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Modeling the impacts of agricultural best management practices on runoff, sediment, and crop yield in an agriculture-pasture intensive watershed

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Abstract
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**Key words:** Watershed, Sediment; Runoff; Crop yield; Conservation; SWAT model; Oklahoma

**Introduction**

Sediments, originating from land use activities including farming and urbanization, constitute one of the major non-point source (NPS) pollutions and have impaired water bodies, reduced reservoir capacity and lifespan, threatened drinking water supply, increased water treatment cost, and reduced the overall ecosystem health globally (Abdulkareem et al., 2018; Falconer et al., 2018, FAO, 2013; Simon and Klimetz, 2008; Palmieri et al., 2001). In the United States of America (USA), more than 50% of water bodies are NPS impaired, with sediment ranking the sixth among the leading causes of water quality impairments (United States Environmental Protection Agency, 2016).

The Great Plains region, characterized by highly intensive agricultural production system in the USA, is subject to water quality issues mostly due to agricultural NPS pollution (Osteen et al., 2012). To reduce agricultural NPS pollution, several management practices including conservation tillage system, are encouraged and adopted in the region. For example, Great Plains Conservation Program provides financial and technical assistance as water quality protection activities in the Great Plains states to ranchers and farmers who adopt total conservation treatment of their entire operation. This approach has increased soil organic carbon in the Great Plains (Lewis et al., 2018). It was reported that replacing just about 10–23% of conventional tillage system to conservation tillage system in Great Plains, could save one billion tons of soil on highly erodible lands (Bernard et al., 1996). Despite the ongoing efforts in reducing agricultural NPS pollution, the water quality issues still persist in the region.

**BMPs for sediment load reduction**

Several studies evaluated the effectiveness of various BMPs in reducing sediment loads from agricultural fields. For example, Zhang and Zhang (2011) reported that the use of sediment ponds as BMPs reduced up to 54-85% sediment from field runoff in Orestimba Creek Watershed, California. Lam and Fohrer (2011) found that the implementation of BMPs related to extensive land use management, grazing management practice, field buffer strip, and nutrient management plan reduced sediment load by 0.8% to 4.9% in a North German lowland catchment. Rousseau et al. (2013) applied vegetated riparian buffer strips, precision slurry application, grassland conversion of cereal and corn fields, and no-till corn in Beaurivage River watershed, Quebec, Canada and found that riparian buffer strips and grassland conversion were highly effective in reducing sediment yield compared to other BMPs. Maharjan et al. (2017) tested three BMPs including split fertilizer application, winter cover crop cultivation, and a combination of the two BMPs in the Haean catchment, South Korea and found that the
combination of split fertilizer application and cover crop cultivation resulted the highest positive
effect in terms of reduced sediment and nitrate loads and increased crop yield. Teshager et al.
(2017) analyzed fourteen scenarios based on systematic combinations of five BMP strategies:
fertilizer/manure management, changing row-crop land to perennial grass, vegetative filter strips,
cover crops and shallower tile drainage systems, in the Raccoon River watershed in west-central
Iowa, USA. Their findings suggest that planting switchgrass in half of the watershed would
reduce the sediment load by up to 67% and meet the drinking water standard. Yang et al. (2009)
estimated about 51.8-71.4% reduction in sediment loads from the Black Brook Watershed in
northwestern New Brunswick, Canada with the implementation of flow diversion terraces.
In this study, we evaluated different agricultural best management practices (BMPs) and
estimated changes in sediment load, surface runoff and crop yield. For this, we selected a rural
agricultural watershed, Fort Cobb Reservoir watershed, located in southwestern Oklahoma,
USA. This watershed is reported to have water quality issues related to sediment, despite of
BMP implementation in most parts of the watershed for years. Therefore, this study area
provides a good site to evaluate how sediment loads alter with the selection and placement of
BMPs in the watershed.
In the Fort Cobb Reservoir watershed, several BMPs such as contour and strip farming, terraces,
conversion of crop land to Bermuda pasture, reduced till and no-till farming, drop structures,
shelter belts, flood retarding structures, etc. have been currently implemented with about 50% of
the cropland under conservation tillage or minimum disturbance tillage (Garbrecht and Starks,
2009). Although hydrological modeling studies of this watershed are available (Storm et al.,
2003; Moriasi et al., 2007, 2008; Mittelstet, 2015), these studies neither included existing BMPs
in their studies nor assessed the effectiveness of these BMPs on water quality. The Oklahoma
Department of Environmental Quality recommended a conversion of 50% of the cultivated area
in the watershed to no-till practices to control sediment and nutrient loads (Oklahoma
Conservation Commission, 2015). Osei et al. (2012) compared the effects of no-till systems on
wheat yield with other tillage systems and found that no-till would be more profitable than
conventional tillage or the current mix of tillage practices in the watershed. On contrary, the
continuous no-till practice showed decreased wheat yield (Decker et al., 2009; Patrignani et al.,
2012), which could be due to increased risk of weeds and diseases cycles associated with wheat
production (Edwards et al., 2006). To the best of our knowledge, there are no studies available in
the study watershed that estimated the impact of rotation of no-till winter wheat with other viable
crops on sediment loads reduction.
Therefore, in this study we estimated the effectiveness of different possible BMPs to reduce
sediment loads while increasing the crop yield. To this end, first, a SWAT hydrological model
was developed and calibrated and validated based on streamflow, sediment, and crop yield data.
Then, the effectiveness of these BMPs was estimated targeted at sediment reduction and
maximization of crop yields.

