

Seedling survival and growth of the invasive species *Prosopis juliflora* (Fabaceae) depend on habitat and seed dispersal mode in the Caatinga dry forest

Clóvis E de Souza Nascimento^{1,2}, Carlos Alberto D da Silva^{3,4}, Inara R Leal⁵, Wagner de S Tavares^{Corresp., 6}, José E Serrão⁷, José C Zanuncio⁸, Marcelo Tabarelli^{Corresp. 5}

¹ Centro de Pesquisa Agropecuária do Trópico Semi-Árido, Empresa Brasileira de Pesquisa Agropecuária, Petrolina, Pernambuco, Brasil

² Departamento de Ciências Humanas, Universidade do Estado da Bahia, Juazeiro, Bahia, Brasil

³ Centro Nacional de Pesquisa de Algodão, Empresa Brasileira de Pesquisa Agropecuária, Campina Grande, Paraíba, Brasil

⁴ Programa de Pós-Graduação em Ciências Agrárias, Universidade Estadual da Paraíba, Campina Grande, Paraíba, Brasil

⁵ Departamento de Botânica, Universidade Federal de Pernambuco, Recife, Pernambuco, Brasil

⁶ Departamento de Fitotecnia/BIOAGRO, Universidade Federal de Viçosa, Viçosa, Minas Gerais, Brasil

⁷ Departamento de Biologia Geral, Universidade Federal de Viçosa, Viçosa, Minas Gerais, Brasil

⁸ Departamento de Entomologia/BIOAGRO, Universidade Federal de Viçosa, Viçosa, Minas Gerais, Brasil

Corresponding Authors: Wagner de S Tavares, Marcelo Tabarelli

Email address: wagner.tavares@ufv.br, mtrelli@ufpe.br

Abstract

Background. Biological invasion is one of the main threats to tropical biodiversity and ecosystem functioning. *Prosopis juliflora* (Sw) DC. (Fabales: Fabaceae: Caesalpinioideae) was introduced in the Caatinga dry forest at early 1940s as a food source for livestock and successfully spread across the region by invading and forming monospecific stands, particularly across former agricultural lands along river banks and flooding plains. As other invasive species, it may benefit from the soils and seed dispersal by livestock but the ultimate forces permitting successful invasion of Caatinga dry forest by this plant needs further studies. The objective was to understand how seed dispersal ecology and soil conditions collectively affect seedling performance and consequently the *P. juliflora* invasive potential.

Methods. Seed germination and seedling performance of *P. juliflora* in 10 plots and its associated soil attributes located across three habitats (floodplain, alluvial terrace and plateau) into a human-modified landscape of the Caatinga dry forest were examined. Seeds of this plant were exposed to four seed dispersal methods: deposition on the soil surface, burial in the soil, passed through cattle digestive tracts and encasing them in cattle manure and passed through mule digestive tracts and mixed with mule manure. The resulting seedlings were monitored through a year and their performance examined with expectancy tables.

Results. Soils differed among habitats, particularly its nutrient availability, texture and water with finely-textured and more fertile soils in the floodplain. Seedling survival rate of *P. juliflora* was highest in the first 30 days and declined between 30 to 60 days with stabilization at 70 days after germination in all seed treatments and habitats. Survival and life expectancy were higher in the floodplain at 75 days and lower in the plateau. *Prosopis juliflora* seedling survival and life expectancy were higher when its seeds were mixed with cattle manure.

Synthesis. *Prosopis juliflora* seedlings are sensitive to water stress and habitat desiccation. Therefore,

they benefit from the humid soils often present across human-disturbed floodplains. This plant also benefits from seed deposition/dispersal by domestic/exotic livestock in these landscapes, since animal manure represents a nutrient-rich and humid substrate for both seeds and seedlings. The quality of the seed dispersal service varies among livestock species, but this key mutualism between exotic species is due to the arillate, hard-coated and palatable seeds. *Prosopis juliflora* traits allow this species to take multiple benefits from human presence and thus operating as a human commensal.

PeerJ

Seedling survival and growth of the invasive species *Prosopis juliflora* (Fabaceae) depend on habitat and seed dispersal mode in the Caatinga dry forest

Clóvis Eduardo de Souza Nascimento^{1,2}, Carlos Alberto Domingues da Silva^{3,4}, Inara Roberta Leal⁵, Wagner de Souza Tavares⁶, José Eduardo Serrão⁷, José Cola Zanuncio⁸, Marcelo Tabarelli⁵

¹ Empresa Brasileira de Pesquisa Agropecuária, Centro de Pesquisa Agropecuária do Trópico Semi-Árido, 56300-970, Petrolina, Pernambuco, Brasil

² Departamento de Ciências Humanas, Universidade do Estado da Bahia, 48900-000, Juazeiro, Bahia, Brasil

³ Empresa Brasileira de Pesquisa Agropecuária, Centro Nacional de Pesquisa de Algodão, 58428-095, Campina Grande, Paraíba, Brasil

⁴ Programa de Pós-Graduação em Ciências Agrárias, Universidade Estadual da Paraíba, 58429-570, Campina Grande, Paraíba, Brasil

⁵ Departamento de Botânica, Universidade Federal de Pernambuco, 50670-901, Recife, Pernambuco, Brasil

⁶ Departamento de Fitotecnia/BIOAGRO, Universidade Federal de Viçosa, 36570-900, Viçosa, Minas Gerais, Brasil

⁷ Departamento de Biologia Geral, Universidade Federal de Viçosa, 36570-900, Viçosa, Minas Gerais, Brasil

⁸ Departamento de Entomologia/BIOAGRO, Universidade Federal de Viçosa, 36570-900, Viçosa, Minas Gerais, Brasil

Corresponding Author:

Wagner de Souza Tavares

Departamento de Fitotecnia/BIOAGRO, Universidade Federal de Viçosa, 36570-900, Viçosa, Minas Gerais, Brasil

E-mail address: wagnermaias@yahoo.com.br

Abstract

Background. Biological invasion is one of the main threats to tropical biodiversity and ecosystem functioning. *Prosopis juliflora* (Sw) DC. (Fabales: Fabaceae: Caesalpinioideae) was introduced in the Caatinga dry forest at early 1940s as a food source for livestock and successfully spread across the region by invading and forming monospecific stands, particularly across former agricultural lands along river banks and flooding plains. As other invasive species, it may benefit from the soils and seed dispersal by livestock but the ultimate forces permitting successful invasion of Caatinga dry forest by this plant needs further studies. The objective was to understand how seed dispersal ecology and soil conditions collectively affect seedling performance and consequently the *P. juliflora* invasive potential.

Methods. Seed germination and seedling performance of *P. juliflora* in 10 plots and its associated soil attributes located across three habitats (floodplain, alluvial terrace and plateau) into a human-modified landscape of the Caatinga dry forest were examined. Seeds of this plant were exposed to four seed dispersal methods: deposition on the soil surface, burial in the soil, passed through cattle digestive tracts and encasing them in cattle manure and passed through mule digestive tracts and mixed with mule manure. The resulting seedlings were monitored through a year and their performance examined with expectancy tables.

Results. Soils differed among habitats, particularly its nutrient availability, texture and water with finely-textured and more fertile soils in the floodplain. Seedling survival rate of *P. juliflora* was highest in the first 30 days and declined between 30 to 60 days with stabilization at 70 days after germination in all seed treatments and habitats. Survival and life expectancy were higher in the floodplain at 75 days and lower in the plateau. *Prosopis juliflora* seedling survival and life expectancy were higher when its seeds were mixed with cattle manure.

Synthesis. *Prosopis juliflora* seedlings are sensitive to water stress and habitat desiccation. Therefore, they benefit from the humid soils often present across human-disturbed floodplains. This plant also benefits from seed deposition/dispersal by domestic/exotic livestock in these landscapes, since animal manure represents a nutrient-rich and humid substrate for both seeds and seedlings. The quality of the seed dispersal service varies among livestock species, but this key mutualism between exotic species is due to the arillate, hard-coated and palatable seeds. *Prosopis juliflora* traits allow this species to take multiple benefits from human presence and thus operating as a human commensal.

Introduction

Biological invasion is an important threat to tropical biodiversity and ecosystem functioning. Increased disturbance by ever-growing human populations will make terrestrial tropical biotas more vulnerable to invasions (Simberloff et al., 2013; Roy et al., 2014; McGeoch et al., 2016). This increased vulnerability is due to alien species benefiting from human disturbances, such as land clearing and soil degradation by agriculture and livestock. Humans and their commensals, whether intentionally or not, favor alien species by creating altered or novel habitats (Almeida et al., 2015; Jauni et al., 2015; Bellard et al., 2016; Malavasi et al., 2016).

The mechanisms driving successful invasion by alien species and those restricting them to particular habitats or conditions are key topics in invasion science (Blackburn et al., 2014; Catford and Jansson, 2014; Hulme, 2015). The life-history strategy or trait package exhibited by alien species affect their competitive performance or adaptability, while environmental conditions, disturbance regimes, habitat degradation and the structure of native communities (e.g. patterns of species richness and functional composition) are external forces controlling invasion success (Gilioli et al., 2014; Goia et al., 2014; Banerjee and Dewanji, 2017). The intrinsic and external factors collectively define the potential for successful invasion and delimit its ecological context, the geographic coverage and potential damage to native biodiversity. Successful invasions only occur when alien species can overcome external forces (Dalmazzone and Giaccaria, 2014; Li et al., 2014; Svenning et al., 2014). These factors explain why a small fraction of introduced alien species become invasive regardless of ecosystem type or habitat integrity (Kalusová et al., 2014; van Wilgen and Richardson, 2014; Novoa et al., 2015).

Successful invasion of tropical plant species usually relies on the key life-history traits across all life-cycle stages (Chapple et al., 2012; Malíková et al., 2012; Mullah et al., 2014), such as vegetative reproduction, massive seed production, effective seed dispersal, high germination success in a wide range of environmental conditions, fast growth and high phenotypic plasticity (Moravcová et al., 2015; van Kleunen et al., 2015; Moran et al., 2017). The effective colonization of human-degraded habitats usually requires dealing with physical stress, particularly reduced soil nutrients and water availability (Boudiaf et al., 2013; Pérez et al., 2015; Rathore et al., 2015). The naturalization and long-term establishment of invasive species depend on integrating adaptive traits as functional strategies (Guo et al., 2018). However, the relative contribution of each trait or strategy vary from case to case (Hulme and Barrett, 2013; Perkins and Nowak, 2013; Rai, 2015).

The Caatinga of northeast Brazil is one of the largest blocks (nearly 1 million Km²) and the world most species-rich seasonally dry tropical forest (Silva et al., 2017). The regional vascular flora reaches 3,000 species with one third endemic, including a myriad of Cactaceae species—making this region the second diversity cactus species center globally (Silva and Souza, 2018; Apgaua et al., 2018; Terra et al., 2018). The Caatinga was inhabited by hunter-gather people for thousands of years (Leal et al., 2005; Mamede and de Araújo, 2008; Silva et al., 2017). The Europeans arrived in the 16th century and the Caatinga dry forest has been converted into human-modified landscapes by a combination of extensive cattle ranching, small-scale subsistence farming and exploitation of forest products such as firewood, fodder, timber, wood for charcoal, fruits and bushmeat (Silva et al., 2014; Leal et al., 2014; Ribeiro et al., 2015; Silva et al., 2017). Nearly 10 million m³ of firewood and charcoal are obtained per year from native species, while goat (i.e. the exotic *Capra aegagrus* subspecies *hircus* L., 1758;

Artiodactyla: Bovidae) herds feeding on native vegetation exceed 16 million heads (Ribeiro et al., 2015; Rito et al., 2017; Sfair et al., 2018). The Caatinga supports around 28 million people, making it one of the most populated semiarid regions and one of the most degraded/vulnerable Seasonally Dry Tropical Forests (SDTFs) globally (Oliveira et al., 2016; Moro et al., 2016; Rito et al., 2016; Schulz et al., 2017; Silva et al., 2017).

