

A peer-reviewed version of this preprint was published in PeerJ on 5 October 2018.

[View the peer-reviewed version](https://peerj.com/articles/5735) (peerj.com/articles/5735), which is the preferred citable publication unless you specifically need to cite this preprint.

Winker K, Glenn TC, Faircloth BC. 2018. Ultraconserved elements (UCEs) illuminate the population genomics of a recent, high-latitude avian speciation event. PeerJ 6:e5735 <https://doi.org/10.7717/peerj.5735>

Ultraconserved elements (UCEs) illuminate the population genomics of a recent, high-latitude avian speciation event

Kevin Winker ^{Corresp., 1}, Travis C Glenn ², Brant C Faircloth ³

¹ University of Alaska Museum & Department of Biology and Wildlife, University of Alaska Fairbanks, Fairbanks, Alaska, USA

² Department of Environmental Health Science and Institute of Bioinformatics, University of Georgia, Athens, United States

³ Department of Biological Sciences and Museum of Natural Science, Louisiana State University, Louisiana State University, Baton Rouge, USA

Corresponding Author: Kevin Winker

Email address: kevin.winker@alaska.edu

Using a large, consistent set of loci shared by descent (orthologous) to study relationships among taxa would revolutionize among-lineage comparisons of divergence and speciation processes. Ultraconserved elements (UCEs), highly conserved regions of the genome, offer such genomic markers. The utility of UCEs for deep phylogenetics is clearly established and there are mature analytical frameworks available, but fewer studies apply UCEs to recent evolutionary events, creating a need for additional example datasets and analytical approaches. We used UCEs to study population genomics in snow and McKay's buntings (*Plectrophenax nivalis* and *P. hyperboreus*). Prior work suggested divergence of these sister species during the last glacial maximum (~18-74 Kya). With a sequencing depth of ~30x from four individuals of each species, we used a series of analysis tools to genotype both alleles, obtaining a complete dataset of 2,635 variable loci (~3.6 single nucleotide polymorphisms [SNPs]/locus) and 796 invariable loci. We found no fixed allelic differences between the lineages, and few loci had large allele frequency differences. Nevertheless, individuals were 100% diagnosable to species, and the two taxa were different genetically ($F_{ST} = 0.034$; $P = 0.03$). The demographic model best fitting the data was one of divergence with gene flow. Estimates of demographic parameters differed from published mtDNA research, with UCE data suggesting lower effective population sizes (~92,500 - 240,500 individuals), a deeper divergence time (~241,000 yrs), and lower gene flow (2.8-5.2 individuals per generation). Our methods provide a framework for future population studies using UCEs, and our results provide additional evidence that UCEs are useful for answering questions at shallow evolutionary depths.

1 **Ultraconserved elements (UCEs) illuminate the population genomics of a recent, high-**
2 **latitude avian speciation event**

3
4 **Kevin Winker¹, Travis C. Glenn², and Brant C. Faircloth³**

5 ¹ University of Alaska Museum, 907 Yukon Drive, Fairbanks, AK 99775, USA

6 ² Department of Environmental Health Science and Institute of Bioinformatics, University of
7 Georgia, Athens, GA 30602, USA

8 ³ Department of Biological Sciences and Museum of Natural Science, Louisiana State
9 University, Baton Rouge, LA 70803, USA

10

11 Corresponding author: Kevin Winker, University of Alaska Museum, 907 Yukon Drive,
12 Fairbanks, AK 99775, USA; kevin.winker@alaska.edu; 907-474-7027; fax 907-474-5469

13

14 Running title: Bunting speciation using UCEs

15

16 Using a large, consistent set of loci shared by descent (orthologous) to study relationships among
17 taxa would revolutionize among-lineage comparisons of divergence and speciation processes.

18 Ultraconserved elements (UCEs), highly conserved regions of the genome, offer such genomic
19 markers. The utility of UCEs for deep phylogenetics is clearly established and there are mature

20 analytical frameworks available, but fewer studies apply UCEs to recent evolutionary events,
21 creating a need for additional example datasets and analytical approaches. We used UCEs to

22 study population genomics in snow and McKay's buntings (*Plectrophenax nivalis* and *P.*

23 *hyperboreus*). Prior work suggested divergence of these sister species during the last glacial

24 maximum (~18-74 Kya). With a sequencing depth of ~30× from four individuals of each species,

25 we used a series of analysis tools to genotype both alleles, obtaining a complete dataset of 2,635

26 variable loci (~3.6 single nucleotide polymorphisms [SNPs]/locus) and 796 invariable loci. We

27 found no fixed allelic differences between the lineages, and few loci had large allele frequency

28 differences. Nevertheless, individuals were 100% diagnosable to species, and the two taxa were

29 different genetically ($F_{ST} = 0.034$; $P = 0.03$). The demographic model best fitting the data was

30 one of divergence with gene flow. Estimates of demographic parameters differed from published

31 mtDNA research, with UCE data suggesting lower effective population sizes (~92,500 - 240,500

32 individuals), a deeper divergence time (~241,000 yrs), and lower gene flow (2.8-5.2 individuals

33 per generation). Our methods provide a framework for future population studies using UCEs,

34 and our results provide additional evidence that UCEs are useful for answering questions at
35 shallow evolutionary depths.

36

37 **Introduction**

38 Among non-model organisms, population genetic studies have used a diverse set of
39 markers, tending to concentrate on those with sufficiently high substitution rates to provide
40 useful data at shallow levels of evolutionary divergence, e.g., from the populations-to-species
41 levels (Awise 1994, Hillis et al. 1996, Pearse & Crandall 2003). This approach usually provides
42 answers to the specific questions asked by researchers, but the historic focus on markers with
43 high substitution rates has produced studies that include relatively few loci and often have little
44 to no overlap with loci used for other taxa. This lack of consistency in the loci used across
45 different studies compromises our ability to make direct comparisons of population genetic
46 parameters among taxa (e.g., in divergence statistics and in estimates of gene flow and effective
47 population sizes). Improvements in sequencing platforms and genomic data collection
48 approaches are changing this general pattern by enabling us to efficiently collect much larger
49 samples of the genome, up to and including whole-genome sequences (Ellegren 2014). However,
50 the sheer quantity of data obtained from whole-genome sequencing can require excessively long
51 computation times, and may be overkill for many questions. The parallel difficulties of collecting
52 a moderate sample of the genome from identical loci across diverse species argue for a sequence
53 data collection approach that a) subsamples the genome to b) obtain orthologous markers across
54 a broad taxonomic scope. This type of approach would provide a tractable number of loci for
55 analyses while improving among-study comparisons and larger-scale comparative meta-
56 analyses. Ultraconserved elements (UCEs) are one class of genome-wide marker that might
57 provide a solution to these problems.

58 UCEs are conserved sequences shared among divergent animal genomes (Bejerano et al.
59 2004, Siepel et al. 2005, Stephen et al. 2008, Janes et al. 2011), and many UCE loci are likely to
60 be involved in controlling gene expression (Marcovitz et al. 2016). UCEs in vertebrates show
61 little overlap with most types of paralogous genes, and, as a marker class, UCE loci are broadly
62 distributed across the genome and are typically transposon-free (Derti et al. 2006, Simons et al.
63 2006, McCormack et al. 2011, Harvey et al. 2016). We focus on the set of UCEs previously
64 defined for tetrapods and now in widespread use (McCormack et al. 2011, Faircloth et al. 2012).