Materials & Methods
A hydrological model of Fort Cobb Reservoir watershed that includes the target study area of Five-Mile Creek sub-watershed was developed using the Soil and Water Assessment Tool (SWAT) modeling framework (Arnold et al., 1998). The model was calibrated both automatically and manually. The verified model was then used to study the impacts of agricultural BMPs on hydrology, sediment, and crop yield of major crops. The steps used in this study as illustrated in Fig. 1 are explained in the sections below.

**Study area**

The selected study area is Five-Mile Creek sub-watershed (FMC) located within Fort Cobb Reservoir watershed in southwestern Oklahoma (Fig. 2). FMC has an area of 113.05 km² with land uses comprised of 50% cropland, 41% pastureland and 9% others. The major crops in FMC include 30% winter wheat, 16% cotton (dryland 3.5%, irrigated 12.5%), and grain sorghum (1.5%). The Five-Mile Creek is one of the four tributaries of the Fort Cobb Reservoir (Fig. 2). The reservoir water quality has been of concern for decades and is included in the impaired and threatened waters, 303(d) list, because of high levels of sedimentation, phosphorous, nitrogen, bacteria, and ammonia caused primarily by intensive agriculture and pastoral activities (Oklahoma Conservation Commission, 2009; Oklahoma Department of Environmental Quality, 2014). The 303(d) list comprises those waters that are in the polluted water category, for which beneficial uses like drinking, aquatic habitat, industrial, recreation and use are impaired by pollution. Despite several additional BMPs being implemented, the issues of sedimentation still exist in the study area.

Figure 2. Five-Mile Creek sub-watershed (FMC) located within the Fort Cobb Reservoir watershed. Land types, slope and soil classes and elevation maps are included.

**Hydrological model**

The Soil and Water Assessment Tool (SWAT) was employed to construct a hydrological model of the study area using the gaging station (USGS 07325800) of the United States Geological Survey as a watershed outlet. This station is the only available monitoring station with continuous records of streamflow. It receives runoff from two sub-watersheds- Cobb Creek and FMC sub-watersheds (Fig. 2). A ten-meter Digital Elevation Model was used for watershed delineation, stream network creation and topographic information. The study area was divided into spatially related 43 sub-basins with an average area of 8 km² (0.2 28 km²). The watershed topography was grouped into four slope classes of 0-2%, 2-4%, 4-6%, and >6%. Existing waterbodies including ponds in the watershed were obtained from U.S. Fish and Wildlife Service (2014) and modeled these waterbodies as ponds in each sub-basin (Appendix A). The SSURGO soil database (Soil Survey Staff, 2015), the finest resolution soil data available, was used to
define soil attributes in the watershed (Appendix B). The land use data were obtained from the 2014 crop data layer (USDA-NASS, 2014). The cultivated land cover types were further separated into irrigated and non-irrigated lands based on the locations of the center pivot irrigation circles. These locations were identified from the 2014 one-meter resolution aerial images (https://datagateway.nrcs.usda.gov/). We found 30 pivot circles encompassing 13.7 km² (12.1%) of irrigated land dedicated for cotton production in the FMC sub-watershed. An Overlay of land use, soil and slope with respective SWAT threshold percentages of 10% for land, 10% for soil and 20% for slope in each sub-basin resulted into 15,217 Hydrologic Response Units (HRUs). An HRU in SWAT captures watershed diversity by combining similar land, soil and slope areas in each sub-basin. In SWAT, loadings of water, sediments, and crop yield are calculated first at HRU level, summed at each sub-basin and then routed to the watershed outlet. These HRUs were assigned agricultural BMPs (conservation tillage, no-till, contouring, crop rotation, and conversion to pasture - Bermuda grass) that are most commonly practiced in the study area. Existing contour in the study watershed were identified by using aerial photographs (Barber and Shortridge, 2005). The broken terraces were recognized using two-meter LiDAR drainage lines from satellite imagery (https://gdg.sc.egov.usda.gov/Catalog/ProductDescription/LIDAR.html). The HRUs with more than 65% contour were classified as being terraced with contour farming. It was found that 8 km² of FMC were terraces and contour without breaking, which modeling existing terraces and contours resulted into 28% reduction in sediment. Information about tillage type and fertilizer application for the selected crops was obtained from relevant literature (Storm et al, 2006; ODEQ, 2006) and consultation with local Oklahoma State University Cooperative Extension Service and Conservation District personnel (Appendix C.1-9). Additionally, cattle information including cattle stocking rate (0.5 head/ha), consumed biomass (3 kg/ha/day), trampled biomass (0.47 kg/ha/day) and deposited manure (1.5 kg/ha/day) were obtained from other sources (USDA-NASS, 2012; Storm et al., 2006) and used in the model.

