

1 **Novel degradation pathway of 2-chloro-4-aminophenol in *Arthrobacter* sp. SPG**

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Abstract

A novel degradation pathway of 2-chloro-4-aminophenol (2C4AP) was studied in an *Arthrobacter* sp. SPG that utilized 2C4AP as a sole carbon and energy source. The 2C4AP degradation was initiated by a 2C4AP-deaminase that catalyzed the conversion of 2C4AP into chlorohydroquinone (CHQ) with removal of ammonium ion. In the next step, a CHQ-dehalogenase catalyzed the dehalogenation of CHQ to hydroquinone (HQ). The further degradation of HQ was proceeded with ring cleavage and the formation of γ -hydroxymuconic semialdehyde. This is the first report of degradation of 2C4AP by any bacteria.

45 **Introduction**

46 Chloroaminophenols (CAPs) are used as hair dyes and have been identified as toxic substances
47 because of their carcinogenicity [1]. Examples are 2-chloro-4-aminophenol (2C4AP) and 4-
48 chloro-2-aminophenol (4C2AP).

49 CAPs are released into the environment because of anthropogenic activities. They have been
50 released into soil and water as by-products during the synthesis of cosmetic dyes and chemicals.
51 They may also release into soil because of microbial degradation of various chemicals [2,3]. The
52 4C2AP may release into the soil due to degradation of 3-chloronitrobenzene and 4-chloro-2-
53 nitrophenol [2,3].

54 Few bacteria capable of utilizing CAPs as the sole carbon and energy source have been
55 identified and characterized [1, 4]. Examples are *Exiguobacterium* sp. PMA [4] and *Burkholderia*
56 sp. RKJ 800 [1]. Both of the strains utilized 4C2AP as the sole source of carbon and energy. The
57 complete mineralization of 4C2AP was studied in *Burkholderia* sp. RKJ 800 that degraded it with
58 release of ammonium and chloride ions [1]. The 4C2AP degradation was initiated with the
59 formation of 4-chlorocatechol that was further degraded via ring cleavage [1].

60 The bacterial degradation of CAPs may be initiated by one of the following mechanisms: (i) the
61 removal of ammonium ion from a CAP by a deaminase [1]; (ii) the ring cleavage of a CAP by
62 an aminophenol dioxygenase [5]; (iii) dehalogenation of a CAP by a dehalogenase [4] (iii)
63 acetylation of a CAP [2,3,6]

64 In this communication, we have reported degradation of 2C4AP by *Arthrobacter* sp. SPG. Strain
65 SPG was previously isolated from the soil collected from a pesticide contaminated site, India by
66 an enrichment method using 4-nitrophenol [7]. Strain SPG utilized 4-nitrophenol, 2-chloro-4-

67 nitrophenol (2C4NP), 2-nitrobenzoate, 3-methyl-4-nitrophenol and nitrocatechol as the sole
68 carbon and energy sources [7]. In addition, *Arthrobacter* sp. SPG is also capable of utilizing
69 2C4AP as a sole carbon and energy source. The aim of this study is to study of the degradation
70 pathway of 2C4AP by strain SPG.

71 **Material and Methods**

72 **Chemicals**

73 2C4AP, Hydroquinone and Chlorohydroquinone were purchased from Sigma-Aldrich. All other
74 chemicals used were of high quality graded.

75 **Growth and Degradation studies**

76 Strain SPG was grown on 500 ml Erlenmeyer flask containing 100 ml minimal medium and 0.3
77 mM 2C4AP as a carbon and energy source. The composition of minimal media was exactly same
78 as described previously [1]. For the growth studies, samples were collected at regular intervals
79 and the growth was measured taking the absorbance at 600 nm. For degradation studies, samples
80 were centrifuged and extracted with ethyl acetate and the extracted samples were dissolved in the
81 methanol and analyzed by the high performance liquid chromatography by previously described
82 method [1].