65 Outside of their functional relevance, UCE loci have demonstrated utility for recovering deeper-
66 level phylogenetic relationships (McCormack et al. 2013, Faircloth et al. 2015, Gilbert et al.
67 2015) and shallower-level genus and population relationships (Smith et al. 2014, Harvey &
68 Brumfield 2015, Leaché et al. 2015, Harvey et al. 2016, Manthey et al. 2016, Oswald et al. 2016,
69 Mason et al. 2018). Although UCEs are highly conserved at their core, which enables universal
70 capture of loci across diverse groups of organisms (Faircloth et al. 2012, 2013, 2015; Starret et
71 al. 2016), lower levels of purifying selection away from the core allow substitutions to
72 accumulate in the flanking regions. Using human genome data, Faircloth et al. (2012)
73 demonstrated that the increased variation in UCE flanking sequence might be adequate to make
74 these loci useful for questions at shallow levels of divergence, and this hypothesis has been
75 supported by subsequent empirical studies (Smith et al. 2014, Harvey & Brumfield 2015, Harvey
76 et al. 2016, Oswald et al. 2016, Mason et al. 2018). However, the utility of UCE loci for studying
77 population genetics, population divergence, and/or incipient speciation is only beginning to be
78 tested, and both the value and challenges of using UCEs at these shallow levels remain
79 underexplored.

80 Here, we examine the utility of UCEs for studying the population genomics of divergence
81 between two bird species, McKay's bunting (*Plectrophenax hyperboreus*) and snow bunting (*P.*
82 *nivalis*). McKay's buntings breed on remote islands in the Bering Sea (where our samples are
83 from) and are the highest-latitude endemic songbirds; their range is restricted to the North Pacific
84 region. Snow buntings breed throughout the rest of the high-latitude Holarctic (our samples are
85 from the southern edge of the Bering sea on the Alaska Peninsula and an Aleutian island; Table
86 S1). McKay's bunting is thought to have arisen ~18-74 Kya during the last glacial maximum
87 (LGM) through divergence from snow buntings, and previous work suggests gene flow between
88 the two may be ongoing (Maley & Winker 2010). These species are interesting to study using
89 UCEs because prior work (Maley & Winker 2010) enables us to compare population genetic
90 statistics derived from UCEs versus traditional population genetic markers (mtDNA sequence
91 and amplified fragment length polymorphisms, AFLPs) and because these species allow us to
92 test the utility of UCEs for studying very shallow divergences between sister lineages where
93 gene flow may be ongoing.

94

95 **Methods**

96 *Laboratory.*—We extracted DNA from muscle tissue of eight specimens (four of each
97 species) studied by Maley & Winker (2010) using proteinase K digestion (100 mM Tris pH 8, 50
98 mM EDTA, 0.5% SDS, 1 mg/mL proteinase K) followed by SPRI bead purification (Rohland &
99 Reich 2012); Supplemental Information, Table S1). We chose this sample size (and our
100 sequencing depth) to ensure that we could confidently call both alleles for each individual in
101 each population to achieve eight sequences per population at each locus, which Felsenstein
102 (2005) considered to be the optimum sample size for coalescent-based analyses. Following DNA
103 extraction, we prepared dual-indexed DNA libraries for each sample using methods described in
104 Glenn et al. (2017). After library preparation, we quantified each library using a Qubit
105 fluorimeter (Invitrogen, Inc.), and we combined eight libraries into equimolar pools of 500 ng
106 each (62.5 ng/library). We enriched the pool of 8 samples for 5,060 UCE loci using the
107 Tetrapods-UCE-5Kv1 kit from MYcroarray following version 1.5 of the UCE enrichment
108 protocol and version 2.4 of the post-enrichment amplification protocol (ultraconserved.org) with
109 HiFi HotStart polymerase (Kapa Biosystems) and 14 cycles of post-enrichment PCR. We then
110 quantified the fragment size distribution of the enriched pool on a Bioanalyzer (Agilent, Inc.) and
111 qPCR quantified the enriched pool using a commercial kit (Kapa Biosystems). We combined the
112 enriched pool of eight bunting samples with enriched pools from other birds at equimolar ratios,
113 and we sequenced the resulting pool using one lane of PE150 sequencing on an Illumina HiSeq
114 2500 (UCLA Neuroscience Genomics Core).

115 *Bioinformatics.*—After sequencing, we demultiplexed the sequencing reads using
116 bcl2fastq version 1.8.4 (Illumina, Inc.), and we trimmed the demultiplexed reads for adapter
117 contamination and low-quality bases using a parallel wrapper (Faircloth 2013) around
118 Trimmomatic (ver. 0.32 Bolger et al. 2014). We then combined singleton reads that lost their
119 mate with read 1 files, combined all individual read 1 files (plus singletons) together and all
120 individual read 2 files together, and assembled these two read 1 and read 2 files *de novo* using
121 Trinity (ver. 2.0.6; Grabherr et al. 2011) on Galaxy (Afgan et al. 2016). After assembling this
122 composite of data from all individuals, we used Phyluce (ver. 1.4.0; Faircloth 2016) to identify
123 FASTA sequences from orthologous UCEs and remove FASTA sequences from non-UCE loci
124 or potential paralogs. We called the resulting file our reference set of UCE loci, which we used
125 as the reference sequence for calling individual variants.

126 Next, we used Phyluce and its program dependencies (BWA 0.7.7, Li & Durbin 2009;
127 SAMtools 0.1.19, Li et al. 2009; Picard 1.106, <http://broadinstitute.github.io/picard>) to align
128 unassembled, raw reads from individual buntings to the reference set of UCE loci. Specifically,
129 this workflow aligned raw reads on a sample-by-sample basis against the composite reference
130 using the bwa-mem algorithm (preferred for reads > 70 bp; Li 2013); added header information
131 to identify alignments from individual samples; cleaned, validated, and marked duplicates in the
132 resulting BAM (Binary Alignment/Map) file using Picard; and merged all individuals into a
133 single BAM file using Picard. Following preparation of the merged BAM, we used GATK (ver.
134 3.4-0; McKenna et al. 2010) to identify and realign indels, call and annotate single nucleotide
135 polymorphisms (SNPs) and indels, and mask SNP calls around indels using a GATK workflow
136 described as part of a population genomics pipeline for UCEs developed by Faircloth and
137 Michael Harvey (https://github.com/mgharvey/seqcap_pop). This included restricting data to
138 high-quality SNPs (Q30) and read-back phasing in GATK. After calling and annotating SNPs,
139 we deviated from this workflow by using VCFtools (ver. 0.1.12b; Danacek et al. 2011) to filter
140 the resulting variant call format (VCF) file with the --max-missing (1.0) and --minGQ (10.0)
141 parameters, which created a complete data matrix with a minimum genotype quality (GQ) of 10.
142 We validated that GQ10 data were present for all individuals at all loci by visually assessing
143 alignment data at 17 SNPs among 10 loci using Tablet (ver. 1.15.09.01; Milne et al. 2013). We
144 used GATK's EMIT_ALL_CONFIDENT_SITES function to ensure that we only retained
145 invariant loci with high quality (rather than missing) data. We then removed variable and
146 invariable loci with incomplete data from downstream analyses, retaining only loci with
147 complete data. This finalized our complete VCF file.

