

1 Where do we go from here: Challenges and the future of endocrine disrupting compound
2 screening and testing

3 Running head: Future challenges for EDC screening and testing

4 Vickie S. Wilson*[†], Gerald A. LeBlanc[‡], Seth Kullman[‡], Kevin Crofton[§], Patricia Schmieder[¶],
5 and Miriam Jacobs[#]

6 [†]U.S. Environmental Protection Agency, Office of Research and Development, National Health
7 and Environmental Effect Research Laboratory, Toxicity Assessment Division, Mail Code:

8 B105-04, Research Triangle Park, NC 27709; Telephone: 919-541-3559; Fax: 919-541-5541;

9 email: wilson.vickie@epa.gov; [‡]Toxicology Program, Department of Biological Sciences, North

10 Carolina State University, Raleigh, NC 27695-7633; email: gal@ncsu.edu; swkullma@ncsu.edu;

11 [§]U.S. EPA, ORD, NCCT, Research Triangle Park, NC; email: crofton.kevin@epa.gov; [¶]U.S.

12 EPA, ORD, NHEERL, Mid-Continent Ecology Division, Duluth, MN; email:

13 schmieder.patricia@epa.gov; [#]Centre for Radiation, Chemical and Environmental Hazards

14 Centre, PHE, UK; email: Miriam.Jacobs@phe.gov.uk.

15

16 * Corresponding Author:

17 Vickie Wilson

18 USEPA, Office of Research and Development, National Health and Environmental Effect Research
19 Laboratory, Toxicity Assessment Division, Mail Code: B105-04, Research Triangle Park, NC,

20 USA 27709;

21 email: wilson.vickie@epa.gov;

22 **ABSTRACT**

23 Worldwide concern about the impacts of endocrine disrupting compounds on both human
24 and environmental health has led to implementation of multiple screening and testing programs.
25 In most cases these programs have focused on impacts to the estrogen, androgen and thyroid
26 hormone (EAT) signaling pathways. The goal of the presentations in session five of the Society
27 of Environmental Toxicology and Chemistry (SETAC) North America Focused Topic Meeting:
28 Endocrine Disruption (February 4 – 6, 2014) was to discuss moving beyond EAT pathways to
29 address current challenges and identify future approaches for the expansion of screening and
30 testing programs. The session was chaired by Drs. Gerald A. LeBlanc and Vickie S. Wilson and
31 included five presentations. Dr. Gerald A. LeBlanc provided insight on non-EAT endocrine
32 targets that are known to be susceptible to endocrine disrupting compounds. Dr. Seth Kullman
33 gave an overview of emerging technologies that hold promise for the screening of chemicals for
34 interaction with EAT and other endocrine pathways. These were followed by two presentations
35 on the current status and future promise of computational (Dr. Kevin Crofton) and *in silico* (Dr.
36 Patricia Schmieder) approaches for screening and ranking chemicals for endocrine activity. Dr.
37 Miriam Jacobs culminated the session with an overview of the current understanding of the role
38 of epigenetics in endocrine regulation and approaches for evaluating chemicals for their ability to
39 disrupt the epigenetic regulation of endocrine processes.

40 Key words: Endocrine disruption, endocrine targets, EDC screening, *in silico*, epigenetics

41 INTRODUCTION

42 The ability of some exogenous chemicals to disrupt normal endocrine function has been
43 well demonstrated. In response, significant efforts are underway to develop screening and
44 testing strategies aimed at detecting and characterizing hazard resulting from exposure to
45 endocrine disrupting chemicals (EDCs). This effort is not trivial. The endocrine system consists
46 of an extensive network of signaling pathways with hundreds of potential targets of disruption.
47 Individual pathways within the network have evolved to respond to different chemical regulators
48 (hormones) and, accordingly, have diverse susceptibilities to exogenous chemicals. Species
49 differences in hormones and their receptors render interspecies extrapolations of chemical
50 susceptibilities challenging. Nonetheless, the endocrine system regulates diverse processes that
51 are critical to an individual's health and population sustainability (*e.g.*, development, growth,
52 metabolism, reproduction). Approaches and assays are urgently needed to screen and test
53 chemicals for such activities.

54 Efforts, to date, have largely focused upon three pathways within the endocrine network:
55 estrogen, androgen, and thyroid hormone (EAT) signaling with the utilization of a relatively
56 limited repertoire of assays and endpoints for each pathway. The goal of session five of the
57 Society of Environmental Toxicology and Chemistry (SETAC) North America Focused Topic
58 Meeting: Endocrine Disruption (February 4 – 6, 2014) was to chart a course for the sagacious
59 expansion of screening and testing efforts. The session was chaired by Drs. Gerald A. LeBlanc
60 and Vickie S. Wilson. Novel endocrine pathways, with known susceptibility of disruption by
61 exogenous chemicals were introduced by Dr. Gerald A. LeBlanc. Dr. Seth Kullman
62 complemented this presentation with a discussion of emerging technologies that hold promise for
63 the screening of chemicals for endocrine disrupting activity. Drs. Kevin Crofton and Patricia

64 Schmieder provided updates on advances in the use of computational and *in silico* approaches for
65 the screening and ranking of chemicals for endocrine disrupting activity. Lastly, Dr. Miriam
66 Jacobs provided an overview of epigenetic regulation of gene expression, the involvement of
67 epigenetics in endocrine regulation, the potential susceptibility of epigenetic mechanisms to
68 chemical interference, and assay approaches that could be applied to the screening and testing of
69 chemicals for epigenetic interactions.

