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# Connecting laboratory behavior to field function through stable isotope analysis

Mael G Glon, Eric R Larson, Kevin L Pangle

Inherent difficulties of tracking and observing organisms in the field often leave researchers with no choice but to conduct behavioral experiments under laboratory settings. However, results of laboratory experiments do not always translate accurately to natural conditions. A fundamental challenge in ecology is therefore to scale up from small area and short-duration laboratory experiments to large areas and long-durations over which ecological processes generally operate. In this study, we propose that stable isotope analysis may be a tool that can link laboratory behavioral observations to past field interactions or function of individual organisms. We conducted laboratory behavioral assays to measure dominance of invasive rusty crayfish, *Orconectes rusticus*, and used stable isotope analysis to hindcast trophic positions of these crayfish under preceding natural conditions. We hypothesized that more dominant crayfish in our assays would have higher trophic positions if dominance were related to competitive ability or willingness to pursue high-risk, high-reward prey. We did not find a relationship between crayfish dominance and trophic position, and therefore infer that laboratory dominance of crayfish may not necessarily relate to their ecology in the field. However, this is to our knowledge the first attempt to directly relate laboratory behavior to field performance via stable isotope analysis. We encourage future studies to continue to explore a possible link between laboratory and field behavior via stable isotope analysis, and propose several avenues to do so.

1      **Connecting Laboratory Behavior to Field Function through Stable Isotope Analysis**

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24 **Abstract**

25 Inherent difficulties of tracking and observing organisms in the field often leave  
26 researchers with no choice but to conduct behavioral experiments under laboratory settings.  
27 However, results of laboratory experiments do not always translate accurately to natural  
28 conditions. A fundamental challenge in ecology is therefore to scale up from small area and  
29 short-duration laboratory experiments to large areas and long-durations over which ecological  
30 processes generally operate. In this study, we propose that stable isotope analysis may be a tool  
31 that can link laboratory behavioral observations to past field interactions or function of  
32 individual organisms. We conducted laboratory behavioral assays to measure dominance of  
33 invasive rusty crayfish, *Orconectes rusticus*, and used stable isotope analysis to hindcast trophic  
34 positions of these crayfish under preceding natural conditions. We hypothesized that more  
35 dominant crayfish in our assays would have higher trophic positions if dominance were related to  
36 competitive ability or willingness to pursue high-risk, high-reward prey. We did not find a  
37 relationship between crayfish dominance and trophic position, and therefore infer that laboratory  
38 dominance of crayfish may not necessarily relate to their ecology in the field. However, this is to  
39 our knowledge the first attempt to directly relate laboratory behavior to field performance via  
40 stable isotope analysis. We encourage future studies to continue to explore a possible link  
41 between laboratory and field behavior via stable isotope analysis, and propose several avenues to  
42 do so.

43

44 **Keywords:** mixing model; dominance; agonistic assays; *Orconectes rusticus*; individual  
45 variation; invasive species

46

## 47 **Introduction**

48 Animal behavior is inherently linked with the fields of ecology and evolution (Sih, Bell  
49 & Johnson, 2004; Réale, Reader & Sol, 2007), and informs applications such as management of  
50 biological invasions (Sih *et al.*, 2010). Owing to logistical difficulties inherent to tracking and  
51 observing organisms without interference in the field, however, many behavioral studies are  
52 conducted *ex situ* in a laboratory setting, where it may be difficult to extrapolate findings to  
53 natural conditions (Niemelä & Dingemanse, 2014; Zavorka *et al.*, 2015). For example, a suite of  
54 often-correlated behaviors including aggression, dominance, and boldness are believed to  
55 contribute to the success of some invasive over native species (Pintor, Sih & Kerby, 2009;  
56 Hudina, Hock & Žganec, 2014), but these same behaviors can be considerably muted in duration  
57 or intensity when observed in the field (Bergman & Moore, 2003; Larson & Magoullick, 2009).  
58 One of ecology's most fundamental challenges is scaling up from the type of small area and  
59 short duration experiments that are easy to conduct, to the larger areas and longer durations over  
60 which ecological processes often operate (Lodge *et al.*, 1998). This same challenge applies when  
61 relating animal behaviors observed in the laboratory to ecological function and intra- or inter-  
62 specific interactions *in situ*.

63 We propose here that linking laboratory behavioral observations to past field interactions  
64 or function of specific, individual organisms may be an overlooked application of stable isotope  
65 analysis. Stable isotopes of elements such as carbon and nitrogen are assimilated into tissues of  
66 consumer organisms relative to their diets in predictable and quantifiable ways (DeNiro &  
67 Epstein, 1978; DeNiro & Epstein, 1981). Importantly, stable isotopes of consumers equilibrate  
68 with those of their diets at different rates for different tissues, giving snapshots of ecological  
69 interactions that may scale from previous days to years (Buchheister & Latour, 2010). Analyzing

70 stable isotope ratios in organisms can provide ecological insights ranging from habitat use and  
71 movement (Hobson, 1999) to trophic position (Post, 2002). For example, stable isotope analysis  
72 of feathers has been used to make inferences about migration and habitat use of several species  
73 of seabirds that spend winter months far from land and are therefore difficult to study during this  
74 period (Phillips *et al.* 2009). In another example, Cherel *et al.* (2008) used stable isotope analysis  
75 to identify the trophic position and diet composition of southern elephant seals (*Mirounga*  
76 *leonina*) which forage at depths exceeding 1000 m and have largely digested their meals by the  
77 time they return to land, precluding them from being studied using traditional methods (e.g.,  
78 direct observation, gut content analysis). Similarly to how these and other studies have applied  
79 stable isotope analysis to infer the influence of past behavior on current success of organisms, we  
80 propose that stable isotope analysis could permit researchers to link laboratory interactions with  
81 previous *in situ* habitat selection, movement, diet choice, or competitive interactions (Figure 1).

82 We conducted laboratory behavioral assays to measure individual dominance of invasive  
83 rusty crayfish, *Orconectes rusticus*, and used stable isotope analysis to hindcast trophic position  
84 of these crayfish under natural field conditions. We predicted that more dominant crayfish in the  
85 behavioral assays would have higher trophic positions if dominance were related to competitive  
86 ability in the field (e.g., ability to access high quality food such as macroinvertebrates; Roth,  
87 Hein & Vander Zanden, 2006) or willingness to pursue high-risk, high-reward prey such as fish  
88 or other crayfish (Taylor & Soucek, 2010). Alternatively, dominance and trophic position may  
89 not be associated if laboratory behaviors are ultimately uninformative with respect to past  
90 interactions of organisms. Numerous previous studies have used stable isotope analysis to infer  
91 various *in situ* behaviors of organisms, such as habitat use and diet preferences (e.g., Hildebrand  
92 *et al.*, 1996; Rubenstein & Hobson, 2004); however, our study is the first to our knowledge to

93 seek a direct relationship for individual organisms between laboratory behaviors and field  
94 function as determined by stable isotope analysis, and proposes the linkage of laboratory  
95 behavioral assays and stable isotopes as a more common practice in the future.

