

Predicting the spatial and temporal effects of climate change on the south american lizard genus *Teius* (Squamata: Teiidae)

Ignacio Minoli $^{\text{Corresp.},-1}$, Pier Cacciali 2,3,4 , Mariana Morando 1 , Luciano Javier Avila 1

Corresponding Author: Ignacio Minoli Email address: minoli@cenpat-conicet.gob.ar

Background. The consequences of past or future climate change have been studied in many physical and biological systems, and their effects could change the ecology and spatial distribution of suitable areas for a wide variety of organisms.

Methods. We analyzed the environmental and geographic space of the current suitable area and we projected these conditions into Mid-Holocene and 2050 RCP8.5 scenarios. Through these projections, we assessed and quantified whether climate change would affect the distribution and size of environmental and geographic space for lizard species of the genus *Teius*.

Results. The potentially suitable geographic area for the Mid-Holocene decreased for T. oculatus (-29.55%) and for T. teyou (-6.82%), but for T. suquiensis it was inferred as a larger suitable area (+26%). For the future scenario all species showed a decrease in the potentially suitable area compared to the present (T. oculatus = -9.30%, T. teyou = -0.79%, T. suquiensis = -37.58%). The first 3 PCA axes in the environmental space explained more than 86% of the variation for each temporal scenario for all species. The higher contribution for PC1-2 in Mid-Holocene and Present were mostly related to temperature and for PC3 with altitude variables: and for the 2050 scenario were temperature for PC1, precipitation for PC2 and altitude-temperature for PC3. The hypervolumen distance for T. oculatus from the Present scenario to Mid-Holocene = 67.80 and for 2050 = 309.10; for T. teyou to Mid-Holocene = 95.67 and for 2050 = 442.67; for T. suquiensis to Mid-Holocene = 275.57 and for 2050 = 248.07.

Discussion. These results suggest that current geographic space versus the other temporal forecast of all *Teius* species, showed different specific magnitude changes in their potentially suitable areas. This work illustrates how ectothermic organisms might have to face major changes in their environmental and geographic space as a consequence of the effect of climate changes.

¹ IPEEC-CONICET, Puerto Madryn, Chubut, Argentina

² Senckenberg Forschungsinstitut und Naturmuseum, Frankfurt, Germany

³ Instituto de Investigación Biológica del Paraguay, Asunción, Paraguay

⁴ Asociación Guyra Paraguay, Asunción, Paraguay



- 1 Predicting the spatial and temporal effects of climate change on the south american lizard
- 2 genus *Teius* (Squamata: Teiidae)
- 3 Ignacio Minoli 1*, Pier Cacciali 2,3,4, Mariana Morando 1, Luciano Javier Avila 1
- 4 ¹ IPEEC, CONICET, Boulevard Almirante Brown 2915, Puerto Madryn (Chubut), Argentina.
- ⁵ Senckenberg Forschungsinstitut und Naturmuseum, Senckenberganlage 25, 60325 Frankfurt
- 6 a.M., Germany.
- 7 ³ Instituto de Investigación Biológica del Paraguay, Del Escudo 1607, 1425 Asunción, Paraguay.
- 8 ⁴ Asociación Guyra Paraguay, Av. Cnel. Carlos Bóveda, Parque Asunción Verde, Viñas Cué,
- 9 Paraguay.
- * Corresponding author. E-mail: minoli@cenpat.conicet.gob.ar



ABSTRACT

- 12 **Background.** The consequences of past or future climate change have been studied in many
- physical and biological systems, and their effects could change the ecology and spatial
- 14 distribution of suitable areas for a wide variety of organisms.
- 15 **Methods.** We analyzed the environmental and geographic space of the current suitable area and
- we projected these conditions into Mid-Holocene and 2050 RCP8.5 scenarios. Through these
- 17 projections, we assessed and quantified whether climate change would affect the distribution and
- 18 size of environmental and geographic space for lizard species of the genus *Teius*.
- 19 **Results.** The potentially suitable geographic area for the Mid-Holocene decreased for *T. oculatus*
- 20 (-29.55%) and for *T. teyou* (-6.82%), but for *T. suquiensis* it was inferred as a larger suitable area
- 21 (+26%). For the future scenario all species showed a decrease in the potentially suitable area
- compared to the present (T. oculatus = -9.30%, T. teyou = -0.79%, T. suguiensis = -37.58%). The
- 23 first 3 PCA axes in the environmental space explained more than 86% of the variation for each
- 24 temporal scenario for all species. The higher contribution for PC1-2 in Mid-Holocene and Present
- 25 were mostly related to temperature and for PC3 with altitude variables: and for the 2050 scenario
- were temperature for PC1, precipitation for PC2 and altitude-temperature for PC3. The
- 27 hypervolumen distance for *T. oculatus* from the Present scenario to Mid-Holocene = 67.80 and
- 28 for 2050 = 309.10; for *T. teyou* to Mid-Holocene = 95.67 and for 2050 = 442.67; for *T. suquiensis*
- 29 to Mid-Holocene = 275.57 and for 2050 = 248.07.
- 30 **Discussion.** These results suggest that current geographic space versus the other temporal
- 31 forecast of all *Teius* species, showed different specific magnitude changes in their potentially
- 32 suitable areas. This work illustrates how ectothermic organisms might have to face major changes
- in their environmental and geographic space as a consequence of the effect of climate changes.
- 34 **Keywords**: niche modeling, climate change, Mid-Holocene, 2050 scenario, reptiles, *Teius*.



INTRODUCTION

36	The concept of ecological niche has long been considered to be an important factor in the species'
37	distribution range and was defined as a set of environmental conditions that maintains each taxon
38	within a certain geographic range, modulated by physiological restrictions under which they can
39	prosper (Grinnell, 1917; James et al., 1984). Numerous works studied this concept in recent
40	decades from the theoretical point of view (Soberón & Peterson, 2005; Peterson et al., 2011) and
41	in general they agree that it includes certain components and they are commonly summarized in
42	the BAM diagram. The BAM framework (Soberón & Peterson, 2005) for a species includes a
43	region with a suitable set of abiotic factors (A) and this represents the geographic expression of
44	the fundamental niche (FN), a region with appropriate biotic interactions (B), a region
45	"accessible" to colonization (M) and a region equivalent to the distribution of the species (P)
46	which is the intersection of the three identified regions ($P = A \cap B \cap M$). The niche concept is
47	commonly separated into two major classes: Grinnellian and Eltonian niches. The Grinnellian
48	niche (Grinnell, 1917; Austin, 2002; Soberón, 2007) is defined mainly by abiotic and non-
49	interactive variables, while biotic interactions and resources define the Eltonian niche (Elton,
50	1927; Vandermeer, 1972; Leibold, 1995). Usually, these concepts and theoretical frameworks are
51	not taken into account by researchers that model species' niches (Townsend Peterson & Soberón,
52	2012).
53	In the past decades, a large number of scientists estimated spatial distributions by calculating
54	and quantifying "environmental" or "ecological" niches (Colwell & Rangel, 2009; Soberón &
55	Nakamura, 2009), mostly through mechanistic and correlative models. The analysis of the niche
56	through mechanistic models allows to evaluate a physiological limitation and to project it into the
57	geographic space (Sinervo et al., 2010), while the correlative models perform correlations
58	between several variables and the distribution of the studied organism (Owens et al., 2013). Both



59 types of models have advantages and disadvantages, but correlative models can analyze habitat suitability considering several environmental variables together. Habitat suitability modeling can 60 61 be subjected to spatial analysis such as area calculation (Di Marco et al., 2016), importance of 62 variables in the geographic distribution (Marino et al., 2011) and interaction of the variables with 63 each other (Hirzel & Le Lay, 2008). 64 The study of present habitat suitability is an effective method to infer the spatial interactions 65 and shifts of organisms with past and future climate changes (Peterson et al., 2011). Niche 66 modeling analysis is frequently used in research dealing with species distribution, and is applied 67 to assess geographic ranges within different evolution scenarios (Wiens & Graham, 2005; 68 Warren, Glor & Turelli, 2008); as a tool for biodiversity conservation (Hortal et al., 2015; Jones 69 et al., 2016; Tulloch et al., 2016), or to predict distributions under diverse climate scenarios 70 events (Barrows, 2011; Anderson, 2013). Given the evidence of impact of climate change on 71 biodiversity and on the fate of species' survivorship, there was an eruption of papers focused on this topic, and many scientists made simulation models and predictions based on the climate shift 72 73 scenarios (Cane, 2005; Parker, 2010). Several resources, software and algorithms were produced 74 in the last two decades to perform modeling and simulations (Thuiller et al., 2009; Di Cola et al., 2017) using climatic layers (e.g., Lima-Ribeiro et al., 2015) to assess the effect of climate change 75 76 on species' distribution. 77 The abiotic factors can determine regional and local patterns by which communities of 78 organisms are assembled, and the resulting degree of similarity or difference between localities 79 with similar environments (Chase, 2003). Studies on changes of the potential suitable abiotic area 80 for a species, have been mostly focused on the geographic space and to a lesser extent in the 81 environmental space (Soberón & Nakamura, 2009; Owens et al., 2013). Ectothermic organisms, 82 and particularly reptiles are susceptible to climate changes (Sinervo et al., 2010; Böhm et al., 83 2016). Lizards are an optimal study object to test the climate changes effects, since this group of



84 ectotherm organisms is highly susceptible to climate change (Kubisch et al., 2015) or habitat 85 destruction, because of their reproductive strategies and low dispersal capacity (Zajitschek, 86 Zajitschek & Clobert, 2012). Therefore, the analysis of displacement and contraction of the 87 appropriate climatic-environmental areas for lizard species with little physiological adaptability is 88 a major research goal (Huey et al., 2012; Ceia-Hasse et al., 2014). 89 In this study, we evaluated the potential effects of climate shifts projected for Mid-Holocene 90 and 2050, in the lizard genus *Teius* using ecological niche modeling (ENM, sensu Townsend 91 Peterson & Soberón, 2012), considering the two aspects of Hutchinson's duality: geographic and 92 environmental spaces (sensu, Soberón & Nakamura, 2009). The geographic projection of present 93 modeled distributions coupled with comparisons with ~6000 years ago conditions versus a future 94 medium-term model of 35 years, could provide solid evidence of the niche's shift magnitude in 95 asymmetric time ranges modulated by climate change. The objectives of this study for these taxa 96 were i) to analyze and quantify the current available area with habitat suitability models, and 97 assess potential spatial changes projecting these conditions onto two temporal scenarios: Mid-98 Holocene (~6000 years ago) and the 8.5 Representative Concentration Pathway greenhouse gas 99 concentration (RCP8.5) for 2050; and ii) evaluate changes in the environmental space through n-100 dimensional hypervolume analysis for the three species, the overall genus, and the three 101 scenarios.

