse & Kotler, 2011) small mammals (Rosenzweig, Abramsky &
211 Subach, 1997; Kotler *et al.*, 2010) and birds (Elgar, 1989; Robinette & Ha, 2001). Vigilance,
212 measured in time, as a direct tool to measure effects of risk in the environment has been applied
213 in roe deer responding to hunting pressures (Benhaiem *et al.*, 2008), duikers responding to
214 habitat heterogeneity (Abu Baker & Brown, 2013), mule deer responding to vegetation density
215 of fir trees compared with juniper bushes (Altendorf *et al.*, 2001)¹ and more.

216 C. Applications

217 The use of LOF follows general trends in ecological research. Ecologists traditionally have
218 an attraction to large megafauna, and these trends carry to the LOF theory as well (Figure 3B).
219 The major group for which the LOF has been applied remains with the initial group for which it
220 was designed, the ungulates (Laundré *et al.*, 2001; Laundré *et al.*, 2010). Similarly, the research
221 has been developed on model organisms for behavioural ecology, species that are readily
222 available and of no particular conservation status, predominantly gerbils, squirrels, heteromyid
223 rodents, voles and lemmings (*e.g.* Brown & Kotler 2004; Bleicher 2014; Eccard & Liesenjohann
224 2014).

¹ Mentioned as surveying behavior and not vigilance

225 Similar to the above, the focal study systems have been drawn to model systems (Figure 3C).
226 For ungulates in North America the predominant systems studied were alpine scrublands and
227 forests, the system in which the theory was developed (*e.g.* Altendorf *et al.* 2001; Laundré *et al.*
228 2001; Ripple & Beschta 2004; Creel *et al.* 2005; Hernández & Laundré 2005). In the small
229 mammals, arid (and tundra) zones (both arid grasslands and hyper-arid environments) have been
230 used largely because of their simplicity (*e.g.* Bleicher *et al.* 2012; Bleicher 2014; Dupuch *et al.*
231 2014; Eccard & Liesenjohann 2014). The use of the theory has been limited to simple systems
232 because of the ease of mapping a LOF on a two dimensional scale with the third being the
233 perception of risk. The one exception to this pattern is the study of habitat selection by primates
234 in which the third dimension of altitude (within trees) was added (Willems & Hill, 2009;
235 Emerson, Brown & Linden, 2011; Coleman & Hill, 2014).

236 On a side note, two trajectories can be expected in the way in which LOFs are being mapped.
237 Initially, three dimensional maps (with the three dimensions of space; longitude X latitude X
238 altitude) are likely to become more popular as landscapes of fear are applied to aquatic systems,
239 alpine ranges, tropical forests and avian species where the populations are not limited in two
240 dimensional vector movement. Similarly, a temporal dimension in the form of time series should
241 gain popularity to study the temporal changes observed in behavior influenced by seasonal,
242 competitive and predation pressures.

243 How have the LOF been applied to research? Given the history of the theory in conservation,
244 the LOF continue to have a significant proportion of manuscripts dedicated to this topic (Figure
245 2). The surprising observation is the small proportion of the research using the LOF that remains
246 in that field. In the conservation literature that used LOF analysis we find a pattern and a trend
247 for future applications. The majority of the papers were using the LOF to monitor the changes in

248 community habitat use and the resulting vegetation changes that occur in the landscape. One
249 example includes the success of willow regeneration in Yellowstone National Park (Ripple &
250 Beschta, 2004a,b, 2006). Other examples study the impact on health of the prey populations
251 through diet quality (Hernández & Laundré, 2005). I believe the LOF, as a population attribute
252 and assessment tool, should be used to a much greater extent in conservation. I found that
253 dasyurid marsupials, during a period of population bust, used artificial shelters only when these
254 were adjacent to natural refuges (Bleicher & Dickman, 2016). The ability to map out the habitat
255 use of species of conservation interest, provides an ability to assess the effectiveness of
256 conservation treatments aimed to increase habitat use by these species. This makes the LOF a
257 powerful conservation tool, which has potential to revolutionize the way in which species of
258 conservation concern are surveyed.

259 Despite the LOF being based in conservation, the majority of applications of the LOF (33
260 manuscripts) appear to be directed towards understanding habitat use and the influence that these
261 have on population and community dynamics (31 of those manuscripts). How are these
262 applications used to interpret population and community dynamics? For example, two competing
263 lemming species showed that the competition for resources in shared habitats had a greater
264 impact on their foraging than did the relative risk of predation (Dupuch et al., 2014). The
265 changes in the LOF of competing rodents from two systems (gerbils and heteromyid) showed
266 how predation risk, and the resulting habitat use, provide a mechanism of coexistence between
267 competing species (Bleicher, 2014). From another example (of many more), predation risk in
268 forest fragments in Australia limited the nest site selection in birds (Zanette & Jenkins, 2000).

