A peer-reviewed version of this preprint was published in PeerJ on 23 May 2018.

<u>View the peer-reviewed version</u> (peerj.com/articles/4857), which is the preferred citable publication unless you specifically need to cite this preprint.

Kim J, Lee S, Shin M, Lee C, Seo C, Eo SH. 2018. Altitudinal patterns in breeding bird species richness and density in relation to climate, habitat heterogeneity, and migration influence in a temperate montane forest (South Korea) PeerJ 6:e4857 https://doi.org/10.7717/peerj.4857



Altitudinal patterns in breeding bird species richness and density in relation to climate, habitat heterogeneity, and migration influence in South Korea

Jin-Yong Kim 1,2, Sanghun Lee 3, Man-Seok Shin 1, Chang-Hoon Lee 3, Changwan Seo 4, Soo Hyung Eo Corresp. 2

Corresponding Author: Soo Hyung Eo Email address: eosh@kongju.ac.kr

Altitudinal patterns in the population ecology of mountain bird species are useful for predicting species occurrence and behavior. Numerous hypotheses about the complex interactions among environmental factors have been proposed; however, these remain controversial. This study investigated the causes of altitudinal patterns in breeding bird species richness and density in relation to climate, habitat heterogeneity, and migration influence in Jirisan National Park, South Korea. We conducted a field survey of 142 plots at altitudes between 200 and 1400 m above mean sea level in the breeding season. A total of 2771 individuals from 53 breeding bird species were recorded. Altitudinal patterns of species richness and density showed a hump-shaped pattern, indicating that the highest richness and density could be observed at moderate altitudes. Models constructed with 16 combinations of six variables demonstrated that species richness was positively correlated with vertical and horizontal habitat heterogeneity, and that density was positively correlated with vertical, but not horizontal, habitat heterogeneity, and negatively correlated with migrant rate. No significant relationships were found between spring temperature and species richness or density. Therefore, the observed patterns in species richness support the hypothesis that habitat heterogeneity, rather than climate, is the main driver of species richness. And neither habitat heterogeneity nor climate hypotheses fully explains the observed patterns in density. However, vertical habitat heterogeneity does likely help explain observed patterns in density. The heterospecific attraction hypothesis did not apply to the distribution of birds along the altitudinal gradient. Appropriate management of vertical habitat heterogeneity, such as vegetation cover, should be maintained for the conservation of bird diversity in this area.

¹ Division of Ecosystem Services and Research Planning, National Institute of Ecology, Seocheon, South Korea

² Department of Forest Resources, Kongju National University, Kongju, South Korea

³ Division of Basic Ecology, National Institute of Ecology, Seocheon, South Korea

⁴ Division of Ecological Survey Research, National Institute of Ecology, Seocheon, South Korea



Altitudinal patterns in breeding bird species richness and density in relation to climate, habitat heterogeneity, and migration influence in South Korea

Jin-Yong Kim^{1,2}, Sanghun Lee³, Man-Seok Shin¹, Chang-Hoon Lee³, Changwan Seo⁴, Soo Hyung Eo²

Corresponding Author:

Soo Hyung Eo²

corresponding author eosh@kongju.ac.kr

¹ Division of Ecosystem Services and Research Planning, National Institute of Ecology, Seocheon, South Korea

² Department of Forest Resources, Kongju National University, Choongnam, South Korea

³ Division of Basic Ecology, National Institute of Ecology, Seocheon, South Korea

⁴ Division of Ecological Survey Research, National Institute of Ecology, Seocheon, South Korea



Abstract

1

23

24

25

26

27

28

29 30

31

32

33

34

35

36

37

38 39

Altitudinal patterns in the population ecology of mountain bird species are useful for 2 predicting species occurrence and behavior. Numerous hypotheses about the complex 3 interactions among environmental factors have been proposed; however, these remain 4 5 controversial. This study investigated the causes of altitudinal patterns in breeding bird species 6 richness and density in relation to climate, habitat heterogeneity, and migration influence in 7 Jirisan National Park, South Korea. We conducted a field survey of 142 plots at altitudes between 200 and 1400 m above mean sea level in the breeding season. A total of 2771 8 9 individuals from 53 breeding bird species were recorded. Altitudinal patterns of species richness and density showed a hump-shaped pattern, indicating that the highest richness and density could 10 be observed at moderate altitudes. Models constructed with 16 combinations of six variables 11 demonstrated that species richness was positively correlated with vertical and horizontal habitat 12 heterogeneity, and that density was positively correlated with vertical, but not horizontal, habitat 13 heterogeneity, and negatively correlated with migrant rate. No significant relationships were 14 found between spring temperature and species richness or density. Therefore, the observed 15 patterns in species richness support the hypothesis that habitat heterogeneity, rather than climate, 16 is the main driver of species richness. And neither habitat heterogeneity nor climate hypotheses 17 fully explains the observed patterns in density. However, vertical habitat heterogeneity does 18 likely help explain observed patterns in density. The heterospecific attraction hypothesis did not 19 apply to the distribution of birds along the altitudinal gradient. Appropriate management of 20 vertical habitat heterogeneity, such as vegetation cover, should be maintained for the 21 conservation of bird diversity in this area. 22

Introduction

Altitudinal changes in bird species diversity provide important information on the limitation of species distribution within mountain areas, and these changes are widely used to identify and study the environmental variables that affect the structure of communities (Adolfo & Navarro, 1992). Recently, there has been an increase in studies on climate change; thus, the data on the distribution of bird species along altitudinal gradients often serve as time-space substitutes and provide valuable predictive information (Chamberlain et al., 2016). Consequently, studies on biodiversity along altitudinal gradients have been of interest to many researchers. One commonly recognized pattern shows decreasing richness with increasing elevation (Terborgh, 1977; Stevens, 1992; Herzog, Kessler & Bach, 2005). However, recent studies have indicated that bird diversity patterns may be more complex (Poulsen &Lambert, 2000; Rahbek, 2005; McCain, 2009). McCain (2009) suggested that, from the point of view of climate zones, four elevational richness patterns are represented. These are 1) decreasing, 2) low plateau, 3) low plateau with a mid-elevational peak, and 4) mid-elevational peak. To explain these altitudinal patterns, numerous hypotheses have been proposed (Rahbek, 2005; Rahbek et al., 2007; McCain, 2009; Pan et al., 2016).