96

97 **Methods**

98 *Orconectes rusticus* was introduced via the bait trade to the Laurentian Great Lakes circa  
99 1960 and has negatively affected fish, macrophytes, and freshwater macroinvertebrates  
100 (McCarthy *et al.*, 2006; Peters *et al.* 2014). The invasion success of this crayfish has made it the  
101 focal point of a large number of laboratory and field studies (e.g., Olsen *et al.*, 1991; Wilson *et*  
102 *al.*, 2004) and hence, a useful organism to test for linkages between field and laboratory  
103 behavior. We collected adult form II (reproductively inactive) male *O. rusticus* (n=40) by hand  
104 on 16 June 2015 in the Chippewa River, Michigan (43.5652°, -84.9183°), where this species is  
105 invasive. Because size influences the outcome of crayfish agonistic trials (Bergman & Moore,  
106 2003), we used crayfish within a carapace length range of 23.41 to 27.53 mm, the smallest size  
107 range for which we could collect 40 crayfish (see supplementary material for additional  
108 morphometrics). Rusty crayfish in this size range are small adults of the same age class (Momot,  
109 1967) and are therefore unlikely to have diets that differ from one another due to ontogenetic  
110 shifts (Bondar *et al.*, 2005; Larson, Olden & Usio, 2010). Immediately following collection, we  
111 placed crayfish in individual 16 oz. plastic containers filled to a depth of 2 cm with river water  
112 and a rock for shelter.

113

114 *Agonistic assays*

115           Laboratory agonistic assays for crayfish are often conducted after isolating individuals  
116    for at least one week to remove possible previous social experience that could influence  
117    interactions (Seebacher & Wilson, 2007). We conducted our experiment directly following  
118    collection (17 June 2015 during daylight hours [07h19-18h59]), but believe that retaining any  
119    existing dominance hierarchies from the field would increase the likelihood of a relationship  
120    between laboratory behaviors and previous field function.

121           We conducted three rounds of twenty, randomized paired assays, with each crayfish  
122    fighting one opponent per round (no interactions between individuals were repeated). In order to  
123    track individual crayfish, we randomly assigned each crayfish a number from 1 to 40, which we  
124    wrote on the dorsal side of its carapace using a permanent marker. Prior to the start of each  
125    assay, crayfish were placed on opposite sides of a separator in a 19 l bucket and allowed to  
126    acclimate for 15 minutes. We then removed the separator and allowed the crayfish to interact for  
127    10 minutes. During each assay, we scored each of the two crayfish individually based on the  
128    interactions that took place when they were within one body length of each other. All agonistic  
129    assays were watched and scored in real time by a single observer to ensure consistency in  
130    scoring. The agonistic assays within each of the three rounds were held in a random order, and  
131    the observer had no knowledge of totaled crayfish scores from previous rounds so as to avoid  
132    bias.

133           The scoring system we used has possible point values ranging from -2 (fast retreat) to 5  
134    (unrestrained fighting) and is based on the ethogram modified from Bruski & Dunham (1987;  
135    Table 1). Following each assay, the participating crayfish were returned to their original holding  
136    container. We then rinsed buckets and refilled them to a depth of 5 cm with fresh water from the  
137    Chippewa River (18-20°C). At the conclusion of all assays, crayfish were placed in individual,

138 labelled bags and euthanized by freezing at -17.8°C. We calculated the dominance score of each  
139 crayfish by first summing its scores from each round, then taking the mean of the three resulting  
140 scores.

141

142 *Stable isotope analysis*

143 Stable isotope analysis is a technique based on the principle that the ratios of heavy to  
144 light isotopes in the tissues of consumers reflect those of their diets in a predictable way (DeNiro  
145 & Epstein, 1978; DeNiro & Epstein, 1981). Stable isotope analysis generally entails drying and  
146 homogenizing tissue or whole-body samples of focal organisms, then using a mass spectrometer  
147 coupled with an elemental analyzer to determine their constituent ratios of heavy to light isotopes  
148 (i.e.,  $R_{sample}$ ). The isotope signatures of samples ( $\delta^x$ ), expressed in per mille (‰), are then  
149 calculated as  $\delta^x = \left( \frac{R_{sample}}{R_{standard}} - 1 \right) * 1000$  where  $R_{standard}$  is the isotopic ratio of a standard (e.g.,  
150 Vienna PeeDee Belemnite for carbon; air for nitrogen). This technique is often used to study the  
151 roles and interactions of organisms in ecosystems, particularly as related to trophic position and  
152 diet composition (Vander Zanden & Rasmussen, 1999; Post, 2002), but patterns of stable isotope  
153 spatial structure can also be applied to study organismal movement and habitat use (Hobson,  
154 1999; Seminoff *et al.*, 2012).

155 In freshwater ecology, the most commonly used stable isotopes have been carbon and  
156 nitrogen (denoted  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$ , respectively). Specifically,  $\delta^{13}\text{C}$  provides a tracer of energy  
157 source origin because it is fixed by primary producers at photosynthesis and is well-conserved up  
158 food chains with little change in value with each increasing trophic level (termed discrimination;  
159 generally 0-1 ‰; Fry & Sherr, 1984). Common sources of primary productivity in freshwater  
160 habitats that can often be distinguished by analyzing  $\delta^{13}\text{C}$  include a benthic algal pathway, an

161 open water phytoplankton pathway, and an allochthonous terrestrial detrital pathway; the  
162 importance of these pathways to consumers can vary depending on habitat attributes (Dekar,  
163 Magoullick & Huxel, 2009; Francis *et al.*, 2011). In contrast to  $\delta^{13}\text{C}$ ,  $\delta^{15}\text{N}$  can be used to estimate  
164 trophic position of organisms as it generally increases or discriminates at a predictable  $3.4 \pm 1.1$   
165 ‰ with each increasing trophic level, from primary producers to primary, secondary, and tertiary  
166 consumers (Minagawa & Wada, 1984). In some cases,  $\delta^{15}\text{N}$  can be used alone to infer trophic  
167 position of organisms; however, this is not the case if different sources of primary productivity  
168 used by a consumer are depleted or enriched in  $\delta^{15}\text{N}$  relative to each other (Vander Zanden &  
169 Rasmussen, 1999; Post, 2002; Figure 2). Under these circumstances, mixing models can be used  
170 to estimate contributions of different energy pathways to consumers, and subsequently correct  
171 for differences in their  $\delta^{15}\text{N}$  enrichment while calculating trophic position of consumers (Post,  
172 2002).

173 For this experiment, we collected snails (*Elimia livescens*; n=45) and mussels (*Elliptio*  
174 *dilatata*; n=5) in the same stretch of the Chippewa River and on the same date as our crayfish  
175 (see above), which we froze at  $-17.8^\circ\text{C}$ , to be used as primary consumer endpoints in a two end-  
176 member stable isotope mixing model related to calculating trophic position of crayfish. We chose  
177 these specific organisms as they are reliable primary consumers (i.e., trophic position = 2) whose  
178 relatively large size and long lives make their isotopic signatures more robust to spatial and  
179 temporal variation than those of primary producers (Cabana & Rasmussen, 1996; Post, 2002).  
180 Specifically, we used snails to represent the isotopic signature of the algal food web, and filter-  
181 feeding mussels as an additional endpoint to represent a broad range of other potential sources of  
182 primary production in lotic systems (e.g., benthic algae, terrestrial detritus, and phytoplankton  
183 from upstream lentic systems; Raikow & Hamilton, 2001; Cole & Solomon, 2002).

