

1	Chemicals associated with plastic packaging: Inventory and hazards
2 3	Ksenia J. Groh ^{1*} , Thomas Backhaus ² , Bethanie Carney-Almroth ² , Birgit Geueke ¹ , Pedro A. Inostroza ² ,
4	Anna Lennquist³, Maricel Maffini⁴, Heather A. Leslie⁵, Daniel Slunge⁶, Leonardo Trasande⁶, A. Michael
5	Warhurst ⁸ and Jane Muncke ¹
6	
7	¹ Food Packaging Forum Foundation, Zurich, Switzerland
8	² Department of Biological and Environmental Sciences, University of Gothenburg, Gothenburg, Sweden
9	³ International Chemical Secretariat (ChemSec), Gothenburg, Sweden
10	⁴ Independent Consultant, Germantown, Maryland, USA
11	⁵ Department of Environment & Health, Vrije Universiteit Amsterdam, The Netherlands
12	⁶ Centre for Sustainable Development (GMV), University of Gothenburg, Gothenburg, Sweden
13	⁷ School of Medicine, New York University, New York, USA
14	⁸ CHEM Trust, London, United Kingdom
15	
16	
17	*Corresponding author:
18	Ksenia J. Groh
19	Food Packaging Forum Foundation
20	Staffelstrasse 8
21	CH-8045 Zurich, Switzerland
22	E-mail: ksenia.groh@fp-forum.org



24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

40

41

42

43

44

45

46

Λ.	Ι.		_		- 4	
/\	n	C.	г	ra	\sim 1	г.
$\overline{}$	L.	ר.ו	l	ıa	١.,	

Global plastics production has reached 380 million metric tons in 2015, with around 40% used for packaging. Plastic packaging is diverse and made of multiple polymers and numerous additives, along with other components, such as adhesives or coatings. Further, packaging can contain residues from substances used during manufacturing, such as solvents, along with non-intentionally added substances (NIAS), such as impurities, oligomers, or degradation products. To characterize risks from chemicals potentially released during manufacturing, use, disposal, and/or recycling of packaging, comprehensive information on all chemicals involved is needed. Here, we present a database of Chemicals associated with Plastic Packaging (CPPdb), which includes chemicals used during manufacturing and/or present in final packaging articles. The CPPdb lists 906 chemicals likely associated with plastic packaging and 3377 substances that are possibly associated. Of the 906 chemicals likely associated with plastic packaging, 63 rank highest for human health hazards and 68 for environmental hazards according to the harmonized hazard classifications assigned by the European Chemicals Agency within the Classification, Labeling and Packaging (CLP) regulation implementing the United Nations' Globally Harmonized System (GHS). Further, 7 of the 906 substances are classified in the European Union as persistent, bioaccumulative, and toxic (PBT), or very persistent, very bioaccumulative (vPvB), and 15 as endocrine disrupting chemicals (EDC). Thirtyfour of the 906 chemicals are also recognized as EDC or potential EDC in the recent EDC report by the United Nations Environment Programme. The identified hazardous chemicals are used in plastics as monomers, intermediates, solvents, surfactants, plasticizers, stabilizers, biocides, flame retardants, accelerators, and colorants, among other functions. Our work was challenged by a lack of transparency and incompleteness of publicly available information on both the use and toxicity of numerous substances. The most hazardous chemicals identified here should be assessed in detail as potential candidates for substitution.

47

48

Keywords

49 Plastics; packaging; chemical composition; additive; harmonized hazard data; environment; human

50 health; substitution; non-intentionally added substances (NIAS)

52 53

54

55

56

57

58

59

60

61

62

63

64

65

66

67

68

69

70

71

72

73

74

75

76

77

78

79

80

81

82

83

84

85

1. Introduction

The use of plastic packaging is on the rise (Smithers Pira, 2018; Van Eygen et al., 2017), explained by the need to reduce food waste or the increased demand due to population growth and market expansion (Andrady and Neal, 2009; Sohail et al., 2018; Thompson et al., 2009b). However, there are also increasing concerns about the harm caused to the environment (North and Halden, 2013) and human health (Halden, 2010). These concerns include littering and accumulation of nondegradable plastics in the environment (Jambeck et al., 2015; Thompson et al., 2009a), generation of secondary microplastics and nanoplastics (Galloway, 2015; Galloway and Lewis, 2016; Revel et al., 2018; Wright and Kelly, 2017), and release of hazardous chemicals during manufacturing and use (Biryol et al., 2017; Caporossi and Papaleo, 2017; Dematteo et al., 2013), as well as following landfilling (Mavakala et al., 2016; Sarigiannis, 2017), incineration (Franchini et al., 2004), or improper disposal leading to pollution of the environment (Gallo et al., 2018; Hahladakis et al., 2018; Hammer et al., 2012; Hermabessiere et al., 2017). In 2015, production of plastics reached 380 million metric tons worldwide, and of these, around 40% were used for packaging (Geyer et al., 2017; PlasticsEurope, 2016). Around 60% of all plastic packaging is used for food and beverages, while the rest covers non-food applications, such as healthcare, cosmetics, consumer, household, apparel, and shipment packaging. To reduce environmental impacts, efforts to drastically increase recycling rates of packaging plastics are currently being undertaken (EU, 2018; European Commission, 2018; Hammer et al., 2012).

Many chemicals used to make plastics, including packaging plastics, are highly hazardous (Lithner et al., 2011) and therefore of significant concern for occupational health (Fucic et al., 2018; Montano, 2014). Moreover, during the subsequent use, disposal, and recycling chemicals present in plastic packaging may transfer into products such as foods or cosmetics, or in the environment (Gallo et al., 2018; Hahladakis et al., 2018; Hermabessiere et al., 2017). With this, plastic packaging is likely to substantially contribute to chemical exposures of the human population and the environment (Biryol et al., 2017; Grob et al., 2006). Recycling can also result in accumulation of hazardous chemicals in secondary materials, negatively affecting their market value and restricting downstream applications (Geueke et al., 2018; Lahl and Zeschmar-Lahl, 2013; Ragossnig and Schneider, 2017). Therefore, a detailed assessment of plastic packaging-associated chemicals may be necessary (Bilitewski et al., 2012b; Bodar et al., 2018; Guzzonato et al., 2017).

Several strategies for assessing and scoring hazardous chemicals in plastics have already been proposed (Lithner et al., 2011; Rossi and Blake, 2014). These studies have so far focused on monomers and a small number of additives used in high concentrations, such as phthalate plasticizers in flexible polyvinyl chloride (PVC). However, to comprehensively inform design, manufacturing, and policy decisions supporting benign alternatives to hazardous chemicals in plastic



packaging, the identity and amounts of all involved substances should be known. This applies to both the intentionally added substances (IAS), such as monomers used to make the polymer, additives added to the polymer to impart a desired property or function, and other chemicals intentionally used during manufacturing (e.g., solvents or processing aids), as well as the non-intentionally added substances (NIAS), such as impurities, reaction by-products, and breakdown products (Bradley and Coulier, 2007; Nerin et al., 2013). Unfortunately, achieving a comprehensive overview of all substances associated with plastic packaging is not a straightforward task. This is due to two main reasons: First, with regard to IAS, packaging plastics are made of multiple polymer types containing a large variety of chemical additives. Second, the chemical identity of many NIAS present in finished plastic packaging is seldom measured and often remains unknown.

The five polymers most commonly applied in plastic packaging include polyethylene (PE), polypropylene (PP), polyethylene terephthalate (PET), polystyrene (PS), and PVC (PlasticsEurope, 2016). Many other polymers, including, for example, polycarbonate (PC), polyamides (PA, nylon), acrylics, polylactic acid (PLA), polyurethanes (PU) and even more specialized polymer types, are also applied for specific packaging applications (PlasticsEurope, 2016; Selke and Culter, 2016). A recent study found that the less common polymers still account for over 10% of post-consumer plastic packaging waste collected for recycling (Brouwer et al., 2018).

The major families of plastics additives (listed in order of decreasing total tonnage) are fillers, plasticizers, flame retardants, colorants, stabilizers, lubricants, foaming agents, and antistatic agents. Stabilizers can be further divided into several groups with more specific functions, including antioxidants, antiozonants, heat stabilizers, UV stabilizers, and biostabilizers (biocides) (Harper, 2006). There are also many other types of additives used in lower amounts (Flick, 2002; Zweifel et al., 2009). In addition to single-material based articles, multimaterial multilayers along with laminated structures are frequently used for packaging as well. In such systems, adhesives are often used to hold the construct together, adding another dimension of chemical diversity (Nerin et al., 2018).

Currently, no publicly available information source exists that offers a one-stop, easily accessible overview of all chemicals associated with plastic packaging. The available information is incomplete and scattered across multiple sources. For example, Annex I of Commission Regulation (EU) No 10/2011 (EU, 2011), also called the Union list, is a positive list of monomers and additives authorized for use in plastic food contact materials (FCMs). The Union list is, however, not specific for plastic food packaging, since it covers all food contact plastics in general, i.e., it includes other food contact articles (FCAs) such as cooking utensils or conveyor belts. Moreover, this list is focused on the IAS only and consequently does not cover most NIAS. Importantly, the EU legislation requires that safety of NIAS should be covered by risk assessment of FCMs but does not provide detailed



guidance. Further, to the best of our knowledge, no lists comparable to the Union list exist for non-food plastic packaging.

To address the current dearth of information, we compiled a list of chemicals associated with plastic packaging. For this, we relied primarily on the information available from publicly accessible sources, but also explored the suitability of several industry-maintained databases on plastics and plastics additives products for finding and extracting information on the exact chemical composition of materials used in plastic packaging. However, we did not include any commercial, paid-for data sources, in order to ensure that the resulting database of chemicals associated with plastic packaging could be made publicly available. Free access to such a resource is highly desirable, as it can inspire and enable further research, regulatory actions, and packaging design innovations, as well as guide citizen activities aimed to tackle the plastic pollution problem. Further, we preliminarily explored the environmental and human health hazards of these chemicals and identified over a hundred substances of high concern that may require further detailed assessment. Numerous data gaps exist with regard to the use patterns and exposure, but also toxicity of plastic packaging-associated substances. These gaps need to be addressed in the future in order to properly evaluate the risks and explore substitution options.

2. Materials and Methods

2.1 Compilation of the database "Chemicals associated with plastic packaging" (CPPdb)

An overview of information sources used to compile the database of chemicals associated with plastic packaging (CPPdb) is given in Figure 1 and the workflow is described below. As a first information source we used the Chemicals and Product Categories database (CPCat; http://actor.epa.gov/cpcat) constructed by the U.S. Environmental Protection Agency (EPA) (Dionisio et al., 2015) (Fig. 1A). From this database, we extracted chemicals assigned to categories we deemed to be potentially plastics- and/or packaging-relevant, such as raw materials or monomers used to make plastics, or various types of additives and process regulators (Flick, 2002; Zweifel et al., 2009). Thus, we extracted the CPCat categories such as "plastics," "manufacturing packing plastics," "food packaging" or "fillers," but omitted the categories clearly irrelevant to plastic packaging, such as "building material plastics" or "wood preservatives." The complete list of all CPCat categories extracted for this project is given in the Supplementary File 1. Since many chemicals can be used in multiple applications, most chemicals in the CPCat database are listed in multiple categories. Therefore, in the next step we fused all extracted categories for each extracted unique CAS number. The extracted chemicals were regarded as associated with plastics ("yes" in the respective column in the CPPdb) whenever the extracted categories contained the indication "plastics" or "manufacturing



plastics," and as associated with "plastic packaging" ("yes" in the respective column in the CPPdb) whenever there were indications for both "plastics" and "packaging."

The second information source consulted for the CPPdb (Fig. 1A) was the "Plastics additives database" by Ernest Flick (2004), which lists around 7000 commercial products being marketed as plastics additives, sorted according to their function in plastics, e.g., stabilizer, antioxidant, filler, or plasticizer (Flick, 2004). Where available, the identity of chemicals associated with the listed products was recorded (withholding the brand name) and corresponding CAS numbers were searched and assigned manually where found. All chemicals referenced in this source were regarded as associated with plastics. The resulting list of chemicals was merged with the CPCat-derived list. For duplicate chemicals (matching by CAS), Flick-derived use/function information was added to CPCat-derived category descriptors and identification as being associated with plastics was assigned if not already given previously based on CPCat data.

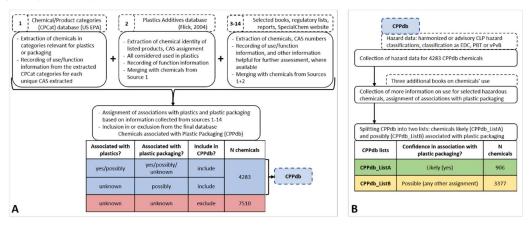


Figure 1. Overview of information sources (dashed-line boxes) and approaches/workflow (solid-line boxes) followed to (A) compile the database of chemicals associated with plastic packaging (CPPdb) and (B) split the full CPPdb into the two lists according to the assigned likely (CPPdb_ListA) or possible (CPPdb_ListB) association with plastic packaging. Abbreviations not explained on the figure: CLP, Classification, Labeling and Packaging; EDC, endocrine disrupting chemical; PBT, persistent, bioaccumulative, toxic; vPvB, very persistent, very bioaccumulative.