The alien flora inhabiting the Caatinga includes 205 species of 135 genera and 48 families, including 61 Poaceae and 33 Fabaceae invasive species (Almeida et al., 2015). This diverse alien flora includes the evergreen tree *Prosopis juliflora* (Sw) DC. (Fabales: Fabaceae: Caesalpinioideae), one of the world top 100 undesired species (Nascimento et al., 2014; Almeida et al., 2015). This species is native to the Caribbean, Central America and South America, but has been intentionally introduced worldwide for its economic value, utility to rural populations and ecological rusticity due to its fast growth, ability to fix nitrogen and tolerance to arid conditions and saline soils (Nascimento et al., 2014; Ilukor et al., 2016; Walkie et al., 2016).

Prosopis juliflora was introduced in the Caatinga at early 1940s as a food source for livestock and successfully spread across the region by invading and forming monospecific stands, particularly across former agricultural lands along river banks and flooding plains (Almeida et al., 2015; Santos and Diodato, 2015; Oliveira et al., 2017). *Prosopis juliflora* forms dense stands, apparently excluding or impelling the reestablishment of species-rich assemblages of tree and shrub species, particularly on river banks degraded by shifting cultivation and livestock overgrazing (Pegado et al., 2006; Andrade et al., 2009; Oliveira et al., 2012). Caatinga dry forests along river banks are structurally complex due the presence of large tree species [e.g., *Libidibia ferrea* (Mart. ex Tul.) L.P. Queiroz, *Piptadenia stipulacea* (Benth.) Ducke (Fabales: Fabaceae) and *Tabebuia aurea* (Silva Manso) Benth. & Hook.f. ex S. Moore (Lamiales:

Bignoniaceae)] and support a diverse flora, including endemic species such as the cactus *Cereus jamacaru* DC. and *Pilosocereus gounellei* (F.A.C. Weber ex K. Schum.) Luetzelb. (Caryophyllales: Cactaceae). *Prosopis juliflora*, as other invasive species, may benefit from the soils and seed dispersal by livestock (Kebede and Coppock, 2015; Pasha et al., 2015; Alvarez et al., 2017) but the ultimate forces permitting successful invasion of Caatinga dry forest by this plant needs further studies (Shackleton et al., 2015; Abdulahi et al., 2017; Naudiyal et al., 2017).

The seed germination, seedling survival and life expectancy in *P. juliflora* across the three main habitats covered by Caatinga dry forest were studied to understand how soil conditions and seed dispersal ecology collectively affect seedling performance and consequently the invasive potential of *P. juliflora*. Seeds exposed to four seed dispersal treatments were deposited across a floodplain, alluvial terrace and plateau, with seed and seedling fate being monitored for a year. The potential mechanisms behind differential seedling performance, especially a positive synergism between human disturbance, key ecological services provided by livestock and the life-history traits exhibited by *P. juliflora* was studied.

Material & Methods

Study site. The study was carried out in 10 sites across a 3.5-Km² human-modified Caatinga landscape (9°00' S, 40°13' W; 377 m altitude) covering three habitats: floodplain, alluvial terrace, and plateau (Supplementary materials 1 and 2). This landscape is typical of the Caatinga, with farming households devoted to livestock production with animals raised extensively and feeding on the Caatinga vegetation plus small patches devoted to subsistence agriculture and remnant patches of Caatinga dry forest; i.e. the traditional Caatinga land use (Sampaio and Costa, 2011). The landscape stretches over sedimentary basins, mountains, plateaus and ravines

covered by cambisols, eutrophic podzols, lithosols, non-calcic brown soils and planosols along the São Francisco River valley (Razanamandranto et al., 2004). The river terrace has alluvial deposits from the valley slopes with sedimentary clay, sandy or silty material in stratified silt layers (Miranda et al., 2014). The floodplain, with slopes between 0 and 2° (Miranda et al., 2011), consists of recent sediments from terraces (Babawi et al., 2016). The alluvial terrace, also called the slope, consists of flat areas or benches, usually situated above the river level, with gravel or thick sediment forming ancient terraces (Mukherjee et al., 2017). The plateau is a flat terrain covered by a sedimentary clay mantle, which spreads following the river terraces (Silva et al., 2008; Ferraz et al., 2013). The regional climate is hot semi-arid, with mean annual temperature of 26.3 °C and relative humidity of 61.7%. The 570 mm annual rainfall is concentrated between January and April (Ramos et al., 2011). The focal landscape was completely covered by Caatinga dry forest prior to European settlement (Silva et al., 2008). Regionally, Cactaceae, Euphorbiaceae and Fabaceae are the most species-rich plant families (Leal et al., 2005; Silva et al., 2008).

Soil attributes. Soil attributes were estimated based on 10 soil samples per habitat type (floodplain, alluvial terrace and plateau). Soil moisture was determined twice a week at 20 cm depth from March 2014 to February 2015 (Francesca et al., 2010). The soil samples were weighed to determine the wet mass (mw, g), dried in an oven at 105 °C for 24 hours and weighed again to determine their dry mass (md, g). Soil moisture was determined by $H (g.g^{-1}) = (md - mh) \div md$ or $H (\%) = (md - mh) \div md \times 100$. Physical (field capacity, sand, silt and clay content) and chemical soil attributes [soil organic matter, phosphorus, potassium, calcium, magnesium, sodium, aluminum and cation exchange capacity (CEC)] were also obtained via samples at 10 soil samples per habitat type, with samples at 0-20 cm depth (Lopes et al., 2013). Soil analyses were

performed in the Laboratory for Soil, Water and Plant Tissue Analysis at EMBRAPA Semi-Arid according to standard international protocols (Marques et al., 2007; Galindo et al., 2008).

Seed germination and seedling performance across habitats. *Prosopis juliflora* seeds were exposed to four seed dispersal treatments as follows: seeds with artificially broken dormancy deposited on the soil surface (T1); seeds with artificially broken dormancy buried in the soil to a depth of 1.0 cm (T2); seeds that passed through the digestive tract of cattle (*Bos taurus* subspecies *indicus* L., 1758; Artiodactyla: Bovidae) and mixed with cattle manure (hereafter cattle-dispersed seeds) (T3) and seeds that passed through the digestive tract of mules (mare, *Equus ferus* subspecies *caballus* L., 1758 × donkey, *Equus africanus* subspecies *asinus* L., 1758; Perissodactyla: Equidae) and mixed with mule manure (hereafter mule-dispersed seeds) (T4). These four seed dispersal treatments resemble natural seed dispersal modes and fates in the Caatinga, as ripe fruits can be (1) consumed by livestock and deposited on the soil surface immersed in manure; (2) remain intact on the ground and as the pod rots, seeds are deposited on the soil surface and (3) carried by runoff and deposited in soil sediments (dos Santos et al., 2006). These fates need further studies, although ripe pods are consumed by livestock including cattle, goats, horses and mules (Nascimento et al., 2014). Seeds were manually collected from mature fruits from several trees collected on the ground throughout 2013. Seeds were used to feed livestock first and their manure collected.

Prosopis juliflora seeds were randomly sown on 120 1.4 m²-field plots across the three habitats (40 plots per habitat, Supplementary material 1). One hundred seeds were sown per treatment and habitat, totaling 12,000 seeds (four treatments × three habitats × 40 replicates= 25 seeds per replicate). Seed dispersal treatments were set up in mid-March (i.e. rainy season) of 2014. Seeds were collected from several trees and animal dungs in the year preceding sowing

experiments. Seed dormancy was broken artificially by making a small incision with blunt scissors on the opposite side of the micropylar region, while those that had passed through the animal digestive tract were extracted from dried dung (Nascimento et al., 2014). Seeds obtained from trees were sown at 10×10 cm spacing and those from the manure separated, counted, mixed with 500 mL of the respective animal dung and sown at 10×10 cm spacing. Seed germination and seedling survival were evaluated for each of the three habitats every five days during the first 30 days and every 15 days thereafter for one year. *Prosopis juliflora* germination is epigeal (Dube et al., 2010) and sprouted seeds (with cotyledons) were marked with toothpicks and protected with wire netting to prevent animal damage. Seeds with their first cotyledon open were classified as seedlings.

Life expectancy and survival tables. Life expectancy and survival tables of *P. juliflora* were calculated (Bogino and Jobbágy, 2011) for each dispersal treatment and habitat considering age classes of 15 days ($x=15$). The number of *P. juliflora* survivors at the beginning of each age class (L_x) and at the germination and seedling phenological stages were estimated based on survival (l_x) in these stages (Caswell, 1996) as follow: (a) number of dead individuals during the age class x (dx): $dx = L_x - L_{x+1}$; (b) mortality ratio for the age class x (qx): $qx = dx \div L_x$; (c) survival rate during age x (sx): $sx = 1 - qx$; (d) age structure (Ex) as the number of alive plants from each

age class: $E_x = (L_x + L_{x+2}) \div 2$; (e) cumulative number of living individuals (T_x) by: $T_x = \sum_{j \geq x}^y E_x$

where j represents any age greater than or equal to that of class x ; and (f) life expectancy for individuals of age class x (ex): $ex = T_x \div L_x$.

Data analysis. Differences in soil attributes were examined via analysis of variance and Tukey's Honest Significant tests ($p < 0.05$) (Tukey, 1949). Non-normal (non-parametric) data obtained

from the three habitats were compared using the Kruskal-Wallis test statistic ($p < 0.05$) (Wallis, 1952) using the BioEstat 5.0 program (Ayres et al., 2007).

Results

Soil physical attributes. Caatinga habitats differed greatly in terms of soil physical attributes. Field capacity was higher in the floodplain ($F = 8.5454$, $df = 2$, $p = 0.0016$; Fig. 1A), with the site and the terrace supporting higher soil moisture ($H = 12.5507$, $df = 2$, $p = 0.0019$) than the plateau (Fig. 1B). Soil clay content was higher in the floodplain followed by the plateau and the alluvial terrace ($H = 9.0142$, $df = 2$, $p = 0.011$). The silt score was highest in the floodplain than in the alluvial terrace and plateau ($F = 6.6178$, $df = 2$, $p = 0.0048$). The plateau and alluvial terrace had the highest sand fraction ($F = 8.2452$, $df = 2$, $p = 0.0019$; Fig. 2A).

Soil chemical attributes. Chemical soil attributes differed across habitats with only the sodium content being similar between them ($H = 5.6227$, $df = 2$, $p = 0.0601$). The phosphorus, calcium, magnesium, aluminum, potassium and CEC concentrations were similar in the floodplain and alluvial terrace and differed from those in the plateau ($p < 0.05$). The organic matter content in the floodplain and plateau was similar but lower than that in the alluvial terrace ($H = 17.8655$, $df = 2$, $p = 0.0001$; Fig. 2B). Moreover, the soil organic matter content in the plateau was extremely variable.

Seedling survival and life expectancy. *Prosopis juliflora* seedling survival was higher in the first 30 days after germination, declined steeply from 30 to 60 days and tended to stabilize 70 days after germination for all seed dispersal treatments and habitats (Figs. 3–5). This plant survival peaks in the floodplain were higher among seedlings from cattle-dispersed seeds, followed by mule-dispersed and those deposited in the soil surface. Life expectancy was highest

approximately 75 days after germination in all seed treatments. The life expectancy in this habitat reached 165 days among seeds deposited on the soil surface and cattle-dispersed ones, but with 210 days among those buried in the soil and those dispersed by mules. Moreover, seedling life expectancy exhibited peaks at 180, 195, 210 and 255 days with buried seeds, cattle-dispersed, seeds on the soil surface and mule-dispersed seeds, respectively. Life expectancy of *P. juliflora* reached 345 days across all treatments (Fig. 3; Supplementary material 3).

The *P. juliflora* survival peak in the terrace was higher among seedlings from cattle-dispersed seeds, followed by those dispersed by mules and with lowest peak by those from seeds buried in the soil. The longest life expectancy occurred within the first 15 days after sowing, without increase from this period onwards. The life expectancy in the alluvial terrace reached 150 days for all seed treatments, except for those buried in the soil surface, which reached 165 days (Fig. 4; Supplementary material 4).