184 We dissected crayfish for abdominal muscle, snails for whole body without shell, and  
185 mussels for foot muscle. We dried samples at 60°C for 24 h, homogenized them in an ethanol-  
186 rinsed mortar and pestle, then weighed and encapsulated aliquots weighing  $.64 \pm .04$  mg of each  
187 sample into tin capsules. We sent these samples to the Stable Isotope Mass Spectrometry Lab at  
188 the University of Florida for analysis on a Micromass Prism II isotope ratio mass spectrometer  
189 coupled with an elemental analyzer. Two internationally recognized standards (l-glutamic acids),  
190 USGS40 (mean  $\pm$  standard deviation  $\delta^{13}\text{C}$ ,  $-26.39\text{ ‰} \pm 0.11$ ;  $\delta^{15}\text{N}$ ,  $-4.53\text{ ‰} \pm 0.12$ ; measured  
191 repeatedly for calibration) and USGS41 ( $\delta^{13}\text{C}$ ,  $47.57\text{ ‰}$ ;  $\delta^{15}\text{N}$ ,  $37.36\text{ ‰}$ ; measured once as a  
192 check standard), were measured during the analysis to ensure precision.

193 We calculated the relative contribution of the primary productivity represented by snails  
194 (SPP) to our crayfish as  $\text{SPP} = \frac{(\delta^{13}\text{C}_{\text{crayfish}} - \delta^{13}\text{C}_{\text{mussel}})}{(\delta^{13}\text{C}_{\text{snail}} - \delta^{13}\text{C}_{\text{mussel}})} * 100$ , where  $\delta^{13}\text{C}_{\text{crayfish}}$  is the  $\delta^{13}\text{C}$  of  
195 each crayfish,  $\delta^{13}\text{C}_{\text{mussel}}$  is the mean  $\delta^{13}\text{C}$  of our mussel samples and  $\delta^{13}\text{C}_{\text{snail}}$  is the mean  $\delta^{13}\text{C}$  of  
196 our snail samples. We then calculated the relative contribution of the primary productivity  
197 represented by mussels (MPP) as  $\text{MPP} = 100 - \text{SPP}$ . Lastly, we calculated the trophic position  
198 (TP) of our crayfish as  $\text{TP} = 2 + \frac{\delta^{15}\text{N}_{\text{crayfish}} - (\delta^{15}\text{N}_{\text{snail}} * \text{SPP} + \delta^{15}\text{N}_{\text{mussel}} * (1 - \text{SPP}))}{\Delta^{15}\text{N}}$ , where  $\delta^{15}\text{N}_{\text{crayfish}}$  is  
199 the  $\delta^{15}\text{N}$  of each crayfish,  $\delta^{15}\text{N}_{\text{snail}}$  is the mean  $\delta^{15}\text{N}$  of the snails,  $\delta^{15}\text{N}_{\text{mussel}}$  is the mean  $\delta^{15}\text{N}$  of  
200 the mussels, and  $\Delta^{15}\text{N}$  is a trophic discrimination factor of 3.4 (Minagawa & Wada, 1984).

201

## 202 *Statistical analysis*

203 We used linear regression to test for a relationship between the mean dominance scores  
204 and calculated trophic positions of our crayfish. Additionally, we performed several linear  
205 regressions controlling for the effect of body size on crayfish dominance by using residuals, as  
206 well as a linear regression testing for a relationship between mean dominance score and

207 unaltered  $\delta^{15}\text{N}$  signatures rather than calculated trophic position (supplementary material). All  
208 analyses were conducted using the R statistical program (R Core Team, 2014).

209

## 210 Results

211 Snails were enriched in  $\delta^{13}\text{C}$  (mean  $\pm$  standard deviation;  $-27.9 \pm 0.9 \text{ ‰}$ ; Figure 2)  
212 relative to mussels ( $-32.0 \pm 0.2 \text{ ‰}$ ). The relatively depleted  $\delta^{13}\text{C}$  signature of the mussels likely  
213 reflects utilization of either phytoplankton or allochthonous terrestrial detritus as food resources,  
214 relative to the generally more  $^{13}\text{C}$  enriched benthic algal pathway (Raikow & Hamilton, 2001;  
215 Cole & Solomon, 2002). Our mixing model allowed us to correct for the relatively depleted  $\delta^{15}\text{N}$   
216 signature of mussels with respect to the trophic positions of our crayfish. The percent reliance of  
217 crayfish (mean  $\pm$  standard deviation) on the algae/snail pathway was  $67.6 \pm 11.3 \text{ %}$ , relative to  
218  $32.4 \pm 11.3 \text{ %}$  on the detritus or phytoplankton/mussel pathway, indicating that most of these  
219 crayfish relied twice as much on the algal than mussel resource pathway. Trophic positions of  
220 crayfish ranged from 2.1 to 2.6 with a mean of  $2.3 \pm 0.1$ , suggesting a range of foraging  
221 behaviors from high reliance on primary producers like benthic algae (i.e., trophic position = 2)  
222 to some predation on primary consumers like snails (i.e., trophic position = 3).

223 The mean crayfish dominance score from the agonistic assays was 29.93 (SD, 28.61;  
224 min, -23.33; max, 80.67). We did not find a significant relationship between dominance and  
225 trophic position ( $y = 0.0005x + 2.32$ ,  $R^2 = 0.0132$ ,  $F_{1,38} = 0.5084$ ,  $p = 0.4802$ ; Figure 3). Our  
226 additional analyses accounting for the role of body size on both dominance and trophic position,  
227 as well as those using an alternative measure of trophic position, did not affect our conclusion  
228 that there is no association between dominance and trophic position (supplementary material).

229

230 **Discussion**

231 We failed to find a relationship between crayfish dominance and trophic position. We  
232 therefore infer that laboratory dominance among these organisms may not necessarily relate to  
233 their dietary preferences in the field, despite our prediction that more dominant crayfish should  
234 be more likely than subordinate crayfish to compete successfully for high quality food or to  
235 pursue high-risk, high reward prey (Roth, Hein & Vander Zanden, 2006; Taylor & Soucek,  
236 2010). However, this is to our knowledge the first attempt to relate laboratory behavior to field  
237 performance via stable isotope analysis; therefore, more studies are warranted to further explore  
238 linkages between these two techniques in light of possible sources of discord.

