

A peer-reviewed version of this preprint was published in PeerJ on 21 March 2019.

[View the peer-reviewed version](https://doi.org/10.7717/peerj.6617) (peerj.com/articles/6617), which is the preferred citable publication unless you specifically need to cite this preprint.

Prieto-Amparán JA, Villarreal-Guerrero F, Martínez-Salvador M, Manjarrez-Domínguez C, Vázquez-Quintero G, Pinedo-Alvarez A. 2019. Spatial near future modeling of land use and land cover changes in the temperate forests of Mexico. PeerJ 7:e6617
<https://doi.org/10.7717/peerj.6617>

Spatial near future modeling of land use and land cover changes in the temperate forests of Mexico

Jesús A Prieto-Amparán¹, Federico Villarreal-Guerrero¹, Martin Martínez-Salvador¹, Carlos Manjarrez-Domínguez², Griselda Vázquez-Quintero², Alfredo Pinedo-Alvarez^{Corresp. 1}

¹ Facultad de Zootecnia y Ecología, Universidad Autónoma de Chihuahua, Chihuahua, Chihuahua, Mexico

² Facultad de Ciencias Agrotecnológicas, Universidad Autónoma de Chihuahua, Chihuahua, Chihuahua, Mexico

Corresponding Author: Alfredo Pinedo-Alvarez
Email address: apinedo@uach.mx

The loss of temperate forests of Mexico has continued in recent decades despite wide recognition of their importance to maintaining biodiversity. This study analyzes land use/land cover change scenarios, using satellite images from the Landsat sensor. Images corresponded to the years 1990, 2005 and 2017. The scenarios were applied for the temperate forests with the aim of getting a better understanding of the patterns in land use/land cover changes. The Support Vector Machine (SVM) multispectral classification technique served to determine the land use/land cover types, which were validated through the Kappa Index. For the simulation of land use/land cover dynamics, a model developed in Dinamica-EGO was used, which uses stochastic models of Markov Chains, Cellular Automata and Weight of Evidences. For the study, a stationary, an optimistic and a pessimistic scenario were proposed. The projections based on the three scenarios were simulated for the year 2050. Five types of land use/land cover were identified and evaluated. They were primary forest, secondary forest, human settlements, areas without vegetation and water bodies. Results from the land use/land cover change analysis show a substantial gain for the secondary forest. The surface area of the primary forest was reduced from 55.8% in 1990 to 37.7% in 2017. Moreover, the three projected scenarios estimate further losses of the surface area for the primary forest, especially under the stationary and pessimistic scenarios. This highlights the importance and probably urgent implementation of conservation and protection measures to preserve these ecosystems and their services. Based on the accuracy obtained and, on the models generated, results from these methodologies can serve as a decision tool to contribute to the sustainable management of the natural resources of a region.

1 **SPATIAL NEAR FUTURE MODELING OF LAND USE AND LAND COVER CHANGES**
2 **IN THE TEMPERATE FORESTS OF MEXICO**

3 Jesús A. Prieto-Amparan¹, Federico Villarreal-Guerrero¹, Martin Martínez-Salvador¹, Carlos
4 Manjarrez-Domínguez², Griselda Vázquez-Quintero², Alfredo Pinedo-Alvarez¹

5 ¹ Facultad de Zootecnia y Ecología, Universidad Autónoma de Chihuahua, Km 1 Perif. R. Almada,
6 Chihuahua, Chih., 31453, México.

7 ² Facultad de Ciencias Agrotecnológicas, Universidad Autónoma de Chihuahua, Km 1 Perif. R.
8 Almada, Chihuahua, Chih., 31453. México.

9

10 Corresponding Author:

11 Alfredo Pinedo-Alvarez¹

12 Perif. R. Almada, Chihuahua, Chih., 31453, México.

13 Email address: apinedo@uach.mx

14

ABSTRACT

The loss of temperate forests of Mexico has continued in recent decades despite wide recognition of their importance to maintaining biodiversity. This study analyzes land use/land cover change scenarios, using satellite images from the Landsat sensor. Images corresponded to the years 1990, 2005 and 2017. The scenarios were applied for the temperate forests with the aim of getting a better understanding of the patterns in land use/land cover changes. The Support Vector Machine (SVM) multispectral classification technique served to determine the land use/land cover types, which were validated through the Kappa Index. For the simulation of land use/land cover dynamics, a model developed in Dinamica-EGO was used, which uses stochastic models of Markov Chains, Cellular Automata and Weights of Evidence. For the study, a stationary, an optimistic and a pessimistic scenario were proposed. The projections based on the three scenarios were simulated for the year 2050. Five types of land use/land cover were identified and evaluated. They were primary forest, secondary forest, human settlements, areas without vegetation and water bodies. Results from the land use/land cover change analysis show a substantial gain for the secondary forest. The surface area of the primary forest was reduced from 55.8% in 1990 to 37.7% in 2017. Moreover, the three projected scenarios estimate further losses of the surface are for the primary forest, especially under the stationary and pessimistic scenarios. This highlights the importance and probably urgent implementation of conservation and protection measures to preserve these ecosystems and their services. Based on the accuracy obtained and, on the models generated, results from these methodologies can serve as a decision tool to contribute to the sustainable management of the natural resources of a region.

38

39 INTRODUCTION

40 Forest ecosystems are important because they provide a wide variety of products and services for
 41 the human well being (Hall et al., 2006; Fischer & Lindenmayer, 2007; Weiskittel et al., 2011)
 42 harvested products (Houghton & Nassikas, 2017), carbon sequestration (Hawkes et al., 2017), soil
 43 retention (Borrelli et al., 2017), water supply (Sun et al., 2006) and are the habitat of many species
 44 of plants and animals. However, anthropogenic activities are the main cause of degradation of
 45 almost half of the world surface in the last three centuries. That has caused the loss of lots of our
 46 precious natural resources. Twenty-five nations have practically degraded 100% of their forests,
 47 and another 29 nations have degraded 10% of their forest areas (Millennium Ecosystem
 48 Assessment, 2005).

49 Temperate forests represent a key element in the carbon cycle (Pan et al., 2011). They are
 50 important carbon dioxide sinks (Ma et al., 2017), offsetting the emissions produced by the world
 51 population (FAO, 2018). Temperate forests store 14% of the planet's carbon (Pan et al., 2011).
 52 However, projections of global environmental change show that temperate forests show high
 53 vulnerability (Gonzalez et al., 2010). This vulnerability can change the productivity of forests by
 54 modifying net carbon sequestration rates (Peters et al., 2013).

55 Temperate forests of Mexico occupy 17% of the national territory, represented by 32 millions
 56 hectares. In this region, the greatest association of pine and oak forests in the world occurs
 57 (González et al., 2012). Around 23 different species of pines and close to 200 species of oaks live
 58 in the ecoregion of Sierra Madre Occidental (Navar, 2009). However, 40 thousand hectares of
 59 forests get on average lost annually. This region has the highest deforestation rate in the world
 60 (Velázquez et al., 2002; Mas et al., 2004).

The study of the land use/land cover changes (LULCC) has become a fundamental research topic, since the change in land use/land cover (LULC) affects forest ecosystems and their biodiversity (Gharun et al., 2017). The LULCC, produced by anthropogenic activities have significantly altered the ecosystems biodiversity and services (Butler & Laurance, 2008; Miles & Kapos, 2008; Miranda-Aragón, 2013). The dynamics of LULCC directly affect the landscape patterns, the biogeochemical cycles, the ecosystems structure and function (Scheffer et al., 2001). Recently, the analysis of the spatio-temporal patterns has been the objective of several research studies (Huang et al., 2009; Manjarrez-Domínguez et al., 2015; Vázquez-Quintero et al., 2016). The models of LULCC commonly employed, quantify deforested surfaces, measuring the degree of change in the ecosystem (Lapola et al., 2011). Regression methods such as the logistic regression have been employed to generate models of LULCC. These models suppose that the relationship between the LULCC and the variables that produce it is a logistic function; however, it has been demonstrated that this relationship is too general (Mas, 2010; Mas, 2014). The dynamics and complexity of the ecosystem requires a more complete evaluation of LULCC. The spatial modeling is a technique contemplating alternative scenarios of LULCC, which could contribute to better explain the key processes influencing LULCC (Pijanowski et al., 2002; Eastman et al., 2005; Torrens, 2006; Perez-Vega et al., 2012). Thus, one of the main functions of the LULCC models is the establishment of scenarios, with the aim of changing policies and inadequate practices for the sustainable management of natural resources (DeFries et al., 2007; Berberoğlu et al., 2016). Several approaches to establish LULCC scenarios have been developed and tested to generate scenarios of LULCC. Ferrerira et al. (2012) generated deforestation scenarios to 2050 in the central Brazilian savanna biome finding the possible increase of 13.5% in deforested areas. Kamusoko et al. (2011) evaluated three scenarios (optimistic, pessimistic and business-as-usual) in the

84 Luangprabang province, Lao People's Democratic Republic, finding decreases in forest areas in
85 the pessimistic and business-as-usual scenarios and an increase in forest areas in the optimistic
86 scenario under a strict regulatory policy. Gago-Silva et al. (2017) used a combination of Bayesian
87 methods and Weights of Evidence to model the probability of change in a western part of
88 Switzerland. Galford et al. (2015) used Bayesian Weights of Evidence for policy scenarios from
89 2010 to 2050 evaluating plans for agriculture and forest in Democratic Republic of Congo.
90 The models to establish reference scenarios of changes in LULCC are based on: systems of
91 equations, statistical models, experts, evolutionary and cellular models, even though there have been
92 efforts to combine platforms in a multiagent system (Mas et al., 2014; Stan et al., 2017). The
93 statistical models employ spatial statistics and regression, in comparison with the expert models,
94 which allow the expert knowledge to lead the model path (Parker et al., 2003; Soares-Filho et al.,
95 2013). The evolutionary or cellular models are very competent to determine the ecological
96 alteration; however, they just provide information about the causality or the decision-making
97 (Parker et al., 2003).
98 The generation of LULCC scenarios for the forest region of the state of Chihuahua, Mexico is
99 necessary because of the higher temperate forest deforestation rates in the country. The generation
100 of the LULCC scenario shows two important aspects: expert knowledge and knowledge based on
101 data. Expert knowledge is useful to establish methodological processes according to the needs of
102 the user (Gounaridis et al., 2018). Knowledge based on data, helps to understand the general
103 behavior between the factors of change of land use in a spatial way (Olmedo et al., 2018). Most
104 studies are based on knowledge of the data (Peagelow and Olmedo, 2005; Kityuttachai et al.,
105 2013), however, few allow the inclusion of both (Soares-Filho et al., 2006; Olmedo et al., 2018).