269 The last application, not well developed at this stage, is the transcendence of environmental
270 risk into a measurable impact on neurologically manifested stress syndromes. Ecologically, when

271 an animal is under stress (risk of predation), the neurological registering of the risk cues causes
272 an increase in stress hormones to be released in the body of the animal (Gross & Canteras,
273 2012). The physiological responses to these stress hormones are energetically costly (Apfelbach
274 et al., 2005) and influence a lowered productivity (*e.g.* Mukherjee *et al.* 2014). An example of
275 the research in the neuro-ecology field showed that sparrows respond with an increase of a
276 variety of stress hormones (plasma total corticosterone, corticosteroid binding globulin (CBG)
277 and free corticosterone) in response to an increase in the risk of predation in the environment
278 (Zanette et al., 2011). In their work, Clinchy *et al.* (2013) suggest that this connection of
279 environmental stress and neurological responses are a fertile ground for research, moving away
280 from the chronic stress studied in laboratory animals.

281 **No longer a theory, now a measurable attribute (Prospectus)**

282 The LOF as a tool, can be applied to understanding the dynamics of change in habitat
283 selection for populations, and how these can be related to ecosystem functions and community
284 structure. The LOF also provides a means of comparison between species on the same physical
285 landscape (or within similar constraints). This tool can, in the future, be used in three major
286 ways.

287 (1) The LOF can reveal spatial intricacies that are driven by the evolutionary games
288 between competitors and between predator and prey species. As a result, I believe that we should
289 study the mechanisms by which both predator and prey interact using a spatial analysis that can
290 highlight the intricacies of the decision-making process (in both players). (2) I would like to
291 encourage the data mining of previous experiments that measured behaviour in vigilance and
292 GUDs. In experiments that used grid formations, a common normalization of the data is usually
293 performed, averaging the GUDs along treatments. These data sets could be reused to explore the

294 spatial distribution of habitat use. I expect many of these old experiments will confirm their
295 findings using the LOF approach; however, many novel patterns are likely to emerge and further
296 our understanding of behavioural patterns on a spatial scale. (3) I expect (stated above) the
297 infiltration of this tool as a measure of conservation success. (4) Last, some experiments have
298 begun to mend the gap between ecological and neurological pathways to the study of fear on the
299 spatial scale (Clinchy et al., 2011). Directly measuring the neurological impacts of risk in the
300 environment is still in its infancy state Clinchy et al (2011, 2013).

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305 **References:**

- 306 Abramsky Z., Strauss E., Subach A., Riechman A., Kotler BP. 1996. The effect of barn owls
307 (Tyto alba) on the activity and microhabitat selection of Gerbillus allenbyi and G.
308 pyramidum. *Oecologia* 105:313–319. DOI: 10.1007/BF00328733.
- 309 Abu Baker MA., Brown JS. 2010. Islands of fear: effects of wooded patches on habitat
310 suitability of the striped mouse in a South African grassland. *Functional Ecology* 24:1313–
311 1322. DOI: 10.1111/j.1365-2435.2010.01757.x.
- 312 Abu Baker M., Brown JS. 2013. Foraging and habitat use of common duikers , Sylvicapra
313 grimmia , in a heterogeneous environment within the Soutpansberg, South Africa. *African*
314 *Journal of Ecology* 52:1–10.
- 315 Ale SB., Brown JS. 2009. Prey Behavior Leads to Predator: A Case Study of the Himalayan Tahr

- 316 and the Snow Leopard in Sagarmatha (Mt. Everest) National Park, Nepal. *Israel Journal of*
317 *Ecology and Evolution* 55:315–327. DOI: 10.1560/IJEE.55.4.315.
- 318 Altendorf KB., Laundré JW., Lopez Gonzalez CA., Brown JS. 2001. Assessing effects of
319 predation risk on foraging behavior of mule deer. *Journal of Mammalogy* 82:430–439.
- 320 Apfelbach R., Blanchard CD., Blanchard RJ., Hayes RA., McGregor IS. 2005. The effects of
321 predator odors in mammalian prey species: a review of field and laboratory studies.
322 *Neuroscience and biobehavioral reviews* 29:1123–44. DOI:
323 10.1016/j.neubiorev.2005.05.005.
- 324 Baker MAA., Brown JS. 2011. Patch use behaviour of *Elephantulus myurus* and *Micaelamys*
325 *namaquensis* : the role of diet , foraging substrates and escape substrates. *African Journal of*
326 *Ecology* 50:167–175.
- 327 Bedoya-Perez MA., Carthey AJR., Mella VSA., McArthur C., Banks PB. 2013. A practical
328 guide to avoid giving up on giving-up densities. *Behavioral Ecology and Sociobiology*
329 67:1541–1553. DOI: 10.1007/s00265-013-1609-3.
- 330 Benhaiem S., Delon M., Lourtet B., Cargnelutti B., Aulagnier S., Hewison AM., Morellet N.,
331 Verheyden H. 2008. Hunting increases vigilance levels in roe deer and modifies feeding site
332 selection. *Animal Behaviour* 76:611–618. DOI: 10.1016/j.anbehav.2008.03.012.
- 333 Bleicher SS. 2014. Divergent behaviour amid convergent evolution: common garden
334 experiments with desert rodents and vipers. University of Illinois at Chicago.
- 335 Bleicher SS., Dickman CR. 2016. Artificial shelters alter dunnart’s landscape of fear. In:
336 Ventura, CA, USA: Gordon Research Conference - Predator-Prey Interactions, 1.

