

The impact of short-term exposure to near shore stressors on the early life stages of the reef building coral *Montipora capitata*

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Successful reproduction and survival of coral is crucial to the continuation and resilience of coral globally. As reef waters warm due to climate change, episodic largescale tropical storms are becoming more frequent, drastically altering the near shore water quality for short periods of time. Therefore, it is critical that we understand the effects warming waters, fresh water input, and run-off have on sexual reproduction of coral. To better understand the effects of these near shore stressors on Hawaiian coral, laboratory experiments were conducted at the Hawai'i Institute of Marine Biology to determine the independent effects of suspended sediment concentrations (100 mg l⁻¹ and 200 mg l⁻¹), lowered salinity (28 ‰), and elevated temperature (31 °C) on the successful fertilization, larval survival, and settlement of the scleractinian coral *Montipora capitata*. In the present study, early developmental stages of coral were exposed to one of three near shore stressors for a period of 24 h and the immediate (fertilization) and latent effects (larval survival and settlement) were observed and measured. Fertilization success and settlement were not affected by any of the treatments; however, larval survival was negatively affected by all of the treatments by 50% or greater (p>0.05). These data suggest that if coral are exposed to any of these stressors during spawning, the subsequent developing larval stages will be severely impacted.

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2 building coral *Montipora capitata*

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25 **Abstract**

26 Successful reproduction and survival of coral is crucial to the continuation and resilience
27 of coral globally. As reef waters warm due to climate change, episodic largescale tropical storms
28 are becoming more frequent, drastically altering the near shore water quality for short periods of
29 time. Therefore, it is critical that we understand the effects warming waters, fresh water input,
30 and run-off have on sexual reproduction of coral. To better understand the effects of these near
31 shore stressors on Hawaiian coral, laboratory experiments were conducted at the Hawai'i
32 Institute of Marine Biology to determine the independent effects of suspended sediment
33 concentrations (100 mg l⁻¹ and 200 mg l⁻¹), lowered salinity (28 ‰), and elevated temperature
34 (31° C) on the successful fertilization, larval survival, and settlement of the scleractinian coral
35 *Montipora capitata*. In the present study, early developmental stages of coral were exposed to
36 one of three near shore stressors for a period of 24 h and the immediate (fertilization) and latent
37 effects (larval survival and settlement) were observed and measured. Fertilization success and
38 settlement were not affected by any of the treatments; however, larval survival was negatively
39 affected by all of the treatments by 50% or greater (p>0.05). These data suggest that if coral
40 are exposed to any of these stressors during spawning, the subsequent developing larval
41 stages will be severely impacted.

42

43 **Introduction**

44 Coral reefs are among the most productive and diverse ecosystems in the world and these
45 vulnerable ecosystems are rapidly experiencing global decline (Bellwood 2004; Wilkinson
46 2000). They provide indispensable ecological services such as shoreline protection, food

47 production, and are highly attractive to tourism (Oliver 2011). There are many stressors
48 impacting the condition of coral reefs, both globally and locally. Global impacts include issues
49 related to climate change such as sea surface temperature rise and ocean acidification
50 (McClanahan et al. 2007; Spalding & Brown 2015). Local impacts include fishing pressure,
51 eutrophication, coastal construction, dredging, increased sedimentation, invasive species, and
52 freshwater runoff. All of the aforementioned stressors have been shown to negatively affect the
53 condition of coral reefs (Banner 1968; Fabricius 2005; Hughes et al. 2007; Ogden & Lobel 1978;
54 Richmond 1993; Rogers 1990).

55 Kāneʻohe Bay, (21°28'N; 157°48'W) – the largest embayment in the Hawaiian islands
56 located on the windward side of Oʻahu – has been experiencing sedimentation and freshwater
57 runoff for decades (Bahr et al. 2015a). Freshwater runoff occurs during storm flooding and is a
58 common event on tropical islands that can temporarily decrease salinity (Banner 1968; Jokiel et
59 al. 1993). Such events have been shown to cause mass mortality of adult coral (Bahr et al. 2015b;
60 Banner 1968; Jokiel et al. 1993) and could pose a threat to the early life history of coral
61 (Babcock et al. 1986; Kolinski & Cox 2003). In the past 100 years, the coral of Kāneʻohe Bay
62 have been chronically impacted by sediment, mainly through dredging and watershed runoff
63 (Bahr et al. 2015a).