83 **Release of Chloride and ammonia**

84 Chloride and ammonia ions were analyzed by the previously described method [1].

85 **Identification of metabolites**

86 Strain SPG was grown on 500 ml Erlenmeyer flask containing 200 ml minimal media and 0.3
87 mM 2C4AP as a carbon and energy source. Samples (50 ml) were collected at regular intervals

88 (0h, 8h, 16h, 24h) and centrifuged. The supernatant was extracted with ethyl acetate and the
89 extracted samples were analyzed using the high performance liquid chromatography (HPLC) and
90 gas chromatography-mass spectrometry (GC-MS) by the previously described method [1]. GC-
91 MS analysis was carried out using a GC-MS-QP5000 instrument (Shimadzu, Tokyo, Japan)
92 equipped with quadrupole mass filter and DB-1 capillary column with ionization of 70 eV, scan
93 interval 1.5 s and mass range of 50–300 m/z. The column temperature was initially increased
94 from 50°C to 190°C at the rate of 5°C/min and then from 190°C to 280°C at the rate of
95 10°C/min. The carrier gas (nitrogen) flow rate was 15 ml/min [1].

96 **Preparation of crude extracts for enzyme assays**

97 For, the preparation of the cell extracts, strain SPG was grown in 500 ml minimal media
98 containing 0.3 mM 2C4AP and 20 mM sodium succinate. Bacterial cells were harvested by
99 centrifugation at the end of the exponential growth phase, washed twice with phosphate buffer
100 (50 mM, pH 7.5), and finally suspended in the same buffer. The cells were broken by sonication
101 and the extract was centrifuged (12,000 x g for 120 min), and the resulting supernatant was used
102 for enzyme assays. All procedures were carried out at 4°C.

103 **Enzyme assays for a 2C4AP-deaminase**

104 The 2C4AP activity was determined by measuring ammonia released from 2C4AP upon
105 incubation with cell-free lysate. The standard reaction mixture contained 50 mM phosphate
106 buffer (pH 7.5), 0.2 mM NADH, 30 mg of cell-free lysate, and 300 µM 2C4AP in a total
107 reaction volume of 2 ml. After, the 5 min, the reaction mixture were centrifuged and 0.5 ml
108 supernatant was subjected to ammonia release and the remaining 1.5 ml was extracted with the
109 equal volume of ethyl acetate and analyzed by the GC-MS.

110 **Enzyme assay for a CHQ-dehalogenase**

111 CHQ dehalogenase activity was determined as the total chloride released at 30°C in a reaction
112 contained 50 mM phosphate buffer (pH 7.5), 0.2 mM NADPH, 30 mg of cell-free lysate and 200
113 µM of CHQ. The final volume of the reaction mixture was 5 ml. Samples were collected at
114 regular intervals and assayed for chloride ions as described above. Samples were also extracted
115 with equal volume of ethylacetate and extracted samples were analyzed by GC-MS to identify
116 the product of reaction.

117 **Enzyme assay for a HQ-dioxygenase**

118 The HQ dioxygenase activity was determined spectrophotometry by monitoring the formation of
119 γ-hydroxymuconic semialdehyde at 320 nm. The reaction mixture contained (in a final volume
120 of 1 ml) 50 mM phosphate buffer, 0.1 mM hydroquinone, 0.1 mM ferrous sulphate and 30 mg
121 crude extracts of the protein. Samples were taken at 0 and 5 min and the UV spectra were
122 recorded.

123 **Results**

124 **Growth and degradation studies**

125 The degradation and growth studies showed that *Arthrobacter* sp. SPG utilized 0.3 mM 2C4AP
126 as a sole carbon and energy source and degraded 2C4AP within 48 h (Fig. 1a). The
127 stoichiometric releases of chloride and ammonia ions were detected during the degradation of
128 2C4AP. Ammonia release occurred before the chloride release that suggested the initiation of the
129 2C4AP degradation via ammonia release (Fig. 1b).

