The effects of sound on the survivorship and embryonic development of a marine gastropod *Stylocheilus striatus*, (aplysiidae).

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How anthropogenic noise pollution affects marine organisms is drawing increasing international concern. There is evidence for anthropogenic noise having negative and harmful effects on the health, development and behavior of many terrestrial species; however, there are few examples of how specific frequencies of sound affect the survivorship and embryonic development of marine invertebrates. This experiment examines the effects of specific frequencies of sound on the survivorship and embryonic development of sound on the survivorship and embryonic development of a marine gastropod, *Stylocheilus striatus* on the island of Mo'orea, French Polynesia. It was found that high frequency sound treatments caused a delay in the embryonic development of *S. striatus* embryos by 3 days while decreasing veliger survivorship by 37%. Additionally, high frequency treatments were shown to cause an observed morphological difference in shell morphology as compared to control and low frequency treatment groups. This study can be used to aid in the management and planning of future conservation polices regarding sound pollution and marine invertebrate gastropods as their presence is crucial for reef health and community structure.

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11 Abstract

12

How anthropogenic noise pollution affects marine organisms is drawing increasing international 13 concern. There is evidence for anthropogenic noise having negative and harmful effects on the 14 health, development and behavior of many terrestrial species; however, there are few examples 15 of how specific frequencies of sound affect the survivorship and embryonic development of 16 marine invertebrates. This experiment examines the effects of specific frequencies of sound on 17 the survivorship and embryonic development of a marine gastropod, Stylocheilus striatus on the 18 island of Mo'orea, French Polynesia. It was found that high frequency sound treatments caused a 19 delay in the embryonic development of S. striatus embryos by 3 days while decreasing veliger 20 survivorship by 37%. Additionally, high frequency treatments were shown to cause an observed 21 22 morphological difference in shell morphology as compared to control and low frequency treatment groups. This study can be used to aid in the management and planning of future 23 24 conservation polices regarding sound pollution and marine invertebrate gastropods as their presence is crucial for reef health and community structure. 25 26

Key words: Anthropogenic, embryonic, *Stylocheilus striatus*, Mo'orea, veliger, survivorshipgastropods, community.

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30 Introduction

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Anthropogenic (human-produced) sound disturbance is drawing increasing concern from the 32 scientific community (McCauley & Fewtrell, 2008; Kight and Swaddle, 2011; Morley et al., 33 34 2014; Nedelec et al., 2014). Since the beginning of the industrial revolution, anthropogenic sound disturbance has continually increased as technology, transportation, resource acquisition 35 and human civilization expands (Morley et al., 2014). This lead to anthropogenic noise becoming 36 classified as a worldwide pollutant within the US National Environmental Policy Act (1972) and 37 the European Commission Marine Strategy Framework Directive (2008). Similarly, the effects 38 39 of noise pollution on terrestrial ecosystems have undoubtedly increased as compared to the past (Barber et al., 2010). 40

Anthropogenic noise pollution research has been suggested to help aid in conservation
management and policy (Kingsford et al., 2002; Barber et al., 2010; Nedelec et al., 2014). For
instance, increases in acoustic pollution reduced the distance and area in which some terrestrial
birds infer their own acoustic signals (Brumm & Slabbekoorn, 2005). Evidence for noise altering
terrestrial vertebrates' foraging behavior, anti-predator responses, reproductive success, and
community structures has also been found (Brumm & Slabbekoorn, 2005; Arch & Narins 2008),
all of which can be assessed for conservation management.

