

Habitat-of-origin predicts degree of adaptation in urban tolerant birds

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Abstract

Urban exploiters and adapters are often coalesced under a term of convenience as ‘urban tolerant’. This useful but simplistic characterisation masks a more nuanced interplay between and within assemblages of birds that are more or less well adapted to a range of urban habitats. Furthermore, cues are generally sought in behavioural ecology and physiology for the degree to which particular bird species are predisposed to urban living. The data in this paper are focused on two assemblages characterised as urban exploiters and suburban adapters from Melbourne, Australia. This study departs from the approach taken in many others of similar kind in that urban bird assemblages that form the basis of the work were identified at the landscape scale and from direct data analyses rather than indirect inference. Further, this paper employs a paired, partitioned analysis of exploiter and adapter preferences for points along the urban-rural gradient that seeks to decompose the overall trend into diagnosable parts for each assemblage. In the present paper I test the hypotheses that the distinct urban exploiter and suburban adapter assemblages within the broad urban tolerant grouping in Melbourne vary in their responses within the larger group to predictor variables, and that the most explanatory predictor variables vary between the two assemblages. In the end, habitat-of-origin better predicts degree of adaptation amongst urban tolerant birds.

Key words

Birds, urban adapter, urban exploiter, urban tolerance, urban-rural gradient, hierarchical Bayesian models, estimating habitat preference, habitat-of-origin

27 **Introduction**

28 The community of ecologists studying urban bird ecology has to a large extent converged
29 on Blair's (1996) typology of 'urban exploiters', 'urban avoiders' and 'suburban
30 adapters', defined by the bird assemblages' biological and behavioural traits (Chace &
31 Walsh 2006; González-Oreja *et al.* 2007; Kark *et al.* 2007; Croci *et al.* 2008). Such
32 assemblages as described here are elsewhere sometimes characterised as 'response
33 guilds' (Leveau 2013). Exploiters and adapters are often coalesced under a term of
34 convenience as 'urban tolerant'. This useful but simplistic characterisation of the urban
35 tolerant subset may mask a more nuanced interplay between and within groups of birds
36 that are more or less well adapted to a range of urban habitats, ranging from the intensely
37 urbanised 'down town' areas of the inner city, out through a fluctuating gradient of
38 generally decreasing urbanisation intensity through the suburbs to the urban fringe. That
39 there are identifiable 'exploiters' and 'adapters' in addition to the 'avoiders' suggests
40 further targeted testing of the urban tolerant grouping may be fruitful in understanding
41 some underlying processes in urban bird ecology.

42 When Kark *et al.* (2007) posed the question "*Can anyone become a urban exploiter?*"
43 they attempted to find the answer in life history traits, phenotypic and behavioural
44 characteristics of individual species. The result was inconclusive, with considerable
45 variability evident in the traits of successful exploiters. They and others concluded that
46 the answer lay in a suite of characters rather than any single factor (Kark *et al.* 2007;
47 Evans *et al.* 2011; Møller 2014). Conole and Kirkpatrick (2011) took a landscape
48 perspective (Snep *et al.* 2009), and found assemblage-specific patterns over a large
49 urbanised landscape in foraging, nesting and mobility guilds that differentiated urban
50 tolerant from urban avoider species, consistent with a number of other studies around the

51 globe (summarised in Chace & Walsh 2006). White *et al.* (2005) identified the non-
52 uniform nature of urban habitats in Melbourne, with species composition between native
53 and exotic birds (within broadly urban tolerant groupings) varying according to habitat
54 structure and floristics. Conole (2011) showed that exotic birds exhibited non-uniform
55 responses to urbanisation in Melbourne in the same way as native birds. Fontana *et al.*
56 (2011) showed that underlying environmental gradients can override the influence of
57 human demographic gradients.

58 A humped distribution of bird species richness has been observed in a number of urban
59 studies, with highest values recorded in the intermediate urbanisation intensity range on
60 the rural-urban gradient (Tratalos *et al.* 2007; Luck & Smallbone 2010; Shanahan *et al.*
61 2014). This pattern has been shown to hold true for all species, but also for urban tolerant
62 species as a subset (Shanahan *et al.* 2014). However, results of earlier data analyses by
63 for this Melbourne study area suggest that the two assemblages within the urban tolerant
64 bird grouping may not show a uniform response trend to urbanisation as has been shown
65 for other cities (Conole 2010; Conole & Kirkpatrick 2010).

66 Gradient analysis has a moderately long history in the relatively young sub-discipline of
67 urban bird ecology (Ruszczyk *et al.* 1987), has been broadly applied in urban ecological
68 studies over the past two decades (McDonnell & Hahs 2008), and much longer in
69 ecology more generally (Whittaker 1967). It is intuitively compatible with a landscape
70 ecology perspective (Snep *et al.* 2009), and despite criticisms of the limitations of
71 gradient analysis as an approach for studying urban ecology, (Catterall 2009; Ramalho &
72 Hobbs 2012a), the potential remains for this approach to be the 'scaffolding' upon which
73 deeper investigations are built (McDonnell *et al.* 2012; Ramalho & Hobbs 2012b).

74 In taking the assemblages identified through gradient analysis (Conole & Kirkpatrick
75 2011) as the basis for this study, I acknowledge the reality that the urban-rural gradient is
76 not simplistically linear (Ramalho & Hobbs 2012a) or neatly concentric around the
77 'down town' centre (Catterall 2009), although in some city layouts it would be correct to
78 assume simple linearity (Santos 2005). Also, these realities do not limit the usefulness of
79 gradient analysis in understanding complexity and nuance in urban bird ecology. I also
80 acknowledge the utility of the urban exploiter/adapter typology, but seek in this paper to
81 deconstruct the concept of 'urban tolerance' for birds, and test a hypothesis which
82 contends that 'urban tolerance' is not monolithic, but multifaceted.

83 The urban tolerance status of birds included in many published studies has been applied *a*
84 *priori*, based on work of others in geographically related systems (such as Kark *et al.*
85 2007), or compiled from secondary or tertiary descriptive sources (such as Bonier *et al.*
86 2007), but see González-Oreja *et al.* (2007). An approach which separates the direct
87 determination of urban tolerance status from the process of analysing species or
88 assemblage responses to urban environmental factors runs the risk of weakening
89 conclusions which may be drawn from such analyses (Conole & Kirkpatrick 2011).

