

Water (or the lack thereof), management, and conservation of an endangered desert wetland obligate, *Lilaeopsis schaffneriana* var. *recurva*

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Huachuca water umbel, *Lilaeopsis schaffneriana* var. *recurva*, is a federally endangered desert wetland plant restricted to southeastern Arizona, USA, and adjacent portions of Mexico. Very little is known about the factors that influence the species presence, absence, and abundance, including the effects of wetland drying caused by drought or water withdrawals. We conducted three lines of research to address the lack of information about the species' ecology using observational field research and controlled experiments. First, we examined ecological correlates of the species presence and absence along a perennial stream in Southeast Arizona where the species has been established for over a decade. Next, we used a full-factorial experiment to examine the species' drought resistance and drought resilience. Last, we monitored transplanting efforts of the species along the upper Rio Yaqui. We found that distance to the nearest water umbel patch centroid—a measure of historical factors—and distance to surface water were the best predictors of the species' presence and leaf density. The drought experiments demonstrated that leaf density is an important factor influencing both resistance and resilience: low leaf-density was associated with replicates reaching critical conditions approximately one full day before high leaf-density plants, and high leaf-density plants rebounded at nearly twice the rate of low-density plants when water became available. Transplants appeared to be most successful in areas where water levels remained relatively constant, herbaceous-layer competitors were rare, and water velocity was low. These results indicate that protecting existing populations by ensuring stable water sources is the most important action for conserving the species in current locations and future transplant locations.

1 **Water (or the lack thereof), management, and conservation of an endangered**
2 **desert wetland obligate, *Lilaeopsis schaffneriana* var. *recurva***

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16

17 **Abstract**

18 Huachuca water umbel, *Lilaeopsis schaffneriana* var. *recurva*, is a federally endangered desert
19 wetland plant restricted to southeastern Arizona, USA, and adjacent portions of Mexico. Very
20 little is known about the factors that influence the species presence, absence, and abundance,
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27 monitored transplanting efforts of the species along the upper Rio Yaqui. We found that distance
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29 surface water were the best predictors of the species' presence and leaf density. The drought
30 experiments demonstrated that leaf density is an important factor influencing both resistance and
31 resilience: low leaf-density was associated with replicates reaching critical conditions
32 approximately one full day before high leaf-density plants, and high leaf-density plants
33 rebounded at nearly twice the rate of low-density plants when water became
34 available. Transplants appeared to be most successful in areas where water levels remained
35 relatively constant, herbaceous-layer competitors were rare, and water velocity was low. These
36 results indicate that protecting existing populations by ensuring stable water sources is the most
37 important action for conserving the species in current locations and future transplant locations.

38

39 **Keywords:** Huachuca water umbel, ecology, drought, competition, endangered species

40

41 Introduction

42 Conserving rare species and their ecosystems requires knowing the threats they face. Desert
43 wetlands are among the most endangered ecosystems in the world because water withdrawals for
44 anthropogenic uses have accelerated over the past century, and will likely continue into the
45 future (Hendrickson and Minckley 1984). Numerous wetland species in the U.S. Southwest and
46 adjacent portions of Mexico are protected under state and federal laws specifically because of
47 altered hydrology (e.g., Rio Yaqui fishes [USFWS 1995], Chiricahua leopard frog [USFWS
48 2002]), with water withdrawals as the primary driver. If regional climate change projections
49 (Seager et al., 2007) are realized, then the threats posed to these species by hydrological changes
50 will be further exacerbated. Understanding the drought ecology of imperiled wetland obligates is
51 needed to devise conservation and management strategies that may allow such species to survive
52 and recover.

53 The Huachuca water umbel (*Lilaeopsis schaffneriana* var. *recurva* (Hill) Affolter;
54 hereafter, water umbel) is listed as an endangered species under the U.S. Endangered Species
55 Act because of, “degradation and destruction of habitat resulting from livestock overgrazing,
56 water diversions, dredging, and groundwater pumping” (USFWS 1997). The species is known
57 from over two dozen sites in the U.S., including San Bernardino and Leslie Canyon National
58 Wildlife Refuges, sites in Coronado National Forest, along the San Pedro River, and adjacent
59 parts of Mexico (USFWS 2014). Desert wetlands (cienegas) in the region have attracted
60 attention for over a century, but the water umbel has not been studied extensively. Research
61 since Affolter’s monograph on the genus (Affolter, 1985) has included the phylogenetic
62 relationships of *Lilaeopsis*, demonstrating the uniqueness of *L. s.* var. *recurva* (Bone et al., 2011,
63 Fehlberg and Allen, 2014), and the threat to *L. s.* var. *recurva* posed by livestock trampling
64 (Malcom & Radke, 2008). But very little quantitative information exists on the basic ecological
65 correlates or requirements of the species. Titus and Titus (Titus & Titus, 2008a,b) examined
66 water umbel monitoring and prospects for reintroducing the species to sites from which it had
67 been extirpated. They found evidence for interspecific competition, negative effects of drought
68 conditions, and general success for transplants as long as drought was not too severe. The same
69 authors also evaluated the seed bank of an historical cienega and showed that water umbel seeds
70 were present and germinated in a greenhouse even though the plant was no longer growing at the
71 site (Titus & Titus, 2009).

72 Here we address three basic questions about water umbel ecology:

- 73 1. What are the ecological correlates of water umbel presence, absence, and abundance
74 in a natural setting?
- 75 2. What characteristics of water umbel morphology are associated with variation in the
76 species' drought resistance and resilience?
- 77 3. What factors influence water umbel transplant success and failure?