49

50 In marine environments anthropogenic sound disturbance is centered between economic

51 transport routes, commercial harvesting fisheries, or underwater drilling sites (Slabbekoorn et al.,

52 2010). Research associated with marine anthropogenic noise pollution has primarily investigated

mammals and fish (Wells et al., 1999; McCauley & Fewtrell, 2008; Morley et al., 2014). Larval

54 stage reef-dominate fish have been found to locate and settle on coral reefs due to sound cues

55 made by fish and shrimp currently living on the reef (Montogomer et al., 2006). This

56 demonstrates how sound is guiding behavior and how anthropogenic noise pollution is

57 interfering with recruitment patterns (Kingsford et al., 2002; Simpson et al., 2004, 2005;

58 Montgomery et al., 2006).

59

There have been considerably fewer investigations into invertebrate behavior and response to increases in anthropogenic sound. One study found coral larvae respond to acoustic cues which aid in their detection of habitat from large distances and from up-current of their preferred settlement locations (Vermeij et al., 2010). This was the first observed auditory response in the invertebrate phylum Cnidaria (Vermeij et al., 2010), demonstrating the urgency needed for additional research on invertebrate response to anthropogenic noise pollution.

66

How anthropogenic noise pollution affects an organism's fitness in early life stages is also an
understudied area (Morley et al., 2014; Nedelec et al., 2014), despite the impacts it might have
on population dynamics (Lindström, 1999; Nedelec et al., 2014). Fortunately, the life cycles of
some marine invertebrates are relatively short (Morley et al., 2014; Nedelec et al., 2014),
allowing fitness and population dynamics to be tested with methods logistically different than
compared to many vertebrates (Morley et al., 2014; Nedelec et al., 2014). The embryonic

73 development of the sea hare *Stylocheilus striatus* (a gastropod in the family Aplysiidae) found on

- 74 Moorea, French Polynesia, was discovered to become impaired by anthropogenic noise playback
- 75 (Nedelec et al., 2014). Increased mortality rates from noise in the veliger (the final larval stage of
- this mollusk, often denoted by small ciliated flaps used for swimming and feeding) stage was
- also observed (Nedelec et al., 2014). It is not known which frequency of sound decreases
- 78 survivorship of the juvenile veligers of *S. striatus*.
- 79
- 80 The biological importance of *Stylocheilus striatus* measures large in terms of herbivorous
- 81 maintenance of algal reefs (Nedelec et al., 2014), and as invertebrates are vital parts of trophic

webs (McCauley & Fewtrell 2008; Morley et al., 2014). It is, therefore, important to understand

the effects of specific frequencies of noise on early life stages in invertebrates and their possible

84 implications in marine ecosystems. This experiment elaborates on Nedelec et al. (2014), while

serving as a quantitative approach to identifying how varying frequencies of sound affect the

survivorship and embryonic development of the early life stages of *S. striatus*. This experiment

will also contribute to understanding invertebrate fitness at crucial, early life stages, potentiallyaiding in future conservation management and policy.

89

90 Materials & Methods

91 Study site and species

92 Forty-six adults of the sea hare *Stylocheilus striatus* were acquired courtesy of the CRIOBE

93 (Centre de Recherches Insulaires et Observatoire de l'Environnement) research facility located

on the Island of Mo'orea in French Polynesia. The specimens acquired from the CRIOBE facility

95 were located from the lagoon around Moorea. *S. striatus* specimens were kept in an aerated

96 flowing seawater aquarium at the University of California Gump Field Research Station. The

97 specimens grazed on cyanobacteria (Lyngyba majuscule) collected from white sandy substrates

98 within the lagoon around Mo'orea at two locations (Figure 1; Table 1).

99 Sea hare reproduction

100 Individual S. striatus specimens were paired with similar sized individuals in 0.5 L cylindrical

101 breeding containers that allowed seawater flow while preventing specimens from mixing with

the main population (Nedelec et al. 2014). Specimens were monitored hourly throughout the
night until copulation was observed (Nedelec et al. 2014). Following mating, sea hare mothers
were separated into 50 mL falcon tubes thus allowing maternity to be known (Nedelec et al.
2014). This opisthobranch gastropod lays a gelatinous string of eggs (an egg ribbon) with each
egg containing 1-6 embryos (Nedelec et al. 2014). The following morning, mother specimens
were removed from the falcon tubes and the egg ribbons were left to develop (Nedelec et al.
2014).