239 For example, other behaviors may correlate better with trophic position than dominance  
240 in paired agonistic assays. Dominance assays may instead be more informative with respect to  
241 acquisition of shelter to avoid fish predation or fitness via sexual selection (Garvey, Stein &  
242 Thomas, 1994; Bergman & Moore, 2003), whereas trophic position in the field might correlate  
243 better with other measures of laboratory behavior, such as boldness. However, dominance and  
244 boldness have been observed to correlate as “behavioral syndromes” in crayfish (Pintor, Sih &  
245 Kerby, 2009), and we would therefore expect boldness and dominance to both correlate with  
246 trophic position. It is also possible that there is a temporal disconnect between our analysis of  
247 crayfish abdominal tissue, which has an isotopic half-life of approximately 20-30 days (Glon,  
248 Larson & Pangle, 2016), and the social memory of our crayfish, which is thought to last from 60  
249 minutes to one week (Bergman *et al.*, 2003). Use of a tissue with a faster turnover rate (e.g.,  
250 haemolymph) may better reflect the most recent *in situ* behavior of crayfish. Further, male  
251 crayfish of the family Cambaridae cycle between a reproductively inactive form II and an active  
252 form I stage. We used form II male crayfish, which are typically less aggressive than crayfish in

253 form I (Bergman *et al.*, 2003). Replicating our experiment with form I individuals might alter the  
254 results of agonistic assays and their relationship to trophic position.

255 Lastly, a possible limitation of our study was our relatively small sample size (n = 40;  
256 Galván, Sweeting & Reid, 2010). In order to determine if our lack of a significant relationship  
257 stemmed from low power, we conducted power analyses using the pwr package in R (Champely,  
258 2015). We found that for our observed effect size (0.013; calculated as  $\frac{R^2}{1-R^2}$  [Cohen, 1988]) and  
259 an alpha of 0.05 and conventional power of 0.80, we would have required 605 crayfish replicates  
260 to observe statistical significance. Conversely, for an alpha of 0.05 and power of 0.80, the  
261 smallest effect size we would have detected as significant with 40 crayfish replicates was 0.21  
262 ( $R^2 = 0.173$ ). We therefore conclude that although the effect size observed here could only be  
263 detected as statistically significant with an uncommonly high level of replication (perhaps  
264 dismissed as statistical significance without biological significance; Nakagawa & Cuthill, 2007),  
265 our level of replication was adequate to find significant relatively weak effect sizes down to an  
266  $R^2 = 0.173$ .

267 Although our study failed to find an association between crayfish dominance and stable  
268 isotope-estimated trophic position, we believe that there are many unexplored and promising  
269 avenues to combine behavioral and isotope ecology in order to learn more about how behavior  
270 observed in laboratories corresponds with movement and organismal interactions in the field.  
271 Laboratory experiments and stable isotope analyses have both separately been used to explore  
272 the “ecology of individuals” or variation within populations and species (Bolnick *et al.*, 2003;  
273 Niemelä & Dingemanse, 2014; Zavorka *et al.*, 2015), yet to our knowledge, researchers have not  
274 previously combined or compared these approaches for the same organisms. For example, stable  
275 isotope analysis and behavioral assays could be combined to together evaluate whether range

276 expansion of invasive species is being driven by subordinate individuals with low trophic  
277 positions that are excluded from core habitats by dominant intraspecific competitors, or instead  
278 bold or aggressive individuals with high trophic positions that are inclined to disperse (Hudina,  
279 Hock & Žganec, 2014). Further, where distinct stable isotope signatures exist over habitat  
280 gradients (Hobson, 1999), researchers could infer whether individuals with or without dispersal-  
281 related behaviors observed in the laboratory were actually recent arrivals or longstanding  
282 residents of their collection locations. We encourage future studies to further explore the possible  
283 insights gained by linking laboratory behavior with field function through stable isotope analysis,  
284 as doing so could contribute meaningfully to an array of ecological and evolutionary questions.

285

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291

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293

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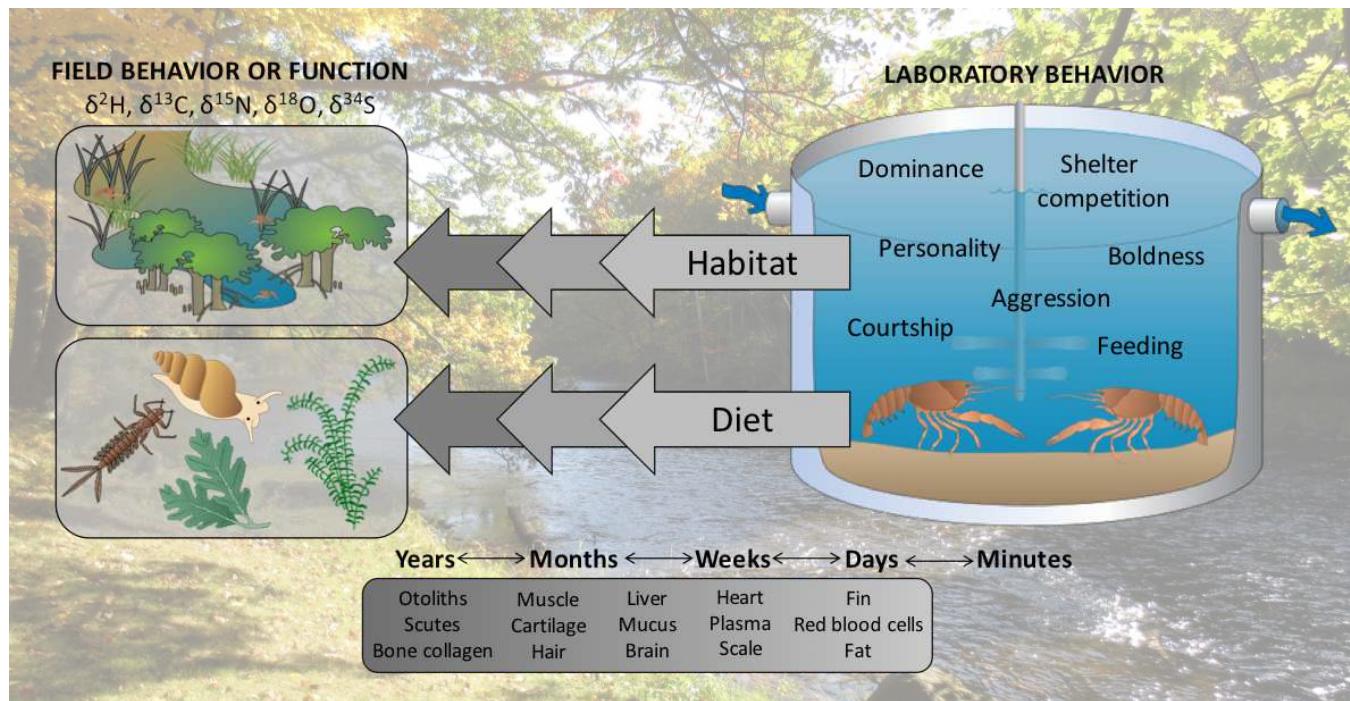
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455 **Figure 1.**

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457 **Figure 1.** Stable isotopes could permit researchers to hindcast the ecological interactions of  
458 organisms, linking behaviors observed in the laboratory with previous field function or behavior.  
459 Symbols courtesy of the Integration and Application Network ([ian.umces.edu/symbols/](http://ian.umces.edu/symbols/)).  
460 Background image is of the study location where organisms were collected (Chippewa River).