130 **Identification of metabolites**

131 HPLC analysis confirmed the degradation of 2C4AP within 48h (Fig. 2). In the sample of 0
132 and 12 h, only peak of 2C4AP was detected with the retention time of 13.36 min. However, in the

133 sample of 24 h, a peak of metabolite 1 was detected along with a peak of 2C4AP. The retention
134 time of metabolite 1 (6.89 min) was exactly match with the authentic standard of
135 chlorohydroquinone. In the sample of 36 h, the peaks of metabolite I and II were detected along
136 with the peak of 2C4AP. The retention time of metabolite 11 (5.09 min) was exactly match with
137 the authentic standard of hydroquinone. In the sample of 48 h, neither the peak of 2C4AP nor
138 the peak of any metabolite was detected that suggested the complete degradation of 2C4AP by
139 *Arthrobacter* sp. SPG.

140 GC-MS analysis showed that the mass fragment of the metabolite I and II were observed at 144
141 m/z and 110 m/z, respectively that were exactly match with authentic standards of CHQ and HQ
142 (Fig. 3). On the basis of the GC-MS, metabolite I and II were identified as CHQ and HQ.

143 **Enzyme assays**

144 We have detected enzyme activities of a 2C4AP-deaminase, a CHQ-dehalogenase and HQ-
145 dioxygenase in the crude extracts of 2C4AP induced cells of strain SPG. The 2C4AP deaminase
146 activity was confirmed by the stoichiometric release of ammonium ions and identification of the
147 product chlorohydroquinone by the GC-MS. The mass fragment of product was observed at 144
148 m/z equivalent to CHQ. The CHQ-dehalogenase activity was confirmed by the stoichiometric
149 release of chloride ions and detection of the product, HQ by the GC-MS. The mass fragment of
150 product was observed at 110 m/z equivalent to HQ. The HQ dioxygenase catalyzed the
151 conversion of HQ to γ -hydroxymuconic semialdehyde via ring cleavage. The spectrophotometric
152 analysis of HQ dioxygenase assay showed that peak of the HQ at 289 nm was disappeared and
153 peak of γ -hydroxymuconic semialdehyde (HMS) around 320 nm was appeared.

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155 Discussion

156 *Arthrobacter* sp. SPG utilized 2C4AP as a sole carbon and energy source and degraded it with
157 release of stoichiometric amounts of chloride and ammonium ions. The CHQ was identified as
158 an intermediate of the degradation pathway of 2C4AP. The enzyme 2C4AP-deaminase was
159 involved in the conversion of 2C4AP into CHQ. Literature studies showed that the CHQ is a
160 common metabolite in the degradation pathway of several chlorinated compounds [8, 9, 10, 11,
161 12]. Reddy et al. [10] reported that the CHQ was degraded either via HQ or via 2-
162 chlorotrihydroxybenzene in degradation pathway of 2,4,6-trichlorophenol in *Phanerochaete*
163 *chrysosporium*. Miyauchi et al. [11, 12] showed that CHQ cleaved to maleylacetate or
164 dechlorinated to HQ in the degradation pathway of gamma-hexachlorocyclohexane in
165 *Sphingomonas paucimibilis* UT260. *Arthrobacter* sp. SJCon degraded 2-chloro-4-nitrophenol via
166 a CHQ pathway in which CHQ was cleaved to maleylacetate by a CHQ-dioxygenase.
167 *Burkholderia* sp. RKJ 800 [8] and *Rhodococcus imtechensis* RKJ 300 [9] degraded 2-chloro-4-
168 nitrophenol via a CHQ pathway in which CHQ dehalogenated to HQ. In this study, we have also
169 detected HQ as a metabolite of degradation of 2C4AP by strain SPG. Furthermore, we have also
170 detected the activity of the CHQ dehalogenase in the crude extracts of the 2C4AP-induced cells
171 of strain SPG that confirmed the formation of the HQ from CHQ with release of chloride ion.
172 The HQ was also detected as a metabolite of degradation pathway of 4-nitrophenol in a various
173 Gram-positive bacteria [7, 13]. In the degradation pathway of 4-nitrophenol, HQ was cleaved
174 into γ -hydroxymuconic semialdehyde by a HQ-dioxygenase [7, 13]. We have also detected the
175 HQ-dioxygenase activity in the crude extracts of the 2C4AP-induced cells of strain SPG that
176 suggested the cleavage of HQ into γ -hydroxymuconic semialdehyde.