MATERIALS AND METHODS

102

103

Studied species and localities data

The genus *Teius* include three species that geographically range from 40° S in Argentina to

Uruguay, Paraguay, and the southeastern region of Brazil and Bolivia at 16°40' S (Cacciali et al.,

2016b,a). These species are relatively large and robust (~144 mm. snout-vent), diurnal,

carnivorous and oviparous (Cei, 1993). Within this genus, *T. oculatus* and *T. teyou* present a wide



108 distribution in different environments in Argentina, Uruguay, Paraguay and Brazil (Cacciali et al., 109 2016a). Teius suquiensis has a restricted distribution in central Argentina that partially overlaps 110 with the other two taxa (Cabrera & Monguillot, 2007). 111 We used 1037 georeferenced presence records (*T. oculatus* = 581; *T. suquiensis* = 52; *T.* 112 teyou = 404) from 675 localities. We obtained this information from the herpetological collection 113 of reptiles (LJAMM-CNP) from Instituto Patagónico para el Estudio de los Ecosistemas 114 Continentales - Centro Nacional Patagónico - CONICET (Puerto Madryn, Chubut, Argentina). 115 Collections acronyms and voucher specimens are cited in Cacciali et al. (2016a). Species 116 sampling, observations and collecting events tend to be clustered around accessible areas and 117 roads. This may overestimate modeling results by including localities that are not spatially 118 independent and do not present environmental variation between them (Dormann et al., 2012). 119 For records with no exact or available geographic locations, we only incorporated them when it 120 was possible to assign approximate coordinates according to the detailed descriptions of locations 121 as they appear in the museums records and literature. 122 Climate and environmental data 123 We modeled habitat suitability for each species through a selection from the 19 bio-climatic 124 variables for current conditions (~1950-2000) and included an altitude variable from the data 125 base WorldClim 1.4 - Global Climate (http://www.worldclim.org). We projected current modeled 126 conditions to BCC-CSM1-1 bioclimatic layers (PMIP2, 2009) for the Mid-Holocene scenario 127 (~6000 years ago) and ACCESS1-0 CMIP5 (2015) layers for modeling 2050 scenarios with 128 RCP8.5 concentrations. The layers for these temporal scenarios were downscaled and calibrated 129 (bias corrected) by their authors through a database using WorldClim 1.4 as baseline 'current' 130 climate (WorldClim - Global Climate Data, 2015). All used layers had a resolution of 30 arc-131 seconds (~1 km²). The IPCC performs simulations of the climate system response to increasing 132 levels of greenhouse gases based on different assumptions about the size of the projected



133 population, technological developments and socio-economic trends. The RCP8.5 scenarios 134 projected relatively large changes based on the most recent observations, which suggested that 135 climate change would be more severe than expected (Meehl et al., 2007; Rahmstorf et al., 2007; 136 Smith et al., 2009). 137 We estimated the accessible area (M; sensu Barve et al., 2011) considering ecological traits, 138 environmental features and potential historical distribution to generate an extent for each taxon. 139 We initially considered the ecoregions as M, but many presence records were located on the edge 140 of these regions, hence we defined M as a minimum convex polygon with an additional spatial 141 buffer of 200 km for each species. Then, we used these extent sizes to clip the environmental 142 layers and generated 5000 random points for each species, and we extracted the pixel values for 143 each bio-climatic variable for both species from these random points and we used them to create 144 a subset of variables. In order to obtain a representation of both types of climatic variables 145 (Barbet-Massin & Jetz, 2014), we divided them in two groups: 1- precipitation and 2-146 temperature. We evaluated the number of variables to use implementing an initial Jackknife test 147 (Pearson et al., 2006). Then, for each variable group we performed a Spearman correlation test 148 (threshold selected > 0.75) to avoid collinearity (Debandi et al., 2012; Dormann et al., 2013; 149 Kershaw et al., 2013) and additionally a selection criterion with biological significance for the 150 studied species (Rissler & Apodaca, 2007; Debandi et al., 2012; Cacciali et al., 2016a) related to 151 the active months period. We performed all correlation analysis with Ggally 1.0.1 R package 152 (Schloerke et al., 2016). 153 **Ecological niche models (ENM)** 154 We generated correlative models of ENM using MaxEnt (3.3.3k version; 2016) algorithm, which 155 seeks maximum entropy density (Merow, Smith & Silander, 2013) through a robust estimation 156 requiring only presence records (Phillips, Anderson & Schapire, 2006; Phillips & Dudík, 2008; Elith et al., 2011) and supports a low number of localities (Pearson et al., 2007; van Proosdij et 157



158 al., 2016). With this software, we calculated the relationship between the current presence of each 159 species and environmental variables, and projected them into the a past and future scenarios. We 160 ran MaxEnt using the selected subset for each species of bio-climatic and altitude variables with 161 the default configurations, except for: 10 replicates, 1000 Maximum iterations, "10 percentile 162 training presence logistic threshold" as selected threshold for each average model. 163 We evaluated the ENMs performance with true skill statistic (TSS) for model's replicates 164 (Allouche, Tsoar & Kadmon, 2006; Lobo, Jiménez-Valverde & Real, 2008) and Area Under the 165 Receiver Operator Curve (further abbreviated to AUC; Phillips, Anderson & Schapire, 2006; Elith et al., 2011) for the averaged models. TSS takes into account both omission and commission 166 167 errors and success as a result of random guessing, hence this evaluation model is not affected by 168 prevalence and also is not affected by the size of the validation set (Allouche, Tsoar & Kadmon, 169 2006). The TSS values ranges from -1 to +1, where +1 indicates perfect agreement and values of 170 zero or less indicate a performance no better than random (Allouche, Tsoar & Kadmon, 2006). 171 We considered these values as TSS model performance indicators: < 0.05 - 0.05 no agreement, 172 0.06 - 0.39 = poor, 0.40 - 0.55 = fair, 0.56 - 0.70 = good, 0.71 - 0.85 = very good, 0.86 - 0.99 = quadrature173 excellent, 0.99 - 1 = perfect (following Monserud & Leemans, 1992). The AUC is a measure of 174 rank-correlation and high AUC values indicates that sites with high predicted suitability values 175 tend to be areas of known presence and locations, whereas a model with lower prediction values 176 tend to be areas where the species is not known to be present (absent or a random point). An AUC 177 score of 0.5 means that the model is as good as a random guess. We considered AUC range 178 values to interpret model performance sensu Araújo et al. (2005): excellent ≥ 0.90 ; good ≥ 0.80 ; 179 fair ≥ 0.70 ; poor ≥ 0.60 ; fail ≤ 0.50 . Finally, we created binary maps based on the selected 180 threshold on the current average models and projected them to onto the Mid-Holocene and 2050 181 RCP8.5 scenarios.