78 Answering the first question is important to managing habitat conditions that can help improve
79 the conservation status of the water umbel. If overhead canopy is the most important
80 determinant of water umbel presence, absence, and abundance, then local vegetation
81 management may be of first order; if soil saturation is more important, then watershed-level
82 management will be critical. Answers to the second question are needed to know what
83 characteristics to measure in monitoring programs and to inform management targets for the
84 umbel. If leaf density is predictive of resistance or resilience to drying conditions, then measures
85 of leaf density are needed to complement measurements of aerial extent in a water umbel status
86 monitoring program. Third, because the water umbel has already been lost from many sites
87 where it was once found, determining the factors that improve transplant success can help
88 identify the best reintroduction sites. We show that historical factors and distance to surface
89 water are primary determinants of water umbel presence and abundance; that leaf density is a
90 strong predictor of drought resistance and resilience; and transplant success is associated with
91 sites exhibiting stable water levels, low-velocity water, and relatively few herbaceous
92 competitors. We discuss how these results inform management and conservation decisions
93 across the species' range.

94

95 **Materials & Methods**

96 Study areas

97 In 2007, Huachuca water umbel occurred in thirteen discrete patches along 100 meters of Leslie
98 Creek in Leslie Canyon National Wildlife Refuge (LCNWR), Cochise County, Arizona, at an
99 elevation of 1,900 m. Most of the extant patches are the result of transplanting that occurred in
100 the mid-1990s from stock collected at San Bernardino National Wildlife Refuge (SBNWR), but
101 no records are available to differentiate transplanted versus natural patches. The riparian canopy
102 of the Creek is dominated by *Fraxinus velutina*, with a few interspersed *Salix gooddingii* and

103 *Populus fremonti*. Understory vegetation includes young *Fraxinus* and *Salix*, but is otherwise
104 lacking. Potential direct competitors of the water umbel include *Rorippa nasturtium-aquaticum*
105 and *Sporobolus giganteus*. Leslie Creek is subject to annual flash flooding, with peak flows of
106 ca. 10,000cfs common, but is also subject to periods of very low water levels during the droughts
107 that characterize the region. Leslie Canyon is deep and steep-walled, but the gradient of Leslie
108 Creek is very low where the water umbel lives.

109 Huachuca water umbel formerly occurred along Black Draw, SBNWR, but was
110 extirpated by the late 1990s as surface waters receded with a long drought. Restoration work,
111 however, has increased and stabilized the local water table to create conditions that may be ideal
112 for re-establishing the water umbel. The transplant sites at SBNWR all occur in or near Black
113 Draw, the northernmost tributary of the Rio Yaqui, at an elevation of 1,190m. Transplants
114 occurred in a *Populus*-dominated corridor with little understory vegetation and in the shallow
115 portion of a palustrine wetland dominated by a mix of *Typha dominigensis* and *Scirpus*
116 *americana*. Black Draw, like Leslie Creek, is subject to annual flooding, with peak flows of
117 >10,000cfs. Black Draw is deeply incised and the terrain adjacent to the stream thalweg is
118 relatively steep compared to the areas of Leslie Creek where the water umbel grows.

119

120 Ecological correlates

121 We studied the ecological correlates of the Leslie Creek water umbel population using a
122 systematic-random sampling design. Samples were taken every 10m along the creek from a
123 random starting point near the downstream end of the plant's local distribution. We took three
124 samples perpendicular at distances of 0.1, 0.25, and 0.75m from the water's edge at each 10m
125 point. We visually estimated canopy cover in five directions for each group of three samples: 45
126 and 135 degrees from horizontal parallel to the stream, 45 and 135 degrees from horizontal
127 perpendicular to the stream, and 90 degrees (directly overhead). That is, we estimated canopy
128 cover for all three samples at a point using five subsamples.

129 At each of the three points we measured the density of water umbel leaves and any
130 competitors, if present, using a 5x5-cm wire frame. Additionally, we collected a 5cm deep,
131 1.75cm diameter soil plug and placed it in a Whirl-Pak that we tightly sealed. We returned the
132 soil samples to the lab and massed them to the nearest milligram the same day as the collection.
133 These samples were dried at 52°C for 48h, then re-massed to the nearest milligram. We

134 determined water content by mass by subtracting the dry mass from the wet mass, then dividing
135 by the wet mass.

136 We hypothesize that history plays a strong role in determining the local patch
137 distribution, but we need to disentangle the roles of spatio-temporal dependency of water umbel
138 presence, absence, and abundance, versus the ecological causes of current variation. To do so,
139 we measured the distance from each sample point to the nearest patch centroid as a surrogate for
140 history. That is, if we assume that patches expand outward from the original transplant location
141 (or a small refuge that survived a drought) as ecological conditions allow, then the closer we are
142 to the nearest patch centroid the more likely water umbel is to be detected. This variable may
143 seem tautological, but the fact that we sampled at sites very near patch centroids that lacked
144 water umbel highlights that informative variation is captured.

145 We defined eight models to examine the relationship between water umbel densities and
146 the environmental variables (Table 1) before any analysis. These models were constructed based
147 on our observations of the conditions in which the water umbel occurs along Leslie Creek. The
148 distribution of water umbel densities highlights the need to consider the factors that affect both
149 presence/absence and leaf densities (Figure 1). We intended to include potential herbaceous-
150 layer competitors in the models to investigate interspecific competition, but when we attempted
151 to use a zero-inflated negative binomial model to fit both the presence/absence and count
152 components simultaneously to all measured variables, the resulting matrices were singular. The
153 offending predictor variable was herbaceous-layer competitor; we found no strong relationship to
154 water umbel densities and therefore excluded competitor from the models (see Article S1 for a
155 variety of analysis supplements). We compared models using Akaike's Information Criterion,
156 corrected for small sample size (AIC_C ; Burnham & Anderson, 2002) using the `AICcmodavg`
157 package (Mazerolle, 2016) for R 3.2.3 (R Core Team, 2015). We calculated the Akaike weights
158 (w_i , for each i^{th} model), to evaluate the degree of support for the models, given the data.