The Dinamica Environment for Geoprocessing Objects (Dinamica-EGO) is a flexible open platform, which allows analyzing distribution, abundance and spatio-temporal dynamic of the landscape (Soares-Filho et al., 2002; Lima et al., 2013). The model incorporated to Dinamica-EGO employs cellular automata to simulate the changes happening in a grid, estimating the transition probability, as well as the direction of changes based in stochastic processes (Rutherford et al., 2008; Arsanjani et al., 2011). Dinamica-EGO allows users to incorporate expert knowledge into the overall statistical analysis based on the spatial data set (Mas et al., 2014). In addition, Dinamica-EGO incorporates the possibility of modifying landscape metrics in the calibration procedure to generate the simulation (Mas et al., 2012). In a comparative evaluation of approaches to modeling LULCC, two key advantages over Dinamica-EGO were emphasized: 1) incorporation of the Patcher and Expander functions. The first function generates new patches in the landscape and the second expands the previously formed patches, 2) Dinamica-EGO allows the incorporation of multiresolution validation by means of the Fuzzy Similarity Index.

The aim of the present study was (a) to evaluate the change dynamics in the period from 1990 to 2017; (b) to simulate the changes of LULCC for the year 2050 and (c) to elaborate a discussion about the impacts of different scenarios, which could happen in the future in a forest region of the state of Chihuahua, Mexico. Specifically three scenarios, pessimistic, optimistic and stationary state. The model will identify where the different types fo LULCC could hapen. This will allow that future studies could determine changes in carbon sequestration in both, on the surface extension and quantity.

MATERIALS & METHODS

Study area

The study area is located in the western part of the state of Chihuahua, Mexico. It is part of the ‘Sierra Tarahumara’ and have a surface area of 497,159 ha. Its extreme coordinates are 108° 00’ W, 29° 00’ N and 107° 10’ W, 27° 30’ N (Figure 1). It is one of the regions of temperate forests, which has experimented the greatest disturbances in the past years in the state of Chihuahua (Herrera, 2002). It belongs to the most extensive forest areas in North America. It is immersed within a complex orography composed of large canyons and deep canyons, which results in a mixture of temperate and tropical ecosystems. It is characterized by its high biodiversity and number of endemic species, estimating the presence of around 4000 species of plants. Also, it is recognized by the International Union for the Conservation of Nature as one of the megacenters of plant diversity (Felger et al., 1995). The main land uses in the area include: pine forests, oak forests, pine-oak and oak-pine forest associations, agriculture and grassland communities. The economic activities in the region are forestry, extensive livestock and rainfed agriculture (INEGI, 2003).

Figure 1. Location and elevations of the study area.

Data source

For the analysis of the LULCC, three scenes of the Landsat sensor (Path 33, Row 41), with a spatial resolution of 30 m, were used. The scenes corresponded to the years 1990, 2005 and 2017 and they were acquired from clear sky days and each of them taken during the same month to reduce the temporal variation. The scenes were downloaded from the United States Geological Survey (USGS, 2018). The characteristics of each scene can be seen in Table 1.

Table 1. Scenes characteristics.

TM= Thematic Mapper, OLI= Operational Land Imager

150 The scenes were radiometrically corrected. The radiometric correction was carried out with the
151 QGIS software v.2.8 through the SemiAutomatic Classification plugin (Congedo, 2013).

152 **Integration and composition of bands**

153 Once the scenes were corrected, they were integrated into a layer stack. False color composites for
154 the Landsat TM5 were then generated, with a combination of the bands 5, 4 and 3. Band 5
155 corresponds to the infrared channel (1.55-1.75 μm), band 4 to the near infrared (0.76-0.90 μm) and
156 the band 3 to the red channel (0.63-0.69 μm). This combination was applied to the scenes of 1990
157 and 2005. Regarding the scene of 2017, the combination for Landsat OLI8 was applied and
158 corresponded to the bands 6, 5 and 4, where band 6 corresponds to the medium infrared channel
159 (1.55-1.65 μm), band 5 to the near infrared channel (0.85-0.88 μm) and band 4 to the red channel
160 (0.64-0.67 μm) (Lillesan and Kiefer, 2000).

161 **Land use and land cover classification**

162 The Support Vector Machine (SVM) classification was applied to the 1990, 2005 and 2017 images
163 through the software R (R Core Team, 2016) with the R package “caret” (Kuhn et al., 2018) to
164 obtain LULC information. The SVM classifier is a supervised technique of nonparametric
165 statistical methods (Mountrakis & Ogle, 2011). The SVM classification has been used in several
166 research studies in the past (Kavzoglu & Colkesen 2009; Otukey and Blashke 2010; Shao &
167 Lunetta, 2012). For the supervised classification, five classes of land use were defined; 1) primary
168 forest, 2) secondary forest, 3) human settlements, 4) areas without vegetation and (5) water bodies
169 (Table 2).

170 Table 2. Land use/land cover types determined through the supervised classification method.

171 **Modeling and spatial simulation with Dinamica-EGO**

The LULCC scenarios were made based on the historical trends of change in forest cover during 1990-2017 of the supervised classifications using Dinamica-EGO (Sohares-Filho et al., 2002). The historical trends of LULCC is based on the transition matrix (Monteiro et al., 2018). Dinamica-EGO uses the algorithm of cellular automata, and the method Weights of Evidence (Olmedo et al., 2018). For the simulation of deforestation, the following steps were undertaken: 1) selection of change drivers as well as transitions, 2) exploratory analysis of the drivers of deforestation, 3) simulation and 4) validation. These four steps are described in the following sections.

Selection of variables and transitions

The selection of the set of exploratory variables to simulate the LULCC is essential for the modeling success (Miranda-Aragón et al., 2012; Perez-Vega et al 2016). In this study, 19 variables were used; 17 static and two dynamic variables. Static variables remain constant during model execution. Dynamic variables change during the execution of the model and they are continuously updated in each iteration (Olmedo et al., 2018). The set of variables used is shown in Table 3.

Table 3. Variables feeding the deforestation model.

The transition refers to the total amount of LULCC that occurred in the simulation period. In this study, the transitions of interest were: a) primary forest to secondary forest, b) primary forest to areas without apparent vegetation, c) primary forest to urban areas and d) secondary forest to areas without apparent vegetation (Table 4).

Table 4. Transitions of land use/land cover

Exploratory analysis of the data

When we modeled LULCC dynamics, Weights of Evidence (WoE) were applied to project transition probabilities. Regarding deforestation, degradation or any other type of change, we previously know about the location of favorable conditions for LULCC. The influence of static

and dynamic variables and the elaboration of the LULC maps was performed with WoE in the Dinamica-EGO software (Soares-Filho et al., 2010).

Positive values of WoE represent an attraction between a transition of land use and a specific variable. The greater the value of W^+ , the greater the probability of transition. Negative values of W^- indicate low probabilities of transition instead (Maeda et al., 2010). By using the WoE values of the variables used in the analysis of LULCC, the Dinamica-EGO model calculates the transition probability of each pixel to change. Thus, the pixels are assigned with a probability value for a given transition and probability maps are generated for the transitions of interest (Soares-Filho et al., 2009 and 2010; Mas and Flamenco, 2011).

Given that the basic hypothesis of the WoE technique is that the driving variables must be independent, for this study the correlation between the variables was tested through the Cramer Coefficient (V), represented by Equation 2.

$$V = \sqrt{\frac{\chi^2}{\Gamma \dots M}} \quad (2)$$

Where: χ^2 = is the chi-square statistic of the contingency between two variables, Γ = denotes the sum of the values of contingency, M = is the minimum of $n-1$ or $m-1$, where n denotes the number of rows and m the number of columns. Bonham-Carter (1994) mentioned that values lower than 0.5 for the Cramer Coefficient (V) suggest independence, while values higher than 0.5 involve a greater association (Almeida et al., 2003, Teixeira et al., 2009).

Simulation of land use and land cover changes

Three types of scenarios were used for 2050; they were called pessimistic, optimistic and stationary. For the three scenarios, the modeling base was the period 1990-2017. The transition matrix of 1990 and 2017 were used to estimate the possible change in forestry coverage in the future, taking 2017 as the beginning year and 2050 as the final year. In the pessimistic scenario,

the transition probability matrix and the change function (patcher and expander) were modified, increasing the deforestation and fragmentation rates between 1990 and 2017. This was done based on the hypothesis that the development of road infrastructure, urban expansion, fires, uncontrolled exploitation, among others, will produce strong spatial changes of land use. For the optimistic scenario, the state and national forest development plans were considered. Such plans promote the protection and conservation of forest resources (CONAFOR, 2001). For this scenario, the conservation and promotion of strategies to protect forests were represented by reducing the transition matrix value, as well as the patcher and expander change functions. Regarding the stationary scenario, transitions or change functions were not modified. In this case, it is assumed that the trend will be the same as the one between 1990 and 2017.

Validation

To evaluate the model performance, we used a Fuzzy Similarity Index (FSI), where the representation of a pixel is influenced by itself and its neighborhood (Ximenes et al 2011; Yanai et al., 2011; Chadid et al., 2015). The FSI employed in this study was developed by Hagen (2003), modified by Soares-Filho (2014) and implemented in Dinamica-EGO. The FSI verifies the agreement between the observed and the simulated land use and land cover datasets by obtaining the number of coincident cells within increasing window sizes of a neighborhood (Costanza 1989; Soares-Filho, 2017). The validation process was carried out by comparing a simulated map and a reference map. The simulation of the 2017 LULCC map was generated. To generate the simulation of 2017, the transition matrix was used between 1990 and 2005. The comparison through the FSI allowed to evaluate the areas of coincidence of change and no change between the real and simulated map of 2017. Finally, the general procedure used in this study is outlined in the flowchart depicted in Figure 2.

Figure 2. Flowchart of the methodological procedure followed to produce the proposed scenarios. Abbreviations: TM: Thematic Mapper, OLI: Operational Land Imager, WoE: Weights of Evidence, LUCC: Land use and cover change.

