

1 **How long should the fully hillside-closed forest protection be implemented on**
2 **the Loess Plateau, Shaanxi, China?**

3 Lin Hou¹, Sijia Hou²

4 1 College of Forestry, Northwest A&F University, Yangling, Shaanxi 712100, China

5 2 College of Transportation, Southeast University, Nanjing, Jiangsu, 211189, China

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18 **Corresponding author**

19 Lin Hou

20 3 Taicheng Road, Yangling, Shaanxi, 712100, China

21 E-mail: houlin1969@163.com

22 **Abstract**

23 **Background.** Restoration of degraded forest ecosystem is crucial for regional sustainable
24 development. To protect the country's fragile and fragmented environment, the Chinese
25 government has initiated an ecological engineering, the Natural Forest Protection Program in
26 seventeen provinces in China since 1998. Fully hillside-closed forest protection (vegetation
27 restoration naturally without any artificial disturbance) was one of vital measures of the Natural
28 Forest Protection Program applied national wide. Whether plant diversity, biomass and age
29 structure of dominant tree species and soil nutrients in protected stands may become better with
30 increase of protected period are still open problems.

31 **Methods.** We investigated community diversity, biomass of dominant tree species, age structures,
32 and analyzed soil chemical properties of a *Pinus tabulaeformis* population at protected sites
33 representing different protected ages at Huanglongshan Forest Bureau on the Loess Plateau, Shaanxi,
34 China.

35 **Results.** Plant species richness of *Pinus tabulaeformis* community was significantly affected
36 ($p < 0.05$) by forest protection and the effect attenuated with protection age. Shannon evenness index
37 of plant species generally increased with protection age. Stands protected for 45 years had the
38 highest tree biomass and considerable natural regeneration capacity. Contents of organic carbon,
39 available phosphorus and available potassium in top soil increased in protected stands less than 45
40 years, however decreased significantly thereafter. Long-term forest protection also decreased
41 content of mineral nitrogen in top soil.

42 **Discussion.** We found that richness of shrubs and herbs was significantly affected by forest

43 protection, and evenness indices of tree, shrub and herb increased inconsistently with protected
44 ages. Forest protection created more complex age structures and tree densities with increasing age
45 of protection. Content of soil mineral nitrogen at 0-20 cm soil depth showed a decreasing trend in
46 stands of up to 30 years. Soil available phosphorus and potassium contents were higher in stands
47 with greater proportions of big and middle trees. Long-term protection (> 45 years) of *Pinus*
48 *tabulaeformis* stand in southeast Loess Plateau, China, may be associated with decreasing plant
49 species richness, proportion of medium to large trees, dominant biomass of *Pinus tabulaeformis*
50 and soil nutrients.

51

52 **Keywords:** Fully Hillside-closed Forest Protection, Optimal Protection Age, *Pinus tabulaeformis*,
53 the Loess Plateau, Restoration

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64 INTRODUCTION

65 Ecological restoration is being recognized as an international priority(Aronson & Alexander
66 2013; Wortley et al. 2013) and it plays a crucial role in rebuilding ecological equilibrium and
67 reversing ecosystem degradation (Ma et al. 2013). As a part of ecological engineering (Mitsch
68 2012), the practice is being widely incorporated into natural resource strategies from the local to
69 global level (Wortley et al. 2013).

70 To protect the country's fragile and fragmented environment, the Chinese government has
71 initiated an ecological engineering, the Natural Forest Protection Program (NFPP) since 1998 (Xu
72 et al. 2006). Logging and harvesting of partial or full timber was prohibited in protected areas from
73 1998 to 2008(Xu et al. 2006). Fully hillside-closed forest protection (vegetation restoration
74 naturally without any artificial disturbance) was applied in national wide. Ecosystems have the
75 capacity to self-organize and the self-design or self-organizational properties of natural systems is
76 an essential component to ecological engineering(Bergen et al. 2001). It is obviously fully
77 hillside-closed forest protection is in accord with the ecological engineering principle self-design.

78 The previous studies regarding NFPP have mainly focused upon the introduction of the
79 related policy issues(Grumbine & Xu 2011; Li 2004; Wang et al. 2015) , the spatial-temporal
80 succession of regional vegetation (Huang et al. 2014) and ecological restoration programs and
81 payments (Yin & Zhao 2012).

82 However, a range of questions remain, particularly in relation to stand function and associated
83 environmental parameters following stand protection. We hypothesize that fully hillside-closed
84 forest protection may promote plant diversity, biomass and age structure of dominant tree species

85 and soil nutrients with increase of protected period.

86 The objectives of this study are to address a few of these key knowledge gaps, including : (i)
87 do the stands exhibit significant differences in plant assemblage; (ii) does soil fertility change with
88 stand age structure;(iii) can a functional relationship be defined regarding length of stand
89 protection and stand quality, i.e., are stands protected for longer timeframes “better” than other
90 stands; and (iv) based on findings of i-iii above, can a preliminary estimate regarding the optimal
91 time span for *Pinus tabulaeformis* stands be recommended to the Natural Forest Protection
92 Program?

93 MATERIALS AND METHODS

94 Site description

95 The study was conducted in Huanglong County (35°28'49"–36°02'01"N, 109°38'49"–
96 110°12'47"E) on the southeast Loess Plateau ,Shaanxi, China. Stands in this area (a part of NFPP
97 area) play key ecological roles in soil erosion abated and sand storm mitigated(Chen et al. 2014).
98 The vegetation type is a northern deciduous broad-leaved forest sub-region. *Pinus tabulaeformis* is
99 dominant tree species in the currently existing stands. The associated tree species are *Quercus*
100 *liaotungensis*, *Syringa oblate*, *Populus davidiana*, *Prunus davidiana*, *Betula platyphylla* and
101 *Toxicodendron vernicifluum*. Shrubs and herb species in understory are abundant. The altitude
102 ranges from 1100 to 1300 m. It is dominated by a warm temperate and semi-humid continental
103 climate. The annual average precipitation is 612 mm and the mean atmospheric temperature is
104 8.6 °C. Cinnamon soil is the main soil type in the forest region.