159

160 *Drought resistance and resilience*

161 During field work with the water umbel we observed that patches or significant portions of
162 patches are often characterized by either high- or low-density leaves. We used a full-factorial
163 experimental design with two treatments to determine if leaf density is an accurate representation
164 of the species' drought resistance (i.e., ability to withstand drying conditions) and resilience (i.e.,

165 ability to recover from drought). The first treatment was high versus low leaf density, reflecting
166 the densities observed in the field. The second treatment was the water umbel's condition in
167 each replicate: green, $\frac{1}{4}$ -wilted, $\frac{1}{2}$ -wilted, wilted, and dead, with plants considered "dead" when
168 no green coloration was visible. We randomized treatments among the replicates. These
169 treatments resulted in forty replicates in full crosses, including four controls per leaf density
170 treatment.

171 Water umbel plugs measuring 5x5cm were taken from plants kept in cultivation at the
172 Wildlife Refuge headquarters starting in 2003. Visual inspection of the stock readily revealed
173 two distinct densities among the stocks. We extracted plugs from the stock and randomly
174 assigned each plug to either the high or low leaf-density treatment, then counted all leaves from
175 the replicates to describe those leaf densities. Mean leaf densities = 8 leaves per cm^2 for high
176 leaf density and 2 leaves per cm^2 for low leaf density (see Article S1). We placed replicates in
177 individual 600-mL containers, maintained the experiment outdoors under a sunshade, and
178 randomized placement on shelves daily to avoid any systematic day-lighting or weather bias.

179 We watered the experimental replicates for three days to ensure that each was well-
180 established, and recorded each plant's condition daily. To test drought resistance, we began
181 withholding water from all replicates (except the eight controls) starting on the fourth day.
182 Water was withheld from each replicate until it reached the pre-assigned critical value for the
183 second treatment. For example, if a replicate was designated as $\frac{1}{2}$ -wilted, then water was
184 withheld until the first day that a $\frac{1}{2}$ -wilted condition was recorded. The time from experiment
185 initiation until the critical value was reached was the response variable "drought resistance"
186 portion of the experiment.

187 After the critical condition was reached, we watered each plant daily for fifteen days to
188 quantify recovery from a given drought-induced condition. We recorded the condition—using
189 the same 5-level index used in the drought resistance experiment—of each plant daily. We
190 watered the controls daily throughout the experiment, until the final replicate had reached the
191 15th day of recovery in the resilience phase of the experiment. Because the resilience portion of
192 the experiment immediately followed the resistance experiment, and because the time-to-critical
193 for individual replicates was staggered, the ending times for the 15-day recovery were staggered
194 as well.

195 The drought response variable of interest was time-to-critical condition, and the drought

196 resilience response was the rate of recovery. We analyzed the data from both experiments using
197 a general linear model and used AIC_C to evaluate model parsimony (see Article S1). Statistical
198 significance was set at $\alpha = 0.05$.

199

200 Transplants

201 We transplanted fourteen water umbel plugs measuring 7.5x15cm at three locations on SBNWR
202 between 2005 and 2007. We placed four at the “Minckley” site in Black Draw, which is a low-
203 gradient lotic system; seven at the “Twin Overflow” site in Black Draw, which is a higher-
204 gradient lotic system; and three in the “Twin II” site, a palustrine lentic site. We monitored plug
205 conditions weekly for the first month following transplant, and monthly or bi-monthly thereafter.
206 The small number of suitable transplant sites, coupled with the small number of plugs available
207 for transplanting, precluded rigorous quantitative analysis; we therefore examine the outcomes
208 qualitatively.

209

210 **Results**

211 Ecological correlates

212 We collected data at 53 points along Leslie Creek in July 2007. Models 4 and 1 accounted for
213 93% of model selection uncertainty and the remaining six were relatively uninformative (Table
214 2). Both distance to nearest centroid (presence-absence of the zero-inflated model; Figure 1) and
215 distance to the edge of the creek (count component of model; Figure 2) were informative
216 variables (Table 3). Although canopy cover measures were included in Model 1, the variables
217 were not well-supported in any models. Our second *post hoc* model (“Post-2” in Table 1) was
218 substantially better than Model 4 ($\Delta AIC_C = 5.02$), and the parameter estimates were very similar
219 to estimates weighted by the w_{AIC} for Models 1 and 4 (see Article S1).

220

221 Drought resistance and resilience

222 We conducted the drought resistance and resilience experiments between June and July 2007.
223 All controls for both high- and low-density leaf treatments survived and maintained their
224 condition during the experiment. The interaction model ($AIC_C = 79.8$) provided a much better fit
225 to the data than the no-interaction model ($AIC_C = 153.1$). Water umbel leaf density was strongly
226 related to drought resistance ($F_{7,24} = 12.3$, $P = 1.38e^{-6}$): high leaf density conferred over one

227 additional day of drought resistance over low leaf densities ($\beta = -1.5$, $SE = 0.46$, $p = 0.003$). On
228 average, the pre-assigned critical condition (treatment 2 in *Materials & Methods*) was reached
229 1.5 d earlier by low leaf density treatments than by high leaf density treatments (Figure 3).