These hypotheses generally fall into four main categories: climate, space, evolutionary history, and biotic processes (Pianka, 1966; Gaston, 2000; McCain, 2009). Climatic hypotheses are based on the theory that species diversity is affected by conditions such as temperature, rainfall, productivity, humidity, and cloud cover (McCain, 2009). Spatial hypotheses suggest that the spatial extent of species distribution is reduced with increasing altitude, and thus, species diversity is simultaneously reduced (Sanders & Rahbek, 2012; Pan et al., 2016). Evolutionary history hypotheses are linked to speciation rates, immigration, extinction rates. and phylogenetic niche conservation (Diamond, 1988; Lomolino, 2001; Allen, Brown & Gillooly, 2002, McCain, 2009), based on the assumption that speciation takes place most rapidly at low altitude and declines with increasing altitude, and that extinction rate is highest at mountaintops and declines with decreasing altitude (McCain, 2009). Finally, biological hypotheses include competition and habitat heterogeneity and complexity (MacArthur & MacArthur, 1961; Terborgh, 1977; McCain, 2009). The biological hypotheses have rarely been tested because of the difficulties in defining the characteristics and performing quantitative measurements in many species (McCain, 2009; Pan et al., 2016).

Among the numerous hypotheses, climatic and biological hypotheses are the most widely supported (Lee et al., 2004; McCain, 2009; Pan et al., 2016). Climatic variables are considered to be the main driver of bird diversity (McCain, 2009), and temperature shows a distinct pattern that decreases with increasing altitude, which directly affects the physiological tolerance of birds (Currie et al., 2004; Pan et al., 2016) and indirectly affects birds by influencing vegetation and food resources. Therefore, the climate hypothesis has been tested in many studies. However, many mechanistic models cannot fully explain the relationship between contemporary climate and species diversity (Currie et al., 2004; Rahbek et al., 2007). Therefore, alternative hypotheses have emerged, and the importance of habitat heterogeneity has been noted (Rahbek et al., 2007). Generally, habitat heterogeneity can positively influence bird species richness (Hurlbert, 2004; Pan et al., 2016); therefore, the hypotheses have been receiving increased attention despite the difficulties in measurement and definition (Pan et al., 2016). Habitat heterogeneity hypothesis proposes that a greater variety of habitat types per unit area and a greater complexity of vegetation structure lead to increased diversity (MacArthur & MacArthur, 1961; Pan et al., 2016). However, most studies have been limited in scope by only employing horizontal factors, such as the variety of habitat types per unit area (Pan et al., 2016).

Although several environmental variables affect species diversity according to altitude, it is important to consider intra- and interspecific relationships. However, hypotheses involving evolutionary history have difficulties in considering the complex dynamics among extinction, immigration, and speciation (Diamond, 1988; McCain, 2009), and most studies have focused on population dynamics over time. Because of these limitations, the importance of spatial distribution and migration has increased, and migration is not only an alternative mechanism to explain birth and death, but also an important process in itself (Dingle & Drake, 2007). According to the heterospecific attraction hypothesis, migrants use residents as a cue to identify sites suitable for breeding because residents occupy higher-quality sites (Mönkkönen et al., 1997;



Mönkkönen & Forsman, 2002). Therefore, increasing migration should positively affect species richness and density of a given site. However, to the best of our knowledge, the heterospecific attraction hypothesis has not yet been applied in advanced studies along an altitudinal gradient.

This study aimed to investigate the altitudinal patterns in breeding bird species richness and density, and their correlations with environmental factors at Jirisan National Park, South Korea, situated in a temperate region. The causes of such patterns were tested in relation to climate (temperature and relative humidity), habitat heterogeneity (vertical coverage of vegetation and horizontal habitat diversity), and migrant rate (migrant rate of species and migrant rate of individuals).

Materials and Methods

Study area

This study was carried out in a forest in Jirisan National Park, the largest national park in South Korea with a total area of 481.022 km² (Fig.1). All field surveys were conducted with the approval and access permits from the Korea National Park Service. The altitude in the park ranges from 110 m to 1915 m above sea level. The vegetation of the subalpine forest (up to 1400 m above mean sea level) is characterized by tree species such as *Betula ermanii*, *Malus baccata*, *Picea jezoensis*, *Pinus koraiensis*, *Abies koreana*, *Quercus mongolica*, *Q. serrata*, *Q. variabilis*, *Stewartia pseudocamellia*, *Pinus densiflora*, and *Abies holophylla* (Gwon et al., 2013). The study focused on montane forest areas between altitudes of 200 and 1400 m above mean sea level, because the altitudes above 1500 m include ridges, most of which are populated by coniferous shrubs.

Bird survey

A total of 142 plots were surveyed along the elevational gradient in mixed or deciduous forested areas, with coniferous forests excluded from the survey area to minimize the differences in bird communities according to forest type (Table 1). We randomly chose 10–12 plots within each 100-m elevation bracket within an altitudinal range of 200–1400 m. The location of each plot was recorded using a Global Positioning System (GPS; Oregon 300, Garmin, Lenexa, KS, USA). Surveys of bird fauna and vertical coverage of vegetation were undertaken in every plot. Point counts of birds (Reynolds, Scott & Nussbaum, 1980) were carried out between late May and June 2015 to account for summer migratory arrivals. At each plot, all breeding bird seen and heard within a 50 m radius (0.8 ha) were recorded the No. of individuals and species using 15 min count period. Point count commenced directly after sunrise and continued until 8 a.m. in good weather conditions (without precipitation, fog and prevalent wind). We did not count chicks, to reduce the change in the number of individuals caused by fledging of chicks. Nonbreeding species, which were classified as passing migrants, were eliminated from the analysis (Table S1).