109 Egg preparation

Egg ribbons were cut into three similarly sized segments using a scalpel and randomly separated 110 and prepped for treatment four days after deposition; denoted by the embryotic formation of villi 111 and defined shell morphology (Figure 2). This stage of the embryonic development was chosen 112 to minimize chances of disease or an early developmental issue causing mortality while allowing 113 accurate counts of hatched and unhatched veligers. Three cross sections of each egg ribbon (the 114 middle and both ends) were taken and the numbers of eggs and veligers were counted using a 115 compound light microscope. The mean of the cross section's counts was used as an extrapolation 116 multiplier against the number of eggs stacked lengthwise within the ribbon to provide an egg 117 number estimation for each segment of ribbon. 118

119 Sound experiments

- 120 To examine the effect of sound on development of *S. striatus* eggs, 6 hour tones of pure sine
- 121 wave function MP3 clips of 100 hertz (low frequency treatment) and 1000 hertz (high frequency
- treatment) were generated using a multi-track audio editor and recorder, Audacity (v2.1.2,
- 123 Carnegie Mellon University, 2016). Tones were played through a waterproof Bluetooth speaker
- 124 (Toshiba DT-B660, 4W, frequency response 0.1-20kHz, Japan) via a Bluetooth (v4.0, A2DP,
- 125 EDR, LE, aptX) connection to a Samsung Galaxy S5-Active phone (Operating system: Android
- 126 2014, Samsung Galaxy, Suwon, South Korea).
- 127 Decibel levels of underwater sound recordings were measured using a digital voice recorder
- 128 converted into a hydrophone (Olympus DM- 620, 3 Microphone, Linear Pulse Code Modulator,
- input level -70dB, Tokyo, Japan). Decibel levels ranged from 76-101 decibels with a reference

level of 20 µPa per second squared per hertz. These measurements were then converted to an
underwater decibel level by adding 61.5 decibels to the readings (DOSITS, 2016; NOAA, 2016).

Petri dishes containing a sea hare egg ribbon segment in 20mL of seawater were placed directly above the speaker and secured to the speaker to prevent the petri dish from vibrating on the speaker. The 100 Hz clip was played to egg ribbon segments between the hours of 12:00 and 18:00 while the 1000 Hz clip was played between 22:00 and 06:00. The control egg ribbon segment did not receive sound treatment. Control and previous treated egg ribbons were kept in a separate laboratory on the Gump facility to control for sound. After treatments were finished, eggs hatched 1-3 days later.

The following categories were counted from October 9th to November 10th, 2016 using the methods adopted and modified from Nedelec et al. (2014): (1) dead eggs or eggs failed to undergo organogenesis (i.e. failed to develop); (2) unhatched eggs with dead developed embryos; (3) total number of dead embryos in unhatched eggs; (4) hatched eggs; and (5) eggs containing veligers with shell abnormalities. Post treatment egg segments were shuffled before counting to reduce counting biases. These egg segment lengths were measured again to identify shuffled segments. Survivorship was measured as the percentage of hatched eggs.

146 Statistical analyses

To avoid pseudoreplication, tests were conducted with each individual sea hare that produced an egg mass (i.e., mother) as the unit of analyses (Nedelec et al. 2014). All tests used the statistical program RStudio (R Core Team, 2013). Kruskal-Wallis analyses were performed to examine differences between means (n = 12 mothers) of control groups and treatment groups; as the survivorship data was not normally distributed. A post-hoc Tukey analysis was performed to identify where survivorship differences occurred between groups.