- 337 Bleicher SS., Kotler BP., Brown JS. 2012. Comparing landscapes of fear in four desert rodents
338 exposed to vipers and owls. :Poster.
- 339 Bouskila A. 1995. Interactions between predation risk and competition : a field study of
340 kangaroo rats and snakes. *Ecology* 76:165–178.
- 341 Brown JS. 1988. Patch use as an indicator of habitat preference, predation risk, and competition.
342 *Behavioral Ecology and Sociobiology* 22:37–47. DOI: 10.1007/BF00395696.
- 343 Brown JS. 1999. Vigilance , patch use and habitat selection : Foraging under predation risk.
344 *Evolutionary Ecology Research* 1:49–71.
- 345 Brown JS. 2010. Ecology of fear. In: Breed MD, Moore J eds. *Encyclopedia of Animal*
346 *Behaviour*. Oxford: Elsevier Ltd, 581–587.
- 347 Brown JS., Kotler BP. 2004. Hazardous duty pay and the foraging cost of predation. *Ecology*
348 *Letters* 7:999–1014. DOI: 10.1111/j.1461-0248.2004.00661.x.
- 349 Brown JS., Kotler BP., Mitchell WA. 1997. Competition between birds and mammals : A
350 comparison of giving-up densities between crested larks and gerbils. *Evolutionary Ecology*
351 11:757–771. DOI: 10.1023/A:1018442503955.
- 352 Brown JS., Vincent TL. 1992. Organization of predator-prey communities as an evolutionary
353 game. *Evolution* 46:1269–1283.
- 354 Clinchy M., Schulkin J., Zanette LY., Sheriff MJ., McGowan PO., Boonstra R. 2011. The
355 Neurological Ecology of Fear: Insights Neuroscientists and Ecologists Have to Offer one
356 Another. *Frontiers in behavioral neuroscience* 5:1–6. DOI: 10.3389/fnbeh.2011.00021.
- 357 Clinchy M., Sheriff MJ., Zanette LY. 2013. Predator-induced stress and the ecology of fear.

- 358 *Functional Ecology* 27:56–65. DOI: 10.1111/1365-2435.12007.
- 359 Coleman BT., Hill RA. 2014. Living in a landscape of fear: the impact of predation, resource
360 availability and habitat structure on primate range use. *Animal Behaviour* 88:165–173. DOI:
361 10.1016/j.anbehav.2013.11.027.
- 362 Creel S., Winnie JJ., Maxwell B., Hamlin K., Creel M. 2005. Elk alter habitat selection as an
363 antipredator response to wolves. *Ecology* 86:3387–3397.
- 364 Davidson DW., Brown JH., Inouye RS. 1980. Competition and the structure of granivore
365 communities. *BioScience* 30:233–238.
- 366 Dickman CR., Predavec M., Downey FJ. 1995. Long-range movements of small mammals in
367 arid Australia: implications for land management. *Journal of Arid Environments* 31:441–
368 452. DOI: 10.1016/S0140-1963(05)80127-2.
- 369 Druce DJ., Brown JS., Kerley GIH., Kotler BP., Mackey RL., Slotow R. 2009. Spatial and
370 temporal scaling in habitat utilization by klipspringers (*Oreotragus oreotragus*) determined
371 using giving-up densities. *Austral Ecology* 34:577–587. DOI: 10.1111/j.1442-
372 9993.2009.01963.x.
- 373 Dupuch A., Morris DW., Ale SB., Wilson DJ., Moore DE. 2014. Landscapes of fear or
374 competition? Predation did not alter habitat choice by Arctic rodents. *Oecologia* 174:403–
375 412. DOI: 10.1007/s00442-013-2792-7.
- 376 Eccard J a., Liesenjohann T. 2014. The importance of predation risk and missed opportunity
377 costs for context-dependent foraging patterns. *PLoS ONE* 9:1–5. DOI:
378 10.1371/journal.pone.0094107.