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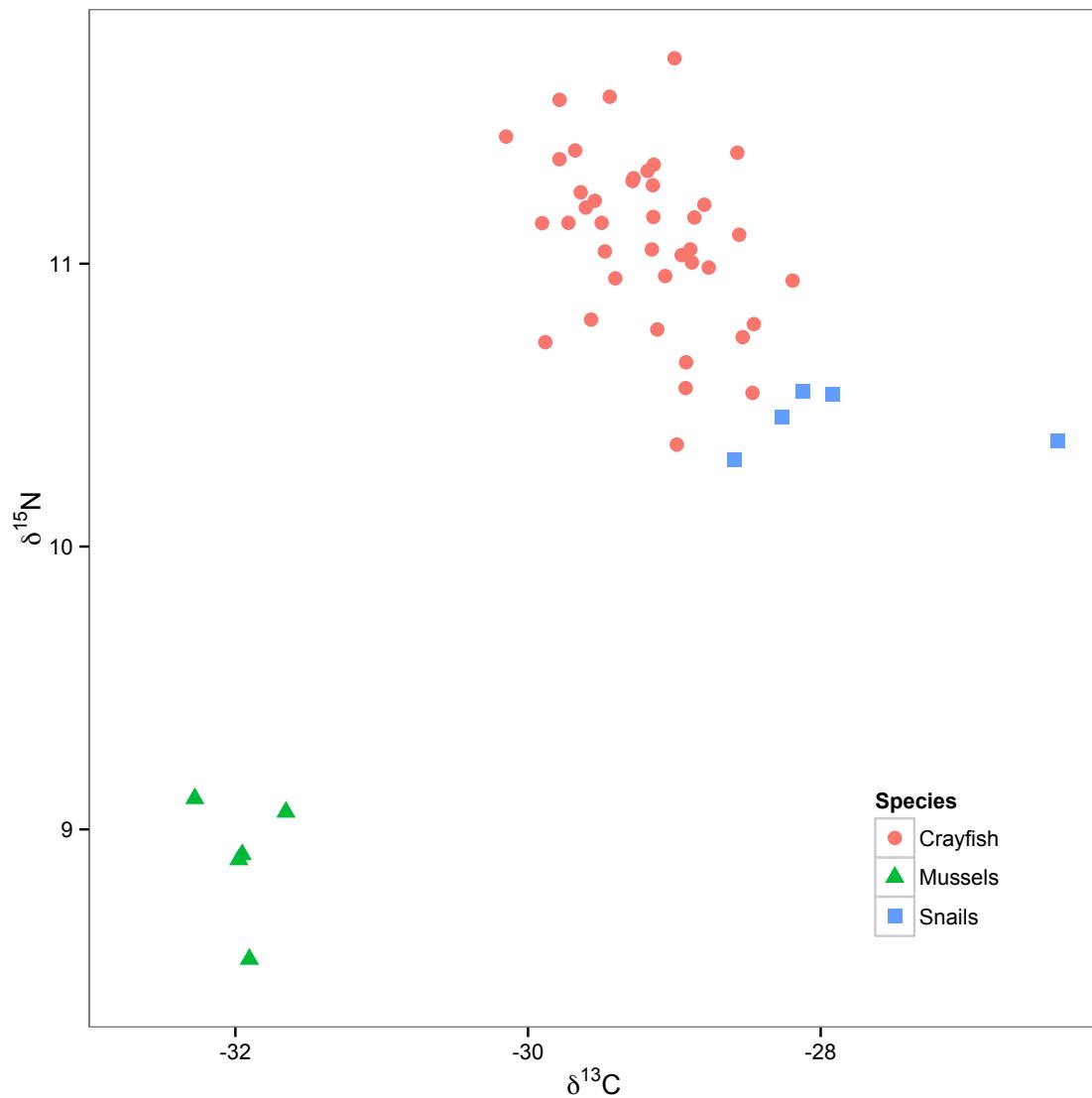
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469 **Figure 2.**

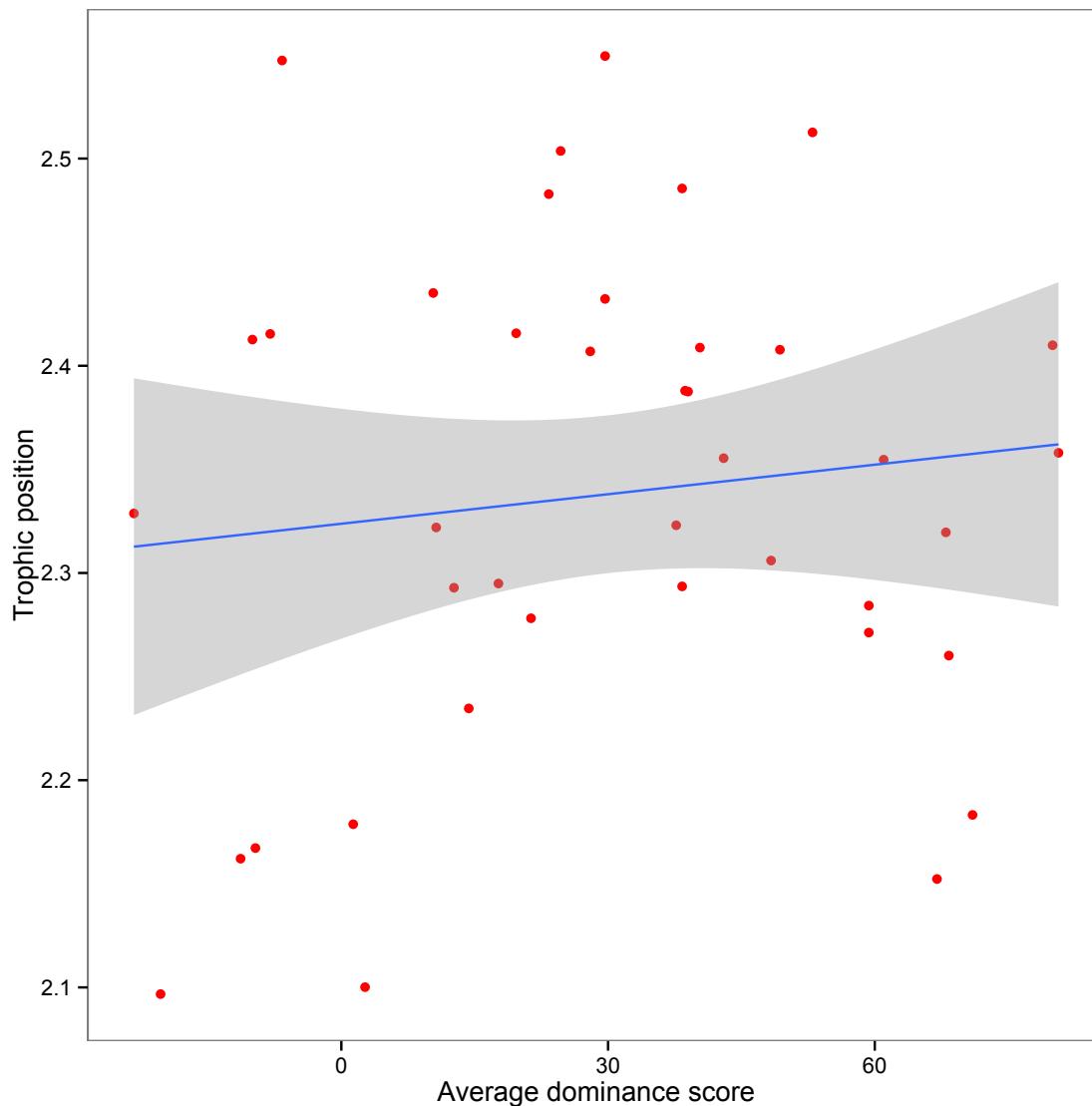
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471 **Figure 2.** Isotopic biplot of  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  for crayfish (red circles), mussels (green triangles),  
472 and snails (blue squares). All values are expressed in per mille (‰) relative to a standard of V-  
473 PDB (Vienna PeeDee Belemnite) for carbon and air for nitrogen.