153 Results

154 Embryonic developmental time

155 *Stylocheilus striatus* eggs hatched 1 to 3 days after sound treatments, totaling 5 to 7 days in

- embryonic development (Figure 3). Egg hatching times were significantly affected by sound
- treatments (Kruskal-Wallis Ranked Sum, $\chi^2 = 9.0474$, df = 2, p < 0.05). High frequency treated

- eggs hatched, on average, 3 days after treatment while the control and low frequency treated
- eggs hatched after 2 and 1.5 days respectively (Figure 3). Post-hoc analysis of means (n = 12)
- 160 indicated eggs treated with high frequency tones hatched significantly later than control and low
- 161 frequency treated eggs (Kruskal-Nemenyi post-hoc analysis, p < 0.005). However,
- 162 developmental time between controls and low frequency treated eggs were not significantly
- affected by sound treatments (Kruskal-Nemenyi, p > 0.05).

164 Egg hatching and percent survivorship

- 165 Of the 12,267 eggs counted, 615 failed to develop to 4 day old embryos (Figure 4). Of the 11,652
- surviving eggs 2,583 failed to hatch (Figure 4). High frequency treated eggs had an average
- survivorship of 52.0% while the control and low frequency treated eggs had a survivorship of
- 168 89.1% and 89.8% respectively (Figure 5). The percentage of hatched eggs (survivorship) was
- significantly affected by high frequency sound treatment (Kruskal-Wallis Ranked Sum test, $\chi^2 =$
- 170 23.447, df = 2, p < 0.005). Additional exploration of the differences between survivorship
- 171 means via a post-hoc pairwise analysis revealed high frequency treatments had the greatest
- impact on survivorship (Kruskal-Nemenyi post hoc analysis, p < 0.005)

173 Interestingly, the pair-wise comparison between low and high frequency groups and the control and high frequency groups indicated the low and high frequency treatments had the greatest 174 175 significance difference between mean survivorship (32.8%) (Kruskal-Nemenyi, p < 0.005). The number of hatched eggs varied greater between the low frequency and high frequency treatments 176 177 (Figure 5). The pair-wise comparison between the low frequency group and the high frequency group demonstrated high frequency treatments had the greatest impact on the number of hatched 178 179 eggs (Kruskal-Nemenyi, p < 0.005). There was no significant difference between the survivorship of the control and low frequency treatments (Kruskal-Nemenyi, p > 0.05). 180

181 Morphological differences

182 Of the 2,583 eggs that failed to hatch, 159 eggs contained veligers with an observed difference in

- their shell morphology as compared to a healthy, surviving veliger (Figure 4). Veligers
- 184 considered morphology different contained abrupt edges or protrusions within their shell (Figure
- 185 4). High frequency treated eggs had a mean of 10 eggs containing veligers with morphological
- differences, while the control and low frequency treated eggs had means of 1 and 2 eggs with

187 deformed shells, respectively (Figure 6). The number of eggs with deformed veligers was

188 significantly affected by high frequency sound treatment (Kruskal-Wallis Ranked Sum test, $\chi^2 =$

189 12.253, df = 2, p < 0.05). Similarly, the pair-wise comparison between deformed shell means

190 revealed high frequency treatments had the greatest impact on the number of eggs containing

191 deformed veligers (Kruskal-Nemenyi, P < 0.05), with no significant difference between control

and low frequency treatment groups (Kruskal-Nemenyi, p > 0.05).

193 Discussion

High frequency noise treatment significantly slowed hatching times of S. striatus eggs (Figure 194 3). Past research has found anothropogenic noise can delay development in New Zealand scallop 195 196 larvae (Aguilar de Soto et al., 2013). This is not the first evidence for sound exposure to decrease developmental life-history rates (see Lagardère, 1982). Lagardère (1982) discovered high level 197 sound exposure also slowed developmental rates of *Crangon crangon*, a species of brown marine 198 shrimp found in the northeastern Atlantic Ocean. However, in situ research also reported 199 anthropogenic sound from boat noise playback may have no apparent effect on embryonic 200 developmental rates (Nedelec et al., 2014). Although the results of the current study contrast 201 with those of Nedelec et al., (2014), a number of methodologically different measures allow the 202 current experiment to provide evidence for anthropogenic noise slowing hatching times of 203 marine invertebrates in a controlled setting. The sea hare embryos were subject to 6 hours of 204 constant sound treatment at 137.5-162.5 decibels with a reference level of 1 µPa per second 205 squared per hertz at either 100 or 1000 hertz frequencies. It is possible that the level of sound and 206 egg ribbon treatment time may be the greatest factors causing the observed delay in hatching 207 time of these ecologically and economically important organisms (McCauley & Fewtrell 2008; 208 209 Wale et al., 2013; Morley et al., 2014).