- 379 Elgar M. 1989. Predator vigilance and group size in mammals and birds: a critical review of the
380 empirical evidence. *Biological reviews of the Cambridge Philosophical Society* 64:13–33.
- 381 Embar K., Mukherjee S., Kotler BP. 2014. What do predators really want? The role of gerbil
382 energetic state in determining prey choice by Barn Owls. *Ecology* 95:280–285.
- 383 Embar K., Raveh A., Burns D., Kotler BP. 2014. To dare or not to dare? Risk management by
384 owls in a predator-prey foraging game. *Oecologia* 175. DOI: 10.1007/s00442-014-2956-0.
- 385 Emerson SE., Brown JS., Linden JD. 2011. Identifying Sykes' monkeys', *Cercopithecus*
386 *albogularis erythrarchus*, axes of fear through patch use. *Animal Behaviour* 81:455–462.
387 DOI: 10.1016/j.anbehav.2010.11.018.
- 388 Fischer J., Lindenmayer DB. 2007. Landscape modification and habitat fragmentation : a
389 synthesis. *Global Ecology and Biogeography* 16:265–280. DOI: 10.1111/j.1466-
390 8238.2006.00287.x.
- 391 Fox JF., Bryant JP. 1984. Instability of the snowshoe hare and woody plant interaction.
392 *Oecologia* 63:128–135. DOI: 10.1007/BF00379794.
- 393 Gross CT., Canteras NS. 2012. The many paths to fear. *Nature reviews. Neuroscience* 13:651–8.
394 DOI: 10.1038/nrn3301.
- 395 Hairston NG., Smith FE., Slobodkin LB. 1960. Community structure , population control, and
396 competition. *American Naturalist* 94:421–425. DOI: 10.1086/521238.
- 397 Haythornthwaite AS. 2005. Microhabitat use and foraging behaviour of *Sminthopsis youngsoni*
398 (Marsupialia : Dasyuridae) in arid central Australia. *Wildlife Research* 32:609–615. DOI:
399 10.1071/WR04126.

- 400 Haythornthwaite AS., Dickman CR. 2006. Distribution, abundance, and individual strategies: A
401 multi-scale analysis of dasyurid marsupials in arid central Australia. *Ecography* 29:285–
402 300. DOI: 10.1111/j.2006.0906-7590.04307.x.
- 403 Hernández L., Laundré JW. 2005. Foraging in the “landscape of fear” and its implications for
404 habitat use and diet quality of elk *Cervus elaphus* and bison *Bison bison*. *Wildlife Biology*
405 11:215–220. DOI: 10.2981/0909-6396(2005)11[215:FITLOF]2.0.CO;2.
- 406 Hewitt CG. 1921. *In The Conservation of the Wild Life of Canada*. New York, New York:
407 Charles Scribner’s Sons.
- 408 Jacob J., Brown JS. 2000. Microhabitat use, giving-up densities and temporal activity as short-
409 and long-term anti-predator behaviors in common voles. *Oikos* 91:131–138. DOI:
410 10.1034/j.1600-0706.2000.910112.x.
- 411 St. Juliana J., Kotler BP., Brown JS., Mukherjee S., Bouskila A. 2011. The foraging response of
412 gerbils to a gradient of owl numbers. *Evolutionary Ecology Research* 13:869–878.
- 413 Kotler BP., Blaustein L., Brown JS. 1992. Predator Facilitation: the combined effect of snakes
414 and owls on the foraging behavior of gerbils. *Annals of Zoology Fennici* 29:199–206.
- 415 Kotler BP., Brown JS., Dall SRX., Gresser S., Ganey D., Bouskila A. 2002. Foraging games
416 between gerbils and their predators : temporal dynamics of resource depletion and
417 apprehension in gerbils. *Evolutionary Ecology Research* 4:495–518.
- 418 Kotler BP., Brown JS., Mukherjee S., Berger-Tal O., Bouskila A. 2010. Moonlight avoidance in
419 gerbils reveals a sophisticated interplay among time allocation , vigilance and state-
420 dependent foraging. *Proceedings of the Royal Society B* 277:1469–1474. DOI:

- 421 10.1098/rspb.2009.2036.
- 422 Kotler B., Dickman CR., Brown JS. 1998. The effects of water on patch use by two Simpson
423 Desert granivores { *Corvus coronoides* and *Pseudomys hermannsburgensis*). *Australian*
424 *Journal of Ecology* 23:574–579.
- 425 Krebs CJ., Boutin S., Boonstra R., Sinclair a R., Smith JN., Dale MR., Martin K., Turkington R.
426 1995. Impact of food and predation on the snowshoe hare cycle. *Science* 269:1112–1115.
427 DOI: 10.1126/science.269.5227.1112.
- 428 Laundré JW., Hernández L., Altendorf KB. 2001. Wolves, elk, and bison: reestablishing the
429 “landscape of fear” in Yellowstone National Park, U.S.A. *Canadian Journal of Zoology*
430 79:1401–1409. DOI: 10.1139/cjz-79-8-1401.
- 431 Laundre JW., Hernandez L., Ripple WJ. 2010. The landscape of fear: Ecological implications of
432 being afraid. *The Open Ecology Journal* 3:1–7. DOI: 10.2174/1874213001003030001.
- 433 Leopold A. 1949. *A Sand County Alamanac and Sketches Here and There*. New York: Oxford
434 University Press.
- 435 Lima SL. 1998. Nonlethal effects in the ecology of predator-prey interactions. *BioScience* 48:25–
436 34. DOI: 10.2307/1313225.
- 437 Lima SL. 2002. Putting predators back into behavioral predator–prey interactions. *Trends in*
438 *Ecology & Evolution* 17:70–75. DOI: 10.1016/S0169-5347(01)02393-X.
- 439 Van Der Merwe M., Brown JS. 2008. Mapping the landscape of fear of the cape ground squirrel
440 (*Xerus inauris*). *Journal of Mammalogy* 89:1162–1169. DOI: 10.1644/08-MAMM-A-035.1.
- 441 Mukherjee S., Heithaus MR., Trexler JC., Ray-Mukherjee J., Vaudo J. 2014. Perceived risk of