RESULTS

Detection of land use/land cover changes

Results from the analysis of LULCC show a considerable gain for secondary forest. The forest cover of the primary forest was reduced from 55.8% of the study area in 1990 to 37.7% in 2017. The areas without vegetation increased their area from 4.11% to 4.87% during 1990-2017 (Table 5). Regarding human settlements and water bodies, they showed a positive trend with an increase from 0.03% and 0.01 in 1990 to 0.1% and 0.03 in 2017, respectively. In general, the primary forest was the land use that experimented a negative trend. The rest of the land uses showed surface gains. The rate of change obtained indicate that the secondary forest, the human settlements and the water bodies were the land uses with the greatest transformation rates, with 8.03, 12.58 and 27.48, respectively, for the period of 1990-2017 and with 10.68, 15.96 and 12.3, respectively, from 2005 to 2017. Figure 3 shows the area occupied by the land uses studied. Likewise, it shows the rate of change of these land use/land cover for the periods 1990-2005 and 2005-2017. The calculated global precision, based on the Kappa Index, presented values of 80%, 85% and 84% for 1990, 2005 and 2017, respectively.

Table 5. Area occupied for five types of land uses during 1990, 2005 and 2017, and rate of change for the periods 1990-2005 and 2005-2017.

Table 6 shows the land use/land cover change dynamics. The primary forest lost the greatest surface area (28,406 ha) during 1990-2005, increasing the surface lost to 63,546 ha during 2005-2017. In contrast, the secondary forest showed the largest increases in area with 87,800 ha in the period 1990-2017.

Figure 3. Land use/land cover of 1990 (a), 2005 (b), 2017 (c), changes during 1990-2005 (d) and changes during 2005-2017. Abbreviations: AWV: areas without vegetation, SF: secondary forest, WB: water bodies, HS: human settlements and PF: primary forest.

Table 6. Land use/land cover change dynamics

Transition matrix

The transition probabilities of LULCC for the periods 1990-2005 and 2005-2017 are shown in Table 7. The diagonal of the matrix represents the permanence probability, i.e. the probability of a LULC type to remain unchanged. The areas without vegetation showed a 90% probability of transition from 1990 to 2005, lowering it to 62% from 2005 to 2017. The areas of primary forest presented a negative trend with a 71% probability of permanence in the period 1990 to 2005, and changing it to 61% for the period 2005-2017.

Table 7. Transition matrix of probability for land use/land cover change (1990-2005, 2005-2017, 1990-2017).

Weights of evidence (WoE) analysis

The WoE of the 19 variables were analyzed to eliminate those values that were above 0.5, based on the Cramer Coefficient (V). The distance to urban locations showed positive values of WoE from 1000 to 9000 m distance and from 42,000 to 47,000 m indicating an influence for cover change from secondary forest to area without vegetation. The distance to rural localities showed positive values of WoE in distances from 0 to 700 m. The topographic position index showed positive values in the ranges of -150 to -60 and 120 to 240. The distance to sawmills indicates that deforestation appears from 0 to 16,000 m with respect to the process of change between secondary forest to areas without vegetation. The transition from primary forest to area without vegetation is likely to occur in distances to the main roads between 13,000 and 21,000 m. The density of main streams such as rivers and creeks had an influence in densities from 0.039 to 0.079 m²/ km².

In the transition from primary forest to secondary forest, the variable altitude showed positive values of WoE in the range of 1,200-1,300 m, suggesting that most of the changes occur in this range. The slope showed that the process of change between primary forest and secondary forest is located on slopes of 45-60 and 60-75 degrees. The transition from primary forest to human settlements was influenced by the distance to secondary streams from 500 to 1000 meters. The distance to sawmills presented an influence from 0 to 6,000 meters. The distance to mines showed that the attraction to change occurs between 2000 and 10,000 m.

Model validation

The model validation was carried with the simulated and the true land use classification of 2017. The FSI was applied for neighborhoods from 1 x 1 to 7 x 7 pixels. The minimum value reported for FSI was 49% in 1x1 pixels, while in 7x7 pixels the value of FSI was 91%. These results indicate that the real and simulated land use changes agree from 49% to 91%. Simulation starts with 49% and adjusts to 91%, reaching a similarity adjustment value at a distance of 210 m. These results agree with that obtained by Ximenes et al. (2011). According to Soares-Filho (2017), and similar studies (Carlson et al., 2012; de Rezende et al., 2015; Elz et al., 2015), for the resolution and the number of transitions considered in the model, the values obtained for the FSI suggest that the models are good and can be used in the simulation of LULCC scenarios. Figure 4 represents the FSI in relation to the size of the window.

Figure 4. Variation of the FSI as a function of different distance.

Scenarios

The LULCC based on the transitions between 1990 and 2017 for the stationary, optimistic and pessimistic scenarios are presented in Table 8.

Table 8. Percentage of surface area occupied by five land use/land cover types and rate of change for 2017-2050 based on three scenarios.

Figure 5 shows the LULC classification of 2017 and the stationary, optimistic and pessimistic scenarios for 2050, after the model calibration.

Figure 5. a) Land use/land cover of 2017 and simulated land use/land cover projected for the year 2050 as a result of the b) Stationary, c) Pessimistic and d) Optimistic scenarios. Abbreviations: AWV: areas without vegetation, SF: secondary forest, WB: water bodies, HS: human settlements and PF: primary forest.

In the stationary scenario the area without vegetation would increase from 4.8% in 2017 to 5.27%. Likewise, the secondary forest would increase from 57.7% (2017) to 73%. For this scenario, the changes in human settlement and water bodies would not increase or reduce their area. Conversely, the rate of change of primary forest and secondary forest were the greatest between 2017 and 2050. Regarding the optimistic scenario, it showed reductions in areas of primary forest; however, in lower magnitudes than for the stationary and pessimistic scenarios. For the pessimistic scenario, the Markov matrix was modified considering a greater pressure on the forest ecosystem. The area without vegetation showed a positive trend, with 4.8% in 2017 and an increase to almost 8% in 2050. The secondary forest would go from 57.7% to 85.6% in 2050. Finally, the primary forest would reduce its area to a 8% and isolated forest areas would appear. The rate of change for this scenario were the ones that showed the highest values. The LULCC dynamics projected for 2050 for the three scenarios (stationary, optimistic, pessimistic) is presented in Table 9.

Table 9. Land use/land cover change dynamics (ha) under three projected scenarios.

DISCUSSION

In this study, scenarios of LULCC for 2017 and 2050 were generated for a temperate forest region of Chihuahua Mexico. The scenarios were developed in Dinamica-EGO. Results were consistent

with the results described by Maeda et al. (2011). For the generation of transitions and simulation of scenarios, LULC of 1990, 2005 and 2017 were determined. In general, proximity to sources with anthropogenic activity as well as topography were important factors influencing the change in forest cover. The exchange between primary forest and secondary forest represented the main transition between 1990 and 2017. This transition produced the greatest impact, in agreement with the results reported by Perez-Vega et al. (2016). Such transition was influenced by the altitude, slope, and density of water streams, in agreement with the results of Armenteras et al. (2006) and Chadid et al. (2015). The transition from primary to secondary forest could be attributed to the reduction in pine vegetation, where shrubs would become dominant. A consequence of the reduction of primary forest is the migration of fauna, which deals with the dispersal of the seeds of large-crowned trees (Lehouck et al., 2009). Other consequences include the change of lands to livestock production systems (Maeda et al., 2010) and the presence of areas with high solar incidence and low coverage, which are prone to fires (de Rezende et al., 2015). Another reason for the reduction of primary forest is the proximity to urban rural localities and roads, which is in agreement with the results reported by Aguiar et al. (2007) and Osorio et al. (2015). The proximity to urban and rural communities indicates the possible extraction of wood for export and also facilitates the expansion of the agricultural or grazing frontier (Chadid et al., 2015). This can be verified by the number of sawmills in the study area. The process of deforestation/degradation is strongly related to this cause. In the forested areas of Chihuahua, the rural localities are in a high degree of marginalization (González et al., 2012) where there exist agricultural incentives PEF 2025 (CONAFOR, 2001), causing the possible increase of the areas without vegetation. Another reason for the degradation may be the distance to the main roads and the topographic position.

The results obtained for the different scenarios showed differences among the surfaces of land use. The stationary scenario resulted in a considerable change in the primary forest, mainly. This scenario considers that the transition values among land use coverages will continue. The long-term impacts of the deforestation/degradation process include increased reservoir sedimentation and decreased flows in the dry season (Gingrich, 1993). Although the optimistic scenario showed increases in non-forested areas, this scenario was the one that showed the greatest resistance to the transitions from primary forest to any other LULC. This scenario considers the strict application of the regulation of forest resources, in agreement with the general trend in the protection of forest ecosystems to degradation (UN, 2015) and the projections of the PEF 2025 (PEF, 2001). The pessimistic scenario showed the greatest losses in the coverage of the primary forest. In addition, the increase in areas without vegetation, which is mainly associated to cropping and the proximity to water currents, is one of the main outputs of the pessimistic scenario, which agrees with the study by Elz et al. (2015). The increase in agricultural areas resulting from this scenario may benefit the inhabitants economically; however, the expansion of this type of land use/land cover could lead to a greater demand of water for irrigation purposes, which could potentially impact water resources (Maeda et al., 2010).

Population growth (Barni et al., 2015), the market demand and the lack of technification for wood processing cause the opening of land and the extraction of wood for self-consumption. Taking these aspects into account, the simulation of changes in forest cover indicates pressure on forest resources, which is consistent with that found by Kamusko et al., (2011). As a consequence, forest degradation could lead to soil loss (Quan et al., 2011), loss in biodiversity (Falcucci et al., 2007) and landscape connectivity (Tambosi et al., 2014), habitat fragmentation (Nagendra et al., 2004), the presence of invasive species (Mas et al., 2012), among others.

The LULCC model of this study incorporated the Markov chains, Cellular Automata and WoE methods. Several transitions were simulated as in the studies by Soares-Filho et al., (2010), Ferreira et al. (2013) and Elz et al. (2015). The validation was carried out based on the FSI, as it was also performed in previous research (Ximenes et al., 2011). The result of this analysis, where the three aforementioned methods are combined, highlighted the variables driving the process of degradation/deforestation, as well as the manipulation based on the knowledge of the transition probabilities, being more suitable for the simulation of LULCC (Mas & Flamenco, 2011). The transition probability matrices revealed that the primary forest has a negative trend in its occupied area, suggesting that degradation will continue over this land use, this area of primary forest changed to secondary forest. Although the other transitions did not produce important changes in the spatial configuration of the landscape, but their cumulative long-term effect could negatively impact the functioning of the ecosystems and their biodiversity (Pompa, 2008).

In this study, we focused on hypothetical scenarios where the pressure of forest resources was controlled by changing the transition probability. However, it is necessary to study scenarios where market demand (Merry et al., 2009) or illegal timber extraction (Chadid et al., 2015) is considered. The wood clandestine corresponds to 30% in the some forest management units of Chihuahua (Silva, 2009).