Climate variables

We used the Weather Research and Forecasting (WRF) version 3.6 model to retrieve climate parameters, including mean spring temperature and relative humidity, on regional and local scales. These parameters were compiled over a three-month period using terrestrial data from the National Center for Environmental Prediction (NCEP) Final (FNL) Operational Global Analysis data. Using these data, climate simulation with WRF was executed for April, May, and June 2015 at time intervals (Δt) of 180 s. Since the NCEP input data resolution of 1° is very coarse for regional or local climate simulations, the domains in this study were downscaled into two-way quadruples of 27, 9, 3, and 1 km with 31 vertical levels in WRF. Simulation outputs were produced every hour with a cumulus parameterization scheme by Kain and Fritsch (Kain & Fritsch, 1993), the WRF Double Moment 6-Class Microphysics Scheme (WSM6) (Hong et al., 2010) to simulate cloud physical processes, and the Yonsei University (YSU) PBL scheme (Lee et al., 2011) to parameterize turbulence in the planetary boundary layer. After simulation, habitat temperatures were extracted based on coordinates.

Habitat heterogeneity

To quantify vertical habitat heterogeneity, we surveyed the vertical coverage of vegetation at each sampling plot within 5-m radii. Within these circles, we classified vertical layers into understory (< 2 m), midstory (2–10 m), and overstory (> 10 m) vegetation. Coverage was classified into the following four categories: 0 (0 % coverage), 1 (1–33 % coverage), 2 (34–66 % coverage), and 3 (67–100 % coverage) (Lee et al., 2011; Rhim, 2012).

For horizontal habitat heterogeneity, we calculated the Shannon-Wiener diversity index (H') Using the area of that particular habitat type (abundance) and number of different habitat types (richness) (Turner & Gardner, 2015; Pan et al., 2016). The area and number of habitat types were extracted from land cover maps (Ministry of Environment, Republic of Korea) within a 150-m radius circle at each plot using ArcGIS 10.3 (ESRI, Redlands, CA, USA). The top categories of habitat type comprised anthropogenic, agricultural, managed, and natural forestry, herbaceous, swamp, barren, and water areas. A total of 15 habitat types were defined and used for the habitat diversity index.

Migrant rate

To calculate migrant rate, all birds detected were classified as residents or summer migrants. Migrants were defined as wintering in the tropical region of Southeast Asia and migrating to the study area for breeding purposes. Twenty-three species were identified as summer migrants and 30 species were defined as residents (Table S1). Generally, a function of the degree of species turnover is used to determine the change in community composition for purposes of measuring beta diversity (Wilson & Shmida, 1984; Harrison, Ross & Lawton, 1992; Lomolino, 2001). To test evolutionary history, the more complex interactions among immigration, emigration, and extinction should be considered. In the present study, we simply



used the migrant rate, which was calculated based on the ratio of the total number of species or individuals and the number of migrant species or individuals in each plot (Helle & Fuller, 1988; Newton & Dale, 1996), and transformed into the arcsine square root as a measurement variable.

Data analyses

To investigate the distribution patterns of breeding bird species richness and individuals along an altitudinal gradient, we used the curve estimation function in SPSS 20. Best-fit curves (linear, quadratic, and exponential) were selected according to the highest R^2 and significant p-values. Using the same method, we verified a linear relationship between the dependent variables (species richness and density) and independent variables (spring temperature, vertical coverage of vegetation, horizontal habitat diversity, and migrant rate).

We used model selection and multimodel inference using a generalized linear model (GLM). We developed a set of 16 candidate models using this GLM, using 16 combinations of variables to identify the causes of altitudinal patterns in bird species richness and density in relation to spring temperature, migrant rate, vertical coverage of vegetation, and horizontal habitat diversity variables. Before adding variables to the model selection, we transformed the relative humidity variable into its arcsine square root as a measurement variable, and replaced correlated predictors ($r \ge |0.7|$) with another variable. Once the models were created, we used information-theoretic methods (Burnham & Anderson, 2002) to choose from among the competing models by converting log-likelihood values calculated using Akaike's information criterion adjusted for small sample sizes (AICc) (Akaike, 1974) and Akaike weights (w_i). The high-confidence set of candidate models consisted of models with Akaike weights within 10% of the highest (Royall, 1997; Lepczyk et al., 2008), and these were used to compute model-averaged parameter estimates. All statistical analyses were performed using R 3.3.2 (packages bbmle, AICcmodavg, and MuMin).

Results

Altitudinal patterns in species richness and density

Fifty-three species were observed in the 142 survey plots during the breeding period surveyed, with a total of 2771 individual birds. To verify the altitudinal patterns in species richness and numbers of individuals, we estimated best-fit curves. Breeding bird species richness showed a hump-shaped pattern along an altitudinal gradient ($R^2 = 0.11$, p < 0.001; Fig. 2a). A linear pattern of species richness was not significant in relation to altitude ($R^2 = 0.00$, p = 0.820). In addition, density showed a hump-shaped pattern ($R^2 = 0.10$, p = 0.002; Fig. 2b), rather than a linear pattern ($R^2 = 0.04$, p = 0.019).

Relationships of species richness and density with different variables Single variable patterns

Pearson's correlation analysis of nine environmental variables showed that spring temperature and relative humidity were highly correlated (r = -0.951; Table S2). Elevation showed strong correlations with spring temperature and relative humidity (r = -0.977, r = 0.938, respectively; Table S2). Although migrant rates of species and individuals were correlated (r = 0.872; Table S2), these were not included in the same model. Therefore, elevation and relative humidity variables were eliminated from the curve estimation and model construction.