- 442 predation affects reproductive life-history traits in *Gambusia holbrooki*, but not in
443 *Heterandria formosa*. *PLoS ONE* 9. DOI: 10.1371/journal.pone.0088832.
- 444 Oyugi JO., Brown JS., Whelan CJ. 2012. Foraging Behavior and Coexistence of Two Sunbird
445 Species in a Kenyan Woodland. *Biotropica* 44:262–269. DOI: 10.1111/j.1744-
446 7429.2011.00785.x.
- 447 Paine RT. 1963. Food web complexity and species diversity. *The American Naturalist* 100:65–
448 75. DOI: 10.2307/2678832.
- 449 Ripple WJ., Beschta RL. 2004a. Wolves and the Ecology of Fear: Can Predation Risk Structure
450 Ecosystems? *BioScience* 54:755–766. DOI: 10.1641/0006-
451 3568(2004)054[0755:WATEOF]2.0.CO;2.
- 452 Ripple WJ., Beschta RL. 2004b. Wolves , elk , willows , and trophic cascades in the upper
453 Gallatin Range of Southwestern Montana , USA. *Forest Ecology and Management*
454 200:161–181. DOI: 10.1016/j.foreco.2004.06.017.
- 455 Ripple WJ., Beschta RL. 2006. Linking wolves to willows via risk-sensitive foraging by
456 ungulates in the northern Yellowstone ecosystem. *Forest Ecology and Management*
457 230:96–106. DOI: 10.1016/j.foreco.2006.04.023.
- 458 Ritchie EG., Johnson CN. 2009. Predator interactions , mesopredator release and biodiversity
459 conservation. *Ecology Letters* 12:982–998. DOI: 10.1111/j.1461-0248.2009.01347.x.
- 460 Robinette RL., Ha JC. 2001. Social and ecological factors influencing vigilance by northwestern
461 crows, *Corvus caurinus*. *Animal Behaviour* 62:447–452. DOI: 10.1006/anbe.2001.1772.
- 462 Rosenzweig ML. 1973. Habitat selection experiments with a pair of coexisting Heteromyid

- 463 rodent species. *Ecology* 54:111–117.
- 464 Rosenzweig ML., Abramsky Z., Subach A. 1997. Safety in numbers: sophisticated vigilance by
465 Allenby's gerbil. *Proceedings of the National Academy of Sciences of the United States of*
466 *America* 94:5713–5715. DOI: 10.1073/pnas.94.11.5713.
- 467 Rösner S., Mussard-Forster E., Lorenc T., Müller J. 2014. Recreation shapes a “landscape of
468 fear” for a threatened forest bird species in Central Europe. *Landscape Ecology* 29:55–66.
469 DOI: 10.1007/s10980-013-9964-z.
- 470 Sandford R. 2013. Effects on Hunting on Roe Deer Habitat Use.
- 471 Schmitz OJ., Price JR. 2011. Convergence of trophic interaction strengths in grassland food
472 webs through metabolic scaling of herbivore biomass. *Journal of Animal Ecology* 80:1330–
473 1336. DOI: 10.1111/j.1365-2656.2011.01882.x.
- 474 Storch I., Voitke E., Krieger S. 2005. Landscape-scale edge effect in predation risk in forest-
475 farmland mosaics of central europe. *Landscape Ecology* 20:927–940. DOI:
476 10.1007/s10980-005-7005-2.
- 477 Tadesse S. 2012. Habitat quality and foraging ecology of Mountain Nyala (*Tragelaphus buxtoni*)
478 in the Munessa Forest and the Bale Mountains National Park, South-Eastern Ethiopia. Ben
479 Gurion University of the Negev.
- 480 Tadesse S., Kotler BP. 2011. Seasonal Habitat Use by Nuian Ibex (*Capra nubiana*) evaluated
481 with behavioral indicators. *Israel Journal of Ecology and Evolution* 57:223–246.
- 482 Vincent TL., Brown JS. 2005. *Evolutionary game theory , natural selection , and Darwinian*
483 *dynamics*. Cambridge University Press.

484 Willems EP., Hill RA. 2009. Predator-Specific Landscapes of Fear and Resource Distribution :
485 Effects on Spatial Range Use. *Ecology* 90:546–555.