105 Due to poor communication and a small population in the past years, stands on some special

106 sites have not been disturbed since 1950, especially since 1998.

107 According to data from Huanglongshan Forest Bureau, Yanan, Shaanxi, China, stands with
108 protected age sequence were found in four forest farms (Table 1).

109 **Field methods**

110 The field investigation and sampling was conducted between June 5 to July 15, 2003. Each
111 plot of tree, shrub and herb was 20 m × 20 m, 2m × 2 m and 1 m × 1 m respectively. Five sub-plots
112 of shrub, herb and regeneration seedlings were arrayed diagonally in each tree plot respectively.
113 The indices, species, number, Height (H), diameter at breast height (DBH) and canopy density of
114 trees, and species, height, cover ratio, number of shrub, herb and regeneration seedlings were
115 measured. All community data were collected from 27 tree plots spreading among the age cohorts
116 and 270 sub-plots (Table 1).

117 Three soil samples were obtained randomly by a special drill in each tree plot. Surface soils
118 (0-20cm depth) at all sites were assessed for soil properties including organic carbon, mineral N,
119 available phosphorous and potassium.

120 **Community diversity**

121 The importance of species richness and evenness in influencing diversity-associated
122 productivity has been demonstrated in a meta-analysis of 54 studies (Zhang et al. 2012). In this
123 study, we chose indices of richness and evenness to reflect characteristics of community. Species
124 richness index (S) was derived from field survey data. To characterize the diversity of the stand
125 community, the Shannon-Wiener index (H') and evenness index (J') were calculated as the
126 following:

127 Shannon-Wiener index $H' = -\sum P_i \ln P_i$

128 Shannon evenness index $J' = \frac{H'}{\ln S}$

129 where P_i is the relative frequency of the i -th species, and S is total number of species in plots
130 and subplots (Magurran 2004).

131 **Biomass of dominant tree species**

132 Average DBH (cm) and height (m) of *Pinus tabulaeformis* in each plot were calculated and
133 living biomass (Mgha^{-1}) of whole trees (*Pinus tabulaeformis*) were estimated (Chen & Peng 1996;
134 Pan et al. 2004).

$$135 \quad Y = 15.525 + 0.6269v$$

$$136 \quad \ln v = 0.99138 \ln(D^2 H) - 10.30211$$

137 where Y is the living biomass of trees (Mgha^{-1}) v is the stand growing stock ($\text{m}^3 \text{ha}^{-1}$), D (cm) is
138 diameter at breast height and H (m) is height.

139 Combining the density of dominant tree species (Table 1) with equations, biomass of *Pinus*
140 *tabulaeformis* in protected stands was determined.

141 **Age structures**

142 DBH of tree species correlate significantly to their ages under the same environmental
143 condition (Parker & Peet 1984). Lacking of analytic wood data, we adopted DBH structures of
144 *Pinus tabulaeformis* population instead of its age structures. Combining DBH to H , age structures
145 of *Pinus tabulaeformis* population were classified as following: I seedling, $H \leq 0.30$ m; II young
146 tree, $0.30 \text{m} < H \leq 2.00 \text{m}$, $\text{DBH} \leq 6.00 \text{cm}$; III small tree, $H > 2.0 \text{m}$, $6.0 \text{cm} < \text{DBH} \leq 12.0 \text{cm}$; IV middle
147 tree, $12.0 \text{cm} < \text{DBH} \leq 20.0 \text{cm}$; V big tree, $\text{DBH} > 20.0 \text{cm}$. The ratio of seedlings, young trees, small

148 trees, middle trees and big trees in stand with same protecting age was used to illustrate age
149 structures. Probable age of individual was determined by their whorled branches.

150 **Chemical analyses**

151 Analyses were made on air-dry soil material that passed through a 2 mm sieve. Soil organic
152 carbon content (SOC) was determined by dry combustion with a TOC/TON analyzer
153 (TOC-VTH-2000A, Shimadzu Corporation, Japan). Soil mineral nitrogen (ammonium nitrogen,
154 $\text{NH}_4^+\text{-N}$ and nitric nitrogen, $\text{NO}_3^-\text{-N}$) content was determined by the colorimetric method with
155 automatic flow injection (AA3, BRAN+LUEBBE, Germany). Available phosphorus content was
156 extracted in 0.5M NaHCO_3 and determined by Mo-Sb colorimetry. Available potassium content
157 was determined by method of flame photometry (Bao 2000).

158 **Data processing and analysis**

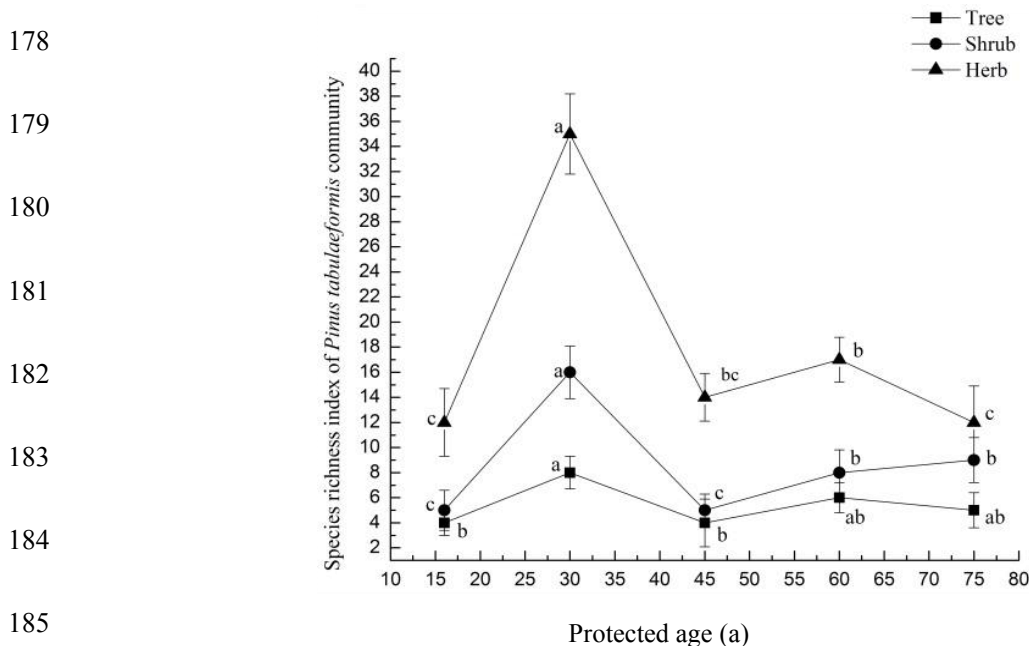
159 SPSS 17.0 and Origin8.0 (OriginLab Corporation) software were used for statistical analysis
160 and plotting. A graphic check of the postulates was performed based on the residual distribution.
161 One-way analysis of variance (ANOVA) following by Fisher's least significant difference (LSD)
162 test ($p < 0.05$) was used to compare the protection age effects on diversity of plant community and
163 soil nutrients respectively.