This study was performed in a laboratory with the goal of discerning which frequencies of noise have the greatest impact on the survivorship *of S. striatus* embryos (Figure 3); however, because it was executed *in vitro*; it may be difficult to pinpoint how specific frequencies of noise pollution affects organisms in their natural environments. Nedelec et al., (2014) suggested it is possible that vibrations of substrates from sound pollution may affect sound transmission differently. Regarding the current experiment, egg ribbons were placed in a petri dish above the speaker, potentially altering the vibrational patterns of sound through the plastic dish into the sea

217 water containing the egg ribbon. Although, an increase in developmental time and a decreased

- survivorship was observed, the laboratory apparatus may have affected sound vibrations,
- 219 pressure and subsequently the hatching time and survivorship of sea hare eggs.

High frequency noise treatment also significantly decreased survivorship of S. striatus veligers. 220 A potential mechanism for this observed response to anthropogenic noise may be due to pressure 221 induced injuries caused by cavitation or barotrauma. Cavitation is when gas bubbles collect and 222 223 implode causing high levels of heat and pressure (Brennen, 1995); while barotrauma is the rupturing of gas filled cavities from changes in pressure (Nedelec et al, 2014). There are few 224 studies of noise-induced cavitation stress and how they affect organisms, yet cavitation can be 225 the result of long exposure times of high frequency sound (Leighton, 1995). S. striatus embryos 226 227 may have gas filled chambers within their tissues where cavitation and barotrauma may occur when they are subject to relatively long term and constant high frequency noise. 228

Another possible mechanism for the observed survivorship and morphological difference is 229 molecular vibrations within the developing organisms (Silva et al., 2002; Aguilar de Soto et al., 230 2013; Nedelec et al., 2014). Low frequency noise and whole-body vibrations caused increased 231 levels of sister chromatid exchange in mice (Silva et al., 2002), and high frequency and sound 232 level caused a deregulation of calcium transport systems in rats (Siegel & Mooney, 1987; 233 Aguilar de Soto, 2013). Because of the small size of scallop larvae and the absence of strong 234 tissue density gradients in early development phases, molecular vibrations have been suggested 235 to decreases survivorship and cause body malformations as well (Aguilar de Soto et al., 2013). 236 According to Aguilar et al., (2013) these observed processes may be related to particle motion 237 rather than to the pressure component of noise exposure. As an herbivore, invertebrates like S. 238 239 striatus often provide a balance between algal and coral health, specifically regarding their specialized grazing of toxic cyanobacteria (Paul & Pennings, 1991; Nedelec et al., 2014). With 240 increasing noise pollution in marine environments, invertebrate response and behavior, like that 241 of S. striatus, require further investigation to determine appropriate mitigation policies. 242

Anthropogenic noise pollution continues to increase across the world. Conservation policy and
management should consider the effects of anthropogenic noise on invertebrates like *S. striatus*,
as their herbivory is important to marine ecosystem health as well as their role within trophic
webs. Research from studies directly accessing how organisms respond to specific frequencies

- 247 of noise, as this experiment has done, are crucial to understanding and implementing future
- 248 domestic and international conservation and management policy, ultimately addressing the
- 249 worldwide concern of sound pollution.