177 On the basis of the identified metabolites and the enzyme assays, we have proposed a pathway of
178 degradation of 2C4AP for *Arthrobacter* sp. SPG (Fig. 4) The 2C4AP degradation was initiated
179 with release of ammonium ion and the formation of CHQ that dehalogenated to HQ with release
180 of chloride. The further degradation of HQ was proceeded via ring cleavage.

181 The 2C4AP degradation pathway identified in *Arthrobacter* sp. SPG was compared with the
182 degradation pathway of 4C2AP in *Burkholderia* sp. RKJ 800 [1]. It was observed that the initial
183 mechanism of degradation of 2C4AP in strain SPG was similar with that of the 4C2AP
184 degradation in strain RKJ 800 [1]. Both stains initiated the degradation of an isomer of CAPs
185 with removal of ammonium ion by a deaminase [1]. The difference in the degradation pathways
186 of 2C4AP and 4C2AP is that the chloride release was occurred before the ring cleavage in the
187 degradation pathway of 2C4AP whereas in the degradation of 4C2AP, chloride was released
188 after the ring cleavage [1].

189 **Conclusion**

190 *Arthrobacter* sp. SPG degraded 2C4AP via a novel pathway in which CHQ and HQ were
191 formed.

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198 **References**

199 1. Arora PK, Srivastava A, Singh VP (2013) Novel degradation pathway of 4-chloro-2-
200 aminophenol via 4-chlorocatechol in *Burkholderia* sp. RKJ 800. Environ Sci Pollut Res Int. Doi:
201 10.1007/s11356-013-2167-y.

202
203 2. Arora PK, Jain RK (2012) Biotransformation of 4-chloro-2-nitrophenol into 5-chloro-2-
204 methylbenzoxazole by a marine *Bacillus* sp. strain MW-1. Biodegradation 23:325–331.

205
206 3. Park HS, Lim SJ, Chang YK, Livingston AG, Kim HS (1999) Degradation of
207 chloronitrobenzenes by a coculture of *Pseudomonas putida* and a *Rhodococcus* sp. Appl Environ
208 Microbiol 65:1083–1091.

209
210 4. Arora PK, Sharma A, Mehta R, Shenoy BD, Srivastava A, Singh VP (2012b) Metabolism of
211 4-chloro-2-nitrophenol in a Gram-positive bacterium, *Exiguobacterium* sp. PMA. Microb Cell
212 Fact 11:150.

213 5. Lendenmann U, Spain JC (1996) 2-Aminophenol 1,6-dioxygenase: a novel aromatic ring
214 cleavage enzyme purified from *Pseudomonas pseudoalcaligenes* JS45. J Bacteriol 178:6227–
215 6232.

216 6. Arora PK (2012) Decolourization of 4-chloro-2-nitrophenol by a soil bacterium, *Bacillus*
217 *subtilis* RKJ 700. PLOS ONE 7(12):e52012.