230 Water umbel leaf density was also directly related to drought resilience. The model with
231 an interaction between leaf density and day of recovery had a substantially better fit ($AIC_C =$
232 1394 ; $F_{6,473} = 102$, $P < 2e^{-16}$) than a purely additive model ($AIC_C = 1398.8$) or a full interaction
233 model ($AIC_C = 1401.5$). On average, high density treatments recovered at twice the rate of low
234 density treatments (Figure 4), with the day estimate = 0.108 ($SE = 0.015$) and the day \times low-
235 density term estimate = -0.056 ($SE = 0.022$, $p = 0.009$).

236

237 Transplants

238 Of the fourteen water umbel plugs transplanted during the study, eleven were extant as of mid-
239 2007 and four showed signs of expansion. The three patches at the Twin II site were apparently
240 out-competed by *Typha* and *Scirpus*, and have disappeared from the site. One of the four 2005
241 transplant plugs was lost at the Twin Overflow site when soil moisture dropped in association
242 with water flow cessation. Three of four 2007 transplant plugs at the Minckley site and one at
243 Twin Overflow, all on relatively level micro-sites adjacent to slow water (water velocity $\bar{x} =$
244 0.01m/s), expanded approximately two-fold. The five remaining Twin Overflow transplants and
245 the one remaining Minckley plug remained alive for several years, but either contracted or never
246 expanded; these patches are on higher-gradient micro-sites and are adjacent to faster-flowing
247 water (water velocity $\bar{x} = 0.21\text{m/s}$). Over the longer term, water umbel transplants were
248 negatively affected by extended drought conditions, herbaceous competitors, trampling by
249 ungulates (*Pecari tajacu*), removed by severe flood events, and buried by sediment loads from
250 floods. By 2015, the transplants survived only at the Minckley site, but had expanded in area
251 occupied.

252

253 **Discussion**

254 Conserving and managing imperiled species and their ecosystems requires understanding the
255 species' ecology. The Huachuca water umbel, a wetland obligate, is clearly limited in its extent
256 by water availability. Despite this general knowledge, very little is known about the ecology of
257 the water umbel. How much soil saturation is enough? How important are competitors to

258 limiting the species' extent? We used a combination of observational field research, greenhouse
259 experiments, and transplants to fill in some of the gaps in our knowledge about the species.

260 The results from the observational research suggests two basic filters act on water umbel
261 presence, absence, and abundance. First, the data illustrate the role of history in determining
262 whether the species is present at any given area. On one hand this isn't terribly surprising, but
263 the biology of the water umbel suggests that colonization might be faster. Specifically, the water
264 umbel's primary mode of reproduction is vegetative, but Leslie Creek is not lined by water
265 umbel even though patches are found along a significant portion of the perennial section. The
266 drying and wetting cycles of the creek may drive water umbel patch contractions and expansions.
267 When the water level of Leslie Creek dropped drastically because of extended drought in 2005
268 and 2006, the water umbel responded by extensive lateral contraction. After 2006, when water
269 levels rose again, the water umbel once again began to spread. This effect is similar to that
270 observed for other riparian species (Stromberg et al. 1996, Busch and Smith 1995), including
271 *Populus* spp. (Rood et al. 2003) and *Carex* sedges (Budelsky and Galatowitsch 2000).

272 The second filter we observed was distance from surface water. Mean water umbel leaf
273 density declines rapidly over a very short distance as one moves laterally away from the edge of
274 Leslie Creek. While we also observed that soil moisture declines with increasing distance, our
275 soil moisture measurements and water umbel leaf density weren't strongly correlated. We do not
276 know if this is because there is one or more other limiting factors that we did not measure. There
277 was limited support for the hypothesis that herbaceous competitors or the extent of canopy cover
278 explained significant variation in either the presence/absence or the density of the water umbel.
279 Given the regular flooding of Leslie Creek, which could homogenize soil conditions, systematic
280 variation in other limiting resources such as nutrients seems unlikely. Future work should assay
281 the water umbel's growth across a variety of realistic nutrient concentrations to determine the
282 extent to which nutrients can be a limiting factor. If a strong factor, then this work would need
283 to be followed by detailed soil analyses from field sites where the species' presence, absence,
284 and abundance varies over short distances.

285 Our drought resistance and resilience experiments further refine our understanding of the
286 species' ecology with relation to water. As Titus and Titus (2008a,b) showed in their work,
287 drought patterns are critical to understanding water umbel ecology. Considerable research has
288 examined the physiological and photosynthetic consequences of drought stress (see Chaves et al.

289 2002 and Chaves et al. 2003 for reviews). The pattern of water umbel leaf senescence observed
290 during the experiments follows the results of Munne-Bosch and Alegre (2004), who describe
291 how senescence aids survival in drought conditions. As the water umbel became drought-
292 stressed, peripheral leaves were the first to begin wilting and witnessing chlorophyll degradation;
293 they may have transferred nutrients to the core leaves and delayed their wilting. The high leaf
294 density treatments likely possessed greater nutrient reserves by virtue of additional leaves,
295 providing a possible explanation of their increased drought resistance and resilience.

296 Hydrological conditions in desert wetlands are critical for many plant species' survival (Busch
297 and Smith 1995), and a variety of adaptations protect individuals against the negative impacts of
298 low water levels (Rood et al. 2003). When water umbel grows in conditions that are better-than-
299 marginal—that is, when water (and presumably nutrients) are readily available, and resultant leaf
300 densities are high—individuals are more resistant to drought. Furthermore, the fact that the
301 water umbel plants recovered from drought conditions twice as quickly at high leaf densities than
302 at low leaf densities underscores the importance of strong growth and establishment. We do not
303 know at this time whether the differences in the low- and high-density water umbels used in the
304 transplant experiments had an effect on their medium- or long-term survival.