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477 **Figure 3.**

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479 **Figure 3.** Scatterplot (with 95% CI) of mean assay dominance score for each crayfish over three  
480 agonistic assays and *in situ* trophic position ( $y = 0.0005x + 2.32$ ,  $R^2 = 0.0132$ ,  $F_{1,38} = 0.5084$ ,  $p =$   
481 0.4802).

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**Table 1.** Ethogram modified from Bruski & Dunham (1987)

Score	Description
-2	Tail flip or fast retreat
-1	Slow retreat
0	Within one body length with no visible interaction
1	Approach without threat display
2	Approach with threat display (e.g., meral spread, antennal whips)
3	Boxing, pushing, or other agonistic interaction with closed chelae
4	Grabbing, tearing, or other agonistic interaction with opened chelae
5	Full out, unrestrained fighting, usually with interlocked chelae

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503 **Supplementary material for “Connecting Laboratory Behavior to Field Function through**  
504 **Stable Isotope Analysis”**

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506 Page Content

507 26 I. Crayfish morphometrics

508 27 Table S1. Crayfish Morphometrics

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510 31 Figure S1. Scatterplot of crayfish carapace length and dominance scores

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512 carapace length regression and calculated trophic position

513 33 Figure S3. Scatterplot of crayfish dominance scores and  $\delta^{15}\text{N}$  signatures

514 34 III. References

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526 **I. Crayfish morphometrics**

527 Size influences the outcome of crayfish agonistic trials (Rubenstein & Hazlett, 1974;  
528 Bergman & Moore, 2003); therefore, to better understand what intrinsic factors might be  
529 affecting the results of our agonistic assays, we used digital calipers to measure carapace length  
530 (CL; from the tip of the rostrum to the posterior edge of the carapace), chelae width (at the  
531 widest point of the palm), and chelae length (from the attachment of the carpus and the propodus  
532 to the most distal point of the fixed finger) to the nearest hundredth of a mm. We used a digital  
533 balance to measure mass to the nearest hundredth of a gram (Table S1). Prior to weighing, we  
534 dabbed all crayfish dry for 10 seconds with a paper towel.

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**Table S1.** Crayfish morphometrics.

Measurement	Mean	Standard Deviation	Minimum	Maximum
Carapace length (mm)	25.38	1.26	23.41	27.53
Chelae length (mm)	17.38	1.70	14.05	21.36
Chelae width (mm)	7.20	0.85	5.02	8.31
Mass (g)	5.06	0.78	3.7	6.5

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567 **II. Alternative comparisons of dominance and trophic position**

568         Body size is a factor that strongly influences the outcome of agonistic encounters in  
569         crayfish, with larger individuals generally being more dominant (Bovbjerg, 1953; Rubenstein &  
570         Hazlett, 1974; Bergman & Moore, 2003). We used as small of a crayfish size range as  
571         logistically possible, but the difference between our largest and smallest study organisms was  
572         still 4.12 mm carapace length (Table S1). Despite this, most paired agonistic interaction trials  
573         were between more closely size-matched crayfish (mean  $\pm$  standard deviation;  $1.44 \pm 1.15$  mm  
574         carapace length). Regardless, we sought to determine if dominance scores might better  
575         correspond with the trophic positions of our crayfish if we corrected for the role of size  
576         differences in determining outcomes of agonistic interactions. We did not correct for potential  
577         ontogenetic effects of crayfish size on trophic position (Bondar *et al.*, 2005; Larson, Olden &  
578         Usio, 2010), because we found no significant relationship between crayfish carapace length  
579         (Table S1) and trophic position ( $y = 0.002x + 2.27$ ,  $R^2 = 0.001$ ,  $F_{1,38} = 0.02$ ,  $p = 0.88$ ). However,  
580         as we anticipated, there was a significant relationship between crayfish carapace length and mean  
581         dominance score ( $y = 11.503x - 261.971$ ,  $R^2 = 0.26$ ,  $F_{1,38} = 13.03$ ,  $p < 0.001$ ; Figure S1). Yet,  
582         when we corrected for the effect of crayfish size on dominance by regressing residuals of the  
583         preceding analysis against trophic position, we still did not find a significant relationship,  
584         consistent with our main text conclusion ( $y = 0.00x + 2.34$ ,  $R^2 = 0.01$ ,  $F_{1,38} = 0.54$ ,  $p = 0.47$ ;  
585         Figure S2). The lack of a relationship between dominance and trophic position is therefore  
586         conserved even when accounting for the potential influence of crayfish size on dominance.