218

- 219 7. Arora PK (2012) Metabolism of *para*-nitrophenol in *Arthrobacter* sp. SPG, E3 J Environ Sci
220 Manag. 3 (2012) 52-57.
221
- 222 8. Arora PK, Jain RK (2012) Metabolism of 2-chloro-4-nitrophenol in a Gram negative bacterium,
223 *Burkholderia* sp. RKJ 800. PLOS ONE 7(6):e38676.
224
- 225 9. Arora PK, Sasikala C, Ramana CV (2012) Degradation of chlorinated nitroaromatic
226 compounds. Appl Microbiol Biotechnol 93:2265–2277.
227
- 228 10. Reddy GV, Gelpke MD, Gold MH (1998) Degradation of 2,4,6- trichlorophenol by
229 *Phanerochaete chrysosporium*: involvement of reductive dechlorination. J Bacteriol 180:5159 -
230 5164.
231
- 232 11. Miyauchi K, Adachi Y, Nagata Y, Takagi M (1999) Cloning and sequencing of a novel meta-
233 cleavage dioxygenase genewhose product is involved in degradation of gamma-
234 hexachlorocyclohexanein *Sphingomonas paucimobilis*. J Bacteriol181:6712–6719.
235
- 236 12. Miyauchi K, Suh SK, Nagata Y, Takagi M (1998) Cloning and sequencing of a 2,5-
237 dichlorohydroquinone reductive dehalogenase gene whose product is involved in degradation of
238 gammahexachlorocyclohexane by *Sphingomonas paucimobilis*. J Bacteriol 180:1354–1359.
239
- 240 13. Ju KS, Parales RE (2010) Nitroaromatic compounds, from synthesis to biodegradation.
241 Microbiol Mol Biol R. 74: 250-272.

242 **Figure legends:**

243 Fig. 1. Utilization and degradation of 2-chloro-4-aminophenol by *Arthrobacter* sp. SPG.

244 Fig 2. High performance liquid chromatography elution profiles of the samples of the
245 degradation of 2-chloro-4-aminophenol by *Arthrobacter* sp. SPG.

246 Fig. 3. Mass spectrum of metabolite I(a), 11(b) and authentic standards chlorohydroquinone (c)
247 and hydroquinone(d).