148 *Data analysis.*—We calculated coverage depths, SNP positions within loci, and SNP-
149 specific and locus-specific F_{ST} values on the complete VCF file using VCFtools (ver. 0.1.12b;
150 Danacek et al. 2011). After thinning the VCF file to 1 SNP/locus (which is required in
151 demographic analyses when unlinked variation is important) and converting the VCF file to
152 STRUCTURE format using PGDSpider (ver. 2.1.0.3; Lischer & Excoffier 2012), we performed
153 tests of Hardy-Weinberg equilibrium and computed observed and expected heterozygosities,
154 homogeneity of variance, population structure (population F_{ST} , including a 10,000-replicate G -
155 test; see Goudet et al. [1996]), and the probabilities of each individual's assignment to a
156 particular population using Discriminant Analysis of Principal Components (DAPC) in adegenet

157 (ver. 2.0.1; Jombart & Ahmed 2011). To calculate nucleotide diversity, we created a
158 concatenated FASTA file of all individual sequences using catfasta2phym1 by Johan Nylander
159 (<https://github.com/nylander/catfasta2phym1>), and we analyzed this file in MEGA (ver. 6;
160 Tamura et al. 2013) using the maximum composite likelihood method.

161 We used Diffusion Approximations for Demographic Inference ($\delta\text{a}\delta\text{i}$; ver. 1.7.0;
162 Gutenkunst et al. 2009) to infer demographic parameters from the data under a variety of
163 divergence scenarios (models) after excluding Z-linked loci (for $\delta\text{a}\delta\text{i}$ analyses only). Z-linked
164 loci in birds are on the sex chromosome, have a different inheritance scalar from autosomal loci,
165 and sample population sex ratios affect allele frequency estimates (e.g., Jorde et al. 2000,
166 Garrigan et al. 2007). We identified Z-linked loci in our data using BLASTn (ver. 2.3.1; Zhang
167 et al. 2000), by aligning the reference set of UCE loci against the zebra finch (*Taeniopygia*
168 *guttata*) genome (NCBI Annotation Release 103). We excluded UCE loci that strongly matched
169 (*E*-values ~ 0.0) the zebra finch Z chromosome. After removing Z-linked loci from our complete
170 VCF file, we converted this reduced dataset to biallelic format (which dropped one locus with >2
171 alleles at a SNP site) and thinned the data to one SNP per locus using VCFtools. Then we
172 converted the resulting VCF file to the joint site frequency spectrum (SFS) format required by
173 $\delta\text{a}\delta\text{i}$ using a PERL script by Kun Wang ([https://groups.google.com/forum/#!msg/dadi-](https://groups.google.com/forum/#!msg/dadi-user/p1WvTKRI9_0/1yQtcKqamPcJ)
174 [user/p1WvTKRI9_0/1yQtcKqamPcJ](https://groups.google.com/forum/#!msg/dadi-user/p1WvTKRI9_0/1yQtcKqamPcJ)). Because we lacked an outgroup, we used a folded SFS in
175 our analyses (Gutenkunst et al. 2009), which lacks polarization of SNPs (Fig. S1).

176 Because we had prior evidence that these species represent two genetic populations
177 (based on taxonomy and results from Maley & Winker 2010), we used $\delta\text{a}\delta\text{i}$ to infer what general
178 two-population divergence model best fit the data. We then used that model to estimate
179 demographic parameters (i.e., effective population sizes, split time, and migration). We ran six
180 different models spanning the standard possible demographic histories of two populations, five
181 basic and one derivative: 1) neutral (no divergence, or still strongly mixing), 2) split with
182 migration, 3) split with no migration, 4) isolation with bidirectional migration and population
183 growth, 5) isolation with population growth and no migration, and 6) a custom split-
184 bidirectional-migration model (a simple derivative of split-migration; Fig. 1). The neutral, split-
185 with-migration, and isolation-with-migration-and-population-growth models are provided in the
186 $\delta\text{a}\delta\text{i}$ file Demographics2D.py as snm, split_mig, and IM, respectively. The no-migration models
187 (3 and 5 above) use the split_mig and IM models, respectively, with migration parameters set to

188 zero. The split-bidirectional-migration model (Supplemental Information) adds bidirectional
189 migration to the split-migration model to examine potential asymmetry in gene flow.

190 We performed a series of optimization runs (8-47 each) of each basic model, adjusting
191 parameters (grid points, upper and lower bounds) to identify high log composite likelihoods. We
192 then ran each model repeatedly, varying parameters within bounds that yielded the highest
193 likelihood during optimization, until three runs yielded the highest observed likelihood value.
194 Our reasoning was that this level of repeatability indicated a best-fit neighborhood for each
195 model. We report this highest observed likelihood, except for poorer models, which yielded
196 variable likelihoods, in which case we averaged and report the highest five values. After
197 identifying the best-fit model based on likelihood values over successive runs, we ran the best-fit
198 model ten times each with jackknifed datasets to estimate the 95% confidence interval (CI) for
199 each parameter.

200 We estimated recombination using the four-gametes test as implemented in IMgc
201 (Woerner et al. 2007), which also produces sequence datasets from which the effects of
202 recombination have been removed. Resulting sequences were used for IMA2p (ver. 1.0,
203 Sethuraman and Hey 2015) analyses to attempt to estimate demographic parameters, but those
204 analyses did not converge under a variety of full- and sub-sampling schemes and are not
205 reported. We nevertheless include the results of IMgc because accounting for recombination is a
206 critical part of workflows using full sequences (i.e., not just SNPs), and these results provide
207 needed insight into the levels of recombination found in UCE loci for studies of this type.

208 We estimated the average per-site substitution rate by BLASTing the FASTA file
209 containing all confidently scored loci (those meeting our quality filters as described in the
210 *Bioinformatics* section, above) for all individuals (3,431 loci) against the budgerigar
211 (*Melopsittacus undulatus*) genome (NCBI release 102) and the rifleman (*Acanthisitta chloris*)
212 genome (NCBI release 100), using time to most recent common ancestor (TMRCA) date
213 estimates of 60.5 Ma (budgerigar) and 53 Ma (rifleman) (Claramunt & Cracraft 2015). These
214 taxa were chosen as the nearest relatives with complete genomes and fossil-dated nodes
215 available. We imported BLAST results (hit table, csv) into a spreadsheet, removed duplicate
216 lower-affinity hits, then summed total length of base pairs, total substitutions, and calculated
217 substitutions per site. This value (substitutions per site) was annualized by multiplying it by 2
218 TMRCA (e.g., 121 Ma for the total time along the branches of divergence between buntings and

219 budgerigars). To account for the uncertainty associated with using divergence time estimates
220 from distant relatives, we averaged the resulting substitutions per site per year rates (6.83×10^{-10}
221 and 6.67×10^{-10} , respectively), and we used the average rate (6.75×10^{-10}) to convert parameter
222 estimates obtained from $\delta a \delta i$ analyses into biologically relevant estimates of effective population
223 size(s) and split time(s). We converted substitution rates to substitutions/site/generation using a
224 generation time of 2.7 yr for snow buntings. We estimated generation time using survival and
225 breeding data from Smith (1994) and the method of Saether et al. (2005), in which generation
226 time (G) is calculated as $G = \alpha + (s/(1-s))$, where α is age of first breeding (1 yr) and s is annual
227 adult survival.