486 Wolf N., Mangel M. 2007. Strategy , compromise , and cheating in predator – prey games.
487 *Evolutionary Ecology Research* 9:1293–1304.

488 Zanette L., Jenkins B. 2000. Nesting success and nest predators in forest fragments : A study
489 using real and artificial nests. *The Auk* 117:445–454.

490 Zanette LY., White AF., Allen MC., Clinchy M. 2011. Percived predation risk reduces the
491 number of offspring songbirds produce per year. *Science* 334:1398–1402.

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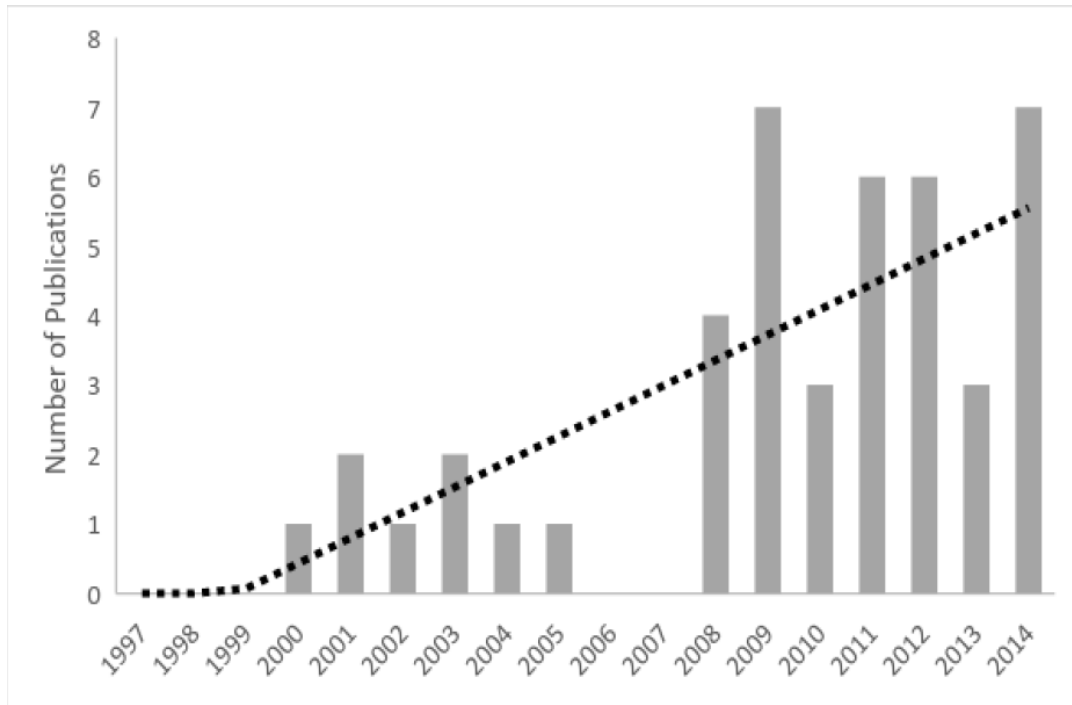
494 Figure Legends

495 **Figure 1.** Cumulative number of manuscripts publishing empirical analysis supporting the theory
496 of LOFs in ecological journals (total of 58). The application of the theory is increasing in
497 applications and use showing higher rate of publication with pulses of increasing magnitude
498 2009, 2012 and 2014 (15 manuscripts). The trend line is for visual emphasis of the increase in
499 use. At the time of writing 4 manuscripts were published in 2015.

500 **Figure 2.** Cumulative number of publications testing the theory of LOF using empirical data
501 organized through the focus of the manuscripts. * A manuscript may be represented in more than
502 one column if the manuscript discusses more than one aspect of the LOF theory.

503 **Figure 3.** Cumulative number of publications testing the theory of LOF using empirical data
504 organized through: (A)* per species group the methods used to quantify fear, (B) the subject
505 species classification (of the author's categorical choice), and (C) the system in which the
506 experiment was performed. * A manuscript may be represented in more than one column if the
507 manuscript uses more than one measurement variable. Abbreviations: avi. = avian, veg.=
508 vegetation l. and s. (mammal) = large and small.