70 This session, ideally, will stimulate discussion for a more inclusive EDC screening and
71 testing paradigm, utilizing technologies for the rapid and effective screening of chemicals, that
72 will prove more efficacious in identifying chemicals that can disrupt endocrine function at
73 relevant exposure levels and preventing such exposures from occurring.

74

75 **SESSION PRESENTATION SUMMARIES**

76 Emerging Targets of Endocrine Disruption: The Xenocrine System, by: Gerald A. LeBlanc.

77 The 2012 Organization for Economic Co-Operation and Development (OECD)
78 monograph entitled *State of the Science on Novel In Vitro and In Vivo Screening and Testing*
79 *Methods and Endpoints for Evaluating Endocrine Disruptors* (OECD 2012) describes results of
80 the formative task of identifying endocrine signaling processes that have known susceptibility to
81 disruption and assays for the evaluation of such disruption. Among the pathways recommended
82 for future priority consideration were the peroxisome proliferator activated receptor (PPAR),
83 vitamin D receptor (VDR), and the retinoic acid receptor (RAR) signaling pathways.
84 Noteworthy, is that none of these pathways utilize a classical hormone. Rather, activating
85 ligands tend to be of exogenous origin related to diet (i.e., PPAR – dietary fatty acids, VDR–

86 dietary vitamin D, RAR - dietary vitamin A). Furthermore, the three signaling pathways utilize a
87 common receptor, the retinoid X receptor (RXR), to partner with the primary receptor to form an
88 active dimeric transcription factor. RXR also can serve as an independent, ligand-activated
89 transcription factor.

90 The existence of a common ligand-binding component (RXR) to the three pathways
91 provides a scenario for possible inter-pathway interactions (Fig. 1). For example, a ligand to
92 RXR may simultaneously activate one signaling pathway, suppress another, and have no effect
93 on the third. Little is currently known of the toxicological consequences of such ligand-mediated
94 inter-pathway interactions. However, it is highly likely that RXR-binding xenobiotics would
95 elicit multiple effects by impacting the signaling of multiple RXR-dependent pathways. For
96 example, tributyltin is a potent RXR agonist and is known to elicit diverse effects spanning
97 reproductive and developmental abnormalities, immunosuppression, and adipogenesis.

98 The provision of two distinct ligand binding sites, one on each receptor protein subunit,
99 within the receptor complex provides the opportunity for intra-pathway interactions between
100 ligand to the primary receptor and ligand to the RXR (Fig. 1). For example, the PPAR ligand
101 clofibrate can independently activate the PPAR:RXR receptor complex, as can the RXR ligand
102 9-*cis* retinoic acid. In combination, these ligands can synergistically activate the receptor
103 complex. Such intra-pathway interactions provide a mechanism by which exposure to chemical
104 combinations can result in greater cellular responses than would be predicted by exposure to the
105 individual chemicals.

106 This signaling network, the “xenocrine system”, regulates various aspects of metabolism
107 and development. Disruptions in the network can have implications on several endemic disease
108 conditions in human populations (*e.g.*, obesity, diabetes, metabolic syndrome) and energy-

109 dependent processes in wildlife populations (*e.g.*, growth, development, and reproduction).
110 Potential disruptors of this network include chemicals that function as agonists/antagonists of the
111 primary receptors, agonists/antagonists of the RXR, modulators of multi-functional co-activators
112 that contribute to receptor activity, and modulators of intracellular receptor levels.

113
114 Emerging Technologies to Assess Endocrine Disruption: Overview of New Assays and
115 Approaches That Show Promise in Evaluating the Endocrine Activity of Chemicals, by: Seth
116 Kullman

117 This presentation examined novel technologies and approaches to assess endocrine active
118 chemicals. Specifically, this presentation related new advances in EDC function beyond standard
119 EAT pathways that are well represented within the EPA/OECD endocrine screening programs.
120 The foundation of the discussion was based upon a recent OECD detailed review paper (DRP)
121 *State of the Science on Novel In Vitro and In Vivo Screening and Testing Methods and Endpoints*
122 *for Evaluating Endocrine Disruptors* (OECD 2012). A short synopsis of the detailed review
123 document was provided highlighting DRP content and focus: 1) categorizing the DRP as a
124 literature-based document designed to identify assays/methods to assess for EDC activity and
125 function with an emphasis on non-EAT pathways, 2) demonstrating both *in vitro* and *in vivo*
126 assay that are sufficiently developed for standardization (not validated), 3) illustrating the
127 importance of neuro-endocrine axis and the key role of nuclear receptor binding and
128 transactivation as primary molecular initiating events and 4) emphasizing the value of
129 presenting pathways in the context of adverse outcome pathways that link molecular initiating
130 events with apical outcomes.