248 Fig. 4. Proposed pathway of degradation of 2-chloro-4-aminophenol for *Arthrobacter* sp. SPG.

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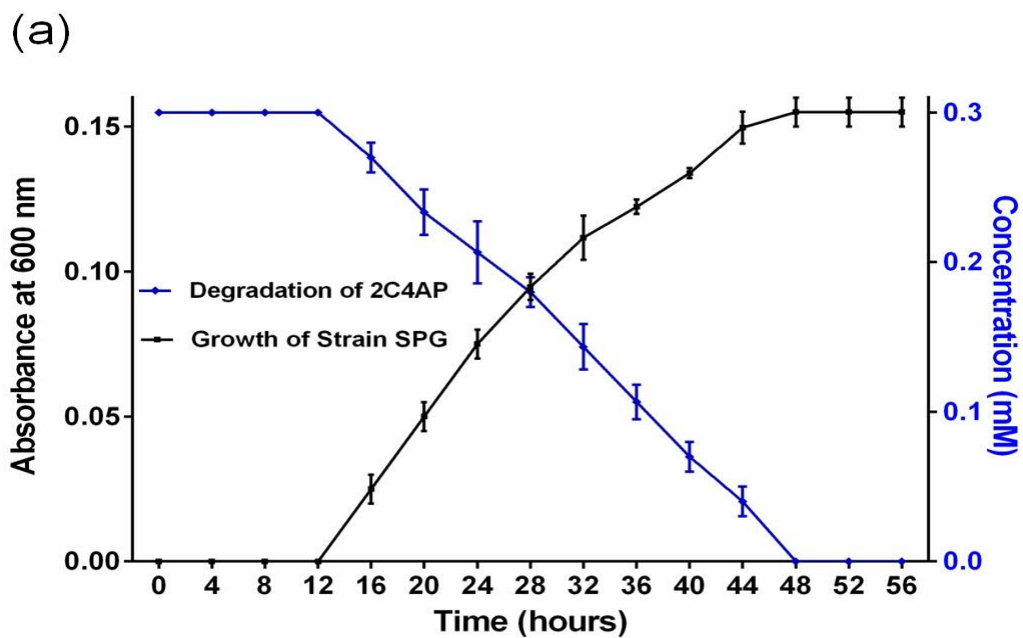
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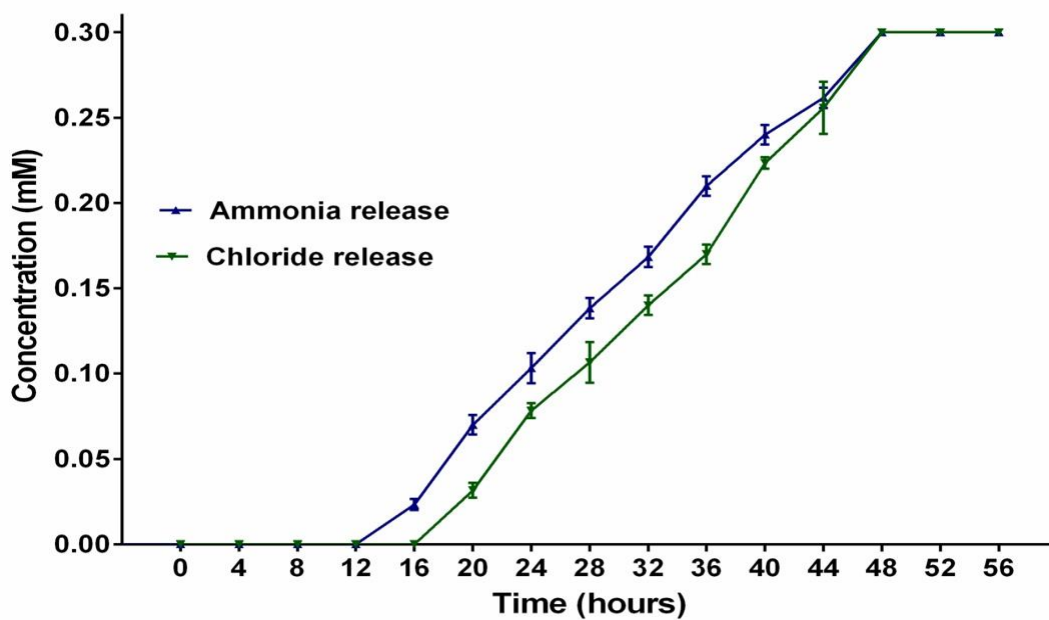
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(b)