In the best-fit curve estimation between species richness, density, and environmental variables, species richness showed significant correlations with spring temperature ($R^2 = 0.08$, p = 0.003; Fig. 3a) and species migrant rate ($R^2 = 0.11$, p < 0.001; Fig. 3b), and were represented by hump-shaped curves. No relationships were observed between species richness and coverage of understory vegetation, midstory vegetation, or habitat diversity (Fig. 3c, d, and f). Species richness and coverage of overstory vegetation showed a significant positive correlation ($R^2 = 0.14$, p < 0.001; Fig. 3e). Moreover, density showed a significant correlation with spring temperature in a hump-shaped pattern ($R^2 = 0.11$, p < 0.001; Fig. 4a). A decreasing pattern was observed between density and migrant rate of individuals ($R^2 = 0.07$, p = 0.006; Fig. 4b), and coverage of under- and overstory vegetation represented a monotonically increasing pattern with increasing density ($R^2 = 0.03$, p = 0.027; $R^2 = 0.40$, p = 0.017; Fig. 4c and e). Other variables, including coverage of midstory vegetation and habitat diversity, did not show any significant correlations.

Model selection and multimodel inference

The set of candidate models with 16 combinations of six variables showed eight models supported for species richness and four models supported for density (Table 2). The best predictors of species richness were overstory vegetation, midstory vegetation, understory vegetation, and habitat diversity ($w_i = 0.336$). A model including habitat diversity was 9.1 times more likely to explain species richness better than models excluding it (w_i 0.336 vs. 0.037; Table 2). Overstory vegetation was included in all supported species richness models. The Akaike weight was 2.9 and 1.4 times higher without the inclusion of spring temperature and species migrant rate, respectively, than when these parameters were included (w_i 0.336 vs. 0.114, w_i 0.336 vs. 0.234, respectively; Table 2). The best model for predicting density included overstory vegetation, midstory vegetation, understory vegetation, habitat diversity, and migrant rate of individuals (w_i = 0.347). When migrant rate of individuals was included in the density model, the Akaike weight was 9.4 times higher than when migrant rate was eliminated from the model (w_i 0.340 vs. 0.036; Table 2), and 1.02 times higher in the absence of spring temperature (w_i 0.347 vs. 0.340; Table 2). Vertical coverage of vegetation variables was included in all supported models.

Multimodel-averaged parameter estimates of species richness, including the eight supported models, showed positive correlations with overstory vegetation, understory vegetation, and habitat diversity (p < 0.001, p = 0.039, p = 0.031, respectively; Table 3). Density showed



positive correlations with overstory vegetation and understory vegetation (p < 0.001, p < 0.001; Table 3) and a negative correlation with migrant rate of individuals (p < 0.001; Table 3).

236

237

Discussion

238

239

240

241

242

243

244

245

246

247

248

249

250

251

252253

254

255

Altitudinal patterns of species richness and density

Altitudinal patterns in breeding bird species richness and density showed a hump-shaped pattern (Fig. 2), as found in previous studies (Poulsen & Lambert, 2000; Lomolino, 2001; Ding et al., 2005; Pan et al., 2016). Four main altitudinal patterns of species richness have been identified for geographical features such as climate type (tropical, subtropical, arid, and temperate), latitude, longitude, landmass type (islands and continents), altitude, and spatial scale (local and regional), but no relationships between elevational species richness and altitude or latitude have been observed (McCain, 2009). Most of the previously studied areas, which demonstrated a mid-peak pattern of species richness, were located in the northern and eastern regions of Asia and consisted of mountain in arid zones (Ding et al., 2005; McCain, 2009; Acharya et al., 2011; Pan et al., 2016). Moreover, a study conducted in Eastern Himalaya showed a hump-shaped pattern (Ding et al., 2005). In contrast, studies from Southeast Asia have shown predominantly decreasing patterns (McCain, 2009)]. Best-fits to the null model generally showed more mid-peaks in local-scale studies than in regional studies (McCain, 2009). Our species richness pattern, found in East Asia and at a local scale, followed the most frequently identified hump-shaped pattern. Further, we identified a hump-shaped pattern of density. Density has been referred to in standardized methods but has not been reported in many studies (Lomolino, 2001) compared to species richness.

256257

258

259

260

261

262

263

264

265

266

267

268

269270

271272

Relationships of species richness and density with different variables

Previous studies have shown that vegetation cover and habitat diversity have strong positive relationships with species richness (MacArthur & MacArthur, 1961; Hurlbert & Haskell, 2003; Hurlbert, 2004; Acharya et al., 2011; Pan et al., 2016). Dense vertical vegetation coverage may play an important role in providing birds with breeding space, shelter, and food resources such as insects, which could contribute positively to bird species richness and density. According to the more-individuals hypothesis, density and energy use in communities is positively correlated with energy availability, and species richness can contribute to density (Goetz et al., 2007). Accordingly, our results showed increasing species richness and density with increasing vertical overstory and understory vegetation cover (Table 3), and species richness showed a significant positive relationship with density (Fig. S3). Further, the present study demonstrated that species richness was affected by horizontal habitat diversity, but density was not (Table 3). High habitat diversity can increase species richness due to niche partitioning and providing habitat edges (Best, Whitmore & Booth, 1990), but high habitat diversity does not necessarily indicate high habitat quality with ample food resources. Therefore, the lack of a relationship

between density and habitat diversity in this study might be because density increased with productivity and habitat quality (Hurlbert, 2004; Goetz et al., 2007). According to the habitat heterogeneity hypothesis (MacArthur & MacArthur, 1961; Pan et al., 2016), greater structural complexity in vegetation and more habitat types likely contributed to species richness in the present study. However, a larger number of habitat types did not influence the density.