131 The thyroid hormone (TH) signaling pathway was presented as an example to illustrate

132 hormone axis complexity and putative sites of action for TH disruptors. Specific mechanisms of
133 chemical action were discussed including disruption in thyrotropin-releasing hormone and
134 thyroid-stimulating hormone synthesis and signaling, inhibition of iodine uptake into the
135 thyrocyte, synthesis of T4/T3 by thyroperoxidases, modification in hormone transport-blood
136 binding proteins, hepatic metabolism of T4/T3, disruption of deiodinase, alteration in cellular
137 uptake/excretion of thyroid hormones, and direct interaction of compounds with the TH and
138 RXR receptors as direct agonists or antagonists. Specific examples detailed novel assays to
139 assess chemical-target interactions and activities. Further attention was given to notable assays
140 developed to assess molecular initiating events associated with thyroid receptor (TR)
141 transactivation/repression including TR expression, protein-protein interactions between TR and
142 the nuclear receptor co-regulators (co-repressors, co-activators), TR-DNA binding, TR agonists
143 and antagonists, and heterodimerization of TR with RXR. To further detail specific mechanistic
144 inferences, TR-DNA binding interaction studies were presented as an example highlighting
145 novel assays to assess chemicals that may disrupt fundamental nuclear receptor (NR) function.
146 Results from two publications were discussed (Ibhazehiebo et al., 2011, Iwasaki et al., 2008)
147 demonstrating the utility of a NR:HRE association/dissociation assay to inform NR-DNA
148 interactions. Results of assay performance were described in context of specific chemical agents
149 that disrupt TR signaling through this mechanism.

150 Select assays from the literature were collated to provide a larger context and to illustrate
151 their utility in establishing an adverse outcome pathway (AOP) for thyroid disrupting compounds
152 and how linking exposure impacts and molecular initiating events to biological outcomes at
153 higher levels of biological organization could inform disruption of the thyroid axis. Several
154 assays were discussed in context of their potential to inform macro-molecular interactions,

155 cellular responses, tissue/organ level responses or organism level responses. Discussion
156 emphasized the ability of AOP's to effectively translate and link mechanistic information at the
157 molecular level (*i.e.* chemical target) to meaningful apical endpoints at the organismic level such
158 as neurobehavioral function.

159 Application of high throughput screening (HTS) technologies was additionally discussed
160 in context of how current HTS programs including ToxCast and Tox21 could inform endocrine
161 disruption. Multiple *in vitro* cell-based assays were highlighted with a focus on nuclear receptor
162 transactivation including the *Attagene* Factorial Trans reporter assay, *NCGC* β -lactamase
163 reporter gene assay and competitive receptor binding assays. Attention was also drawn to the
164 need to incorporate *in vivo* translational models for validation of cell-based HTS assays. Several
165 novel assays exploiting the use of endocrine sensitive transgenic reporter lines in zebrafish and
166 medaka were demonstrated. These systems emphasized the utility of incorporating comparative
167 *in vivo* approaches that leverage the advantages of alternative models in combination with
168 emerging high-throughput and computational technologies that may facilitate extrapolation of
169 chemical action to human toxicity studies.

170 The conclusion reiterated the importance of establishing and validating assays that inform
171 EDC function and activity beyond EAT. Emphasis was placed on the ability to incorporate
172 assays into defined adverse outcome pathways with the ability to identify novel molecular
173 initiating events and link these activities to apical endpoints. The importance of integrating high
174 throughput assays informative of EDC function was also discussed in context of developing a
175 global assessment of endocrine active compounds. And finally, the necessity to utilize
176 translational *in vivo* models was also highlighted to better establish means to extrapolate cell-
177 based data to human toxicity.