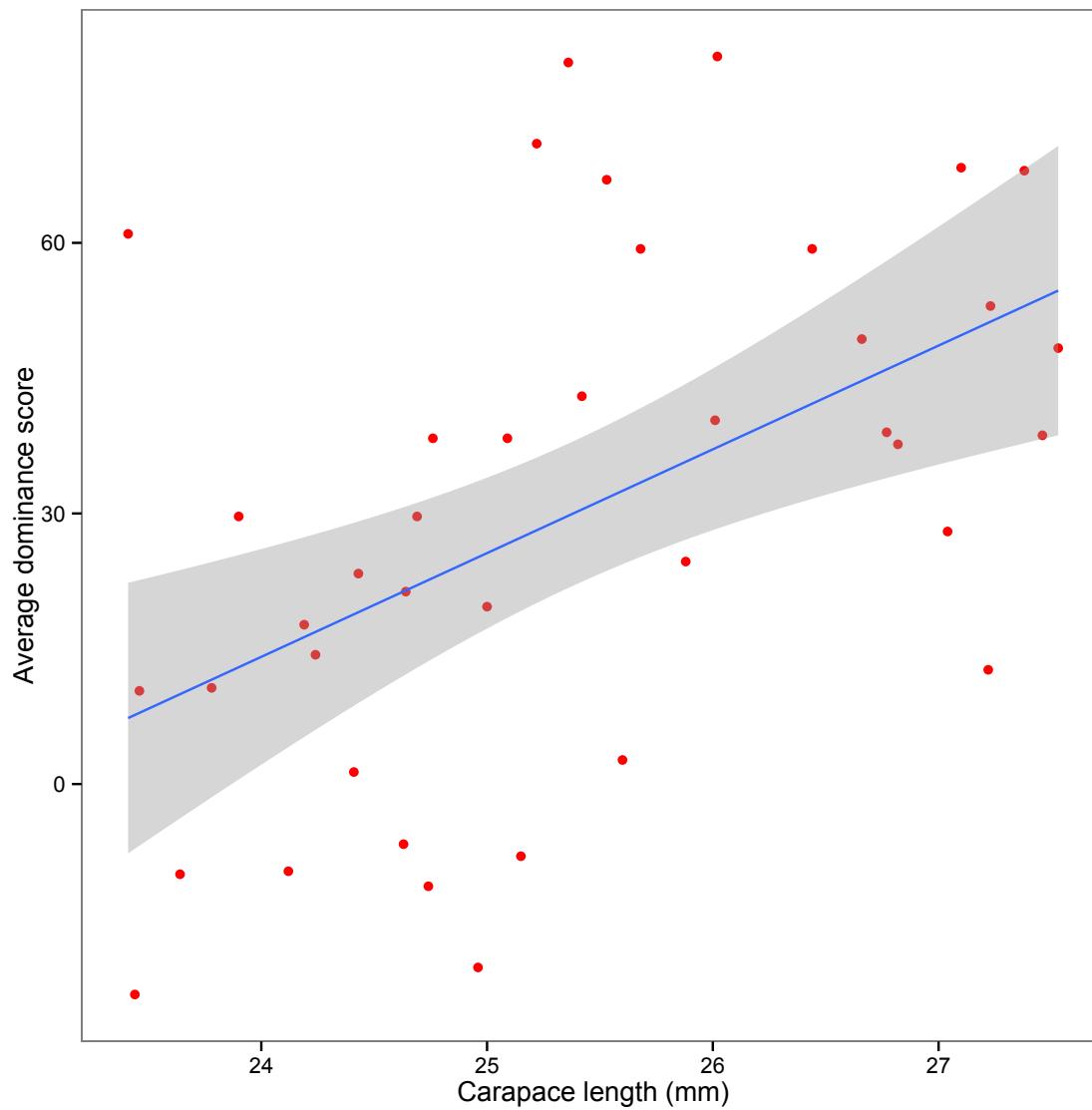
587         Carapace length is the most commonly used size metric for crayfish; however, chelae size  
588         has been shown to dictate success in agonistic encounters and may be a better measure of  
589         dominance in crayfish (Garvey & Stein, 1993). We therefore ran two additional iterations of the

590 analysis presented above, using chelae length and width instead of carapace length. We found  
591 significant relationships between mean dominance scores and both chelae length ( $y = 9.125x -$   
592  $128.686$ ,  $R^2 = 0.29$ ,  $F_{1,38} = 15.72$ ,  $p < 0.001$ ) and chelae width ( $y = 16.040x - 85.562$ ,  $R^2 = 0.23$ ,  
593  $F_{1,38} = 11.07$ ,  $p = 0.002$ ). Yet again, regressing residuals from the chelae length or width and  
594 dominance score analyses against trophic position did not change our main text conclusion that  
595 that dominance and trophic position are unrelated (chelae length residuals vs trophic position;  $y$   
596  $= 0.0002x - 2.34$ ,  $R^2 = 0.001$ ,  $F_{1,38} = 0.04$ ,  $p = 0.85$ ; chelae width residuals vs trophic position;  $y$   
597  $= -0.0001x - 2.34$ ,  $R^2 = 0.01$ ,  $F_{1,38} = 0.02$ ,  $p = 0.89$ ).

598 The use of isotopic mixing models, applied here as a step in calculating trophic position  
599 (Post, 2002), is dependent on a number of assumptions. For example, stream and river  
600 ecosystems can have extremely high spatiotemporal variation in the  $\delta^{13}\text{C}$  and  $\delta^{15}\text{N}$  values of  
601 sources of primary production owing to a number of factors (Fry & Sherr, 1984; Finlay, 2001;  
602 Trudeau & Rasmussen, 2003). Accordingly, we followed convention in using primary  
603 consumers rather than primary producers in mixing model calculations of trophic position (see  
604 section IV), as long-lived organisms like mussels or snails can integrate and correct for this  
605 variability (Post, 2002; Cabana & Rasmussen, 1996). However, we cannot exclude that our field  
606 sampling of primary consumer endpoints for our mixing model could have missed some such  
607 variability inherent to heterogeneous lotic ecosystems, and our collection of potential prey  
608 resources concurrent with crayfish consumers does not necessarily reflect isotopic values of prey  
609 items for *Orconectes rusticus* over preceding weeks or months (Moore & Semmens, 2008).  
610 Another assumption of mixing models is that constant discrimination factors can be used for  
611 each trophic step and between different taxonomic groups and diet items. However,  
612 discrimination factors can vary across taxa, diets, and tissues used (e.g., Stenroth *et al.*, 2006;

613 Caut, Angulo & Courchamp, 2009; Phillips *et al.*, 2014), and consequently may misrepresent  
614 trophic position of a focal organism (Bond & Diamond, 2011). Due to the potential vulnerability  
615 of our model to the preceding assumptions, we also conducted a simpler analysis using crayfish  
616 dominance scores and unaltered  $\delta^{15}\text{N}$  values to determine if our results were dependent on our  
617 specific trophic position calculations. Doing so did not alter our overall nonsignificant result and  
618 conclusion ( $y = 0.002x - 11.04$ ,  $R^2 = 0.03$ ,  $F_{1,38} = 1.292$ ,  $p = 0.26$ ; Figure S3). We therefore  
619 conclude that our result of a lack of relationship between crayfish dominance in the laboratory  
620 and trophic position in the field is robust to our measures of both crayfish dominance and trophic  
621 position.