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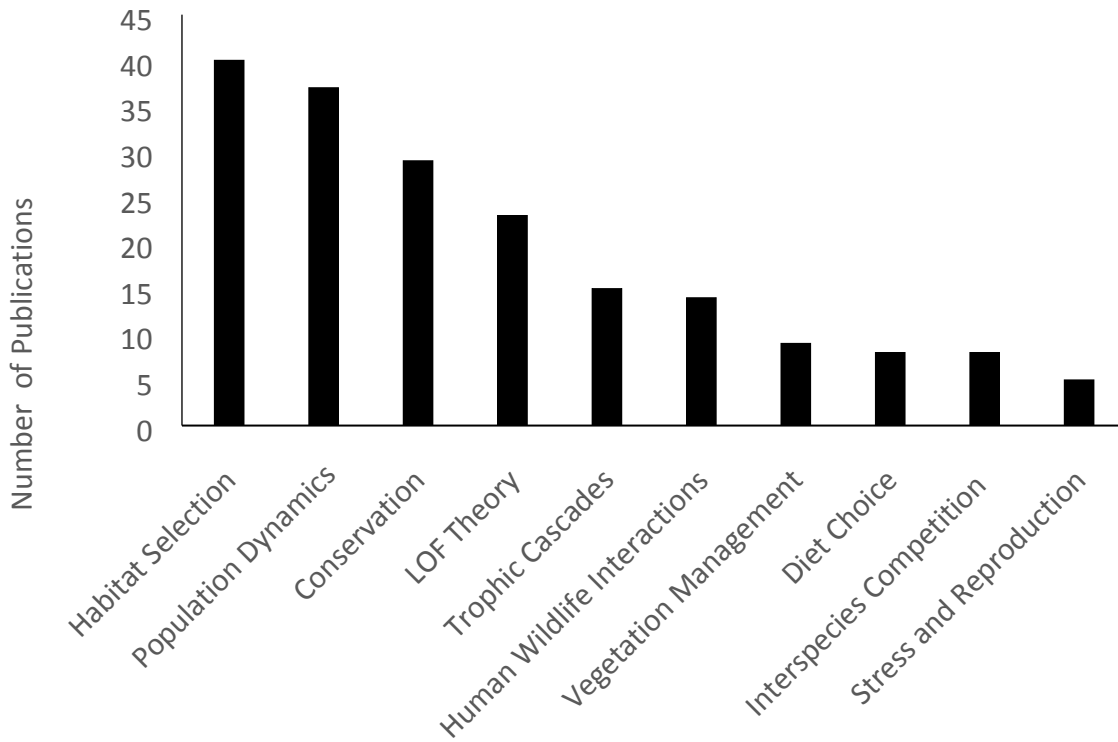