164 **RESULTS**

165 **Diversity of plants in protected stands**

166 Richness index of tree, shrub and herb was highest in the stand protected for 30 years (Table
167 2). A significant difference in the tree species richness index was observed in the 30 year protected
168 stand compared to stands protected for 16 years ($n=5$, $p < 0.05$) and 45 years ($n=8$, $p < 0.05$), but not

169 in other stands with different protected ages (Figure 1). The richness index of within stand shrubs
 170 differed significantly between stands protected for 30 years compared to stands protected for 16
 171 years ($n=25, p<0.05$), 45 years ($n=40, p<0.05$), 60 years ($n=20, p<0.05$) and 75 years ($n=20,$
 172 $p<0.05$) (Figure 1). Significant differences in the within-stand shrub richness index were also
 173 found in stands protected (i)16 years and 60, 75 years, (ii) 45 years and 60 years, 75years (Figure
 174 1). The within stand herb richness index in the stands protected for 30 years differentiated
 175 significantly to stands protected 16 years ($n=25, p<0.05$), 45 years ($n=40, p<0.05$), 60 years ($n=20,$
 176 $p<0.05$) and 75 years ($n=20, p<0.05$) (Figure 1). Richness of within stand herb at stands protected
 177 for 16 years also varied significantly to stands protected for 60 years and 75 years (Figure 1).

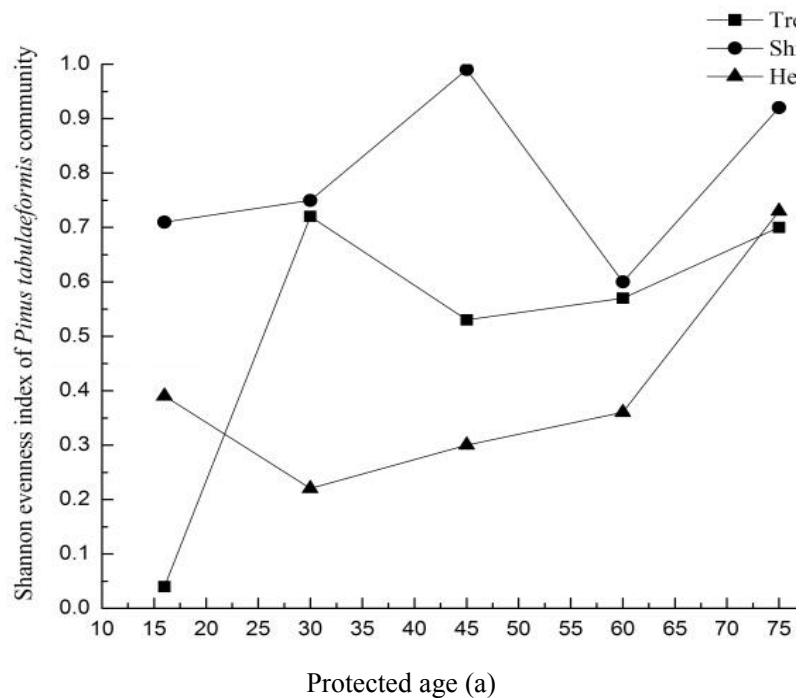


185
 186 **Figure 1 Richness of *Pinus tabulaeformis* community in protected stands.** The values in figures are the mean \pm SD.

187 Different letters in the same layer in the figure indicate significant differences between groups based on LSD ($p<0.05$).

188 Shannon-Wiener evenness index of tree, shrub and herb was the highest in stands protected
 189 for 30 years, 45 years and 75 years respectively (Figure 2). The index of herb generally increased

190 with protected ages except in stands protected for 16 years to 30 years (Figure 2). However, the
 191 index of tree and shrub fluctuated with stand protected years and did not follow a trending
 192 relationship (Figure 2). Tree and shrub Shannon-Wiener index increased with stand protection age,
 193 with the exception of tree index 30-45 year stand protection and shrub index 45-60 year stand
 194 protection (Figure 2).



203 **Figure 2 Shannon-Wiener evenness index of *Pinus tabulaeformis* community in protected stands**

204 **Biomass of *Pinus tabulaeformis* in protected stands**

205 Biomass of *Pinus tabulaeformis* increased in stands until 45 years of forest protection;
 206 however, for sites older than this protection age, stand biomass decreased (Figure 3). Peak biomass
 207 was $70.60 \pm 8.00 \text{ t} \cdot \text{ha}^{-1}$ in the stand protected for 45 years, while biomass in the stand protected for
 208 75 years ($19.90 \pm 9.20 \text{ t} \cdot \text{ha}^{-1}$) was lower than the stand protected for 16 years ($23.70 \pm 17.10 \text{ t} \cdot \text{ha}^{-1}$)
 209 (Figure 3).