The scenarios are not exact projections of the future state of the environment (Feng and Liu, 2016). However, it is an alternative means of supporting forest managers, which can serve as a valuable tool for studying political decisions (Kolb and Galicia, 2018). That would lead to a better knowledge of forest exploitation and protection. Managers can take into account the proposed scenarios and take decisions based on the one with the most promising results.

Due to the distribution of economic information (municipality based) and the lack of information from georeferenced illicit extractions, we believe an approach such as agent-based models would help to improve the study and address these issues. Finally, the model did not consider climatic variations such as precipitation and temperature, which can affect patterns and dynamics in recovery zones. That should be implemented in future studies.

CONCLUSIONS

The use of scenarios as a methodology to study LULCC has been studied in depth at different scales and in different areas. However, several improvements can be implemented. This study presents an approach that integrates expert knowledge, and geospatial technologies such as geographic information systems and spatial simulation. The developed scenarios were based on the application of the forestry law (non-spatially) as well as the state of the landscape, and not only on the extrapolation of past trends. In addition, the scenarios are spatially explicit, which allow identifying the spatial pattern of change and the possible critical areas of change in forest cover. Finally, this study contributes to the understanding of the future fragmentation of the forest cover. Therefore, the current decisions in the field of forest management and land use/land cover influence the future of our forests and can probably be represented in one of the three proposed scenarios.

REFERENCES

- Aguiar APD, Câmara G, Escada, MIS. 2007. Spatial statistical analysis of land-use determinants in the Brazilian Amazonia: Exploring intra-regional heterogeneity. *Ecological modelling* 209(2-4):169-188 DOI <https://doi.org/10.1016/j.ecolmodel.2007.06.019>.
- Almeida CM, Batty M, Monteiro AMV, Camara G, Soares-Filho BS, Cerqueira GC, Pennachin, CL. 2003. Stochastic cellular automata modelling of urban land use dynamics: Empirical development and estimation. *Computers, Environment and Urban Systems* 27:481-509 DOI [https://doi.org/10.1016/S0198-9715\(02\)00042-X](https://doi.org/10.1016/S0198-9715(02)00042-X).

- 430 Armenteras D, Rudas G, Rodriguez N, Sua S, Romero M. 2006. Patterns and causes of
431 deforestation in the Colombian Amazon. *Ecological Indicators* 6(2):353-368 DOI
432 <https://doi.org/10.1016/j.ecolind.2005.03.014>.
- 433 Arsanjani JJ, Helbich M, Kainz W, Boloorani AD. 2013. Integration of logistic regression, Markov
434 chain and cellular automata models to simulate urban expansion. *International Journal of Applied*
435 *Earth Observation and Geoinformation*, 21: 265-275 DOI
436 <https://doi.org/10.1016/j.jag.2011.12.014>.
- 437 Barni PE, Fearnside, PM, de Alencastro Graça PML. 2015. Simulating deforestation and carbon
438 loss in Amazonia: impacts in Brazil's Roraima state from reconstructing Highway BR-319
439 (Manaus-Porto Velho). *Environmental management*, 55(2): 259-278 DOI
440 <https://doi.org/10.1007/s00267-014-0408-6>.
- 441 Berberoğlu, S, Akın, A, Clarke, KC. 2016. Cellular automata modeling approaches to forecast
442 urban growth for adana, Turkey: A comparative approach. *Landscape and Urban Planning* 153:
443 11-27 DOI <https://doi.org/10.1016/j.landurbplan.2016.04.017>.
- 444 Bonham-Carter G. 1994. *Geographic Information Systems for Geoscientists: Modelling with GIS*.
445 New York: Pergamon.
- 446 Borrelli P, Robinson DA, Fleischer LR, Lugato E, Ballabio C, Alewell C, Meusburger K,
447 Modungo S, Schütt B, FerroV, Bagarello V, Oost K, Montanarella L, Panagos P. 2017. An
448 assessment of the global impact of 21st century land use change on soil erosion. *Nature*
449 *communications* 8:2013 DOI <https://doi.org/10.1038/s41467-017-02142-7>
- 450 Butler, RA, Laurance WF. 2008. New strategies for conserving tropical forests. *Trends in Ecology*
451 *& Evolution* 23:469-472 DOI <https://doi.org/10.1016/j.tree.2008.05.006>.
- 452 Carlson KM, Curran LM, Ratnasari D, Pittman AM, Soares-Filho, BS, Asner GP, Lawrence D,
453 Rodrigues HO. 2012. Committed carbon emissions, deforestation, and community land conversion
454 from oil palm plantation expansion in West Kalimantan, Indonesia. *Proceedings of the National*
455 *Academy of Sciences* 109(19): 7559-7564 DOI <https://doi.org/10.1073/pnas.1200452109>.
- 456 Chadid MA, Dávalos LM, Molina J, Armenteras D. 2015. A Bayesian spatial model highlights
457 distinct dynamics in deforestation from coca and pastures in an Andean biodiversity
458 hotspot. *Forests* 6(11): 3828-3846 DOI [doi:10.3390/f6113828](https://doi.org/10.3390/f6113828)
- 459 Comisión Nacional Forestal (CONAFOR). 2001. Programa Estratégico Forestal 2025. Available
460 at <http://www.conafor.gob.mx:8080/documentos/ver.aspx?articulo=307&grupo=4>.
- 461 Congedo, L. 2017. Semi-Automatic Classification Plugin for QGIS, Technical Report, Sapienza
462 University, ACC Dar Project: Rome, Italy. Available at
463 [https://media.readthedocs.org/pdf/semiautomaticclassificationmanual-](https://media.readthedocs.org/pdf/semiautomaticclassificationmanual-v3/latest/semiautomaticclassificationmanual-v3.pdf)
464 [v3/latest/semiautomaticclassificationmanual-v3.pdf](https://media.readthedocs.org/pdf/semiautomaticclassificationmanual-v3/latest/semiautomaticclassificationmanual-v3.pdf)
- 465 Costanza R. 1989. Model goodness of fit: a multiple resolution procedure. *Ecological Modelling*
466 47: 199-215 DOI [https://doi.org/10.1016/0304-3800\(89\)90001-X](https://doi.org/10.1016/0304-3800(89)90001-X).
- 467 de Rezende CL, Uezu A, Scarano FR, Araujo DSD. 2015. Atlantic Forest spontaneous
468 regeneration at landscape scale. *Biodiversity and conservation* 24(9): 2255-2272 DOI
469 <https://doi.org/10.1007/s10531-015-0980-y>.

- 470 DeFries R, Hansen A, Turner BL, Reid, R, Liu J. 2007. Land use change around protected areas:
471 management to balance human needs and ecological function. *Ecological Applications* 17(4):
472 1031-1038 DOI <https://doi.org/10.1890/05-1111>.
- 473 Eastman JR, Solorzano LA, Van Fossen ME. 2005. Transition potential modeling for land-cover
474 change In: GIS, spatial analysis, and modeling, eds. Maguire D Batty JM, Goodchild MF. ESRI
475 Press, California.
- 476 Elz I, Tansey K, Page SE, Trivedi M. 2015. Modelling deforestation and land cover transitions of
477 tropical peatlands in Sumatra, Indonesia using remote sensed land cover data sets. *Land* 4(3): 670-
478 687 DOI [10.3390/land4030670](https://doi.org/10.3390/land4030670).
- 479 Falcucci A, Maiorano L, Boitani L. 2007. Changes in land-use/land-cover patterns in Italy and
480 their implications for biodiversity conservation. *Landscape ecology* 22(4): 617-631 DOI
481 <https://doi.org/10.1007/s10980-006-9056-4>.
- 482 FAO, 2018: The state of world's forests 2018 - Forest pathways to sustainable development. FAO,
483 Rome. Available at <http://www.fao.org/publications/sofo/en/> (Accessed 7 October 2018).
- 484 Felger R S, Wilson MF 1995. Northern Sierra Madre Occidental and its Apachian outliers: a
485 neglected center of biodiversity. Biodiversity and Management of the Madrean Archipelago: The
486 Sky Islands of Southwestern United States and Northwestern Mexico, 36-51.
- 487 Ferreira ME, Ferreira Jr LG, Mizziara, F, Soares-Filho, BS. 2013. Modeling landscape dynamics in
488 the central Brazilian savanna biome: future scenarios and perspectives for conservation. *Journal*
489 *of Land Use Science* 8(4): 403-421 DOI <https://doi.org/10.1080/1747423X.2012.675363>.
- 490 Fischer J, Lindenmayer DB. 2007. Landscape modification and habitat fragmentation: a
491 synthesis. *Global ecology and biogeography* 16(3): 265-280 DOI <https://doi.org/10.1111/j.1466-8238.2007.00287.x>.
- 493 Gago-Silva A, Ray N, Lehmann A. 2017. Spatial Dynamic Modelling of Future Scenarios of Land
494 Use Change in Vaud and Valais, Western Switzerland. *ISPRS International Journal of Geo-*
495 *Information* 6:115 DOI: [10.3390/ijgi6040115](https://doi.org/10.3390/ijgi6040115).
- 496 Galford GL, Soares-Filho BS, Sonter LJ, Laporte, N. 2015. Will passive protection save Congo
497 forests?. *PloS one*, 10: e0128473 DOI <https://doi.org/10.1371/journal.pone.0128473>.
- 498 Gharun M, Possell M, Bell TL, Adams MA. 2017. Optimisation of fuel reduction burning regimes
499 for carbon, water and vegetation outcomes. *Journal of environmental management* 203:157-170
500 DOI: <https://doi.org/10.1016/j.jenvman.2017.07.056>.
- 501 Gingrich RW. 1993. The political ecology of deforestation in the Sierra Madre Occidental of
502 Chihuahua (Master's thesis, University of Arizona).
- 503 González BA. 2012. La Sierra Tarahumara, el bosque y los pueblos originarios: estudio de caso
504 de Chihuahua México. Available at: [http://www.fao.org/forestry/17194-](http://www.fao.org/forestry/17194-0381f923a6bc236aa91ecf614d92e12e0.pdf)
505 [0381f923a6bc236aa91ecf614d92e12e0.pdf](http://www.fao.org/forestry/17194-0381f923a6bc236aa91ecf614d92e12e0.pdf)
- 506 Gonzalez P, Neilson RP, Lenihan JM, Drapek RJ. 2010. Global patterns in the vulnerability of
507 ecosystems to vegetation shifts due to climate change. *Global Ecology and Biogeography* 19:755-
508 768 DOI <https://doi.org/10.1111/j.1466-8238.2010.00558.x>.