178

179 Using High-Throughput Methods to Conduct Risk-based Prioritization of Chemicals for the
180 EDSP, by: Kevin Crofton

181 A major challenge for regulatory authorities is assessing risk for the large universe of
182 untested chemicals for which there is human or ecological exposure. This chemical universe has
183 been estimated to be 30K-50K unique substances. In particular, the U. S. EPA Endocrine
184 Disruption Screening Program (EDSP) is required to test 5K-10K chemicals due to pesticidal use
185 or their potential to contaminate drinking water. The U.S. EPA Computational Toxicology
186 program is developing methods and models to prioritize this large chemical universe for further
187 testing based on estimation of risk. The prioritization framework includes consideration of
188 hazard, exposure and dosimetry. Hazard estimation combines *in vitro* high-throughput screening
189 (HTS) assays plus QSAR and docking models. The U. S. EPA's ToxCast program, together with
190 the U.S. Interagency Tox21 program, has generated data for over 8,500 chemicals using a variety
191 of HTS assays for endocrine activity (*e.g.* ER, AR, TR). Additionally, a large-scale
192 multinational effort is developing and evaluating QSAR and docking models to use and extend
193 the HTS data, initially for the estrogen receptor. Results of this modeling effort will, in turn,
194 drive further *in vitro* testing. U. S. EPA's ExpoCast program is developing quantitative exposure
195 prediction models based on chemical properties and use patterns. These models allow rapid
196 estimates of exposure potential for thousands of chemicals. Finally, for dosimetry, we are using
197 a combined *in vitro* and modeling approach (called Reverse Toxicokinetics (RTK)) to provide
198 quantitative estimates of concentration-to-dose scaling. By combining quantitative *in vitro*
199 potency estimates from ToxCast, concentration-to-dose scaling from RTK, quantitative exposure
200 values from ExpoCast, and estimates of uncertainty, we can provide quantitative risk metrics at

9

201 the pathway level (estrogen, androgen, thyroid) for hundreds to thousands of chemicals. The first
202 use of these estimates will be in prioritizing chemicals for inclusion in the EDSP Tier 1 assay
203 battery.

204

205 Developing In Silico Approaches to Chemical Prioritization for ED Testing Within an AOP
206 Context, by: Patricia Schmieder

207 There are several key things the U.S. EPA must consider in implementing the Endocrine
208 Disruptor Screening Program (EDSP), particularly with regard to the use of predictive tools to
209 wisely screen chemicals and identify which are most likely to initiate toxicity due to interference
210 with endocrine pathways. The presentation, summarized herein, described important aspects of
211 building a predictive tool specifically for prioritization of the EPA's EDSP chemical universe of
212 ~10,000 chemicals (US EPA 2012). Key considerations in tool development are clarity of the
213 programmatic application of the tool; (*i.e.*, to determine the order in which EDSP universe
214 chemicals are to be screened) and defining the domain of application where the regulatory
215 decisions are needed so that the tool is developed to cover those types of chemicals. The OECD
216 principles for QSAR validation serve as a guide to ensuring scientific soundness, transparency
217 and maximum utility of the tool (US EPA 2012, OECD 2007). An estrogen receptor expert
218 system (ERES) developed for ED chemical prioritization following these guidelines was
219 presented to an EPA Scientific Advisory Panel (SAP) in 2009 (US EPA, 2009) and again in 2013
220 (US EPA 2013a) as it was expanded to cover the EDSP chemical universe. The ERES was
221 developed in the context of an ER-mediated adverse outcome pathway (AOP) (Hornung et al
222 2014, Schmieder et. al 2004, Schmieder et. al 2014, US EPA 2009, and US EPA 2013a) using *in*
223 *vitro* assays at the molecular and tissue levels of biological organization (see Figure 1 in

224 Schmieder et. al 2004). The *in vitro* assays used to develop the ERES were specifically
225 optimized to identify the types of low affinity binding that are characteristic of EDSP chemicals,
226 whereas off-the-shelf assays developed for different types of chemicals and a different purpose
227 were found to be limited (Schmeider 2004, US EPA 2013a, US EPA 2013b).

228 An extensive database was developed using the *in vitro* rainbow trout ER binding and
229 liver slice gene activation assays (Hornung et al 2014, Schmeider 2004) which are considered the
230 gold standard in detecting ER activation for this AOP (US EPA 2013b). Once the chemical
231 universe for tool application was defined, a systematic process was used to develop the database
232 by testing the boundaries of where chemical-ER binding occurs and where there is no binding.
233 By striving for a mechanistic understanding of ER binding grounded in theory and confirmed in
234 practice, it was possible to define seven major ‘effects-based chemical categories’ (Figure 2; I to
235 VII) where the effect in question is binding to the ER and potentially initiating an ER-mediated
236 adverse outcome pathway (Hornung et al 2014, Schmieder et. al 2004, Schmieder et. al 2014, US
237 EPA 2009, US EPA 2013a). The coding of logic rules in a decision tree was automated into an
238 ERES that not only provides a prediction of ER binding potential but also allows the user to
239 examine the empirical evidence within an effects-based chemical category to which an untested
240 chemical is assigned based upon its similar chemical features and properties pertinent to ER
241 binding that the untested chemical shares with the training set chemicals (US EPA 2013a, US
242 EPA 2013b). Applying the ERES to the initial chemical inventories it was built to prioritize
243 resulted in only ~5% of the chemicals being flagged as potential ER binders (US EPA 2009).
244 The expansion of the ERES to cover the discrete chemicals in the more recently defined EDSP
245 universe (OECD 2007) was demonstrated for the 2013 SAP review (US EPA 2013a) where
246 again only 5-10% of the universe is prioritized for further examination based upon potential to

247 initiate an ER-mediated AOP. The approach, found to be transparent, scientifically sound, and
248 relevant to EPA needs by SAP panels (US EPA 2009, US EPA 2013a) and an OECD expert
249 consultation (OECD, 2009) is equally applicable to building prioritization tools for additional
250 ED pathways (US EPA 2013b).