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254 **References**

- Aguilar DS, Delorme N, Atkins J, Howard S, Williams J, Johnson M. 2013. Anthropogenic
 Noise causes body malformations and delays development in marine larvae. *Nature*.
 Scientific Reports 3: 2831
- Arch VS, Narins PM. 2008. 'Silent' signals: selective forces acting on ultrasonic communication
 systems in terrestrial vertebrates. *Animal Behavior* 76: 1423-1428.
- Barber JR, Crooks KR, Fristrup KM. 2010. The costs of chronic noise exposure for terrestrial
 organisms. *Trends in ecology & evolution* 25:180-189.
- 262 Brennen C, 2015. Cavitation and Bubble Dynamics. Oxford University Press. p. 21
- Brumm H & Slabbekorn H. 2005. Acoustic communication in noise. *Adv. Study. Behav* 35: 151264 209.
- Discovery of Sound in the Sea (DOSITS) 2016. University of Rhode Island, Graduate School of
 Oceanography. http://www.dosits.org/science/soundsinthesea/airwater/. Accessed 10
- 267 November 2016.
- European Parliament 2008 Directive 2008/49/EC of the European Parliament and the Council.
 Official J. Eur. Communities L 189:12-26.
- 270 Gross JA, Irvine KM, Wilmoth S, Wagner TL, Shields PA, Fox JR. 2013. The effects of pulse
- pressure from seismic water gun technology on Northern Pike. *Transactions of the American Fisheries Society* 142(5):1335-1346.
- 273 Kight CR & Swaddle JP. 2011. How and why environmental noise impacts animals: an
- integrative, mechanistic review. *Ecology letters* 14(10):1052-1061.

275	Kingsford MJ, Leis JM, Shanks A, Lindeman KC, Morgan SG, Pineda J. 2002. Sensory		
276	environments, larval abilities and local self-recruitment. Bulletin of Marine		
277	Science 70(1):309-340.		
278	Lindström J. 1999. Early development and fitness in birds and mammals. Trends in Ecology &		
279	Evolution 14(9)343-348.		
280	McCauley RD & Fewtrell J. 2008. Marine invertebrates, intense anthropogenic noise, and squid		
281	response to seismic survey pulses. Bioacoustics 17:315-318.		
282	Montgomery JC, Jeffs A, Simpson SD, Meekan M, Tindle C. 2006. Sound as an orientation cue		
283	for the pelagic larvae of reef fishes and decapod crustaceans. Advances in Marine		
284	<i>Biology</i> 51:143-196.		
285	Morley EL, Jones G, Radford AN. 2014. The importance of invertebrates when considering the		
286	impacts of anthropogenic noise. Proc. Royal Society B 281:2013-2683.		
287	National Ocean and Atmospheric Administration (NOAA) 2016.		
288	$\underline{http://ocean explorer.noaa.gov/explorations/sound01/background/acoustics/acoustics.html}$		
289	Accessed: 10 November 2016.		
290	Nedelec SL, Radford AN, Simpson SD, Nedelec B, Lecchini D, Mills SC. 2014. Anthropogenic		
291	noise playback impairs embryonic development and increases mortality in a marine		
292	invertebrate. Scientific Reports 2831.		
293	Patek SN, & Caldwell RL. 2005. Extreme impact and cavitation forces of a biological hammer:		
294	strike forces of the peacock mantis shrimp Odontodactylus scyllarus. Journal of		
295	Experimental Biology 208(19):3655-3664.		
296	Radford AN, Morley EL, Jones G. 2012. The effects of noise on biodiversity, Defra Report		
297	NO0235. Department for Environment. Food, & Rural Affairs, London.		
298	Shieh BS, Liang SH, Chen CC, Loa HH, Liao CY. 2012. Acoustic adaptations to anthropogenic		
299	noise in the cicada Cryptotympana takasagona Kato (Hemiptera: Cicadidae). acta		
300	<i>ethologica</i> 15(1):33-38.		
301	Siegel M. & Mooney MP. 1987. Perinatal stress and increased fluctuating asymmetry of		
302	dental calcium in the laboratory rat. Am. J. Phys. Anthropol 73:267-270.		
303	10.1002/ajpa.1330730213		
304	Simpson SD, Meekan M, Montgomery J, McCauley R, Jeffs A. 2005. Homeward		
305	sound. Science 308(5719):221-221.		