The survival peak for *P. juliflora* seedlings in the plateau was highest for those that were cattle-dispersed followed by the mule-dispersed ones. Seedlings from soil surface seeds or buried ones exhibited no survival peaks. The life expectancy peak among seedlings from soil surface or buried was up to 15-day old, while this was up to 30- and 60-day among seedlings from mule- and cattle-dispersed seeds, respectively. *Prosopis juliflora* seedling life expectancy reached 45 days among those deposited on the soil surface and dispersed by mules. However, it reached 120 and 165 days among seeds buried and cattle-dispersed, respectively (Fig. 5; Supplementary material 5).

Discussion

The main habitats covered by the Caatinga dry forest differ in soil attributes or conditions. Overall soils along floodplains and terraces have finer texture and are more humid and fertile than those on the plateau, which are sandy and xeric. The way *P. juliflora* seeds are dispersed and deposited (i.e. seed dispersal treatments) appears to respond to soil conditions and greatly influences seedling performance across the main Caatinga habitats. The chances of seedling survival when coated in manure and across floodplains are higher, while those deposited on the soil surface on the plateau have a lower survival probability. The seedling life span was longer from those originated with seeds immersed into cattle manure. Moreover, the benefit for seedling dispersal depends on animal species (cattle vs. mule) and the habitat. Interactions between soil attributes and seed dispersal by exotic species apparently define the *P. juliflora* potential as an invasive species across human-modified Caatinga landscapes. The dominant species *Stipa breviflora* Griseb. (Poales: Poaceae) expansion potential depends on climate changes in Chinese temperate grasslands (Lv and Zhou, 2018). Extreme climatic change in Mediterranean rivers is suggested to promote the disappearance of the pioneer and young succession stages of riparian woodlands (Rivaes et al., 2013), particularly evident in Europe (Rivaes et al., 2014).

Soil differs across habitats, through which *P. juliflora* has been documented to exhibit extreme variations in establishment and stand density in the Caatinga dry forest (Bailis and McCarthy, 2011). *Prosopis juliflora* stands have been reported with densities as high as 140.39 stems per ha across floodplains and alluvial deposits at the Keoladeo National Park in Rajasthan, India (Mukherjee et al., 2017) and in Tamil Nadu, India (Gandhi and Pandian, 2014). Such dense monospecific stands suggest soil conditions and seedling ecology as potential key drivers for this invasive species as previously documented for other regions in which *Prosopis* species are invasive (Ansley et al., 2018). *Prosopis* species largely benefit from seed dispersal by domestic

animals via dung pellets, improving seed germination and seedling performance and leading to higher life expectancy (Miranda et al., 2011, 2014; Razanamandranto et al., 2014). Seeds deposited on the soil surface without manure had the worse seedling performance and survival in our focal landscape; similar to that documented across a variety of habitats and biotas (Babawi et al., 2016; Mukherjee et al., 2017).

Prosopis juliflora as an invasive plant species benefiting from dispersal services offered by exotic animals is far from novel, with a number of examples for species of this genus (Abbas et al., 2018). Ingestion and deposition by livestock is not required for *Prosopis* seeds to germinate, but it improved seedling performance and long-distance seed dispersal of this plant as it has a patchy distribution and no native vertebrate species have been found to disperse *P. juliflora* seeds in the Caatinga region. Seed dispersal by native ungulates [(i.e. deers, *Mazama americana* Erxleben, 1777 and *Mazama gouazoubira* Fischer, 1814 (Artiodactyla: Cervidae)] is theoretically possible but large vertebrates have been intensively extirpated in the Caatinga region (Bragagnolo et al., 2019; de Oliveira et al., 2019). Caatinga supports nearly 30 million cattle, goats and mules (Sampaio and Costa, 2011; Santos et al., 2017) as a large number of active seed dispersers which spread through diverse habitat types, from old-growth forest stands to degraded areas (Sampaio and Costa, 2011). Livestock frequently access humid habitats such as river banks on their daily search for fresh water and native forage (Dias et al., 2019; Lopes et al., 2019). Moreover, fodder availability is seasonal and scarce (Costa et al., 2016), while *P. juliflora* fruits and foliage represent a permanent and appreciable food source for livestock, likely stimulating its fruit consumption.

Prosopis has relatively hard pods and its sweet-arillate seeds probably evolved as a response to consumption by native ungulates [e.g. cattle, goats, horses, mules and sheep *Ovis*

aries L., 1758 (Artiodactyla: Bovidae)], as an adaptation for contemporary consumption and seed dispersal by domestic livestock (Nascimento et al., 2014; Almeida et al., 2015). Although a substantial seed portion may be damaged or digested when consumed by livestock, the benefits from dispersal outweigh this drawback. This may be particularly true in the Caatinga as the native flora (mostly abiotically-dispersed) does not benefit from the dispersal services offered by livestock.

Higher nutrient and water content in both floodplain soils and livestock manure may enhance seedling performance of *Prosopis* species. This was suggested for *Prosopis africana* (Guill. & Perr.) Taub. seeds in sheep and zebu manure in the Sudanian savanna (Razanamandranto et al., 2014) and *P. juliflora* in animal manure in the Caatinga in northeast Brazil (Miranda et al., 2011, 2014). Causal mechanisms responsible for differential seedling performance of that plant across seed dispersal modes and habitats were not explicitly examined in our focal landscape. Seeds deposited on the soil surface rarely germinate due to dehydration as observed for *Prosopis ferox* Griseback and *Prosopis flexuosa* DC. across semiarid regions of Argentina and Sudan (Campos et al., 2011; Morandini et al., 2013; Yoda et al., 2015). High *Prosopis caldenia* Burkart seed mortality has also been related to soil water availability reduction across semiarid rangelands of Central Argentina (de Villalobos et al., 2005; de Villalobos and Peláez, 2015; Risio et al., 2016).

Fine-textured soils with greater silt and clay quantities, as observed in the floodplain, are expected to increase water retention and storage capacity, and thus may increase plant survival. A similar phenomenon was observed for *P. caldenia* and *P. glandulosa* since allometric growth and survival was directly proportional to the water availability along a topographic gradient in the Chihuahuan Desert, Mexico, in the semiarid rangelands of Central Argentina and in the arid

and semiarid ecosystems of the United States of America (USA), respectively (Martínez and López-Portillo, 2003; de Villalobos et al., 2005; Maestre and Reynolds, 2006). Water storage capacity (Arcone et al., 2008) and organic matter, phosphorus and calcium accumulation are usually higher in fine-textured soils with greater silt and clay quantities, found mainly in gallery forests (Jiménez et al., 2008), which gives these soils a greater CEC (Shiferaw et al., 2004; Silva et al., 2008). In synthesis, *P. juliflora* seeds and seedlings are sensitive to water stress and habitat desiccation.

The higher nutrient availability in the floodplain and terrace, along with the low aluminum content in the floodplain, may also be important because excess of this element in the soil can impair the calcium, magnesium, phosphorous and potassium absorption by *Prosopis* plants, as reported for *P. juliflora* in the Caatinga (Jiménez et al., 2008; Oliveira et al., 2017). However, calcium and magnesium counteract this detrimental effect as observed for *Prosopis alba* Griseb., *Prosopis pubescens* Benth. and *Prosopis ruscifolia* G. in the central and western Gran Chaco phytogeographical region, Argentina, and in El Paso, Texas, USA, respectively (Velarde et al., 2005; Meloni et al., 2012; Zapala et al. 2014). The soil nutrient availability improved plant vigor, making them less vulnerable to drought stress in our focal landscape is similar to that reported for *Prosopis cineraria* (L.) Druce (Ghaff) in United Arab Emirates (Gil and Al-Shankiti, 2015; Song et al., 2017). Nutrient-rich soils increase the *P. juliflora* seedling establishment, allowing its embryos to grow quickly (Nasr et al., 2012; Patil and Karadge, 2012; El-Keblawy and Abdelfatah, 2014).

The high nutrient and water availability in animal manure also apparently creates suitable microhabitats for seeds and seedlings. This substrate is favorable during the critical germination and emergence periods, especially considering the spatially erratic rainfall in the Caatinga even

in the rainy season (Sampaio and Costa, 2011). Such services provided by livestock manure benefit the *Prosopis* seeds of different species from South America to the Middle East and Africa (Majd et al., 2013; Westphal et al., 2015; Araujo et al., 2017). On the other hand, the rapid dehydration and hardening of animal manure reduced seed germination of *P. glandulosa* in the arid and semiarid regions of northeastern Mexico and in the drylands of Japan (Garza et al., 2013; Abdalla et al., 2017). The lower soil moisture in the plateau of the Caatinga accelerated the drying and hardening process of cattle and mule manure (CE de SN, personal observation). This adversely affected the *P. juliflora* germination and emergence and contributed to its lower survival and life expectancy compared to the floodplain and terrace habitats.

The higher survival rate and life expectancy of *P. juliflora* seedlings germinated from seeds mixed with cattle manure compared to those mixed with mule manure in the plateau can be attributed to differences in the composition of these animal dungs (Campos et al., 2008; Nascimento et al., 2014). Mule manure has a lower water content and higher fibrous sheath content encasing seeds (Gonçalves et al., 2013). The fibrous sheath covering the seeds probably reduces seed imbibition due to the fact that these animals are not ruminants like cattle (Gonçalves et al., 2013). Both exotic and native ungulates can favor invasive plant species by (1) creating disturbed habitats via feeding, trampling and movement, (2) controlling or eliminating palatable native species and (3) dispersing seeds via endozoochory and epizoochory (Vavra et al., 2007). The colonization and direct seed dispersal services provided by domestic ungulate livestock was confirmed, and evidences for improved germination and seedling performance associated with fruit consumption and the seed dispersal, coated in protective manure, to suitable habitats was also confirmed.

Caatinga dry forest has been converted to small-scale farming for the agriculture and livestock subsistence, as found in other dry forests and savannas globally (Miccolis et al., 2017; Pérez-Marin et al., 2017). Farming has a continuous demand for the intentional introduction of useful alien species such as cattle, goats and the multiple-use shrub *P. juliflora*, which has now established throughout the Caatinga (Nascimento et al., 2014; Almeida et al., 2015; Ilukor et al., 2016; Walkie et al., 2016). Seed and seedling sensitivity to water stress and desiccation limits the *P. juliflora* proliferation and restricts its establishment to humid/more fertile soils even when its seeds are coated in livestock manure. *Prosopis juliflora* has spread throughout the Caatinga region establishing monospecific stands across floodplains, river banks and sedimentary deposits despite such a life-history limitation. This invasion relies on a close mutualism between human populations and their commensals as agriculture provided degraded habitats, affect soil conditions across river banks via deposition of fine material, while livestock provide seed dispersal services due to the presence of arillated and hard-coated seeds; i.e. a high-quality seed dispersal service that is not available for the native flora in the Caatinga, thereby conferring an adaptive advantage to *Prosopis* species. *Prosopis juliflora* is expected to spread in the Caatinga with dense populations locally because the rural populations continue to convert Caatinga dry forest into degraded habitats (Sfair et al., 2018). The mutualism with human populations favors *P. juliflora*, as found for other invasive *Prosopis* species across semiarid rangelands globally (Busso et al., 2013). A general perspective on how some *P. juliflora* life-history traits confer invasive capacity has been defined, but other traits (e.g. resprouting capacity, seedling tolerance to herbivory and trampling, allelopathic abilities) enabling *P. juliflora* to become a human commensal deserves further investigation.

Conclusions

The *Prosopis juliflora* invasion capacity promoted by its seedling survival and growth depend on habitat and seed dispersal mode in the Caatinga dry forest as this species benefits from human-disturbed, humid soils and manure-involved seeds provided by livestock. This perspective highlights plant-animal interactions among exotic species as potential driver for successful invasion, proliferation and habitat distribution of invasive plants in human-modified landscapes or rangelands.

Acknowledgments

To Asia Science Editing (<http://www.asiascienceediting.com/>) of Republic of Ireland and Kieran Withey (Lancaster University in Lancaster, United Kingdom) for English editing and correction of an early and final version of this manuscript, respectively. Three anonymous Referees offered constructive criticism.