228

229 Results

230 Assembly produced 632,401 contigs (min = 224 bp, max = 17,453 bp) with a mean
231 length of 396.6 bp (± 0.27 bp 95% CI) for a total of 250,802,355 bp. Fully 9,194 contigs were
232 over 1 Kb in length. After identifying UCE loci and removing potential paralogs, we recovered
233 4,018 UCE loci. After filtering UCE loci for quality, calling SNPs, phasing (reconstructing
234 haplotypes), and applying additional quality filters, we identified 2,635 loci that contained data
235 for all individuals and were variable. This complete matrix of variable loci included a total of
236 9,449 SNPs (averaging 3.6 sites per locus). Per-site sequencing depth for these SNPs averaged
237 26.3 reads (± 16.9 SD). An additional 587 loci exhibited variation but the data were not of
238 sufficient quality (i.e., $GQ < 10$) among all individuals to confidently call both alleles. There
239 were 796 high-quality invariant loci (loci with invariant data, rather than an absence of data),
240 providing a full dataset of 3,431 loci with mean length of 1,153.6 bp (± 4.95 bp 95% CI). The
241 shortest locus was 228 bp, the longest 2,543 bp, and 2,482 loci were longer than 1 Kb (Fig. S2).
242 The total length of these loci was 3,957,876 bp. The distribution of SNP variation among loci
243 confidently called for all individuals is given in Fig. 2. Nucleotide diversity (π) was 0.000519
244 overall, 0.000523 for snow buntings, and 0.000493 for McKay's buntings.

245 No alleles showed fixed differences ($F_{ST} = 1.0$) between the two populations, and few
246 alleles showed strong segregation. No variable sites had an F_{ST} value above 0.9, and there were
247 only three each at 0.86 and 0.72 (Fig. S3; two of these sites were on the same locus). One of the
248 five loci with the highest F_{ST} values was Z-linked; all of the others were on different
249 chromosomes (Supplemental Information). There were 128 Z-linked loci among the 2,635

250 variable loci. As noted, only one showed high F_{ST} between the two species. The two populations
251 had an overall $F_{ST} = 0.034$, which was significant ($P = 0.03$). Discriminant Analysis of Principal
252 Components (DAPC in adegenet) assigned all individuals to their correct taxon of origin, with
253 100% probabilities for each, indicating a high level of genomic diagnosability (Fig. S4).

254 Fully 2,510 loci were in Hardy-Weinberg equilibrium; 124 were not (one was triallelic).
255 McKay's buntings had fewer unique alleles (4,238) than snow buntings (4,389), concordant with
256 the smaller population size of McKay's buntings. Bartlett's test rejected homogeneity of
257 variance between observed heterozygosity ($H_o = 0.18, 0.19$) and expected heterozygosity ($H_e =$
258 $0.20, 0.22$), but H_o did not differ from H_e ($t = -3.1653$, $df = 2633$, $P = 1.0$).

259 The four-gametes test suggested that recombination occurred in hundreds of loci. For 405
260 loci, locus lengths were shortened by IMgc to meet the four-gametes test, and for 252 loci one or
261 more individuals were removed to meet the same criteria (a few of these loci had both done;
262 IMgc automatically performs one or the other or both operations to obtain non-recombinant
263 sequence data). There were thus 15.4% to 24.9% of variable loci exhibiting patterns indicative of
264 recombination. As noted in the Methods, these sequence data, together with all other unchanged
265 sequences, were not used further; we used only SNP data for further analyses.

266 In testing our six, two-population models with $\delta a \delta i$, the highest maximum log composite
267 likelihood values were obtained for the split-with-migration model (-112.76), which made it the
268 best-fitting model for these data (model 2 in Fig. 1). We obtained successively lower likelihood
269 values for the neutral (-588.45), isolation with bidirectional migration and population growth (-
270 803.30), and isolation with population growth and no migration (-2,026.93) models. The final
271 model tested, split-bidirectional-migration, had an intermediate likelihood of -286.49. The split-
272 with-no-migration model was unstable under all conditions tried, and we could not get it to run
273 to convergence. We provide jackknifed estimates and confidence intervals for the best-fitting,
274 split-with-migration model in Table 1.

275

276 Discussion

277 Our data provided sufficient variation to answer fundamental questions about these two
278 recently diverged taxa, despite a lack of fixed genetic differences and evidence for moderate
279 levels of gene flow. Thus, our study adds to evidence showing the utility of UCEs for
280 illuminating key evolutionary attributes among populations with shallow levels of divergence

281 (e.g., Table 2). These data also provide a direct comparison to markers previously used to
282 investigate recent divergence (i.e., mtDNA, AFLPs) for these same taxa (see below). As UCEs
283 are used more frequently for population genomics, in addition to systematics, new actions
284 become desirable (Table 2). Some of the key approaches are: sequencing at increased depth,
285 genotyping individuals (determining both alleles of a locus), implementing genotype quality
286 filters, accounting for recombination, improving mutation rate estimates, and implementing
287 population genomics analytical pipelines rather than those oriented more typically toward
288 systematics. Questions often differ at population levels, but researchers are successfully applying
289 a variety of approaches that demonstrate the utility of UCEs in population genomics (Table 2).

290 In considering UCEs as a class of markers that subsamples the genome, it is useful to
291 note that our estimated substitution rates (mean of 6.75×10^{-10} substitutions per site per year) are
292 roughly an order of magnitude slower than the mutation rate estimated across the entire genome
293 of three generations of *Ficedula* flycatchers (Smeds et al. 2016). This is perhaps not surprising
294 given the conserved nature of these loci. Using a very different method to estimate substitution
295 rates (scaling UCE results to an mtDNA molecular clock), Harvey et al. (2016) estimated rates of
296 1.74×10^{-12} to 2.32×10^{-11} substitutions per site per year for UCE loci, one to two orders of
297 magnitude slower than comparable RAD-seq data from the same animals and also slower than
298 the rates we estimated here for buntings. In addition to differences in methodology, Harvey et al.
299 (2016) had shorter loci on average (mean locus length 604 bp) than loci in our study.

300 Nevertheless, more study of substitution rates in loci with UCEs is warranted because these
301 estimates are important when converting modeled demographic parameters into biological units.
302 The effect of our estimated substitution rates on our demographic estimates (if, for example, our
303 substitution rate estimates are wrong) is that for some, they are positively correlated; lower
304 substitution rates would drive effective population sizes and split times lower (N_e : *nu1*, *nu2*, and
305 *Nref* and *T* in Table 1). Migration rate (*m*) estimates in Table 1 are unaffected by substitution
306 rates.

307 The nucleotide diversity levels that we observed are approximately an order of magnitude
308 lower than typical levels across the avian genome (0.0011~0.005; Ellegren 2013). This is likely
309 the result of purifying selection acting on UCE loci, effecting an apparent lower substitution rate.
310 Our values are more similar to the values for Z-linked loci in other bird species (e.g.,
311 Balakrishnan & Edwards 2009, Huynh et al. 2010, Lavretsky et al. 2015).

312 When applied to this relatively recently derived pair of taxa, UCE results raise the
313 question of whether McKay's bunting is a full biological species. Although McKay's bunting is
314 taxonomically recognized as a species, this dataset shows substantial levels of gene flow (see
315 Wright 1943, Cabe and Alstad 1994, Winker 2010), and the lack of fixed alleles is surprising
316 given that we sampled thousands of loci and four individuals from each of two putative species;
317 we would expect several fixed differences to occur by chance through neutral processes. There
318 are some noted plumage differences between the two taxa (Maley & Winker 2007), but while our
319 results enabled 100% diagnosability (which might decline with broader sampling), they also
320 suggest widespread genomic similarity between McKay's and snow buntings (e.g., relatively low
321 F_{ST}). Given phenotypic differences between the taxa, it seems likely that there are fixed allelic
322 differences in portions of the genome not included in our data that could be detected by more
323 extensive surveys of each species' genome. The status of the taxa as biological species, however,
324 is more likely to hinge on gene flow (i.e., the geographic partitioning of traits that may be
325 responding to adaptation is not equivalent to speciation).