251

252 Endocrine Disruptors and the Epigenome: Potential Regulatory Applications for Chemical Safety
253 Testing, by: Miriam N. Jacobs

254 Epigenetic modulations underlie critical developmental processes and contribute to
255 determining adult phenotype. Alterations to the phenotype, due to exposure to environmental
256 insults during sensitive periods of development, are mediated through alterations in epigenetic
257 programming in affected tissues. This presentation evaluated the potential role of chemical-
258 induced epigenetic modifications to endocrine signaling pathways during sensitive windows of
259 exposure as a mechanism of endocrine disruption, and examined potential methods for assessing
260 such disruption, building upon the Annex to OECD 2012 and Greally and Jacob (2013) including
261 table references therein. Potential targets of disruption along putative AOPs associated with the
262 signaling pathways were identified as were assays that showed promise in evaluating the target
263 in a screening and testing program. *In vitro* methods are used where possible, and animal
264 experiments are used only when *in vitro* methods were inadequate. Monitoring such epigenetic
265 changes in response to toxicant exposure may provide a valuable tool for predicting adverse later
266 life outcomes. Indeed, preliminary work in which embryonic stem cells (ESCs) have been
267 differentiated into primordial germ cells (PGCs) in the presence of 17-beta estradiol and dibutyl
268 phthalate suggests that altered microRNA expression centered around PPAR alpha and the
269 reproductive system may be involved in such chemically-induced epigenetic toxicity. Unlimited

270 stocks of clonal PGCs can be differentiated from ESCs in the presence of environmental
271 chemicals. These clones can be used to investigate mechanisms of chemically-induced epigenetic
272 changes that may lead to altered phenotypes in unexposed future generations. MicroRNAs are
273 important epigenetic regulators that might mediate the adverse effects of environmental
274 chemicals across multiple generations and the in vitro PGC system is a valuable tool for
275 investigating such mechanisms of multigenerational epigenetic inheritance.

276 A more robust basis for Test Guideline (TG) recommendations, however, is still needed.
277 Although there is evidence to suggest that epigenomic dysregulation might mediate effects of
278 exposures to endocrine disruptors, it is uncertain as to whether these changes are truly predictive
279 of adverse outcome(s). Adverse effects observed in the OECD transgenerational assays could be
280 used to inform future tests specifically designed to investigate the epigenetic mechanism of
281 action. Collection and identification of RNA/DNA for genomic and epigenetic analysis from the
282 higher level tests would assist in the mechanistic understanding of the epigenetic contributions to
283 an adverse outcome pathway as well as making better use of the animal tissues. Follow-up
284 studies should include both an epigenetic as well as a genomic component to differentiate
285 between the contributions of potentially compensatory mechanisms.

286 Other particularly pertinent test method models to develop include the zebrafish. The use
287 of zebrafish for studies of reproductive and developmental screening of endocrine active
288 substances is now well-established. Zebrafish appear to represent a suitable model organism
289 available for studies replicating the effects of endocrine active substances and endocrine
290 disruptors on vertebrates and aquatic wildlife. This model also has great potential for rapid,
291 reliable, and less expensive exploration of the role of epigenetics, aging, senescence, and cancer
292 outcomes in relation to endocrine endpoints. This area is increasingly important to address in

293 regulatory toxicology, but the current higher level (level 5) *in vivo* TGs are unable to address it
294 for reasons of high cost, extended time, and the humane concerns related to the extension of such
295 tests.

296 There are clearly well supported reasons for further development of the zebrafish model
297 which has already been successfully applied by many laboratories. It could be further developed
298 to specifically address epigenetic endpoints in relation to endocrine activity and phenotypic
299 consequences in the model, to assess the quantitative and predictive capabilities for later adverse
300 outcomes. It might also be a useful model to assist in the discussion on the treatment of
301 functional genomics in TGs. A better understanding of the mechanisms and consequences of
302 epi-mutations is vital for assessing the risk of environmental human exposures.

303 **CONCLUSION**

304 In summary, the topics presented during session five of the SETAC Focus Topic
305 Meeting: Endocrine Disruption highlighted important factors to consider as we move forward
306 with the screening and testing of endocrine disrupting compounds. While the traditional focus of
307 these screens has been on classic EAT pathways, it is clear that other endocrine signaling
308 processes are susceptible to disruption. Utilization of assays that address such disruption may be
309 relevant in understanding xenobiotic contributions to adverse conditions in human and wildlife
310 populations. Furthermore, the inclusion of novel assays that provide additional information on
311 the EAT pathways should be considered, such as those that disrupt nuclear receptor function.
312 Other assays that can evaluate the effects of chemicals on the epigenome, particularly those
313 involving zebrafish, should also be considered as the role of epigenetic modulations becomes an
314 area of increasing importance. In addition, modifications to existing assays to include genomic-

315 and epigenetic-relevant endpoints can further expand the information that can be obtained from
316 animal tissues used in higher level tests.