The current climate pattern (1982-2016) in the watershed was represented by six climate variables: precipitation, minimum temperature, maximum temperature, solar radiation, relative humidity and wind speed. The climate data at daily scale were collected from a combination sources including the USDA Agricultural Research Service (USDA-ARS) (http://globalweather.tamu.edu/), the Oklahoma MESONET (https://www.mesonet.org/). Between 1982 and 2016, the study area received 2.2 mm/day precipitation with daily average temperature (15.8°C), solar radiation (16.9 MJ/m²), relative humidity (0.6 fractional), and wind speed (4.3 m/s).

Model calibration and validation
First, the model was calibrated manually to improve the model performance based on operation management parameters and associated cropping schedules and then automated iterative calibration was performed using SWAT-CUP tool (Abbaspour et al., 2007) for three important
components: streamflow, sediment, and crop yield. Crop operation management parameters and
associated cropping schedules were adjusted manually. Model sensitivity was tested prior to
model calibration to determine the most sensitive parameters. Observed data on streamflow, crop
yields and sediment loads from 1990 to 2010 were used for model calibration and validation.
Three different statistical matrices- coefficient of determination ($R^2$), Nash-Sutcliffe efficiency
(NSE) and percent bias (Pbias) were used to evaluate the model performance.

**Streamflow**

Monthly streamflow observed at the USGS gaging station- Cobb Creek near Eakely gage (USGS
07325800) for a ten-year period (1991–2000) was used for model calibration. Prior to model
calibration, the sensitivity of the model to streamflow was tested in SWAT-CUP for 17
parameters. The p-value and t-state indicators were used to identify the most sensitive parameters
in the watershed. The smaller the p-value and the larger the absolute value of t-state, the more
sensitive the parameter is. The six parameters related to water balance: Curve number (CN), soil
evaporation compensation factor (ESCO), groundwater delay (GW_DELAY), deep aquifer
percolation fraction (RCHRG_DP), Manning’s n value for the main channel (CH_N2), and
available water capacity of soil layer (SOL_AWC) were found to be the most sensitive
(Appendix D), similar to what other studies found (Moriasi et al., 2008; Storm et al., 2006).
According to Moriasi et al. (2015), model performance can be judged “satisfactory” for flow
simulations if daily, monthly, or annual $R^2 > 0.60$, NSE > 0.50, and $PBIAS \leq \pm 15\%$ for
watershed-scale models. The model was calibrated satisfactorily for streamflow with values of
$R^2$ (0.64) and NSE (0.61) and Pbias (5.1%) (Fig. 3). The validation of the model with an
independent set of monthly observed streamflow at the same gage station for a different ten-year
period (2001–2010) indicated a robust model performance with values of $R^2$ (0.79) and NSE
(0.62) and Pbias (-15%) (Fig. 3). Calibrated parameters and their final value ranges are listed in
Table 1.

**Sediment**

Suspended sediment was calibrated for ten years (1991–2000) and validated for another ten years
(2001–2010) at the watershed outlet. For this, grab suspended sediment sample data that were
available from 2004 to 2012 (usually 1 to 3 samples per month with a few months missing) was
used. This grab sample data provided us an opportunity to estimate sediment loads for the time
period that lacked observations using sediment rating curve method as suggested by Horowitz
(2003). This method is a regression relationship between the observed streamflow and sediment
data used popularly to generate sediment information for missing period in many modeling
studies (Salimi et al., 2013; Shabani, 2012; Jothiprakash and Grag, 2009; Sarkar et al., 2008;
Gray and Simoes, 2008). A strong correlation ($R^2=0.9$) between the observed grab sample
sediment data and runoff in the study watershed (Fig. 4) was observed. This regression
relationship was used to estimate the missing sediment data for the model simulation period. Then these data were used to calibrate the model by modifying ten parameters that were related to sediment load (Storm et al., 2006; Moriasi et al., 2008). The model calibration with values of $R^2$ (0.30), NSE (0.35) and Pbias (<20%), (Fig. 5) and validation with values of $R^2$ (0.33), NSE (0.43) and Pbias (<55%) (Fig. 5) was considered acceptable. Calibrated parameters and their final value ranges are listed in Table 1.

We found that the largest errors in sediment prediction were associated with errors of peak flow estimation. This could be due to the “second storm effect” problem in hydrological models, including SWAT (Abbaspour et al. 2007). The first storm event causes a larger sediment transport and makes remaining surface layers difficult to mobilize. As a result, the second and third storm events regardless of their event sizes, will result in smaller sediment loads. For this study area, the “second storm effect” was not tested since there were no observed sediment data representing flood events (May 1993, June 1995, June 2007) during model calibration-validation period. The simulated sediment data failed to accurately capture these events, resulting uncertainty in sediment calibration. The over-and under-estimation of sediment during flood events was reported in other SWAT based studies (Oeurng et al., 2011).