- 509 Gounaridis D, Chorianopoulos I, Koukoulas, S. 2018. Exploring prospective urban growth trends
510 under different economic outlooks and land-use planning scenarios: the case of Athens. *Applied*
511 *Geography* 90:134-144 DOI <https://doi.org/10.1016/j.apgeog.2017.12.001>.
- 512 Hall RJ, Skakun RS, Arsenault EJ, Case BS. 2006. Modeling forest stand structure attributes using
513 Landsat ETM+ data: Application to mapping of aboveground biomass and stand volume. *Forest*
514 *ecology and management* 225(1-3): 378-390 DOI <https://doi.org/10.1016/j.foreco.2006.01.014>
- 515 Hawkes CV, Waring BG, Rocca JD, Kivlin SN. 2017. Historical climate controls soil respiration
516 responses to current soil moisture. *Proceedings of the National Academy of Sciences* 114:6322-
517 6327 DOI: <https://doi.org/10.1073/pnas.1620811114>.
- 518 Herrera A. 2002. Situación actual de los bosques de Chihuahua. *Madera y bosques* 8(1) DOI
519 <http://dx.doi.org/10.21829/myb.2002.811302>.
- 520 Houghton RA, Nassikas, AA. 2017. Global and regional fluxes of carbon from land use and land
521 cover change 1850–2015. *Global Biogeochemical Cycles* 31:456-472 DOI
522 <https://doi.org/10.1002/2016GB005546>.
- 523 Huang B, Zhang L, Wu B. 2009. Spatiotemporal analysis of rural–urban land
524 conversion. *International Journal of Geographical Information Science* 23(3): 379-398 DOI
525 <https://doi.org/10.1080/13658810802119685>.
- 526 Instituto Nacional de Estadística, Geografía e Informática (INEGI). 2003. Síntesis de Información
527 Geográfica del Estado de Chihuahua.
- 528 Kamusoko C, Oono K, Nakazawa A, Wada Y, Nakada R, Hosokawa T, Someya T. 2011. Spatial
529 simulation modelling of future forest cover change scenarios in Luangprabang province, Lao PDR.
530 *Forests* 2(3): 707-729 DOI 10.3390/f2030707.
- 531 Kavzoglu, T, Colkesen I. 2009. A kernel functions analysis for support vector machines for land
532 cover classification. *International Journal of Applied Earth Observation and*
533 *Geoinformation*, 11(5): 352-359 DOI <https://doi.org/10.1016/j.jag.2009.06.002>.
- 534 Kityuttachai K, Tripathi, NK, Tipdecho T, Shrestha R. 2013. CA-Markov analysis of constrained
535 coastal urban growth modeling: Hua Hin seaside city, Thailand. *Sustainability* 5(4): 1480-1500
536 DOI <https://doi.org/10.3390/su5041480>.
- 537 Kuhn M, Wing J, Weston S. 2015. Package ‘caret’. Classification and regression training.
- 538 Lapola D, Schaldach MR, Alcamo J, Bondeau A, Msangi S, Priess JA, Soares-Filho BS. 2011.
539 Impacts of climate change and the end of deforestation on land use in the Brazilian Legal
540 Amazon. *Earth Interactions*. 15:1-29 DOI <https://doi.org/10.1175/2010EI333.1>.
- 541 Lehouck V, Spanhove T, Colson L, Adringa-Davis A, Cordeiro NJ, Lens L. 2009. Habitat
542 disturbance reduces seed dispersal of a forest interior tree in a fragmented African cloud forest.
543 *Oikos* 118(7): 1023-1034 DOI <https://doi.org/10.1111/j.1600-0706.2009.17300.x>.
- 544 Lillesand TM, Kiefer RW. Remote Sensing and Image Interpretation, 4th ed.; John Wiley and
545 Sons: New York, NY, USA, 2000.
- 546 Lima, T C, Guilen-Lima CM, Oliveira MS, Soares-Filho B. 2013. DINAMICA EGO e Land
547 Change Modeler para simulação de desmatamento na Amazônia brasileira: análise comparativa.
548 In: Anais XVI Simpósio Brasileiro de Sensoriamento Remoto: Foz do Iguaçu, INPE, 6379-6386.

- 549 Ma W, Jia G, Zhang A. 2017. Multiple satellite-based analysis reveals complex climate effects of
550 temperate forests and related energy budget. *Journal of Geophysical Research: Atmospheres*,
551 122(7): 3806-3820 DOI <https://doi.org/10.1002/2016JD026278>.
- 552 Maeda EE, Clark B, Pellikka P, Siljander M. 2010. Driving forces of land-use change in the Taita
553 Hills, Kenya. In: 13th AGILE International Conference on Geographic Information Science, 2-5.
- 554 Maeda EE, Clark B, Pellikka P, Siljander M. 2010. Modelling agricultural expansion in Kenya's
555 eastern arc mountains biodiversity hotspot. *Agricultural Systems* 103: 609-620 DOI
556 <https://doi.org/10.1016/j.agsy.2010.07.004>.
- 557 Maeda EE, De Almeida C M, de Carvalho Ximenes A, Formaggio AR, Shimabukuro YE, Pellikka
558 P. 2011. Dynamic modeling of forest conversion: Simulation of past and future scenarios of rural
559 activities expansion in the fringes of the Xingu National Park, Brazilian Amazon. *International
560 Journal of Applied Earth Observation and Geoinformation* 13(3): 435-446 DOI
561 <https://doi.org/10.1016/j.jag.2010.09.008>.
- 562 Manjarrez-Dominguez C, Pinedo-Alvarez A, Pinedo-Alvarez C, Villarreal-Guerrero F, Cortes-
563 Palacios L. 2015. Vegetation landscape analysis due to land use changes on arid lands. *Polish
564 Journal of Ecology* 63: 167-174 DOI <https://doi.org/10.3161/15052249PJE2015.63.2.001>.
- 565 Mas J, Kolb M, Houet T, Paegelow M, Olmedo MC. 2010. Una comparación de diferentes
566 enfoques de modelación de cambios de cobertura/uso del suelo. In: Proceedings of the XIV
567 Simposio Internacional SELPER 2010.
- 568 Mas JF, Flamenco A. 2011. Modelación de los cambios de coberturas/uso del suelo en una región
569 tropical de México. *GeoTrópico* 5(1): 1-24 DOI <http://dx.doi.org/10.5154/r.rchscfa.2014.10.049>.
- 570 Mas JF, Kolb M, Paegelow M, Olmedo MTC, Houet T. 2014. Inductive pattern-based land
571 use/cover change models: A comparison of four software packages. *Environmental Modelling &
572 Software* 51: 94-111 DOI <https://doi.org/10.1016/j.envsoft.2013.09.010>.
- 573 Mas JF, Pérez-Vega A, Clarke KC. 2012. Assessing simulated land use/cover maps using
574 similarity and fragmentation indices. *Ecological Complexity* 11: 38-45 DOI
575 <https://doi.org/10.1016/j.ecocom.2012.01.004>.
- 576 Mas JF, Velázquez A, Díaz-Gallegos JR, Mayorga-Saucedo R, Alcántara C, Bocco G, Castro R,
577 Fernandez T, Pérez-Vega A. 2004. Assessing land use/cover changes: a nationwide multidecadate
578 spatial database for Mexico. *International Journal of Applied Earth Observation and
579 Geoinformation* 5(4): 249-261 DOI <https://doi.org/10.1016/j.jag.2004.06.002>.
- 580 Merry F, Soares-Filho BS, Nepstad D, Aamacher G, Rodrigues H. 2009. Balancing, conservation
581 and economic sustainability: the future of the amazon timber industry. *Environmental
582 Management* 44(3): 395-407 DOI <https://doi.org/10.1007/s00267-009-9337-1>.
- 583 Miles L, Kapos V. 2008. Reducing greenhouse gas emissions from deforestation and forest
584 degradation: global land-use implications. *Science* 320(5882): 1454-1455 DOI
585 [10.1126/science.1155358](https://doi.org/10.1126/science.1155358).
- 586 Millennium Ecosystem Assessment. 2005. Ecosystems and Human Well-Being: Synthesis. Island
587 Press, Washington, DC.
- 588 Miranda-Aragón L, Treviño-Garza EJ, Jiménez-Pérez J, Aguirre-Calderón OA, González-Tagle
589 MA, Pompa-García M, Aguirre-Salado CA. 2012. Modeling susceptibility to deforestation of

- 590 remaining ecosystems in North Central Mexico with logistic regression. *Journal of forestry*
591 *research* 23(3): 345-354 DOI <https://doi.org/10.1007/s11676-012-0230-z>.
- 592 Monteiro Junior JJ, Silva E., De Amorim Reis AL, Mesquita Souza Santos, JP. 2018. Dynamical
593 spatial modeling to simulate the forest scenario in Brazilian dry forest landscapes. *Geology,*
594 *Ecology, and Landscapes*, 1:7 DOI <https://doi.org/10.1080/24749508.2018.1481658>.
- 595 Mountrakis G, Im J, Ogole C. 2006. Support vector machines in remote sensing: A review. ISPRS
596 J. Photogramm. Remote Sens. 2011, 66, 247–259 DOI
597 <https://doi.org/10.1016/j.isprsjprs.2010.11.001>
- 598 Nagendra H, Munroe DK, Southworth J. 2004. From pattern to process: landscape fragmentation
599 and the analysis of land use/land cover change. *Agriculture, Ecosystems & Environment* 101(2):
600 111-115 DOI <https://doi.org/10.1016/j.agee.2003.09.003>
- 601 Navar J. 2009. Allometric equations for tree species and carbon stocks for forests of northwestern
602 Mexico. *Forest ecology and Management* 257(2) 427-434 DOI
603 <https://doi.org/10.1016/j.agee.2003.09.003>.
- 604 Olmedo MTC, Paegelow M, Mas JF, Escobar F. 2018. Geomatic Approaches for Modeling Land
605 Change Scenarios. Switzerland: Springer.
- 606 Osorio LP, Mas JF, Guerra F, Maass M. 2015. Análisis y modelación de los procesos de
607 deforestación: un caso de estudio en la cuenca del río Coyuquilla, Guerrero,
608 México. *Investigaciones geográficas* (88): 60-74 DOI <http://dx.doi.org/10.14350/rig.43853>.
- 609 Otukei JR, Blaschke T. 2010. Land cover change assessment using decision trees, support vector
610 machines and maximum likelihood classification algorithms. *International Journal of Applied*
611 *Earth Observation and Geoinformation* 12: S27-S31 DOI
612 <https://doi.org/10.1016/j.jag.2009.11.002>
- 613 Paegelow M, Olmedo, MTC. 2005. Possibilities and limits of prospective GIS land cover
614 modelling—a compared case study: Garrotxes (France) and Alta Alpújarra Granadina (Spain).
615 *International Journal of Geographical Information Science* 19(6): 697-722 DOI
- 616 Pan Y, Birdsey RA., Fang J, Houghton R, Kauppi PE, Kurz WA, Oliver L, Shvidenko PA, Lewis
617 SL, Canadell JG, Ciais, P, Jackson RB, Pacala S, McGuire AD, Piao s, Rautiainen A, Sitch s,
618 Hayes D. 2011. A large and persistent carbon sink in the world's forests. *Science*, 1201609 DOI:
619 10.1126/science.1201609.
- 620 Parker D, Manson S, Jansen M, Hoffman M, Deadman P. 2003. Multi-Agent systems for the
621 simulation of land-Use and land-Cover change: a review. *Annals of the American Association of*
622 *Geographers* 93(2): 314–337 DOI <https://doi.org/10.1111/1467-8306.9302004>.
- 623 Pérez-Vega A, Mas JF, Ligmann-Zielinska A. 2012. Comparing two approaches to land use/cover
624 change modeling and their implications for the assessment of biodiversity loss in a deciduous
625 tropical forest. *Environmental Modelling & Software* 29(1): 11-23 DOI
626 <https://doi.org/10.1016/j.envsoft.2011.09.011>.
- 627 Pérez-Vega A, Rocha Álvarez F, Regil García HH. 2016. Distribución espacial del uso/cubierta
628 del suelo y degradación forestal en la reserva de la biosfera Sierra Gorda de Guanajuato. *Acta*
629 *Universitaria* 26(2) DOI 10.15174/au.2016.1500.