317 As the number of suggested screening assays increases to address these new potential
318 targets, the need to devise rapid screening methods becomes increasingly important. Thoughtful
319 and strategic utilization of high throughput screening assays in conjunction with alternative
320 models and computational technologies could add significant value to the overall effectiveness of
321 this approach and provide a method for rapidly screening hundreds to thousands of chemicals;
322 and framing these data within the context of adverse outcome pathways will provide the most
323 effective approach in the prioritization of chemicals. Therefore, the utilization of transparent,
324 scientifically sound, and relevant tools to assess these pathways will be key as we move forward
325 in this effort.

326

327 **ACKNOWLEDGMENTS**

328 The authors would like to thank the following collaborators: John M Greally from Albert Einstein
329 College of Medicine and Emma L Marczyklo from Public Health England. The authors would also
330 like to thank Drs Justin Conley and Mary Ann Ottinger for providing technical reviews of this
331 manuscript

332 Disclaimer: This manuscript has been reviewed in accordance with the policy of the National
333 Health and Environmental Effects Research Laboratory, U.S. Environmental Protection Agency,
334 and approved for publication. Approval does not signify that the contents necessarily reflect the
335 views or policy of the Agency nor does mention of trade names or commercial products constitute
336 endorsement or recommendation for use.

337

338 **REFERENCES**

339 Greally JM, Jacobs MN. 2013. *In vitro* and *in vivo* testing methods of epigenomic endpoints
340 for evaluating endocrine disruptors. *ALTEX* 30:445-71.

341 Hornung MW, Tapper MA, Denny JS, Kolanczyk RC, Sheedy BR, Hartig PC, Aladjov H,
342 Schmieder PK. 2014. Effects-Based Chemical Category Approach for Prioritization of
343 Low Affinity Estrogenic Chemicals. *QSAR/SAR in Environ Res*, 25:289-323.

344 Ibhazehiebo K, Iwasaki T, Kimura-Kuroda J, Miyazaki W, Shimokawa N, Koibuchi N. 2011.
345 Disruption of Thyroid Hormone Receptor–Mediated Transcription and Thyroid
346 Hormone–Induced Purkinje Cell Dendrite Arborization by Polybrominated Diphenyl
347 Ethers. *Environmental Health Perspectives* 119:168-175.

348 Iwasaki T, Miyazaki W, Rokutanda N, Koibuchi N. 2008. Liquid chemiluminescent DNA pull-
349 down assay to measure nuclear receptor-DNA binding in solution. *BioTechniques*
350 45:445-448.

351 [OECD] Organisation for Economic Co-Operation and Development. 2007. Guidance
352 Document on the Validation of (Quantitative) Structure-Activity Relationships
353 [(Q)SARs] Models, Series on Testing and Assessment. No. 69 (2007).
354 ENV/JM/MONO(2007)2.

355 [OECD] Organisation for Economic Co-Operation and Development. 2009, Report of the
356 Expert Consultation to Evaluate an Estrogen Receptor Binding Affinity Model for Hazard
357 Identification. Series on Testing and Assessment, No. 111 (2009),
358 ENV/JM/MONO(2009)1.

359 [OECD] Organisation for Economic Co-Operation and Development. 2012. Health and Safety
360 Publications Series on Testing and Assessment, Detailed Review Paper State of the
361 Science on Novel *In Vitro* and *In Vivo* Screening and Testing Methods and Endpoints for
362 Evaluating Endocrine Disruptors. ENV/JM/MONO(2012)23. No. 178.

363 [OECD] Organisation for Economic Co-Operation and Development. 2014. Detailed Review
364 Paper on the State of the Science on Novel In Vitro and In Vivo Screening and Testing