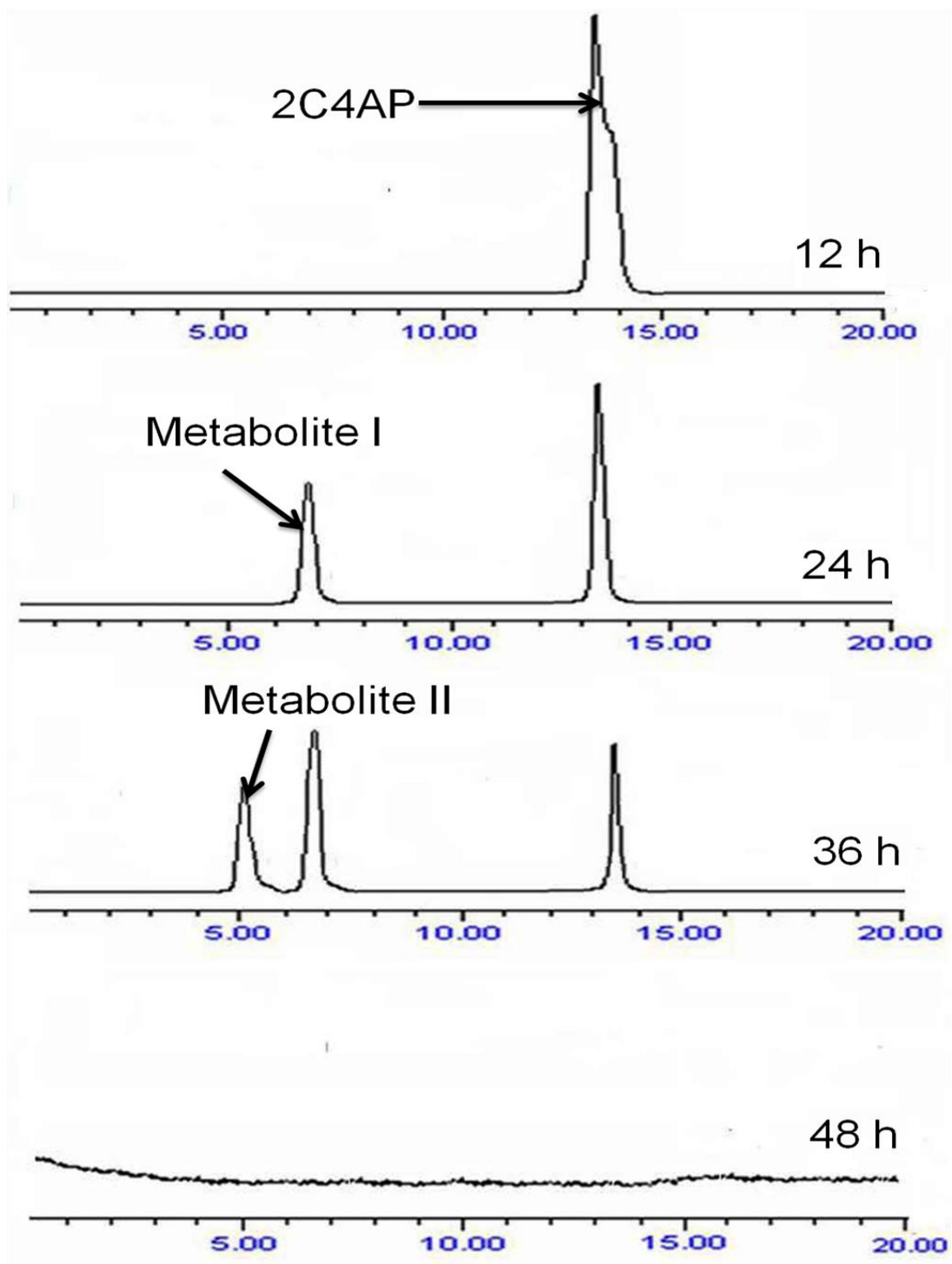
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Figure 1