References

- Abbas AM, Mancilla-Leytón JM, Castillo JM. 2018. Can camels disperse seeds of the invasive tree *Prosopis juliflora*? *Weed Research* 58,221–228. DOI 10.1111/wre.12298
- Abdalla MA, Elkarim AHA, Taniguchi T, Endo T, Yamanaka N. 2017. Phytoremediation of calcareous saline-sodic soils with mesquite (*Prosopis glandulosa*). *Acta Agriculturae Scandinavica, Section B — Soil & Plant Science* 67,352–361. DOI 10.1080/09064710.2017.1281432
- Abdulahi MM, Ute JA, Regasa T. 2017. *Prosopis juliflora* L: Distribution, impacts and available control methods in Ethiopia. *Tropical and Subtropical Agroecosystems* 20,75–89.

- Almeida WR, Lopes AV, Tabarelli M, Leal IR. 2015. The alien flora of Brazilian Caatinga: Deliberate introductions expand the contingent of potential invaders. *Biological Invasions* 17,51–56. DOI 10.1007/s10530-014-0738-6
- Alvarez M, Leparmarai P, Heller G, Becker M. 2017. Recovery and germination of *Prosopis juliflora* (Sw.) DC seeds after ingestion by goats and cattle. *Arid Land Research and Management* 31,71–80. DOI 10.1080/15324982.2016.1234521
- Alves FC, Rossetti DF. 2017. Late Holocene coastal dynamics and relative sea-level changes in the littoral of Paraíba, northeastern Brazil. *Progress in Physical Geography: Earth and Environment* 41,375–392. DOI 10.1177/0309133317709744
- Ansley RJ, Zhang T, Cooper C. 2018. Soil moisture, grass production and mesquite resprout architecture following mesquite above-ground mortality. *Water* 10,1243. DOI 10.3390/w10091243
- Apgaua DMG, Pereira DGS, Santos RM, Menino GCO, Pires GG, Fontes MAL, Tng DYP. 2018. Floristic variation within seasonally dry tropical forests of the Caatinga Biogeographic Domain, Brazil, and its conservation implications. *International Forestry Review* 17,33–44. DOI 10.1505/146554815815834840
- Araujo MER, Pérez DR, Bonvissuto GL. 2017. Seed germination of five *Prosopis* shrub species (Fabaceae-Mimosoideae) from the Monte and Patagonia phytogeographic provinces of Argentina. *Journal of Arid Environments* 147,159–162. DOI 10.1016/j.jaridenv.2017.07.019
- Arcone S, Grant S, Boitnott G, Bostick B. 2008. Complex permittivity and clay mineralogy of grain-size fractions in a wet slit soil. *Geophysics* 73,J1–J13. DOI 10.1190/1.2890776

- 454 Arruda DM, Schaefer CEGR, Moraes MLB. 2015. Relationship between soil properties and
455 vegetation in the ecotone region of the Middle São Francisco River Valley, Brazil.
456 *Revista Brasileira de Ciências do Solo* 39,1524–1532. DOI
457 10.1590/01000683rbc20150241
- 458 Ayres M, Ayres Júnior M, Ayres DL, Santos AS. 2007. BioEstat: Aplicações estatísticas nas
459 áreas das ciências biológicas e médicas. Versão 5.0. Belém, Pará: Sociedade Civil
460 Mamirauá, MCT-CNPq. 364 p.
- 461 Babawi FF, Campbell SD, Mayer RJ. 2016. Seed fall, seed predation, twigging and litter fall of
462 *Cascabela thevetia* (L.) Lippold. *Rangeland Journal* 38,569–577. DOI
463 10.1071/RJ16021
- 464 Bailis R, McCarthy H. 2011. Carbon impacts of direct land use change in semiarid woodlands
465 converted to biofuel plantations in India and Brazil. *Global Change Biology Bioenergy*
466 3,449–460. DOI 10.1111/j.1757-1707.2011.01100.x
- 467 Banerjee AK, Dewanji A. 2017. Native exotic relationships in plant communities: The role of
468 exotic dominance in framing community composition. *Ecological Research* 32,653–
469 665. DOI 10.1007/s1128
- 470 Bellard C, Cassey P, Blackburn TM. 2016. Alien species as a driver of recent extinctions.
471 *Biological Letters* 12,20150623. DOI 10.1111/oik.01416
- 472 Blackburn TM, Essl F, Evans T, Hulme PE, Jeschke JM, Kühn I, Kumschick S, Marková Z,
473 Mrugała A, Nentwig W, Pergl J, Pyšek P, Rabitsch W, Ricciardi A, Richardson DM,
474 Sendek A, Vilà M, Wilson JR, Winter M, Genovesi P, Bacher S. 2014. A unified
475 classification of alien species based on the magnitude of their environmental impacts.
476 *PLoS Biology* 12,e1001850. DOI 10.1371/journal.pbio.1001850

- 477 Bragagnolo C, Gama GM, Vieira FAS, Campos-Silva JV, Bernard E, Malhado ACM, Correia
478 RA, Jepson P, de Carvalho SHC, Efe MA, Ladle RJ. 2019. Hunting in Brazil: What are
479 the options? *Perspectives in Ecology and Conservation* 17,71–79. DOI
480 10.1016/j.pecon.2019.03.001
- 481 Busso CA, Bentivegna DJ, Fernández OA. 2013. A review on invasive plants in rangelands of
482 Argentina. *Interciência* 38,95–103.
- 483 Bogino SM, Jobbágy EG. 2011. Climate and groundwater effects on the establishment, growth
484 and death of *Prosopis caldenia* trees in the Pampas (Argentina). *Forest Ecology and*
485 *Management* 262,1766–1774. DOI 10.1016/j.foreco.2011.07.032
- 486 Boudiaf I, Baudoin E, Sanguin H, Beddiar A, Thioulouse J, Galiana A, Prin Y, Le Roux C,
487 Lebrun M, Duponnois R. 2013. The exotic legume tree species, *Acacia mearnsii*, alters
488 microbial soil functionalities and the early development of a native tree species,
489 *Quercus suber*, in North Africa. *Soil Biology and Biochemistry* 65,172–179. DOI
490 10.1016/j.soilbio.2013.05.003
- 491 Campos CM, Campos VE, Mongeaud A, Borghi CE, de los Ríos C, Giannoni SM. 2011.
492 Relationships between *Prosopis flexuosa* (Fabaceae) and cattle in the Monte Desert:
493 Seeds, seedlings and saplings on cattle-use site classes. *Revista Chilena de Historia*
494 *Natural* 84:289–299.
- 495 Campos CM, Peco B, Campos VE, Malo JE, Giannoni SM, Suárez F. 2008. Endozoochory by
496 native and exotic herbivores in dry areas: Consequences for germination and survival of
497 *Prosopis* seeds. *Seed Science Research* 18,91–100. DOI 10.1017/S0960258508940344

- 498 Caswell H. 1996. Analysis of life table response experiments. 2. Alternative parameterizations
499 for size- and stage-structured models. *Ecological Modelling* 88,73–82. DOI
500 10.1016/0304-3800(95)00070-4
- 501 Catford JA, Jansson R. 2014. Drowned, buried and carried away: Effects of plant traits on the
502 distribution of native and alien species in riparian ecosystems. *New Phytologist* 205,19–
503 36. DOI 10.1111/nph.12951
- 504 Chapple DG, Simmonds SM, Wong BBM. 2012. Can behavioral and personality traits influence
505 the success of unintentional species introductions? *Trends in Ecology and Evolution*
506 27,57–64. DOI 10.1016/j.tree.2011.09.010
- 507 Cipriano-Silva R, Valladares GS, Pereira MG, Anjos LHC. 2014. Characterization of histosols in
508 floodplain environments in the northeast region of Brazil. *Revista Brasileira de Ciência*
509 *do Solo* 38,26–38. DOI 10.1590/S0100-06832014000100003
- 510 Costa CAG, Araújo JC, Lopes JWB, Pinheiro EAR. 2016. Permanence of water effectiveness in
511 the root zone of the Caatinga biome. *Revista Caatinga* 29,692–699. DOI 10.1590/1983-
512 21252016v29n320rc
- 513 Dalmazzone S, Giaccaria S. 2014. Economic drivers of biological invasions: A worldwide, bio-
514 geographic analysis. *Ecological Economics* 105,154–165. DOI
515 10.1016/j.ecolecon.2014.05.008
- 516 Dias DM, Massara RL, Bocchiglieri A. 2019. Use of habitats by donkeys and cattle within a
517 protected area of the Caatinga dry forest biome in northeastern Brazil. *Perspectives in*
518 *Ecology and Conservation* 17,64–70. DOI 10.1016/j.pecon.2019.04.005
- 519 Dube S, Mlambo D, Sebata A. 2010. Response of *Faidherbia albida* (Del.) A. Chev., *Acacia*
520 *nigrescens* Oliver. and *Acacia nilotica* (L.) Willd ex Del. seedlings to simulated

- 521 cotyledon and shoot herbivory in a semi-arid savanna in Zimbabwe. *African Journal of*
- 522 *Ecology* 48,361–367. DOI 10.1111/j.1365-2028.2009.01119.x
- 523 Dupont LM, Schlütz F, Ewah CT, Jennerjahn TC, Paul A, Behling H. 2010. Two-step vegetation
- 524 response to enhanced precipitation in Northeast Brazil during Heinrich event 1. *Global*
- 525 *Change Biology* 16,1647–1660. DOI 10.1111/j.1365-2486.2009.02023.x
- 526 El-Keblawy A, Abdelfatah MA. 2014. Impacts of native and invasive exotic *Prosopis congeners*
- 527 on soil properties and associated flora in the arid United Arab Emirates. *Journal of Arid*
- 528 *Environments* 100–101,1–8. DOI 10.1016/j.jaridenv.2013.10.001
- 529 Ferraz EMN, Rodal MJN, Sampaio EVSB. 2003. Physiognomy and structure of vegetation along
- 530 an altitudinal gradient in the semi-arid region of northeastern Brazil. *Phytocoenologia*
- 531 33,71–92. DOI 10.1127/0340-269X/2003/0033-0071
- 532 Francesca V, Osvaldo F, Stefano P, Paola RP. 2010. Soil moisture measurements: Comparison of
- 533 instrumentation performances. *Journal of Irrigation and Drainage Engineering* 136,81–
- 534 89. DOI 10.1061/(ASCE)0733-9437(2010)136:2(81)
- 535 Galindo ICL, Ribeiro MR, Santos MAV, Lima JFWF, Ferreira RFAL. 2008. Soils and
- 536 vegetation relations in areas under desertification in Jataúba county, Pernambuco State,
- 537 Brazil. *Revista Brasileira de Ciência do Solo* 32,1283–1296. DOI 10.1590/S0100-
- 538 06832008000300036
- 539 Garza JAV, Estrada AR, Cárdenas-Ávila ML, Limón SM, Álvarez MG, López VV. 2013.
- 540 Morphometric characteristics, viability and germination of mesquite and sweet acacia
- 541 seeds in northeastern Mexico. *Phyton* 82,169–174.