Geographic space, spatial analyses and potential effects on adequate areas



- 183 We used the suitable habitat from the final output models to quantified the area for each species 184 and scenario. From these surfaces, we calculated impact metric to estimate specific distribution 185 changes of predicted suitable areas and the exposure to climate change per species and the overall 186 genus. We analyzed the relative exposure to climate change (Recc) between the current surface 187 models (Cm) versus past (Pm) and future models (Fm) as: Recc = (Pm * 100 / Cm) - 100 and 188 Recc = (Fm * 100 / Cm) - 100 respectively. Thus, high negative relative exposure values indicate 189 that a species decreases the suitable area in that scenario, whereas higher positive values indicate 190 an increase of the suitable area. The estimation and comparison of areas to estimate changes in 191 the different modeling scenarios is a widely implemented method in ENM (e.g., Nori et al., 2013; 192 Nori, Carrasco & Leynaud, 2014; Bonino et al., 2015; Rubio-Salcedo et al., 2016). This might be 193 interpreted as a rigid and conservative estimate of changes from the predicted suitable areas, 194 since we did not considered areas of recolonization after a potential disappearance of suitable 195 environment. Given that there is almost no published literature on the spatial ecology of these 196 species, our analyzes were based on the assumption that none of these lizards would be able to 197 disperse to new potentially suitable areas within the time period of the model. These assumptions 198 were considered due to lack of ecological information on its dispersal ability and adaptability to 199 all the bio-climatic variables considered in the models. These kind of assumptions are commonly 200 associated with forecasting modeling niche techniques (Araújo & Pearson, 2005; Elith & 201 Leathwick, 2009). 202 We performed all spatial analyses and maps with "raster 2.5-2" (Hijmans, 2015b), "maptools 203 0.8-39" (Bivand & Lewin-Koh, 2016), "sp 1.2-2" (Pebesma & Bivand, 2016), "rgdal 1.1-7" 204 (Bivand, Keitt & Rowlingson, 2016), "geosphere 1.5-1" (Hijmans, 2015a), "rgeos 0.3-17" 205 (Bivand & Rundel, 2016) R packages. 206 Environmental space and n - dimensional hypervolumes

We evaluated changes in the environmental space for each of the species considering



209

210

211

212

213

214

215

216

217

218

219

220

221

222

223

224

225

226

227

228

229

230

231

232

visualization results and metrics between the n - dimensional hypervolumes obtained from the geographic suitable outcome area for each of the model's scenarios. The environmental space state is a complex system that involves interactions for more than two variables and can be described as an n - dimensional cloud of points or an n - dimensional hypervolume (Soberón, Osorio-Olvera & Peterson, 2017). If these environmental space conditions are disturbed, it might lead these variables to take other values and to produce a different n - dimensional hypervolume; hence the comparison of two or more hypervolumes could provide an assessment of the magnitude of changes that the ecosystem suffered, such as shifts from the initial state (Blonder et al., 2014). In order to do this, we followed the "framework to study ecosystem stability in face of environmental changes using n - dimensional hypervolumes" (sensu, Barros et al., 2016). Within this framework, we calculated two metrics to made hypervolumes comparisons: 1- the proportion of overlap between pre and post perturbation hypervolumes will reflect overall differences between the two corresponding states, 2- the distance between the centroids of the pre and post perturbation hypervolumes will reflect how much mean values of the ecosystem components have departed from their pre perturbation levels (changes in mean values) and 3- the contribution of the variables to each retained component. To accomplish these analyses, we randomly sampled points (= number of localities * 10) from each of the species' suitable area, and we used these random points to extract the information from the sum of the individual layers employed in MaxEnt for each taxon. Then, all this environmental information for each species was compiled and we performed a Principal Component Analysis (PCA) per scenario, to quantify the ecological multivariate space from the variables that determine the geographic models, and explore how these spaces interacted among the three taxa through time. To improve the visualization of the points, we incorporated gridded ellipses including 95% of the observations for each factor level. We used "The Scree Test" and "Proportion of Variance Accounted For" criteria and components that explain a minimum of



~10% of the variation (O'Rourke & Hatcher, 2013), to decide how many PCA axes we had to consider for this analysis. For these given dimensions, we evaluated the variables' contributions with the function "fviz_contrib". This function calculates any row/column with a contribution above the reference threshold, under which it could be considered as an important variable in contributing to the n - dimensions taken into account (see, Kassambara & Mundt, 2017). We compared the hypervolumes implementing the algorithm "hypervolume_distance" which calculates the Euclidean distance between two hypervolumes centroids (see, Blonder, 2015). We performed all environmental space analyses with "FactoMineR 1.34" (Lê, Josse & Husson, 2008), "factoextra 1.0.4.999" (Kassambara & Mundt, 2017), "hypervolume 1.4.1" (Blonder, 2015), "car 2.1-4" (Fox & Weisberg, 2011), "ade4 1.7-5" (Dray & Dufour, 2007), "rgl 0.96.0" (Adler & Murdoch, 2017) R packages.

RESULTS

ENM, spatial analyses and potential temporal shifts in suitable areas

Considering the bio-climatic correlation analysis outcome, we selected altitude and these environmental variables for each species: *Teius oculatus*: bio1 (annual mean temperature), bio3 (Isothermality = (bio2 / bio7 * 100), bio5 (maximum temperature of warmest month), bio7 (temperature annual range = bio5–bio6), bio9 (mean temperature of driest quarter), bio17 (precipitation of driest quarter), bio18 (precipitation of warmest quarter); *T. suquiensis*: bio2 (mean diurnal range = (mean of monthly (maximum temp – minimum temp)), bio3, bio4 (temperature seasonality = (standard deviation*100)), bio5, bio6 (minimum temperature of coldest month), bio15 (precipitation of seasonality = (coefficient of variation)), bio18; *T. teyou*: bio2, bio3, bio4, bio5, bio15, bio17, bio18. The replicates performance through TSS indicated models fitted above "fair" or "good", with relatively high values for most of the replicates in the three species (Fig. S.1, supporting information). The average test AUC for replicate runs for



258 SD) for *T. teyou*. The suitable climate area in Present scenario was 748,911.98 km² for *Teius oculatus*, 259 260 104,209.16 km² for *T. suquiensis* and 926,690.31 km² for *T. teyou* (Fig. 1). The Recc for *T.* oculatus Mid-Holocene scenario suitable area showed a reduction of 29.5 % and 9.30 % for 2050 261 262 RCP8.5 (Fig. 1); but *T. teyou* Recc presented a moderate decrease in both past and future 263 scenarios with a reduction of 6.82 % and 0.79 % respectively (Table 1, Fig. 1). The Mid-264 Holocene models Recc for T. suquiensis indicated a 26 % higher suitable area than the Present 265 scenario and a reduction of 37.58 % for 2050 RCP8.5 (Table 1, Fig. 1). All species did not show 266 future considerable differences in changes of latitude or longitude (Fig. 1). The relative exposure to climate change (Recc) for different scenarios and suitable area shifts for *Teius* are detailed in 267 268 Table 1. 269 **Environmental space** 270 We observed changes in the environmental space for all the studied species. Principal component 271 analysis indicated that most of the hypervolumes variation could be explained by the first three 272 components for the 3 studied scenarios. The first three axes accounted for all scenarios a variance 273 over 86% of the variation for each temporal scenario (Table 2). The variables that most 274 contributed in Mid-Holocene and Present were BIO6, BIO7 and BIO 4 for PC1; BIO17, BIO1, 275 BIO15 and BIO5 for PC2; and BIO3, BIO5, and altitude for PC3. For the 2050 scenario the 276 variables with higher contribution for PC1 were BIO7, BIO9, BIO4, BIO6; for PC2 BIO, BIO17, 277 BIO15; and altitude, BIO5 for PC3 (Fig. 2). The hypervolumen distance for *T. oculatus* from the Present scenario to the Mid-Holocene = 67.80 and to the 2050 = 309.10; for T. teyou to the Mid-278 279 Holocene = 95.67 and for the 2050 = 442.67; and for *T. suquiensis* to the Mid-Holocene = 275.57280 and to the 2050 = 248.07. These changes in the ellipses for each of the analyzed angles (Fig. 3), 281 were moderate between the Mid-Holocene and the Present scenario. On the other hand, the

Teius oculatus was 0.763 (\pm 0.028 SD), 0.922 (\pm 0.043 SD) for T. suquiensis, and 0.797 (\pm 0.028



ellipses of the Present and 2050 scenarios showed considerable changes between them, especially for *T. teyou* and *T. suquiensis* (Fig. 3).

DISCUSSION

284

285 We found strong support for Grinnellian niche shifts in the past and future scenarios for *Teius* 286 species. The geographic space from an ENM does not exclusively underlie on abiotic variables 287 and, on the contrary there are several factors that determine it (Peterson et al., 2005; Soberón & 288 Peterson, 2005). Despite this, the Grinnellian niche is a good approach to study the relationship 289 between climatic changes and areas suitable for the species' presence without introducing biases 290 of non-climatic aspects (Larson, Olden & Usio, 2010; Chalghaf et al., 2016). Our results, in 291 congruence with other published works related to climate change in lizards (Ballesteros-Barrera, 292 MartÍnez-Meyer & Gadsden, 2007; Kaliontzopoulou et al., 2008; Winck, Almeida-Santos & 293 Rocha, 2014) and amphibians (Lemes, Melo & Loyola, 2014; Zank et al., 2014; Vasconcelos & Do Nascimento, 2016), showed a general tendency toward a higher potentially adequate habitat 294 295 in the past scenario (Fontanella et al., 2012; Sillero & Carretero, 2013; Breitman et al., 2015) and 296 a reduction for the future (Araújo, Thuiller & Pearson, 2006; Bonino et al., 2015; Minoli & Avila, 297 2017). 298 Although the studied genus inhabits environments from a wide geographic space, our 299 findings supported the hypothesis that the suitable area changes as the result of exposure to 300 climate change (Vera-Escalona et al., 2012; Minoli & Avila, 2017). Comparing the spatial 301 forecasts among the studied species, T. teyou and T. oculatus showed less reduction of their 302 potential future area than *T. suquiensis*, which presented high Recc values resulting in a major 303 reduction of the potential future area. In addition, these findings support the theoretical 304 framework which states that each species has its own niche area (Medina, Ponssa & Aráoz, 2016) 305 and magnitude of response to potential changes in their Grinnellian niche (Colwell & Rangel,