Figure 4. Observed daily discharge and observed daily suspended sediment concentration trend

Table 1. Streamflow and sediment calibration parameter values in study area

Figure 5. Calibration and validation monthly time series (2000–2010) for observed and SWAT simulated suspended sediment concentration at the Cobb Creek near Eakley, Oklahoma gauging station

Crop yield

Crop yield and biomass production affect watershed hydrology through altered erosion and water balance (Hu et al. 2007; Ng et al. 2010a; Andersson et al. 2011; Nair et al. 2011). A combination of the Oklahoma State University variety trial data from 2001 to 2016 (http://croptrials.okstate.edu/), and the county level data (1986–2005) obtained from the USDA National Agricultural Statistics Service were used to calibrate yield of crops (winter wheat, grain sorghum, cotton- both dry and irrigated) (USDA-NASS, 1986 to 2005, http://digitalprairie.ok.gov/cdm/ref/collection/stgovpub/id/11177). The variety trial crop yields were collected from sites in seven Oklahoma counties (Apache, El Reno, Homestead, Chickasha, Altus, Tipton, and Thomas) that are located within and nearby the study area. A list of crop yield parameters with their initial and calibrated values is provided in Appendix E.1 and E.2. In this study the PB was used as an indicator to compare the SWAT simulated yield with the observation. Ten crop model parameters were selected (Appendix E.1 and E.2) and their associated value ranges were set based on recommendation made by Sinnathamby, et al., (2017) and Nair et al. (2011). The values were then manually adjusted until the PB for the crop models
reached satisfactory values for cotton (-4.5%), grain sorghum (-27.3%) and winter wheat (-6.0%) over the years 1986 to 2010 (Figure 4).

Agricultural Best Management Practices Scenarios

Studies identified sedimentation as one of the water quality issues in the region with the associated ecological and economic impacts (Zhang et al., 2015). As a result, various agricultural BMPs have been implemented in the watershed to abate sediment loading and transport (Becker, 2011). Despite these efforts, there are still soil erosion problem in agricultural fields causing degraded water quality.

Often, conservation tillage and no-till practices can be employed to improve the success of new cropping systems and help assure the sustainability of the land. No-till cropping systems in Oklahoma have proved important resources for the economic viability of producers and landowners operations (Malone, J., 2008). Conversion to no-till practices on at least half of the cultivated area in the study watershed was one of the recommendations to reduce sediment and nutrient loadings for this Watershed (OCC, 2015). Conservation practices such as contour and strip farming, terraces, conversion of crop land to Bermuda pasture, reduced till and no-till farming, drop structures, shelter belts, flood retarding structures, etc. have been currently implemented in the study region as the effective BMPs for mitigating NPS pollution (Garbrecht and Starks, 2009). However, records detailing types and time of installation of these management practices prior to the 1990s are not readily available in either the state offices of the Natural Resources Conservation Service (NRCS) or the local conservation districts. According to Garbrecht and Starks (2009), 80%-90% of cropland in the study area that needed terraces, has been terraced over the last 50 years. Over the last decade, about 50% of the cropland was in conservation tillage or minimum disturbance tillage. In addition to these management practices, gully reshaping and grad stabilization structures were implemented by conservation funds. Other conservation practices have been implemented without cost sharing assistance. Also, some selected channel bank sections were stabilized and some channels have been fenced to prohibit cattle from eroding banks, small impoundments were constructed, and a number of gravel roads were paved to control cropland erosion in this watershed. Despite these efforts, there are issues of NPS pollution in the region. Therefore, we developed five scenarios that reflect the commonly used agricultural BMPs in the study area and throughout the Great Plains region (Table 2). These BMPs included practices of conservation tillage and no till on both contouring and no-contouring along with the rotation of winter wheat with other crops. The BMPs were applied to three major crops- cotton, grain sorghum and winter wheat. Because of weed and disease problems associated with continuous no-till wheat, wheat was rotated/cover cropped with canola, cotton and grain sorghum. A combination of land use and these five scenarios resulted into 22 SWAT model simulations. In scenarios 1–4, the study area was simulated for one crop at a time by converting all crops into one (for example, all crops converted to wheat and so on). In scenario 5, all the cropland in the study area was converted to Bermuda grass because of its popularity in the study watershed (Moriiasi et al. 2008).
Table 2. Agricultural Best Management Practices (BMPs) scenarios simulated for, cotton, grain sorghum and winter wheat to evaluate their impacts on hydrology, sediment and crop yield in the study area.