- 630 Peters EB, Wythers KR, Bradford JB, Reich PB. 2013. Influence of disturbance on temperate
631 forest productivity. *Ecosystems* 16(1):95-110 DOI <https://doi.org/10.1007/s10021-012-9599-y>.
- 632 Pijanowski BC, Brown DG, Shellito BA, Manik, GA. 2002. Using neural networks and GIS to
633 forecast land use changes: a land Transformation Model. *Computers, Environment and Urban
634 Systems* 26:553–575 DOI [https://doi.org/10.1016/S0198-9715\(01\)00015-1](https://doi.org/10.1016/S0198-9715(01)00015-1).
- 635 Pompa M. 2008. Análisis de la deforestación en ecosistemas montañosos del noroeste de México.
636 *Avances en Investigación Agropecuaria*, 12(2).
- 637 Quan B, Römken MJM, Li R, Wang F, Chen J. 2011. Effect of land use and land cover change
638 on soil erosion and the spatio-temporal variation in Liupan Mountain Region, southern Ningxia,
639 China. *Frontiers of Environmental Science & Engineering in China* 5(4): 564-572 DOI
640 <https://doi.org/10.1007/s11783-011-0348-9>.
- 641 R Core Team. 2016. R: A language and environment for statistical computing. Vienna: R
642 Foundation for Statistical Computing.
- 643 Rutherford GN, Bebi P, Edwards PJ, Zimmermann NE. 2008. Assessing land-use statistics to
644 model land cover change in a mountainous landscape in the European Alps. *Ecological
645 modelling* 212(3-4):460-471 DOI <https://doi.org/10.1016/j.ecolmodel.2007.10.050>.
- 646 Scheffer M, Carpenter S, Foley JA, Folke C, Walker B. 2001. Catastrophic shifts in
647 ecosystems. *Nature* 413(6856):591 DOI 10.1038/35098000
- 648 Shao Y, Lunetta RS. 2012. Comparison of support vector machine, neural network, and CART
649 algorithms for the land-cover classification using limited training data points. *ISPRS Journal of
650 Photogrammetry and Remote Sensing* 70:78-87 DOI
651 <https://doi.org/10.1016/j.isprsjprs.2012.04.001>.
- 652 Silva RS. 2009. Estudio Regional Forestal “Unidad de Manejo Forestal San Juanito A.C.
653 UMAFOR San Juanito A. C.
- 654 Soares-Filho B, Rodrigues H, Follador M. 2013. A hybrid analytical-heuristic method for
655 calibrating land-use change models. *Environmental Modelling & Software* 43: 80-87 DOI
656 <https://doi.org/10.1016/j.envsoft.2013.01.010>.
- 657 Soares-Filho BS, Cerqueira GC, Pennachin CL, 2002. DINAMICA — a stochastic cellular
658 automata model designed to simulate the landscape dynamics in an Amazonian colonization
659 frontier. *Ecological Modelling* 154: 217–235 DOI [https://doi.org/10.1016/S0304-3800\(02\)00059-](https://doi.org/10.1016/S0304-3800(02)00059-5)
660 5.
- 661 Soares-Filho BS, Ferreira BM, Filgueira DS; Rodrigues HO, Hissa LBV, Lima LS, Machado RF,
662 Costa WLS. Dinamica project. Remote Sensing Center. Federal University of Minas Gerais
663 (UFMG), Belo Horizonte, MG, Brazil. <http://www.csr.ufmg.br/dinamica/>. Accessed 24 Jul 2017.
- 664 Soares-Filho BS, Moutinhob P, Nepstad D, Anderson A, Rodrigues H, Garcia R, 2010. Role
665 of Brazilian amazon protected areas in climate change mitigation. *Proceedings of the National
666 Academy of the United States of America* 107(24): 10821-10826; DOI:
667 <https://doi.org/10.1073/pnas.0913048107>
- 668 Soares-Filho BS, Nepstad DC, Curran LM, Cerqueira GC, Garcia RA, Ramos CA, Voll E,
669 McDonald A, Lefebvre P, Schlesinger, P. 2006. Modelling conservation in the Amazon
670 basin. *Nature* 440(7083):520 DOI <https://doi.org/10.1038/nature04389>.

- 671 Soares-Filho BS, Rodrigues H, Costa, W. 2009. Modeling Environmental Dynamics with
672 Dinamica Ego.
673 Available at: http://www.lapa.ufscar.br/geotecnologias-1/Dinamica_EGO_guidebook.pdf
- 674 Stan K D, Sanchez-Azofeifa A. 2017. The Edmonton–Calgary corridor: Simulating future land
675 cover change under potential government intervention. *Land Use Policy* 63:356-368 DOI
676 <https://doi.org/10.1016/j.landusepol.2017.01.039>.
- 677 Sun G, Zhou G, Zhang Z, Wei X, McNulty SG, Vose JM. 2006. Potential water yield reduction
678 due to forestation across China. *Journal of Hydrology* 328(3-4):548-558 DOI
679 <https://doi.org/10.1016/j.jhydrol.2005.12.013>.
- 680 Tambosi LR, Martensen AC, Ribeiro MC, Metzger JP. 2014. A framework to optimize
681 biodiversity restoration efforts based on habitat amount and landscape connectivity. *Restoration*
682 *Ecology* 22(2): 169-177 DOI <https://doi.org/10.1111/rec.12049>.
- 683 Teixeira AM, Soares-e-Filho BS, Freitas S, Metzger JPW. 2009. Modeling landscape dynamics in
684 the Atlantic Rainforest domain: implications for conservation. *Forest Ecology and Management*
685 257:1219e1230 DOI 10.1016/j.foreco.2008.10.011.
- 686 Torrens PM. 2006. Geosimulation and its application to urban growth modeling. Springer-Verlag:
687 London, U. K.
- 688 United Nations (UN) 2015. Transforming our World: The 2030 Agenda for Sustainable
689 Development Available at
690 [https://sustainabledevelopment.un.org/content/documents/21252030%20Agenda%20for%20Sust](https://sustainabledevelopment.un.org/content/documents/21252030%20Agenda%20for%20Sustainable%20Development%20web.pdf)
691 [ainable%20Development%20web.pdf](https://sustainabledevelopment.un.org/content/documents/21252030%20Agenda%20for%20Sustainable%20Development%20web.pdf)
- 692 United States Geological Survey. Available at: <http://glovis.usgs.gov>.
- 693 Vázquez-Quintero G, Solís-Moreno R, Pompa-García M, Villarreal-Guerrero F, Pinedo-Alvarez
694 C, Pinedo-Alvarez A. 2016. Detection and projection of forest changes by using the Markov Chain
695 Model and cellular automata. *Sustainability* 8(3):236 DOI 10.3390/su8030236.
- 696 Velázquez A, Mas JF, Díaz Gallegos JR, Mayorga Saucedo R, Alcántara PC, Castro R, Fernández
697 T, Bocco G, Ezcurra E, Palacio JL. 2002. Patrones y tasas de cambio de uso del suelo en
698 México. *Gaceta ecológica* (62).
- 699 Weiskittel AR, Crookston NL, Radtke PJ. 2011. Linking climate, gross primary productivity, and
700 site index across forests of the western United States. *Canadian Journal of Forest Research* 41(8):
701 1710-1721 DOI <https://doi.org/10.1139/x11-086>.
- 702 Ximenes AC, Almeida CM, Amaral S, Escada MIS, Aguiar APD. 2011. Spatial dynamic
703 modelling of deforestation in the Amazon. In Cellular Automata-Simplicity Behind Complexity.
704 *InTech* DOI 10.5772/16137.
- 705 Yanai AM, Fearnside PM, de Alencastro Graça PML, Nogueira EM. 2012. Avoided deforestation
706 in Brazilian Amazonia: simulating the effect of the Juma Sustainable Development
707 Reserve. *Forest Ecology and Management* 282:78-91 DOI
708 <https://doi.org/10.1016/j.foreco.2012.06.029>.