- 365 Methods and Endpoints for Evaluating Endocrine Disruptors. OECD Series on Testing
366 and Assessment, No. 178, OECD Publishing, Paris. DOI:
367 <http://dx.doi.org/10.1787/9789264221352-en>
- 368 Schmieder PK, Tapper MA, Denny JS, Kolanczyk RC, Sheedy BR, Henry TR, Veith GD. 2004.
369 Use of trout liver slices to enhance mechanistic interpretation of estrogen receptor
370 binding for cost-effective prioritization of chemicals within large inventories. *Environ.*
371 *Sci. Technol* 38:6333-6342.
- 372 Schmieder PK, Kolanczyk RC, Hornung MW, Tapper MA, Denny JS, Sheedy BR, Aladjov H.
373 2014. A Rule-based Expert System for Chemical Prioritization Using Effects-based
374 Chemical Categories. *QSAR/SAR in Environ Res*, 25:253-287.
- 375 [USEPA] U.S. Environmental Protection Agency. 2013b. Prioritization of the Endocrine
376 Disruptor Screening Program Universe of Chemicals for an Estrogen Receptor Adverse
377 Outcome Pathway Using Computational Toxicology Tool. FIFRA Science Advisory
378 Panel Meeting, Arlington, VA, 2013, Panel Report listed as 'Meeting Minutes'.
379 Available at <https://www.regulations.gov/document?D=EPA-HQ-OPP-2012-0818-0037>.
- 380 [USEPA] U.S. Environmental Protection Agency. 2009. An Effects-based Expert System to
381 Predict Estrogen Receptor Binding Affinity for Food Use Inert Ingredients and
382 Antimicrobial Pesticides: Application in a Prioritization Scheme for Endocrine Disruptor
383 Screening. FIFRA Science Advisory Panel Meeting, Arlington, VA, 2009. White paper
384 listed as 'Meeting Materials'. Available at
385 <https://www.regulations.gov/document?D=EPA-HQ-OPP-2009-0322>
- 386 [USEPA] U.S. Environmental Protection Agency. 2012. EDSP Universe of Chemicals and
387 General Validation Principles, November 2012.
388 [https://www.epa.gov/sites/production/files/2015-](https://www.epa.gov/sites/production/files/2015-07/documents/edsp_chemical_universe_and_general_validations_white_paper_11_12.pdf)
389 [07/documents/edsp_chemical_universe_and_general_validations_white_paper_11_12.pdf](https://www.epa.gov/sites/production/files/2015-07/documents/edsp_chemical_universe_and_general_validations_white_paper_11_12.pdf)
- 390 [USEPA] U.S. Environmental Protection Agency. 2013a. Prioritization of the Endocrine
391 Disruptor Screening Program Universe of Chemicals for an Estrogen Receptor Adverse
392 Outcome Pathway Using Computational Toxicology Tool. FIFRA Science Advisory

393 Panel Meeting, Arlington, VA, 2013, White paper listed as ‘Meeting Materials’.
394 Available at <https://www.regulations.gov/document?D=EPA-HQ-OPP-2012-0818-0017>
395 West JA, Park I, Daley GQ, Geijsen N. 2006. *In vitro* generation of germ cells from murine
396 embryonic stem cells. *Nature Protocols* 1:2026-2036. doi:10.1038/nprot.2006.303.
397
398

399 Figure 1. Pleiotropic consequences of RXR ligand binding. Xenobiotic ligand binding (triangle)
400 to the RXR subunit may stimulate differential dimerization with partner receptor subunits (A)
401 resulting in the differential activation of signaling pathways. In addition, combinations of RXR
402 ligands with partner receptor subunit ligands (circles) can result in multiple intra-pathway
403 regulatory activities (B).