Results

Surface runoff

All five scenarios except for S3 with wheat-cotton and wheat-canola rotations and cotton in S1 and S3 decreased surface runoff compared with the baseline scenario (Fig. 6). When contouring was applied in conservation tillage (S2), surface runoff reduced by 18.4% for cotton and grain sorghum and by 19.2% for winter wheat. Similarly, implementation of contouring on the existing no-till BMP (S4) led to surface runoff reduction by 18.4% (cotton and grain sorghum) and 19.4% for wheat compared to the no-till BMP (S3). Between the three major crops in scenarios 1 to 4, grain sorghum was the least runoff generator followed by winter wheat and cotton. When all crops were converted to Bermuda grass (S5) surface runoff reduced by 31.7% as compared to rest of the scenarios. Application of different grazing operations and stocking rates in S5 resulted virtually the same runoff generation (37.96 to 38.08 mm) with less than one-third of a percentage point difference between them. Of the 22 combinations of agricultural BMPs simulated in all five scenarios, wheat rotated with cotton under no-till resulted in the highest runoff followed by wheat rotated with canola. We found that there was virtually no change in surface runoff between the conservation and no-till systems. But, the implementation of contouring reduced surface runoff in both conservation and no-till systems.

Figure 6. Changes in surface runoff generation under different scenarios of Best Management Practices.

Sediment

We found that implementation of contouring on conservation tillage (S2) and on no-till (S4) reduced sediment loss nearly by half (Fig. 7 and Table 3). Between all 22 combinations of BMPs, cotton was the lead contributor to sediment. For cotton, contouring on no-till practice generated the least sediment (1.27 tons/ha) while the conservation tillage with no contouring released most sediment (3.01 ton/ha). Wheat’s contribution to sediment loss was as half as that of grain sorghum and one-fourth of that of cotton (S1–S4). Wheat, under the conservation tillage with contour (S2), was the least contributor of sediment (0.4 ton/ha). Rotating wheat with canola was found to be the most effective in controlling sediment loss under no-till system with only 0.87 ton/ha loss as compared to wheat as a cover crop for cotton (2.0 ton/ha) and grain sorghum.
Converting all crops to Bermuda grass pasture with combinations of different grazing time and stocking rate (S5) released only 0.10 to 0.12 ton/ha sediment. We found virtually no difference in simulated sediment loss between the combination of grazing timings and stocking rates applied.

Table 3. Sediment reduction in percentage as a result of contouring on conservation tillage and no-till practices for cotton, grain sorghum and winter wheat

In the business-as-usual baseline scenario (BL), the four out of 11 sub-basins (#7, 15, 16, 18) generated sediments at an average of 1.2–1.5 ton/ha (Fig. 8.a). These four sub-basins have erosive soil texture (fine sandy loam and silty clay loam) with wheat (28.5%) and cotton (18.5%) as major crops. The amount and location of sediment loadings varied between the scenarios. For example, 90% of sediment load was reduced when the crops were converted to Bermuda grass (Fig. 8 b), while the sediment load was increased by 76% and 135% with cotton under no-till and under conservation tillage respectively (Fig. 8 e-h).

Figure 8. Sub-basin level SWAT simulated sediment loadings (tons/ha) in Five-Mile Creek sub-watershed under different BMP scenarios

**Crop yield**

We found no significant effect of contouring and tillage systems on the simulated yields of cotton, grain sorghum and winter wheat. However, we found differences in yields of these crops when they were used as cover crop or in rotation. For example, under the no-till practice, the yield of grain sorghum when wheat was used as a cover crop decreased by 28.4% (S3) and once there was no-till plus contour farming it decreased by 14.8% (S4). It was found that covering/rotation with winter wheat resulted into reduced yield for both cotton and grain sorghum regardless of contouring (S3 and S4). When covering/rotated with winter wheat, cotton yield decreased by 52% with or without contouring while grain sorghum yield decreased by 28.4% (no contour) and by 14.8% with contour (S3 and S4). This decreased yield is attributed to the presence of wheat residues and lack of available soil moisture for the second crop. We found that cotton yield decreased more than that of grain sorghum when wheat was used as a cover crop. We found virtually no effect of stocking rate and grazing start months on pasture yield (Fig. 9).

Figure 9. Crop yields under different scenarios of Best Management Practices

**Discussion**

Five Mile Creek is one of the main contributing sub-watersheds of the Fort Cobb Reservoir watershed. It is a typical example of agriculture-pasture intensive watershed in the US Great Plains that may present a test bed for simulating the impacts of agricultural activities in
combination with various BMPs on crop yield, water quality and quantity. In order to reduce
erosion in the Fort Cobb Reservoir watershed, several BMPs and conservation measures
including terraces, changing cropping patterns, and progressive adoption of no-till and
conservation tillage systems among others have been implemented (OCC, 2014). There are
conservation programs with financial and technical assistance available to install new tillage or
cropping systems in the study region (USDA, Farm Service Agency, 2016). Some farmers have
converted the highly erosive parts of their crop land to Bermuda grass pastureland (USDA-FSA,
2015). These initiatives reduced sediment loadings by three to five times as compared to the time
prior to 1963 (Zhang et al. 2015). Garbrecht et al. (2008) stated that there was substantial
reduction in sediment yield in the Five Mile Creek sub-watershed in the second half of the 20th
century mainly due to conversion of cropland to pasture land. However, the sediment loads in the
study area are still high and need to be reduced (ODEQ, 2014). Therefore, in this study, we
evaluated the effectiveness of agricultural BMPs on surface runoff, and crop and sediment
yields.