- Gandhi DS, Pandian SS. 2014. Inventory of trees in tropical dry deciduous forests of Tiruvannamalai district, Tamil Nadu, India. *Biodiversitas* 15,169–179. DOI 10.13057/biodiv/d150208
- Gil S, Al-Shankiti A. 2015. Priming of *Prosopis cineraria* (L.) Druce and *Acacia tortilis* (Forssk) seeds with fulvic acid extracted from compost to improve germination and seedling vigor. *Global Journal of Environmental Science and Management* 1,225–232. DOI 10.7508/gjesm.2015.03.005
- Gilioli G, Schrader G, Baker RHA, Ceglarska E, Kertész VK, Lövei G, Navajas M, Rossi V, Tramontini S, van Lenteren JC. 2014. Environmental risk assessment for plant pests: A procedure to evaluate their impacts on ecosystem services. *Science of the Total Environment* 468–469,475–486. DOI 10.1016/j.scitotenv.2013.08.068
- Goia I, Ciocanea C-M, Gavrilidis AA. 2014. Geographic origins of invasive alien species in “Iron Gates” Natural Park (Banat, Romania). *Transylvanian Review of Systematical and Ecological Research* 16,115–130. DOI 10.1515/trser-2015-0036
- Gonçalves GS, Andrade LA, Gonçalves EP, Oliveira LSB, Dias JT. 2013. Physiological quality of algaroba seeds recovered from faeces of mules. *Semina-Ciências Agrárias* 34,593–602. DOI 10.5433/1679-0359.2013v34n2p593
- Guo W-Y, van Kleunen M, Winter M, Weigelt P, Stein A, Pierce S, Pergl J, Moser D, Maurel N, Lenzner B, Kreft H, Essl F, Dawson W, Pyšek P. 2018. The role of adaptive strategies in plant naturalization. *Ecology Letters* 21,1380–1389. DOI 10.1111/ele.13104
- Hulme PE. 2015. Invasion pathways at a crossroad: Policy and research challenges for managing alien species introductions. *Journal of Applied Ecology* 52,1418–1424. DOI 10.1111/1365-2664.12470

- Hulme PE, Barrett SCH. 2013. Integrating trait- and niche-based approaches to assess contemporary evolution in alien plant species. *Journal of Ecology* 101,68–77. DOI 10.1111/1365-2745.12009
- Ilukor J, Rettberg S, Treydte A, Birner R. 2016. To eradicate or not to eradicate? Recommendations on *Prosopis juliflora* management in Afar, Ethiopia, from an interdisciplinary perspective. *Pastoralism* 6,1–8. DOI 10.1186/s13570-016-0061-1
- Jauni M, Gripenberg S, Ramula S. 2015. Non-native plant species benefit from disturbance: A meta-analysis. *Oikos* 124,122–129. DOI 10.1111/oik.01416
- Jiménez JJ, Lal R, Russo RO, Leblanc HA. 2008. The soil organic carbon in particle-size separates under different regrowth forest stands of north eastern Costa Rica. *Ecological Engineering* 34,300–310. DOI 10.1016/j.ecoleng.2008.07.001
- Kalusová V, Chytrý M, Kartesz JT, Nishino M, Pyšek P. 2013. Where do they come from and where do they go? European natural habitats as donors of invasive alien plants globally. *Diversity and Distributions* 19,199–214. DOI 10.1111/ddi.12008
- Kebede AT, Coppock DL. 2015. Livestock-mediated dispersal of *Prosopis juliflora* imperils grasslands and the endangered grevy's zebra in Northeastern Ethiopia. *Rangeland Ecology and Management* 68,402–407. DOI 10.1016/j.rama.2015.07.002
- Leal LC, Andersen AN, Leal IR. 2014. Anthropogenic disturbance reduces seed-dispersal services for myrmecochorous plants in the Brazilian Caatinga. *Oecologia* 174,173–181. DOI 10.1007/s00442-013-2740-6
- Leal LC, Meiado MV, Lopes AV, Leal IR. 2013. Germination responses of the invasive *Calotropis procera* (Ait.) R. Br. (Apocynaceae): Comparisons with seeds from two

ecosystems in northeastern Brazil. *Anais da Academia Brasileira de Ciências* 85,275–
284. DOI 10.1590/S0001-37652013000300013

Leal IR, Silva JMC, Tabarelli M, Lacher Jr TE. 2005. Changing the course of biodiversity
conservation in the Caatinga of northeastern Brazil. *Conservation Biology* 19,701–706.
DOI 10.1111/j.1523-1739.2005.00703.x

Li Y, Liu X, Li X, Petitpierre B, Guisan A. 2014. Residence time, expansion toward the equator
in the invaded range and native range size matter to climatic niche shifts in non-native
species. *Global Ecology and Biogeography* 23,1094–1104. DOI 10.1111/geb.12191

Lopes HL, Cabral JJSP, Araújo Filho JC, Montenegro SMGL. 2013. Mapping alluvial areas in
semi-arid region of Brazil through collateral data and satellite images. *Revista
Brasileira de Engenharia Agrícola e Ambiental* 17,763–769. DOI 10.1590/S1415-
43662013000700011

Lopes I, Montenegro AAA, Lima JLMP. 2019. Performance of conservation techniques for
semiarid environments: Field observations with Caatinga, mulch, and cactus forage
palma. *Water* 11,792. DOI 10.3390/w11040792

Lv X, Zhou G. 2018. Climatic suitability of the geographic distribution of *Stipa breviflora* in
Chinese temperate grassland under climate change. *Sustainability* 10,3767. DOI
10.3390/su10103767

Maestre FT, Reynolds JF. 2006. Small-scale spatial heterogeneity in the vertical distribution of
soil nutrients has limited effects on the growth and development of *Prosopis glandulosa*
seedlings. *Plant Ecology* 183,65–75. DOI 10.1007/s11258-005-9007-1

- 608 Majd R, Aghaie P, Monfared EK, Alebrahim MT. 2013. Evaluating of some treatments on
609 breaking seed dormancy in mesquite. *International Journal of Agronomy and Plant*
610 *Production* 4,1433–1439.
- 611 Malavasi M, Santoro R, Cutini M, Acosta ATR, Carranza ML. 2016. The impact of human
612 pressure on landscape patterns and plant species richness in Mediterranean coastal
613 dunes. *Plant Biosystems - An International Journal Dealing with all Aspects of Plant*
614 *Biology* 150,73–82. DOI 10.1080/11263504.2014.913730
- 615 Malíková L, Mudrák O, Klimešová J. 2012. Adventitious sprouting enables the invasive annual
616 herb *Euphorbia geniculata* to regenerate after severe injury. *Ecological Research*
617 27,841–847. DOI 10.1007/s11284-012-0960-6
- 618 Mamede MA, Araújo FS. 2008. Effects of slash and burn practices on a soil seed bank of
619 caatinga vegetation in Northeastern Brazil. *Journal of Arid Environments* 72,458–470.
620 DOI 10.1016/j.jaridenv.2007.07.014
- 621 Marques FA, Ribeiro MR, Bittar SMB, Tavares Filho AN, Lima JFWF. 2007. Characterization
622 and classification of neosols on the Fernando de Noronha Island, Pernambuco, Brazil.
623 *Revista Brasileira de Ciência do Solo* 31,1553–1562. DOI 10.1590/S0100-
624 06832007000600032
- 625 Martínez AJ, López-Portillo J. 2003. Allometry of *Prosopis glandulosa* var. *torreyana* along a
626 topographic gradient in the Chihuahuan desert. *Journal of Vegetation Science* 4,111–
627 120.
- 628 McGeoch MA, Genovesi P, Bellingham PJ, Costello MJ, McGrannachan C, Sheppard A. 2016.
629 Prioritizing species, pathways, and sites to achieve conservation targets for biological
630 invasion. *Biological Invasions* 18,299–314. DOI 10.1007/s10530-015-1013-1

- 631 Meloni DA. 2012. Physiological responses to calcium supplementation of NaCl-stressed vinal
632 (*Prosopis ruscifolia* G.) seedlings. *Revista de la Facultad de Ciencias Agrarias*
633 *UNCuyo* 44,79–88.
- 634 Miccolis A, Peneireiro FM, Vieira DLM, Marques HR, Hoffmann MRM. 2017. Restoration
635 through agroforestry: Options for reconciling livelihoods with conservation in the
636 Cerrado and Caatinga biomes in Brazil. *Experimental Agriculture* 1,1–18. DOI
637 10.1017/S0014479717000138
- 638 Miranda RQ, Correia RM, Almeida-Cortez JS, Pompelli MF. 2014. Germination of *Prosopis*
639 *juliflora* (Sw.) D.C. seeds at different osmotic potentials and temperatures. *Plant*
640 *Species Biology* 29,E9–E20. DOI 10.1111/1442-1984.12025
- 641 Miranda RQ, Oliveira MTP, Correia RM, Almeida-Cortez JS, Pompelli MF. 2011. Germination
642 of *Prosopis juliflora* (Sw) DC seeds after scarification treatments. *Plant Species Biology*
643 26,186–192. DOI 10.1111/j.1442-1984.2011.00324.x
- 644 Moran EV, Reid A, Levine JM. 2017. Population genetics and adaptation to climate along
645 elevation gradients in invasive *Solidago canadensis*. *PLoS ONE* 12,e0185539. DOI
646 10.1371/journal.pone.0185539
- 647 Morandini MN, Giamminola EM, Viana ML. 2013. Desiccation tolerance in seeds of *Prosopis*
648 *ferox* and *Pterogyne nitens* (Fabaceae). *Revista de Biología Tropical* 61,335–342. DOI
649 10.15517/RBT.V61I1.11132
- 650 Moravcová L, Pyšek P, Jarošík V, Pergl J. 2015. Getting the right traits: Reproductive and
651 dispersal characteristics predict the invasiveness of herbaceous plant species. *PLoS*
652 *ONE* 10,e0123634. DOI 10.1371/journal.pone.0123634

- 653 Moro MF, Lughadha EN, Araújo FS, Martins FR. 2016. A phytogeographical metaanalysis of
654 the Semiarid Caatinga Domain in Brazil. *Botanical Review* 82,91–148. DOI
655 10.1007/s12229-016-9164-z
- 656 Mukherjee AM, Velankar AD, Kumara HN. 2017. Invasive *Prosopis juliflora* replacing the
657 native floral community over three decades: A case study of a world heritage site,
658 Keoladeo National Park, India. *Biodiversity and Conservation* 26,2839–2856. DOI
659 10.1007/s10531-017-1392-y
- 660 Mullah CJA, Klanderud K, Totland Ø, Odee D. 2014. Community invasibility and invasion by
661 non-native *Fraxinus pennsylvanica* trees in a degraded tropical forest. *Biological*
662 *Invasions* 16,2747–2755. DOI 10.1007/s10530-014-0701-6
- 663 Nascimento CED, Tabarelli M, Silva CA, Leal IR, Souza Tavares W, Serrão JE, Zanuncio JC.
664 2014. The introduced tree *Prosopis juliflora* is a serious threat to native species of the
665 Brazilian Caatinga vegetation. *Science of the Total Environment* 481,108–113. DOI
666 10.1016/j.scitotenv.2014.02.019
- 667 Nasr SMH, Parsakhoo A, Naghavi H, Koohi SKS. 2012. Effect of salt stress on germination and
668 seedling growth of *Prosopis juliflora* (Sw.). *New Forests* 43,45–55. DOI
669 10.1007/s11056-011-9265-9
- 670 Naudiyal N, Joachim S, Stefanie G. 2017. What influences the plant community composition on
671 Delhi ridge? The role played by *Prosopis juliflora* and anthropogenic disturbances.
672 *Tropical Ecology* 58,33–43.
- 673 Novoa A, Le Roux JJ, Robertson MP, Wilson JR, Richardson DM. 2015. Introduced and
674 invasive cactus species: A global review. *AoB Plants* 7,plu078. DOI
675 10.1093/aobpla/plu078

- 676 Oliveira JV, Lopes SF, Barboza RRD, Trovão DMMB, Ramos MB, Alves RRN. 2019. Wild
677 vertebrates and their representation by urban/rural students in a region of northeast
678 Brazil. *Journal of Ethnobiology and Ethnomedicine* 15,1–23. DOI 10.1186/s13002-018-
679 0283-y
- 680 Oliveira LSB, Andrade LA, Fabricante JR, Gonçalves GS. 2012. Structure of a *Prosopis juliflora*
681 (Sw.) DC. population established in a temporary riverbed in the Microregion of Cariri in
682 the State of Paraíba. *Semina: Ciências Agrárias* 33,1769–1778. DOI 10.5433/1679-
683 0359.2012v33n5p1769
- 684 Oliveira MT, Matzek V, Medeiros CD, Rivas R, Falcão HM, Santos MG. 2014. Stress tolerance
685 and ecophysiological ability of an invader and a native species in a seasonally dry
686 tropical forest. *PLoS ONE* 9,e105514. DOI 10.1371/journal.pone.0105514
- 687 Oliveira MT, Souza GM, Pereira S, Oliveira DAS, Figueiredo-Lima KV, Arruda E, Santos MG.
688 2017. Seasonal variability in physiological and anatomical traits contributes to invasion
689 success of *Prosopis juliflora* in tropical dry forest. *Tree Physiology* 37,326–337. DOI
690 10.1093/treephys/tpw123
- 691 Oliveira OF, Santos MVF, Cunha MV, Dubeux Júnior JCB, Muir JP, Mello ACL, Lira MA,
692 Barros GFNP. 2016. Botanical composition of Caatinga rangeland and diets selected by
693 grazing sheep. *Tropical Grasslands* 4,71–81. DOI 10.17138/TGFT(4)71-81
- 694 Pasha SV, Satish KC, Reddy CS, Jha CS. 2015. Massive invasion of mesquite (*Prosopis*
695 *juliflora*) in Wild Ass Wildlife Sanctuary, India. *National Academy Science Letters*
696 38,271–273. DOI 10.1007/s40009-014-0321-9