307 possible sampling bias (Syfert, Smith & Coomes, 2013), the distribution employed here is the complete currently known for the genus. 308 309 Spatial barriers in reptiles (Sahlean et al., 2014) and potential changes in the distribution of 310 suitable areas and their inter-species overlaps (Barbosa et al., 2012), have been extensively 311 documented and related to hypotheses about hybridization events as a consequence of this 312 phenomenon (Engler et al., 2013; Ahmadzadeh et al., 2013). The most important finding in the 313 geographic context, were the changes in the potential suitable habitat of the *T. suquiensis*, which 314 currently has a broader distribution range. This parthenogenetic species showed significantly 315 changes in both geographically and ecological space with a noticeable tendency towards the 316 reduction of adequate area through temporal scenarios (Fig. 1, Fig. 3, see supplementary files). 317 Our results support that this species showed the greatest decrease of potentially suitable areas for 318 the future, and even the suitable northern sector for the present model would disappear in the 319 future scenario. The unisexual species could arise from hybrid individuals and they are from a 320 recent origen, perhaps within ten thousands of years (Moritz et al., 1992). The unisexual species 321 would be expected to have a shorter persistence than sexual species, and to colonize and occupy 322 ecotone habitats more quickly compared with their sexual relatives (Pough et al., 2015). 323 Moreover, potential contractions for the species's suitable areas should be taken with great 324 attention as there are multiple factors (e.g., anthropic changes, such as oil extraction, cultivated 325 soils, dams, desertification) that could play a decisive role increase the effects to materialize these 326 shifts (Bastos, Araújo & Silva, 2005; Nori et al., 2013). We consider that new field surveys are 327 needed because this is one of the sectors that would be lost in the 2050 scenarios. 328 The limited knowledge of these species' dispersion or fossil records combined with the great 329 fragmentation and degradation of the environments they inhabit (Pelegrin et al., 2009, 2013; 330 Pelegrin & Bucher, 2012), leads us to think that some of the suitable climate spaces predicted in

2009; Soberón & Nakamura, 2009). Although the differences between the results may be due to



332

333

334

335

336

337

338

339

340

341

342

343

344

345

346

347

348

349

350

351

352

353

354

355

our results for temporal scenarios remains uncertain whether they can be occupied or not (Araújo & New, 2007; Colwell & Rangel, 2009). Despite this, past or future impacts of climate change keep a moderate uncertainty level being influenced by interactions and effects of biotic, abiotic, and dispersal considerations (Townsend Peterson, Papes & Soberón, 2016), the robustness of the models presented here were supported by the definition of the segment (i.e., Grinnelian) of the niche studied (Soberón, Osorio-Olvera & Peterson, 2017) and by the use of open algorithms which explains in detail their calculations (Phillips et al., 2017). These constraints of employed models could be mitigated and provide accurate forecasts of the lost-gain of the suitable geographic and climate spaces with a good sample of species presences and correct methodological choices, which increase the certainty of the predictions (Radosavljevic & Anderson, 2014). **CONCLUSIONS** In the last two decades several authors have done review works of ENM concepts (Soberón & Nakamura, 2009; Rangel & Loyola, 2012; Townsend Peterson & Soberón, 2012; McInerny & Etienne, 2012a,b,c), advantages and disadvantages of correlative, mechanistic and processoriented models (Rangel & Loyola, 2012). Moreover, some authors propose blurry boundaries among these approaches (Dormann et al., 2012), and consider that these three approaches do overlap and interconnect (e.g., correlational niche estimates being used in process-oriented approaches). The additional advantages of using correlative models that we have implemented is to analyze a wider range of dimensions that define the niche the study object, but the biggest disadvantage is not being able to evaluate physiological thresholds as mechanistic models (Townsend Peterson, Papes & Soberón, 2016). Until present, there have been numerous publications that used ENM as a tool to predict temporal changes in the distribution of suitable environments in the geographic space and projected them onto another scenarios (Schwartz, 2012; Lemes, Melo & Loyola, 2014; Ruete & Leynaud, 2015; Jones et al., 2016; Allen &



356 Lendemer, 2016). However, the future challenge in these studies should be to perform methods or 357 analyses with very detailed explanations in order to be repeatable in both geographical and 358 ecological space to also obtain quantified results on how and how much the climate change 359 affects species' distribution. 360 The information provided in this work highlights the importance of potential distribution in the geographic and ecological spaces as tools to better understand the processes linked to recent 361 362 species contractions and expansions at the regional and local context. The information provided 363 here also allows adding testable data for the use of some species as proxies for climatic changes 364 in the past and in the future. Moreover, understanding the processes that are behind this kind of 365 phenomenon in such a wide area of South America, is essential to understand biota changes 366 through time in this subcontinent. 367 **ACKNOWLEDGMENTS** 368 We thank C.H.F. Perez, J. Grummer, M. Femenias, J. Villamil Lamas, N. Frutos, N. Feltrin, F. 369 Werneck, A. Camargo, L. Belver, J.C. Acosta, D.R. Perez, R. Martori, F. Palacios, L. Acker for 370 help in field trips. We thank S.I. Tiranti donated some *Teius* samples from NE Argentina. This 371 research did not receive any specific grant from funding agencies in the public, commercial, or 372 not-for-profit sectors. 373 **REFERENCES** Adler D., Murdoch D. 2017.rgl: 3D Visualization Using OpenGL. Available at https://cran.r-374 375 project.org/web/packages/rgl/index.html (accessed February 13, 2017). 376 Ahmadzadeh F., Flecks M., Carretero MA., Mozaffari O., Böhme W., Harris DJ., Freitas S., 377 Rödder D. 2013. Cryptic Speciation Patterns in Iranian Rock Lizards Uncovered by 378 Integrative Taxonomy. PLoS ONE 8:e80563. DOI: 10.1371/journal.pone.0080563.



379	Allen JL., Lendemer JC. 2016. Climate change impacts on endemic, high-elevation lichens in a
380	biodiversity hotspot. Biodiversity and Conservation 25:555-568. DOI: 10.1007/s10531-
381	016-1071-4.
382	Allouche O., Tsoar A., Kadmon R. 2006. Assessing the accuracy of species distribution models:
383	prevalence, kappa and the true skill statistic (TSS). Journal of Applied Ecology 43:1223-
384	1232. DOI: 10.1111/j.1365-2664.2006.01214.x.
385	Anderson RP. 2013. A framework for using niche models to estimate impacts of climate change
386	on species distributions. Annals of the New York Academy of Sciences 1297:8–28. DOI:
387	10.1111/nyas.12264.
388	Araújo MB., New M. 2007. Ensemble forecasting of species distributions. <i>Trends in Ecology &</i>
389	Evolution 22:42–47. DOI: 10.1016/j.tree.2006.09.010.
390	Araújo MB., Pearson RG. 2005. Equilibrium of species' distributions with climate. <i>Ecography</i>
391	28:693–695. DOI: 10.1111/j.2005.0906-7590.04253.x.
392	Araújo MB., Pearson RG., Thuiller W., Erhard M. 2005. Validation of species-climate impact
393	models under climate change. Global Change Biology 11:1504–1513. DOI:
394	10.1111/j.1365-2486.2005.01000.x.
395	Araújo MB., Thuiller W., Pearson RG. 2006. Climate warming and the decline of amphibians and
396	reptiles in Europe. Journal of Biogeography 33:1712–1728. DOI: 10.1111/j.1365-
397	2699.2006.01482.x.
398	Austin MP. 2002. Spatial prediction of species distribution: an interface between ecological
399	theory and statistical modelling. <i>Ecological Modelling</i> 157:101–118. DOI:
400	10.1016/S0304-3800(02)00205-3.
401	Ballesteros-Barrera C., MartÍnez-Meyer E., Gadsden H. 2007. Effects of land-cover
402	transformation and climate change on the distribution of two microendemic lizards, genus
403	Uma, of northern mexico. Journal of Herpetology 41:733-740. DOI: 10.1670/06-276.1.
404	Barbet-Massin M., Jetz W. 2014. A 40-year, continent-wide, multispecies assessment of relevant
405	climate predictors for species distribution modelling. Diversity and Distributions
406	20:1285–1295. DOI: 10.1111/ddi.12229.



407	Barbosa AM., Sillero N., Martínez-Freiría F., Real R. 2012. Ecological niche models in						
408	Mediterranean herpetology: past, present, and future. In: Zhang W-J ed. Ecological						
409	Modeling. Hauppauge, N.Y: Nova Science Publishers, 173-204.						
410	Barros C., Thuiller W., Georges D., Boulangeat I., Münkemüller T. 2016. N-dimensional						
411	hypervolumes to study stability of complex ecosystems. <i>Ecology letters</i> 19:729–742.						
412	DOI: 10.1111/ele.12617.						
413	Barrows CW. 2011. Sensitivity to climate change for two reptiles at the Mojave–Sonoran Desert						
414	interface. Journal of Arid Environments 75:629-635. DOI:						
415	10.1016/j.jaridenv.2011.01.018.						
416	Barve N., Barve V., Jiménez-Valverde A., Lira-Noriega A., Maher SP., Peterson AT., Soberón J.,						
417	Villalobos F. 2011. The crucial role of the accessible area in ecological niche modeling						
418	and species distribution modeling. Ecological Modelling 222:1810–1819. DOI:						
419	10.1016/j.ecolmodel.2011.02.011.						
420	Bastos EG de M., Araújo AFB de., Silva HR da. 2005. Records of the rattlesnakes Crotalus						
421	durissus terrificus (Laurenti) (Serpentes, Viperidae) in the State of Rio de Janeiro, Brazil:						
422	a possible case of invasion facilitated by deforestation. Revista Brasileira de Zoologia						
423	22:812–815. DOI: 10.1590/S0101-81752005000300047.						
424	Bivand R., Keitt T., Rowlingson B. 2016.rgdal: bindings for the geospatial data abstraction						
425	library. R package version 1.1-3. Available at https://CRAN.R-project.org/package=rgdal						
426	(accessed March 28, 2016).						
427	Bivand R., Lewin-Koh N. 2016.maptools: tools for reading and handling spatial objects. R						
428	package version 0.8-39. Available at https://CRAN.R-project.org/package=maptools						
129	(accessed March 28, 2016).						
430	Bivand R., Rundel C. 2016.rgeos: interface to geometry engine - open source (GEOS). R package						
431	version 0.3-17. Available at https://CRAN.R-project.org/package=rgeos (accessed March						
432	28, 2016).						
433	Blonder B. 2015.hypervolume: High-Dimensional Kernel Density Estimation and Geometry						
434	Operations. Available at https://cran.r-project.org/web/packages/hypervolume/index.html						