90 It is also the case that many urban bird studies are largely descriptive or narrowly site-
91 specific (Marzluff *et al.* 2001; McDonnell & Hahs 2013), lacking either a theoretical
92 underpinning or focus (Scheiner 2013), and there have been calls to formulate research
93 questions designed to develop a greater mechanistic understanding of the underlying
94 ecological processes operating in urban landscapes (Shochat *et al.* 2006; McDonnell &
95 Hahs 2013), and move towards generalizable concepts (Mac Nally 2000).

96 Part of the process of moving towards generalisable concepts in urban bird ecology
97 involves gaining a better understanding of the extent to which the degree of adaptation to

98 urban environments progresses from intolerance to the high level of adaptation that
99 characterises exploiters. How similar are the responses of the adapters and exploiters to
100 different aspects of the urban-rural gradient?

101 The data in this paper are focused on two assemblages characterised by the author as
102 urban exploiters and suburban adapters from Melbourne, Australia (Figure S1) (Conole
103 & Kirkpatrick 2011). The present study departs from the approach taken in many others
104 of similar kind in that urban bird assemblages that form the basis of the work were
105 identified at the landscape scale and from direct data analyses (Conole & Kirkpatrick
106 2011) rather than indirect inference or *a priori* assignment. Further, this paper attempts a
107 paired, partitioned analysis of exploiter and adapter preferences for points along the
108 urban-rural gradient which seeks to decompose the overall trend into diagnosable parts
109 for each urban tolerant response guild, in a way not previously seen in the literature.

110 In the present paper I test the hypotheses that the distinct urban exploiter and suburban
111 adapter assemblages within the broad urban tolerant grouping in Melbourne vary in their
112 responses within the larger group to predictor variables, and that the most explanatory
113 predictor variables vary between the two assemblages. I also test the hypothesis that
114 habitat-of-origin has predictive utility in determining which urban tolerant birds become
115 exploiters or adapters.

116 **Materials & Methods**

117 Detailed descriptions of the study area and methodology used to derive the urban bird
118 assemblages can be found in (Conole & Kirkpatrick 2011).

119 *Study area and data handling*

120 The study area is metropolitan Melbourne; capital city of the State of Victoria in coastal
121 south-eastern Australia, within a 50 km radius of its Central Business District (Figure S1)

122 (37° 49' S and 144° 58' E). Approximately 220,000 records of birds were extracted from
123 the Birds Australia 'New Atlas of Australian Birds' database (Barrett *et al.* 2003), and
124 intersected with a 1 × 1 km grid (Hahs & McDonnell 2006) to produce a matrix of grid
125 cells by species presence/absence. Species and sites were filtered out according to criteria
126 for representativeness to arrive at a final list of 141 species and 390 cells (Conole &
127 Kirkpatrick 2011).

128 *Environmental and demographic indices*

129 Spatial data on the degree of urbanisation of the study area employed in this study were
130 developed at ARCUE and are discussed in detail by (Hahs & McDonnell 2006); a brief
131 summary of the two selected factors follows. Frequency Greenspace (hereafter
132 greenspace) is the reciprocal of the average amount of impervious surface calculated at
133 the sub-pixel level from the impervious surface fraction image created during the spectral
134 mixture analysis of the 2000 Landsat ETM+ image (Hahs & McDonnell 2006).

135 Combined index ($Index_{combined}$) is the average value of $Index_{image}$ and $Index_{census}$; where
136 $Index_{image}$ is calculated from fraction images produced by the spectral mixture analysis of
137 the 2000 Landsat ETM+ image, and $Index_{census}$ = the total number of people multiplied by
138 the proportion of males employed in non-agricultural work, as enumerated in the 2001
139 census (Hahs & McDonnell 2006).

140 *Data analysis*

141 Statistical analyses were performed in R (R Core Team 2013) using core functions and
142 procedures from the R-packages 'vegan' (Oksanen *et al.* 2013) and 'bayespref' (Fordyce
143 *et al.* 2011). Figures were drawn using R core functions and R-packages 'vegan' and
144 'ggplot2' (Wickham 2009; Oksanen *et al.* 2013).

145 An earlier assemblage analysis (Conole & Kirkpatrick 2011) was the basis for
146 partitioning the total bird data sets for this study; detailed methodology is described
147 therein. Adapter and exploiter species were further partitioned into two new matrices for
148 this study, and separate non-metric multidimensional scaling (NMDS) ordinations
149 performed for each.

150 Species richness of exploiter and adapter species was enumerated for each of 390 grid
151 cells (Conole & Kirkpatrick 2011), along with an index of urbanisation intensity
152 ($\text{Index}_{\text{Combined}}$ – hereafter urbanisation index) and cover of vegetation (greenspace). Data
153 were then modelled as hierarchical Bayesian models using R-package ‘bayespref’
154 (Fordyce *et al.* 2011) to test the preferences of exploiters and adapters for partitioned
155 urban habitats. Model parameters were estimated using a Markov Chain Monte Carlo
156 (MCMC) approach, with 10,000 MCMC steps following a burn-in of 1,000 generations.
157 The parameters estimated in this way are intended to directly address the hypothesis
158 (Fordyce *et al.* 2011), namely that adapter and exploiter bird assemblages show
159 preferences for urban habitat characterised by differing levels of urbanisation intensity or
160 vegetation cover. The hierarchical Bayesian approach has the advantage of directly
161 estimating the parameter of interest (in this case preference for levels of urbanisation or
162 green space by urban tolerant bird assemblages), and models the uncertainty around those
163 parameters as well as allowing comparisons between *a priori* identified groups, in
164 contrast to methods such as ANOVA or t-tests, which assess whether the mean difference
165 is different from zero (Fordyce *et al.* 2011). The estimates are population-level
166 preferences (Fordyce *et al.* 2011).