Impacts of contouring and tillage on runoff, sediment and crop yield:
Contouring and terracing are popularly used practices to control erosion in the study region
(Garbrecht, et al., 2009). We found that contouring with either conservation tillage or no-till
farming prevented sediment yield by almost half while the surface runoff was reduced by at least
18% in the watershed. Compared to the conservation tillage practice, no-till farming decreased
sediment yield by 25.3% and 9.0% for cotton and grain sorghum respectively. In several other
watersheds, no-till practice was found to generate less sediment yield (Dickey et al., 1983; Olson
et al., 2010; Parajuli et al., 2013; Sharpley and Smith, 1994). We found virtually no difference in
surface runoff and yields of cotton and grain sorghum between the conservation till and no till
practices similar to what was observed by Sharpley and Smith (1994) in the Southern Plains
region of Kansas, Oklahoma, and Texas. However, Fawcett et al. (1994) in their review of
several paired watersheds reported that conservation tillage usually led to reduced sediment and
surface runoff.

Impacts of crop rotation/cover on runoff, sediment and crop yield:
We found differences in runoff and crop yields as a result of crop rotation. Surface runoff
decreased for sorghum (-4.6% vs. -8.1% with contour) and increased for cotton (+5% regardless
of contouring) when these crops were rotated with winter wheat. The effect of wheat as cover
crop for grain sorghum generated lowest runoff followed by its rotation with canola and cotton
regardless of contouring. Sediment yield increased for sorghum (13.7% vs. 8.0% with contour)
and it decreased for cotton (11.0% regardless of contouring) when these crops were rotated with
winter wheat. The sediment yield was the highest for cotton followed by grain sorghum and
canola when rotated with winter wheat regardless of contouring.
Yields of both cotton and grain sorghum decreased when winter wheat was used as a cover crop.
Cotton yields decreased by 52.2% regardless of contouring (51% dry land cotton and 62%
irrigated lands cotton). Grain sorghum yields decreased by 28.4% vs. 14.8% under contour farming. Winter wheat yield remained virtually the same when rotated with canola and used as a cover crop for grain sorghum and cotton regardless of contouring. Osei (2016) applied three conservation practices in the FCR watershed to find the optimal distribution of conservation practices and indicated that no-till winter wheat production in central Oklahoma results in a small cost reduction while maintaining yields and is the win-win option. But since continuous no-till wheat is not possible because of weeds and other disease, it is not the good scenarios for adoption.

Impacts of crop conversion to pasture on runoff, sediment and crop yield:
We found that converting all the crops in the watershed into Bermuda grass would significantly reduce runoff by 6.8 to 38.5%, and decrease sediment loss by 72.5 to 96.3%. We did not find major difference on surface runoff and sediment loss due to two different stocking rates (1,200 and 1,600) on three grazing timings. Although conversion to pasture may be costly without government incentives, it leads substantial and consistent reductions in all environmental indicators through reduced sediment and nutrient losses (Osei, 2016).

Success of the BMP installation in the FCR watershed is of interest to many groups because erosion and transport of sediment and associated nutrients are common problems in the surrounding agricultural watersheds (Becker, 2011). Moreover, state and federal funding has supported the implementation of conservation practices in the watershed (Steiner and others, 2008). Boyer et al. (2017) stated that farming experience, gender and attitudes towards soil and water conservation increases the total number of practices adopted. According to Tong et al. (2017), negative externalities are the main challenges for adoption of conservation practices in the FCR watershed, and this point indicates the need for new extension educational efforts, economic incentives from government, and research efforts to reduce to negative externalities. These negative effects of sediment and other NPS pollutions are not paid for by the producers and landowners. Instead, downstream users (e.g. recreationists and municipal systems) face the costs. The principal approach for adoption of conservation practices for reduction of NPS pollution from agricultural fields in the USA is subsidizing adoption of conservation practices instead of taxing inputs like sediment and phosphorous. So, there should be motivations from government for landowners and producers to implement conservation practices. In this regard, apart from the environmental impact of different agricultural BMPs, there should be economic consideration of these management practices for selecting the most cost efficient BMPs since funding agencies are better appreciating the link between farm economics and producer adoption of the conservation practices.