- Patil AV, Karadge BA. 2012. Effect of NaCl salinity on the growth and mineral nutrition of one month old *Prosopis juliflora* (Sw.) DC seedlings. *Pharmacognosy Journal* 4,63–66. DOI 10.5530/pj.2012.31.12
- Pegado CMA, Andrade LA, Félix LP, Pereira IM. 2006. Effects of the biological invasion of algaroba - *Prosopis juliflora* (Sw.) DC. on composition and structure of the shrubtree stratum of the Caatinga in Monteiro Municipality, Paraíba State, Brazil. *Acta Botanica Brasilica* 20,887–898. DOI 10.1590/S0102-33062006000400013
- Pérez EA, Téllez TR, Maqueda SR, Linares PJC, Pardo FMV, Medina PLR, Moreno JL, Gallego FL, Cortés JG, Guzmán JMS. 2015. Seed germination and risks of using the invasive plant *Eichhornia crassipes* (Mart.) Solms-Laub. (water hyacinth) for composting, ovine feeding and biogas production. *Acta Botanica Gallica* 162,203–214. DOI 10.1080/12538078.2015.1056227
- Pérez-Marin AM, Rogé P, Altieri MA, Forer LFU, Silveira L, Oliveira VM, Domingues-Leiva BE. 2017. Agroecological and social transformations for coexistence with semi-aridity in Brazil. *Sustainability* 9,990. DOI 10.3390/su9060990
- Perkins LB, Nowak RS. 2013. Invasion syndromes: Hypotheses on relationships among invasive species attributes and characteristics of invaded sites. *Journal of Arid Land* 5,275–283. DOI 10.1007/s40333-013-0161-3
- Rathore VS, Singh JP, Bhardwaj S, Nathawat NS, Kumar M, Roy MM. 2015. Potential of native shrubs *Haloxylon salicornicum* and *Calligonum Polygonoides* for restoration of degraded lands in Arid Western Rajasthan, India. *Environmental Management* 55,205–216. DOI 10.1007/s00267-014-0372-1

- 719 Rai PK. 2015. What makes the plant invasion possible? Paradigm of invasion mechanisms,
720 theories and attributes. *Environmental Skeptics and Critics* 4,36–66.
- 721 Ramos CMC, Silva AF, Sartori AAC, Zimback CRL, Bassoi LH. 2011. Modeling the hourly
722 variation of air temperature at Petrolina and Botucatu, Brazil. *Revista Brasileira de*
723 *Engenharia Agrícola e Ambiental* 15,959–965. DOI 10.1590/S1415-
724 43662011000900012
- 725 Razanamandranto S, Tigabu M, Neya S, Odén PC. 2004. Effects of gut treatment on recovery
726 and germinability of bovine and ovine ingested seeds of four woody species from the
727 Sudanian savanna in West Africa. *Flora - Morphology, Distribution, Functional*
728 *Ecology of Plants* 199,389–397. DOI 10.1078/0367-2530-00167
- 729 Ribeiro EMS, Arroyo-Rodríguez V, Santos BA, Tabarelli M, Leal IR. 2015. Chronic
730 anthropogenic disturbance drives the biological impoverishment of the Brazilian
731 Caatinga vegetation. *Journal of Applied Ecology* 52,611–620. DOI 10.1111/1365-
732 2664.12420
- 733 Risio L, Calama R, Bogino SM, Bravo F. 2016. Inter-annual variability in *Prosopis caldenia* pod
734 production in the Argentinean semiarid pampas: A modelling approach. *Journal of Arid*
735 *Environments* 131,59–66. DOI 10.1016/j.jaridenv.2016.03.007
- 736 Rito KF, Arroyo-Rodríguez V, Queiroz RT, Leal IR, Tabarelli M. 2016. Precipitation mediates
737 the effect of human disturbance on the Brazilian Caatinga vegetation. *Journal of*
738 *Ecology* 105,828–838. DOI 10.1111/1365-2745.12712
- 739 Rivaes R, Rodríguez-González PM, Albuquerque A, Pinheiro AN, Egger G, Ferreira MT. 2013.
740 Riparian vegetation responses to altered flow regimes driven by climate change in
741 Mediterranean rivers. *Ecohydrology* 6,413–424. DOI 10.1002/eco.1287

- Rivaes RP, Rodríguez-González PM, Ferreira MT, Pinheiro AN, Politti E, Egger G, García-Arias A, Francés F. 2014. Modeling the evolution of riparian woodlands facing climate change in three European rivers with contrasting flow regimes. *PLoS ONE* 9,e110200. DOI 10.1371/journal.pone.0110200
- Roy HE, Peyton J, Aldridge DC, Bantock T, Blackburn TM, Britton R, Clark P, Cook E, Dehnen-Schmutz K, Dines T, Dobson M, Edwards F, Harrower C, Harvey MC, Minchin D, Noble DG, Parrott D, Pocock MJO, Preston CD, Roy S, Salisbury A, Schönrogge K, Sewell J, Shaw RH, Stebbing P, Stewart AJA, Walker KJ. 2014. Horizon scanning for invasive alien species with the potential to threaten biodiversity in Great Britain. *Global Change Biology* 20,3859–3871. DOI 10.1111/gcb.12603
- Sampaio EVSB, Costa TL. 2011. Stocks and fluxes of carbon in semiarid northeast Brazil: Preliminary estimates. *Revista Brasileira de Geografia Física* 6,1275–1291.
- Santos DM, Santos JMFF, Silva KA, Araújo VKR, Araújo EL. 2016. Composition, species richness, and density of the germinable seed bank over 4 years in young and mature forests in Brazilian semiarid regions. *Journal of Arid Environments* 129,93–101. DOI 10.1016/j.jaridenv.2016.02.012
- Santos JPO, Silva Júnior JM, Silva Filho JA, Silva JLC, Ribeiro JKN, Oliveira FIF. 2017. Environmental degradation and risk of desertification in Alto Sertão Region, Alagoas State, Brazil. *Revista Geama* 3,191–195.
- Santos JPS, Diodato MA. 2015. Carob tree (*Prosopis juliflora*) in north Rio Grande caatinga: Analysis of the municipality of Fernando Pedroza. *Nature and Conservation* 8,13–21. DOI 10.6008/SPC2318-2881.2015.001.0002

- 764 Schulz C, Koch R, Cierjacks A, Kleinschmit B. 2017. Land change and loss of landscape
765 diversity at the Caatinga phytogeographical domain – Analysis of pattern-process
766 relationships with MODIS land cover products (2001–2012). *Journal of Arid*
767 *Environments* 136,54–74. DOI 10.1016/j.jaridenv.2016.10.004
- 768 Sfair JC, Bello F, França TQ, Baldauf C, Tabarelli M. 2018. Chronic human disturbance affects
769 plant trait distribution in a seasonally dry tropical forest. *Environmental Research*
770 *Letters* 13,025005. DOI 10.1088/1748-9326/aa9f5e
- 771 Shackleton RT, Le Maitre DC, Richardson DM. 2015. *Prosopis* invasions in South Africa:
772 Population structures and impacts on native tree population stability. *Journal of Arid*
773 *Environments* 114,70–78. DOI 10.1016/j.jaridenv.2014.11.006
- 774 Shiferaw H, Teketay D, Nemomissa S, Assefa F. 2004. Some biological characteristics that
775 foster the invasion of *Prosopis juliflora* (Sw.) DC. at Middle Awash Rift Valley Area,
776 north-eastern Ethiopia. *Journal of Arid Environments* 58,135–154. DOI
777 10.1016/j.jaridenv.2003.08.011
- 778 Silva AC, Souza AF. 2018. Aridity drives plant biogeographical sub regions in the Caatinga, the
779 largest tropical dry forest and woodland block in South America. *PLoS ONE*
780 13,e0196130. DOI 10.1371/journal.pone.0196130
- 781 Silva JMC, Leal IR, Tabarelli M. 2017. Caatinga: The largest dry forest region in South
782 America. Cahm: Springer International Publishing. 482 p. DOI 10.1007/978-3-319-
783 68339-3
- 784 Silva LCR, Sternberg L, Haridasan M, Hoffmann WA, Miralles-Wilhelm F, Franco AC. 2008.
785 Expansion of gallery forests into central Brazilian savannas. *Global Change Biology*
786 14,2108–2118. DOI 10.1111/j.1365-2486.2008.01637.x

- 787 Silva FKG, Lopes SF, Lopez LCS, Melo JIM, Trovão DMBM. 2014. Patterns of species richness
788 and conservation in the Caatinga along elevational gradients in a semiarid ecosystem.
789 *Journal of Arid Environments* 110,47–52. DOI 10.1016/j.jaridenv.2014.05.011
- 790 Simberloff D, Martin JL, Genovesi P, Maris V, Wardle DA, Aronson J, Courchamp F, Galil B,
791 García-Berthou E, Pascal M, Pyšek P, Sousa R, Tabacchi E, Vilà M. 2013. Impacts of
792 biological invasions: What’s what and the way forward. *Trends in Ecology and*
793 *Evolution* 28,58–66. DOI 10.1016/j.tree.2012.07.013
- 794 Song G, Li X, Hui R. 2017. Effect of biological soil crusts on seed germination and growth of an
795 exotic and two native plant species in an arid ecosystem. *PLoS ONE* 12,e0185839. DOI
796 10.1371/journal.pone.0185839
- 797 Svenning J-C, Gravel D, Holt RD, Schurr FM, Thuiller W, Münkemüller T, Schiffrers KH,
798 Dullinger S, Edwards Jr TC, Hickler T, Higgins SI, Nabel JEMS, Pagel J, Normand S.
799 2014. The influence of interspecific interactions on species range expansion rates.
800 *Ecography* 37,1198–1209. DOI 10.1111/j.1600-0587.2013.00574.x
- 801 Terra MCN, Santos RM, Prado Júnior JA, Mello JM, Scolforo JRS, Fontes MAL, Schiavini I,
802 Reis AA, Bueno IT, Magnago LFS, ter Steege H. 2018. Water availability drives
803 gradients of tree diversity, structure and functional traits in the Atlantic–Cerrado–
804 Caatinga transition, Brazil. *Plant Ecology* 11,803–814. DOI 10.1093/jpe/rty017
- 805 Tukey JW. 1949. Comparing individual means in the Analysis of Variance. *Biometrics* 5,99–
806 114.
- 807 Vavra M, Parks CG, Wisdom MJ. 2007. Biodiversity, exotic plant species, and herbivory: The
808 good, the bad, and the ungulate. *Forest Ecology and Management* 246,66–72. DOI
809 10.1016/j.foreco.2007.03.051