 Further information sources (see Fig. 1A, sources numbered 3 to 13) that we consulted and integrated into the CPPdb included: (3) the chemicals listed in the book on chemical analytics of plastics and polymers (Bolgar et al., 2016); (4) the list of plastic monomers and selected additives along with some other substances used in plastics manufacturing (Lithner et al., 2011); (5) the chemicals listed on the European Flavours, Additives, and food Contact materials Exposure Task (FACET) list for food contact materials (Oldring et al., 2014); (6-8) three reports on hazardous chemicals associated with plastics (Hansen et al., 2013; Klar et al., 2014; Stenmarck et al., 2017); (9) the list of 1009 food contact substances for which Biryol and colleagues provided high-throughput estimates of their initial concentration in food contact articles made of "plastics, coatings, or



silicones" (Biryol et al., 2017); (10) the chemicals discussed in two reviews on NIAS (Geueke, 2018; Nerin et al., 2013); (11) the list of extractables from pharmaceutical packaging and devices made of polyolefins, PET, PS, PVC, PA, ethylene-vinyl acetate (EVA), and PU (Jenke, 2009); (12) Annex I of Commission Regulation (EU) No 10/2011, listing plastics monomers and additives permitted for use in plastic food contact materials in the European Union (EU, 2011); and (13) EPA's list of 16 priority polycyclic aromatic hydrocarbons (PAHs) (Keith and Telliard, 1979; Keith, 2015). The latter list was included because some PAHs have been detected as contaminants in plastic packaging made of PS (Li et al., 2017; Rochman et al., 2013). Apart from the chemicals' identity extracted from these sources, we also recorded, as far as available, other relevant information, such as product or polymer types where a given substance has been used or found. The resulting lists of chemicals were merged with the list generated at the previous step. For CAS-matched chemicals, additional information was recorded and identification as being associated with plastics and/or plastic packaging was assigned where appropriate.

The 14th and final information source consulted for the initial compilation of the CPPdb (Fig. 1A) was the SpecialChem website ("The Universal Selection Source of Chemical & Materials," https://www.specialchem.com, accessed November 24, 2017). Some content on this website can be viewed for free upon registration; additional features are provided through paid upgrades, but this option was not used in our research. We performed a search in the subsection on plastics additives (https://polymer-additives.specialchem.com) with the filter "recommended for packaging applications." For the products returned with this search, we extracted chemical identity and assigned CAS numbers manually, where found. The resulting list was fused with the list generated at the previous step. For CAS-matched chemicals, additional identification as being associated with plastics and/or plastic packaging was assigned if not given previously.

All chemicals identified as likely or possibly associated with plastics were kept in the final CPPdb inventory, regardless of whether or not clear indications of their association with plastic packaging were available at this stage (Fig. 1A). This was done because we reasoned that all chemicals associated with plastics could potentially be associated with plastic packaging as well, and we lacked the information necessary to systematically exclude substances that might be applied only in plastics not used for packaging purposes. However, all chemicals for which we could not find at this stage any indication of their association with plastics or plastic packaging were removed from the final CPPdb inventory (Fig. 1A). All CPPdb chemicals were then taken further to the next stage of this project, that is, the exploration of human health and environmental hazards (see Subsection 2.2). For selected hazardous chemicals identified at this later stage, three additional books on chemicals' use (Ash and Ash, 2008; Milne, 2005; Sheftel, 2000) were then consulted to obtain more information regarding their potential use in plastic packaging, and additional identification as being associated with plastic



218

219

220

221

222223

224

225

226

227

228

229

230

231

232

233

234

235

236

237

238

239

240

241

242

243

244

245

246

247

248

249

250

251

packaging was then assigned where justified. After this, depending on a substance's assigned association with plastic packaging, the full CPPdb was split into the CPPdb ListA and CPPdb ListB, containing lists of chemicals likely and possibly associated with plastic packaging, respectively (Fig. 1B, see also Subsection 3.1). 2.2 Exploration of CPPdb chemicals' hazards To explore the environmental and human health hazards of the CPPdb chemicals, we applied, with some modifications, the methodology published by Lithner and colleagues (2011). This approach uses the harmonized hazard classifications assigned by the European Chemical Agency (ECHA) under the European Classification, Labeling and Packaging (CLP) legislation (EU, 2008). The CLP governs the EU implementation of the Globally Harmonized System of Classification and Labeling of Chemicals (GHS), adopted by the United Nations (UN) in 2002 (see http://www.unece.org/trans/danger/publi/ghs/ghs welcome e.html, accessed June 5, 2018). Harmonized CLP classifications (hazard category and class) were extracted from the ECHA-provided file "Annex VI to CLP ATP10" (updated version of Table 3.1 of Annex VI to CLP, 10th Adaptation to Technical Progress (ATP10; in force from Dec 1, 2018), accessed at https://echa.europa.eu/information-on-chemicals/annex-vi-to-clp on Nov 2, 2017). These classifications were matched to the CPPdb chemicals by CAS numbers. Every classification was assigned a numerical hazard grade score according to the previously proposed gradation (Lithner et al., 2011). For example, Carcinogenicity 1A (Carc. 1A) and Carc. 1B received the highest hazard grade score of 10 000, while that of Carc. 2 was 100. Acute Toxicity 1 (Acute Tox. 1) and Acute Tox. 2 both received a hazard grade score of 1000, while grade scores of Acute Tox. 3 and 4 were 100 and 10, respectively. For aquatic toxicity, Aquatic Acute 1 (Aq. Acute 1) classification was assigned a hazard grade score of 100 and Aq. Chronic 1 received a hazard grade score of 1000. A table listing all CLP classifications with corresponding hazard grade scores as adopted in this work is given in the Supplementary File 2. Then, a sum hazard score was calculated for each chemical, separately for environmental and human health hazards. For the former, hazard grade scores assigned in Aq. Acute and Ag. Chronic categories were summed. For the latter, hazard grade scores assigned in all human health-related toxicity categories were summed. Since many chemicals in the CPPdb did not have any harmonized CLP classifications assigned by ECHA, we also extracted advisory CLP classifications for human health hazards as assigned by the Danish Environmental Protection Agency (Danish EPA) in the online database at http://mst.dk/kemi/kemikalier/stoflister-og-databaser/vejledende-liste-til-selvklassificering-af-

farlige-stoffer/clp/, accessed March 28, 2018. Each advisory CLP classification was then assigned the

same hazard grade score as that assigned to a matching harmonized CLP classification, and sum



hazard scores were calculated as described above for harmonized CLP classifications. We did not examine advisory CLP classifications for environmental hazards, because updated classifications for the Aq. Chronic category were not available at the time.

In addition to the hazard categories currently covered by the CLP system, we also considered whether the substance is classified as persistent, bioaccumulative, and toxic (PBT), or very persistent, very bioaccumulative (vPvB), or endocrine disrupting chemical (EDC). This type of information was searched and added for the CPPdb chemicals by consulting sources such as PBT and vPvB status of a substance in the EU (https://echa.europa.eu/information-on-chemicals/pbt-vpvb-assessments-under-the-previous-eu-chemicals-legislation, accessed December 7, 2017), EDC classifications assigned by December 2017 within the Registration, Evaluation, Authorization and restriction of Chemicals (REACH) regulation in the EU (EU, 2006), recognition as an EDC or a potential EDC listed in the recent report by the United Nations Environment Programme (UNEP) on EDCs (UNEP, 2018), the Substitute It Now! (SIN) list maintained by the non-governmental organization (NGO) International Chemical Secretariat (ChemSec; http://chemsec.org/business-tool/sin-list/, accessed November 24, 2017) and ChemSec's identification of REACH-relevant EDCs (ChemSec, 2015), and The Endocrine Disruption Exchange (TEDX) list of potential EDCs (https://endocrinedisruption.org/interactive-tools/tedx-list-of-potential-endocrine-disruptors/, accessed November 24, 2017), maintained by the U.S. NGO TEDX.

To complete the CPPdb with information which could be useful for further assessment, we also documented additional aspects, for example, a substance's production tonnage band in the EU as documented in its REACH registration dossier (resource accessed September 11, 2017, database containing information from 62200 dossiers on 16402 unique substances), a substance's presence in the U.S. Toxic Substances Control Act inventory (accessed January 16, 2018), or whether it is included in a biomonitoring program in the U.S. (Centers for Disease Control and Prevention (CDC), 2018), performed with the blood and urine samples collected within the National Health and Nutrition Examination Survey (NHANES) (Sobus et al., 2015), or similar programs in Canada, i.e., within Canada Health Measures Survey (Haines et al., 2017), or Germany (Kolossa-Gehring et al., 2017). With regard to the regulatory status, we recorded a substance's inclusion on the Candidate, Authorization, or Restriction lists (status December 2017) within REACH (EU, 2006), on the California Proposition 65 list in the U.S. (http://oehha.ca.gov/proposition-65/proposition-65-list, accessed March 16, 2018, and in the Occupational Chemical Database maintained by the U.S. Occupational Safety and Health Administration (OSHA) (http://www.osha.ogv/chemicaldata/, accessed November 25, 2017. A detailed description of all information sources used for populating different columns in the CPPdb is given in the "READ ME" worksheet of the Supplementary File 3 presenting the CPPdb.



3. Results

3.1 The CPPdb and its information content

To populate the CPPdb we relied on 14 publicly available information sources and followed the strategy described in the subsection 2.1 and summarized in the Figure 1A. Based on this, we identified and included in the CPPdb 4283 substances that are likely or possibly associated with plastic packaging, both in terms of use during the manufacturing and/or presence in the final products. Additional 7510 substances, most of them originally retrieved from our first source, the CPCat database, were excluded from the CPPdb list at that stage, because the 14 initial information sources did not contain any indication of these substances' use in plastics and/or plastic packaging (Fig. 1A).

The 4283 substances included in the CPPdb cover raw materials and chemicals used in plastics manufacturing, such as monomers, polymerization aids, solvents or catalysts, along with additives such as pigments, fillers, antioxidants, stabilizers, plasticizers, antistatic agents, slip agents and others (Zweifel et al., 2009). We also added several well-known NIAS such as impurities or degradation products of some polymers and stabilizers (Nerin et al., 2013). However, the coverage of NIAS was knowingly incomplete. Many more substances have been reported to be present in different plastics (Bradley and Coulier, 2007), but data characterizing their complete chemical identity, hazard, and exposure are often lacking (Muncke et al., 2017).

When compiling the CPPdb, we sought to determine the substance's relevance specifically for plastic packaging, whenever any such indications could be found in the original 14 data sources. However, information on the substance's association with plastic packaging was even scarcer than that on the substance's association with plastics in general. For example, the CPCat database does contain a category "manufacturing packing plastics," but this category includes only 79 substances, although many more chemicals are likely used (Biryol et al., 2017; EU, 2011; Muncke et al., 2017; Oldring et al., 2014). Most other data sources, while indicating that a given chemical is used in plastics in general, often did not further specify the types of polymers, products, or applications using this chemical. Thus, based on the 14 initial information sources, the likely association with plastic packaging could be assigned for only a few hundred CPPdb substances. Therefore, for a subset of substances we considered the most important, we consulted three additional literature sources on chemical uses (Ash and Ash, 2008; Milne, 2005; Sheftel, 2000) in order to further explore their potential association with plastic packaging. These substances were those identified to be of the highest environmental and/or human health hazard based on harmonized and/or advisory CLP classifications (see below), and/or classified in the EU as EDC, PBT, or vPvB substance, or included in biomonitoring programs in the U.S., Canada or Germany (Fig. 1B).



321

322

323

324

325

326

327

328

329

330

331

332

333

334

335

In the end, we were able to identify 906 substances to be likely associated with plastic packaging in terms of being used during the manufacture of plastic packaging or being present in the final packaging articles. These substances are listed on the CPPdb ListA (Fig. 1B). All remaining substances (3377 chemicals), jointly considered as possibly associated with plastic packaging, are listed on the CPPdb_ListB (Fig. 1B). This list includes substances for which only 'possible' association with plastic packaging could be assigned, or no information regarding the association with plastic packaging was available in the information sources referenced above, with or without the three additional books on chemicals' use. Since these three last sources were checked only for selected chemicals, it is possible that some of the chemicals on the CPPdb_ListB are in fact also 'likely' associated with plastic packaging, but this could not be assigned based on the information collected. Note, although a few substances on the CPPdb ListB are currently banned in Europe or in the U.S., they have been kept on the list because their relevance to plastic packaging worldwide could not be excluded. The two CPPdb lists can be found in the Supplementary File 3 or accessed at the DOI 10.5281/zenodo.1287773. Additionally, the CPPdb lists will be uploaded into the Chemical Hazard Data Commons resource, maintained by the non-governmental organization (NGO) Healthy Building Network (HBN) (https://commons.healthymaterials.net/home, accessed June 5, 2018).

336337

338

339

340

341

342

343

344

345

346

347

348

349

350

351

352

353

354

3.2 Exploration of CPPdb chemicals' hazards, identification of the most hazardous substances To explore the hazards of the CPPdb chemicals with the goal to identify the most hazardous substances, we consulted both the harmonized CLP classifications (i.e. ECHA-assigned) and the advisory CLP classifications assigned by the Danish EPA based on in silico models. In addition, we considered EU-accepted classifications as an EDC, PBT, or vPvB substance, and recognition as an EDC or a potential EDC in the 2018 UNEP report on EDCs (UNEP, 2018). The UNEP report recognizes three stakeholders and consequently the EDC assessments that they performed as "robust," namely the REACH EDC classifications, the ChemSec's assessments for SIN list, and the assessments by the Danish Centre for Endocrine Disruptors. The "thorough scientific assessments" performed by these stakeholders were carried out based on the definition of an EDC or a potential EDC given by the World Health Organization (WHO)/International Program on Chemical Safety (IPCS) in 2002 (WHO/IPCS, 2002). Further, the UNEP report differentiates between chemicals identified as EDCs or potential EDCs with the involvement of "multiple stakeholders" (this group includes REACH EDC classifications only) and "at least one stakeholder" (this group includes EDC identifications by ChemSec and/or Danish Centre for Endocrine Disruptors). For 60% of the 4283 CPPdb substances no hazard data were available in the sources we reviewed. A detailed breakdown of the number of chemicals having at least one of the reviewed hazard classifications is given in Table 1. To select the most hazardous substances as candidates for further assessment and potential substitution with



regard to their use in plastic packaging, we next focused on the substances identified as likely associated with plastic packaging (CPPdb_ListA), as discussed in the next subsections.