64 More recently, Kāneʻohe Bay has been experiencing warmer summer temperatures (2-4°
65 C above the summer average of 27° C) that have resulted in coral bleaching. In 2014 and 2015,
66 Kāneʻohe Bay experienced consecutive warming events that resulted in widespread bleaching of
67 coral. With increased intensity and frequency of bleaching events coral are less likely to recover.
68 Bahr et al. (2017) surveyed coral before and after both the 2014 and 2015 bleaching events. They

69 found that overall coral mortality was higher after the 2015 bleaching event (5.5% in 2014 and
70 16.0% in 2015) which suggests that consecutive bleaching events affect coral resilience.

71 The effects of local near shore stressors such as sedimentation, freshwater runoff, and
72 elevated temperatures on adult coral have been thoroughly studied (Erftemeijer et al. 2012;
73 Fabricius 2005; Humphrey et al. 2008; Jokiel et al. 2014; Rogers 1990; Te 2001). More recently,
74 studies have looked at the effects of near shore stressors on the early life stages of coral as well
75 (Edmunds et al. 2001; Hedouin et al. 2015; Humanes et al. 2017; Jones et al. 2015; Ricardo et al.
76 2015; Ricardo et al. 2018; Ricardo et al. 2017). However, very little is known about how these
77 near shore stressors affect the early life stages in Hawaiian coral. These near shore stressors were
78 selected for this study because all three have been shown to negatively affect adult coral
79 worldwide (Douglas 2003; Fabricius 2005; Rogers 1990) and in Hawai'i (Bahr et al. 2015b;
80 Jokiel & Brown 2004; Jokiel et al. 1993; Jokiel et al. 2014).

81 In this study we exposed the early life stages (eggs – round stage) of the scleractinian
82 coral *M. capitata* to three different near shore stressors for 24 h: high sediment concentrations
83 (100 mg l⁻¹ and 200 mg l⁻¹), lowered salinity (28 ‰), and elevated temperature (31° C) (Fig.1).
84 Exposure to stressors was independent and none were combined. Criteria selection for stressor
85 levels include: 1) accurate representation of Kāne'ohe Bay conditions (Table 1), and 2)
86 comparability with historical studies conducted with other Pacific species.

87 The purpose of this study was to determine whether the different stressors have
88 deleterious effects on the early life stages of *M. capitata*: 1) fertilization, 2) larval survival, and
89 3) settlement. Maintenance of early life stages of coral are very important for reef resilience and
90 recovery. Therefore, understanding how these common stressors affect the early life stages is
91 imperative.

92

93 **Methods and Materials**94 *Location and study species*

95 These experiments were conducted at the Hawai'i Institute of Marine Biology (HIMB) in
96 Kāneʻohe Bay (21°28'N; 157°48'W) on the windward side of Oʻahu, under the Hawai'i
97 Department of Land and Natural Resources Special Activity Permit No. SAP 2015-48. *M.*
98 *capitata* is the second most common coral species in Kāneʻohe Bay and the third most common
99 throughout the Hawaiian Islands (Jokiel et al. 2004; Rodgers 2005). *M. capitata* is a
100 hermaphroditic broadcast-spawner, releasing positively buoyant egg and sperm bundles during
101 the months of May through August between the hours of 22:45 and 22:30 on and , 2–3 days
102 following the new moon (Kolinski & Cox 2003). Thirty-four adult coral were collected from
103 multiple patch reefs throughout Kāneʻohe Bay on 14 May and 12 June, 2015 and transported
104 back to tanks at HIMB. Coral were held in seawater tables continuously supplied with 27° C
105 seawater pumped from the adjacent reef at 2 m depth. Approximately one hour prior to
106 spawning, coral colonies were isolated into individual containers. The exposure experiment was
107 replicated on three spawning nights during July 2015 with a total starting sample size of
108 approximately 33 for each treatment (Table 2). Variability in the quality and the resultant
109 sensitivity of coral gametes and their larvae from different nights of spawning is well known
110 (Hédouin & Gates 2013). However, this study attempted to compensate for this variability by
111 having as many unique parental crosses as possible (approx. 33). Ideally,
112 this physiological variation would be averaged and thus minimized when studied across all of the
113 crosses. Vials with less than 10 eggs were not included for analysis due to the high likelihood
114 that only one egg-sperm bundle was added.