Conclusions
We employed SWAT model to estimate changes in surface runoff, sediment load and crop-yield under five different scenarios of agricultural BMPs in an agriculture-pasture intensive watershed located in southwestern Oklahoma. We found that no-till system released less sediment load than
conservation tillage system. Compared to the conservation tillage practice, no-till system 
decreased sediment load by 25.3% and 9.0% for cotton and grain sorghum respectively. The 
contour farming with either conservation tillage or no-till practice significantly reduced sediment 
load. Similarly, contour tillage practices reduced surface runoff by more than 18% in both 
conservation tillage and no-till practices for all crops. We found varying impacts of wheat used 
as a cover crop on surface runoff, sediment load and crop yield. We found decreased runoff for 

grain sorghum and increased runoff for cotton when wheat was used as a cover crop with no-till 
system. However, we found increase in sediment load for both cotton and grain sorghum when 
no-till wheat was used as a cover crop. A hypothetical conservation scenario that converted all 
crops to Bermuda grass pasture land reduced runoff sediment yield significantly but the 
practicality of this scenario can be realized only with financial incentive programs.

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submitting this work. The rest of the authors would like to express their gratitude and admiration 
to him and also expect that this article serves as a tribute to his memory.

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Table 1 (on next page)

Streamflow and sediment calibration parameter values in study area
Table 1. Streamflow and sediment calibration parameter values in study area

<table>
<thead>
<tr>
<th>Component</th>
<th>Parameter</th>
<th>Parameter value range</th>
<th>Calibrated value</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>V__GWQMN.gw</td>
<td>0.20_0.60</td>
<td>0.60</td>
</tr>
<tr>
<td>Streamflow</td>
<td>V__GW_REVAP.gw</td>
<td>0.02_0.03</td>
<td>0.02</td>
</tr>
<tr>
<td></td>
<td>V__REVAPMN.gw</td>
<td>0.50_1.50</td>
<td>1.38</td>
</tr>
<tr>
<td></td>
<td>V__RCHRG_DP.gw</td>
<td>0.10_0.50</td>
<td>0.47</td>
</tr>
<tr>
<td></td>
<td>V__GW_DELAY.gw</td>
<td>320_390</td>
<td>376</td>
</tr>
<tr>
<td></td>
<td>R__CN2.mgt</td>
<td>-0.16_-0.13</td>
<td>-0.13</td>
</tr>
<tr>
<td></td>
<td>V__ALPHA_BF.gw</td>
<td>0.80_1.00</td>
<td>0.95</td>
</tr>
<tr>
<td></td>
<td>V__ESCO.hru</td>
<td>0.80_0.90</td>
<td>0.83</td>
</tr>
<tr>
<td></td>
<td>V__CH_K1.sub</td>
<td>0.00_0.40</td>
<td>0.09</td>
</tr>
<tr>
<td></td>
<td>V__SURLAG.bsn</td>
<td>0.50_4.00</td>
<td>3.05</td>
</tr>
<tr>
<td></td>
<td>V__EVRCCH.bsn</td>
<td>0.00_0.50</td>
<td>0.34</td>
</tr>
<tr>
<td></td>
<td>V__TRNSRCH.bsn</td>
<td>0.00_0.10</td>
<td>0.10</td>
</tr>
<tr>
<td></td>
<td>V__ALPHA_BNK.rte</td>
<td>0.60_1.00</td>
<td>0.84</td>
</tr>
<tr>
<td></td>
<td>R__SOL_AWC(..).sol</td>
<td>-0.02_0.06</td>
<td>0.04</td>
</tr>
<tr>
<td></td>
<td>V__CH_N2.rte</td>
<td>0.05_0.30</td>
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<tr>
<td></td>
<td>V__CH_K2.rte</td>
<td>1.85_2.15</td>
<td>1.98</td>
</tr>
<tr>
<td></td>
<td>R__USLE_P.mgt</td>
<td>-1.000_0.000</td>
<td>-0.240</td>
</tr>
<tr>
<td>Sediment</td>
<td>R__SLSUBBSN.hru</td>
<td>0.000_0.230</td>
<td>0.217</td>
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<tr>
<td></td>
<td>R__USLE_K().sol</td>
<td>-0.500_0.300</td>
<td>-0.247</td>
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<td></td>
<td>V__RSDCO.bsn</td>
<td>0.010_0.100</td>
<td>0.083</td>
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<td></td>
<td>V__BIOMIX.mgt</td>
<td>0.000_0.300</td>
<td>0.297</td>
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<td></td>
<td>V__SPCON.bsn</td>
<td>0.000_1.000</td>
<td>0.009</td>
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<tr>
<td></td>
<td>V__SPEXP.bsn</td>
<td>1.000_2.000</td>
<td>1.714</td>
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<tr>
<td></td>
<td>V__CH_ERODMO(..).rte</td>
<td>0.050_0.700</td>
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<tr>
<td></td>
<td>V__CH_COV1.rte</td>
<td>0.001_0.800</td>
<td>0.518</td>
</tr>
<tr>
<td></td>
<td>V__CH_COV2.rte</td>
<td>0.001_0.800</td>
<td>0.332</td>
</tr>
</tbody>
</table>