167 Adapter and exploiter species' habitats of origin were determined by reference to the
168 literature (Marchant & Higgins 1993; Higgins & Davies 1996; Higgins 1999; Schodde &

169 Mason 1999; Higgins *et al.* 2001; Higgins & Peter 2002; Higgins *et al.* 2006), shown in
170 Table 1. The data for cluster analysis consisted of a standard 'r x c' array, with species
171 as rows and habitats of origin as columns (forest, woodland, heath, scrub, urban, farm,
172 air). A Bray-Curtis distance matrix was prepared, and groups of species were formed by
173 hierarchical agglomerative clustering using Ward's algorithm performed on the distance
174 matrix, using core R-function 'hclust' (R Core Team 2013).

175 **Results and Discussion**

176 *Results*

177 In an earlier ordination of all bird species from the Melbourne study, urban exploiters
178 and adapters are shown as overlapping but distinct clusters in ordination space (Figure
179 S2) (Conole & Kirkpatrick 2011). When the exploiters and adapters were partitioned
180 from the avoiders and run as separate ordinations, different pictures of response to urban
181 environmental factors became apparent (Figures 1 & 2).

182 For exploiters the observed species richness vector (S_{obs}) was orthogonal with both
183 greenspace and the urbanisation index (Figure 2). The equivalent vector for adapters
184 (Figure 1) was orthogonal with the urbanisation index, but almost aligned with that for
185 greenspace (Figure 2). Greenspace and the urbanisation index were chosen as
186 representative of structural and demographic aspects of urbanisation intensity even
187 though other parameters were included in the initial analyses, and further analyses were
188 limited to these two factors.

189 The same data plotted as binned boxplots showed that adapter species richness was
190 positively associated with increasing greenspace, but exploiter species richness was flat
191 across the range (Figure 3). Whilst broadly similar trends were evident for both groups
192 as binned boxplots plotted against the urbanisation index (Figure 5), adapters trended to

193 zero species richness at the highest levels, whilst 10 - 15 species of exploiters persisted at
194 the same level. Peak species diversity of urban adapter birds occurred in the middle of
195 the range of urbanisation intensity (Figure 5). Adapter richness peaked at approximately
196 0.8 frequency green-space; exploiters at around 0.55 (Figure 4).

197 The hierarchical Bayesian models for greenspace showed a relatively flat preference by
198 urban exploiters across the range; though increasing preference by urban adapters for
199 higher levels of greenspace (median = 0.46; credible intervals 0.424 – 0.494) almost
200 match exploiter preference (0.54; 0.506 – 0.576) in the highest bin (Figure 4; Table S1).

201 The preferences of urban exploiters and adapters did not overlap in any of the greenspace
202 bins.

203 Hierarchical Bayesian models for the combined index showed a joint preference by urban
204 adapters and exploiters in the middle of the range of the urbanisation index (20.0 – 29.9).
205 Areas of low (0 – 19.9) and high (30.0 – 50.0) urbanisation index were strongly preferred
206 by urban exploiters but not adapters (Figure 6; Table S2).

207 The cluster analysis of adapters and exploiters by habitat of origin returned a dendrogram
208 showing two clear major clusters. All of the adapters clustered together in a woody
209 vegetation habitat group, along with a group of exploiters; five indigenous nectarivores
210 (Red Wattlebird *Anthochaera carunculata* (Shaw 1790), Little Wattlebird *A. chrysoptera*
211 (Latham 1802), White-plumed Honeyeater *Lichenostomus penicillatus* (Gould 1837),
212 Musk Lorikeet *Glossopsitta concinna* (Shaw 1791), Little Lorikeet *G. pusilla* (Shaw
213 1970)), two indigenous avivorous raptors (Australian Hobby *Falco longipennis* Swainson
214 1837, Brown Goshawk *Accipiter fasciatus* (Vigors and Horsfield, 1827)) and two exotic
215 species which are not exclusively synanthropic (Common Blackbird *Turdus merula*,
216 Linnaeus 1758, Common Starling *Sturnus vulgaris* Linnaeus 1758) (Conole 2010). The

217 cluster of exclusively exploiter species were characterised by those originating from open
218 grassy or urban habitats.

219 The boxplots (Figures 3 and 5) and the hierarchical Bayesian models (Figures 4 and 6)
220 showed clear but distinct trends of urban habitat preference by urban exploiter and
221 adapter bird assemblages against two these representative urban habitat measures. The
222 landscape scale preferences of urban adapters and urban exploiters for levels of
223 greenspace never overlap, though they come close tother at the highest values as
224 exploiter preference declines and adapter preference increases. In contrast, landscape
225 preferences for urbanisation intensity measured by the urbanisation index overlap
226 strongly in the middle of the range but are strongly divergent at the lowest and highest
227 values.

228 *Discussion*

229 The diversity of urban adapters on the gradient of urban intensity follows a humped
230 distribution (Figure 5); the trend even more strongly humped when viewed as landscape
231 scale preference (Figure 6). This is consistent with the trend seen for urban tolerant birds
232 in other studies (Tratalos *et al.* 2007). The inverted, humped curve for exploiters is not
233 consistent with the trends for urban tolerant bird species richness seen in other studies
234 (Tratalos *et al.* 2007; Luck & Smallbone 2010; Shanahan *et al.* 2014).

235 This quadratic trend in diversity also resembles that described by the Intermediate
236 Disturbance Hypothesis (IDH), where diversity peaks at a mid point along a gradient of
237 disturbance (Catford *et al.* 2012; Fox 2013). The urban-rural gradient is however not a
238 true analogue of a disturbance gradient. Suburban areas are more stable habitats than
239 either the developing fringe or the intensely re-shaped core of the city, and so disturbance
240 itself shows a quadratic distribution along the urban-rural gradient. Also implicit within

241 IDH is a notion of competition/colonisation trade-off amongst species more or less
242 adapted to disturbed environments, and at least for urban adapted birds it has been
243 suggested that competition is not important (Mikami & Nagata 2013) except for specific
244 cases such as the 'despotic' Noisy Miner (*Manorina melanocephala*) (Kath *et al.* 2009;
245 Maron *et al.* 2013; Robertson *et al.* 2013).