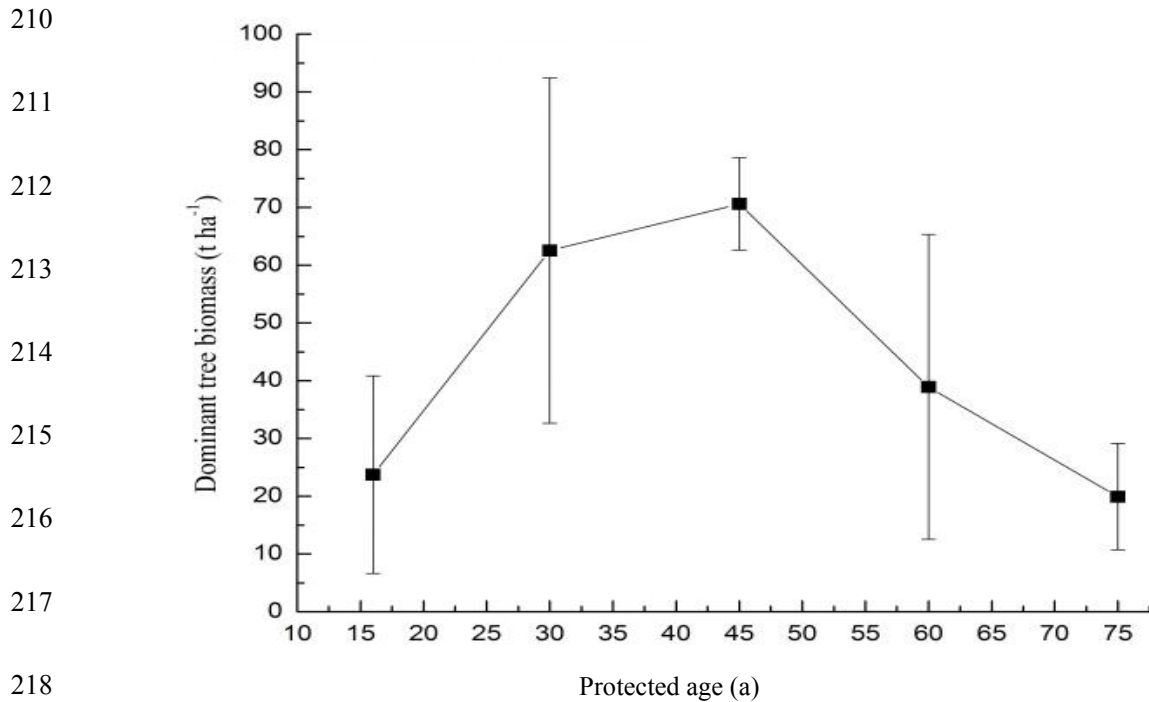


Figure 3 Biomass of *Pinus tabulaeformis* in protected stands

Age structure of *Pinus tabulaeformis* population in protected stands

Although age classes of *Pinus tabulaeformis* occurred in protected stands, they varied greatly (Figure 4). Only young (II) and small trees (III) were found in the stand protected for 16 years, small (III) and middle trees (IV) dominated the stand protected for 30 years (Figure 4). For the stand protected for 45 years, big (V) and middle trees (IV) were main components, but seedlings (I) and young trees were considerable also (Figure 4). In contrast, for stands protected for 60 and 75 years, seedlings (I) were the dominant component, followed by young (II) and small trees (III), with big trees (V) lowest in distribution (Figure 4).