Note: “R” before the parameter name stands for relative change (the parameter is multiplied by 1+value); “V” stands for replacement (the parameter is replaced by a value within the range)
Agricultural Best Management Practices (BMPs) scenarios simulated for cotton, grain sorghum and winter wheat to evaluate their impacts on hydrology, sediment and crop yield in the study area
Table 2. Agricultural Best Management Practices (BMPs) scenarios simulated for, cotton, grain sorghum and winter wheat to evaluate their impacts on hydrology, sediment and crop yield in the study area

<table>
<thead>
<tr>
<th>Code</th>
<th>BMP Scenario</th>
<th>Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>BL</td>
<td>Baseline</td>
<td>Simulation under the calibrated and validated model with 14 land uses, 8 km2 FMC under contour farming</td>
</tr>
<tr>
<td>S1</td>
<td>Conservation tillage and strip cropping</td>
<td>BMP applied to cotton, grain sorghum, and winter wheat. No changes made to hay and alfalfa. Data obtained from NASS (2014), Storm et al. (2003) and Storm et al. (2006). Total three simulations, one for each crop.</td>
</tr>
<tr>
<td>S2</td>
<td>Conservation tillage on contour</td>
<td>Applied contour on scenarios 1; 97 km2 additional contour as compared to the baseline scenario. Resulted three simulations, one for each crop.</td>
</tr>
<tr>
<td>S3</td>
<td>No-till and strip cropping</td>
<td>All tillage practices were removed while management practices were kept the same; applied to cotton, grain sorghum and winter wheat. Because of weed and disease problems associated with continuous no-till wheat, wheat was rotated/cover cropped with (i) canola, (ii) cotton and (iii) grain sorghum. Total five simulations, one for each crop.</td>
</tr>
<tr>
<td>S4</td>
<td>No-till on contour</td>
<td>Applied contour on Scenario 3. Resulted five simulations, one for each crop.</td>
</tr>
<tr>
<td>S5</td>
<td>Conversion to pasture</td>
<td>All crops were converted to Bermuda grass pasture. A combination of three grazing start months (May, June and July) and two stocking rates (1,200 and 1,600 kg) were applied. Total of six simulations.</td>
</tr>
</tbody>
</table>

Note: Details of each scenario are provided in Appendix F
Table 3 (on next page)

Sediment reduction in percentage as a result of contouring on conservation tillage and no-till practices for cotton, grain sorghum and winter wheat
Table 3. Sediment reduction in percentage as a result of contouring on conservation tillage and no-till practices for cotton, grain sorghum and winter wheat

<table>
<thead>
<tr>
<th>Grain sorghum</th>
<th>Cotton</th>
<th>Wheat</th>
</tr>
</thead>
<tbody>
<tr>
<td>Conservation tillage</td>
<td>Conservation tillage</td>
<td>Conservation tillage</td>
</tr>
<tr>
<td>44</td>
<td>44</td>
<td>45</td>
</tr>
</tbody>
</table>


Figure 1 (on next page)

Schematic representation of Best management practices (BMP) implementation in a watershed
Figure 1. Schematic representation of Best management practices (BMP) implementation in a watershed
Figure 2 (on next page)

Five-Mile Creek sub-watershed (FMC) located within the Fort Cobb Reservoir watershed. Land types, slope and soil classes and elevation maps are included.
Figure 2. Five-Mile Creek sub-watershed (FMC) located within the Fort Cobb Reservoir watershed. Land types, slope and soil classes and elevation maps are included.
Figure 3 (on next page)

Calibration and validation monthly time series (2000–2010) for observed and SWAT predicted streamflow at the Cobb Creek near Eakley, Oklahoma gauging station
Figure 3. Calibration and validation monthly time series (2000–2010) for observed and SWAT predicted streamflow at the Cobb Creek near Eakley, Oklahoma gauging station.
Figure 4 (on next page)

Observed daily discharge and observed daily suspended sediment concentration trend
Figure 4. Observed daily discharge and observed daily suspended sediment concentration trend
Figure 5 (on next page)

Calibration and validation monthly time series (2000–2010) for observed and SWAT simulated suspended sediment concentration at the Cobb Creek near Eakley, Oklahoma gauging station
Figure 5. Calibration and validation monthly time series (2000–2010) for observed and SWAT simulated suspended sediment concentration at the Cobb Creek near Eakley, Oklahoma gauging station.
Changes in surface runoff generation under different scenarios of Best Management Practices
Figure 6. Changes in surface runoff generation under different scenarios of Best Management Practices.
Average annual sediment loss (tons per hectare) under each five agricultural Best Management Practices scenarios compared with the baseline scenario
Figure 7. Average annual sediment loss (tons per hectare) under each five agricultural Best Management Practices scenarios compared with the baseline scenario.
Figure 8 (on next page)

Sub-basin level SWAT simulated sediment loadings (tons/ha) in Five-Mile Creek sub-watershed under different BMP scenarios
Figure 8. Sub-basin level SWAT simulated sediment loadings (tons/ha) in Five-Mile Creek sub-watershed under different BMP scenarios.
Figure 9 (on next page)

Crop yields under different scenarios of Best Management Practices
Figure 9. Crop yields under different scenarios of Best Management Practices.