115

116 ***Sediment treatment***

117 Three concentrations of sediment were used: 0 mg l⁻¹, 100 mg l⁻¹, and 200 mg l⁻¹. These
118 concentrations were chosen to mimic suspended sediment concentrations during/after a
119 largescale rain event (Table 1). Terrigenous red clay was collected from a historically
120 undisturbed hillside at the highest elevation on Moku o Lo'e Island at HIMB. Sediment was
121 sorted using a standard sieve to <63 µm. This sediment size was chosen to mimic natural
122 suspended sediment grain size. This clay/silt fraction was added to a 1 L beaker of seawater,
123 allowed to settle and the clear supernatant decanted to concentrate into a sediment slurry that
124 could be more accurately measured into 100 and 200 mg l⁻¹. The remaining sediment slurry was
125 used to make sediment solutions. Sediments were not allowed to desiccate completely, which
126 can alter the chemical composition and also impede re-suspension (Jokiel 1986). A wet weight to
127 dry weight ratio was determined in order to obtain an accurate suspended sediment
128 concentration. The wet slurry was weighed and dried several times to obtain an accurate wet to
129 dry weight ratio. The day of the experiment, the sediment slurry was weighed to the nearest
130 milligram on a Mettler Toledo XS403S scale (Columbus, OH) in a plastic weigh pan and added
131 to 1 L of filtered seawater (FSW). This was repeated for each sediment concentration. All FSW
132 was filtered through a Millipore Type GS 0.22 µm filter.

133

134 ***Salinity treatment***

135 Two salinities were used (ambient 34 ‰ and 28 ‰) to determine the effect of lowered
136 salinity (28 ‰) on fertilization success of *M. capitata*. Treatment seawater was obtained from the
137 seawater system at HIMB and salinity was measured prior to use. Salinity was measured in parts

138 per thousand (‰) using a YSI Model 556 conductivity meter (Yellow Springs, OH). The 28 ‰
139 treatment was prepared using filtered Kāneʻohe Bay seawater diluted with filtered freshwater to
140 obtain the desired salinity. The salinity of 28 ‰ was selected based on previous studies which is
141 representative of salinities measured on nearshore reefs during flood events (Hedouin et al. 2015;
142 Humphrey et al. 2008).

143

144 *Temperature treatment*

145 Two temperature treatments were used in these experiments, ambient (27° C during the
146 summer months), and an elevated temperature of 31° C. The elevated temperature represents 2°
147 C above the summer thermal maximum, a temperature that elicits the stress response of
148 bleaching in adult coral over a short time period (Jokiel 1977). The elevated temperature of 31°
149 C was chosen because during the 2014 bleaching event in Kāneʻohe Bay the maximum mid-day
150 temperature was between 30 and 31° C (Bahr et al. 2015b).

151 All ambient, sediment, and salinity vials containing gametes were secured in floating
152 foam racks and were placed in water tables to maintain temperature (27° C) and simulate mild
153 wave motion similar to field conditions. The elevated temperature treatment (31° C) vials were
154 placed in a heated water bath and were also secured in floating foam racks. Temperature was
155 controlled with an aquarium heater and an Onset HOBO Pendant[®] measured the temperature
156 over the time of exposure. Loggers have an accuracy of $\pm 0.21^{\circ}$ C from 0° to 50° C ($\pm 0.38^{\circ}$ F
157 from 32° to 122° F). Additional laboratory calibrations were conducted at 0 and 35° C to assure
158 precision and account for any drift in calibrations. Mild agitation was achieved using aquarium
159 power heads that gently circulated water in the bath.