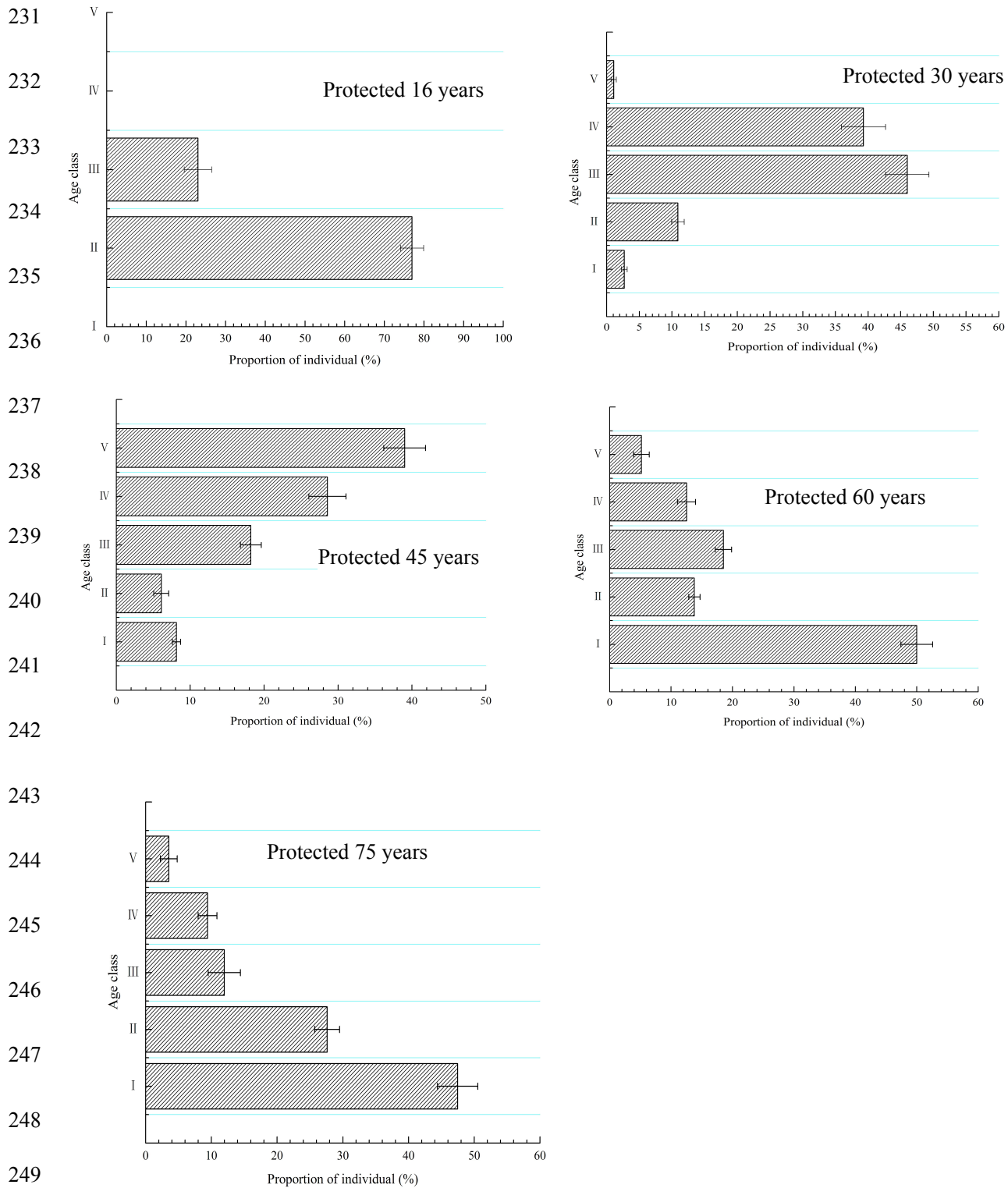


Figure 4 Age structure of *Pinus tabulaeformis* population in protected stands

252 **Soil nutrients**

253 Significant differences of soil organic carbon content at 0-20 cm soil depth were observed
 254 between the stands, with higher soil organic carbon content observed in stands protected for longer
 255 than 30 years (Figure 5 A). Content of mineral nitrogen at 0-20 cm soil depth demonstrated

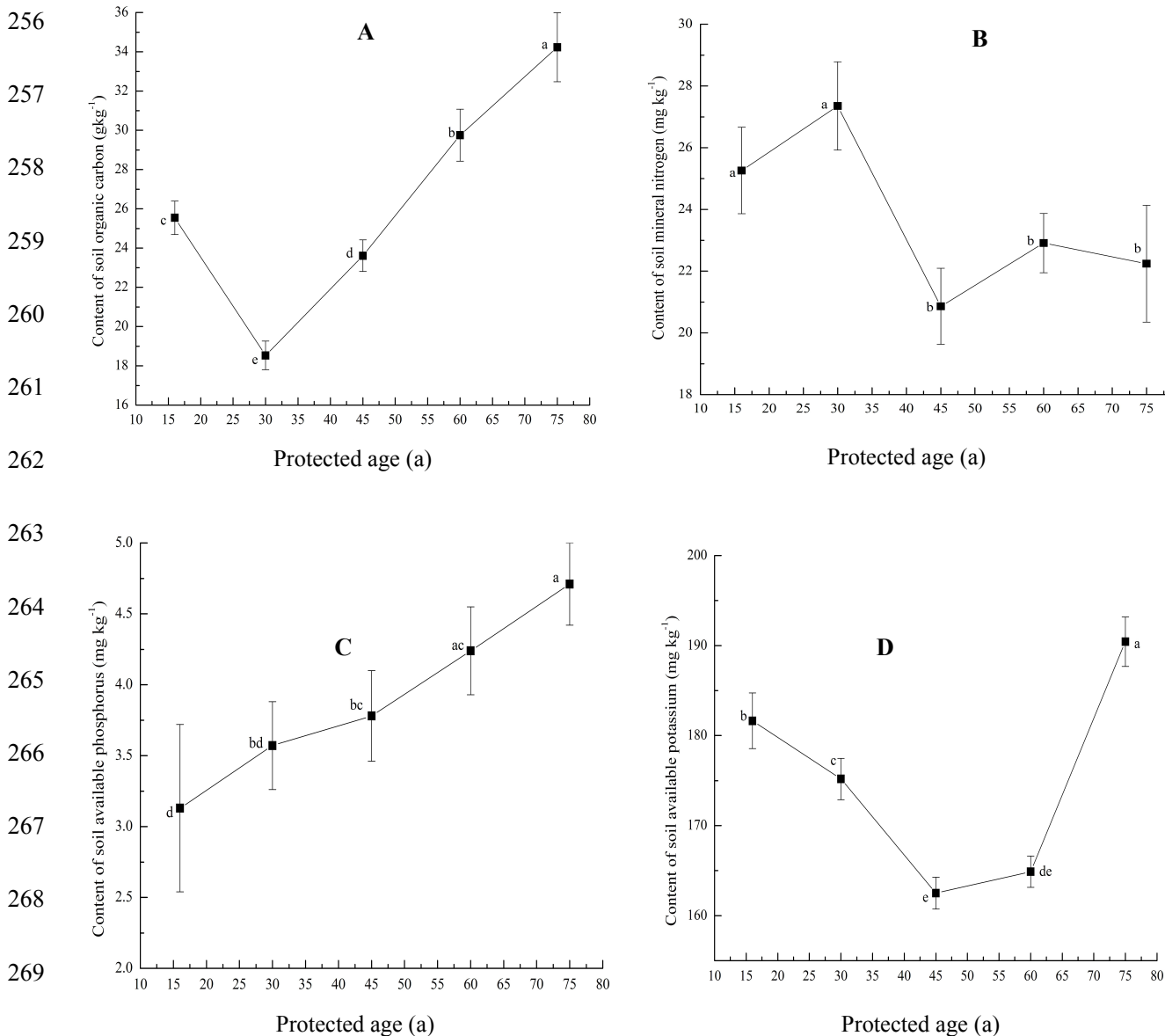


Figure 5 Soil nutrients in protected stands

271 significant differences in stands before and after protected 30 years (Figure 5 B). No
 272 significant differences were found between protected for 16 years and 30 years, and among stands

273 after protected for 30 years (Figure 5 B).Content of available phosphorus at 0-20 cm soil depth
274 increased as protection of stand age increased, with significant differences observed mostly at
275 youngest and oldest stand ages (Figure 5 C). No significant difference in available phosphorus was
276 observed in stands between 45 and 30 years protection age, and between 45 and 60 years
277 protection age (Figure 5 C).Content of available potassium at 0-20 cm soil depth decreased in
278 stands younger than 45 years forest protection and thereafter increased (Figure 5 D). Significant
279 differences were demonstrated among stands with different protection ages, except at 45 years and
280 60 years (Figure 5 D).