Table 1. Availability of hazard information for CPPdb chemicals in the sources consulted

Likely Possible CPPdb_ListA CPPdb_ListA CPPdb_List CPPdb_ListA CPPdb_List CPPd_List CP
Total 4283 906 3377 Harmonized CLP for environmental hazards 332 (7.8%) 121 (13.4%) 211 (6.2%) sum 200 1 1 0 0 0 66 30 66 score 10 62 22 40 0 0, i.e., no classifications 3951 (92.2%) 785 (86.6%) 3166 (93.8%) Harmonized CLP for human health hazards 718 (16.8%) 245 (27.0%) 473 (14.0%) 1000 – 3210 (34 210)* 205 63 142 sum 1000 – 3210 200 67 133 hazard 100 – 400 182 70 112 score 10 – 50 131 45 86
Harmonized CLP for environmental hazards 332 (7.8%) 121 (13.4%) 211 (6.2%)
sum hazard score 1100 173 68 105 hazard score 100 96 30 66 score 10 62 22 40 0, i.e., no classifications 3951 (92.2%) 785 (86.6%) 3166 (93.8%) Harmonized CLP for human health hazards 718 (16.8%) 245 (27.0%) 473 (14.0%) sum hazard hazard 1000 – 3210 200 67 133 hazard score 10 – 400 182 70 112 score 10 – 50 131 45 86
sum hazard hazard 200 1 1 0 score 10 96 30 66 0, i.e., no classifications 3951 (92.2%) 785 (86.6%) 3166 (93.8%) Harmonized CLP for human health hazards 718 (16.8%) 245 (27.0%) 473 (14.0%) sum hazard hazard 1000 – 32 100 (34 210)* 205 63 142 sum hazard score 100 – 400 182 70 112 score 10 – 50 131 45 86
hazard score 10 96 30 66 score 10 62 22 40 0, i.e., no classifications 3951 (92.2%) 785 (86.6%) 3166 (93.8% Harmonized CLP for human health hazards 718 (16.8%) 245 (27.0%) 473 (14.0% sum 1000 – 32 100 (34 210)* 205 63 142 sum 1000 – 3210 200 67 133 hazard 100 – 400 182 70 112 score 10 – 50 131 45 86
score 10 62 22 40 0, i.e., no classifications 3951 (92.2%) 785 (86.6%) 3166 (93.8%) Harmonized CLP for human health hazards 718 (16.8%) 245 (27.0%) 473 (14.0%) sum 10 000 – 32 100 (34 210)* 205 63 142 sum 1000 – 3210 200 67 133 hazard 100 – 400 182 70 112 score 10 – 50 131 45 86
0, i.e., no classifications 3951 (92.2%) 785 (86.6%) 3166 (93.8%) Harmonized CLP for human health hazards 718 (16.8%) 245 (27.0%) 473 (14.0%) sum 10 000 – 32 100 (34 210)* 205 63 142 sum 1000 – 3210 200 67 133 hazard 100 – 400 182 70 112 score 10 – 50 131 45 86
Harmonized CLP for human health hazards 718 (16.8%) 245 (27.0%) 473 (14.0%) sum 10 000 – 32 100 (34 210)* 205 63 142 sum 1000 – 3210 200 67 133 hazard 100 – 400 182 70 112 score 10 – 50 131 45 86
sum 10 000 - 32 100 (34 210)* 205 63 142 sum 1000 - 3210 200 67 133 hazard 100 - 400 182 70 112 score 10 - 50 131 45 86
sum 1000 – 3210 200 67 133 hazard 100 – 400 182 70 112 score 10 – 50 131 45 86
hazard 100 – 400 182 70 112 score 10 – 50 131 45 86
score 10 – 50 131 45 86
0, i.e., no classifications 3565 (83.2%) 661 (73.0%) 2904 (86.0%)
Advisory CLP for human health hazards 950 (22.2%) 202 (22.3%) 748 (22.1%)
1000 – 2110 (2220) 304 102 202
sum 100 – 300 (200) 264 59 205
score 10 – 20 382 41 341
0, i.e., no classifications 3333 (77.8%) 704 (77.7%) 2629 (77.9%)
EDC classification, REACH 17 (0.4%) 15 (1.7%) 2 (0.06%)
EDC identification, UNEP report, 'multiple' 16 (0.4%) 14 (1.5%) 2 (0.06%)
EDC identification, UNEP report, 'at least one' 25 (0.6%) 20 (2.2%) 5 (0.15%)
PBT and/or vPvB classification, EU 31 (0.7%) 7 (0.8%) 24 (0.7%)
None of the above classifications 2568 (60.0%) 433 (47.6%) 2135 (63.2%)

^{*}The number in brackets shows the upper border value for this range on the CPPdb_ListB, if different from CPPdb_ListA.

3.2.1 Identification of the most hazardous substances based on harmonized CLP and EU classifications

The sum hazard scores derived based on harmonized CLP classifications for environmental or human health hazards were used to sort the lists of chemicals from highest to lowest sum hazard scores, corresponding to higher and lower hazards, respectively. In addition to harmonized CLP-based ranking, substances classified within the EU as EDC, PBT, or vPvB, and recognized as EDCs or potential EDCs in the UNEP report (2018) were also considered to be highly hazardous for human health and/or the environment. A detailed breakdown of the number of substances with each classification and within several sum hazard score ranges is given in Table 1. Based on these sources,

369

370

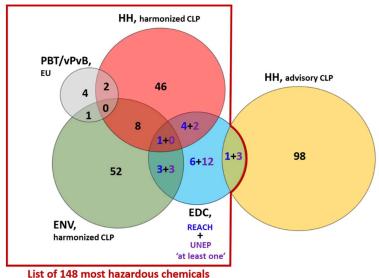
371

372

373

374

we identified 148 substances as the most hazardous ones among the 906 substances on the CPPdb ListA, i.e., likely associated with plastic packaging; 24 of these substances had more than one of the reviewed hazard classifications in the chosen 'most hazardous' range (Fig. 2). These 148 substances are listed in the Supplementary File 4. Among the 3377 substances on the CPPdb ListB, i.e., possibly associated with plastic packaging, 214 substances were identified as the most hazardous based on the same criteria, with 57 substances showing an overlap between different classifications in the 'most hazardous' range.



375 376

377

378

379

380

381

382

383

384

385

386

387

388

389

390

Figure 2. Overlap between the groups of the most hazardous substances likely associated with plastic packaging (i.e., on the CPPdb ListA), identified according to (i) environmental (ENV) hazards based on harmonized Classification, Labeling and Packaging (CLP) classifications (green circle, list "ENV, harmonized CLP," 68 substances in total); (ii) human health (HH) hazards based on harmonized CLP classifications (red circle, list "HH, harmonized CLP," 63 substances in total); (iii) classification in the European Union (EU) as a substance with persistent, bioaccumulative, and toxic (PBT) properties and/or very persistent, very bioaccumulative (vPvB) properties (gray circle, list "PBT and/or vPvB, EU," 7 substances in total); (iv) classification as an endocrine disrupting chemical (EDC) within Registration, Evaluation, Authorization and restriction of CHemicals (REACH) legislation in the EU [font color blue] or recognition in the 2018 United Nations Environment Programme (UNEP) report as an EDC or a potential EDC identified by at least one robust stakeholder following a thorough scientific assessment based on a definition of an EDC or a potential EDC postulated by the World Health Organization (WHO)/International Program on Chemical Safety (IPCS) in 2002 [font color purple] (blue circle, list "EDC, REACH + UNEP 'at least one'," 35 substances in total; n.b., all but one REACH-classified EDC is also recognized in the UNEP report as an EDC or a potential EDC identified with involvement of multiple stakeholders); and (v) HH hazards based on advisory CLP classifications (yellow circle, list "HH, advisory CLP," 102 substances in total). The red frame encloses the substances included in the final list of 148 most hazardous chemicals likely associated with plastic packaging. The 98 substances identified as 'most hazardous' for HH based only on advisory CLP classifications are not included in this final list. Sizes of the circles are not exactly to scale.

391 392 393

394

395

396

397

398

For environmental hazards, there were only 4 different sum hazard score values (Table 1), because only two relevant categories, Aq. Acute and Aq. Chronic, were available to calculate this score. Sixty-eight substances on the CPPdb_ListA had the highest sum hazard score for environmental hazards (1100) and were thus considered to be of the highest environmental hazard based on harmonized CLP classifications.



For human health hazards, there were 80 different sum hazard score values based on harmonized CLP classifications (ranging from 32 100 to 10) within the CPPdb_ListA (Table 1). Here, 63 substances on the CPPdb_ListA with sum hazard scores equal to or higher than 10 000 were considered to have the highest human health hazard based on harmonized CLP classifications (Table 1). This range was chosen because the substances that would be included within it are certain to have at least one classification with the highest hazard grade score of 10 000, given to the most severe CLP classifications only, namely the 1A and 1B hazard classes in the carcinogenicity (Carc.), mutagenicity (Muta.), and reproductive toxicity (Repr.) categories, reflective of the so-called CMR properties (Lithner et al., 2011).

Fifteen CPPdb_ListA substances have been identified as substances of very high concern (SVHC) within REACH for their endocrine disrupting properties with regard to human health and/or the environment, either as an individual substance or belonging to a group of related substances. Fourteen of these 15 substances are also listed in the UNEP report (2018) as EDCs or potential EDCs identified after a thorough scientific assessment with involvement of multiple stakeholders (with the exception of BPA, see above). The fifteenth chemical, not inventoried by the UNEP report but classified as an EDC within REACH, is bisphenol A (BPA, CAS 80-05-7). This chemical was not included in the UNEP report because it was officially classified as an EDCs only after the final drafting of the UNEP report in July 2017. Twenty additional substances on the CPPdb_ListA are listed in the UNEP report (2018) as EDCs or potential EDCs identified after a thorough scientific assessment by at least one robust stakeholder. Further 7 substances on the CPPdb_ListA have been classified as having PBT and/or vPvB properties in Europe. Among the two latter categories, three substances have both PBT and vPvB classifications.

3.2.2 Identification of the substances most hazardous for human health based on advisory CLP classifications

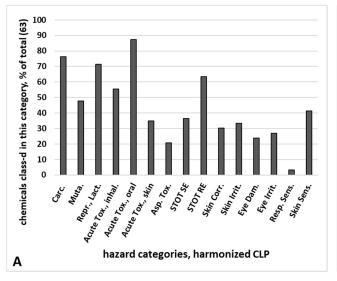
Similar to the approach taken with harmonized CLP classifications, the sum hazard scores derived based on advisory CLP classifications were used to sort the chemical lists from the highest to the lowest hazard. Based on this ranking, 102 CPPdb_ListA substances having sum hazard score values equal to or higher than 1000 were considered to exhibit the highest human health hazard. This cut-off was chosen because it covered all CMR-classified substances along with a Skin Sensitization 1 classification, also regarded as a crucial human health-related hazard (Lithner et al., 2011). Overall, CPPdb_ListA chemicals had 15 sum hazard score values calculated based on advisory CLP classifications for human health hazards, ranging from 2110 to 10 (Table 1). Thus, the maximal sum hazard scores in this case were much lower than those calculated based on the harmonized CLP classifications. This is because the most severe hazard classifications—those reflective of the CMR properties—are never assigned when only *in silico* estimations of toxicity are considered. Therefore,



modeling-based advisory CLP classifications by default cannot rise above "possible" or "predicted" status. Among the 102 most hazardous substances identified based on advisory CLP classifications, 4 chemicals have been identified as EDCs by REACH and/or UNEP report (2018) (Fig. 2).

3.2.3 Distribution of CLP hazard categories among the most hazardous substances

The 68 CPPdb_ListA substances identified as the most hazardous for the environment according to harmonized CLP data are assigned in the highest hazard class for both acute and chronic aquatic toxicity (Aq. Acute 1 and Aq. Chronic 1). Among the 63 most hazardous CPPdb_ListA substances identified based on harmonized CLP classifications for human health hazards, CMR properties were represented most often, followed by acute toxicity after oral intake or inhalation, specific target organ toxicity after a single or repeated exposure (STOT SE and STOT RE categories), and skin sensitization (Fig. 3A). Other hazard categories, such as acute toxicity after dermal exposure, skin irritation or corrosion, eye damage or irritation, were also assigned often, but these have much less weight (lower numerical hazard grade score) and therefore contributed the least to the overall sum hazard score for human health hazards. On the contrary, among the most hazardous chemicals identified based on the advisory CLP classifications for human health hazards, classifications in the CMR-related categories were assigned relatively rarely, while classifications for skin sensitization were the most frequent (Fig. 3B).



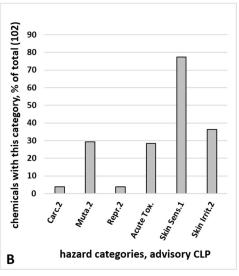


Figure 3. Distribution of hazard categories among the most hazardous substances identified based on harmonized classification, labeling and packaging (CLP) classifications for human health hazards (A, 63 substances in total) or advisory CLP classifications for human health hazards (B, 102 substances in total). Hazard categories given without numbers include all classes within the given category. Hazard categories given with numbers include on the respective class. Abbreviations: Carc.: Carcinogenicity, Muta.: Mutagenicity, Repr.: Reproductive Toxicity, Lact.: Lactation toxicity, Acute Tox., inhal.: Acute Toxicity, inhalation; Acute Tox., oral: Acute Toxicity, oral; Acute Tox., skin: Acute Toxicity, dermal; Asp. Tox.: Aspiration Toxicity; STOT SE: Specific Target Organ Toxicity Single Exposure; STOT RE: Specific Target Organ Toxicity Repeated Exposure; Skin Corr.: Skin Corrosion; Skin Irrit.: Skin Irritation; Eye Dam.: Eye Damage; Eye Irrit.: Eye Irritation; Resp. Sens.: Respiratory Sensitization; Skin Sens.: Skin Sensitization.



3.2.4 Distribution of functions among the most hazardous substances

Table 2 presents an overview of the main functions performed in plastics by the most hazardous chemicals likely associated with plastic packaging, identified as described in the subsection 3.2.1, along with the specific hazards and sum hazard scores calculated for each substance. This table also shows the specific hazards and sum hazard scores calculated for each substance. The functions represented most often were (in the order of decreasing number of substances in a group):

Monomers and intermediates, solvents, surfactants and their degradants, stabilizers, plasticizers, biocides, fire retardants, accelerators, and colorants (Table 2).