- van Kleunen M, Dawson W, Maurel WD. 2015. Characteristics of successful alien plants.
Molecular Ecology 24,1954–1968. DOI 10.1111/mec.13013
- van Wilgen BW, Richardson DM. 2014. Challenges and trade-offs in the management of
invasive alien trees. *Biological Invasions* 16,721–734. DOI 10.1007/s10530-013-0615-8
- Velarde M, Felker P, Gardiner D. 2005. Influence of elemental sulfur, micronutrients,
phosphorus, calcium, magnesium and potassium on growth of *Prosopis alba* on high pH
soils in Argentina. *Journal of Arid Environments* 62,525–539. DOI
10.1016/j.jaridenv.2005.01.022
- Villalobos AE, Peláez DV, Elia OR. 2005. Factors related to establishment of *Prosopis caldenia*
Burk. seedlings in central rangelands of Argentina. *Acta Oecologica* 27,99–106. DOI
10.1016/j.actao.2004.10.001
- Villalobos AE, Peláez DV. 2015. Functional responses of woody *Prosopis caldenia* seedlings to
drought and livestock grazing in semiarid rangelands of Argentina. *Arid Land Research
and Management* 29,487–502. DOI 10.1080/15324982.2015.1030798
- Wallis K. 1952. Use of ranks in one-criterion variance analysis. *Journal of the American
Statistical Association* 47,583–621. DOI 10.1080/01621459.1952.10483441
- Walkie TT, Hoag D, Evangelista PH, Luizza M, Laituri M. 2016. Is control through utilization a
cost effective *Prosopis juliflora* management strategy? *Journal of Environmental
Management* 168,74–86. DOI 10.1016/j.jenvman.2015.11.054
- Westphal C, Gachón P, Bravo J, Navarrete C, Salas C, Ibáñez C. 2015. The potential of
algarrobo (*Prosopis chilensis* (Mol.) Stuntz) for regeneration of desertified soils:
Assessing seed germination under saline conditions. *Environmental Management*
56,209–220. DOI 10.1007/s00267-015-0490-4

- 833 Yoda K, Tsuji W, Inoue T, Saito T, Elbasit MAMA, Eldoma AM. 2015. Evaluation of the effect
834 of a rain pulse on the initial growth of *Prosopis* seedlings. *Arid Land Research and*
835 *Management* 29,210–221. DOI 10.1080/15324982.2014.943376
- 836 Zapala MN, Ellzey JT, Bader J, Peralta-Videa JR, Gardea-Torresdey J. 2014. Effects of copper
837 sulfate on seedlings of *Prosopis pubescens* (screwbean mesquite). *International Journal*
838 *of Phytoremediation* 16,1031–1041. DOI 10.1080/15226514.2013.810582

Figure 1

Field capacity (A) and soil moisture (B) in floodplain, alluvial terrace and plateau habitats, at 0-20 cm depth. Petrolina, Pernambuco State, Brazil.

Tukey's Honest Significant (field capacity) and Kruskal-Wallis (soil moisture) tests. Columns followed by the same letter are similar ($p < 0.05$). Error bars indicate standard deviation.

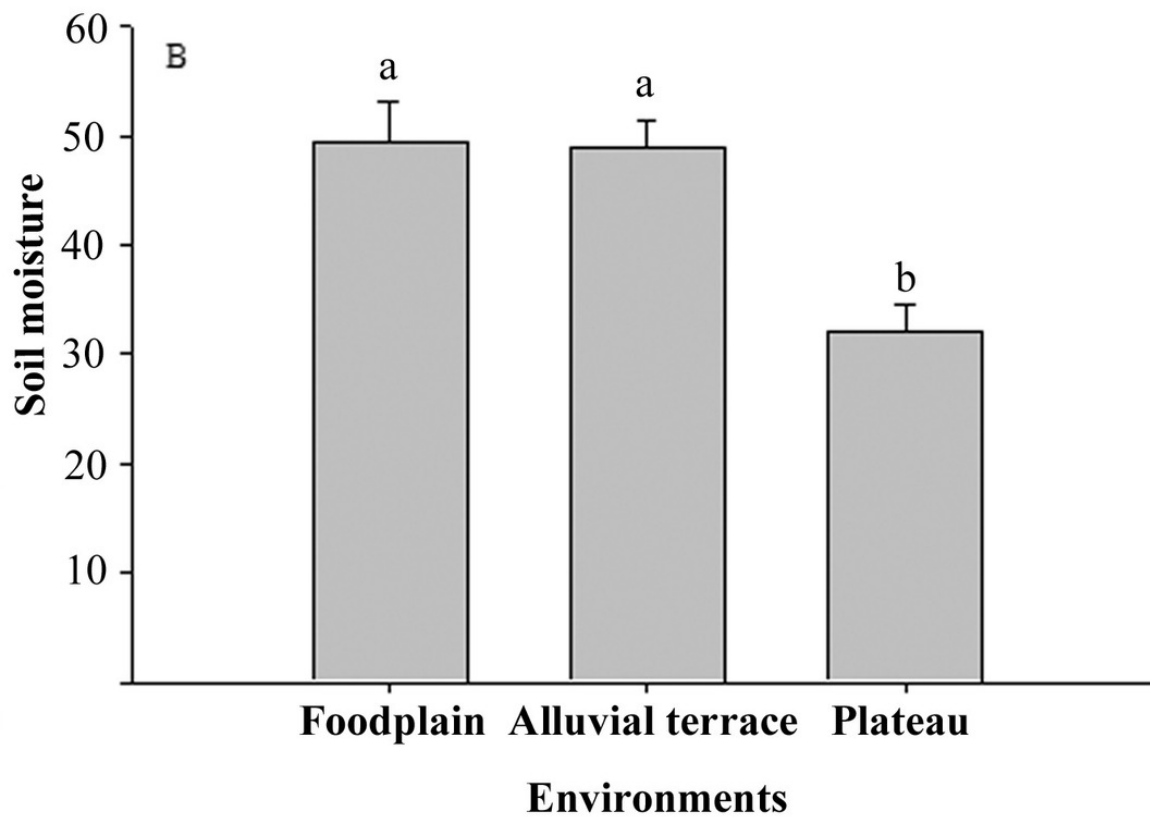
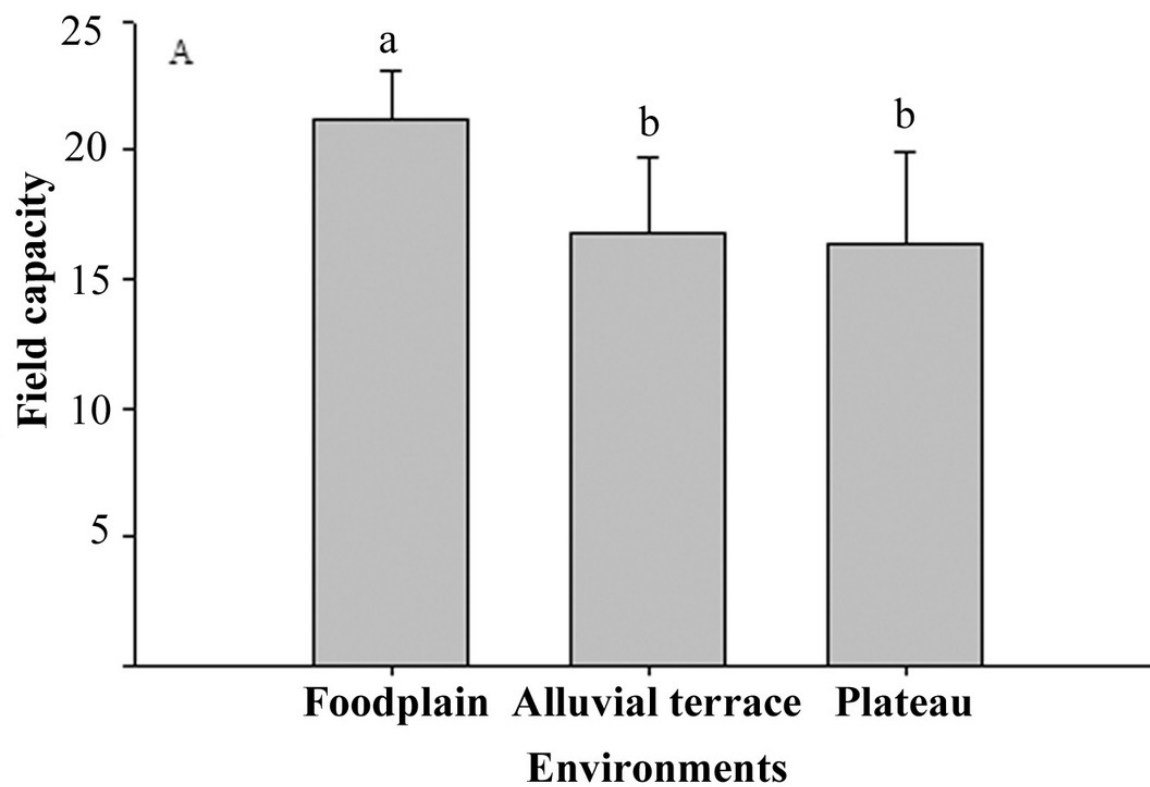


Figure 2

Particle size (A) and organic matter (OM) (g.Kg^{-1}), phosphorus (P) (mg.dm^{-3}), potassium (K) (cmolc.dm^{-3}), calcium (Ca) (cmolc.dm^{-3}), magnesium (Mg) (cmolc.dm^{-3}), aluminum (Al) (cmolc.dm^{-3}) and ca

Tukey's Honest Significant (sand and silt) and Kruskal-Wallis (clay, OM, P, K, Ca, Mg, Al and CEC) tests. Columns followed by same letter are similar ($p= 0.05$). Error bars indicate standard deviation.