160

161 ***Fertilization***

162 Coral were placed into individual 11 L containers prior to spawning and water was
163 allowed to flow in and around the containers to maintain constant water temperature. Unlike
164 other fertilization experiments (Gilmour 1999; Hedouin et al. 2015), egg bundles were not
165 pooled since *M. capitata* eggs contain a toxin that will effectively kill sperm within minutes if
166 their membrane is even slightly damaged (Hagedorn et al. 2015). Pooling eggs requires more
167 handling and increases the chance of damaging membranes and releasing the toxin. Therefore,
168 fertilization was accomplished by egg-sperm bundle crosses from two individuals in 15 ml
169 scintillation vials (Maté et al. 1997). A “cross” consisted of two parent colonies and one egg-
170 sperm bundle from each individual *M. capitata* colony was carefully placed into 4.9 ml of
171 filtered seawater (FSW) using a pipette. Each bundle addition contained 0.05 ml of seawater for
172 a final volume of 5 ml in each 15 ml glass scintillation vial. To determine the level of self-
173 fertilization, two egg-sperm bundles from the same individual were placed in a vial – no colonies
174 used in this experiment self-fertilized. The average sperm concentration of an *M. capitata* egg-
175 sperm bundle is approximately 5×10^5 cells/ml (Hagedorn et al. 2016). The bundle-bundle cross
176 method has been shown to produce the optimal sperm concentration ($\sim 1 \times 10^6$ cells/ml) for
177 fertilization in *M. capitata* (Maté et al. 1997). Over the three nights of spawning, 33 unique
178 crosses were obtained. Each cross had 6 scintillation vials containing 1 bundle from each parent:
179 1) ambient conditions; 2) medium suspended sediment; 3) high suspended sediment; 4) low
180 salinity; and 5) high temperature (Table 3), and 6) self fertilization control (2 bundles from same
181 colony). Each treatment was separate and exposed to each cross in a scintillation vial for 24 h
182 (Fig. 1).

183 Bundles separated approximately 20 minutes after spawning and the number of eggs per vial
184 was recorded using a Wild M5 dissecting microscope at 100x magnification. The fertilization
185 count was recorded the following morning, approximately eight hours after spawning.

186

187 *Larval survival*

188 Twenty-four hours after spawning with continued exposure to the treatments, the
189 embryos were moved to clean vials with 5 ml of FSW in each vial. The number of swimming
190 larvae was determined 48 hours following spawning. The percent larval survival was calculated
191 by dividing the number of swimming larvae by the number of fertilized eggs.

192

193 *Settlement*

194 After counting the number of swimming larvae in each scintillation vial, larvae were
195 moved into 10 ml petri dishes. A 1 cm² chip of crustose coralline algae (CCA) and
196 approximately 10 ml of FSW were added to each dish and were covered loosely with a lid. Every
197 two days fresh FSW was added to petri dishes and each dish and CCA chip were checked for
198 settlement during water changes. Fourteen days after spawning, settlement was determined by
199 counting the number of settled larvae. Negri et al. (2001) define settlement as a planula that is
200 still pear-shaped but has attached its aboral end to a hard substrate, whereas metamorphosis
201 involves a morphological and physiological change (i.e., flattened and with septal mesentery
202 radiating from mouth). Larvae were considered “settled” if they had metamorphosed. The
203 percent settlement was calculated by dividing the number of metamorphosed coral by the
204 number of swimming larvae.

205

206 *Statistical Analyses*

207 All count data for fertilization, larval survival, and settlement were counted and
208 represented as percentages. These values were transformed using an arcsine square root
209 transformation as recommended for proportional data. Data are graphically displayed as percent
210 in figures to better visualize the proportions and all error bars are Standard Error of the Mean
211 (SEM). Each treatment was individually compared to the control using a One-way ANOVA with
212 Dunnett's Method using the statistical software program JMP Pro 12. P-values ≤ 0.05 were
213 considered statistically significant.

214

215 **Results**

216 *Fertilization*

217 None of the treatments were statistically higher or lower than control fertilization success
218 of *M. capitata* (ANOVA, $p > 0.05$). The mean fertilization success for ambient conditions was
219 approximately $66.8 \pm 5.7\%$. All suspended sediment treatments showed slightly higher mean
220 fertilization success; $77.8 \pm 4.05\%$ in the 200 mg l^{-1} treatment and $68.7 \pm 6.3\%$ fertilization in the
221 100 mg l^{-1} treatment (Fig. 2). The low salinity treatment (28 ‰, $n = 31$) decreased mean
222 fertilization success to $52.2 \pm 5.9\%$, but was not statistically different from the ambient salinity
223 of 34 ‰ at $66.8 \pm 5.9\%$. The elevated temperature treatment (31°C , $n = 29$) also produced a
224 lower mean fertilization to $50.0 \pm 6.6\%$, but was not different than the control.

225

226 *Larval survival*

227 Although fertilization appeared to be resilient to the treatments, larval survival (% of
228 swimming larvae from fertilized eggs) decreased drastically in all of the treatments, especially

229 hypo-osmotic conditions (Fig. 2). Control larvae under ambient conditions had the highest
230 survival, $67.3 \pm 10.7\%$ (Fig. 2) and were statistically different than all of the treatments
231 (ANOVA, $p < 0.05$). Embryos at the round, non-motile stage were removed from the treatments
232 and placed into clean seawater for the remainder of the experiment; differences in larval survival
233 were due to latent effects of the treatments.