The overview in Table 2 is intended to be illustrative only and is by no means comprehensive. For clarity, we chose to show only one, presumably main, function for each chemical, although some substances could perform several different functions in plastics. Further, to avoid overcrowding, we excluded the 22 substances containing one or more of the 4 heavy metals considered most hazardous according to the European Directive 94/62/EC on packaging and packaging waste (EU, 1994), i.e., cadmium, lead, mercury, and hexavalent chromium. Therefore, only 126 out of the 148 most hazardous substances identified in 3.2.1 are listed in Table 2. Among the 22 excluded metal-containing substances, all except one (cadmium sulfide, CAS 1306-23-6) are classified into the top category for environmental hazards (sum hazard score 1100), and some are also highly hazardous to human health due to CMR properties (see Supplementary File 4).

Table 2. Main performance functions and chemical groups represented among the 111 most hazardous chemicals likely associated with plastic packaging (CPPdb_ListA), identified based on selected harmonized hazard data(*).

function	chemical group	CAS name	name		ard scores, zed CLP(**)	EDC, PBT/vPvB(***)
				ENV	HH	
accelerat	dithiocar	14324-55-1	zinc bis(diethyldithiocarbamate)	1100	1040	-
ors	bamate	136-23-2	zinc bis(dibutyldithiocarbamate)	1100	1030	-
	thiazole	97-77-8	disulfiram	1100	1110	-
	and	120-78-5	di(benzothiazol-2-yl) disulphide	1100	1000	-
	thiuram	149-30-4	2-mercaptobenzothiazole	1100	1000	-
biocides	carbamat	137-30-4	ziram	1100	2220	EDC (UNEP)
	e	55406-53-6	3-iodo-2-propynyl-N-butyl carbamate	1100	2210	-
		137-26-8	thiram	1100	1140	EDC (UNEP)
	phenolics	3380-34-5	triclosan	1100	20	EDC (UNEP)
		97-23-4	dichlorophen	1100	20	-
	organo	76-87-9	fentin hydroxide	1100	2520	-
	metallic	900-95-8	fentin acetate	1100	2520	-
		1338-02-9	naphthenic acids, copper salts	1100	10	
		155925-27-2	silver sodium hydrogen zirconium phosphate	1100	-	
	parabens	99-76-3	methyl 4-hydroxybenzoate	-	-	EDC (UNEP)
		120-47-8	ethyl 4-hydroxybenzoate	-	-	EDC (UNEP)
		94-13-3	propyl 4-hydroxybenzoate	-	-	EDC (UNEP)
colorants	dye, azo	101-77-9	4,4-methylenedianiline	100	13100	-
		95-80-7	4-methyl-m-phenylenediamine	100	12310	-
		1937-37-7	disodium 4-amino-3-[[4-[(2,4- diaminophenyl)azo][1,1-biphenyl]-4-yl]azo]-5- hydroxy-6-(phenylazo)naphthalene-2,7- disulphonate	-	10100	-



		573-58-0	disodium 3,3-[[1,1-biphenyl]-4,4-diylbis(azo)]bis(4-aminonaphthalene-1-sulphonate)	-	10100	-
	pigment, Co	71-48-7	cobalt(II) diacetate	1100	23000	-
fire	boron	10043-35-3	boric acid	-	10000	-
retardant		12179-04-3	sodium tetraborate, pentahydrate	-	10000	-
S		1303-96-4	sodium borate, decahydrate	-	10000	-
		1330-43-4	sodium tetraborate, anhydrous	-	10000	-
	organo	115-96-8	tris(2-chloroethyl) phosphate	100	10110	-
	phosphat	115-86-6	triphenyl phosphate	-	-	EDC (UNEP)
	е	25155-23-1	trixylyl phosphate	-	10000	-
	other	117-08-8	tetrachlorophthalic anhydride	1100	2100	-
		79-94-7	2,2',6,6'-tetrabromobisphenol A	1100	-	-
foaming agents	simple hydrocar bon	75-28-5	isobutane	-	20000	-
monome	acrylic	29590-42-9	isooctyl acrylate	1100	30	-
rs and		107-13-1	acrylonitrile	100	11420	-
intermed		106-91-2	glycidyl methacrylate	-	23320	-
iates		79-06-1	acrylamide	-	22240	-
	amine	108-45-2	m-phenylenediamine	1100	2310	-
		151-56-4	aziridine	100	23100	-
		111-41-1	2-((2-aminoethyl)amino)ethanol	-	11100	-
	bispheno I	80-05-7	bisphenol A	-	11110	EDC ENV, EDC HH (REACH)
		620-92-8	bisphenol F	-	-	EDC (UNEP)
		80-09-1	bisphenol S	-	-	EDC (UNEP)
	other	91-20-3	naphthalene	1100	110	-
	hydrocar	92-52-4	biphenyl	1100	30	-
	bons	75-56-9	propylene oxide	-	20230	-
		96-18-4	1,2,3-trichloropropane	-	20030	-
		106-97-8	butane	-	20000	-
		106-99-0	1,3-butadiene	-	20000	-
		50-00-0	formaldehyde	-	12400	-
		106-89-8	epichlorohydrin	-	11400	-
		126-99-8	2-chlorobuta-1,3-diene	-	10150	-
		96-09-3	(epoxyethyl)benzene	-	10020	-
		75-01-4	chloroethylene	-	10000	-
	zinc	7646-85-7	zinc chloride	1100	110	-
		7733-02-0	zinc sulphate	1100	110	-
		7440-66-6	zinc	1100	-	-
		7779-90-0	trizinc bis(orthophosphate)	1100	-	-
plasticize	chlorinat	85535-84-8	alkanes, C10-13, chloro	1100	100	PBT, vPvB
rs	ed paraffin	85535-85-9	medium-chain chlorinated paraffins, >17 carbon atoms	1100	100	-
	phthalate	85-68-7	benzyl butyl phthalate	1100	10000	EDC HH (REACH); EDC (UNEP)
		131-17-9	diallyl phthalate	1100	10	-
		84-74-2	dibutyl phthalate	100	10000	EDC HH (REACH); EDC (UNEP)
		117-81-7	bis(2-ethylhexyl) phthalate	-	10000	EDC ENV, EDC HH (REACH); EDC (UNEP)
		84-69-5	diisobutyl phthalate	-	10000	EDC HH (REACH); EDC (UNEP)
		117-82-8	dimethoxyethyl phthalate	-	10000	-
		68515-42-4	1,2-Benzenedicarboxylic acid, di-C7-11- branched and linear alkyl esters	-	10000	-
		71888-89-6	diisoheptyl phthalate	-	10000	-
		84-61-7	dicyclohexyl phthalate	-	11000	EDC (UNEP)
		84-75-3	dihexyl phthalate	-	10000	EDC (UNEP)
		84-66-2	diethyl phthalate	-	-	EDC (UNEP)
		117-84-0	dioctyl phthalate	-	-	EDC (UNEP)
		3648-40-0	diundecyl phthalate	-	-	EDC (UNEP)
		26761-40-0	di-isodecyl phthalate	-	-	EDC (UNEP)
solvents	limonene	138-86-3	dipentene	1100	1010	_



		5989-27-5	(R)-p-mentha-1,8-diene	1100	1010	-
	naphtha-	8052-41-3	stoddard solvent	-	21100	-
	related	64741-65-7	naphtha (petroleum), heavy alkylate	-	20100	-
		64741-66-8	naphtha (petroleum), light alkylate	-	20100	-
		64742-49-0	naphtha (petroleum), hydrotreated light	-	20100	_
		64742-95-6	solvent naphtha (petroleum), light arom.	_	20100	_
		8030-30-6	naphtha	_	20100	_
		64742-46-7	distillates (petroleum), hydrotreated middle	-	10000	
		64742-52-5	distillates (petroleum), hydrotreated heavy		10000	
			naphthenic	-		-
		64742-54-7	distillates (petroleum), hydrotreated heavy paraffinic	-	10000	-
		8009-03-8	petrolatum	-	10000	-
	pure	110-82-7	cyclohexane	1100	120	-
	hydrocar	142-82-5	heptane	1100	120	-
	bons	79-01-6	trichloroethylene	10	11030	_
		71-43-2	benzene	-	21120	_
		109-86-4	2-methoxyethanol	-	10030	-
			dimethylformamide	-		-
	_	68-12-2		-	10030	-
stabilizer	tin	683-18-1	dibutyltin dichloride	1100	13210	-
S		77-58-7	dibutyltin dilaurate	-	12000	-
		15571-58-1	2-ethylhexyl 10-ethyl-4,4-dioctyl-7-oxo-8-oxa-	-	10000	-
			3,5-dithia-4-stannatetradecanoate			
	organic	26523-78-4	tris(nonylphenyl) phosphite	1100	1000	-
	phosphit e	101-02-0	triphenyl phosphite	1100	20	-
	hindered	128-37-0	butylated hydroxytoluene (BHT)	-	_	EDC (UNEP)
	phenol	25013-16-5	butylated hydroxyanisole (BHA)	_	_	EDC (UNEP)
	-	131-56-6	2,4-dihydroxybenzophenone	-	_	EDC (UNEP)
	benzoph enone				-	` '
	enone	131-55-5	2,2',4,4'-tetrahydroxybenzophenone	-	-	EDC (UNEP)
		131-57-7	2-hydroxy-4-methoxybenzophenone	-	-	EDC (UNEP)
	benzotria zol	36437-37-3	2-(2H-benzotriazol-2-yl)-4-(tert-butyl)-6-(sec- butyl)phenol	-	-	vPvB
		3864-99-1	Phenol, 2-(5-chloro-2H-benzotriazol-2-yl)-4,6-bis(1,1,-dimethylethyl)-	-	-	vPvB
		25973-55-1	2-(2'-hydroxy-3,5'-di-t- amylphenyl)benzotriazole	-	-	PBT, vPvB
		3846-71-7	2-benzotriazol-2-yl-4,6-di-tert-butylphenol			PBT, vPvB
	other	2451-62-9	1,3,5-tris(oxiranylmethyl)-1,3,5-triazine-	10	11400	-
	Other		2,4,6(1H,3H,5H)-trione			
		122-39-4	diphenylamine	1100	400	-
		1314-13-2	zinc oxide	1100	-	-
surfactan ts and	NP, OP, and NP-	84852-15-3	phenol, 4-nonyl-, branched	1100	210	EDC ENV (REACH); EDC (UNEP)
their	related	25154-52-3	nonylphenol	1100	210	Same as above
degradati		140-66-9	4-tert-Octylphenol	1100	110	Same as above
on		104-40-5	p-nonylphenol	-	-	Same as above
products		127087-87-0	4-nonylphenol, branched, ethoxylated	-	-	Same as above
		26027-38-3	nonoxynol-1	_	_	Same as above
		37205-87-1	Isononylphenol ethoxylate	-	-	Same as above
		7311-27-5	2-[2-[2-(4-	-	-	Same as above
			nonylphenoxy)ethoxy]ethoxy]ethoxy]ethanol			
		68412-54-4	nonylphenol, branched, ethoxylated	-	-	Same as above
		9016-45-9	nonylphenol, ethoxylated	-	-	Same as above
	amine	61788-46-3	amines, coco alkyl	1100	320	-
	and N-	112-90-3	(Z)-octadec-9-enylamine	1100	320	-
	containin	61788-45-2	amines, hydrogenated tallow alkyl	1100	310	-
	g	61790-33-8	amines, tallow alkyl	1100	310	-
		2687-96-9	N-dodecyl-2-pyrrolidone	1100	1100	-
		107-64-2	dimethyldioctadecylammonium chloride	1100	100	-
1	PFAS	335-67-1	perfluorooctanoic acid	100	11320	PBT
		3825-26-1	ammonium pentadecafluorooctanoate	100		

(*) Harmonized hazard data reviewed included: (i) harmonized classification, labeling and packaging (CLP) classifications for environmental hazards (ENV) and human health (HH) hazards; (ii) classification as endocrine disrupting chemical (EDC) for ENV or HH effects within Registration, Evaluation, Authorization and restriction of Chemicals (REACH) legislation or recognition in the 2018 report by the United Nations Environment Programme (UNEP) as EDC or potential EDC identified with involvement of multiple stakeholders (corresponds to all but one REACH-identified EDCs) or at least one robust stakeholder following a thorough scientific assessment based on a definition of an



491 EDC or a potential EDC postulated by the World Health Organization (WHO)/International Program on Chemical Safety (IPCS) in 2002; and (iii) classification in the European Union (EU) as a substance with persistent, bioaccumulative, and toxic (PBT) and/or very persistent, very bioaccumulative (vPvB) properties.

(**) Red background highlights properties exhibited by the most hazardous substances selected according to (sum hazard score ENV)=1100 and/or (sum hazard score HH)≥10,000; yellow background highlights ranges 100≤(sum hazard score ENV)<1000 and 1000≤(sum hazard score HH)<10,000; blue background highlights ranges (sum hazard score ENV)<100 and (sum hazard score HH)<1000; gray background: no classifications found.

(***) Red background identifies the most hazardous substances selected based on the classification/identification as EDC, PBT, or vPvB substance; gray background: no classifications/identifications found.

500

501

502

503

504

505

506

507

494

495

496

497

498

499

4. Discussion

This study aimed to (i) compile a comprehensive and publicly accessible database of chemicals used or found in plastic packaging, (ii) provide hazard information for these chemicals with regard to human health and the environment, and (iii) identify the most hazardous substances in need of further assessment as potential candidates for substitution. Achieving the study's aims was significantly hindered by substantial information and data gaps on chemical use, levels of chemicals in finished plastic packaging products, and hazards of chemicals associated with plastic packaging.