306	Simpson SD, Meekan MG, McCauley RD, Jeffs A. 2004. Attraction of settlement-stage coral		
307	reefs fishes to ambient reef noise. Mar Ecol Prog Ser 276:263-268.		
308	Silva MJ, Dias A, Barreta A, Nogueira PJ, Castelo-Branco NAA & Boavida MG. 2002. Low		
309	frequency noise and whole-body vibration cause increased levels of sister chromatid		
310	exchange in splenocytes of exposed mice. Teratogenesis, carcinogenesis, and		
311	mutagenesis 22(3):195-203.		
312	Slabbekoorn H, Bouton N, van Opzeeland I, Coers A, ten Cate C, Popper AN. 2010. A noisy		
313	spring: the impact of globally rising underwater sound levels on fish. Trends in Ecology		
314	& Evolution 25(7):419-427.		
315	US National Environmental Policy Act. <u>http://www.gsa.gov/portal/content/100896</u> Accessed: 3		
316	October 2016.		
317	Vermeij MJ, Marhaver KL, Huijbers CM, Nagelkerken I, Simpson SD. 2010. Coral larvae move		
318	toward reef sounds. <i>PloS one</i> , 5(5):e10660.		
319	Wale MA, Simpson SD & Radford AN. 2013. Size-dependent physiological responses of shore		
320	crabs to a single and repeated playback of ship noise. Biology Letters 9:20121194		
321	Wells RS, Boness DJ, Rathbun GB, 1999. Biology of marine mammals. Behavior 324-422 in JE		

322 Reynolds III and SA Rommel edition.

323 Figures and Tables

- **Figure 1:** Map of the island of Mo'orea (Scale bar = 5km). The stars and numbers represented
- field collection sites corresponding with Table 1. The grey line around the island represents coralreefs.



329

Collection Site		GPS Coordinates
Temae Public Beach	1	S 17° 29' 56.0544" W 149° 45' 32.778"
Piha'ena Public Beach	2	S 17° 29' 0.6125" W 149° 50' 0.9318"

Table 1: Cyanobacteria collection sites and GPS Coordinates.

Figure 2: Pretreatment *Stylocheilus striatus* embryos four days after copulation. (Scale bar =

331 0.5mm).



- **Figure 3**. A boxplot of *Stylocheilus striatus* egg development time in days, grouped by
- frequency treatment (n =12). A frequency of 0 Hz represents the control, while 100 Hz and 1000
- Hz represent the low and high frequency treatment groups respectively.



Frequency (Hz)

- **Figure 4.** *Stylocheilus striatus* egg ribbon segment post high frequency (1000 Hz) treatment. Red
- arrows indicate eggs that failed to reach day 4 of development. Light blue arrows specify
- unhatched eggs with dead embryos. Yellow arrows indicate eggs with abnormal shell
- morphology denoted by abrupt edges or protrusions within the shell (Scale bar = 0.5 mm).



- **Figure 5.** A boxplot of hatched eggs as percent survivorship (n=12 mothers), grouped by
- 343 frequency treatment. A frequency of 0 Hz represents the control, while 100 Hz and 1000 Hz
- 344 represent the low and high frequency treatment groups respectively.



- 346 Figure 6. A boxplot of the number of eggs containing deformed veligers (n=12 mothers),
- 347 grouped by frequency treatment. A frequency of 0 Hz represents the control, while 100 Hz and
- 348 1000 Hz represent the low and high frequency treatment groups respectively. The hollow circle
- 349 represents an outlier.