234 Both elevated sediment treatments had much lower larval survival when compared to the
235 control ($p < 0.05$, $F = 13.9$). The 200 mg l^{-1} sediment treatment had higher larval survival than the
236 100 mg l^{-1} treatment ($24.47 \pm 5.5\%$, $11.6 \pm 5.2\%$). Elevated temperature produced the highest
237 mean larval survival of all treatments, $36.4 \pm 8.6\%$. Salinity had the most dramatic effect on
238 larval survival. The low salinity treatment had the lowest percent larval survival of all the
239 treatments, $1.14 \pm 6.1\%$ (ANOVA, $p < 0.05$, $F = 38.8$).

240

241 *Settlement*

242 There was no difference in percent settlement when treatments were compared with the
243 control, $21.1 \pm 3.8\%$ (Fig. 2). However, the means were up to 50% different in the respective
244 treatments. The 100 mg l^{-1} sediment treatment had a greater mean percent settlement than the
245 control ($39.5 \pm 14.3\%$), and the 200 mg l^{-1} sediment treatment had a lower mean settlement (20.0
246 $\pm 6.1\%$) (Fig. 2). There was no settlement in the salinity treatment and, due to low survival, no
247 statistical analysis was conducted. The high temperature treatment had a lower mean percent
248 settlement than the control ($25.3 \pm 8.3\%$) (Fig. 2).

249

250 **Discussion**

251 The purpose of the present study was to identify the direct and latent effects of a 24-hour
252 exposure to suspended sediment (100 mg l⁻¹ and 200 mg l⁻¹), lowered salinity (28 ‰), and
253 elevated temperature (31 °C) on the early life stages of the Hawaiian scleractinian coral *M.*
254 *capitata*. The stressors directly affected fertilization while larval survival and settlement were
255 due to latent effects of the 24-hour exposure. Results of this experiment show that fertilization
256 and settlement were not affected by any of the stressors, and larval survival was negatively
257 affected by all of the near shore stressors (suspended sediment, lowered salinity, and elevated
258 temperature).

259

260 ***Effects of increased sediment on early life stages***

261 Neither suspended sediment treatment (100 and 200 mg l⁻¹) had an effect on fertilization
262 but both decreased larval survival (Fig. 2). Previous and concurrent studies from the Great
263 Barrier Reef found that sediment decreased both fertilization and larval survival (Gilmour 1999;
264 Humphrey et al. 2008). Gilmour (1999) found that suspended sediment as low as 50 mg l⁻¹
265 inhibited fertilization in *Acropora digitifera*. Additionally, Humphrey et al. (2008) saw reduced
266 fertilization in *A. millepora* when exposed to suspended sediments (100 mg l⁻¹ and 200 mg l⁻¹).
267 Gilmour (1999) exposed *Acropora digitifera* larvae to suspended sediment (20 mg l⁻¹ and 50 mg
268 l⁻¹) and there was significantly greater mortality (>98%) in the sediment treated larvae.

269 Ricardo et al. (2015) found that fertilization decreased in the presence of suspended
270 sediment, which was compounded by lowering the concentration of sperm available. When
271 exposed to suspended sediment concentrations of 230 mg l⁻¹ and 700 mg l⁻¹, they determined that
272 2-37 fold more sperm was needed in order to equal fertilization rates seen in sediment-free

273 treatments. In this study we used an optimal sperm concentration for fertilization in *M. capitata*
274 (Maté et al. 1997) which may have made fertilization more resilient to the sediment treatments.

275 Humanes et al. (2017) performed similar experiments to this study but also looked at the
276 combined effects of sediment, temperature, and nutrients on the early life stages of *Acropora*
277 *tenuis*. They found that fertilization was most sensitive to high suspended sediments (100 mg l⁻¹)
278 while larval survival and settlement were not affected.