508 509

510 511

512

513

514

515 516

517

518

519

520

521

522

523

524

525 526

527

528

4.1 Challenges and information requirements

Two major challenges hamper the identification of chemicals associated with plastic packaging: (i) The lack of publicly accessible comprehensive registries for chemicals used in plastic packaging, and (ii) use restrictions for commercial data sources. This also impacted on the verification of a substance's current use in plastic packaging. Multiple commerce-oriented online-accessible inventories exist, such as global product databases or smaller repositories maintained by individual companies producing plastics and plastics additives. However, such sources usually give little information on the chemical composition of formulated products, concentrating almost exclusively on physical properties such as material performance or compatibility of different products. Furthermore, industry sources tend to limit their users in accessing and retrieving chemicals-related information (whichever is available) for multiple products or product categories simultaneously. Instead, users can view or extract information for only a few products at a time. In addition, users are often requested to demonstrate a commercial interest in order to be allowed to retrieve detailed chemical information or continue any data-mining on the website. Thus, even though commercial websites and companies clearly could be or are in possession of potentially detailed chemicalsrelated information, they strongly limit access to this information for non-commercial, academic research purposes. On the other hand, information contained in publicly accessible sources that offer a comparatively easy retrieval of data, such as the CPCat database (Dionisio et al., 2015), is often incomplete or insufficiently detailed. Filling the data gaps by collecting and recording information scattered across multiple literature sources such as research manuscripts, books, or reports is highly



time-consuming and not achievable within the scope of one project. Furthermore, information collected from such sources could still prove to be incomplete, incorrect, or outdated. For example, several publications summarizing marketing information for plastics additives do provide some use statistics, but only for groups of additives and not for individual chemicals or specific applications (Levy et al., 2001; PlasticsEurope, 2016).

Consequently, in the course of this study we were able to assign with some certainty the likely association with plastic packaging to less than a quarter (906 substances) of the 4283 chemicals included in the CPPdb. For the remaining 3377 CPPdb substances, no final conclusions regarding their association with plastic packaging could be drawn. Moreover, when compiling the CPPdb, we excluded 7510 more substances originally retrieved from the CPCat database, because no indication of their use in plastics was found in the 14 initial information sources we have consulted (Fig. 1A). However, some of these chemicals could still prove to be used in plastic packaging upon closer examination. Adhesives, coatings, and inks represent product categories exhibiting particularly severe information gaps with regard to their constituent chemicals' use in plastic packaging.

In agreement with earlier studies (Bilitewski et al., 2012a; Rossi and Blake, 2014), our work demonstrated a significant lack of detailed information concerning use of chemicals in plastics manufacturing and the chemicals' presence in final products, especially for some additives and the often unpredictable NIAS (Nerin et al., 2013; van Oers et al., 2011). This lack of publicly accessible information is prominent even for those products that are in direct contact with foods and therefore can be assumed to directly contribute to population-wide human exposure (Muncke et al., 2017). More transparency regarding the exact chemical composition of marketed products might improve the situation in the future, but currently the ability to perform accurate risk assessment, i.e., incorporating an exposure component, is rather limited. Therefore, hazard-based assessment remains the approach of choice when dealing with large numbers of chemicals potentially present in consumer products.

Significant data gaps also exist for hazard information. For example, less than a third of the chemicals likely associated with plastic packaging had harmonized hazard classifications in the consulted sources (see Table 1). However, over two hundred plastic packaging-associated chemicals lacking a harmonized CLP classification had advisory CLP classifications assigned by the Danish EPA based on quantitative structure-activity relationship (QSAR) models. Thus, these chemicals might be hazardous as well, but are not yet officially classified as such, possibly due to the fact that confirming the predicted hazardous properties requires experimental toxicity testing data which are usually lacking. This example shows that a comprehensive hazard assessment should include data sources other than the harmonized hazard classifications.



For instance, hazard or risk assessments, conducted based on peer-reviewed literature and then made publicly available, could serve to inform and guide substitution efforts. An example of such an initiative is the SIN list (http://chemsec.org/sin-list/) maintained by the NGO *ChemSec*. However, there are significant obstacles to such an approach, such as funding limitations, or issues related to primary research communication and study quality (Agerstrand et al., 2017), but also integration and regulatory acceptance of alternative toxicity assessment methods, such as *in silico* predictions or *in vitro* tests (Piersma et al., 2018). Furthermore, peer-reviewed literature also tends to be biased towards the better-known substances, and switching the academic attention to the next group of 'emerging' substances tends to have a lengthy lag period (Bao et al., 2015; Li et al., 2018). Despite these challenges, peer-reviewed literature deserves to be given proper consideration in hazard assessment and prioritization studies (Kaltenhäuser et al., 2017; Myers et al., 2010). Increasingly, this is being done using a systematic review methodology, originally developed in the medical field, and now being gradually implemented for toxicological assessments as well (Birnbaum et al., 2013; Hoffmann et al., 2017; Morgan et al., 2016; Vandenberg et al., 2016).

Efficient communication and collaboration are also of crucial importance for ensuring the success of current efforts aimed at identification of hazardous chemicals and promoting their substitution with safer alternatives. One resource developed to facilitate the exchange of information on the use patterns and hazards of chemicals is the Chemical Hazards Data Commons (https://healthybuilding.net/content/data-commons/). This resource brings together the information from multiple lists and resources on hazardous chemicals and provides features supporting communication between different stakeholders. The CPPdb lists will be uploaded to this resource to enable collaborations and open dialogue to support further refinement of the inventory of plastic packaging-associated chemicals, as well as understanding the associated hazards and supporting the search for safer alternatives.

With regard to the generation of new data on the health hazards of plastic packaging-associated chemicals, the focus of toxicity testing may need to shift from assessing individual substances towards looking at the mixtures of chemicals present in a finished packaging article (Muncke, 2014). The advantage would be the ability to evaluate the effects of unknown NIAS, and account for mixture toxicity of multiple substances. Thus, the overall chemical extracts or migrates from a given packaging article could be tested for multiple types of toxicity using cell-based *in vitro* systems (Groh and Muncke, 2017; Severin et al., 2017). Besides the ability to increase the testing throughput, use of *in vitro* systems also allows achieving insights into additional effects and toxicity mechanisms not yet accounted for by the hazard categories covered in the harmonized hazard classification systems and traditional toxicity tests, such as neurobehavioral disorders (Maffini and Neltner, 2014), gut-related



ailments (Groh et al., 2017), metabolic disruption (Heindel et al., 2017), and endocrine disruption (Wagner, 2016).

4.2 Overview of the most hazardous chemicals associated with plastic packaging The discussion below will focus on the most hazardous chemicals identified based on the harmonized CLP classifications for environmental and human health hazards, as well as EU-accepted EDC and PBT/vPvB classifications, and recognition as an EDC or potential EDC in the UNEP report on EDCs (2018). One major limitation of this study is its reliance on harmonized hazard data which was available for less than a half of all substances included in the CPPdb. Consequently, the performed ranking could identify only the already known hazardous chemicals, while other chemicals with equal or more severe hazard properties may be overlooked due to the absence of harmonized hazard data.

Over 20 chemicals used as monomers and intermediates in plastics production, and roughly the same number of substances likely used as solvents, were among the most hazardous plastic packaging-associated substances identified in this work. Environmental and human health hazards of plastics monomers and solvents frequently used in plastics manufacturing have been addressed before (Lithner et al., 2011). When judged based on their monomers, some polymers used in packaging, including PS, PVC, PC, and PU (the latter used often in adhesives), are regarded as highly hazardous, while polyolefins and PLA are considered to be of lower hazard (Rossi and Blake, 2014). However, uncompounded polymers are rarely used in final applications, as various additives are usually added to modify polymer properties. If hazardous, these substances can lend hazard properties to even a seemingly safe polymeric material. Indeed, we observed that the majority of plastic packaging-associated substances identified as the most hazardous for environmental and human health were in fact plastics additives representing diverse chemical groups used for a variety of functions.

A particularly prominent group of hazardous additives consisted of substances containing metals, including cadmium, chromium, lead, mercury, cobalt, tin, and zinc. Apart from frequent classifications for aquatic toxicity, some of these chemicals also ranked high with regard to human health hazards, mostly because of CMR properties. Cadmium- and zinc-containing substances have been used as stabilizers in PVC and some other plastics (zinc is safer but less efficient than cadmium), while mercury-containing chemicals can be applied as catalysts in the production of some plastics and rubbers. Many metal-containing substances are also used as colorants, and some as antimicrobials and accelerators (Zweifel et al., 2009). The use and presence of the four most hazardous metals—cadmium, hexavalent chromium, lead, and mercury—in packaging is regulated in the EU (EU, 1994) and 19 U.S. states. However, despite the regulations, toxic metals, especially cadmium, have been detected at levels exceeding the regulatory limits in some PVC packaging



633

634

635

636

637

638

639

640

641

642

643

644

645

646

647

648

649

650

651

652

653

654

655

656

657

658

659

660

661

662

663

664

665

666

samples obtained from U.S. retailers (Toxics in Packaging Clearinghouse, 2017; van Putten, 2011). Notably, most of the non-compliant packaging items identified in the U.S. appeared to be imported, often from China. Indeed, in other parts of the world, especially in developing countries, the use and presence of heavy metals in plastic packaging is either not regulated, or regulated insufficiently, or regulations are not properly enforced. For example, PE bags in Uganda contain cadmium, chromium, cobalt, and lead, found to contaminate food cooked in these bags (Musoke et al., 2015), while in Brazil lead was detected in some HDPE packaging samples (Kiyataka et al., 2014). Plastics beached on the shores of a fresh water lake in Europe (Filella and Turner, 2018) or the Pacific ocean (Munier and Bendell, 2018), most often composed of PVC, polyolefins, and PS, were also found to contain multiple heavy metals, including cadmium, mercury, and lead. Although plastics fragments are known to absorb metals present in the environment (Munier and Bendell, 2018), it was also suggested that some of the detected metals could in fact be "legacy" chemicals contained in the nondegraded plastics originating from the times before regulatory restrictions (Filella and Turner, 2018). Surfactants and their degradation products form another large group of highly hazardous substances likely associated with plastic packaging. Surfactants are used in a variety of applications in plastics, for example, as wetting or antistatic agents. Some surfactants can also be used as dispersion agents in biocidal or colorant formulations. Both biocidal and coloring chemicals are also well represented among the hazardous substances likely associated with plastic packaging. Most biocides are classified as environmental hazards, but also have some classifications for human health hazards. Commodity plastics used in packaging applications are usually based on polymers that are not biodegradable and therefore resistant to the attack by microorganisms. Biocides added to such plastics, if any, are intended to prevent microbial degradation of some of the additives, thus their amounts are usually small. However, in PVC plastics, which can contain close to 50% by weight of phthalate plasticizers (Freire et al., 2006; Kawakami et al., 2011), known to be susceptible to microbial attack (Latorre et al., 2012), application of biocides can also be more significant. In addition, biocides are intentionally used to impart antimicrobial properties in the products belonging to a growing family of functional, smart, or active, packaging (Larson and Klibanov, 2013; Malhotra et al., 2015; Nguyen Van Long et al., 2016). Furthermore, biocides can be added to biodegradable plastics to protect the product from premature degradation (Fink, 2014), however, information on the types and amounts of biocides used in biodegradable plastics is rather scarce (Harrison et al., 2018; Lambert and Wagner, 2017). This area requires more research. Apart from the substances showing a high hazard score based on the harmonized CLP classifications, substances possessing endocrine disrupting, PBT, or vPvB properties are also considered hazardous and therefore of immediate concern (Halden, 2010; Muncke et al., 2014; North and Halden, 2013). To identify the EDCs among the chemicals associated with plastic



668

669

670

671

672

673

674

675

676

677

678

679

680

681

682

683

684

685

686

687

688

689

690

691

692

693

694

695

696

697

698

699

700

701

packaging, we relied on the substance's identification as an EDC within the REACH legislation (EU, 2008) along with a recognition as an EDC or potential EDC in the UNEP report on EDCs (2018). Based on the REACH EDC classification source (also recognized in the UNEP report as EDCs assessed with involvement of multiple stakeholders), 15 substances likely associated with plastic packaging were identified as EDCs. Only 8 of these substances also had a high CLP-based sum hazard score for environmental or human health hazards. Thus, without an additional consideration of endocrine disrupting properties, 7 of these 15 chemicals would not have been ranked as hazardous. Similarly, of the 20 additional chemicals recognized in the UNEP report (2018) as EDCs or potential EDCs identified by at least one robust stakeholder, 15 chemicals would not have been ranked as hazardous if the endocrine disrupting properties were not considered. In this regard, it is important to point out that some chemicals that have been suggested as substitutes for the recognized hazardous substances happen to be structurally similar to the original offenders, suggesting that they could have similar hazardous properties. This notion proved to be particularly true with regard to endocrine disruption. For example, BPA, a monomer in some PC plastics and epoxy coatings, has recently been classified as an EDC under REACH, in addition to several harmonized CLP classifications for human health hazards this chemical already had. To replace BPA in some applications, other bisphenols are now used as alternatives, for example, bisphenol S (BPS, CAS 80-09-1) and bisphenol F (BPF, CAS 620-92-8). However, since the endocrine activities displayed by these (and other) bisphenols are similar to those of BPA (Goldringer et al., 2015; Rochester and Bolden, 2015), the safety of such substitutions may be questioned (Rosenmai et al., 2014). Indeed, these substances are already recognized in the UNEP report (2018) as EDCs or potential EDCs identified by at least one robust stakeholder, however, the REACH classification is apparently lagging behind. Similar considerations concern the ortho-phthalate group, as some phthalates are classified as EDCs within REACH, while others, many of them recognized in the UNEP report (2018), lack such a classification to date.

Several other groups of plastic packaging-associated chemicals that have not yet been identified as EDCs within the REACH regulation, but are recognized in the UNEP report (2018), include parabens used as preservatives (Berger et al., 2015), benzophenones used as UV stabilizers (Cwiek-Ludwicka and Ludwicki, 2014; Simon et al., 2016), and two hindered phenols used as antioxidants, butylated hydroxytoluene (BHT, CAS 128-37-0) and butylated hydroxyanisole (BHA, CAS 25013-16-5). Among these groups, the two latter substances are used most extensively in high production volume commodity plastics such as polyolefins (Tolinski, 2009). Both BHT and BHA are estrogenic *in vitro* (Miller et al., 2001; Pop et al., 2018), and several other endocrine-related effects have been observed as well (Rajamani et al., 2017). *In vivo*, variable alterations have been reported, requiring further research to better understand their significance (Pop et al., 2013).