404

405

406

407

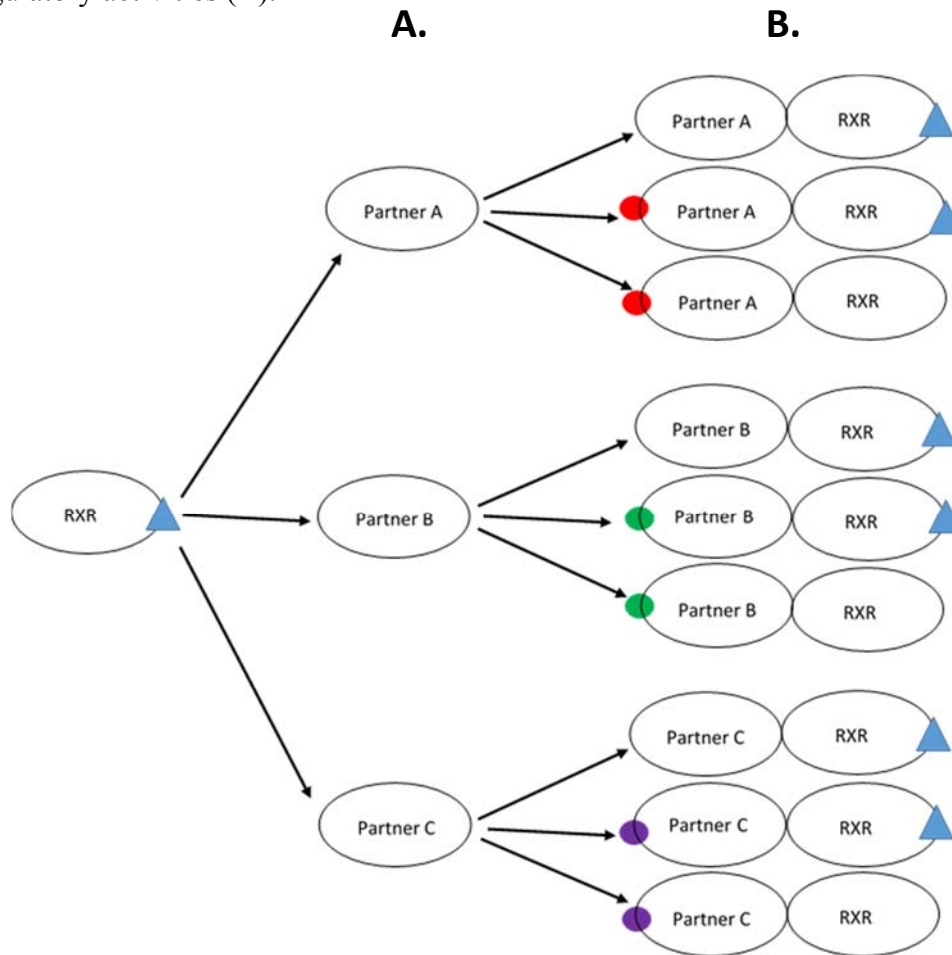


Figure 2. The rule-based ERESv1 for predicting binding potential for low affinity chemicals to the rER. The decision tree contains seven major nodes and multiple effects-based chemical categories within the nodes. The ERES is built on rainbow trout ER binding data using the cyto rER $\alpha\beta$ in combination with rainbow trout tissue slice gene activation data.

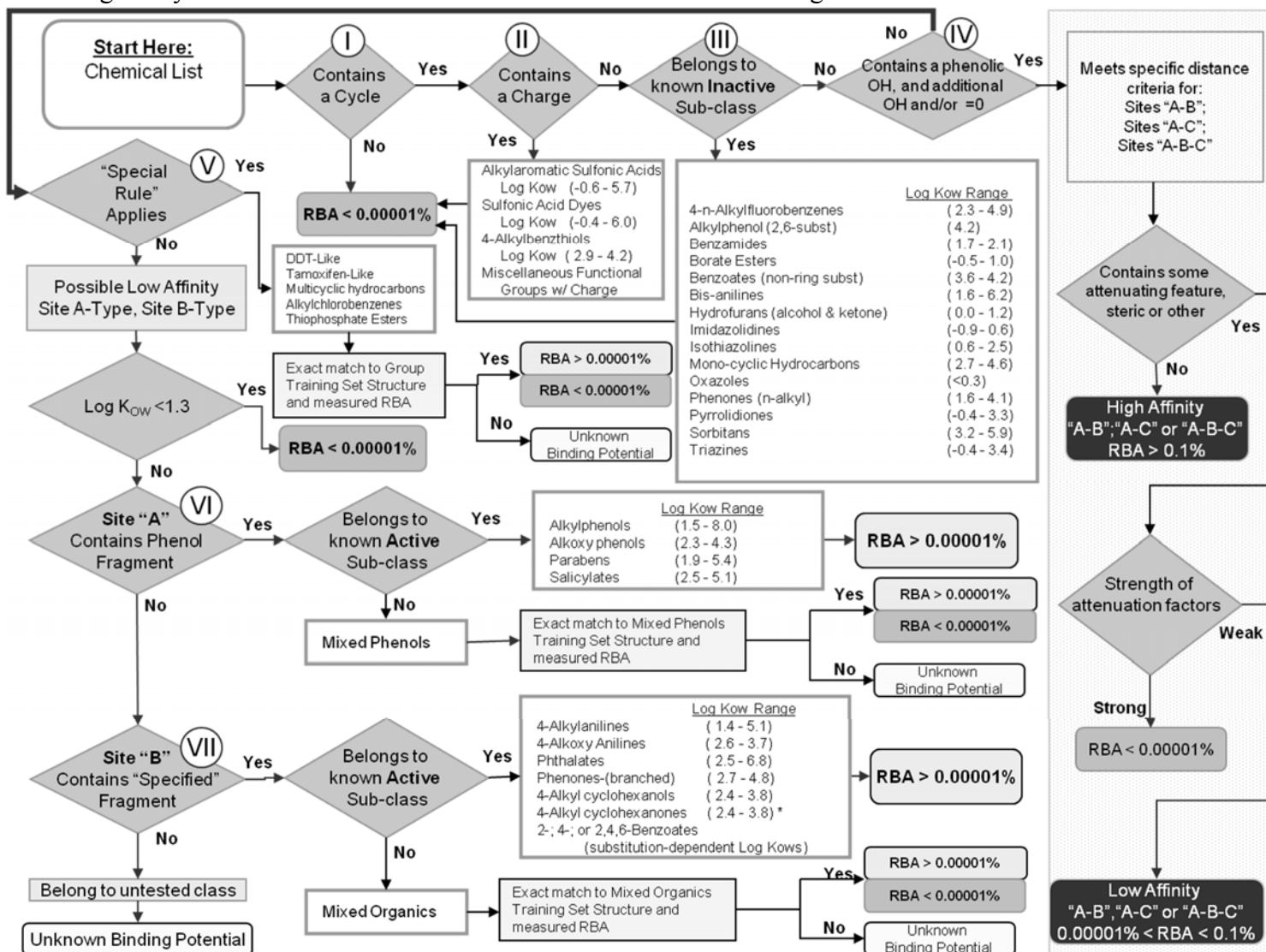


Figure 3. Current OECD (and related stakeholder activities) Workplan Timeline for Future of Endocrine Disruptor Screening and Testing