Table 1 (on next page)

Scenes characteristics

Sensor	Date	Characteristics
Landsat TM 5	1990	7 spectral bands, 30 m resolution
Landsat TM 5	2005	7 spectral bands, 30 m resolution
Landsat OLI 8	2017	8 spectral bands, 30 m resolution; 1 panchromatic band 15 m resolution

1 TM= Thematic Mapper, OLI= Operational Land Imager

2

Table 2 (on next page)

Land use/land cover types determined through the supervised classification method

Land use and land cover	Acronym	Description
Primary forest	PF	Forest fully covered with canopy
Secondary forest	SF	Forest partially covered with canopy
Human settlements	HS	Residential areas
Areas without vegetation	AWV	Areas without vegetation, agriculture areas or induced grasslands
Water bodies	WB	Water bodies

1

Table 3(on next page)

Variables feeding the deforestation model

No	Variable type	Name	Unit	Acronym
1	Density	Density of main roads	m ² /Km ²	Denmr
2		Density of secondary roads	m ² /Km ²	Densr
3		Density of main streams	m ² /Km ²	Denms
4		Density of secondary streams	m ² /Km ²	Denss
5		Density of rural settlements	m ² /Km ²	Denrs
6	Proximity	Distance to sawmills	m	Diss
7		Distance to water bodies	m	Diswb
8		Distance to main roads	m	Dismr
9		Distance to secondary roads	m	Dissr
10		Distance to main streams	m	Disms
11		Distance to secondary streams	m	Disss
12		Distance to rural settlements	m	Disrs
13		Distance to urban settlements	m	Disus
14		Distance to mines	m	Dism
15		Distance to areas without apparent vegetation	m	Disawav
17	Topographic	Altitude	m	Alt
18		Slope	°	Slop
19		Topographic position index	Dimensionless	TPI

Table 4(on next page)

Transitions of land use/land cover

		To				
From		PF	SF	HS	AWV	WB
	PF		✓	✓	✓	
	SF				✓	
	HS					
	AWV					
	WB					

1

Table 5(on next page)

Area occupied for five types of land uses during 1990, 2005 and 2017, and rate of change for the periods 1990-2005 and 2005-2017.

Land Use	Occupied area (Ha)			Occupied area (%)			Rate of change	
	1990	2005	2017	1990	2005	2017	1990-2005	2005-2017
AWV	20444.18	23828.59	24101.92	4.11	4.79	4.85	8.33	8.43
SF	199121.38	223948.16	286922.04	40.05	45.05	57.72	8.03	10.68
HS	154.60	272.35	521.65	0.03	0.05	0.10	12.58	15.96
WB	26.9712	103.76	153.91	0.01	0.02	0.03	27.48	12.36
PF	277380.46	248973.97	185427.79	55.80	50.08	37.30	6.41	6.21

AWV=Areas without vegetation, SF= Secondary forest, HS= Human settlements, WB=Water bodies, PF= Primary forest.

Table 6(on next page)

Land use/land cover change dynamics

Land use	Difference 1990-2005 (ha)	Difference 2005-2017(ha)	Overall Difference (ha)	Type of change	1990-2005 (ha)	2005-2017 (ha)
AWV	3384.40	273.34	3657.74	Deforestation	3120.40	7283.95
SF	24826.78	62973.88	87800.66	Degradation	54455.78	73904.27
HS	117.74	249.31	367.05	Other	76.22	219.41
WB	76.79	50.15	126.94	Recovery	27128.71	20204.13
PF	-28406.49	-63546.18	-91952.66	--	--	--

AWV=Areas without vegetation, SF= Secondary forest, HS= Human settlements, WB=Water bodies, PF= Primary forest.

Table 7 (on next page)

Transition matrix of probability for land use/land cover change (1990-2005, 2005-2017, 1990-2017)

	Periodo	AWV	PF	HS	WB	PF
AWV	1990-2005	0.9000	0.0250	0.0250	0.0250	0.0250
	2005-2017	0.6250	0.3504	0.0108	0.0029	0.0109
	1990-2017	0.6615	0.3124	0.0120	0.0035	0.0106
SF	1990-2005	0.0222	0.7516	0.0008	0.0005	0.2248
	2005-2017	0.0557	0.8116	0.0004	0.0000	0.1323
	1990-2017	0.0654	0.7945	0.0012	0.0006	0.1384
HS	1990-2005	0.0452	0.0645	0.8806	0.0000	0.0097
	2005-2017	0.0557	0.2479	0.6959	0.0000	0.0004
	1990-2017	0.0651	0.0774	0.8575	0.0000	0.0000
WB	1990-2005	0.0000	0.1254	0.0000	0.8553	0.0193
	2005-2017	0.0095	0.1684	0.0000	0.8030	0.0191
	1990-2017	0.0000	0.1868	0.0000	0.7957	0.0175
PF	1990-2005	0.0020	0.2865	0.0000	0.0000	0.7115
	2005-2017	0.0056	0.3798	0.0003	0.0000	0.6144
	1990-2017	0.0071	0.4419	0.0002	0.0000	0.5508

AWV=Areas without vegetation, SF= Secondary forest, HS= Human settlements, WB=Water bodies, PF= Primary forest.

Table 8(on next page)

Percentage of surface area occupied by five land use/land cover types and rate of change for 2017-2050 based on three scenarios

Land Use	Occupied surface area (%)				Change rate		
	2017	2050s	2050o	2050p	2017-2050s	2017-2050o	2017-2050p
AWV	4.848	5.275	5.017	7.695	3.40	3.23	4.96
SF	57.716	73.721	61.863	83.628	3.99	3.35	4.53
HS	0.105	0.105	0.105	0.105	3.13	3.13	3.13
WB	0.031	0.031	0.031	0.031	3.12	3.13	3.12
PF	37.300	20.868	32.983	8.541	1.75	2.76	0.72

AWV=Areas without vegetation, SF= Secondary forest, HS= Human settlements, WB= Water bodies, PF= Primary forest, S= Stationary, O= Optimistic, P= Pessimistic.

Table 9(on next page)

Land use/land cover change dynamics (ha) under three proyected scenario.

Land use	2017-2050 _s	2017-2050 _o	2017-2050 _p
AWV	2121.97	840.44	14150.28
SF	79565.57	20617.02	128818.00
HS	0.87	1.25	1.05
WB	0.46	0.32	0.10
PF	-81688.00	-21459.04	-142969.31

1 AWV=Areas without vegetation, SF= Secondary forest, HS= Human settlements, WB= Water bodies, PF= Primary
2 forest, S= Stationary, O= Optimistic, P= Pessimistic.
3

Figure 1

Location and elevations of the study area

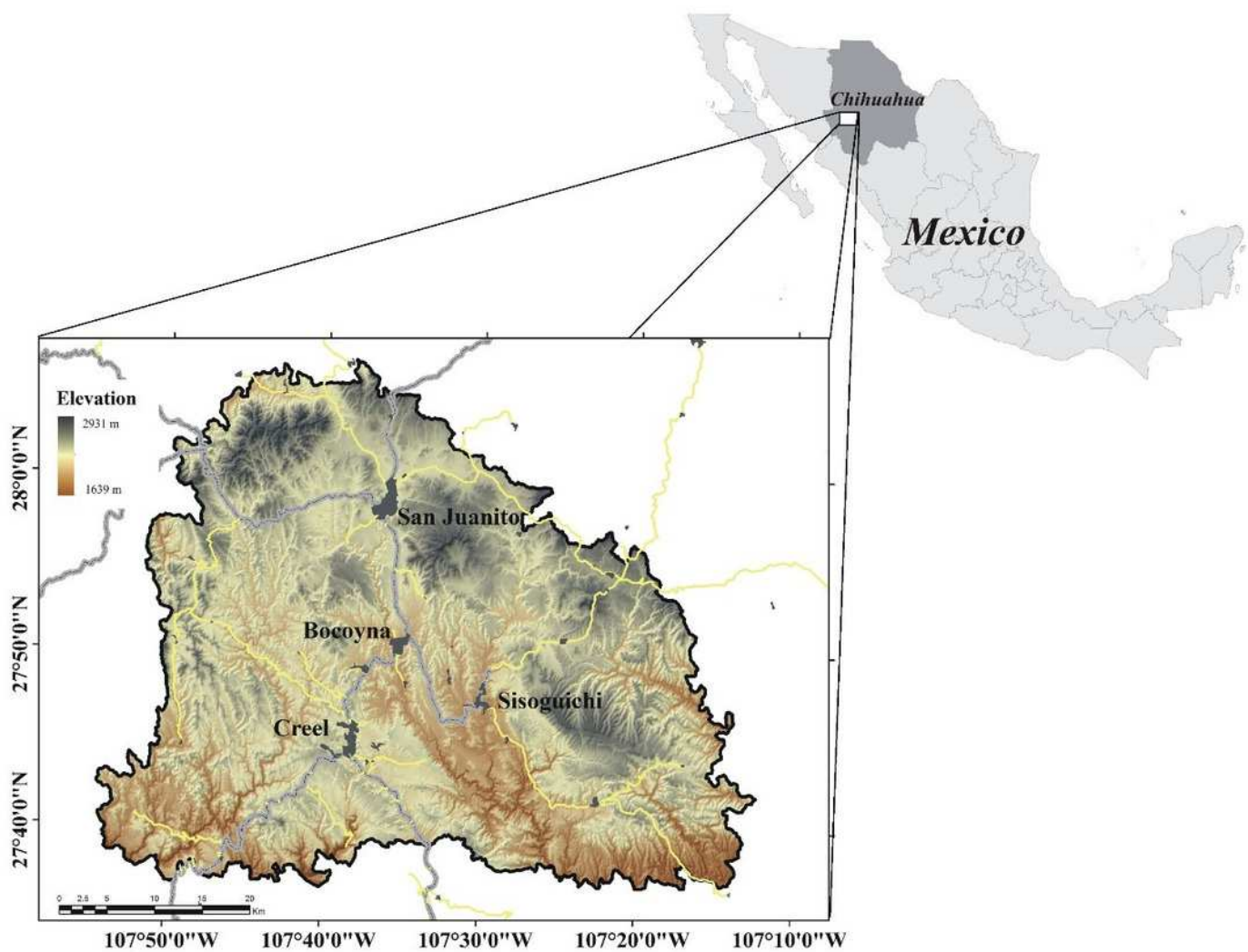


Figure 2

Flowchart of the methodological procedure followed to produce the proposed scenarios. Abbreviations: TM: Thematic Mapper, OLI: Operational Land Imager, WoE: Weights of Evidence, LUCC: Land use and cover change