281 **Discussion**

282 **Response of plant diversity to forest protection**

283 Species richness is one measure of biodiversity and is very important for ecosystem
284 functioning, stability and integrity(Coroi et al. 2004). We found that richness of shrubs and herbs
285 was significantly affected by forest protection, although richness (Figure 1) and evenness (Figure 2)
286 indices of tree, shrub and herb increased inconsistently with protected age. The richness of plant
287 species increased in stands protected for 16 years to 30 years, decreased in stands protected for 30
288 years to 45 years and remained fairly stable in stands protected for longer than 45 years (Table 2).
289 Due to adequate sunlight and growing spaces, some pioneer tree species (*Populus davidiana*,
290 *Betula platyphylla*) and drought resistant shrubs (*Rubus corchorifolius*, *Rosa hugonis*,etc.) and
291 herbs (*Artemisia gmelinii*, *Saussurea petrovii*, etc.) were more prevalent in the younger forest
292 protection sites, increasing plant species richness of these stands (Table 2). With the growth of
293 trees, canopy density increased and some drought resistant plant species disappeared. The naturally

294 regenerating seedlings in stands protected for 45 years and older occupied a large proportion of the
295 species observed, impeding invasive plant species and stabilizing plant diversity of the community
296 assemblage. Inherent spatial variability within the landscape may provide a possible explanation
297 for this pattern observed, since *Pinus tabulaeformis* stands are distributed across variable site
298 conditions within the region. Soil moisture is considered to be the key limiting factor on the Loess
299 Plateau for differences in plant species growth and regeneration(Chen et al. 2014) and it is possible
300 that the differences in soil properties observed in this study reflect variation in plant-available
301 moisture.

302 Forest protection in Huanglongshan forest region, Yanan, Shaanxi, China was initiated in
303 1950. Stand structure within the protection area under the natural restoration condition differed.
304 Stands with age diversity generally were more species rich than stands with less age structure
305 (Thompson 2012). Findings in this study partly support this notion. Stands protected for great than
306 16 years had more diverse age structures and plant species richness (Figure 1). Age class structure
307 in stands protected for 30 years were generally simpler than stands protected for longer periods
308 (Figure 4). However, stands older than 30 years forest protection had lower richness index of tress
309 and understory species (Figure 2).

310 Our results suggest that sustainable forest protection can potentially contribute to plant
311 diversity conservation by increasing species richness generally (Table 2) and promoting
312 distribution of trees and herbs more even (Figure 2).

313 **Response of age structure to forest protection**

314 Forest protection created more complex age structures (Figure 3) and tree densities with

315 increasing age of protection (Table1). Seedlings, middle and big trees were absent in younger
316 stands (Figure 4) which indicated tree biomass was low (Figure 3) and lacked natural regeneration
317 capacity. Although plants species were most abundant in stands protected for 30 years (Figure 1),
318 this protection age contained the lowest proportion of big trees (V) among age classes (Figure 4)
319 limiting tree biomass. Both seedlings (Figure 4) and density of trees (Figure 1) in older (> 60 years)
320 stands protection were higher than younger stands, suggesting that a better natural regeneration
321 capacity. However, more seedlings and small trees without adequate big trees (Figure 4) in the
322 older stands also unveiled their insufficient productivity (Figure 3).

323 Our results support the widely accepted view that the rate of stand biomass accumulation
324 peaks in the early stage of development, usually at the time of canopy closure, and declines
325 thereafter (Acker et al. 2002; McMahon & Schlesinger 2010; Sarah Lesley Taylor 2005; Xu et al.
326 2012). The stand protected for 45 years had not only the highest canopy density (Table 1), but also
327 the highest proportion of big trees and tree biomass and considerable seedling density (Figure 4),
328 suggesting adequate regeneration capacity at this age.

329 **Response of soil nutrients to forest protection**

330 Vegetation plays a key role in maintaining the soils in which they grow (Mishra et al. 2003),
331 directly influencing soil nutrients accumulation and consequently soil development by above
332 ground inputs (Blazewski et al. 2009; Drouin et al. 2011; Giese et al. 2000). Litter fall and its
333 decomposition is an important mechanism governing soil chemical properties (Mishra et al. 2003),
334 especially the upper soil layer (Ma et al. 2007).

335 In the present study, *Pinus tabulaeformis* tree growth (Figure 3) and understory plant species

336 richness increased quickly for stands protected less than 30 years (Figure 1), however litterfall
337 input to soil was lower due to the absence of big trees in these stands (Figure 4). Tree and canopy
338 density (Table 1) decreased in stands protected for more than 30 years , with highest values
339 observed in stands protected for 45 years (Figure 3). Increased litterfall input, decomposition rate
340 and higher soil organic carbon contents were also observed at older forest sites (Figure 5A).

341 Content of soil mineral nitrogen at 0-20 cm soil depth showed an decreasing trend in stands of
342 up to 30 years protection although no significant differences were found among stands (Figure 5B).
343 This trend supports previous studies which have observed that young or developing stands
344 accumulate forest floor nitrogen, tending towards relatively stable conditions in undisturbed
345 mature forests (Johnson & Turner 2014; Miller 1981; Turner 1981).

346 The primary source of phosphorus and potassium in terrestrial ecosystems derives from
347 mineral materials in weathering parent rock (Filippelli 2008; Sheng 2005; Smeck 1985; Tiessen et
348 al. 1984). A proportion of there leased phosphorus and potassium, available in exchangeable and
349 soluble (available) fractions, can be assimilated by plants and soil microorganisms directly
350 (Schachtman et al. 1998; Sheng 2005). Soil phosphorus availability is also enhanced through
351 phosphorus solubilizing and mineralizing microbial biomass (Richardson & Simpson 2011). Many
352 soil microorganisms excrete organic acids to directly dissolve rock potassium to bring the
353 potassium into solution (Bennett 1998; Friedrich et al. 1991; Groudev 2010; Ullman et al. 1996).