246 The zone of overlap in habitat preference along the human demographic gradient accords
247 broadly with the inner ring of suburbs in Melbourne; long established and heavily
248 vegetated (Hahs & McDonnell 2006). At the extremes of this gradient lie the new
249 suburbs/exurbia at the fringe, and the central business districts ('down town') at various
250 central locations - either lightly vegetated or with largely treeless vegetation (lawns and
251 pasture) (Hahs & McDonnell 2006). The overlap represents depressed preference by
252 exploiters coincident with greatest preference shown by adapters.

253 The response of urban tolerant birds to increasing Frequency Greenspace is consistent
254 with wider trends in other cities (Chace & Walsh 2006). The distinct responses between
255 adapters and exploiters is also less marked with respect to greenspace than urbanisation
256 intensity.

257 The responses of the two assemblages to two simple measures of urban habitat character
258 were divergent, consistent with the study's main hypothesis. Though the larger group of
259 urban tolerant bird species may occasionally be treated as one entity, it is clear from this
260 study and others (Crocchi *et al.* 2008; Catterall 2009; Conole 2011; Conole & Kirkpatrick
261 2011) that the two groups within it are sufficiently distinct in their responses to
262 urbanisation to caution against using pooled data for urban tolerant species in future
263 studies.

264 The response of urban adapter species to the urbanisation index is consistent with what
265 we broadly understand them to be; adapted to suburbanisation (Blair & Johnson 2008).
266 Greenspace typically increases in old suburbs versus the exurban fringe or downtown
267 areas (Hahs & McDonnell 2006). The strong depression in exploiter preference for mid-
268 range urbanisation intensity (versus the extremes) is less expected. At least with the
269 Melbourne data, there is not a single generalised urban tolerant group of birds. The
270 adapters and exploiters share ecological traits with each other but also with avoiders
271 (Conole & Kirkpatrick 2010).

272 In part the contemporary avifauna of an urbanised area is a legacy of the species present
273 in the former landscape, rather than solely being the product of invasion or colonisation
274 (*sensu* Møller *et al.* 2012). As urban areas progressively come to resemble woodland,
275 structurally if not floristically (Kirkpatrick *et al.* 2007), it makes sense that the urban
276 tolerant bird species are likely to include legacy woodland adapted species. Despite the
277 findings of Blair and Johnson (2008) in North American urban areas, it does not appear
278 that suburban areas within a previously forested landscape in Melbourne are *loci* for
279 indigenous woodland bird extirpation or exotic bird invasion (Conole & Kirkpatrick
280 2011). Instead the reverse seems to be true, and they are sites for colonisation and
281 expansion of some indigenous woodland birds (adapters) and where exotic exploiters are
282 less abundant.

283 Exploiters are mostly indigenous species derived from open environments such as
284 grassland and grassy open-woodland (Møller *et al.* 2012), with a small cohort of
285 synanthropic exotic species and indigenous dietary specialists (avivorous raptors,
286 nectarivores) more typical of forest/woodland habitats (Table 1; Figure 7) (Conole &
287 Kirkpatrick 2011). Adapters as a group are all indigenous species of forest, woodland

288 and riparian scrub origins (Table 1; Figure 7), and they have closer affinities with the
289 riparian and bush remnant urban avoiders than the exploiters (Conole & Kirkpatrick
290 2011). It is therefore remnants of the former indigenous avifauna of wooded parts of
291 Melbourne that are the source of the emerging group of urban adapted species, though
292 none are yet as successful as the aptly named urban exploiters. The adapters are
293 essentially the vanguard of a group of semi-specialised bird species that utilise particular
294 niches within urban matrix habitats, but are not yet ubiquitous across the matrix in the
295 way that exploiters are.

296 The responses observed here of each group to both degree of urbanisation and greenspace
297 are largely explained by their ecological histories. The exploiters are able to use
298 disturbed habitats across the matrix analogous to their original habitats, and many of
299 them were established in Melbourne during the early stages of urban expansion and
300 consolidation of the city. As suburban parts of the city became more heavily vegetated
301 and less open, a group of species from analogous riparian/forest habitats became
302 increasingly well established in parts of the city proximate to their source natural
303 habitats. Many parts of the urban matrix are now at or close to the point of saturation
304 with members of the exploiter assemblage due to their ubiquity, but the number of
305 adapter species contributing to bird species richness at points across the matrix is likely
306 to increase on a site by site basis as the process of afforestation of the older suburbs
307 continues. It follows then that the distribution of exploiter species may decline in more
308 established suburban parts of the city over time, though expanding in range and
309 continuing to dominate in developing areas of the city at or near the fringe.

310 The broad linear trends documented by Blair and Johnson (2008) for overall indigenous
311 woodland bird richness falling as urban tolerant bird richness increases holds for this
312 study area too (Conole & Kirkpatrick 2010). However, the partitioning of adapters and

313 exploiters within the urban tolerant grouping in this study reveals the fallacy of assuming
314 uniformity of response of all 'urban tolerant' species, thereby overlooking a key to
315 understanding how habitat origins may be important for understanding how species adapt
316 to urban environments. Other workers have examined the importance of a variable suite
317 of physiological and behavioural traits that may predispose birds to urban adaptability
318 (e.g. Kark *et al.* 2007; Møller 2009; Evans *et al.* 2011). This study has examined the
319 higher order habitat filtering mechanism that may be influential in this regard, and more
320 broadly generalisable as a conceptual model at the scale of the landscape and the
321 assemblage.

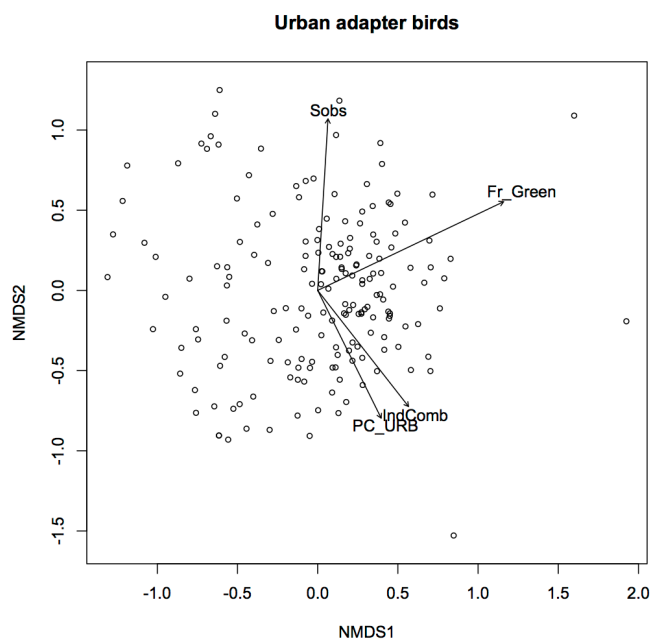
322 **Conclusion**

323 In a similar way to that in which time since establishment has been found to be related to
324 high urban densities of some bird species (Møller *et al.* 2012), spatial or habitat origins of
325 members of bird assemblages influence the degree to which they become urban tolerant;
326 ranging from not at all through to ubiquitous. Bird species that classify as urban tolerant
327 will further classify as either exploiters or adapters according to their habitats of origin.