305 Our qualitative analysis of transplants suggests that water permanence is of greatest
306 importance to Huachuca water umbel re-establishment. Several transplant patches were lost
307 when water from a wetland outflow was suppressed, and growth did not resume when water was
308 eventually re-introduced. This finding is in line with the drought experiment, but contrasts with
309 the results of Titus and Titus (2009), who were able to sprout the water umbel from seeds in the
310 seed bank. It may be that our transplants had not yet fruited by the time the artificial drought
311 occurred, and therefore the species was not part of the seed bank. Patches introduced adjacent to
312 slow water appeared to survive and expand, while patches introduced adjacent to faster-flowing
313 water typically survived for a time but did not expand. These transplant results reinforce the idea
314 that soil must remain saturated for the species to survive and expand, but that the details of the
315 water movement may be important.

316 Although our ecological correlates research did not detect a strong relationship between
317 water umbel presence or density and the presence or abundance of competitors, the transplant
318 experiments suggest a relationship. Patches introduced at lentic sites were consistently out-
319 competed by *Scirpus*, *Typha*, and other herbaceous species. The degree of water umbel exposure

320 to direct sunlight throughout the year also appears to impact survivability. Patches restricted to
321 shaded sites appear to lose viability over time and eventually disappearing, and patches exposed
322 to direct sunlight or mixed shade typically surviving well. Competition among herbaceous-layer
323 waterline species is well documented (see, e.g., Gaudet and Keddy 1995), and is suggested as a
324 potential negative influence on water umbel (USFWS 1997, Zuhlke et al. 2002, Titus and Titus
325 2008a). Thus, despite the lack of statistical evidence of the interaction, our transplant
326 experiments suggest that competition has a real negative effect on water umbel.

327 Previous workers have suggested that monsoon-associated flooding is important to the
328 water umbel because such floods disproportionately affect potential competitors (e.g., Stromberg
329 and Tellman 2009). We have made similar observations, and the fact that transplants protected
330 from seasonal flooding appears to have been out-competed provides further support for the
331 proposed relationship of the water umbel to flooding. However, the transplant experiments
332 indicate that flood events can negatively affect the water umbel. Plants that are otherwise
333 thriving can be buried by aggraded sediment and killed, plants can be directly uprooted and
334 relocated downstream, and severe erosion can remove plants from water saturated sites and
335 ultimately deposit them on drier locations where they desiccate and die. Additionally, we have
336 observed water umbel that was uprooted during a severe flood event and subsequently re-
337 deposited at a suitable location downstream, where it survived to root and establish a new
338 population. Additional work is needed to understand the full diversity of the water umbel's
339 relationship to seasonal flooding that is characteristic across its range.

340

341 **Conclusion**

342 The importance of protecting water sources for conserving and recovering the Huachuca water
343 umbel is well recognized (USFWS 2014). Our research reinforces this conclusion, and suggests
344 that maintaining water levels over extended periods is important because of the role history plays
345 in shaping presence and absence. While small, short-duration events such as trampling by
346 javelina or cattle can have a negative impact on the species, water loss and wide fluctuations
347 likely have a larger impact. In fact, the livestock damage described by Malcom and Radke
348 (2008) was no longer visible along Leslie Creek by late 2008. While competitors may have local
349 negative effects on water umbel, the effects might be managed by transplanting in areas that
350 regularly flood to minimize the establishment of potential competitors (but see below). Because

351 the subsurface (root competition) impacts are difficult to ascertain, the relationship between leaf
352 density and drought resistance and resilience underscores the need to measure water umbel leaf
353 density as a part of monitoring programs because it is predictive of survival in fluctuating
354 environments. Lastly, we propose that using high leaf density plugs for transplants may be
355 preferable to low density plugs because high leaf density plugs exhibit better short-term survival.

356 Numerous ecological research questions surrounding Huachuca water umbel
357 remain. Although we have observed flowering of the water umbel plugs kept in cultivation
358 (during one season only), what are the environmental conditions that trigger flowering? Rare
359 species buffer ecosystems against non-native species invasion (Lyons and Schwartz 2001) and
360 can be important in wetland resilience to disturbance (Englehardt et al. 2001); does the water
361 umbel assist in these respects? In areas where the water umbel occurs, and where stream flows
362 can be managed, is there a specific flood regime that benefits the species (Richter and Richter
363 2000)? To what degree does leaf litter from trees along Leslie Creek and Black Draw affect the
364 soil nutrients available to the water umbel, as Andersen and colleagues (2003) investigated in a
365 general sense? Is nutrient deficiency a significant issue for strong growth and
366 development? Answers to these questions can only help with the recovery of this rare desert
367 wetland species.

368

369 **Acknowledgements**

370 We thank B. Hill for field assistance. *N* anonymous reviewers provided helpful
371 comments on the manuscript. This work does not necessarily reflect the views of Defenders of
372 Wildlife, Virginia State Parks, or the U.S. Fish and Wildlife Service.