703

704

705

706

707

708

709

710

711

712

713

714

715

716

717

718

719

720

721

722

723

724

725

726

727

728

729

730

731

732

733

734

735

736

The PBT and/or vPvB substances associated with plastic packaging include the group of four benzotriazole stabilizers, two of them classified as PBT and two as vPvB substances. These chemicals were shown to accumulate in fish and birds, with patterns of bioaccumulation suggested to depend on food sources and the presence of different plastics within those food sources (Lu et al., 2018). The two other PBT-classified chemicals belong to the family of per-and polyfluoroalkyl substances (PFAS). Many other PFAS show capacity for persistence, and possibly bioaccumulation and toxicity as well (Blum et al., 2015; Cousins et al., 2016; Scheringer et al., 2014; Wang et al., 2015), however, not all of them have been classified or identified as SVHCs in the EU for these properties yet (Brendel et al., 2018). Similarly, chlorinated paraffins with the carbon chain length of 10-13 (short-chain) have a PBT classification, but not the same type of chemicals with longer carbon chains (Glüge et al., 2018). These examples and those discussed above for groups of structurally similar chemicals showing comparable endocrine disrupting properties underscore the importance of assessing and classifying substances by looking at the groups formed according to structural and/or functional similarity and not only one-by-one at individual substances (Blum, 2016). Currently, tools are being developed allowing to weed-out structurally similar chemicals with likely similar toxicity when searching for alternatives, for example, the SINimilarity tool developed by ChemSec (http://sinimilarity.chemsec.org). Regulatory agencies are also increasingly applying grouping-based approaches (OECD, 2014; Schultz et al., 2015) to perform hazard and risk assessments for certain types of chemicals, for example, pesticides (Boobis et al., 2008; EFSA, 2014) and food flavorings (Schrankel, 2004), and there are plans to extend the application of this approach to cover more chemicals in the future (ECHA, 2018; Kienzler et al., 2016; Swedish Chemicals Agency (KEMI), 2015).

Apart from intentionally used substances and well characterized break-down products of plastics additives, important environmental or human health hazards may also be associated with (other) NIAS. These include a multitude of substances not yet completely identified (Bradley and Coulier, 2007; Paseiro-Cerrato et al., 2016; Vera et al., 2018), but also some better known chemicals that may nonetheless be difficult to manage, such as impurities and contaminants. For example, PAHs are often found in the PS polymer or in some additives, e.g., carbon black (Li et al., 2017), and 'legacy' chemicals such as flame retardants have been detected in recycled plastics (Andra et al., 2012; Filella and Turner, 2018; Leslie et al., 2016; Pivnenko et al., 2017; Rani et al., 2014; Samsonek and Puype, 2013). Many of these chemicals are classified as carcinogenic or have other highly hazardous properties, while others have never been tested. Though it can be difficult to control these impurities in the first place, their presence in plastics should not be simply disregarded but rather seen as an additional aspect to be considered when making decisions on the (future) use of a particular polymer for a specific application. Paying attention to impurities and contaminants is also important when evaluating the potential health effects of increasing levels of recycled content in



738

739

740

741

742

743

744

745

746

747

748

749

750

751

752

753

754

755

756

757

products (Geueke et al., 2018). Awareness of contamination may incentivize modification of the manufacturing or recycling processes, substitution of hazardous additives, or even reconsideration of the likely contaminated plastics types for use in packaging applications as a whole.

Some of the substances highlighted above are already regulated to some extent in the EU or U.S., but regulations worldwide may vary widely. For substances regulated within REACH, use in Europe may be on the decline. However, REACH only partially regulates food packaging products in the EU, as these are covered instead by a separate food contact legislation (EU, 2004, 2011). Consequently, such products are not subject to some of the regulations enacted within REACH. For example, some phthalates restricted within REACH are still authorized for use in food contact and thus could still be used in food packaging (an ongoing EFSA-led assessment of several phthalates has not been concluded yet). Many more chemicals are pending REACH classification or have not yet been evaluated under REACH. Furthermore, REACH regulation only applies to products marketed in the EU, but plastic packaging is produced and consumed globally, and distribution of environmental pollutants does not observe country borders (Gallo et al., 2018). Therefore, sharing information on (hazardous) chemicals in these products is highly relevant worldwide, and concerted actions are needed to tackle the associated health problems. We expect that the publication of the CPPdb on Chemical Hazards Data Commons resource will enable continuation of the work initiated here. Urgent collaborative efforts, ideally involving many different stakeholders including industry, are required to decrease the uncertainty with regard to chemical composition of packaging plastics, and the chemical industry has already signaled its intent for information sharing across sectors (Stringer, 2018). This will enable improved understanding, assessment, and management of the environmental and human health risks posed by plastic packaging-associated chemicals.

758759

760 761

762

763

764

765

766

767

768

769

770

771

5. Conclusions

In this study we compiled a database, the CPPdb, listing hundreds of substances that are likely or possibly associated with plastic packaging. The CPPdb contains substances used in manufacturing and/or present in final plastic packaging articles, including selected NIAS. Some of the substances in the CPPdb are known to be hazardous for environmental and/or human health, with harmonized hazard classification data available. For some of the key hazardous chemicals identified in this study, more detailed analyses should be performed in the future, including an assessment of the availability of alternatives. However, we faced numerous data gaps that hinder a comprehensive hazard and risk assessment of chemicals in plastic packaging. First, there is a substantial shortage of, or lack of access to, information on how specific chemicals are used, or which chemicals are used in what application and in what quantities, and at which levels they are present in finished plastic packaging. Insufficient information on chemicals' use patterns prevents any scientific, exposure-based assessments, since



773

774

775776

777

778

779

780

781

782

783

784

785

786

787

788

789

790

791

filling these data gaps using a systematic, scientific approach is nearly impossible for anyone outside industry. However, chemical risk assessments are essential for assessing impacts on human health and the environment. Reliable risk assessments should be based on actual data and not on estimates or assumptions. Therefore, there is an urgent need for publicly available information on the use of chemicals in plastics, and the exact chemical composition of finished plastics articles. Second, harmonized toxicological information, such as CLP hazard classifications, is currently not available for many chemicals that are associated with plastic packaging, even for substances for which hazards have been identified and characterized in academic studies. The incompleteness of harmonized CLP classifications for many chemicals affects the hazard ranking performed in this study, as such substances are excluded. The peer-reviewed literature and in silico predictions can provide hazard information for additional substances and should therefore be integrated into comprehensive hazard assessments. Third, many of the substances that are likely or possibly associated with plastic packaging lack any publicly available hazard data at all. Fourth, the chemical inventory-based approach taken in this study does not comprehensively address the issue of NIAS, since many of them remain unidentified and therefore cannot be risk-assessed as individual substances. Novel approaches, such as toxicity testing for the overall migrate from finished plastic packaging, address the issue of unidentified NIAS, and in addition deal with the challenge of mixture toxicity. In the future, in vitro toxicity assays could be used to test the safety of finished packaging articles. Therefore, broader application and further development of such effect-based testing approaches is desirable to guide the substitution efforts and ensure the toxicological safety of plastic packaging in the circular economy.

792 793

794 795

796

797

798

Acknowledgements

Work on this project was financed by a grant from MAVA foundation. K.G., B.G., and J.M. are employees of Food Packaging Forum Foundation (FPF). They were not restricted in any way in their work on this article. FPF is funded by unconditional donations and project-based grants. Donors and funders have no influence on the conception and execution of FPF's work. B.C.A. is financed by the Swedish research council, FORMAS grant 216-2012-625.

799 800 801

802

803

804 805

806

807

808

References

Agerstrand, M., Sobek, A., Lilja, K., Linderoth, M., Wendt-Rasch, L., Wernersson, A.-S., Ruden, C., 2017. An academic researcher's guide to increased impact on regulatory assessment of chemicals. ES:PI 19, 644-655.

Andra, S.S., Makris, K.C., Shine, J.P., Lu, C., 2012. Co-leaching of brominated compounds and antimony from bottled water. Environ. Int. 38, 45-53.

Andrady, A.L., Neal, M.A., 2009. Applications and societal benefits of plastics. Philosophical Transactions of the Royal Society B: Biological Sciences 364, 1977-1984.



Ash, M., Ash, I., 2008. Handbook of food packaging chemicals and materials. Synapse Information Resources, Endicott, NY, USA.

Bao, L.-J., Wei, Y.-L., Yao, Y., Ruan, Q.-Q., Zeng, E.Y., 2015. Global trends of research on emerging contaminants in the environment and humans: a literature assimilation. Environ. Sci. Pollut. Res. 22, 1635-1643.

Berger, E., Potouridis, T., Haeger, A., Puettmann, W., Wagner, M., 2015. Effect-directed identification of endocrine disruptors in plastic baby teethers. J. Appl. Toxicol. 35, 1254-1261.

Bilitewski, B., Darbra, R.M., Barcelo, D., (eds.), 2012a. Global risk-based management of chemical additives I: Production, usage and environmental occurrence. Springer, Berlin Heidelberg.

Bilitewski, B., Darbra, R.M., Barcelo, D., (eds.), 2012b. Global risk-based management of chemical additives II: Risk-based assessment and management strategies. Springer-Verlag Berlin Heidelberg, Berlin.

Birnbaum, L.S., Thayer, K.A., Bucher, J.R., Wolfe, M.S., 2013. Implementing Systematic Review at the National Toxicology Program: Status and Next Steps. Environmental Health Perspectives 121, a108-a109.

Biryol, D., Nicolas, C.I., Wambaugh, J., Phillips, K., Isaacs, K., 2017. High-throughput dietary exposure predictions for chemical migrants from food contact substances for use in chemical prioritization. Environ. Int. 108, 185-194.

Blum, A., 2016. Tackling toxics. Science 351, 1117-1117.

Blum, A., Balan, S.A., Scheringer, M., Trier, X., Goldenman, G., Cousins, I.T., Diamond, M., Fletcher, T., Higgins, C., Lindeman, A.E., Peaslee, G.F., de Voogt, P., Wang, Z., Weber, R., 2015. The Madrid statement on poly- and perfluoroalkyl substances (PFASs). Environ. Health Perspect. 123, A107-A111.

Bodar, C., Spijker, J., Lijzen, J., Waaijers-van der Loop, S., Luit, R., Heugens, E., Janssen, M., Wassenaar, P., Traas, T., 2018. Risk management of hazardous substances in a circular economy. J. Environ. Manag. 212, 108-114.

Bolgar, M., Hubball, J., Groeger, J., Meronek, S., 2016. Handbook for the chemical analysis of plastic and polymer additives. CRC Press, Taylor & Francis Group, Boca Raton, U.S.

Boobis, A.R., Ossendorp, B.C., Banasiak, U., Hamey, P.Y., Sebestyen, I., Moretto, A., 2008. Cumulative risk assessment of pesticide residues in food. Toxicology Letters 180, 137-150.

Bradley, E.L., Coulier, L., 2007. An investigation into the reaction and breakdown products from starting substances used to produce food contact plastics. Food Standards Agency FD 07/01.

Brendel, S., Fetter, É., Staude, C., Vierke, L., Biegel-Engler, A., 2018. Short-chain perfluoroalkyl acids: environmental concerns and a regulatory strategy under REACH. Environmental Sciences Europe 30, 9.

Brouwer, M.T., Thoden van Velzen, E.U., Augustinus, A., Soethoudt, H., De Meester, S., Ragaert, K., 2018. Predictive model for the Dutch post-consumer plastic packaging recycling system and implications for the circular economy. Waste Management 71, 62-85.

Caporossi, L., Papaleo, B., 2017. Bisphenol A and Metabolic Diseases: Challenges for Occupational Medicine. International Journal of Environmental Research and Public Health 14, 959.

Centers for Disease Control and Prevention (CDC), 2018. Fourth National Report on human exposure to environmental chemicals. Updated Tables, March 2018, in: U.S.D.o.H.a.H. Services (Ed.), U.S.

ChemSec, 2015. The 32 to leave behind: The most well-founded list of EDCs relevant for REACH, in: International Chemical Secretariat (ChemSec) (Ed.).

Cousins, I.T., Vestergren, R., Wang, Z., Scheringer, M., McLachlan, M.S., 2016. The precautionary principle and chemicals management: The example of perfluoroalkyl acids in groundwater. Environment International 94, 331-340.

Cwiek-Ludwicka, K., Ludwicki, J.K., 2014. Endocrine disruptors in food contact materials: is there a health threat? Rocz Panstw Zakl Hig 65, 169-177.

Dematteo, R., Keith, M.M., Brophy, J.T., Wordsworth, A., Watterson, A.E., Beck, M., Rochon, A., Michael, F., Jyoti, G., Magali, P., Dayna, R., Scott, N., 2013. Chemical Exposures of Women

Workers in the Plastics Industry with Particular Reference to Breast Cancer and Reproductive Hazards. NEW SOLUTIONS: A Journal of Environmental and Occupational Health Policy 22, 427-448.

Dionisio, K.L., Frame, A.M., Goldsmith, M.-R., Wambaugh, J.F., Liddell, A., Cathey, T., Smith, D., Vail, J., Ernstoff, A.S., Fantke, P., Jolliet, O., Judson, R.S., 2015. Exploring consumer exposure pathways and patterns of use for chemicals in the environment. Tox. Rep. 2, 228-237.

ECHA, 2018. Authorities to focus on substances of potential concern. Roadmap for SVHC identification and implementation of REACH risk management measures., Annual Report.

EFSA, 2014. Scientific Opinion on the identification of pesticides to be included in cumulative assessment groups on the basis of their toxicological profile. EFSA J. 11, 3293.