328 **Acknowledgments**

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333 database. Dr Amy Hahs (ARCUE, University of Melbourne, Australia) provided access
334 to her dataset of remotely sensed landscape metrics. Dr Kath Handasyde (Department of
335 Zoology, University of Melbourne) provided support during the writing of this paper.

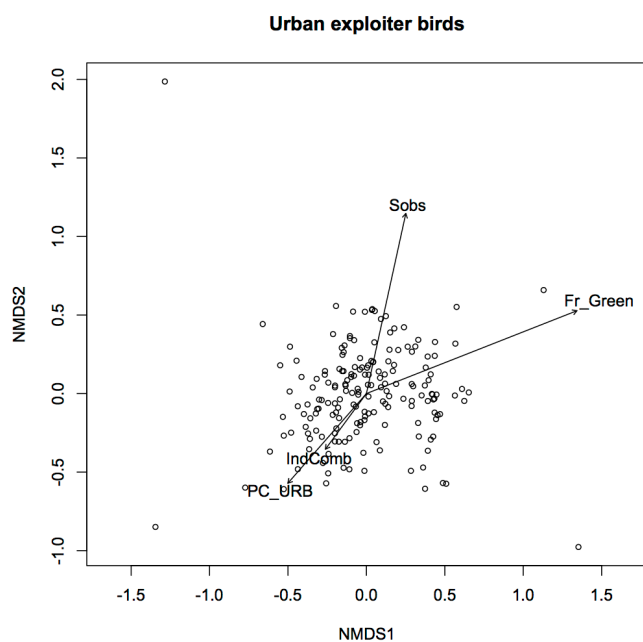
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337 **Figure 1:** NMDS ordination, urban adapters – fitted vectors for which $p \leq 0.01$.

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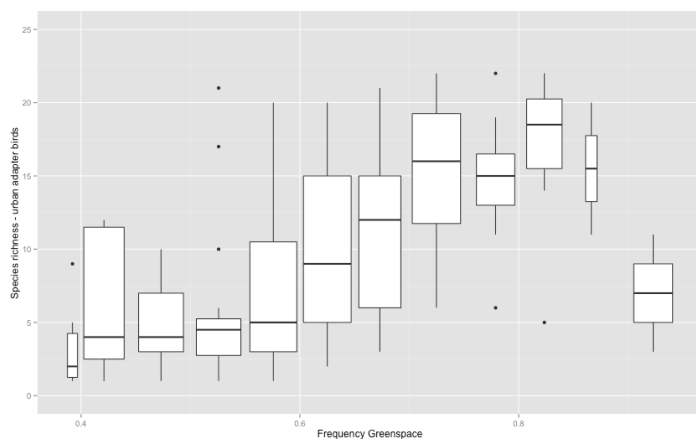


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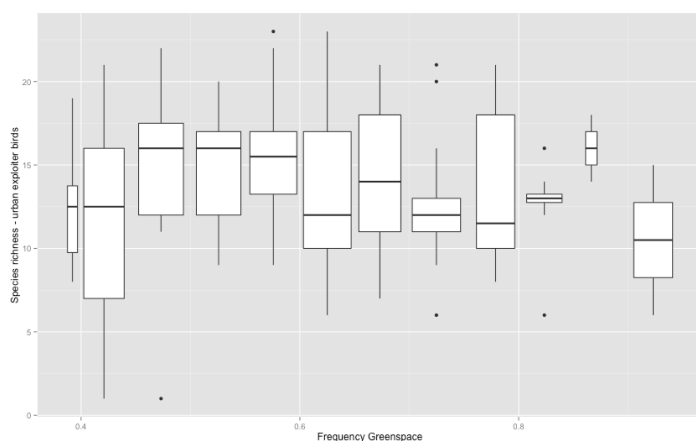
340 **Figure 2:** NMDS ordination, urban exploiters – fitted vectors for which $p \leq 0.01$.

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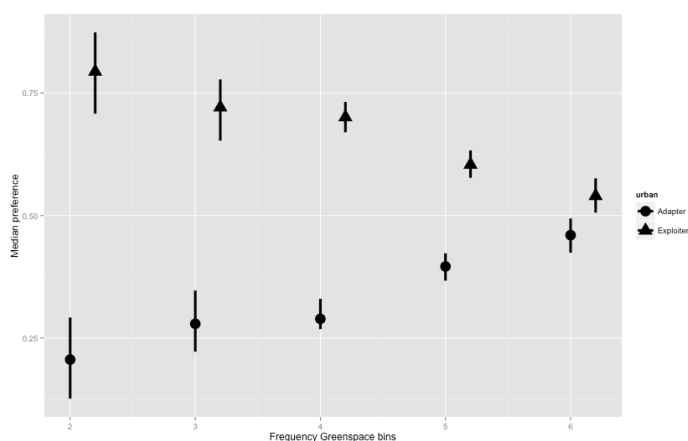
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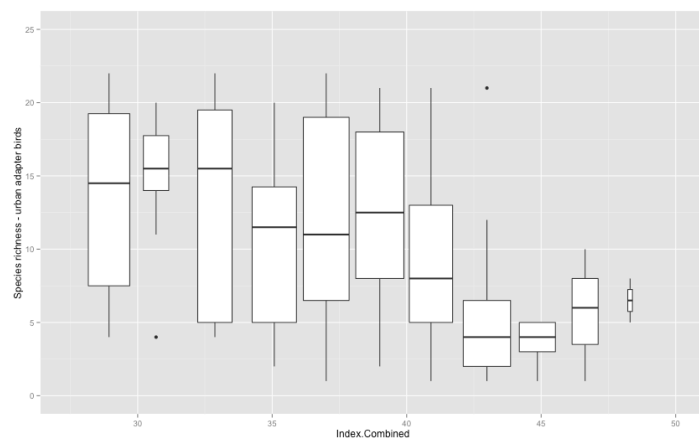
344 **Figure 3:** Species richness of (a) urban adapter and (b) urban exploiter bird species as a
 345 function of the proportion of green space at urbanised sites.