279 In Hawai‘i, studies have shown that the coral in Kāne‘ohe Bay show some resilience to
280 sediment during gametogenesis and larval survival but not settlement. Padilla-Gamiño et al.
281 (2014) looked at the effect of sediment on gametogenesis in *M. capitata* and found no difference
282 in gamete production between sites with high and low sediment regimes. Perez III et al. (2014)
283 exposed larvae of *Pocillopora damicornis* to substrate covered in varying levels of fine sediment
284 (0.008-0.08 mm). They found that larval survival was not impacted by sediment but a thin layer
285 (>0.9 mg cm⁻²) of fine sediment could completely block settlement of larvae. Ricardo et al.
286 (2017) also found that very thin layers of deposited sediment can block settlement and this was
287 consistent regardless of sediment type (carbonate and siliciclastic) and particle size (fine and
288 coarse silt).

289 Sediment type and composition has been shown to have varied effects on coral
290 fertilization (Ricardo et al. 2018). Sediments with high organic-clay or certain minerals (i.e.,
291 Bentonite) decreased fertilization even at low suspended sediment concentrations. In contrast,
292 terrigenous sediments with low organic matter only decreased fertilization at high suspended
293 sediment levels (>100 mg l⁻¹). In this study we used terrigenous red clay but did not analyze the
294 sediment for its organic or mineral composition. Low organic composition could explain
295 fertilization resilience to suspended sediments seen in this study.

296

297 ***Effects of low salinity on early life stages***

298 Scleractinian corals are known to be stenohaline and osmoconformers. Corals do not
299 have a developed physiological regulatory system thus, osmotic stress on corals may cause
300 damage at the cellular level. A rapid increase in the induction of heat shock proteins may result
301 from changes in salinity (Seveso et al. 2013). Lowered salinity did not have an effect on
302 fertilization but did decrease larval survival and settlement. Kāneʻohe Bay has nine perennial
303 streams that feed directly into it, which could explain why *M. capitata* appear to have some
304 resilience to low salinity during fertilization. A few other studies have examined how low
305 salinity affects fertilization, larval survival, and settlement but none have included Hawaiian
306 coral. Humphrey et al. (2008) exposed *Acroporid* gametes to different salinities (28 to 36 ‰) and
307 documented reduced fertilization at 30 ‰ and no fertilization at 28 ‰. Similarly, Hedouin et al.
308 (2015) exposed gametes of two *Acroporid* species to different levels (26 to 36 ‰) of salinity and
309 found that salinities ≤ 28 ‰ (26.6 and 27.1 ‰) reduced fertilization success in both species.
310 Hedouin et al. (2015) also found that lowered salinity decreased larval survival. The results from
311 the present study are more consistent with those of Chui & Ang Jr. (2015). They exposed
312 gametes of *Platygyra acuta* to several different salinities and found that fertilization success was
313 statistically the same from 32 to 28 ‰ with significant decreases at 26 ‰, suggesting that some
314 species of coral may be more tolerant to lowered salinities.

315

316 ***Effects of high temperature on early life stages***

317 Exposing the early life stages of *M. capitata* to elevated temperature for 24 h following
318 spawning had negative latent effects on larval survivorship but did not directly impact

319 fertilization. Gametes and embryos held at high temperature during this experiment had a lower
320 mean fertilization but it was not different from the percent fertilization of the control. The results
321 from the present study are consistent with studies performed on coral from Okinawa where Negri
322 et al. (2007) found that *Favites chinesis* had high fertilization success even at 31.8° C (~79-
323 91%). Other studies have shown that elevated temperature negatively affects larval survival and
324 settlement. Bassim & Sammarco (2003) and Edmunds et al. (2001) found elevated temperature
325 treatments increased mortality in coral larvae. Coral settlement was also found to be negatively
326 affected by elevated temperature (Bassim & Sammarco 2003; Randall & Szmant 2009).
327 Humanes et al. (2017) found that elevated temperature (31 and 32° C) decreased fertilization,
328 larval development, and settlement in *Acropora tenuis*.

329 In this study, there was no statistical difference in the percent settlement of *M. capitata*
330 larvae between the control and the treatments. However, the results of the settlement experiment
331 are lacking due to low replication at the settlement stage (Table 2). Following the same cohort of
332 gametes and embryos through fertilization, larval survival, and settlement resulted in important
333 information about the lasting effects of a short-term exposure to stressors. However, this led to
334 an inherent loss of sample size for settlement results.