We observed a negative relationship between density and migrant rate of individuals, and no relationship was observed between species richness and species migrant rate (Fig. 4b and Table 3). Based on the heterospecific attraction hypothesis (Mönkkönen et al., 1997; Mönkkönen, Forsman & Heterospecific, 2002), we predicted that the migrant rate would have a positive effect on migrant species richness and density, and that the migrant rate would increase with resident species richness and density. However, in the present study, a reduction in the individual migrant rate led to an increase in density. Additionally, migrant species richness and density showed an increasing pattern along the altitude gradient, whereas resident species richness and density showed a mid-peak pattern along the altitude gradient (Fig. S4). It is unlikely that the migrants could choose a mid-elevation with higher vegetation coverage than the residents could (Fig. S4). Migrant species and individuals did not positively influence species richness and density, and they were not attracted to resident species. Therefore, the heterospecific attraction hypothesis was not applicable along the altitude gradient surveyed in the present study.

No relationships were found between species richness or density and climatic factors (Table 3 and Table S2); however, a decreasing pattern of spring temperature along the altitudinal gradient was identified (Table S2). Numerous studies have shown a positive relationship between temperature and diversity (McCain, 2009). However, significantly stronger relationships between temperature and diversity can be found in humid mountain habitats than in dry mountain habitats (McCain, 2009). Furthermore, a negative relationship between density and climatic factors was found in a study conducted in Asia (Pan et al., 2016). Despite numerous studies on this phenomenon, the pattern has not been adequately explained (Currie et al., 2004; Rahbek et al., 2007). Most studies used the average annual temperature from the WorldClim database and conducted bird surveys across all seasons using considerably larger datasets that have constrained accuracy due to the sampling effort involved (Lomolino, 2001; Ding et al., 2005). However, in the present study, we used spring temperature values derived for micro-scale studies, and focused on breeding bird survey on a local scale in a short period in mixed and deciduous forest areas; this approach may have led to the variation in the findings. Another possible explanation is that birds are restricted more by habitat quality for chick rearing than by temperature during the breeding season.

A single variable analysis showed no significant relationships between species richness and understory vegetation or habitat diversity (Fig. 3c and f); however, a significant relationship was observed in the modeling approach. Additionally, no differences in density were observed either in the single variable or in the modeling approach. Null hypothesis testing, similar to a simple linear correlation, has been used in many ecological studies and is currently being used in many areas. However, almost ecological phenomena has been often represented by nonlinear and



multiple interaction among variables (Landis et al., 2013). For example, species should live at the proper temperature for the optimal thermal fitness during breeding season. But if there is no proper nesting resources, food and shelter, the species should choose a different habitat even the proper temperature for breeding. Consequentially, each variables does not affect the dependent variable, but multiple interactions of the variables. Therefore, alternative modeling approach is required for ecological studies and considered to be a more reliable method that avoiding uninformative, logical deficiencies and common misinterpretations of null hypothesis testing (Anderson, Burnham & Thompson, 2000; Mönkkönen, Forsman & Bokma, 2006). In order to understand the complex ecological phenomena, the use of multimodels is more reasonable and needs more efforts to clarify the relationship of the causative variables.

Conclusion

Trends in species richness showed hump-shaped patterns along altitudinal gradients and were related to vertical vegetation coverage and horizontal habitat diversity. In addition, trends in density also showed hump-shaped patterns, with density related to vertical vegetation coverage and migrant rate, but not to habitat diversity. No significant relationships were found between spring temperature and species richness or density. The results on species richness support the habitat heterogeneity hypothesis rather than the climate hypothesis, whereas those of species density do not support fully either hypothesis, and they were related to species richness and vertical vegetation coverage. The heterospecific attraction hypothesis was not applicable to the distribution of birds along the altitudinal gradient studied. Taken together, our findings indicate that management of vegetation cover would be an appropriate strategy for avian conservation in this region. To achieve a better understanding of the specific reasons for the distribution of birds along altitudinal gradients, further studies on the interactions among species related to niche and competition are required.

References

- 1. Acharya BK, Sanders NJ, Vijayan L, Chettri B. 2011. Elevational gradients in bird diversity in the Eastern Himalaya: an evaluation of distribution patterns and their underlying mechanisms. *PLoS One* 6:e29097.
- Adolfo G, Navarro S. 1992. Altitudinal distribution of birds in the Sierra Madre del Sur,
 Guerrero, Mexico. *Condor* 94:29-39.
- 3. Akaike H. 1974. A new look at the statistical model identification. *IEEE Transactions on Automatic Control* 19:716-723.
- 4. Allen AP, Brown JH, Gillooly JF. 2002. Global biodiversity, biochemical kinetics, and the energetic-equivalence rule. *Science* 297:1545-1548.
- 5. Anderson DR, Burnham KP, Thompson WL. 2000. Null hypothesis testing: problems, prevalence, and an alternative. *J Wildlife Manage* 64:912-923.
- Best LB, Whitmore RC, Booth GM. 1990. Use of cornfields by birds during the breeding season: the importance of edge habitat. *Am Midl Nat* 123:84-99.