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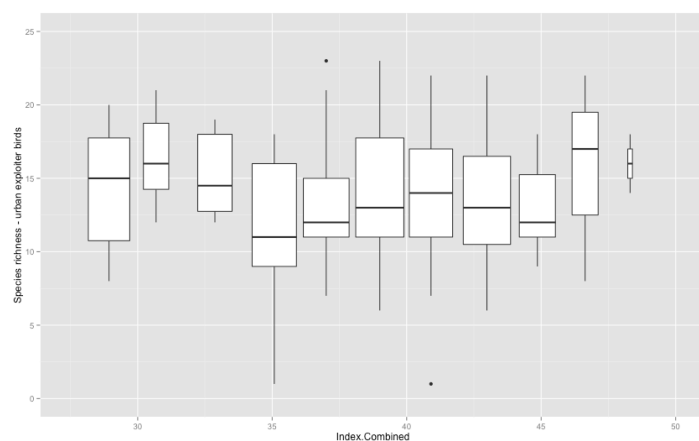


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348 **Figure 4:** Posterior density for landscape-scale preferences of urban adapter and exploiter
 349 bird assemblages (median preference and 95% credible intervals) binned by Frequency
 350 Greenspace at urbanised sites. Posterior densities estimated from 10,000 MCMC steps
 351 following a burn-in of 1,000 generations.

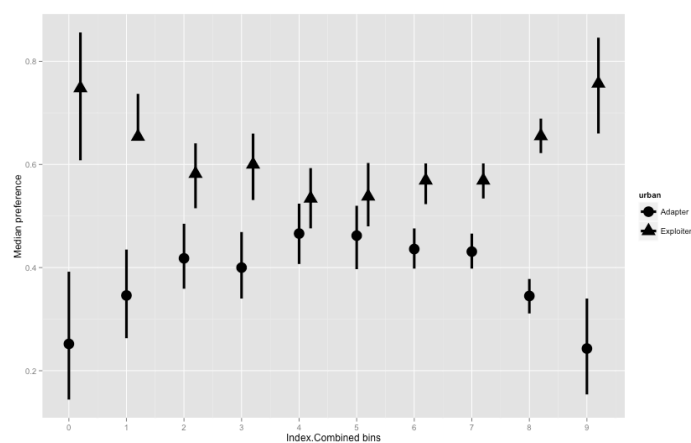


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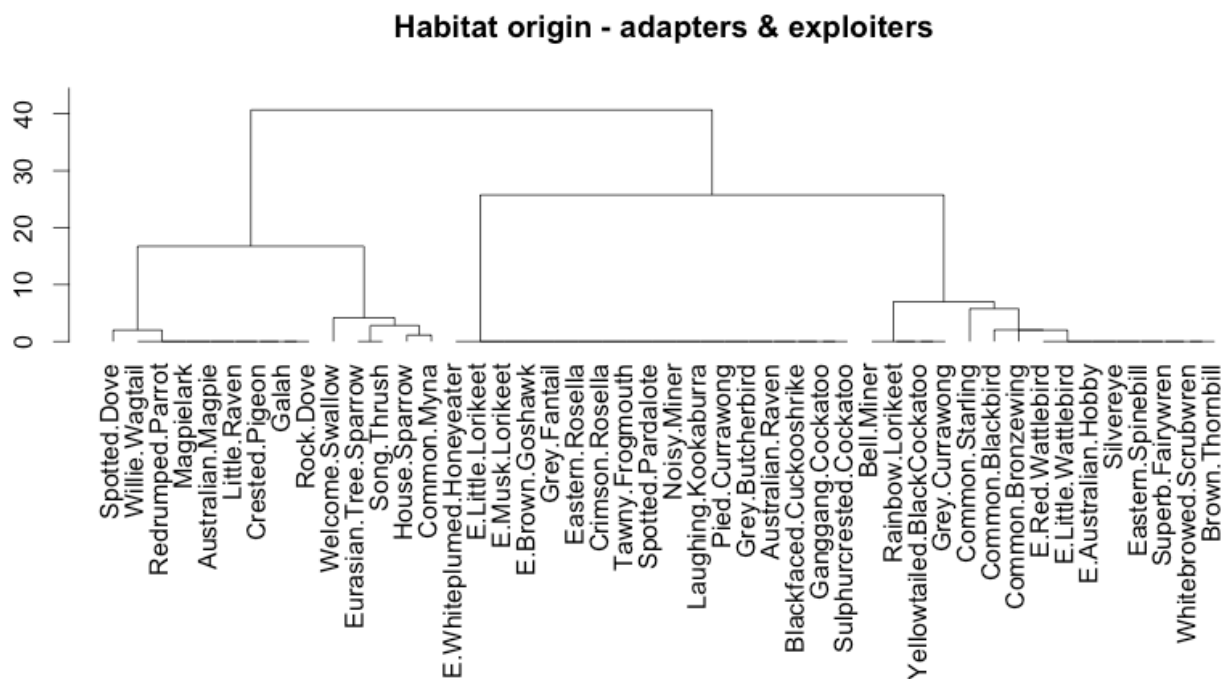
354 **Figure 5:** Species richness of (a) urban adapter and (b) urban exploiter bird species as a
 355 function of urbanisation intensity ($Index_{Combined}$) at urbanised sites.



356

357 **Figure 6:** Posterior density for landscape-scale preferences of urban adapter and exploiter
 358 bird assemblages (median preference and 95% credible intervals) binned by urbanisation
 359 intensity (as $Index_{Combined}$) at urbanised sites. Posterior densities estimated from 10,000
 360 MCMC steps following a burn-in of 1,000 generations.