354 In the present study, soil available phosphorus (Figure 5C) and potassium (Figure 5D)
355 contents were higher in stands with greater proportions of big and middle trees. We suggest that the
356 stands with greater biomass accumulated more litter and humic mineral in the top soil, which

357 provided a substantial energy source and favorable conditions for microbial activity (Fontaine et al.
358 2003). In younger stands, more nutrients may be taken up by the vegetation during intense tree
359 growth phase than can be replaced within the soil from mineral weathering and litter
360 decomposition (Brais et al. 1995) which may explain why soil available potassium decreased in
361 stands of up to 45 years protection in the present study (Figure 5D).

362 **The optimal age for the fully hillside-closed forest protection**

363 No restoration project is undertaken in a social vacuum (Knight et al. 2010). The goods and
364 services provided by forests are an important source of income for local people in the rural part of
365 China (Ma et al. 2013). Even when the intentions of ecological restoration are good and the
366 restoration strategy suitable for the environmental conditions (Ma et al. 2013), restoration action
367 will not be sustainable if it does not take into account the profit potential of local people

368 Our results showed that long-term protection (> 45 years) of *Pinus tabulaeformis* stand in
369 southeast Loess Plateau, China, may be associated with decreasing plant species richness (Table 2),
370 proportion of medium to large trees (Figure 4) , dominant tree biomass(Figure 3) and soil nutrients
371 (Figure 5). We suggest that it is possible, based on the findings above, to couple forest
372 management policy without exacerbating the poverty of local people, through the promotion of
373 measured forest indices as evidence-based support for forest protection and use. For this region,
374 we suggest the optimum forest protection age of 45 years would encourage maximum plant
375 diversity and productivity, while supporting the socio-economic conditions of the local population
376 for sustainable land use.

377 **Conclusions**

378 The present study has reported differences of plant diversity, changes in forest age structure
379 and soil nutrients of *Pinus tabulaeformis* stands restoring in chronosequence on the southeast
380 Loess Plateau, China. The richness of plant species significantly differed with age of forest
381 protection, attenuating towards more even distribution with increasing age of forest protection.
382 Sustainable forest protection not only hindered from increased organic carbon content, available
383 phosphorus and potassium in top soil, but also abated tree biomass. Our findings have practical
384 implications. By using measured forest indices as evidence-based support for balancing forest
385 management policy, ecological restoration and local economy development including sustainable
386 timber harvesting, we conclude that the preliminary optimal age for forest protection in this area
387 should be not more than 45 years.

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392 **ADDITIONAL INFORMATION AND DECLARATIONS**

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395 **Competing Interests**

396 The author declares there are no competing interests.

397 **Author Contribution**

398 Lin Hou conceived and designed the experiments, performed the experiments, and wrote the

399 paper. Sijia Hou analyzed the data, prepared figures and tables, and reviewed drafts of the paper.

400 **Data Availability**

401 The following information was supplied regarding data availability. The summary data on
402 which these analyses are based are available in Table 1 and Table 2.

403

Table 1 General information of plots

Forest farm	Plot No.	Altitude (m)	Direction	Gradient (°)	Position	Density of <i>P.tabulaeformis</i> (stems·ha ⁻¹)	Canopy density	Protected age(a)
Shibu	1	1170	Southeast	27	Upper	1025	0.30	16
	2	1150	Southeast	29	Middle	1075	0.35	
	3	1165	Northeast	24	Middle	1075	0.30	
	4	1135	Northeast	26	Lower	1100	0.30	
	5	1295	North	19	Middle	1050	0.30	
Guanzhuang	6	1154	North	22.3	Middle	1050	0.60	30
	7	1167	Northwest	24	Lower	675	0.50	
	8	1180	South	35	Upper	700	0.70	
	9	1165	South	35	Lower	1350	0.60	
	10	1180	South	22	Upper	1375	0.60	
	11	1163	South	25	Upper	1050	0.50	
Wazijie	12	1170	North	24	Upper	750	0.60	45
	13	1160	North	22.3	Middle	800	0.70	
	14	1175	North	21	Upper	625	0.60	
	15	1163	North	23	Middle	650	0.70	
	16	1154	North	26	Lower	600	0.70	
	17	1120			Gully bottom	730	0.70	
	18	1130	North	10	Lower	760	0.60	
19	1150	North	8	Ridge top	640	0.70		
Caijiachuan	20	1200	Northeast	19	Middle	1525	0.60	75
	21	1155	Northern	5	Ridge top	1550	0.40	
	22	1150	North	18	Upper	1725	0.40	
	23	1130	North	16	Lower	1475	0.40	
	24	1205	Northeast	10	Middle	1900	0.30	
	25	1200	Northeast	5	Lower	1700	0.30	
	26	1185			Mesa	1425	0.30	
	27	1135	North	18	Middle	1400	0.40	