EU, 1994. European Parliament and Council Directive 94/62/EC of 20 December 1994 on packaging and packaging waste. Official Journal L 365, Brussels.

EU, 2004. REGULATION (EC) No 1935/2004 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL of 27 October 2004 on materials and articles intended to come into contact with food and repealing Directives 80/590/EEC and 89/109/EEC, in: T.E.P.a.t. Council (Ed.), 2004R1935, Brussels.

EU, 2006. Regulation (EC) No 1907/2006 of the European Parliament and of the Council of 18 December 2006 concerning the Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH), establishing a European Chemicals Agency, amending Directive 1999/45/EC and repealing Council Regulation (EEC) No 793/93 and Commission Regulation (EC) No 1488/94 as well as Council Directive 76/769/EEC and Commission Directives 91/155/EEC, 93/67/EEC, 93/105/EC and 2000/21/EC, Official Journal of the European Union.

EU, 2008. REGULATION (EC) No 1272/2008 OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL of 16 December 2008 on classification, labelling and packaging of substances and mixtures, amending and repealing Directives 67/548/EEC and 1999/45/EC, and amending Regulation (EC) No 1907/2006. Official Journal of the European Union.

EU, 2011. COMMISSION REGULATION (EU) No 10/2011 of 14 January 2011 on plastic materials and articles intended to come into contact with food. Offic. J. EU L 12/1.

EU, 2018. Directive (EU) 2018/852 of the European Parliament and of the Council of 30 May 2018 amending Directive 94/62/EC on packaging and packaging waste (Text with EEA relevance). Official Journal of the European Union, Brussels.

European Commission, 2018. Communication from the Commission to the European Parliament, the Council, the European Economic and Social Committee and the Committee of the Regions: A European strategy for plastics in a circular economy, Brussels.

Filella, M., Turner, A., 2018. Observational Study Unveils the Extensive Presence of Hazardous Elements in Beached Plastics from Lake Geneva. Front. Environ. Sci. 6.

Fink, J.K., 2014. The chemistry of bio-based polymers. Scrivener Publishing, Wiley, Beverly, USA.

Flick, E.W., 2002. Plastics Additives: An industrial guide. Noyes Publications William Andrew Publishing, Norwich, New York, USA.

Flick, E.W., 2004. Plastics additives database. Plastics Design Library.

Franchini, M., Rial, M., Buiatti, E., Bianchi, F., 2004. Health effects of exposure to waste incinerator emissions: A review of epidemiological studies. Ann. Ist. Super Sanita 40, 101-115.

Freire, M.T.D.A., Santana, I.A., Reyes, F.G.R., 2006. Plasticizers in Brazilian food-packaging materials acquired on the retail market. Food Add. & Contam. 23, 93-99.

Fucic, A., Galea, K.S., Duca, R.C., Yamani, M.E.I., Frery, N., Godderis, L., Halldorsson, T.I., lavicoli, I., Ndaw, S., Ribeiro, E., Viegas, S., Moshammer, H., 2018. Potential health risk of endocrine disruptors in construction sector and plastics industry: A new paradigm in occupational health. J. Environ. Res. Public Health 15, 1229.

Gallo, F., Fossi, C., Weber, R., Santillo, D., Sousa, J., Ingram, I., Nadal, A., Romano, D., 2018. Marine litter plastics and microplastics and their toxic chemicals components: the need for urgent preventive measures. Environ. Sci. Eur. 30, 13.

Galloway, T.S., 2015. Micro- and nano-plastics and human health, in: M. Bergmann, L. Gutow, M. Klages (Eds.), Marine Anthropogenic Litter. Springer International Publishing, Cham, pp. 343-366.



- Galloway, T.S., Lewis, C.N., 2016. Marine microplastics spell big problems for future generations. PNAS 113, 2331-2333.
- Geueke, B., 2018. Dossier- Non-intentionally added substances (NIAS), 2 ed. Food Packaging Forum Foundation, Zurich, Switzerland.
 - Geueke, B., Groh, K., Muncke, J., 2018. Food packaging in the circular economy: Overview of chemical safety aspects for commonly used materials. J. Clean. Prod.
- 919 https://doi.org/10.1016/j.jclepro.2018.05.005.
 - Geyer, R., Jambeck, J.R., Law, K.L., 2017. Production, use, and fate of all plastics ever made. Sci. Adv. 3, 31700782.
 - Glüge, J., Schinkel, L., Hungerbühler, K., Cariou, R., Bogdal, C., 2018. Environmental Risks of Medium-Chain Chlorinated Paraffins (MCCPs): A Review. Environmental Science & Technology 52, 6743-6760.
 - Goldringer, D.m., Demierre, A.L., Zoller, O., Rupp, H., Reinhard, H., Magnin, R., Becker, T.A., Bourqui-Pittet, M., 2015. Endocrine activity of alternatives to BPA found in thermal paper in Switzerland. Regul. Toxicol. Pharmacol. 71, 453-462.
 - Grob, K., Biedermann, M., Scherbaum, E., Roth, M., Rieger, K., 2006. Food contamination with organic materials in perspective: packaging materials as the largest and least controlled source? A view focusing on the European situation. Crit. Rev. Food Sci. Nutr. 46, 529-535.
 - Groh, K.J., Geueke, B., Muncke, J., 2017. Food contact materials and gut health: Implications for toxicity assessment and relevance of high molecular weight migrants. Food Chem. Toxicol. 109, 1-18.
 - Groh, K.J., Muncke, J., 2017. In Vitro Toxicity Testing of Food Contact Materials: State-of-the-Art and Future Challenges. Compr. Rev. Food Sci. Food Saf. 16, 1123-1150.
 - Guzzonato, A., Puype, F., Harrad, S.J., 2017. Evidence of bad recycling practices: BFRs in children's toys and food-contact articles. ES:PI 19, 956-963.
 - Hahladakis, J.N., Velis, C.A., Weber, R., Iacovidou, E., Purnell, P., 2018. An overview of chemical additives present in plastics: Migration, release, fate and environmental impact during their use, disposal and recycling. J. Hazard. Mater. 344, 179-199.
 - Haines, D.A., Saravanabhavan, G., Werry, K., Khoury, C., 2017. An overview of human biomonitoring of environmental chemicals in the Canadian Health Measures Survey: 2007–2019. Int. J. Hyg. Environ. Health 220, 13-28.
 - Halden, R.U., 2010. Plastics and Health Risks. Annu. Rev. Public Health 31, 179-194.
 - Hammer, J., Kraak, M.H., Parsons, J.R., 2012. Plastics in the marine environment: the dark side of a modern gift. Rev. Environ. Contam. Toxicol. 220, 1-44.
 - Hansen, E., Nilsson, N.H., Lithner, D., Lassen, C., 2013. Hazardous substances in plastic materials. COWI and Danish Technological Institute.
 - Harper, C.A.e., 2006. Handbook of plastic technologies: The complete guide to properties and performance, 2nd Revised edition ed. McGraw-Hill Education, New York, USA.
 - Harrison, J.P., Boardman, C., O'Callaghan, K., Delort, A.M., Song, J., 2018. Biodegradability standards for carrier bags and plastic films in aquatic environments: a critical review. R. Soc. Open Sci. 5, 171792.
 - Heindel, J.J., Blumberg, B., Cave, M., Machtinger, R., Mantovani, A., Mendez, M.A., Nadal, A., Palanza, P., Panzica, G., Sargis, R., Vandenberg, L.N., vom Saal, F., 2017. Metabolism disrupting chemicals and metabolic disorders. Reprod. Toxicol. 68, 3-33.
 - Hermabessiere, L., Dehaut, A., Paul-Pont, I., Lacroix, C., Jezequel, R., Soudant, P., Duflos, G., 2017. Occurrence and effects of plastic additives on marine environments and organisms: A review. Chemosphere 182, 781-793.
 - Hoffmann, S., de Vries, R.B.M., Stephens, M.L., Beck, N.B., Dirven, H.A.A.M., Fowle, J.R., Goodman, J.E., Hartung, T., Kimber, I., Lalu, M.M., Thayer, K., Whaley, P., Wikoff, D., Tsaioun, K.,
- 2017. A primer on systematic reviews in toxicology. Archives of Toxicology 91, 2551-2575.
 Jambeck, J.R., Geyer, R., Wilcox, C., Siegler, T.R., Perryman, M., Andrady, A., Narayar
 - Jambeck, J.R., Geyer, R., Wilcox, C., Siegler, T.R., Perryman, M., Andrady, A., Narayan, R., Law, K.L., 2015. Plastic waste inputs from land into the ocean. Science 347, 768-771.



Jenke, D., 2009. Compatibility of pharmaceutical products and contact materials: Safety considerations associated with extractables and leachables. John Wiley & Sons, Inc., Hoboken, New Jersey, U.S.

Kaltenhäuser, J., Kneuer, C., Marx-Stoelting, P., Niemann, L., Schubert, J., Stein, B., Solecki, R., 2017. Relevance and reliability of experimental data in human health risk assessment of pesticides. Regul. Toxicol. Pharmacol. 88, 227-237.

Kawakami, T., Isama, K., Matsuoka, A., 2011. Analysis of phthalic acid diesters, monoester, and other plasticizers in polyvinyl chloride household products in Japan. J. Environ. Sci. Health A Tox. Hazard. Subst. Environ. Eng. 46, 855-864.

Keith, L., Telliard, W., 1979. ES&T Special Report: Priority pollutants: I-a perspective view. Environ. Sci. Technol. 13, 416-423.

Keith, L.H., 2015. The Source of U.S. EPA's Sixteen PAH Priority Pollutants. Polycycl. Arom. Comp. 35, 147-160.

Kienzler, A., Bopp, S.K., van der Linden, S., Berggren, E., Worth, A., 2016. Regulatory assessment of chemical mixtures: Requirements, current approaches and future perspectives. Regulatory Toxicology and Pharmacology 80, 321-334.

Kiyataka, P.H.M., Dantas, S.T., Pallone, J.A.L., 2014. Method for assessing lead, cadmium, mercury and arsenic in high-density polyethylene packaging and study of the migration into yoghurt and simulant. Food Addit. & Contam.: Part A 31, 156-163.

Klar, M., Gunnarsson, D., Prevodnik, A., Hedfors, C., Dahl, U., 2014. Everything you (don't) want to know about plastics. Naturskyddsfoereningen.

Kolossa-Gehring, M., Fiddicke, U., Leng, G., Angerer, J., Wolz, B., 2017. New human biomonitoring methods for chemicals of concern—the German approach to enhance relevance. Int. J. Hyg. Environ. Health 220, 103-112.

Lahl, U., Zeschmar-Lahl, B., 2013. Risk based management of chemicals and products in a circular economy at a global scale (risk cycle), extended producer responsibility and EU legislation. Environ. Sci. Eur. 25, 3.

Lambert, S., Wagner, M., 2017. Environmental performance of bio-based and biodegradable plastics: the road ahead. Chem. Soc. Rev. 46, 6855-6871.

Larson, A.M., Klibanov, A.M., 2013. Biocidal Packaging for Pharmaceuticals, Foods, and Other Perishables. Annu. Rev. Chem. Biomolec. Eng. 4, 171-186.

Latorre, I., Hwang, S., Sevillano, M., Montalvo-Rodriguez, R., 2012. PVC biodeterioration and DEHP leaching by DEHP-degrading bacteria. Int. Biodeter. Biodegr. 69, 73-81.

Leslie, H.A., Leonards, P.E.G., Brandsma, S.H., de Boer, J., Jonkers, N., 2016. Propelling plastics into the circular economy - weeding out the toxics first. Environ. Int. 94, 230-234.

Levy, J.H., Anderson, E., Sasano, T., 2001. Plastics additives, Chemical Economics Handbook (CEH) Marketing Research Reports.

Li, J., Su, G., Letcher, R.J., Xu, W., Yang, M., Zhang, Y., 2018. Liquid Crystal Monomers (LCMs): A New Generation of Persistent Bioaccumulative and Toxic (PBT) Compounds? Environ. Sci. Technol. 52, 5005-5006.

Li, S.-Q., Ni, H.-G., Zeng, H., 2017. PAHs in polystyrene food contact materials: An unintended consequence. Science of The Total Environment 609, 1126-1131.

Lithner, D., Larsson, Å., Dave, G., 2011. Environmental and health hazard ranking and assessment of plastic polymers based on chemical composition. Science of The Total Environment 409, 3309-3324.

Lu, Z., De Silva, A.O., McGoldrick, D.J., Zhou, W., Peart, T.E., Cook, C., Tetreault, G.R., Martin, P.A., de Solla, S.R., 2018. Substituted Diphenylamine Antioxidants and Benzotriazole UV Stabilizers in Aquatic Organisms in the Great Lakes of North America: Terrestrial Exposure and Biodilution. Environmental Science & Technology 52, 1280-1289.

Maffini, M.V., Neltner, T.G., 2014. Brain drain: the cost of neglected responsibilities in evaluating cumulative effects of environmental chemicals. J Epidemiol Community Health 69, 496-499.



Malhotra, B., Keshwani, A., Kharkwal, H., 2015. Antimicrobial food packaging: potential and pitfalls. Front. Microbiol. 6, 611.

Mavakala, B.K., Le Faucheur, S., Mulaji, C.K., Laffite, A., Devarajan, N., Biey, E.M., Giuliani, G., Otamonga, J.-P., Kabatusuila, P., Mpiana, P.T., Poté, J., 2016. Leachates draining from controlled municipal solid waste landfill: Detailed geochemical characterization and toxicity tests. Waste Management 55, 238-248.

Miller, D., Wheals, B.B., Beresford, N., Sumpter, J.P., 2001. Estrogenic activity of phenolic additives determined by an *in vitro* yeast bioassay. Environ. Health Perspect. 109, 133-138.

Milne, G.W.A.e., 2005. Gardner's commercially important chemicals. Synonyms, trade names, and properties. John Wiley & Sons, Hoboken, New Jersey, USA.

Montano, D., 2014. Chemical and biological work-related risks across occupations in Europe: a review. J. Occup. Med. Toxicol. 9, 28-28.