373

374

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- 437

438 **Tables**

439 **Table 1. Eight models relating predictor variables to *Lilaeopsis schaffneriana* var. *recurva***
 440 **presence / absence and density using zero-inflated count models.**

Model	Description
1	distance to water + soil moisture + canopy PC1 + canopy PC2 + dist. to centroid
2	soil moisture + canopy PC1 + canopy PC2 + dist. to centroid ¹
3	dist. to water + canopy PC1 + canopy PC2 + dist. to centroid
4	dist. to water + dist. to centroid
5	soil moisture + dist. to centroid
6	soil moisture + canopy PC1 + canopy PC2
7	canopy PC1 + canopy PC2
8	can_45perp + can_135perp + can_45para + can_135para + can_90deg ²
Post-1	soil moisture + dist. to centroid dist. to centroid ³
Post-2	dist. to water + dist. to centroid dist. to centroid ³

441
 442 ¹ Canopy PC1 and PC2 are the first two principal components of the five riparian canopy
 443 measures.
 444 ² “perp” = measurement perpendicular to Leslie Creek; “para” = measurement on the axis
 445 parallel to Leslie Creek; “deg” = measurement 90 degrees (directly over) each patch
 446 ³ Post-hoc models separated the count portion of the zero-inflated model before the vertical bar, |,
 447 and the presence / absence portion after the bar.
 448

449 **Table 2. Model selection results for the eight *a priori* models evaluated, as described in**
 450 **Table 1.**

Model	K	AICc ¹	Δ AICc ²	Mod Lik. ³	AICcWt ⁴	logLikel. ⁵
Mod 4	9	151.46	0.00	1.00	0.64	-64.53
Mod 1	15	153.10	1.64	0.44	0.28	-54.69
Mod 3	13	155.83	4.37	0.11	0.07	-60.00
Mod 5	7	174.37	22.92	0.00	0.00	-78.88
Mod 2	11	179.04	27.58	0.00	0.00	-75.14
Mod 7	7	195.17	43.71	0.00	0.00	-89.28
Mod 6	9	195.35	43.89	0.00	0.00	-86.48
Mod 8	13	206.42	54.97	0.00	0.00	-85.29