433	(accessed March 9, 2017).						
436	Blonder B., Lamanna C., Violle C., Enquist BJ. 2014. The n-dimensional hypervolume. <i>Global</i>						
437	Ecology and Biogeography 23:595–609. DOI: 10.1111/geb.12146.						
438	Böhm M., Cook D., Ma H., Davidson AD., García A., Tapley B., Pearce-Kelly P., Carr J. 2016.						
439	Hot and bothered: Using trait-based approaches to assess climate change vulnerability in						
440	reptiles. Biological Conservation 204:32–41. DOI: 10.1016/j.biocon.2016.06.002.						
441	Bonino MF., Azócar DLM., Ii JAS., Cruz FB. 2015. Climate change and lizards: changing						
442	species' geographic ranges in Patagonia. Regional Environmental Change 15:1121–1132.						
443	DOI: 10.1007/s10113-014-0693-x.						
444	Breitman MF., Bonino MF., Sites JW., Avila LJ., Morando M. 2015. Morphological variation,						
445	niche divergence, and phylogeography of lizards of the Liolaemus lineomaculatus section						
446	(Liolaemini) from southern Patagonia. Herpetological Monographs 29:65-88. DOI:						
447	10.1655/HERPMONOGRAPHS-D-14-00003.						
448	Cabrera MR., Monguillot JC. 2007. Reptilia, Squamata, Teiidae, Teius suquiensis: new evidence						
449	of recent expansion of this parthenogenetic lizard? Check List 3:180–184. DOI:						
450	10.15560/3.3.180.						
451	Cacciali P., Morando M., Köhler G., Avila L. 2016a. On the distribution of the genus Teius						
452	Merrem, 1820 (Reptilia: Squamata: Teiidae). Zootaxa 4136:491–514. DOI:						
453	10.11646/zootaxa.4136.3.4.						
454	Cacciali P., Scott NJ., Aquino Ortíz AL., Fitzgerald LA., Smith P. 2016b. The reptiles of						
455	Paraguay: literature, distribution, and an annotated taxonomic checklist. Special						
456	Publication of the Museum of Southwestern Biology: 1–373.						
457	Cane MA. 2005. The evolution of El Niño, past and future. Earth and Planetary Science Letters						
458	230:227–240. DOI: 10.1016/j.epsl.2004.12.003.						
459	Cei JM. 1993. Reptiles del nordeste y este de la Argentina. Herpetofauna de las selvas						
460	subtropicales, Puna y Pampas. Museo Regionale di Scienze Naturali Torino.						
461	Ceia-Hasse A., Sinervo B., Vicente L., Pereira HM. 2014. Integrating ecophysiological models						
462	into species distribution projections of European reptile range shifts in response to climate						

change. *Ecography* 37:1–10. DOI: 10.1111/j.1600-0587.2013.00600.x. 463 464 Chalghaf B., Chlif S., Mayala B., Ghawar W., Bettaieb J., Harrabi M., Benie GB., Michael E., 465 Salah AB. 2016. Ecological niche modeling for the prediction of the geographic 466 distribution of cutaneous leishmaniasis in Tunisia. The American Journal of Tropical 467 *Medicine and Hygiene* 94:844–851. DOI: 10.4269/ajtmh.15-0345. Chase JM. 2003. Community assembly: when should history matter? *Oecologia* 136:489–498. 468 469 DOI: 10.1007/s00442-003-1311-7. 470 CMIP5 2015.CMIP5 Coupled Model Intercomparison Project. Available at http://cmip-471 pcmdi.llnl.gov/cmip5/ (accessed October 23, 2015). 472 Colwell RK., Rangel TF. 2009. Hutchinson's duality: the once and future niche. *Proceedings of* 473 the National Academy of Sciences 106:19651–19658. DOI: 10.1073/pnas.0901650106. 474 Debandi G., Corbalán V., Scolaro JA., Roig-Juñent SA. 2012. Predicting the environmental niche 475 of the genus *Phymaturus*: Are *palluma* and *patagonicus* groups ecologically 476 differentiated? Austral Ecology 37:392–400. DOI: 10.1111/j.1442-9993.2011.02295.x. 477 Di Cola V., Broennimann O., Petitpierre B., Breiner FT., D'Amen M., Randin C., Engler R., Pottier J., Pio D., Dubuis A., Pellissier L., Mateo RG., Hordijk W., Salamin N., Guisan A. 478 479 2017. ecospat: an R package to support spatial analyses and modeling of species niches 480 and distributions. *Ecography* 40:1–14. DOI: 10.1111/ecog.02671. 481 Di Marco M., Watson JEM., Possingham HP., Venter O. 2016. Limitations and trade-offs in the 482 use of species distribution maps for protected area planning. Journal of Applied Ecology 483 54:402–411. DOI: 10.1111/1365-2664.12771. 484 Dormann CF., Elith J., Bacher S., Buchmann C., Carl G., Carré G., Marquéz JRG., Gruber B., Lafourcade B., Leitão PJ., Münkemüller T., McClean C., Osborne PE., Reineking B., 485 486 Schröder B., Skidmore AK., Zurell D., Lautenbach S. 2013. Collinearity: a review of 487 methods to deal with it and a simulation study evaluating their performance. *Ecography* 488 36:27–46. DOI: 10.1111/j.1600-0587.2012.07348.x. 489 Dormann CF., Schymanski SJ., Cabral J., Chuine I., Graham C., Hartig F., Kearney M., Morin 490 X., Römermann C., Schröder B., Singer A. 2012. Correlation and process in species



491 distribution models: bridging a dichotomy. Journal of Biogeography 39:2119–2131. DOI: 492 10.1111/j.1365-2699.2011.02659.x. 493 Dray S., Dufour AB. 2007. The ade4 package: implementing the duality diagram for ecologists. 494 *Journal of Statistical Software* 22:1–20. 495 Elith J., Leathwick JR. 2009. Species distribution models: ecological explanation and prediction 496 across space and time. Annual Review of Ecology, Evolution, and Systematics 40:677– 497 697. DOI: 10.1146/annurev.ecolsys.110308.120159. 498 Elith J., Phillips SJ., Hastie T., Dudík M., Chee YE., Yates CJ. 2011. A statistical explanation of 499 MaxEnt for ecologists. Diversity and Distributions 17:43–57. DOI: 10.1111/j.1472-500 4642.2010.00725.x. 501 Elton CS. 1927. Animal ecology. New York,: Macmillan Co.,. 502 Engler JO., Rödder D., Elle O., Hochkirch A., Secondi J. 2013. Species distribution models 503 contribute to determine the effect of climate and interspecific interactions in moving 504 hybrid zones. Journal of Evolutionary Biology 26:2487–2496. DOI: 10.1111/jeb.12244. 505 Fontanella FM., Feltrin N., Avila LJ., Sites JW., Morando M. 2012. Early stages of divergence: 506 phylogeography, climate modeling, and morphological differentiation in the South 507 American lizard Liolaemus petrophilus (Squamata: Liolaemidae). Ecology and Evolution 508 2:792-808. DOI: 10.1002/ece3.78. 509 Fox J., Weisberg S. 2011. An R Companion to Applied Regression. Thousand Oaks CA: Sage. 510 Grinnell J. 1917. The niche-relationships of the california thrasher. *The Auk* 34:427–433. DOI: 511 10.2307/4072271. 512 Hijmans RJ. 2015a.geosphere: spherical trigonometry. R package version 1.5-1. Available at 513 https://CRAN.R-project.org/package=geosphere (accessed March 28, 2016). 514 Hijmans RJ. 2015b.raster: geographic data analysis and modeling. R package version 2.5-2. 515 Available at https://CRAN.R-project.org/package=raster (accessed March 28, 2016). 516 Hirzel AH., Le Lay G. 2008. Habitat suitability modelling and niche theory. Journal of Applied 517 Ecology 45:1372–1381. DOI: 10.1111/j.1365-2664.2008.01524.x. Hortal J., Bello F de., Diniz-Filho JAF., Lewinsohn TM., Lobo JM., Ladle RJ. 2015. Seven 518