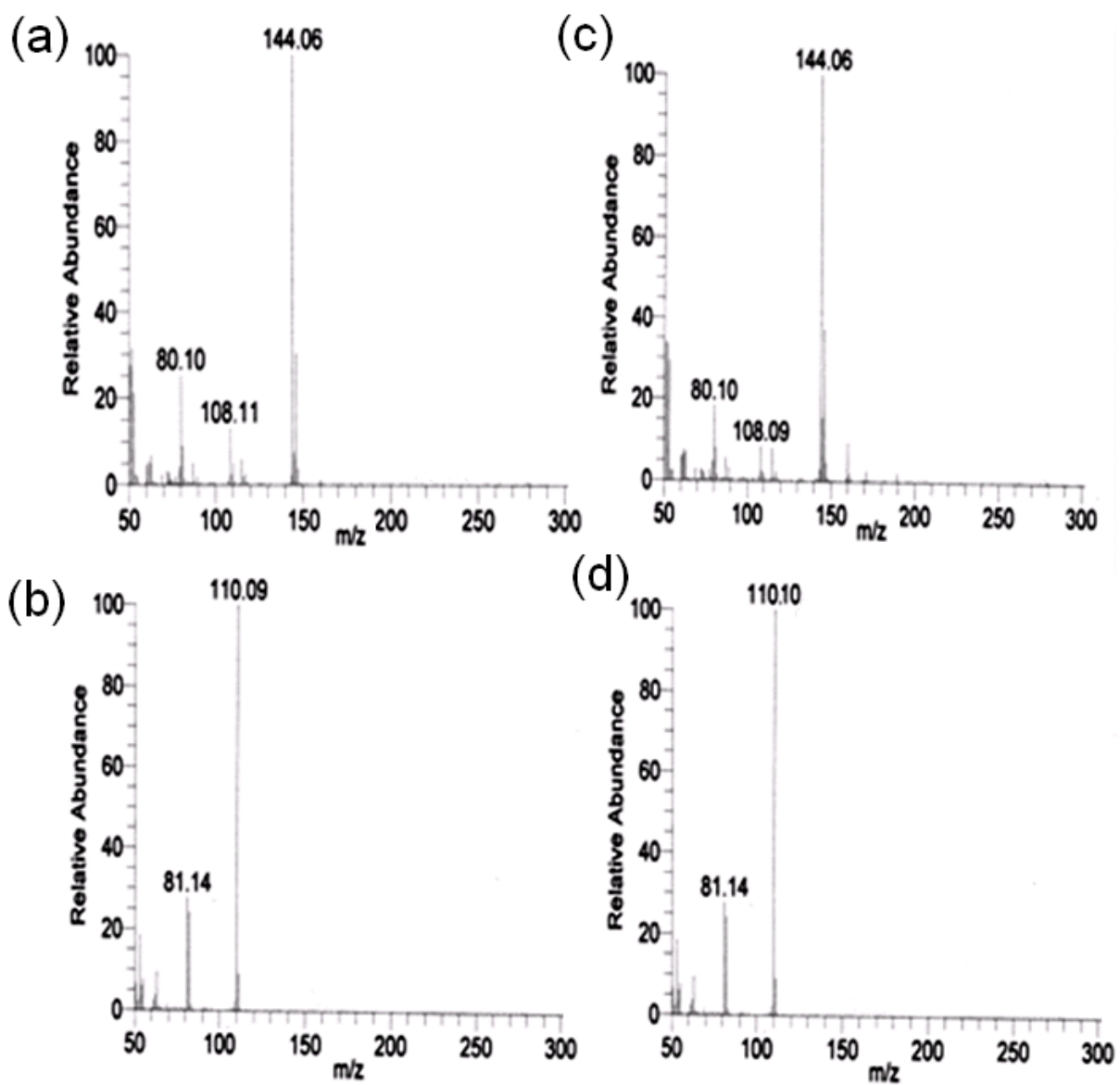
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Figure 2



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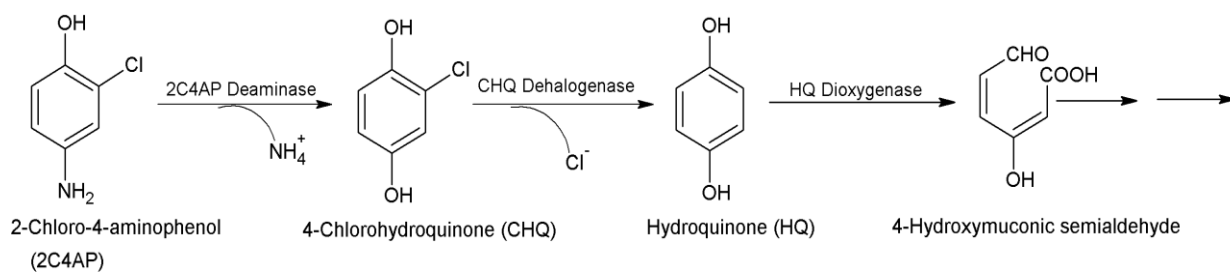
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Figure 3

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Figure 4