451 ¹ AIC_C = Akaike's Information Criterion, adjusted for small samples

452 ² The AIC_C difference from the best (lowest AIC) model

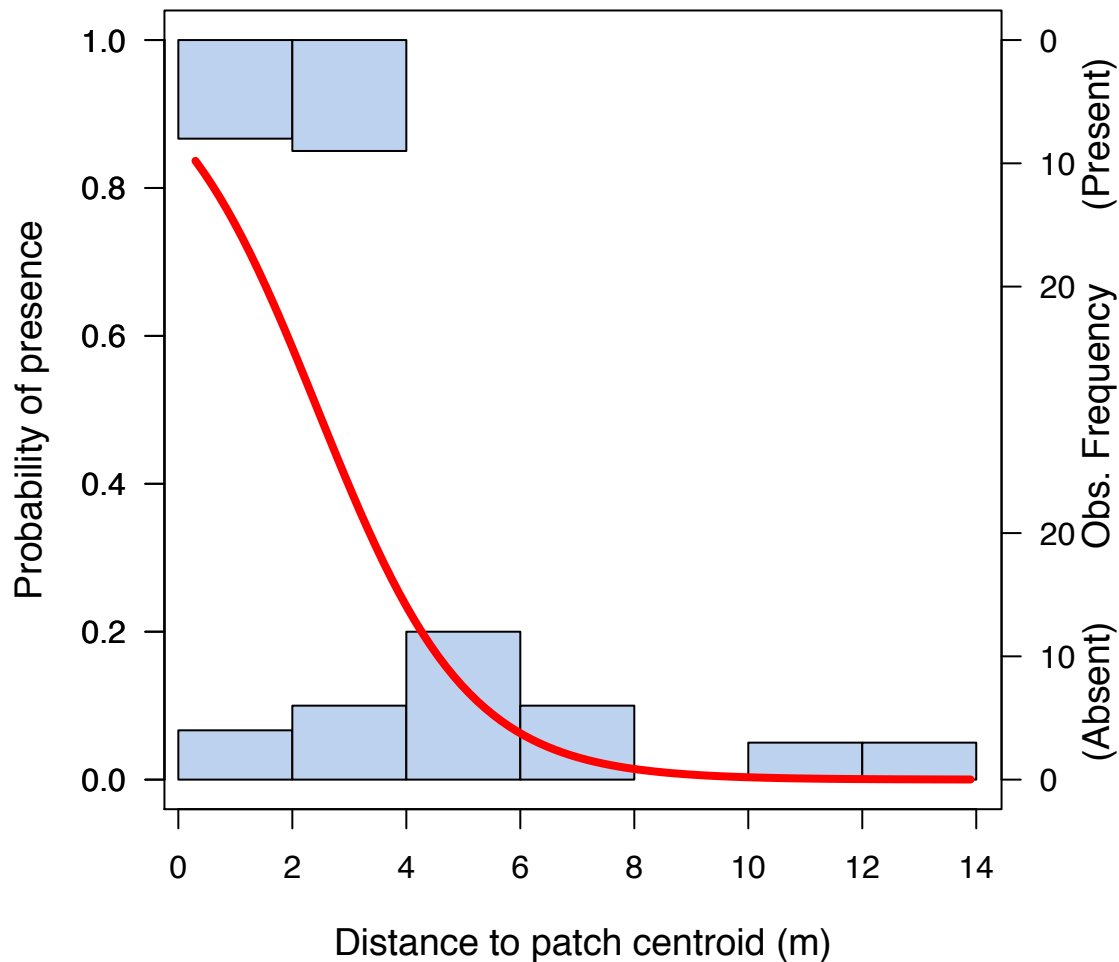
453 ³ Unscaled likelihood of the model

454 ⁴ Scaled model likelihood

455 ⁵ log maximum-likelihood of the model

456

457 Figure 1.



458

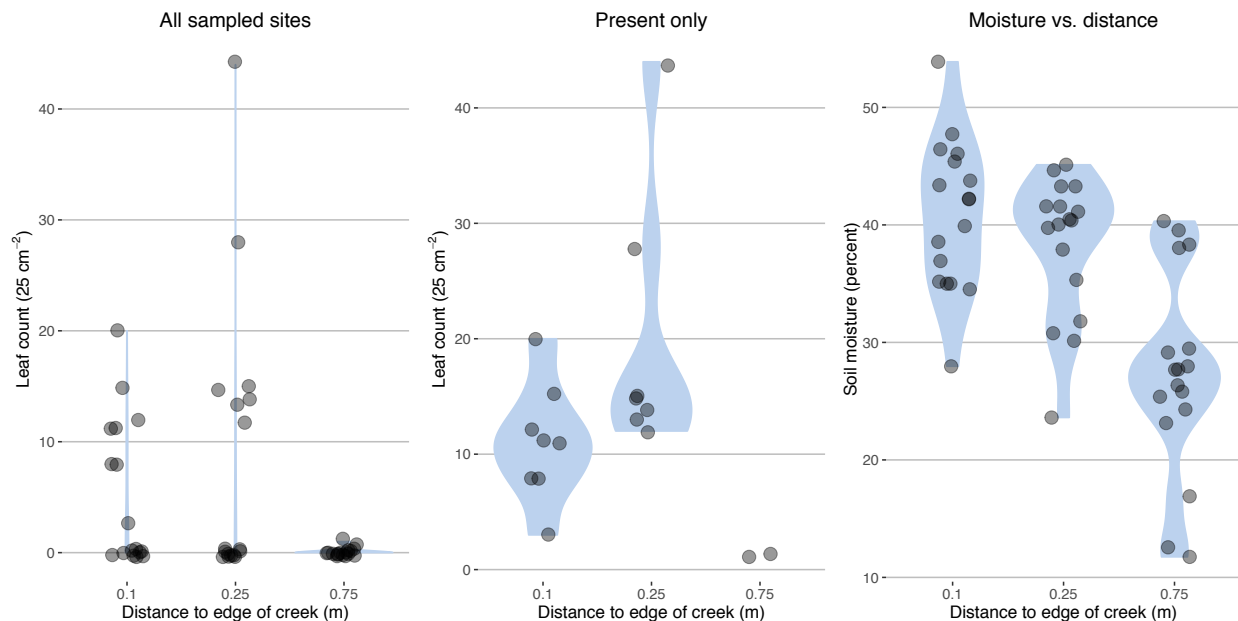
459 **Figure 1. The probability of Huachuca water umbel presence and absence appears to be**
 460 **strongly constrained by history.** The probability of occurrence (red line) drops to 50% by
 461 approximately 3m from the nearest patch centroid, which likely represents transplant locations
 462 from the early-mid 1990s. The histogram bars show the counts of occurrences (top) and
 463 absences (bottom) as a function of distance from the nearest patch centroid.

464

465

466

467 Figure 2.



468

469 **Figure 2. The density of Huachuca water umbel leaves covaries with distance from surface**

470 **water.** The density of Huachuca water umbel leaves peaked 0.25m from the creek (left and

471 center panels) and the distance from the creek corresponds to a strong soil moisture gradient

472 (right panel). The left panel highlights that several sample sites at the water's edge lacked the

473 water umbel, as did almost all sample sites 0.75m from the water. Excluding samples lacking

474 water umbel the relationship to distance from Leslie Creek is a bit clearer. And the left panel

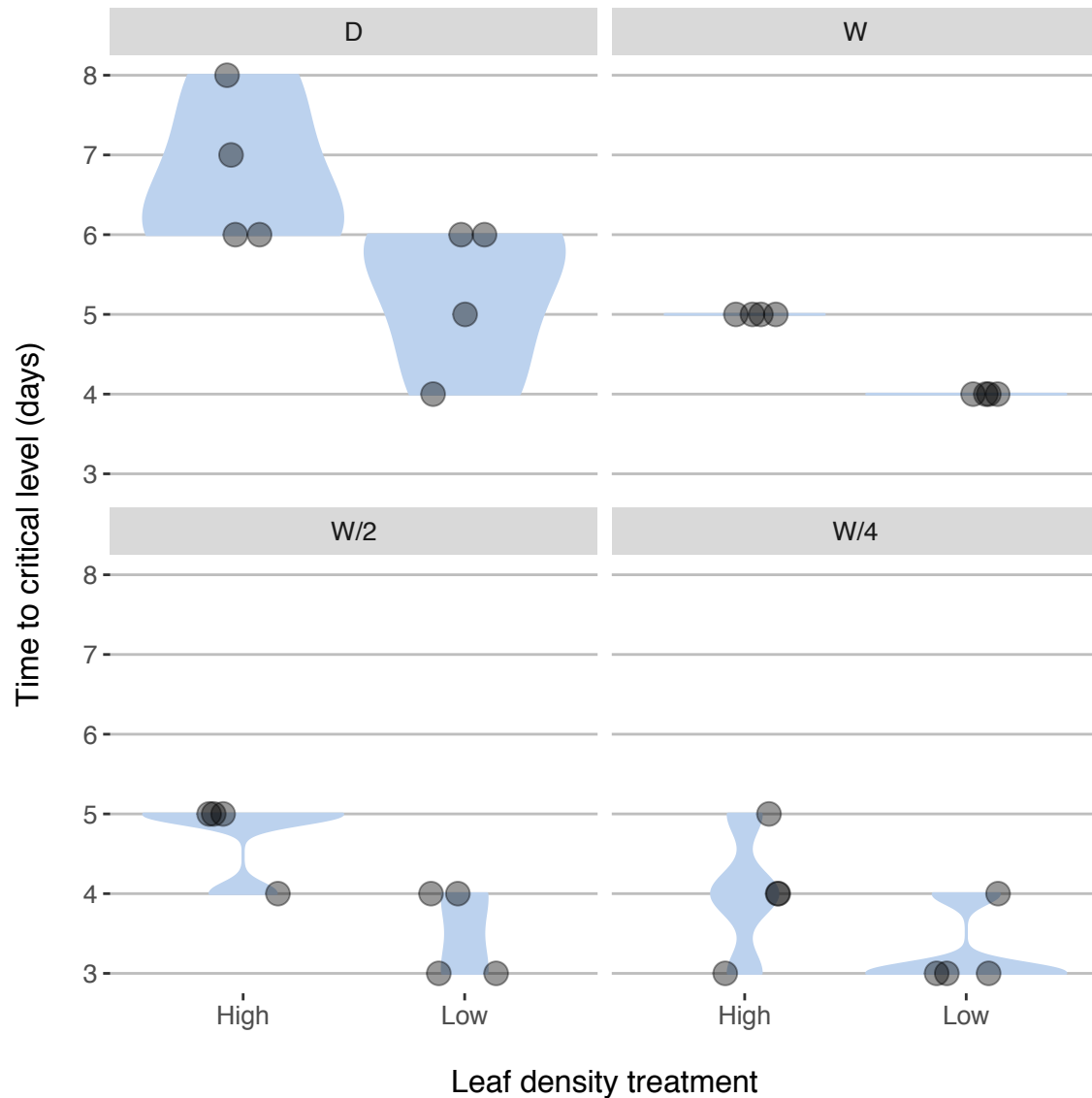
475 shows that soil moisture content covaries strongly with distance from the creek (as expected).