326 There are reports of male McKay's buntings present outside their breeding range and
327 possible hybridization between McKay's and snow buntings (Sealy 1967, 1969). Snow buntings
328 are also common on the breeding range of McKay's buntings at St. Matthew Island prior to and
329 during early portions of the breeding season, although most individuals leave before fledging
330 (Winker et al. 2002). Just one pair of snow buntings has been recorded on the island during
331 fledging (Winker et al. 2002). Observations thus suggest the possibility of hybridization; our data
332 provide a confirmation and a quantification of it. The levels of gene flow that we found, 2.8 – 5.2
333 individuals per generation (Table 1), seem rather high for two putative biological species (Rice
334 & Hostert 1993, Hostert 1997, Winker 2010). Further study will be needed to determine species
335 limits between these taxa, including larger sample sizes, broader genomic coverage, and proper
336 caution for interpreting genomic results in terms of species delimitation (Robinson et al. 2013,
337 Sukumaran & Knowles 2017).

338 In comparing UCE-based estimates of demographic parameters with those based on
339 mtDNA sequence (Maley & Winker 2010), we find little overlap (Table 3), and our UCE-based
340 split-time estimate is an order of magnitude earlier. Although effective population size estimates
341 for McKay's buntings are close (though non-overlapping), those for snow buntings are one-to-
342 two orders of magnitude smaller, a difference that is only partially explained by differences in

343 effective population size for autosomal and mtDNA estimates. These differences may also be
344 driven by the different selection regimes operating on the two marker classes. For example,
345 purifying selection on UCEs will result in background selection on linked variation in flanking
346 regions, reducing (through hitchhiking) the effective population size (Charlesworth &
347 Charlesworth 2016). Previously, mtDNA results suggested that gene flow was highly
348 asymmetric (Maley & Winker 2010), concordant with what was likely a post-glacial
349 introgression of McKay's buntings into snow buntings during a snow bunting range expansion
350 into Beringia. Our UCE-based estimates have much narrower confidence limits (and without the
351 strong asymmetry found in mtDNA; Table 3), but they do suggest moderate levels of gene flow
352 between the two species.

353

354 **Conclusions**

355 Although more work is needed to understand demographic estimates made using UCEs
356 relative to those obtained using other markers, UCEs provide rich, high-quality data for
357 population genomic studies (Table 4). They are thus an important new class of genomic marker
358 that should provide broad comparative value among diverse population genomics studies, with
359 ever-increasing value as additional studies using UCEs (or whole genomes from which UCEs
360 can be obtained) are conducted.

361

362 **Acknowledgments**

363 Oralee Nudson provided invaluable guidance for supercomputer use. Matthew Miller, Mike
364 Harvey, and Naoki Takebayashi provided helpful suggestions in bioinformatics, and Jessica
365 McLaughlin provided a script for Fig. S2. Kevin Hawkins, Phil Lavretsky, Jeff Peters, and Ryan
366 Gutenkunst provided help in running *daði*. We also thank Iris Cato, Maryanne Evans, Katie
367 Everson, Angela Gastaldi, Jessica McLaughlin, Kendall Mills, Katie Shink, Naoki Takebayashi,
368 and Sara Wilbur for helpful comments.

369

370 **Literature Cited**

371

372 Afgan E, Baker D, van den Beek M, Blankenberg D, Bouvier D, Cech M, Chilton J, Clements D,
373 Coraor N, Eberhard C, Grüning B, Guerler A, Hillman-Jackson J, Von Kuster G, Rasche E,
374 Soranzo N, Turaga N, Taylor J, Nekrutenko A, Goecks J. 2016. The Galaxy platform for

- 375 accessible, reproducible and collaborative biomedical analyses: 2016 update. *Nucleic Acids*
376 *Research* 44:W3-W10. doi: 10.1093/nar/gkw343
377
- 378 Avise JC. 1994. *Molecular Markers, Natural History, and Evolution*. Chapman & Hall, New
379 York.
380
- 381 Balakrishnan CN, Edwards SV. 2009. Nucleotide variation, linkage disequilibrium and founder-
382 facilitated speciation in wild populations of the zebra finch (*Taeniopygia guttata*). *Genetics*
383 181:645–60.
384
- 385 Bolger AM, Lohse M, Usadel B. 2014. Trimmomatic: A flexible trimmer for Illumina Sequence
386 Data. *Bioinformatics* 30:2114-2120. <http://dx.doi.org/10.1093/bioinformatics/btu170>.
387
- 388 Cabe PR, Alstad DN. 1994. Interpreting population differentiation in terms of drift and selection.
389 *Evolutionary Ecology* 8:489–492.
390
- 391 Charlesworth B, Charlesworth D. 2016. Population genetics from 1966 to 2016. *Heredity*
392 2016:1-8.
393
- 394 Danecek P, Auton A, Abecasis G, Albers CA, Banks E, DePristo MA, Handsaker RE, Lunter G,
395 Marth GT, Sherry ST, McVean G, Durbin R, 1000 Genomes Project Analysis Group. 2011.
396 The Variant Call Format and VCFtools. *Bioinformatics* 27:2156–2158.
397
- 398 Ellegren H. 2013. The evolutionary genomics of birds. *Annual Review of Ecology Evolution &*
399 *Systematics* 44:239–59.
400
- 401 Ellegren H. 2014. Genome sequencing and population genomics in non-model organisms.
402 *Trends in Ecology & Evolution* 29:51-63.
403
- 404 Faircloth BC. 2013. illumiprocessor: a trimmomatic wrapper for parallel adapter and quality
405 trimming. <http://dx.doi.org/10.6079/J9ILL>.
406
- 407 Faircloth BC 2016. PHYLUCE is a software package for the analysis of conserved genomic loci.
408 *Bioinformatics* 32:786-788. doi: 10.1093/bioinformatics/btv646.
409
- 410 Faircloth BC, McCormack JE, Crawford NG, Harvey MG, Brumfield RT, Glenn TC. 2012.
411 Ultraconserved elements anchor thousands of genetic markers for target enrichment
412 spanning multiple evolutionary timescales. *Systematic Biology* 61:717-726.
413 doi:10.1093/sysbio/SYS004
414
- 415 Faircloth BC, Branstetter MG, White ND, Brady SG. 2015. Target enrichment of ultraconserved
416 elements from arthropods provides a genomic perspective on relationships among
417 Hymenoptera. *Molecular Ecology Research* 15:489-501. doi:10.1111/1755-0998.12328
418
- 419 Felsenstein J. 2005. Accuracy of coalescent likelihood estimates: Do we need more sites, more
420 sequences, or more loci? *Molecular Biology and Evolution* 23:691-700.