Table 2 Richness of plant species in protected stands

Protected age(a)	Tree	Shrub	Herb
16	<i>Pinus tabulaeformis</i>	<i>Lespedeza dahurica</i>	<i>Artemisia gmelinii</i>
	<i>Populus davidiana</i>	<i>Lespedeza floribunda</i>	<i>Astragalus kifonsanicus</i>
	<i>Syringa oblata</i>	<i>Rosa hugonis</i>	<i>Artemisia mongolica</i>
		<i>Sophora japonica</i>	<i>Bothriochloa ischaemum</i>
			<i>Bupleurum chinense</i>
			<i>Kengia serotina</i>
			<i>Lamium barbatum</i>
			<i>Patrinia heterophylla</i>
			<i>Rhaponticum uniflorum</i>
			<i>Scutellaria baicalensis</i>
		<i>Viola chaerophylloides</i>	
		<i>Viola yedoensis</i>	
30	<i>Betula platyphylla</i>	<i>Acer ginnala</i>	<i>Adenophora potaninii</i>
	<i>Conus walteri wanger</i>	<i>Berberis dielsiana</i>	<i>Adenophora stricta</i>
	<i>Pinus tabulaeformis</i>	<i>Clematis brevicaudata</i>	<i>Agrimonia pilosa</i>
	<i>Prunus davidiana</i>	<i>Clematis fruticosa</i>	<i>Anaphalis margaritacea</i>
	<i>Prunus tomentosa</i>	<i>Cotoneaster multiflorus</i>	<i>Artemisia giraldii</i>
	<i>Xanthoceras sorbifolia</i>	<i>Lespedeza dahurica</i>	<i>Artemisia gmelinii</i>
	<i>Quercus Liaotungensis</i>	<i>Lonicera ferdinandii</i>	<i>Artemisia mongolica</i>
	<i>Syringa oblata</i>	<i>Ostryopsis davidiana</i>	<i>Aster tataricus</i>
		<i>Periploca sepium</i>	<i>Bothriochloa ischaemum</i>
		<i>Rhamnus davurica</i>	<i>Bupleurum chinense</i>
		<i>Rhamnus utilis</i>	<i>Carpesium divaricatum</i>
		<i>Rosa hugonis</i>	<i>Discorea nipponica</i>
		<i>Rubus corchorifolius</i>	<i>Gentiana macrophylla</i>
		<i>Sophora japonica</i>	<i>Kengia serotina</i>
		<i>Spiraea fritschiana</i>	<i>Leontopodium leontopodioides</i>
		<i>Ziziphus jujube</i>	<i>Lilium pumilum</i>
		var. <i>spinosus</i>	<i>Lysimachia barystachys</i>
			<i>Melissitus ruthenicus</i>
		<i>Patrinia heterophylla</i>	
		<i>Pennisetum clandestinum</i>	
		<i>Polygonatum odoratum</i>	
		<i>Potentilla supina</i>	

			<i>Sanguisorba officinalis</i>
			<i>Saussurea morifolia</i>
			<i>Saussurea nivea</i>
			<i>Saussurea petrovii</i>
			<i>Saussurea salsa</i>
			<i>Scutellaria baicalensis</i>
			<i>Spodiopogon sibiricus</i>
			<i>Thalictrum przewalskii</i>
			<i>Urena lobata</i>
			<i>Vicia unijuga</i>
			<i>Viola chaerophylloides</i>
			<i>Viola selkirkii</i>
			<i>Viola yedoensis</i>
	<i>Betula platyphylla</i>	<i>Acer ginnala</i>	<i>Agrimonia pilosa</i>
	<i>Pinus tabulaeformis</i>	<i>Lespedeza dahurica</i>	<i>Anaphalis margaritacea</i>
	<i>Quercus Liaotungensis</i>	<i>Lonicera maccki</i>	<i>Artemisia mongolica</i>
	<i>Syringa oblata</i>	<i>Rubus corchorifolius</i>	<i>Aster tataricus</i>
		<i>Spiraea fritschiana</i>	<i>Bothriochloa ischaemum</i>
			<i>Kengia serotina</i>
			<i>Neottianthe cucullata</i>
			<i>Potentilla discolor</i>
			<i>Spodiopogon sibiricus</i>
			<i>Urena lobata</i>
			<i>Viola chaerophylloides</i>
			<i>Viola japonica var. stenopetala</i>
			<i>Viola selkirkii</i>
			<i>Viola yedoensis</i>
45			
	<i>Pinus tabulaeformis</i>	<i>Acer ginnala</i>	<i>Adenophora stricta</i>
	<i>Populus davidiana</i>	<i>Berberis dolichobotrys</i>	<i>Anaphalis margaritacea</i>
	<i>Prunus tomentosa</i>	<i>Cotoneaster zbakelii</i>	<i>Artemisia gmelinii</i>
	<i>Quercus Liaotungensis</i>	<i>Lespedeza dahurica</i>	<i>Artemisia mongolica</i>
	<i>Syringa oblata</i>	<i>Ostryopsis davidiana</i>	<i>Aster tataricus</i>
60	<i>Toxicodendron vernicifluum</i>	<i>Rosa hugonis</i>	<i>Bothriochloa ischaemum</i>
		<i>Rubus corchorifolius</i>	<i>Bupleurum chinense</i>
		<i>Spiraea fritschiana</i>	<i>Kengia serotina</i>
			<i>Neottianthe cucullata</i>
			<i>Polygonatum sibircum</i>
			<i>Potentilla discolor</i>

			<i>Scutellaria baicalensis</i>
			<i>Spodiopogon sibiricus</i>
			<i>Vicia cracca</i>
			<i>Viola chaerophylloides</i>
			<i>Viola selkirkii</i>
			<i>Urena lobata</i>
	<i>Pinus tabulaeformis</i>	<i>Acer ginnala</i>	<i>Anaphalis margaritacea</i>
	<i>Populus davidiana</i>	<i>Clematis fruticosa</i>	<i>Artemisia mongolica</i>
	<i>Quercus Liaotungensis</i>	<i>Cotoneaster zbakelii</i>	<i>Aster tataricus</i>
	<i>Syringa oblata</i>	<i>Indigofera amblyantha</i>	<i>Bothriochloa ischaemum</i>
	<i>Toxicodendron vernicifluum</i>	<i>Lespedeza dahurica</i>	<i>Kengia serotina</i>
75		<i>Lonicera maccki</i>	<i>Sanguisorba officinalis</i>
		<i>Ostryopsis davidiana</i>	<i>Spodiopogon sibiricus</i>
		<i>Rubus corchorifolius</i>	<i>Thalictrum przewalskii</i>
		<i>Spiraea fritschiana</i>	<i>Urena lobata</i>
			<i>Viola chaerophylloides</i>
			<i>Vicia unijuga</i>
			<i>Viola selkirkii</i>

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