476

477

478

479 Figure 3.



480

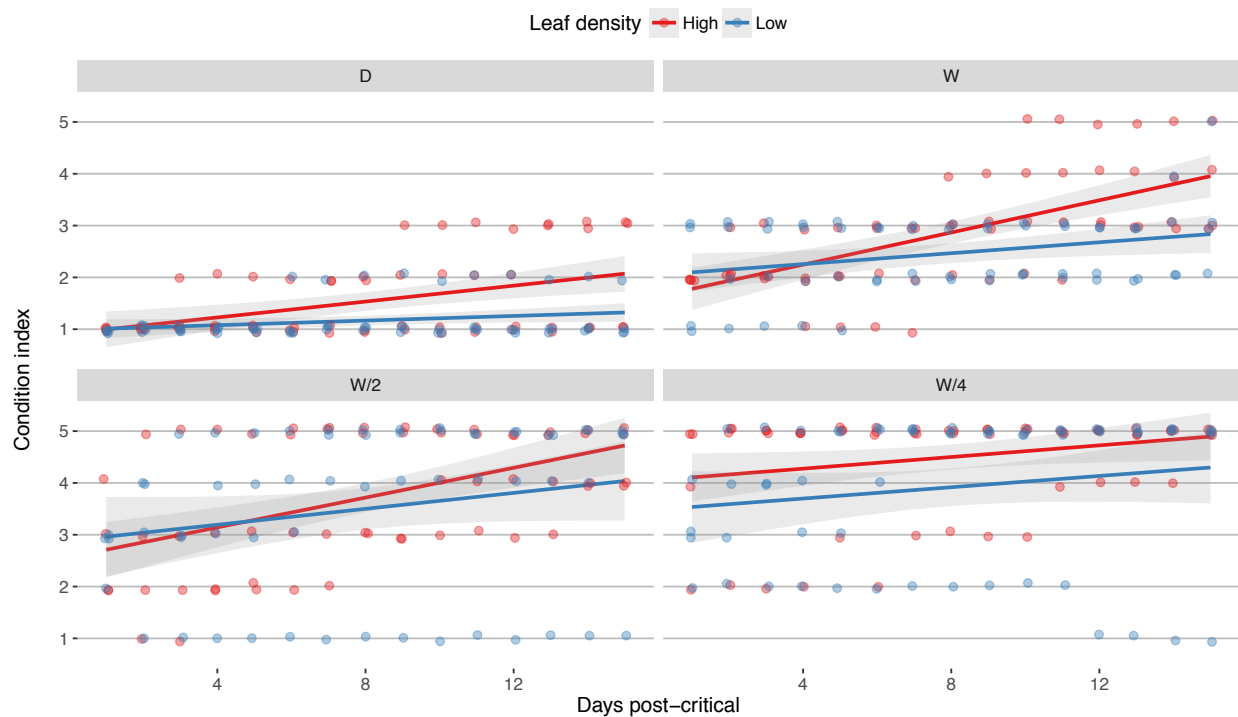
481 **Figure 3. High leaf density confers resistance to drought conditions for Huachuca water**
 482 **umbel.** High density Huachuca water umbel reaches apparent leaf death (D), fully wilted (W),
 483 and $\frac{1}{2}$ -wilted (W/2) conditions significantly later than low-density water umbel, but the time to
 484 $\frac{1}{4}$ -wilted (W/4) is similar.

485

486

487

488 Figure 4.



489

490 **Figure 4. High leaf density confers resilience to drought conditions for Huachuca water**
 491 **umbel.** High density Huachuca water umbel recovers significantly faster from apparent leaf
 492 death (D), fully wilted (W), and $\frac{1}{2}$ -wilted (W/2) conditions low-density water umbel, but the
 493 recovery rate from $\frac{1}{4}$ -wilted (W/4) is similar.