361



362
 363 **Figure 7:** Cluster dendrogram (Ward method) of adapters and exploiters by habitats of origin.
 364 Exploiters that cluster within the adapters are prefixed with the letter “E”.

365

365 **Table 1:** List of bird species analysed in this study

366

Urban status	Common name	Scientific name	Family	Feral	Original habitat
Adapter	White-browed Scrubwren	<i>Sericornis frontalis</i>	Acanthizidae		Forest, woodland, heath, scrub
Adapter	Brown Thornbill	<i>Acanthiza pusilla</i>	Acanthizidae		Forest, woodland, heath, scrub
Adapter	Yellow-tailed Black-Cockatoo	<i>Calyptorhynchus funereus</i>	Cacatuidae		Forest, woodland, heath
Adapter	Gang-gang Cockatoo	<i>Callocephalon fimbriatum</i>	Cacatuidae		Forest, woodland
Adapter	Sulphur-crested Cockatoo	<i>Cacatua galerita</i>	Cacatuidae		Forest, woodland
Adapter	Black-faced Cuckoo-shrike	<i>Coracina novaehollandiae</i>	Campephagidae		Forest, woodland
Adapter	Common Bronzewing	<i>Phaps chalcoptera</i>	Columbidae		Forest, woodland, scrub
Adapter	Australian Raven	<i>Corvus coronoides</i>	Corvidae		Forest, woodland
Adapter	Grey Butcherbird	<i>Cracticus torquatus</i>	Artamidae		Forest, woodland
Adapter	Pied Currawong	<i>Strepera graculina</i>	Artamidae		Forest, woodland
Adapter	Grey Currawong	<i>Strepera versicolor</i>	Artamidae		Forest, woodland, heath
Adapter	Laughing Kookaburra	<i>Dacelo novaeguineae</i>	Halcyonidae		Forest, woodland
Adapter	Rainbow Lorikeet	<i>Trichoglossus haematodus</i>	Loriidae		Forest, woodland, heath
Adapter	Superb Fairy-wren	<i>Malurus cyaneus</i>	Maluridae		Forest, woodland, heath, scrub
Adapter	Eastern Spinebill	<i>Acanthorhynchus tenuirostris</i>	Meliphagidae		Forest, woodland, heath, scrub
Adapter	Bell Miner	<i>Manorina melanophrys</i>	Meliphagidae		Forest, woodland, scrub
Adapter	Noisy Miner	<i>Manorina melanocephala</i>	Meliphagidae		Forest, woodland
Adapter	Spotted Pardalote	<i>Pardalotus punctatus</i>	Pardalotidae		Forest, woodland
Adapter	Tawny Frogmouth	<i>Podargus strigoides</i>	Podargidae		Forest, woodland
Adapter	Crimson Rosella	<i>Platycercus elegans</i>	Psittacidae		Forest, woodland
Adapter	Eastern Rosella	<i>Platycercus eximius</i>	Psittacidae		Forest, woodland
Adapter	Grey Fantail	<i>Rhipidura albiscapa</i>	Rhipiduridae		Forest, woodland
Adapter	Silveryeye	<i>Zosterops lateralis</i>	Timaliidae		Forest, woodland, heath, scrub
Exploiter	Brown Goshawk	<i>Accipiter fasciatus</i>	Accipitridae		Forest, woodland
Exploiter	Galah	<i>Eolophus roseicapillus</i>	Cacatuidae		Woodland, grassland
Exploiter	Rock Dove	<i>Columba livia</i>	Columbidae	Y	Grassland
Exploiter	Spotted Dove	<i>Streptopelia chinensis</i>	Columbidae	Y	Forest, woodland

Urban status	Common name	Scientific name	Family	Feral	Original habitat
Exploiter	Crested Pigeon	Ocyphaps loquax	Columbidae		Woodland, grassland
Exploiter	Little Raven	Corvus mellori	Corvidae		Woodland, grassland
Exploiter	Australian Magpie	Cracticus tibicen	Artamidae		Woodland, grassland
Exploiter	Australian Hobby	Falco longipennis	Falconidae		Forest, woodland, heath, scrub
Exploiter	Welcome Swallow	Hirundo neoxena	Hirundinidae		Aerial
Exploiter	Musk Lorikeet	Glossopsitta concinna	Loriidae		Forest, woodland
Exploiter	Little Lorikeet	Glossopsitta pusilla	Loriidae		Forest, woodland
Exploiter	White-plumed Honeyeater	Lichenostomus penicillatus	Meliphagidae		Forest, woodland
Exploiter	Little Wattlebird	Anthochaera chrysoptera	Meliphagidae		Forest, woodland, heath, scrub
Exploiter	Red Wattlebird	Anthochaera carunculata	Meliphagidae		Forest, woodland, heath, scrub
Exploiter	Magpie-lark	Grallina cyanoleuca	Monarchidae		Woodland, grassland
Exploiter	House Sparrow	Passer domesticus	Passeridae	Y	Urban, farm
Exploiter	Eurasian Tree Sparrow	Passer montanus	Passeridae	Y	Urban
Exploiter	Red-rumped Parrot	Psephotus haematonotus	Psittacidae		Woodland, grassland
Exploiter	Willie Wagtail	Rhipidura leucophrys	Rhipiduridae		Woodland, grassland
Exploiter	Common Starling	Sturnus vulgaris	Sturnidae	Y	Urban, farm, woodland, heath, scrub
Exploiter	Common Myna	Sturnus tristis	Sturnidae	Y	Urban, farm, woodland
Exploiter	Common Blackbird	Turdus merula	Turdidae	Y	Forest, woodland, heath, scrub, urban
Exploiter	Song Thrush	Turdus philomelos	Turdidae	Y	Urban

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368 Habitat data from (Marchant & Higgins 1993; Higgins & Davies 1996; Higgins 1999;
 369 Schodde & Mason 1999; Higgins et al. 2001; Higgins & Peter 2002; Higgins et al. 2006).