Morgan, R.L., Thayer, K.A., Bero, L., Bruce, N., Falck-Ytter, Y., Ghersi, D., Guyatt, G., Hooijmans, C., Langendam, M., Mandrioli, D., Mustafa, R.A., Rehfuess, E.A., Rooney, A.A., Shea, B., Silbergeld, E.K., Sutton, P., Wolfe, M.S., Woodruff, T.J., Verbeek, J.H., Holloway, A.C., Santesso, N., Schünemann, H.J., 2016. GRADE: Assessing the quality of evidence in environmental and occupational health. Environ. Int. 92-93, 611-616.

Muncke, J., 2014. Hazards of food contact material: Food packaging contaminants. Ref. Module Food Sci. 2, 430-437.

Muncke, J., Backhaus, T., Geueke, B., Maffini, M., Martin, O., Myers, J.P., Soto, A., Trasande, L., Trier, X., Scheringer, M., 2017. Scientific challenges in the risk assessment of food contact materials. Env. Health Persp. 125, 095001.

Muncke, J., Myers, J.P., Scheringer, M., Porta, M., 2014. Food packaging and migration of food contact materials: will epidemiologists rise to the neotoxic challenge? J. Epidemiol. Commun. Health 68, 592-594.

Munier, B., Bendell, L.I., 2018. Macro and micro plastics sorb and desorb metals and act as a point source of trace metals to coastal ecosystems. PLoS ONE 13, e0191759.

Musoke, L., Banadda, N., Sempala, C., Kigozi, J., 2015. The migration of chemical contaminants from polyethylene bags into food during cooking. Open Food Sci. J. 9, 14-18.

Myers, J.P., vom Saal, F.S., Akingbemi, B.T., Arizono, K., Belcher, S., Colborn, T., Chahoud, I., Crain, D.A., Farabollini, F., Guillette, L.J.J., Hassold, T., Ho, S.-m., Hunt, P.A., Iguchi, T., Jobling, S., Kanno, J., Laufer, H., Marcus, M., McLachlan, J.A., Nadal, A., Oehlmann, J., Olea, N., Palanza, P., Parmigiani, S., Rubin, B.S., Schoenfelder, G., Sonnenschein, C., Soto, A.M., Talsness, C.E., Taylor, J.A., Vandenberg, L.N., Vandenbergh, J.G., Vogel, S., Watson, C.S., Welshons, W.V., Zoeller, R.T., 2010. Why Public Health Agencies Cannot Depend on Good Laboratory Practices as a Criterion for Selecting Data: The Case of Bisphenol A. 117, 309-315.

Nerin, C., Alfaro, P., Aznar, M., Domeno, C., 2013. The challenge of identifying non-intentionally added substances from food packaging materials: A review. Anal. Chim. Acta 775, 14-24.

Nerin, C., Canellas, E., Vera, P., Garcia-Calvo, E., Luque-Garcia, J.L., Cámara, C., Ausejo, R., Miguel, J., Mendoza, N., 2018. A common surfactant used in food packaging found to be toxic for reproduction in mammals. Food Chem. Toxicol. 113, 115-124.

Nguyen Van Long, N., Joly, C., Dantigny, P., 2016. Active packaging with antifungal activities. Int. J. Food Microbiol. 220, 73-90.

North, E.J., Halden, R.U., 2013. Plastics and Environmental Health: The Road Ahead. Rev. Int. Health 28, 1-8.

OECD, 2014. Guidance on grouping of chemicals, second edition, in: OECD (Ed.), Series on Testing & Assessment. OECD, Paris.

Oldring, P.K.T., O'Mahony, C., Dixon, J., Vints, M., Mehegan, J., Dequatre, C., Castle, L., 2014. Development of a new modelling tool (FACET) to assess exposure to chemical migrants from food packaging. Food Additives & Contaminants: Part A 31, 444-465.



Paseiro-Cerrato, R., MacMahon, S., Ridge, C.D., Noonan, G.O., Begley, T.H., 2016. Identification of unknown compounds from polyester cans coatings that may potentially migrate into food or food simulants. Journal of Chromatography A 1444, 106-113.

Piersma, A.H., Burgdorf, T., Louekari, K., Desprez, B., Taalman, R., Landsiedel, R., Barroso, J., Rogiers, V., Eskes, C., Oelgeschläger, M., Whelan, M., Braeuning, A., Vinggaard, A.M., Kienhuis, A., van Benthem, J., Ezendam, J., 2018. Workshop on acceleration of the validation and regulatory acceptance of alternative methods and implementation of testing strategies. Toxicol. In Vitro 50, 62-74.

Pivnenko, K., Granby, K., Eriksson, E., Astrup, T.F., 2017. Recycling of plastic waste: Screening for brominated flame retardants (BFRs). Waste Management 69, 101-109.

PlasticsEurope, 2016. Plastics -- the Facts 2016: An analysis of European plastics production, demand and waste data, Brussels, Belgium.

Pop, A., Drugan, T., Gutleb, A.C., Lupu, D., Cherfan, J., Loghin, F., Kiss, B., 2018. Estrogenic and anti-estrogenic activity of butylparaben, butylated hydroxyanisole, butylated hydroxytoluene and propyl gallate and their binary mixtures on two estrogen responsive cell lines (T47D-Kbluc, MCF-7). J. App. Toxicol. 38, 944-957.

Pop, A., Kiss, B., Loghin, F., 2013. Endocrine disrupting effects of butylated hydroxyanisole (BHA - E320). Clujul Medical 86, 16-20.

Ragossnig, A.M., Schneider, D.R., 2017. What is the right level of recycling of plastic waste? Waste Manag. & Res. 35, 129-131.

Rajamani, U., Gross, A.R., Ocampo, C., Andres, A.M., Gottlieb, R.A., Sareen, D., 2017. Endocrine disruptors induce perturbations in endoplasmic reticulum and mitochondria of human pluripotent stem cell derivatives. Nat. Comm. 8, 219.

Rani, M., Shim, W.J., Han, G.M., Jang, M., Song, Y.K., Hong, S.H., 2014. Hexabromocyclododecane in polystyrene based consumer products: An evidence of unregulated use. Chemosphere 110, 111-119.

Revel, M., Châtel, A., Mouneyrac, C., 2018. Micro(nano)plastics: A threat to human health? Curr. Opin. Environ. Sci. Health 1, 17-23.

Rochester, J.R., Bolden, A.L., 2015. Bisphenol S and F: A systematic review and comparison of the hormonal activity of bisphenol A substitutes. Environ. Health Perspect. 123, 643-650.

Rochman, C.M., Manzano, C., Hentschel, B.T., Simonich, S.L.M., Hoh, E., 2013. Polystyrene Plastic: A Source and Sink for Polycyclic Aromatic Hydrocarbons in the Marine Environment. Environ. Sci. Technol. 47, 13976-13984.

Rosenmai, A.K., Dybdahl, M., Pedersen, M., van Vugt-Lussenburg, A.B.M., Wedebye, E.B., Taxvig, C., Vinggaard, A.M., 2014. Are structural analogues to bisphenol a safe alternatives? Toxicol. Sci. 139, 35-47.

Rossi, M.S., Blake, A., 2014. The plastics scorecard: Evaluating the chemical footprint of plastics, Version 1.0 ed. Clean Production Action.

Samsonek, J., Puype, F., 2013. Occurrence of brominated flame retardants in black thermo cups and selected kitchen utensils purchased on the European market. Food Addit. Contam.: Part A 30, 1976-1986.

Sarigiannis, D.A., 2017. Assessing the impact of hazardous waste on children's health: The exposome paradigm. Environmental Research 158, 531-541.

Scheringer, M., Trier, X., Cousins, I.T., de Voogt, P., Fletcher, T., Wang, Z., Webster, T.F., 2014. Helsingør Statement on poly- and perfluorinated alkyl substances (PFASs). Chemosphere 114, 337-339.

Schrankel, K.R., 2004. Safety evaluation of food flavorings. Toxicology 198, 203-211.

Schultz, T.W., Amcoff, P., Berggren, E., Gautier, F., Klaric, M., Knight, D.J., Mahony, C.,

Schwarz, M., White, A., Cronin, M.T.D., 2015. A strategy for structuring and reporting a read-across prediction of toxicity. Regulatory Toxicology and Pharmacology 72, 586-601.

Selke, S.E.M., Culter, J.D., 2016. Plastics packaging: Properties, processing, applications, and regulations. Carl Hanser Verlag, Munich, Germany.



- Severin, I., Souton, E., Dahbi, L., Chagnon, M.C., 2017. Use of bioassays to assess hazard of food contact material extracts: State of the art. Food Chem. Toxicol. 105, 429-447.
- Sheftel, V.O., 2000. Indirect food additives and polymers. Migration and toxicology. Lewis Publishers, CRC Press, Boca Raton, Florida, USA.
 - Simon, C., Onghena, M., Covaci, A., Van Hoeck, E., Van Loco, J., Vandermarken, T., Van Langenhove, K., Demaegdt, H., Mertens, B., Vandermeiren, K., Scippo, M.-L., Elskens, M., 2016. Screening of endocrine activity of compounds migrating from plastic baby bottles using a multireceptor panel of *in vitro* bioassays. Toxicol. In Vitro 37, 121-133.
 - Smithers Pira, 2018. The future of rigid plastic packaging to 2022, USA.
 - Sobus, J.R., DeWoskin, R.S., Tan, Y.-M., Pleil, J.D., Phillips, M.B., George, B.J., Christensen, K., Schreinemachers, D.M., Williams, M.A., Hubal, E.A.C., Edwards, S.W., 2015. Uses of NHANES biomarker data for chemical risk assessment: Trends, challenges, and opportunities. Env. Health Persp. 123, 919-927.
 - Sohail, M., Sun, D.-W., Zhu, Z., 2018. Recent developments in intelligent packaging for enhancing food quality and safety. Critical Reviews in Food Science and Nutrition 10.1080/10408398.2018.1449731, 1-13.
- Stenmarck, A., Belleza, E.L., Frane, A., Busch, N., Larsen, A., Wahlstroem, M., 2017. Hazardous substances in plastics -- ways to increase recycling, in: N.C.o. Ministers (Ed.), Denmark.
 - Stringer, L., 2018. Cross-sector initiative sets full materials disclosure goal, Chemical Watch. Swedish Chemicals Agency (KEMI), 2015. Grouping of chemical substances in the REACH and CLP regulations.
 - Thompson, R.C., Moore, C.J., vom Saal, F.S., Swan, S.H., 2009a. Plastics, the environment and human health: current consensus and future trends. Phil. Transact. R. Soc. B: Biol. Sci. 364, 2153-2166.
 - Thompson, R.C., Swan, S.H., Moore, C.J., vom Saal, F.S., 2009b. Our plastic age. Philosophical Transactions of the Royal Society B: Biological Sciences 364, 1973-1976.
 - Tolinski, M., 2009. Additives for polyolefins. Elsevier Inc, USA.
 - Toxics in Packaging Clearinghouse, 2017. XRF screening of packaging components: Cadmium continues to be present in flexible polyvinyl chloride (PVC), USA, p. 9.
 - UNEP, 2018. Overview Report I: Worldwide initiatives to identify endocrine disrupting chemicals (EDCs) and potential EDCs. International Panel on Chemical Pollution (IPCP).
 - Van Eygen, E., Feketitsch, J., Laner, D., Rechberger, H., Fellner, J., 2017. Comprehensive analysis and quantification of national plastic flows: The case of Austria. Resourc. Conserv. Recycl. 117, 183-194.
 - van Oers, L., van der Voet, E., Grundmann, V., 2011. Additives in the plastics industry, in: B. Bilitewski, R.M. Darbra, D. Barcelo (Eds.), Global risk-based management of chemical additives I: Production, usage and environmental occurrence. Springer, Berlin, pp. 133-149.
 - van Putten, E.M., 2011. Heavy metals in packaging: A literature survey, This report is a translation of the RIVM Report 609021111/2011 "Zware metalen in verpakkingen" ed. National Institute for Public Health and the Environment, Ministry of Health, Welfare and Sport.
 - Vandenberg, L.N., Ågerstrand, M., Beronius, A., Beausoleil, C., Bergman, Å., Bero, L.A., Bornehag, C.-G., Boyer, C.S., Cooper, G.S., Cotgreave, I., Gee, D., Grandjean, P., Guyton, K.Z., Hass, U., Heindel, J.J., Jobling, S., Kidd, K.A., Kortenkamp, A., Macleod, M.R., Martin, O.V., Norinder, U., Scheringer, M., Thayer, K.A., Toppari, J., Whaley, P., Woodruff, T.J., Rudén, C., 2016. A proposed framework for the systematic review and integrated assessment (SYRINA) of endocrine disrupting chemicals. Environ. Health 15, 74.
- 1165 Vera, P., Canellas, E., Nerín, C., 2018. Identification of non volatile migrant compounds and 1166 NIAS in polypropylene films used as food packaging characterized by UPLC-MS/QTOF. Talanta 188, 1167 750-762.
- Wagner, M., 2016. Know thy unknowns: why we need to widen our view on endocrine disruptors. J. Epidemiol. Commun. Health 71, 209-212.



1170	Wang, Z., Cousins, I.T., Scheringer, M., Hungerbuehler, K., 2015. Hazard assessment of
1171	fluorinated alternatives to long-chain perfluoroalkyl acids (PFAAs) and their precursors: Status quo,
1172	ongoing challenges and possible solutions. Environment International 75, 172-179.
1173	WHO/IPCS, 2002. Global assessment of the state-of-the-science of endocrine disruptors, in:
1174	T. Damstra, S. Barlow, A. Bergman, R. Kavlock, G. Van Der Kraak (Eds.).
1175	Wright, S.L., Kelly, F.J., 2017. Plastic and Human Health: A Micro Issue? Environ. Sci. Technol.
1176	51, 6634-6647.
1177	Zweifel, H., Meier, R.D., Schiller, M., (eds.), 2009. Plastics additives handbook. Carl Hanser
1178	Verlag, Munich, Germany.