- 421
422 Garrigan D, Kingan SB, Pilkington MM, Wilder JA, Cox MP, Soodyall H, Strassmann B,
423 Destro-Bisol G, de Knijff P, Novelletto A, Friedlaender J, Hammer MF. 2007. Inferring
424 human population sizes, divergence times and rates of gene flow from mitochondrial, X and
425 Y chromosome resequencing data. *Genetics* 177:2195-2207.
426
- 427 Gilbert PS, Chang J, Pan C, Sobel E, Sinsheimer JS, Faircloth BC, Alfaro ME. 2015. Genome-
428 wide ultraconserved elements exhibit higher phylogenetic informativeness than
429 traditional gene markers for the fish series Percomorpha. *Molecular Phylogenetics &*
430 *Evolution* 92:140-146. doi:10.1016/j.ympev.2015.05.027
431
- 432 Glenn TC, Nilsen R, Kieran TJ, Finger Jr JW, Pierson TW, Bentley KE, Hoffberg SL, Louha S,
433 García-De-Leon FJ, Portilla MAR, Reed K, Anderson JL, Meece JK, Aggrey S, Rekaya R,
434 Alabady M, Belanger M, Winker K, Faircloth BC. 2017. Adapterama I: Universal stubs and
435 primers for thousands of dual-indexed Illumina libraries (iTru & iNext). *Molecular Ecology*
436 *Resources* (accepted), available at <http://biorxiv.org/content/early/2016/06/15/049114>
437
- 438 Goudet J, Raymond M, de Meeüs T, Rousset F. 1996. Testing differentiation in diploid
439 populations. *Genetics* 144:1933-1940.
440
- 441 Grabherr MG, Haas BJ, Yassour M, Levin JZ, Thompson DA, Amit I, Adiconis X, Fan L,
442 Raychowdhury R, Zeng Q, Chen Z, Mauceli E, Hacohen N, Gnirke A, Rhind N, di Palma F,
443 Birren BW, Nusbaum C, Lindblad-Toh K, Friedman N, Regev A. 2013. Trinity:
444 reconstructing a full-length transcriptome without a genome from RNA-Seq data. *Nature*
445 *Biotechnology* 29:644–652.
446
- 447 Gutenkunst RN, Hernandez RD, Williamson SH, Bustamante CD. 2009. Inferring the joint
448 demographic history of multiple populations from multidimensional SNP data. *PLoS*
449 *Genetics* 5:e1000695.
450
- 451 Harvey MG, Brumfield RT. 2015. Genomic variation in a widespread Neotropical bird (*Xenops*
452 *minutus*) reveals divergence, population expansion, and gene flow. *Molecular Phylogenetics*
453 *& Evolution* 83:305-316.
454
- 455 Harvey MG, Smith BT, Glenn TC, Faircloth BC, Brumfield RT. 2016. Sequence capture versus
456 restriction site associated DNA sequencing for shallow systematics. *Systematic Biology*
457 65:910-924.
458
- 459 Harvey MG, Aleixo A, Ribas CC, Brumfield RT. 2016. Habitat preference predicts genetic
460 diversity and population divergence in Amazonian birds. *American Naturalist* 190:631-648.
461
- 462 Hillis DM, Moritz C, Mable BK (eds). 1996. *Molecular Systematics, 2nd ed.* Sinauer Associates,
463 Inc. Sunderland, Massachusetts.
464
- 465 Hostert EE. 1997. Reinforcement: A new perspective on an old controversy. *Evolution* 51:697-
466 702.

- 467
468 Huynh LY, Maney DL, Thomas JW. 2010. Contrasting population genetic patterns within the
469 white-throated sparrow genome (*Zonotrichia albicollis*). *BMC Genetics* 11:96.
470
- 471 Jombart T, Ahmed I. 2011. adegenet 1.3-1: new tools for the analysis of genome-wide SNP data.
472 *Bioinformatics* 27:3070-3071. doi:10.1093/bioinformatics/btr521
473
- 474 Jorde LB, Watkins WS, Bamshad MJ, Dixon ME, Ricker CE, Seielstad MT, Batzer MA. 2000.
475 The distribution of human genetic diversity: A comparison of mitochondrial, autosomal,
476 and Y-chromosome data. *American Journal of Human Genetics* 66:979-988.
477
- 478 Lavretsky P, Peters JL, Winker K, Bahn V, Kulikova I, Zhuravlev YN, Wilson RE, Barger C,
479 Gurney K, McCracken KG. 2016. Becoming pure: identifying generational classes of
480 admixed individuals within lesser and greater scaup populations. *Molecular Ecology*
481 25:661-674.
482
- 483 Leaché AD, Chavez AS, Jones LN, Grummer JA, Gottscho AD, Linkem CW 2015.
484 Phylogenomics of Phrynosomatid lizards: Conflicting signals from sequence capture versus
485 restriction site associated DNA sequencing. *Genome Biology and Evolution* 7:706-719.
486
- 487 Li H. 2013. Aligning sequence reads, clone sequences and assembly contigs with BWA-MEM.
488 <https://arxiv.org/abs/1303.3997>
489
- 490 Li H, Durbin R. 2010. Fast and accurate long-read alignment with Burrows-Wheeler transform.
491 *Bioinformatics* 26:89-95.
492
- 493 Li H, Handsaker B, Wysoker A, Fennell T, Ruan J, Homer N, Marth G, Abecasis G, Durbin R,
494 1000 Genome Project Data Processing Subgroup. 2009. The sequence alignment/map
495 format and SAMtools. *Bioinformatics* 25:2078-2079.
496
- 497 Lischer HEL, Excoffier L. 2012. PGDSpider: An automated data conversion tool for connecting
498 population genetics and genomics programs. *Bioinformatics* 28:298-299.
499
- 500 Maley JM, Winker K. 2007. The utility of juvenal plumage in diagnosing species limits: An
501 example using buntings in the genus *Plectrophenax*. *Auk* 124:907-915.
502
- 503 Maley JM, Winker K. 2010. Diversification at high latitudes: Speciation of buntings in the genus
504 *Plectrophenax* inferred from mitochondrial and nuclear markers. *Molecular Ecology*
505 19:785-797.
506
- 507 Manthey JD, Campillo LC, Burns KJ, Moyle RG. 2016. Comparison of target-capture and
508 restriction-site associated DNA sequencing for phylogenomics: A test in cardinalid tanagers
509 (Aves, genus: *Piranga*). *Systematic Biology* 65:640-650.
510
- 511 Marcovitz A, Jia R, Bejerano G. 2016. "Reverse genomics" predicts function of human
512 conserved noncoding elements. *Molecular Biology & Evolution* 33:1358-1369.
513