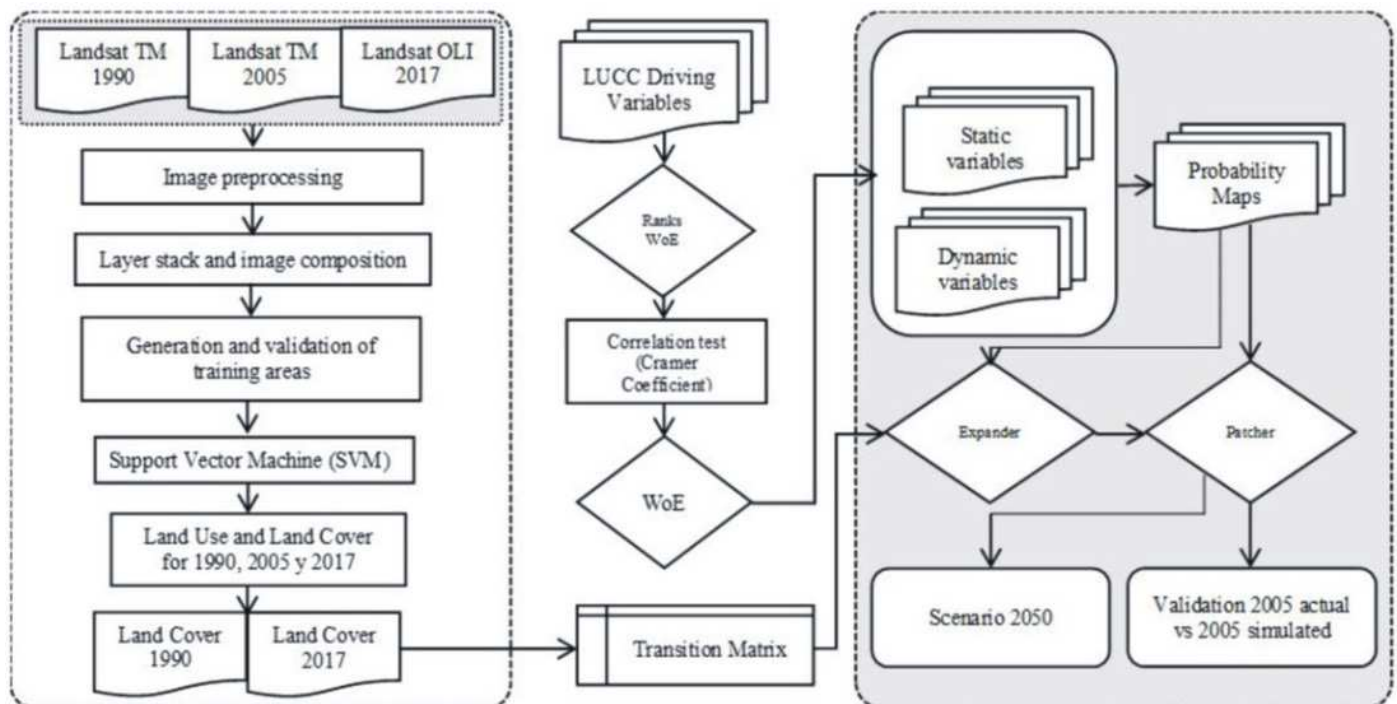


Figure 3

Land use/land cover of 1990 (a), 2005 (b), 2017 (c), changes during 1990-2005 (d) and changes during 2005-2017 (e). Abbreviations: AWW: areas without vegetation, SF: secondary forest, WB: water bodies, HS: human settlements and PF: primary forest.

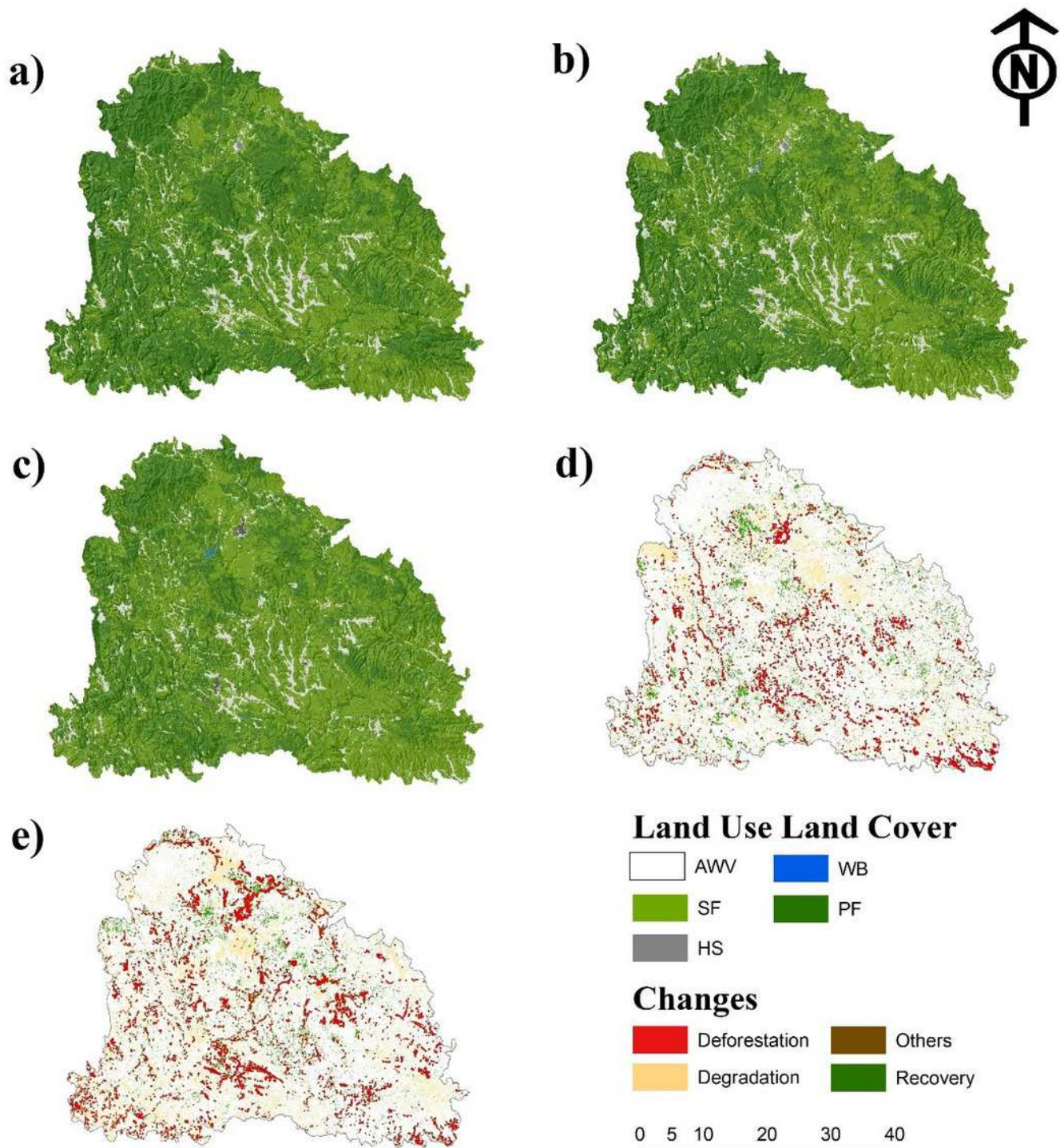


Figure 4

Variation of the FSI as a function of different distance

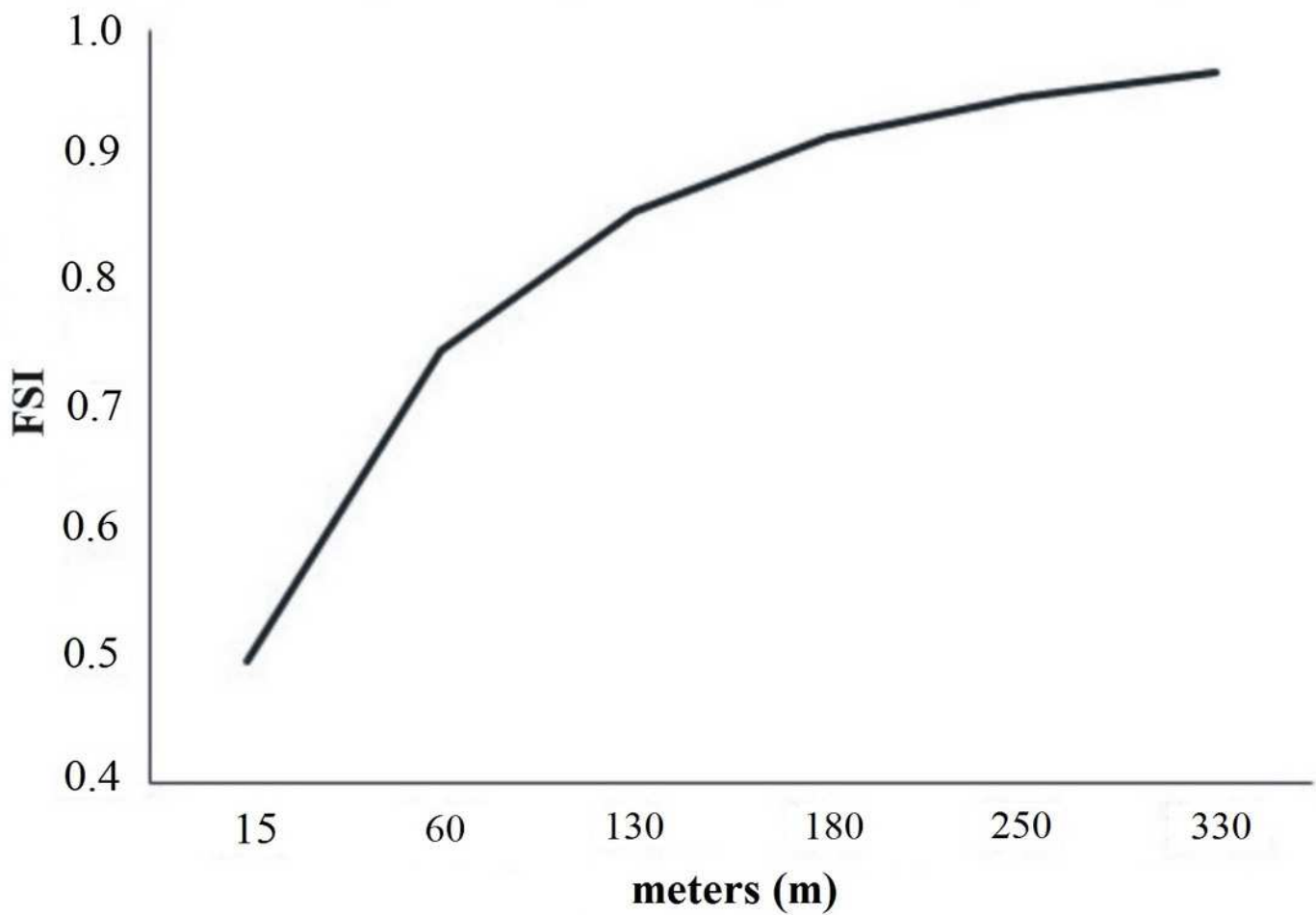


Figure 5

a) Land use/land cover of 2017 and simulated land use/land cover projected for the year 2050 as a result of the b) Stationary, c) Pessimistic and d) Optimistic scenarios.
Abbreviations: AWW: areas without vegetation, SF: secondary forest, WB: water bodies