- 352 7. Burnham KP, Anderson DR. 2002. Information and likelihood theory: a basis for model selection and inference. *Model selection and multimodel inference: a practical information-theoretic approach* 2:49-97.
- Burnham KP, Anderson DR, Huyvaert KP. 2011. AIC model selection and multimodel inference in behavioral ecology: some background, observations, and comparisons. *Behav Ecol Sociobiol* 65:23-35.
- Chamberlain D, Brambilla M, Caprio E, Pedrini P, Rolando A. 2016. Alpine bird distributions along elevation gradients: the consistency of climate and habitat effects across geographic regions. *Oecologia* 181:1139-1150.
- Currie DJ, Mittelbach GG, Cornell HV, Field R, Guégan JF, Hawkins BA, Kaufman DM,
 Kerr JT, Oberdorff T, O'Brien E, Turner JRG. 2004. Predictions and tests of
 climate-based hypotheses of broad-scale variation in taxonomic richness. *Ecol Lett* 7:1121-1134.
- Diamond J. 1988. Factors controlling species diversity: overview and synthesis. *Ann Missouri Bot Gard* 117-129.
- Ding TS, Yuan HW, Geng S, Lin YS, Lee PF. 2005. Energy flux, body size and density in relation to bird species richness along an elevational gradient in Taiwan. *Glob Ecol Biogeogr* 14:299-306.
- 370 13. Dingle H, Drake VA. 2007. What is migration? *Bioscience* 57:113-121.
- 371 14. Gaston KJ. 2000. Global patterns in biodiversity. *Nature* 405:220-227.
- Goetz S, Steinberg D, Dubayah R, Blair B. 2007. Laser remote sensing of canopy habitat
 heterogeneity as a predictor of bird species richness in an eastern temperate forest, USA.
 Remote Sens Environ 108:254-263.
- 375 16. Gwon JH, Sin MK, Kwon HJ, Song HK. 2013. A study on the forest vegetation of Jirisan National Park. *J Korea Soc Environ Restor Technol* 16:93-118.
- Harrison S, Ross SJ, Lawton JH. 1992. Beta diversity on geographic gradients in Britain.
 J Anim Ecol 61:151-158.
- Herzog SK, Kessler M, Bach K. 2005. The elevational gradient in Andean bird species richness at the local scale: a foothill peak and a high-elevation plateau. *Ecography* 28:209-222.
- Helle P, Fuller RJ. 1988. Migrant passerine birds in European forest successions in relation to vegetation height and geographical position. *J Anim Ecol* 57:565-579.
- Hong SY, Noh Y, Dudhia J. 2006. A new vertical diffusion package with an explicit treatment of entrainment processes. *Mon Weather Rev* 134:2318-2341.
- Hong SY, Lim KSS, Lee YH, Ha JC, Kim HW, Ham SJ, Dudhia J. 2010. Evaluation of the WRF double-moment 6-class microphysics scheme for precipitating convection. *Adv Meteorol.* DOI:10.1155/2010/707253.
- Hurlbert AH, Haskell JP. 2003. The effect of energy and seasonality on avian species richness and community composition. *Amer Nat* 16:83-97.
- 391 23. Hurlbert AH. 2004. Species-energy relationship and habitat complexity in bird



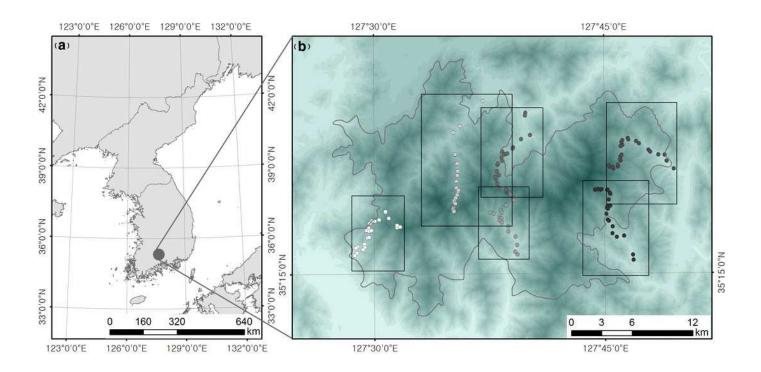
- 392 communities. *Ecol Lett* 7: 714-720.
- 393 24. Kain JS, Fritsch JM. 1993. Convective parameterization fir mesoscale models: The Kain-394 Fritsch scheme. In: Emanuel KA, Raymond DJ, editors. The representation of cumulus 395 convection in numerical models. *Boston: American Meteorological Society* 165-170.
- Landis WG, Durda JL, Brooks ML, Chapman PM, Menzie CA, Stahl RG, Stauber J L.
 2013. Ecological risk assessment in the context of global climate change. *Environmental toxicology and chemistry* 32(1):79-92.
- Lee PF, Ding TS, Hsu FH, Geng S. 2004. Breeding bird species richness in Taiwan:
 distribution on gradients of elevation, primary productivity and urbanization. *J Biogeogr* 31:307-314.
- 402 27. Lee EJ, Lee WS, Son SH, Rhim SJ. 2011. Differences in bird communities in postfire 403 silvicultural practices stands within pine forest of South Korea. *Landsc Ecol Eng* 7:137-404 143.
- Lepczyk CA, Flather CH, Radeloff VC, Pidgeon AM, Hammer RB, Liu J. 2008. Human impacts on regional avian diversity and abundance. *Conserv Biol* 22: 405-416.
- 407 29. Lomolino MV. 2001. Elevation gradients of species-density: historical and prospective views. *Glob Ecol Biogeogr* 10:3-13.
- 409 30. MacArthur RH, MacArthur JW. 1961. On bird species diversity. *Ecology* 42:594-598.
- 410 31. McCain CM. 2009. Global analysis of bird elevational diversity. *Glob Ecol Biogeogr* 411 18:346-360.
- 412 32. Mönkkönen M, Helle P, Niemi GJ, Montgomery K. 1997. Heterospecific attraction affects community structure and migrant abundances in northern breeding bird communities. *Can J Zool* 75:2077-2083.
- 33. Mönkkönen M, Forsman JT. 2002. Heterospecific attraction among forest birds: a review.
 416 *Ornithol Sci* 1:41-51.
- Mönkkönen M, Forsman JT, Bokma F. 2006. Energy availability, abundance, energy-use
 and species richness in forest bird communities: a test of the species-energy theory. *Glob Ecol Biogeogr* 15:290-302.
- 35. Newton I, Dale L. 1996. Relationship between migration and latitude among west European birds. *J Anim Ecol* 65:137-146.
- 422 36. Pan X, Ding Z, Hu Y, Liang J, Wu Y, Si X, Guo M, Hu H, Jin K. 2016. Elevational pattern of bird species richness and its causes along a central Himalaya gradient, China. *Peer J* 4:e2636.
- 425 37. Pianka ER. 1966. Latitudinal gradients in species diversity: a review of concepts. *Amer*426 *Nat* 100:33-46.
- 427 38. Poulsen MK, Lambert FR. 2000. Altitudinal distribution and habitat preferences of forest 428 birds on Halmahera and Buru, Indonesia: implications for conservation of Moluccan 429 avifaunas. *Ibis* 142:566-586.
- 39. Rahbek C. 2005. The role of spatial scale and the perception of large-scale speciesrichness patterns. *Ecol Lett* 8:224-239.