- 514 Mason NA, Olvera-Vital A, Lovette IJ, Navarro-Sigüenza AG. 2018. Hidden endemism, deep
515 polyphyly, and repeated dispersal across the Isthmus of Tehuantepec: Diversification of the
516 White-collared Seedeater complex (Tharupide: *Sporophila torqueola*). *Ecology and*
517 *Evolution* 8:1867-1881.
- 518
- 519 McCormack JE, Harvey MG, Faircloth BC, Crawford NG, Glenn TC, Brumfield RT. 2013. A
520 phylogeny of birds based on over 1,500 loci collected by target enrichment and high-
521 throughput sequencing. *PLoS ONE* 8:e54848. doi:10.1371/journal.pone.0054848
522
- 523 McKenna A, Hanna M, Banks E, Sivachenko A, Cibulskis K, Kernytsky A, Garimella K,
524 Altshuler D, Gabriel S, Daly M, DePristo MA. 2010. The Genome Analysis Toolkit: a
525 MapReduce framework for analyzing next-generation DNA sequencing data. *Genome*
526 *Research* 20:1297-303.
- 527
- 528 Milne I, Stephen G, Bayer M, Cock PJA, Pritchard L, Cardle L, Shaw PD, Marshall D. 2013.
529 Using Tablet for visual exploration of second-generation sequencing data. *Briefings in*
530 *Bioinformatics* 14:193-202.
- 531
- 532 Oswald JA, Harvey MG, Rensen RC, Foxworth DU, Cardiff SW, Dittmann DL, Megna LC,
533 Carling MD, Brumfield RT. 2016. Willet be one species or two? A genomic view of the
534 evolutionary history of *Tringa semipalmata*. *Auk* 133:593-614.
- 535
- 536 Pearse DE, Crandall KA. 2004. Beyond F_{ST} : Analysis of population genetic data for
537 conservation. *Conservation Genetics* 5:585-602.
- 538
- 539 Rohland N, Reich D. 2012. Cost-effective, high-throughput DNA sequencing libraries for
540 multiplexed target capture. *Genome Research* 22:939-946.
- 541
- 542 Rice WR, Hostert EE. 1993. Laboratory experiments on speciation: What have we learned in 40
543 years? *Evolution* 47:1637-1653.
- 544
- 545 Robinson, J. D., A. C. Coffman, M. J. Hickerson, and R. N. Gutenkunst. 2014. Sampling
546 strategies for frequency spectrum-based population genomic inference. *BMC Evolutionary*
547 *Biology* 14:254.
- 548
- 549 Sæther B-E, Lande R, Engen S, Weimerskirch H, Lillegård M, Altwegg R, Becker PH,
550 Bregnballe J, Brommer JW, McCleery RH, Merilä J, Nyholm E, Rendell W, Robertson
551 RR, Tryjanowski P, Visser ME. 2005. Generation time and temporal scaling of bird
552 population dynamics. *Nature* 436:99-102.
- 553
- 554 Sethuraman A, Hey J. 2015. IMA2p—parallel MCMC and inference of ancient demography under
555 the Isolation with migration (IM) model. *Molecular Ecology Resources* 16:206-215. DOI:
556 <http://dx.doi.org/10.1111/1755-0998.12437>
557
- 558 Sealy SG. 1967. The occurrence and possible breeding of McKay's bunting on St. Lawrence
559 Island, Alaska. *Condor* 69:531-532.

- 560
561 Sealy SG. 1969. Apparent hybridization between snow bunting and McKay's bunting on St.
562 Lawrence Island, Alaska. *Auk* 86:350-351.
563
- 564 Smeds L, Qvarnström A, Ellegren H. 2016 Direct estimate of the rate of germline mutation in a
565 bird. *Genome Research* 26:1211-1218.
566
- 567 Smith RD. 1994. Snow buntings *Plectrophenax nivalis*: the behavioural ecology and site use of
568 an itinerant flock species in the nonbreeding season. PhD thesis, University of Glasgow.
569
- 570 Smith BT, Harvey MG, Faircloth BC, Glenn TC, Brumfield RT. 2014. Target capture and
571 massively parallel sequencing of ultraconserved elements (UCEs) for comparative studies at
572 shallow evolutionary time scales. *Systematic Biology* 63: 83-95. doi:10.1093/sysbio/syt061
573
- 574 Sukumaran J, Knowles LL. 2017. Multispecies coalescent delimits structure, not species.
575 *Proceedings of the National Academy of Sciences* 114: 1607–12.
576
- 577 Tamura K, Stecher G, Peterson D, FilipSKI A, Kumar S. 2013. MEGA6: Molecular Evolutionary
578 Genetics Analysis version 6.0. *Molecular Biology and Evolution* 30:2725-2729.
579
- 580 Winker K. 2010. Subspecies represent geographically partitioned variation, a goldmine of
581 evolutionary biology, and a challenge for conservation. *Ornithological Monographs* 67:6-
582 23.
583
- 584 Winker K, Gibson DD, SOWLS AL, Lawhead BE, Martin PD, Hoberg EP, Causey D. 2002. The
585 birds of St. Matthew Island. *Wilson Bulletin* 114:491-509.
586
- 587 Woerner AE, Cox MP, Hammer MF. 2007. Recombination-filtered genomic datasets by
588 information maximization. *Bioinformatics* 23:1851-1853.
589
- 590 Wright S. 1943. Isolation by distance. *Genetics* 28:114–138.
591
- 592 Zarza E, Faircloth BC, Tsai WLE, Bryson Jr. RW, Klicka J, McCormack JE. 2016. Hidden
593 histories of gene flow in highland birds revealed with genomic markers. *Molecular Ecology*
594 25:5144-5157.
595
- 596 Zhang Z, Schwartz S, Wagner L, Miller W. 2000. A greedy algorithm for aligning DNA
597 sequences. *Journal of Computational Biology* 7:203-14.
598

Figure 1(on next page)

Population divergence models tested using $\delta a \delta i$

Population divergence models tested using $\delta a \delta i$, varying from 1) neutral (no divergence), to a series of different two-population models with an ancestral population diverging into two populations (at time T): 2) split with migration (gene flow), 3) split with no migration, 4) isolation with bidirectional migration and population growth, 5) isolation with population growth and no migration, and 6) a derivative of model 2 with bidirectional migration.