519	snortialis that beset large-scale knowledge of blodiversity. Annual Review of Ecology,
520	Evolution, and Systematics 46:523–549. DOI: 10.1146/annurev-ecolsys-112414-054400.
521	Huey RB., Kearney MR., Krockenberger A., Holtum JAM., Jess M., Williams SE. 2012.
522	Predicting organismal vulnerability to climate warming: roles of behaviour, physiology
523	and adaptation. Philosophical Transactions of the Royal Society B: Biological Sciences
524	367:1665–1679. DOI: 10.1098/rstb.2012.0005.
525	James FC., Johnston RF., Wamer NO., Niemi GJ., Boecklen WJ. 1984. The grinnellian niche of
526	the wood thrush. <i>The American Naturalist</i> 124:17–47.
527	Jones KR., Watson JEM., Possingham HP., Klein CJ. 2016. Incorporating climate change into
528	spatial conservation prioritisation: A review. Biological Conservation 194:121–130. DOI:
529	10.1016/j.biocon.2015.12.008.
530	Kaliontzopoulou A., Brito JC., Carretero MA., Larbes S., Harris DJ. 2008. Modelling the
531	partially unknown distribution of wall lizards (Podarcis) in North Africa: ecological
532	affinities, potential areas of occurrence, and methodological constraints. Canadian
533	Journal of Zoology 86:992–1001. DOI: 10.1139/Z08-078.
534	Kassambara A., Mundt F. 2017.factoextra: Extract and Visualize the Results of Multivariate Data
535	Analyses. R package version 1.0.4.999. Available at
536	http://www.sthda.com/english/rpkgs/factoextra (accessed February 20, 2017).
537	Kershaw F., Waller T., Micucci P., Draque J., Barros M., Buongermini E., Pearson RG., Mendez
538	M. 2013. Informing conservation units: barriers to dispersal for the yellow anaconda.
539	Diversity and Distributions 19:1164–1174. DOI: 10.1111/ddi.12101.
540	Kubisch EL., Corbalán V., Ibargüengoytía NR., Sinervo B. 2015. Local extinction risk by global
541	warming of three species of lizards from Patagonia. Canadian Journal of Zoology. DOI:
542	10.1139/cjz-2015-0024.
543	Larson ER., Olden JD., Usio N. 2010. Decoupled conservatism of Grinnellian and Eltonian
544	niches in an invasive arthropod. <i>Ecosphere</i> 1:1–13. DOI: 10.1890/ES10-00053.1.
545	Lê S., Josse J., Husson F. 2008. FactoMineR: An R package for multivariate analysis. <i>Journal of</i>
546	statistical software 25:1–18.



547 Leibold MA. 1995. The niche concept revisited: mechanistic models and community context. 548 Ecology 76:1371–1382. DOI: 10.2307/1938141. 549 Lemes P., Melo AS., Loyola RD. 2014. Climate change threatens protected areas of the Atlantic 550 Forest. Biodiversity and Conservation 23:357–368. DOI: 10.1007/s10531-013-0605-2. 551 Lima-Ribeiro MS., Varela S., González-Hernández J., Oliveira G de., Diniz-Filho JAF., Terribile 552 LC. 2015. EcoClimate: a database of climate data from multiple models for past, present, 553 and future for macroecologists and biogeographers. *Biodiversity Informatics* 10:1–21. 554 DOI: 10.17161/bi.v10i0.4955. 555 Lobo JM., Jiménez-Valverde A., Real R. 2008. AUC: a misleading measure of the performance of 556 predictive distribution models. *Global Ecology and Biogeography* 17:145–151. DOI: 557 10.1111/j.1466-8238.2007.00358.x. 558 Marino J., Bennett M., Cossios D., Iriarte A., Lucherini M., Pliscoff P., Sillero-Zubiri C., Villalba 559 L., Walker S. 2011. Bioclimatic constraints to Andean cat distribution: a modelling 560 application for rare species. *Diversity and Distributions* 17:311–322. DOI: 561 10.1111/j.1472-4642.2011.00744.x. McInerny GJ., Etienne RS. 2012a. Stitch the niche - a practical philosophy and visual schematic 562 563 for the niche concept. Journal of Biogeography 39:2103–2111. DOI: 10.1111/jbi.12032. 564 McInerny GJ., Etienne RS. 2012b. Pitch the niche - taking responsibility for the concepts we use 565 in ecology and species distribution modelling. Journal of Biogeography 39:2112–2118. 566 DOI: 10.1111/jbi.12031. 567 McInerny GJ., Etienne RS. 2012c. Ditch the niche - is the niche a useful concept in ecology or 568 species distribution modelling? *Journal of Biogeography* 39:2096–2102. DOI: 569 10.1111/jbi.12033. 570 Medina RG., Ponssa ML., Aráoz E. 2016. Environmental, land cover and land use constraints on 571 the distributional patterns of anurans: *Leptodacylus* species (Anura, Leptodactylidae) 572 from Dry Chaco. *PeerJ* 4:e2605. DOI: 10.7717/peerj.2605. 573 Meehl GA., Stocker TF., Collins WD., Friedlingstein P., Gaye AT., Gregory JM., Kitoh A., Knutti 574 R., Murphy JM., Noda A., Raper SCB., Watterson IG., Weaver AJ. 2007. Regional



3/3	Climate Projections. In: Solomon S, Qin D, Manning M, Chen Z, Marquis M, Averyt KB,
576	Tignor M, Miller HL eds. Climate Change 2007: The Physical Science Basis.
577	Contribution of Working Group I to the Fourth Assessment Report of the
578	Intergovernmental Panel on Climate Change. Cambridge, United Kingdom and New
579	York, NY, USA: Cambridge University Press, 748–846.
580	Merow C., Smith MJ., Silander JA. 2013. A practical guide to MaxEnt for modeling species'
581	distributions: what it does, and why inputs and settings matter. Ecography 36:1058–1069.
582	DOI: 10.1111/j.1600-0587.2013.07872.x.
583	Minoli I., Avila LJ. 2017. Conservation assessments in climate change scenarios: spatial
584	perspectives for present and future in two Pristidactylus (Squamata: Leiosauridae) lizards
585	from Argentina. Zootaxa 4237:91–111. DOI: 10.11646/zootaxa.4237.1.5.
586	Monserud RA., Leemans R. 1992. Comparing global vegetation maps with the Kappa statistic.
587	Ecological Modelling 62:275–293. DOI: 10.1016/0304-3800(92)90003-W.
588	Moritz C., Uzzell T., Spolsky C., Hotz H., Darevsky I., Kupriyanova L., Danielyan F. 1992. The
589	material ancestry and approximate age of parthenogenetic species of Caucasian rock
590	lizards (Lacerta: Lacertidae). <i>Genetica</i> 87:53–62. DOI: 10.1007/BF00128773.
591	Nori J., Carrasco PA., Leynaud GC. 2014. Venomous snakes and climate change: ophidism as a
592	dynamic problem. Climatic Change 122:67-80. DOI: 10.1007/s10584-013-1019-6.
593	Nori J., Lescano JN., Illoldi-Rangel P., Frutos N., Cabrera MR., Leynaud GC. 2013. The conflict
594	between agricultural expansion and priority conservation areas: Making the right
595	decisions before it is too late. <i>Biological Conservation</i> 159:507–513. DOI:
596	10.1016/j.biocon.2012.11.020.
597	O'Rourke N., Hatcher L. 2013. A step-by-step approach to using SAS for Factor Analysis and
598	Structural Equation Modeling. Cary, North Carolina, USA: SAS Institute.
599	Owens HL., Campbell LP., Dornak LL., Saupe EE., Barve N., Soberón J., Ingenloff K., Lira-
600	Noriega A., Hensz CM., Myers CE., Peterson AT. 2013. Constraints on interpretation of
601	ecological niche models by limited environmental ranges on calibration areas. Ecological
602	Modelling 263:10–18. DOI: 10.1016/j.ecolmodel.2013.04.011.



003	Parker w.S. 2010. Predicting weather and climate: Uncertainty, ensembles and probability.
604	Studies in History and Philosophy of Science Part B: Studies in History and Philosophy
605	of Modern Physics 41:263–272. DOI: 10.1016/j.shpsb.2010.07.006.
606	Pearson RG., Raxworthy CJ., Nakamura M., Townsend Peterson A. 2007. Predicting species
607	distributions from small numbers of occurrence records: a test case using cryptic geckos
608	in Madagascar. Journal of Biogeography 34:102–117. DOI: 10.1111/j.1365-
609	2699.2006.01594.x.
610	Pearson RG., Thuiller W., Araújo MB., Martinez-Meyer E., Brotons L., McClean C., Miles L.,
611	Segurado P., Dawson TP., Lees DC. 2006. Model-based uncertainty in species range
612	prediction. Journal of Biogeography 33:1704-1711. DOI: 10.1111/j.1365-
613	2699.2006.01460.x.
614	Pebesma E., Bivand R. 2016.sp: classes and methods for spatial data in R. Available at
615	http://cran.r-project.org/doc/Rnews/ (accessed March 28, 2016).
616	Pelegrin N., Bucher EH. 2012. Effects of habitat degradation on the lizard assemblage in the Arid
617	Chaco, central Argentina. Journal of Arid Environments 79:13-19. DOI:
618	10.1016/j.jaridenv.2011.11.004.
619	Pelegrin N., Chani JM., Echevarría AL., Bucher EH. 2009. Effects of forest degradation on
620	abundance and microhabitat selection by ground dwelling Chaco lizards. Amphibia-
621	Reptilia 30:265–271.
622	Pelegrin N., Chani JM., Echevarria AL., Bucher EH. 2013. Habitat degradation may affect niche
623	segregation patterns in lizards. Acta Oecologica 51:82-87. DOI:
624	10.1016/j.actao.2013.06.006.
625	Peterson AT., Soberón J., Pearson RG., Anderson RP., Martínez-Meyer E., Nakamura M., Araújo
626	MB. 2011. Ecological niches and geographic distributions. United Kingdom: Princeton
627	University Press.
628	Peterson AT., Tian H., Martínez-Meyer E., Soberón J., Sánchez-Cordero V., Weaver AJ. 2005.
629	Modeling distributional shifts of individual species and biomes. In: Lovejoy TE, Hannah
630	L eds. Climate Change and Biodiversity. New Halen & London, UK: Yale University