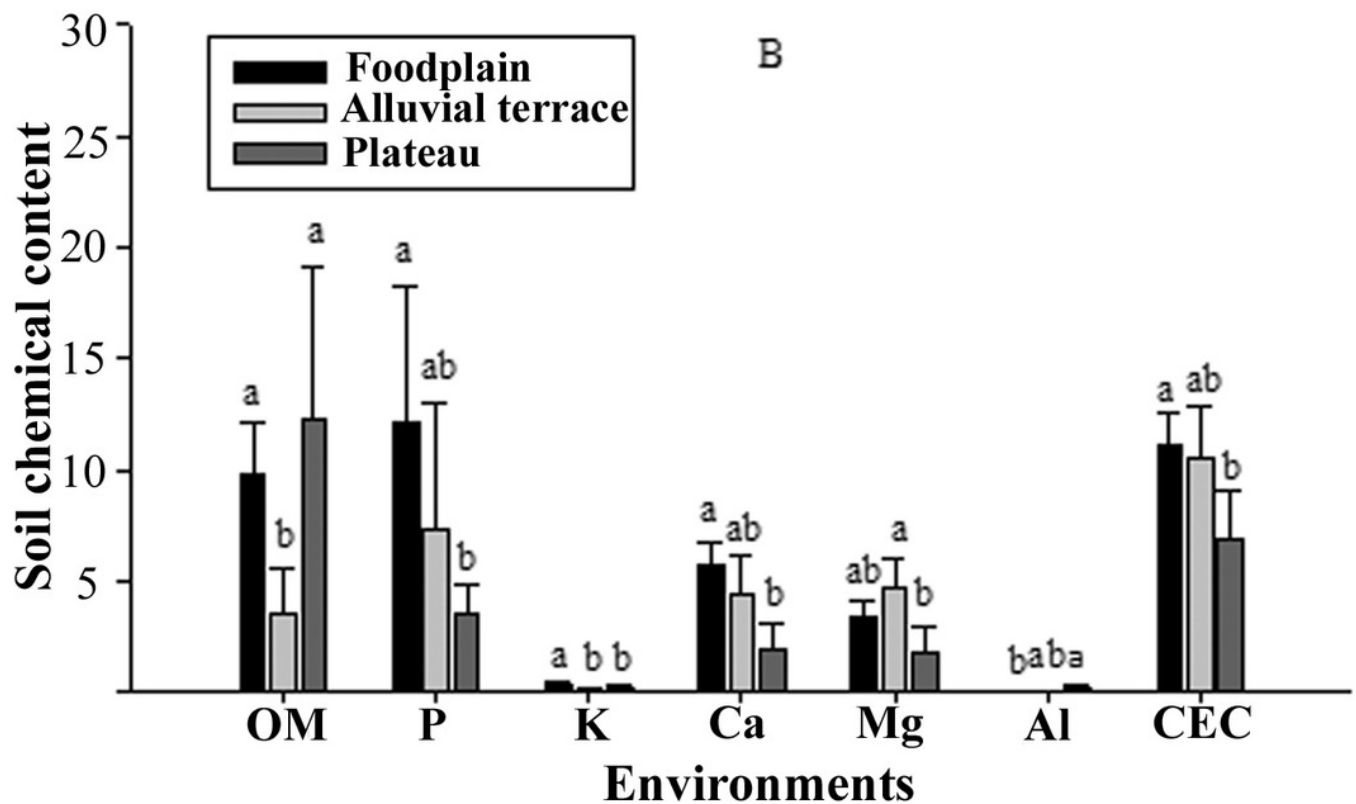
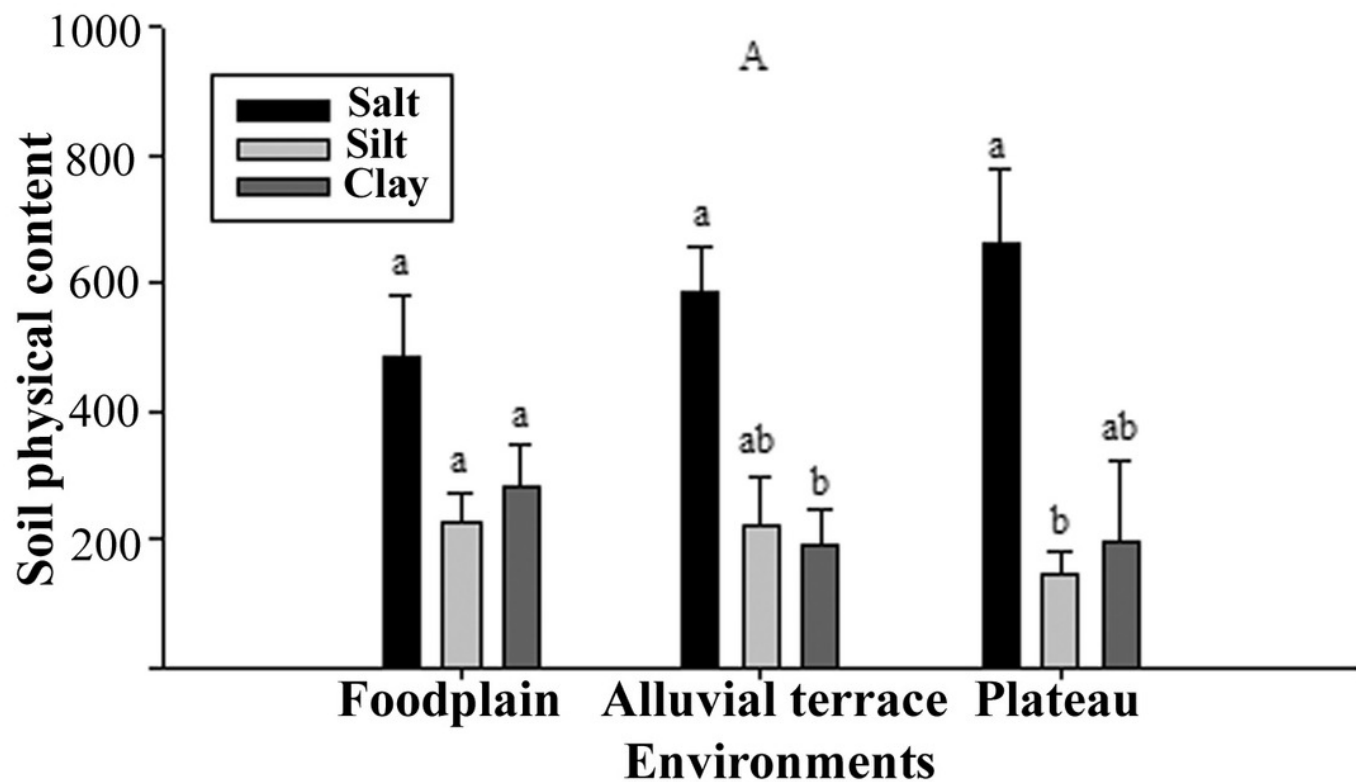


Figure 3

Survival (l_x) and life expectancy (e_x) of *Prosopis juliflora* sown on the soil surface (A), buried (B) or mixed with cattle (C) or mule (D) manure in the floodplain habitat up to 345 days. Petrolina, Pernambuco State,

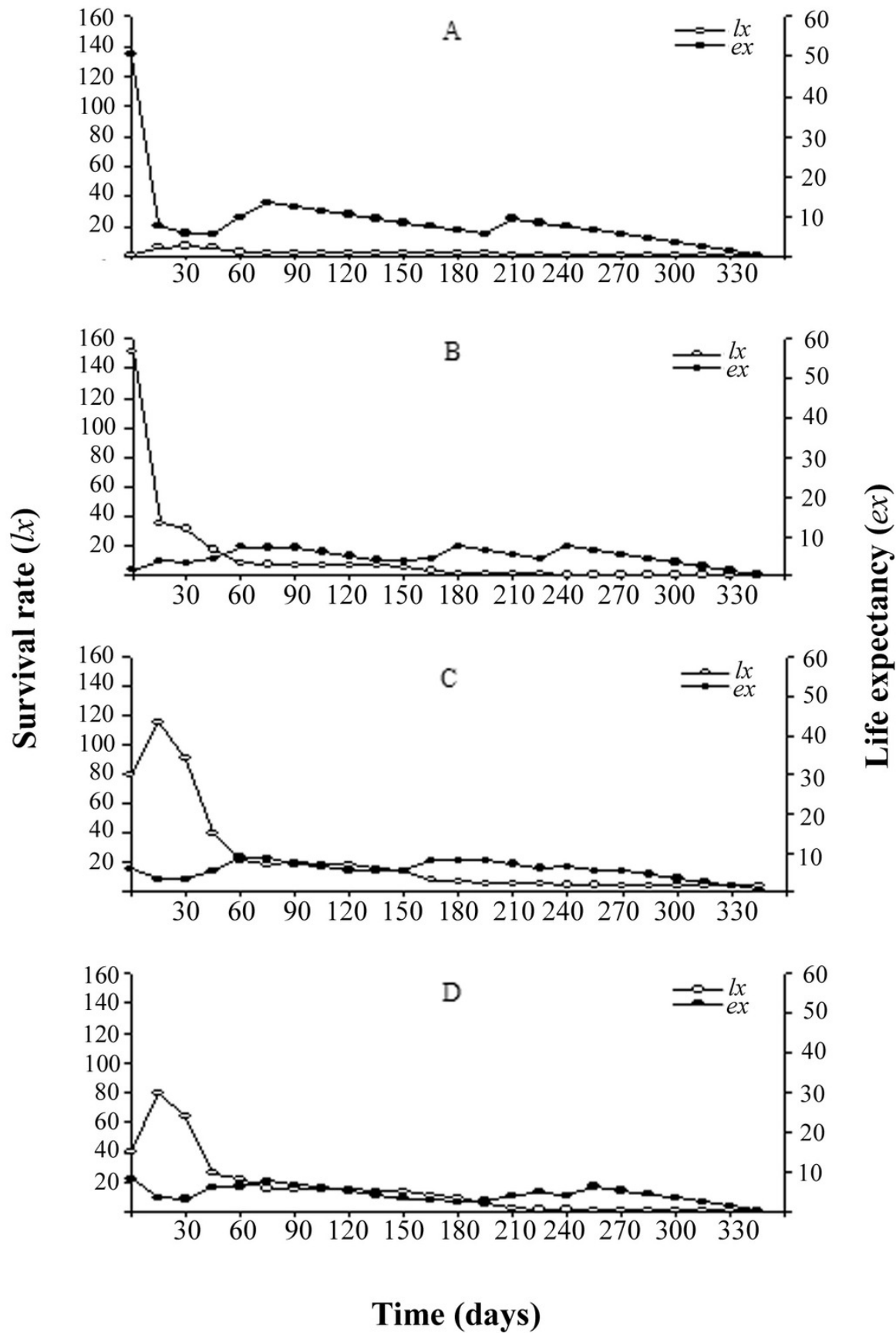


Figure 4

Survival (l_x) and life expectancy (e_x) of *Prosopis juliflora* sown on the soil surface (A), buried (B) or mixed with cattle (C) or mule (D) manure in the alluvial terrace habitat up to 165 days. Petrolina, Pernambuco S

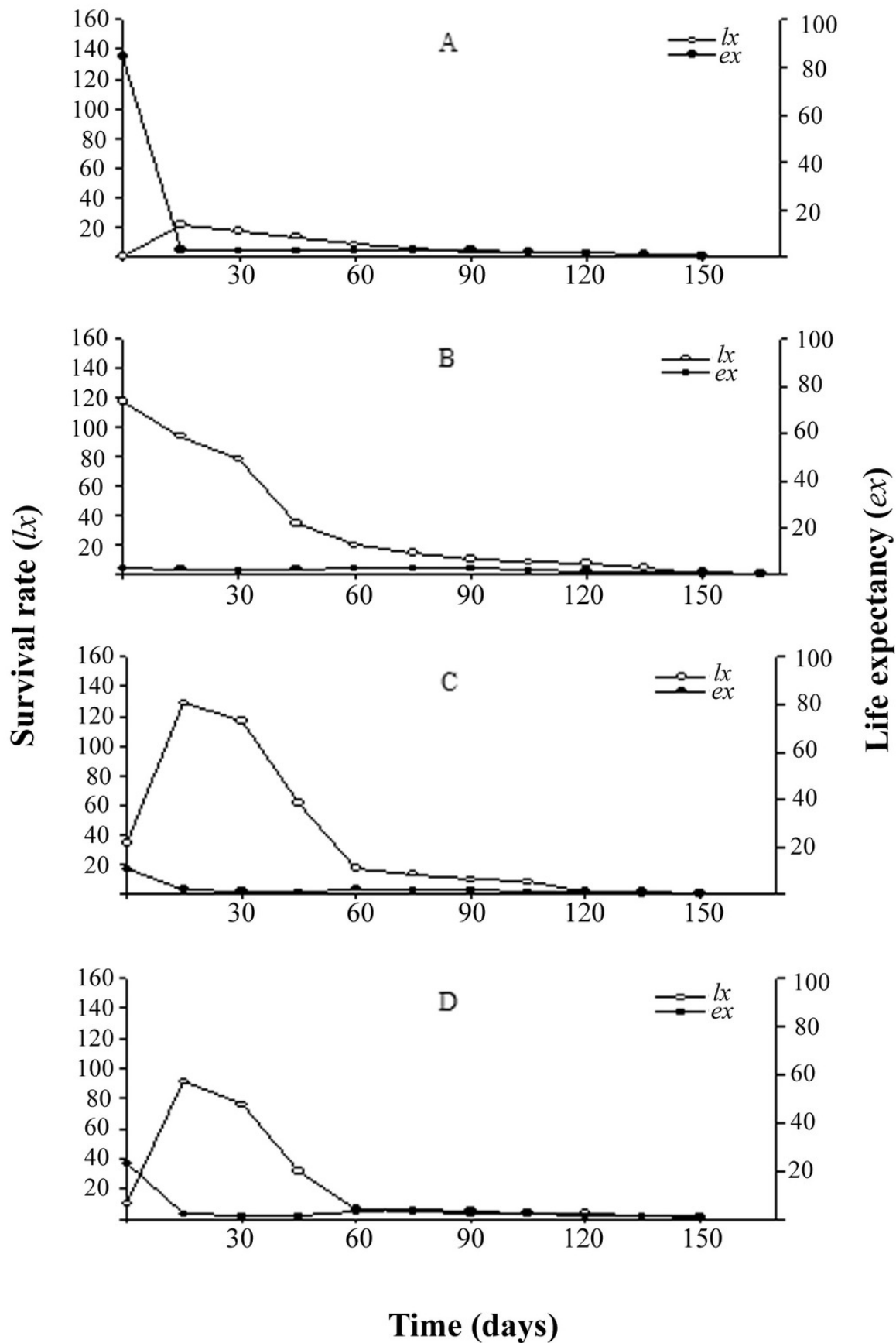


Figure 5

Survival (l_x) and life expectancy (e_x) of *Prosopis juliflora* sown on the soil surface (A), buried (B) or mixed with cattle (C) or mule (D) manure in the plateau habitat up to 165 days. Petrolina, Pernambuco State, Bra