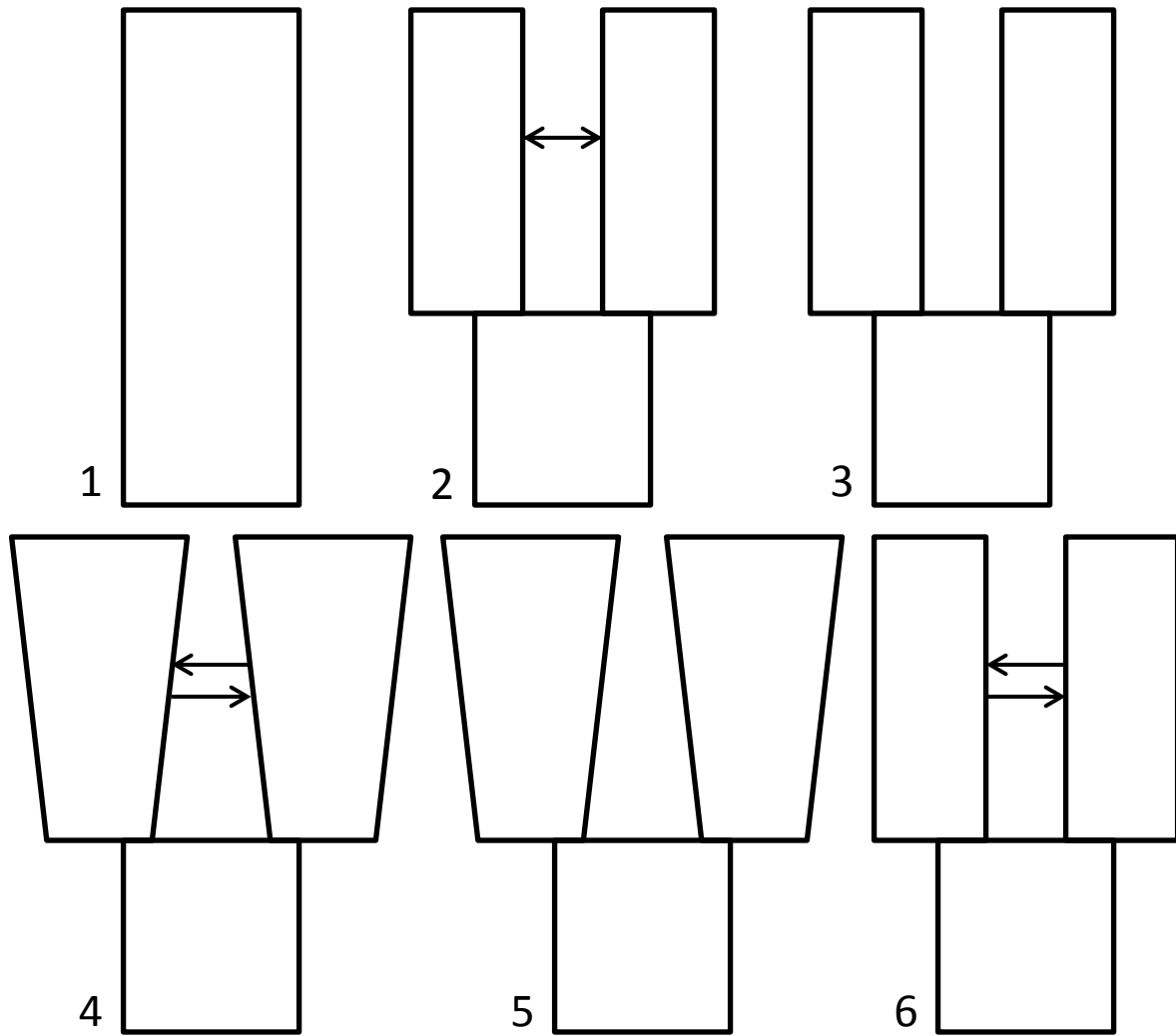


Figure 2 (on next page)

Distribution of single nucleotide polymorphisms (SNPs) per locus

Distribution of single nucleotide polymorphisms (SNPs) per locus among 3,431 confidently called loci.

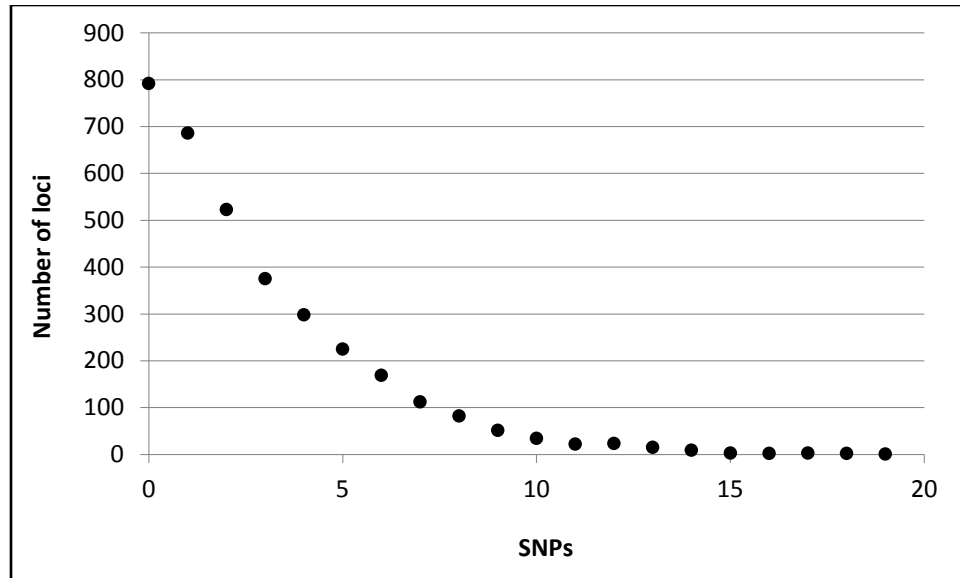


Table 1 (on next page)

Demographic model parameters

Demographic model parameters from the $\delta a \delta i$ split-migration model (variables in first column) and estimates in biological units (defined in final column), with 95% CIs determined by jackknifed datasets. The two migration rates use the two different effective population sizes in their calculation.

1 **Table 1.** Demographic model parameters from the $\delta a \delta i$ split-migration model (variables in first
 2 column) and estimates in biological units (defined in final column), with 95% CIs determined by
 3 jackknifed datasets. The two migration rates use the two different effective population sizes in
 4 their calculation.

5

Model parameters	Parameter (+ 95% CI)	Estimates (+ 95% CI)	Lower-upper bounds	Biological units
nu1 (pop size McKay's)	3.52 (\pm 0.54)	109,330 (\pm 16,790)	92,540 - 126,120	individuals McKay's
nu2 (pop size snow)	5.95 (\pm 1.79)	184,991 (\pm 55,523)	129,467 - 240,514	individuals snow
T (split time)	1.44 (\pm 0.37)	241,491 (\pm 62,429)	179,061 - 303,920	years
m_1 (migration)	1.65 (\pm 0.39)	2.90 (\pm 0.10)	2.8 - 3.0	individuals using nu1
m_2 (migration)	1.65 (\pm 0.39)	4.90 (\pm 0.35)	4.6 - 5.2	individuals using nu2
theta	249.97 (\pm 32.71) ^a	31,072 (\pm 4,066) ^a	27,006 - 35,138	ancestral population individuals

^a - Nref

6

Table 2 (on next page)

Some bioinformatic and analytical attributes typical of population genomics studies using UCEs.

Some bioinformatic and analytical attributes typical of population genomics studies and some of the variation among researchers in applying them to different questions using UCE data.

1 **Table 2.** Some bioinformatic and analytical attributes typical of population genomics studies
2 and some of the variation among researchers in applying them to different questions using UCE
3 data.

4
5

Popgen attribute	Smith et al. 2014	Harvey et al. 2016	Zarza et al. 2016	Oswald et al. 2016	this study
Genotyping	no	yes	yes	yes	yes
GQ filters	no	yes	yes	?	yes
recombination	no	no	no	no	yes
substitution rates	yes	yes	no	yes	yes
population differentiation	yes	yes	yes	yes	yes
gene flow rates	yes	no	no	yes	yes
effective pop. sizes	yes	yes	no	yes	yes
heterozygosity	no	yes	no	no	yes

6
7
8
9
10

Table 3 (on next page)

Comparing UCE and mtDNA demographic estimates

Comparing bunting UCE results of demographic estimates with those obtained using mtDNA by Maley & Winker (2010).

1 **Table 3.** Comparing bunting UCE results of demographic estimates with those obtained using
2 mtDNA by Maley & Winker (2010).

3

Parameter	UCEs	mtDNA
split time	179-304 Kyr	18-74 Kyr
N_e McKay's	93-126 K	170-680 K
N_e snow	0.13-0.24 million	6-24 million
m	2.8-5.2	0.05-753

4

5

6

Table 4 (on next page)

Comparing avian UCE population genomic characteristics.

1 **Table 4.** Comparing avian UCE population genomic characteristics.

2

Species	# loci	% polymorphic	Nucleotide diversity	Heterozygosity	Source
<i>Tringa semipalmata</i>	4635	94	n.a.	n.a.	Oswald et al. 2016
<i>Cymbilaimus lineatus</i>	776	53	0.0019	n.a.	Smith et al. 2014
<i>Xenops minutus</i>	1368	73	0.0019	n.a.	Smith et al. 2014
<i>Schiffornis turdina</i>	851	77	0.0003	n.a.	Smith et al. 2014
<i>Querula purpurata</i>	1516	58	0.0013	n.a.	Smith et al. 2014
<i>Microcerculus marginatus</i>	1077	60	0.0015	n.a.	Smith et al. 2014
<i>Plectrophenax</i> spp. (2)	3431	77	0.0005	0.20-0.22	this study
average of 40 Amazonian species	2416	varied	0.0011	~0.44 (1 sp.)	Harvey et al. 2016

3