631 Press, 211–231. 632 Phillips SJ., Anderson RP., Dudlk M., Schapire RE., Blair ME. 2017. Opening the black box: an 633 open-source release of Maxent. Ecography 40:887–893. DOI: 10.1111/ecog.03049. 634 Phillips SJ., Anderson RP., Schapire RE. 2006. Maximum entropy modeling of species 635 geographic distributions. *Ecological Modelling* 190:231–259. DOI: 636 10.1016/j.ecolmodel.2005.03.026. 637 Phillips SJ., Dudík M. 2008. Modeling of species distributions with Maxent: new extensions and 638 a comprehensive evaluation. Ecography 31:161–175. DOI: 10.1111/j.0906-639 7590.2008.5203.x. 640 Phillips SJ., Dudík M., Schapire RE. 2016. Maxent software and datasets. 3.3.3k version. 641 Available at https://www.cs.princeton.edu/~schapire/maxent/ (accessed March 10, 2016). 642 PMIP2 2009 Paleoclimate Modelling Intercomparison Project Phase II. Available at 643 https://pmip2.lsce.ipsl.fr/ (accessed February 8, 2017). 644 Pough FH., Andrews RM., Crump ML., Savitzky AH., Wells KD., Brandley MC. 2015. 645 Herpetology. Sunderland, Massachusetts, USA: Sinauer Associates is an imprint of 646 Oxford University Press. 647 van Proosdij ASJ., Sosef MSM., Wieringa JJ., Raes N. 2016. Minimum required number of 648 specimen records to develop accurate species distribution models. Ecography 39:542– 649 552. DOI: 10.1111/ecog.01509. 650 Radosavljevic A., Anderson RP. 2014. Making better Maxent models of species distributions: 651 complexity, overfitting and evaluation. *Journal of Biogeography* 41:629–643. DOI: 652 10.1111/jbi.12227. Rahmstorf S., Cazenave A., Church JA., Hansen JE., Keeling RF., Parker DE., Somerville RCJ. 653 654 2007. Recent climate observations compared to projections. *Science* 316:709. DOI: 655 10.1126/science.1136843. 656 Rangel TF., Loyola RD. 2012. Labeling Ecological Niche Models. *Natureza & Conservação* 657 10:119–126. DOI: 10.4322/natcon.2012.030. 658 Rissler LJ., Apodaca J. 2007. Adding more ecology into species delimitation: ecological niche



039	models and phylogeography nelp define cryptic species in the black salamander (Aneldes
660	flavipunctatus). Systematic Biology 56:924–942. DOI: 10.1080/10635150701703063.
661	Rubio-Salcedo M., Psomas A., Prieto M., Zimmermann NE., Martínez I. 2016. Case study of the
662	implications of climate change for lichen diversity and distributions. Biodiversity and
663	Conservation:1–21. DOI: 10.1007/s10531-016-1289-1.
664	Ruete A., Leynaud GC. 2015. Identification of limiting climatic and geographical variables for
665	the distribution of the tortoise Chelonoidis chilensis (Testudinidae): a baseline for
666	conservation actions. PeerJ 3:e1298. DOI: 10.7717/peerj.1298.
667	Sahlean TC., Gherghel I., Papeş M., Strugariu A., Zamfirescu ŞR. 2014. Refining climate change
668	projections for organisms with low dispersal abilities: a case study of the caspian whip
669	snake. PLOS ONE 9:e91994. DOI: 10.1371/journal.pone.0091994.
670	Schloerke B., Crowley J., Cook D., Briatte F., Marbach M., Thoen E., Elberg A. 2016.GGally:
671	extension to ggplot2. Available at https://cran.r-
672	project.org/web/packages/GGally/index.html (accessed April 19, 2016).
673	Schwartz MW. 2012. Using niche models with climate projections to inform conservation
674	management decisions. Biological Conservation 155:149-156. DOI:
675	10.1016/j.biocon.2012.06.011.
676	Sillero N., Carretero MA. 2013. Modelling the past and future distribution of contracting species.
677	The iberian lizard Podarcis carbonelli (Squamata: Lacertidae) as a case study.
678	Zoologischer Anzeiger - A Journal of Comparative Zoology 252:289–298. DOI:
679	10.1016/j.jcz.2012.08.004.
680	Sinervo B., Méndez de la Cruz F., Miles DB., Heulin B., Bastiaans E., Villagrán Santa Cruz M.,
681	Lara Resendiz R., Martinez Méndez N., Calderón Espinosa ML., Meza Lazaro RN.,
682	Gadsden H., Avila LJ., Morando M., De la Riva IJ., Sepúlveda PV., Rocha CFD.,
683	Ibarguengoytía N., Puntriano CA., Massot M., Lepetz V., Oksanen TA., Chapple DG.,
684	Bauer AM., Branch WR., Clobert J., Sites JW. 2010. Erosion of lizard diversity by climate
685	change and altered thermal niches. Science 328:894–899. DOI: 10.1126/science.1184695.
686	Smith JB., Schneider SH., Oppenheimer M., Yohe GW., Hare W., Mastrandrea MD., Patwardhan



08/	A., Burton I., Corree-Moriot J., Magadza CH. 2009. Assessing dangerous climate change
688	through an update of the Intergovernmental Panel on Climate Change (IPCC)"reasons for
689	concern." Proceedings of the National Academy of Sciences 106:4133–4137. DOI:
690	10.1073/pnas.0812355106.
691	Soberón J. 2007. Grinnellian and Eltonian niches and geographic distributions of species.
692	Ecology Letters 10:1115–1123. DOI: 10.1111/j.1461-0248.2007.01107.x.
693	Soberón J., Nakamura M. 2009. Niches and distributional areas: concepts, methods, and
694	assumptions. Proceedings of the National Academy of Sciences 106:19644–19650. DOI:
695	10.1073/pnas.0901637106.
696	Soberón J., Osorio-Olvera L., Peterson T. 2017. Diferencias conceptuales entre modelación de
697	nichos y modelación de áreas de distribución. Revista Mexicana de Biodiversidad:1-5.
698	DOI: 10.1016/j.rmb.2017.03.011.
699	Soberón J., Peterson AT. 2005. Interpretation of models of fundamental ecological niches and
700	species' distributional areas. <i>Biodiversity Informatics</i> 2:1–10. DOI: 10.17161/bi.v2i0.4.
701	Syfert MM., Smith MJ., Coomes DA. 2013. The effects of sampling bias and model complexity
702	on the predictive performance of MaxEnt species distribution models. PloS One
703	8:e55158. DOI: 10.1371/journal.pone.0055158.
704	Thuiller W., Lafourcade B., Engler R., Araújo MB. 2009. BIOMOD – a platform for ensemble
705	forecasting of species distributions. <i>Ecography</i> 32:369–373. DOI: 10.1111/j.1600-
706	0587.2008.05742.x.
707	Townsend Peterson A., Papeş M., Soberón J. 2016. Mechanistic and correlative models of
708	ecological niches. European Journal of Ecology 1:28–38. DOI: 10.1515/eje-2015-0014.
709	Townsend Peterson A., Soberón J. 2012. Species distribution modeling and ecological niche
710	modeling: getting the concepts right. Natureza & Conservação 10:102-107. DOI:
711	10.4322/natcon.2012.019.
712	Tulloch AIT., Sutcliffe P., Naujokaitis-Lewis I., Tingley R., Brotons L., Ferraz KMPMB.,
713	Possingham H., Guisan A., Rhodes JR. 2016. Conservation planners tend to ignore
714	improved accuracy of modelled species distributions to focus on multiple threats and

715 ecological processes. *Biological Conservation* 199:157–171. DOI: 716 10.1016/j.biocon.2016.04.023. 717 Vandermeer J. 1972. Niche Theory. *Annual Review of Ecology and Systematics* 3:107–132. DOI: 718 10.1146/annurev.es.03.110172.000543. 719 Vasconcelos TS., Do Nascimento BTM. 2016. Potential climate-driven impacts on the 720 distribution of generalist treefrogs in South America. *Herpetologica* 72:23–31. DOI: 721 10.1655/HERPETOLOGICA-D-14-00064. 722 Vera-Escalona I., D'Elía G., Gouin N., Fontanella FM., Muñóz-Mendoza C., Sites Jr JW., 723 Victoriano PF. 2012. Lizards on ice: evidence for multiple refugia in *Liolaemus pictus* 724 (Liolaemidae) during the last glacial maximum in the southern andean beech forests. 725 PLoS ONE 7:e48358. DOI: 10.1371/journal.pone.0048358. 726 Warren DL., Glor RE., Turelli M. 2008. Environmental niche equivalency versus conservatism: 727 quantitative approaches to niche evolution. Evolution 62:2868–2883. DOI: 728 10.1111/j.1558-5646.2008.00482.x. 729 Wiens JJ., Graham CH. 2005. Niche Conservatism: Integrating Evolution, Ecology, and 730 Conservation Biology. Annual Review of Ecology, Evolution, and Systematics 36:519– 731 539. 732 Winck G., Almeida-Santos P., Rocha C. 2014. Potential distribution of the endangered endemic 733 lizard Liolaemus lutzae Mertens, 1938 (Liolaemidae): are there other suitable areas for a 734 geographically restricted species? *Brazilian Journal of Biology* 74:338–348. DOI: 735 10.1590/1519-6984.18612. 736 WorldClim - Global Climate Data 2015.CMIP5 30-seconds | WorldClim - Global Climate Data. 737 Available at http://www.worldclim.org/cmip5 30s#com (accessed June 15, 2015). 738 Zajitschek SR., Zajitschek F., Clobert J. 2012. The importance of habitat resistance for movement 739 decisions in the common lizard, *Lacerta vivipara*. *BMC ecology* 12:1–9. DOI: 740 10.1186/1472-6785-12-13. 741 Zank C., Becker FG., Abadie M., Baldo D., Maneyro R., Borges-Martins M. 2014. Climate 742 change and the distribution of neotropical red-bellied toads (*Melanophryniscus*, Anura,