- 432 40. Rahbek C, Gotelli NJ, Colwell RK, Entsminger GL, Rangel TFL, Graves GR. 2007.
 433 Predicting continental-scale patterns of bird species richness with spatially explicit
 434 models. *Proc Royal Soc Lond B: Biol Sci* 274:165-174.
- 435 41. Reynolds RT, Scott JM, Nussbaum RA. 1980. A variable circular-plot method for estimating bird numbers. *Condor* 309-313.
- 42. Rhim SJ. 2012. Ecological factors influencing nest survival of hazel grouse *Bonasa* bonasia in a temperate forest, South Korea. *Forest Ecol Manag* 282:23-27.
- 439 43. Royall RM. 1997. Statistical evidence: a likelihood paradigm. London: Chapman and Hall.
- 44. Sanders NJ, Rahbek C. 2012. The patterns and causes of elevational diversity gradients.
 442 *Ecography* 35:1-3.
- 443 45. Stevens GC. 1992. The elevational gradient in altitudinal range: an extension of Rapoport's latitudinal rule to altitude. *Amer Nat* 140:893-911.
- 445 46. Terborgh J. 1977. Bird species diversity on an Andean elevational gradient. *Ecology* 446 58:1007-1019.
- 47. Turner MG, Gardner RH. 2015. Landscape ecology in theory and practice: pattern and process. 2nd edition. New York: Springer.
- 48. Wilson MV, Shmida A. 1984. Measuring beta diversity with presence-absence data. *J* 450 *Ecol* 72:1055-1064.



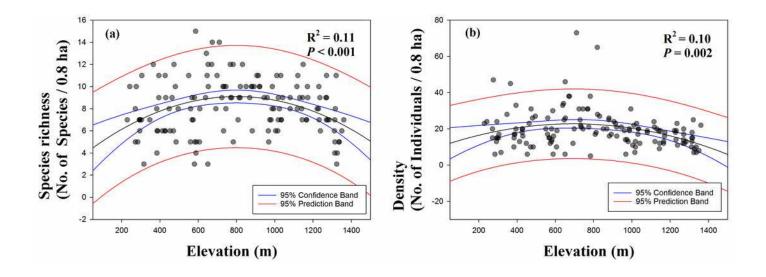
Location of Study site and survey plots





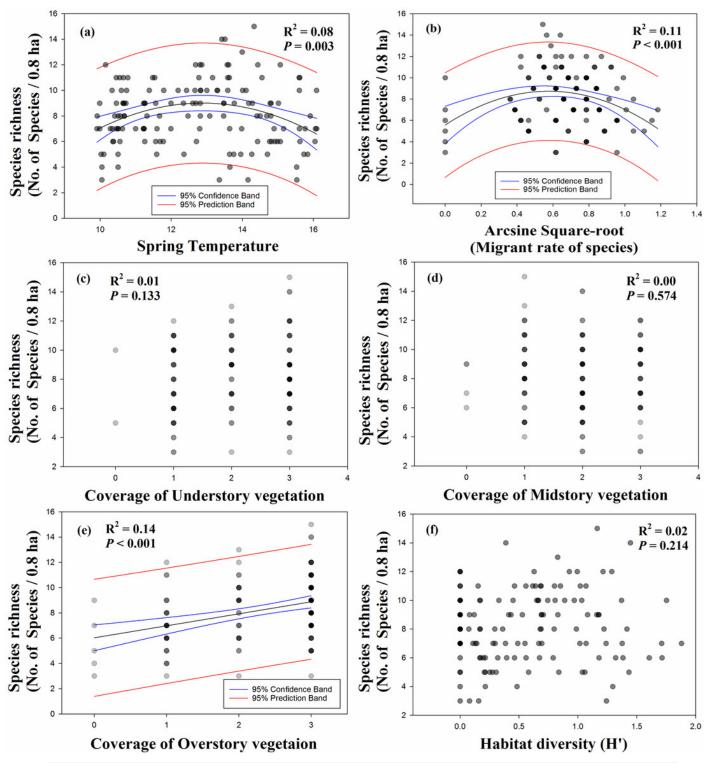
Distribution patterns of (a) species richness and (b) density along an altitudinal gradient.

Best-fit curves (linear, quadratic, and exponential) were selected according to the highest R^2 and significant p -values.





Singlevariable patterns using best-fit curve function between species richness and variables.



Singlevariable patterns using best-fit curve function between density and variables.