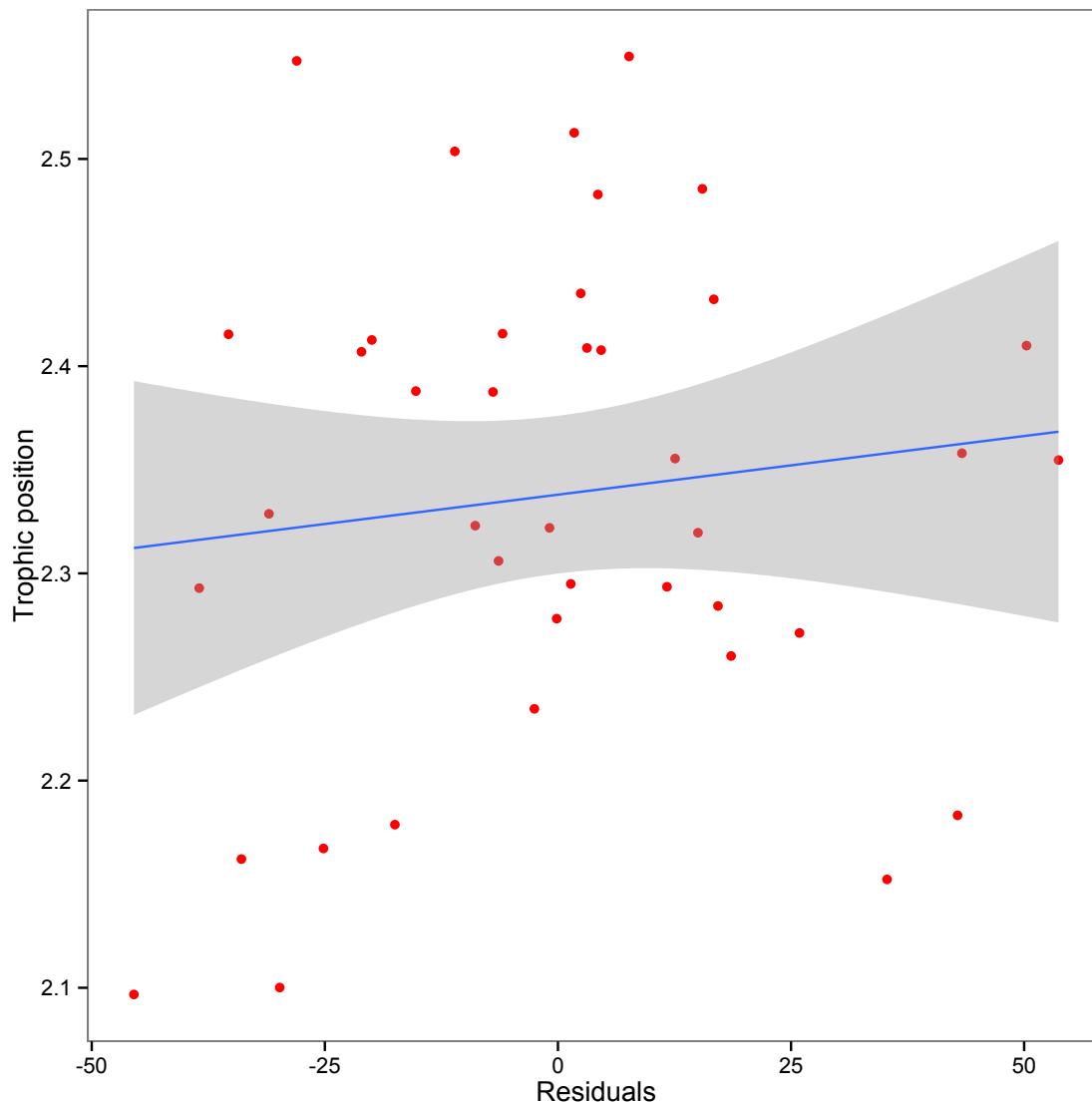
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624 **Figure S1.** Scatterplot (with 95% CI) showing significant relationship between crayfish carapace  
625 length and dominance score from behavioral assays ( $y = 11.503x - 261.971$ ,  $R^2 = 0.26$ ,  $F_{1,38} =$   
626 13.03,  $p < 0.001$ ).

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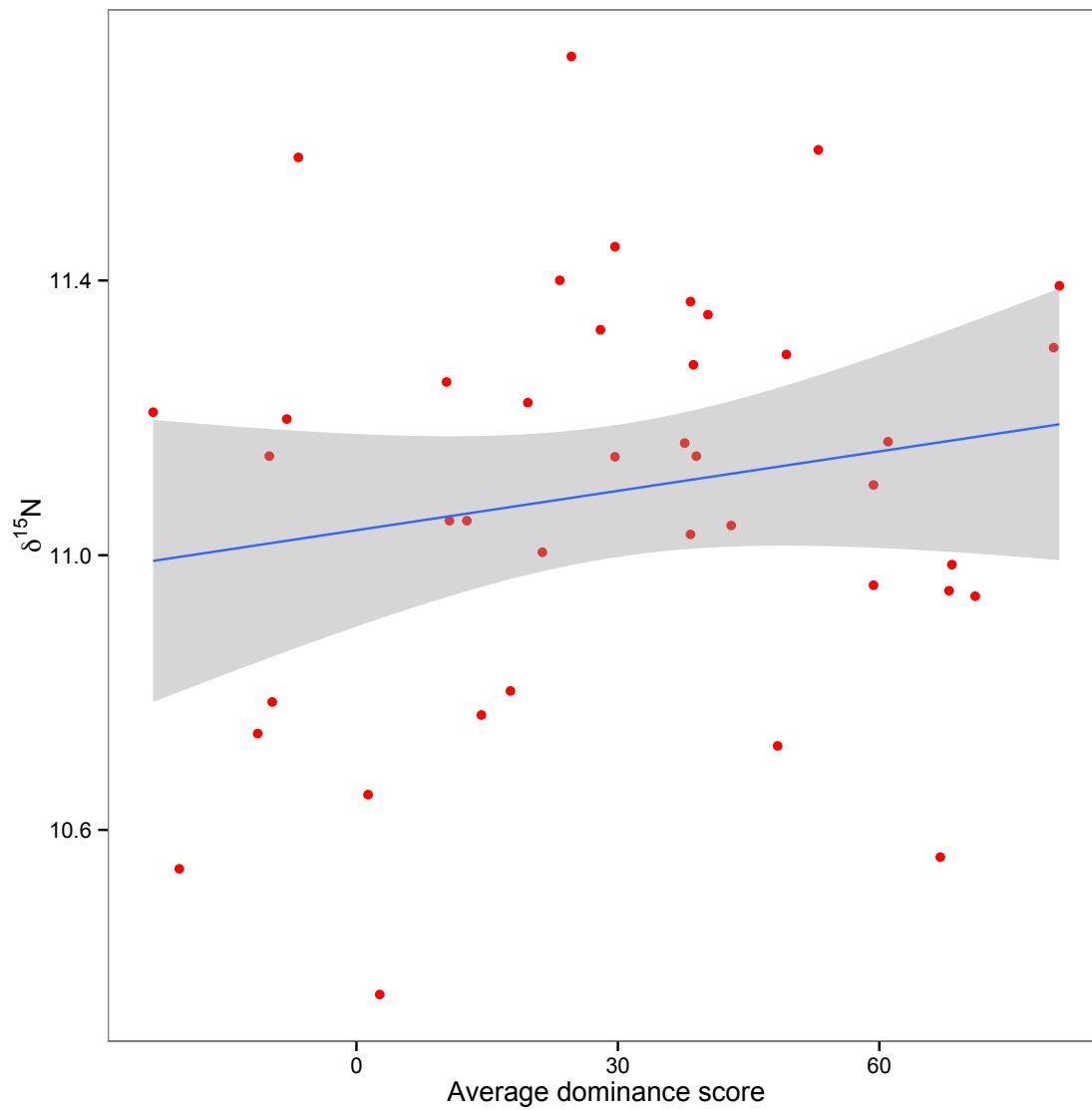
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629 **Figure S2.** Scatterplot (with 95% CI) of the residuals from crayfish dominance and carapace  
630 length regression against calculated trophic position ( $y = 0.00x + 2.34$ ,  $R^2 = 0.01$ ,  $F_{1,38} = 0.54$ ,  $p$   
631  $= 0.47$ ).

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