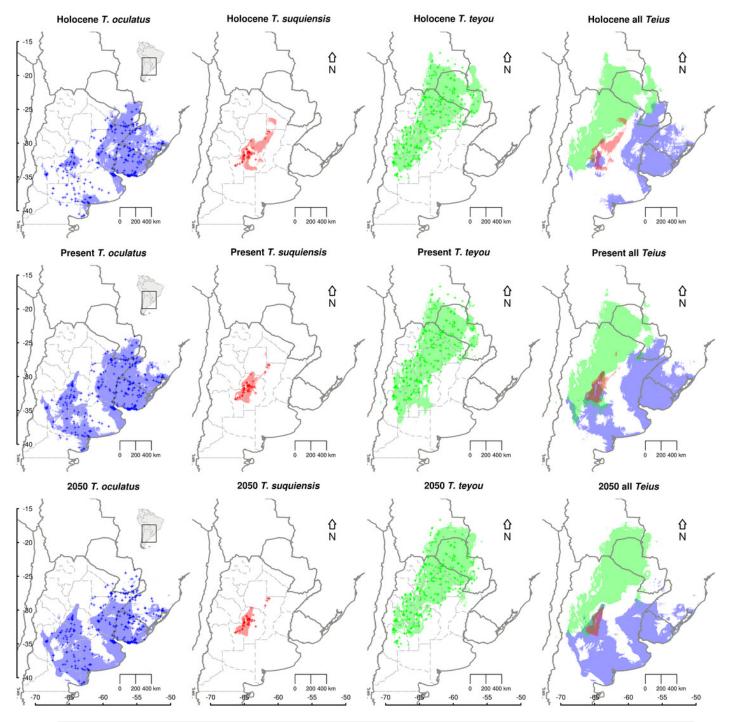


- Amphibia): how to prioritize species and populations? *PLoS ONE* 9:e94625. DOI:
- 744 10.1371/journal.pone.0094625.

Figure 1

Area models for suitability habitat from the averaged replications output.

Teius oculatus in blue, *T. suquiensis* in red, *T. teyou* in green and all the genus. Projection: WGS84, coordinate system: EPSG 4326. Crosses represent localities from each species.



PeerJ Preprints | https://doi.org/10.7287/peerj.preprints.26560v1 | CC BY 4.0 Open Access | rec: 22 Feb 2018, publ: 22 Feb 2018



Figure 2

Barplot of variables contribution for the multivariate environmental space

Reference: dashed line corresponds to the expected value if the contribution where uniform between the considered dimensions.



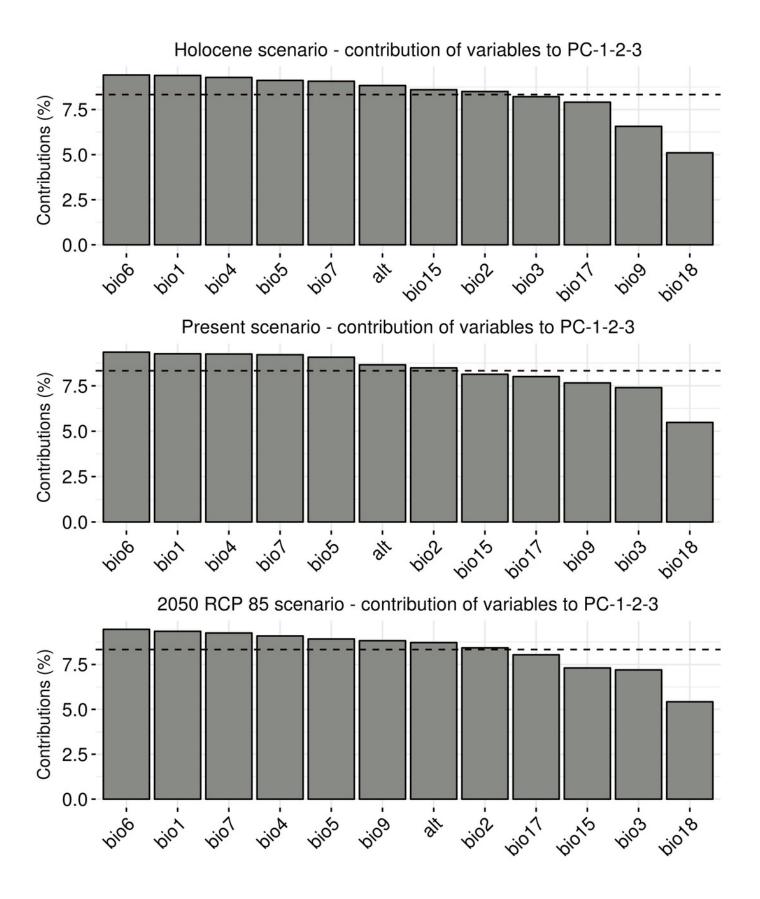




Figure 3

Changes of the environmental space in a multivariate context from the *Teius* studied scenarios

References: A) environmental space in a 0° degree view; B) environmental space in a 37° degree view; C) environmental space in a 74° degree view; D) environmental space in a 111° degree view. Colors: blue: *T. oculatus*; green: *T. teyou* and red: *T. suquiensis*.



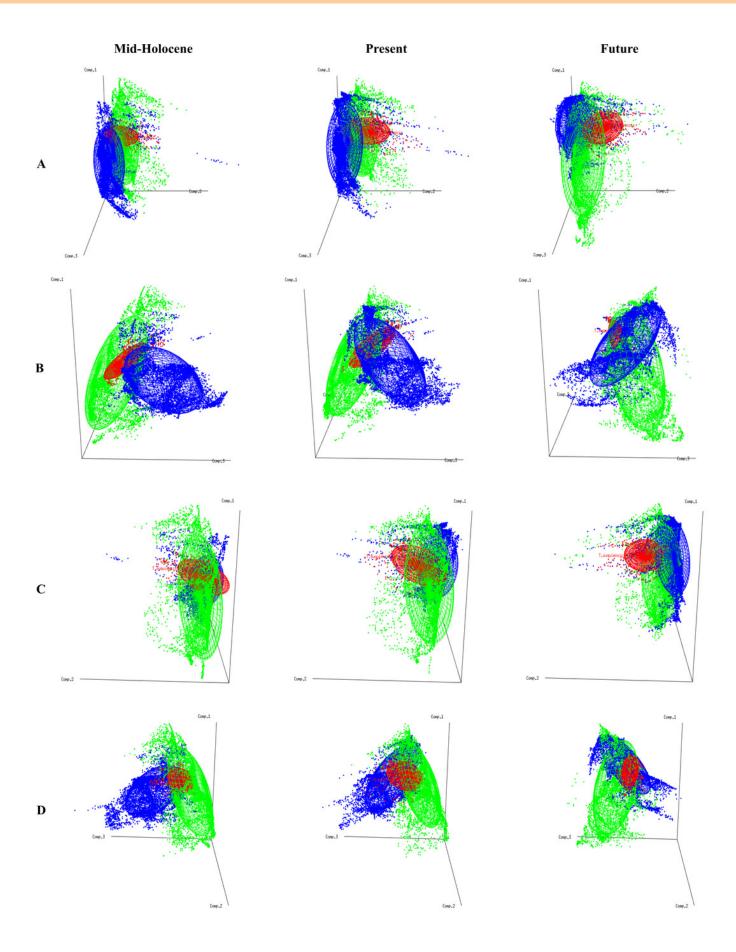




Table 1(on next page)

Spatial analysis performed on output models for each time scenario.

References: % relative exposure to climate change (Recc) from present area.



Area (km²)

Scenarios	Teius	%	T. oculatus	%	T. suquiensis	%	T. teyou	%
Mid-Holocene	1,7174,23.23	-12.48	748,911.98	-29.55	104,209.16	26.00	926,690.31	-6.82
Present	1,962,427.11	-	1,063,106.79	-	82,708.09	-	994,568.05	-
2050	1,892,090.42	-3.58	964,195.47	-9.30	51,624.04	-37.58	986,697.77	-0.79



Table 2(on next page)

Analysis of the environmental space.

Principal Component Analysis of the environmental space for each temporal scenario.



Holocene						
Axis	xis Eigenvalue Variance % Cum		Cumulative Variance %			
PC 1	5.62	46.80	46.80			
PC 2	3.62	30.20	77.00			
PC 3	1.11	9.21	86.21			
		Present				
Axis	Eigenvalue	Variance %	Cumulative Variance %			
PC 1	6.10	50.84	50.84			
PC 2	3.16	26.34	77.18			
PC 3	1.24	10.36	87.55			
Future						
Axis	Eigenvalue	Variance %	Cumulative Variance %			
PC 1	6.13	51.07	51.07			
PC 2	3.19	26.59	77.66			
PC 3	1.15	9.56	87.23			