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371 **SUPPLEMENTARY MATERIAL**

372 **Table S1:** Posterior density for landscape-level preferences of urban adapter bird assemblages
 373 (species richness) in Frequency Greenspace bins

Frequency Greenspace		Urban adapter		Urban exploiter	
Bin name	Range	Median preference	Credible intervals	Median preference	Credible intervals
2	20 – 29.9	0.206	0.126, 0.292	0.794	0.708, 0.874
3	30 – 39.9	0.279	0.222, 0.347	0.721	0.653, 0.778
4	40 – 49.9	0.289	0.268, 0.33	0.701	0.67, 0.732
5	50 – 59.9	0.396	0.367, 0.423	0.604	0.577, 0.633
6	60 – 69.9	0.46	0.424, 0.494	0.54	0.506, 0.576

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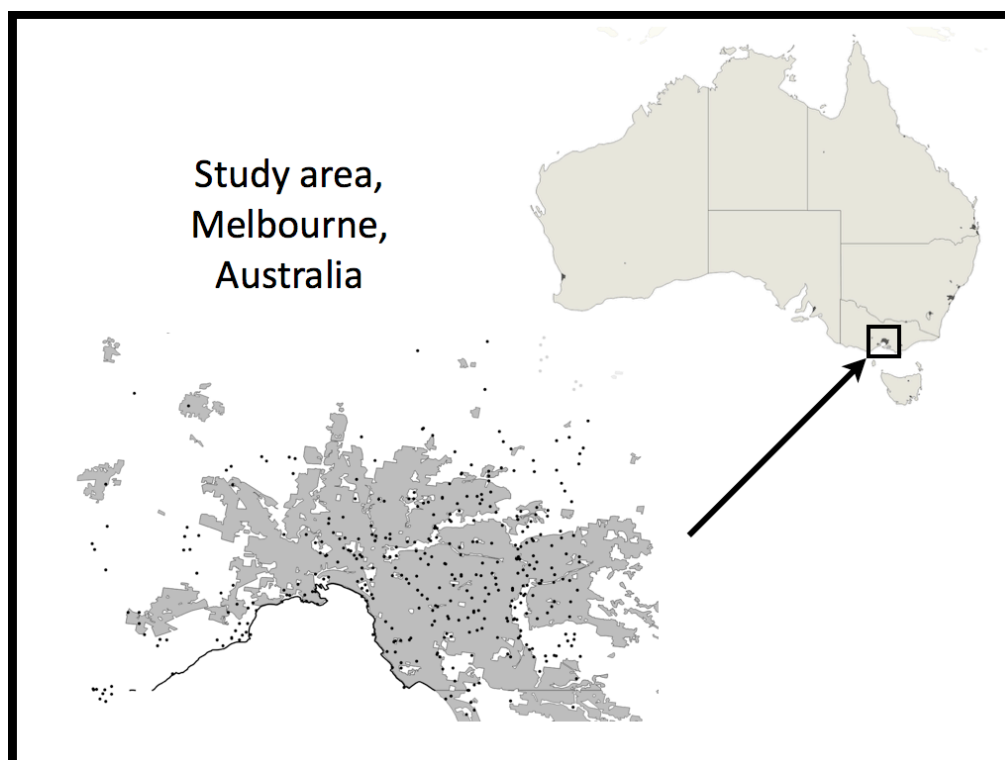
376 **Table S2:** Posterior density for landscape-level preferences of urban adapter bird assemblages
 377 (species richness) in Index_{Combined} bins

Index _{Combined}		Urban adapter		Urban exploiter	
Bin name	Range	Median preference	Credible intervals	Median preference	Credible intervals
0	0 – 4.9	0.252	0.144, 0.392	0.748	0.608, 0.856
1	5.0 – 9.9	0.346	0.263, 0.435	0.654	0.565, 0.737
2	10 – 14.9	0.418	0.359, 0.485	0.582	0.515, 0.641
3	15 – 19.9	0.4	0.34, 0.469	0.6	0.531, 0.66
4	20 – 24.9	0.466	0.407, 0.524	0.534	0.476, 0.593
5	25 – 29.9	0.462	0.397, 0.52	0.538	0.48, 0.603
6	30 – 34.9	0.436	0.398, 0.476	0.569	0.523, 0.602
7	35 – 39.9	0.431	0.398, 0.466	0.569	0.534, 0.602
8	40 – 44.9	0.345	0.311, 0.378	0.655	0.622, 0.689
9	45 – 50	0.243	0.154, 0.34	0.757	0.66, 0.846

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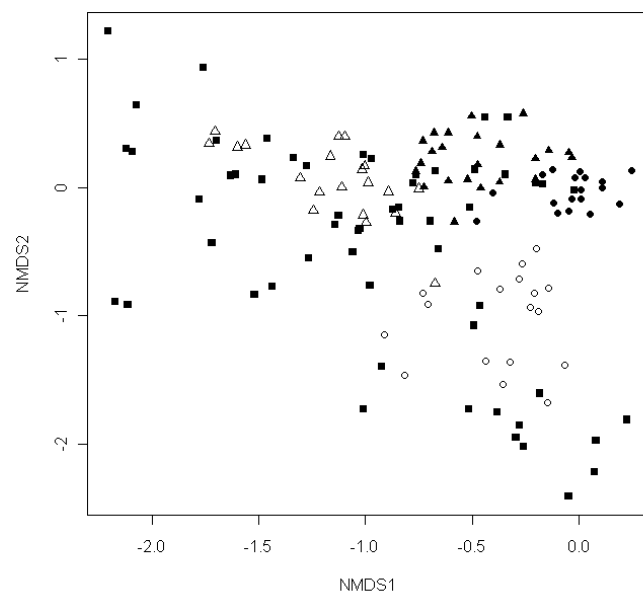
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381 **Figure S1:** Map of Melbourne study area, Australia). Grey areas show built-up areas of
 382 Melbourne. Black circles are survey sites.



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384 **Figure S2:** Plot of non-metric multidimensional scaling ordination results for species. Urban
 385 tolerant species: ▲ = Urban adapters, ● = urban exploiters. Urban avoiders: ■ = Assemblage
 386 4; ○ = Assemblage 3, △ = Assemblage 1 (Conole and Kirkpatrick 2011).

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