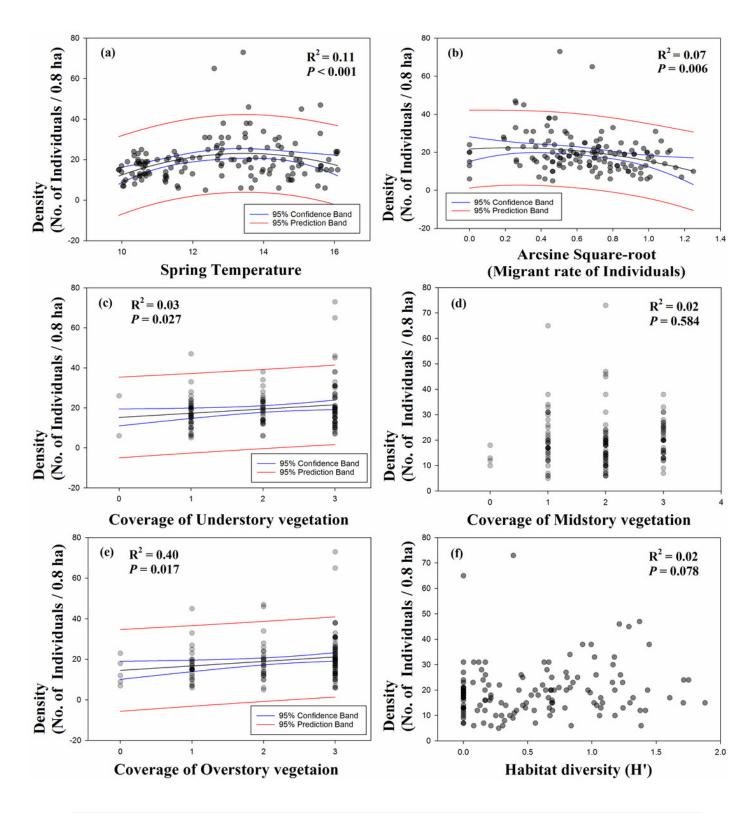




Table 1(on next page)

Studyplots along altitudinal gradient.



Elevation (m)	Latitude (°N)	Longitude (°E)	No. of plots
200–300	35°23'46.3"–35°15'09.3"	127°49'56.2"–127°28'02.9"	10
300–400	35°24'42.5"–35°16'07.1"	127°48'59.7"–127°28'03.6"	11
400–500	35°23′45.7″–35°16′04.7″	127°47′56.9″–127°29′03.3″	13
500-600	35°21′58.8″–35°17′01.9″	127°47′57.2″–127°29′02.5″	12
600–700	35°22′42.2″–35°17′04.6″	127°46′59.7″–127°29′03.5″	12
700–800	35°22′46.6″–35°17′08.2″	127°46′39.1″–127°29′00.5″	12
800–900	35°22′57.3″–35°18′00.1″	127°46′57.6″–127°29′04.0″	12
900–1000	35°21′50.4″–35°18′01.1″	127°45′56.2″–127°29′00.3″	12
1000-1100	35°21′37.9″–35°19′05.3″	127°45′56.9″–127°30′03.8″	12
1100–1200	35°20′59.1″–35°18′08.9″	127°45′54.3″–127°30′02.0″	12
1200-1300	35°20′50.8″–35°17′28.3″	127°45′38.4″–127°31′02.0″	12
1300-1400	35°20′51.2″–35°17′31.6″	127°45′34.8″–127°31′04.8″	12



Table 2(on next page)

Modelselection for predicting species richness and density according to springtemperature, migrant rate, vertical coverage of vegetation, and horizontalhabitat diversity.

Model selection for predicting species richness and density according to spring temperature, migrant rate, vertical coverage of vegetation, and horizontal habitat diversity.



Response variables	Candidate models	AICc	ΔAICc	df	Wi
Species richness (No. of species / 0.8 ha)	(Best model) Intercept + OV + MV + UV + HD	638.9	0	6	0.336
	Intercept + OV + MV + UV + HD + MRS	639.7	0.7	7	0.234
	Intercept + OV + MV + UV + HD + ST	641.1	2.2	7	0.114
	(Full model) Intercept + OV + MV + UV + HD + ST + MRS	641.8	2.9	8	0.08
	Intercept + OV + MV + UV + MRS	642.1	3.2	6	0.069
	Intercept + OV	642.2	3.2	3	0.067
	Intercept + OV + MV + UV + ST	642.3	3.4	6	0.062
	Intercept + OV +MV + UV	643.4	4.4	5	0.037
Density (No. of individuals / 0.8 ha)	(Best model) Intercept + OV + MV+ UV + HD + MRI	1038.1	0	7	0.347
	(Full model) Intercept + OV + MV + UV + HD + ST + MRI	1038.1	0	8	0.34
	Intercept + OV + MV + UV + MRI	1038.6	0.5	6	0.264
	Intercept + OV + MV + UV + ST	1042.6	4.5	6	0.036

Coverage of understory vegetation = UV, midstory vegetation = MV, overstory vegetation = OV, spring temperature = ST, habitat diversity = HD, migrant rate of species = MRS; migrant rate of individuals = MRI.



Table 3(on next page)

Resultsof AICc-based multimodel inference of species richness and density.



Parameter	Model-averaged	SE	<i>P</i> -value	Importance value	
1 at affected	estimates	SE.	1 -value		
Species richness					
Intercept	4.807	1.605	0.003 **	-	
Spring temperature	0.046	0.170	0.789	0.26	
Migrant rate of species	-1.272	0.993	0.204	0.38	
Understory vegetation	0.464	0.223	0.039 *	0.93	
Midstory vegetation	-0.113	0.235	0.632	0.93	
Overstory vegetation	1.089	0.222	< 0.001 ***	1.00	
Habitat diversity	0.948	0.435	0.031 *	0.76	
Density				•	
Intercept	6.053	10.838	0.578	-	
Spring temperature	1.043	0.699	0.138	0.38	
Migrant rate of individuals	-12.406	3.624	< 0.001 ***	0.96	
Understory vegetation	3.537	0.929	< 0.001 ***	1.00	
Midstory vegetation	0.086	0.951	0.929	1.00	
Overstory vegetation	3.452	0.886	< 0.001 ***	1.00	
Habitat diversity	1.832	2.084	0.383	0.70	

^{***}p < 0.001; **p < 0.01; *p < 0.05. SE = standard error.