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511 **Figure 1.**

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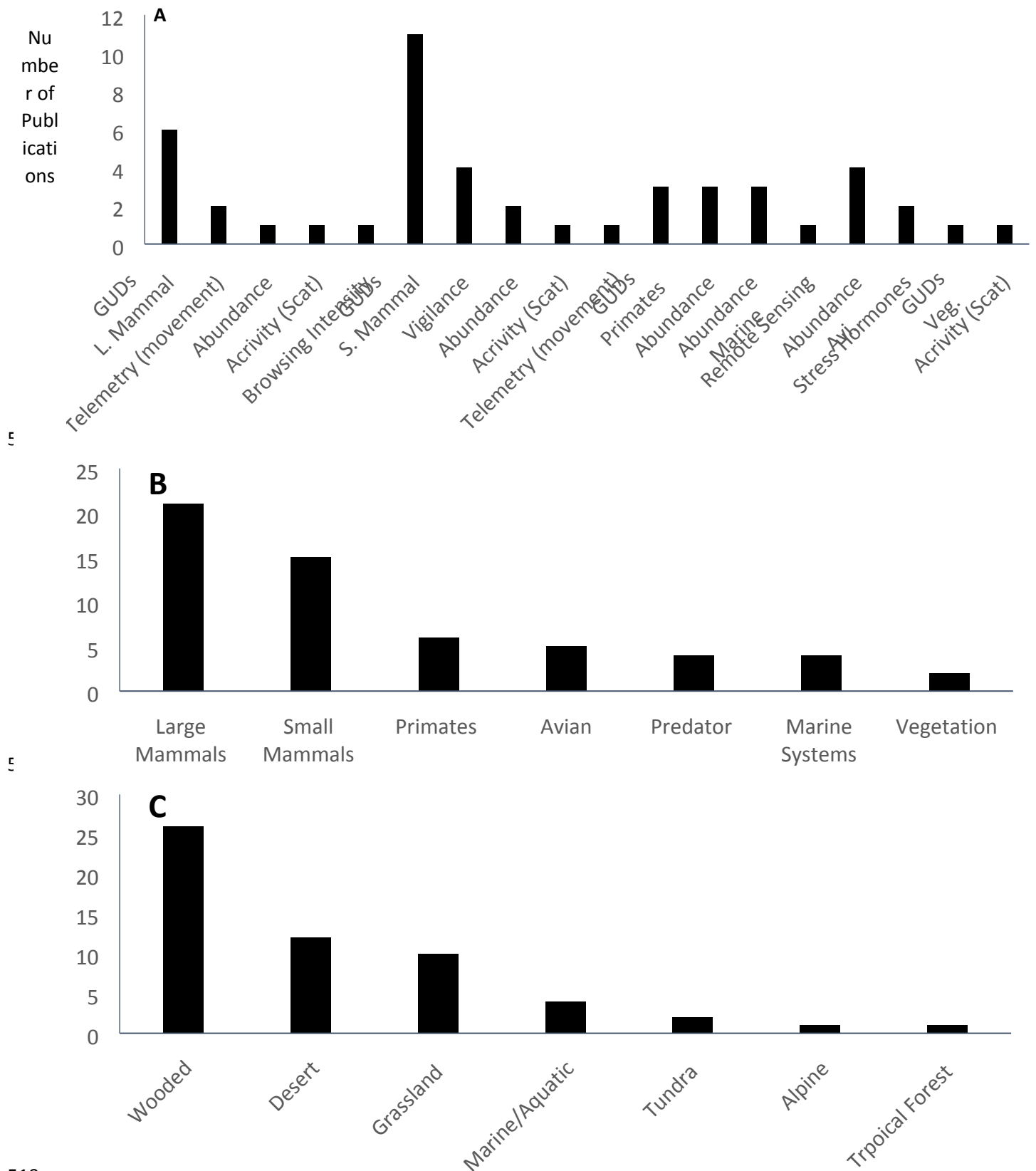
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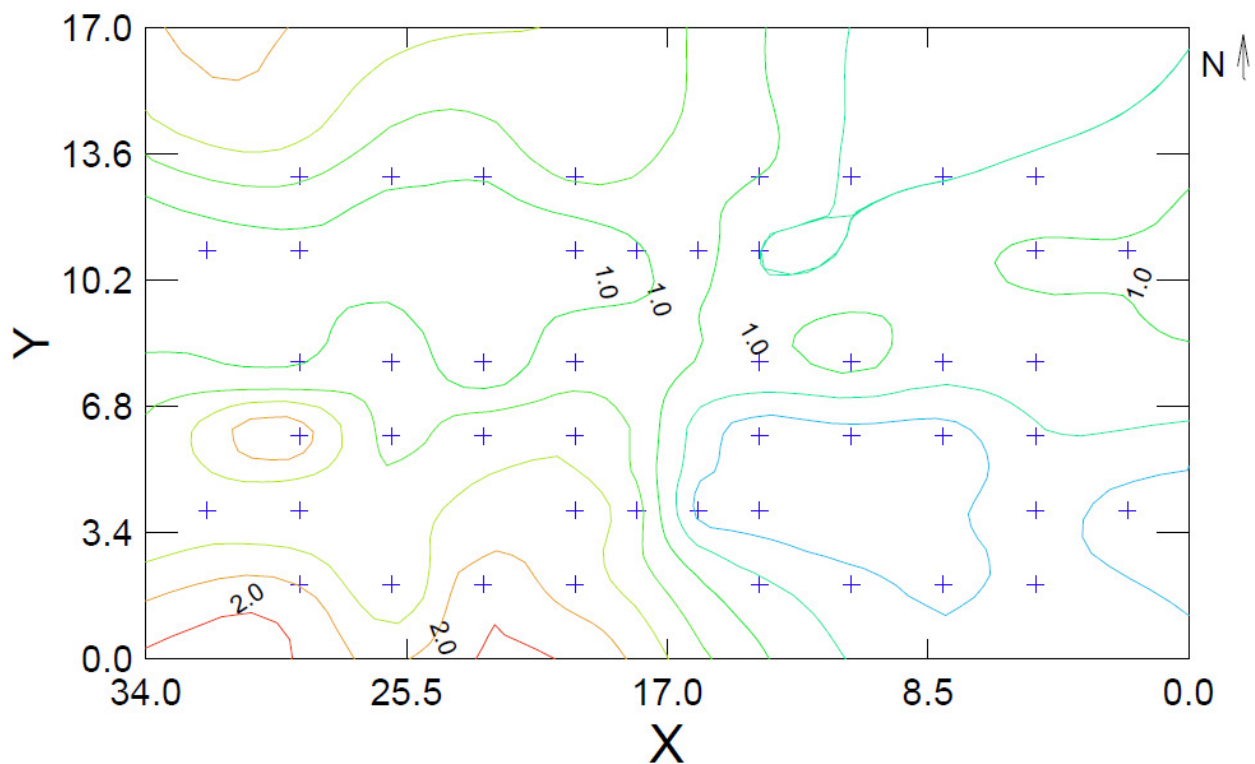
515 **Figure 2.**

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519
520 **Figure 3.**

521 Appendix I. Example of Landscape of Fear Map using a dataset adapted from Bleicher et al.
522 (2016). The map shows the distribution of risk using giving up densities (GUDs) for a population
523 of Allenby's gerbils (*G. andersoni allenbyi*) in a controlled enclosure in Sde Boker, Israel. The
524 contour lines are derived using the distance weighted least squares (DWLS) smoothing function
525 at a tension of 0.5. GUD values above 2.0g (orange and red) reflect areas that are perceived as
526 dangerous by the gerbils while areas below 1.0g (green and blue) reflect zones of safety. The +
527 signs are the locations in which the data was collected and both x and y -axes are measuring the
528 enclosure in meters.



529