335

336 **Conclusion**

337 Kāneʻohe Bay is a calm, protected lagoon and *M. capitata* is one of the major reef
338 building coral there. As climate change accelerates, more frequent episodic largescale storms and
339 hurricanes will impact the main Hawaiian Islands (Li et al. 2018). Storms may produce a lens of
340 warm hyposaline water with increased suspended sediments that result in high mortality in the
341 early life stages of *M. capitata* when spawning is synchronous to flooding. Our study showed

342 that the stressors did not affect fertilization, but there were negative latent effects on larval
343 survival. Most notably, gametes and embryos exposed to salinity of 28 ‰ for 24 hours had less
344 than 1% larval survival while those exposed to sediment decreased larval survival by
345 approximately 55-81% and increased temperature reduced larval survival by 48% as compared to
346 the control.

347 This study reveals new, valuable information on how near shore stressors such as runoff
348 and elevated temperature affect the early life stages of a Hawaiian coral. Results that increase the
349 understanding of the impact of local stressors on early life stages can provide managers with
350 sound science to develop management strategies for the conservation and protection of coral
351 reefs. Managers will be able to use this information in coral reef management programs such as
352 outplanting coral fragments. Outplanting of reef building corals into areas that have lost cover or
353 onto artificial reefs has become a popular method of reef restoration. Growth of adult colonies
354 and asexual reproduction through fission or fragmentation can increase coral cover, but the long-
355 term success of these outplanted populations depends on the genetic diversity and successful
356 recruitment of sexually produced offspring. Understanding how the early life stages of corals are
357 affected by near shore stressors will assist managers with outplanting corals in suitable habitat
358 for adult growth and reproduction as well as recruit new larval corals.

359 Future studies involving all of the near shore stressors should include varying
360 concentrations of sperm, combinations of stressors, species with different reproductive strategies
361 (brooding vs. spawning), and multiple coral species. It is important to use different
362 concentrations of sperm because an optimal sperm concentration is 1) not realistic for *in situ*
363 concentrations and 2) could mask deleterious effects of the near shore stressors. Also,
364 availability of sperm has been shown to have a strong influence on successful fertilization in

365 corals (Ricardo et al. 2015). It is also important to see whether the effects of near shore stressors
366 change when combined. Therefore, a factorial design where near shore stressors are combined to
367 determine if effects are additive, synergistic, or antagonistic (Chui & Ang Jr. 2015; Humanes et
368 al. 2017). Lastly, it is important to study the effects of near shore stressors on different species of
369 coral and corals with different reproductive strategies. Other studies have shown that different
370 coral species from the same reef system respond differently to stressors (Hedouin et al. 2015;
371 Negri et al. 2007).

372 It is important for the conservation and protection of coral reefs that effects of near shore
373 stressors on early life stages of corals be studied. The resilience and recovery of coral reefs is
374 highly dependent on successful reproduction and settlement of larval corals.

375

376

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380

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382

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Table 1 (on next page)

Summary of normal and extreme water quality for Kāneʻohe Bay, Oʻahu

This table summarizes the average/normal values and extreme (i.e., post-storm or bleaching) for suspended sediment (mg l^{-1}), salinity (‰) and temperature ($^{\circ}\text{C}$) in Kāneʻohe Bay, Oʻahu.

1

	Normal	Extreme (i.e., storm or bleaching event)
Sediment	13.1 ± 0.45 mg l ⁻¹ (Uchino 2004)	600-800 mg l ⁻¹ (Hoover & Mackenzie 2009)
	1-3 NTU (De Carlo et al. 2007)	8 NTU (De Carlo et al. 2007)
Salinity	34-35 ‰ (De Carlo et al. 2007)	>20 ‰ (De Carlo et al. 2007)
Temperature	27 °C – summer avg (Bahr et al. 2015)	30 -31 °C (Bahr et al. 2015)

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Table 2 (on next page)

Summary of sample sizes for the treatments

This table lists the sample sizes of the different treatments at the three early life stages: fertilization, larval survival, and settlement. There was an inherent loss of sample size throughout the experiment. *Due to low sample size, differences in settlement were not analyzed statistically.

1

	Fertilization	Larval survival	Settlement*
Control	33	28	22
Medium suspended sediment	28	24	5
High suspended sediment	30	30	15
Low salinity	31	29	1
High temperature	29	21	12

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Table 3 (on next page)

Summary of conditions for the treatments and control

There were five treatments - medium suspended sediment, high suspended sediment, low salinity, and high temperature - and the control. This table lists the suspended sediment, salinity and temperature values for each treatment and control.

1

	Suspended sediment (mg l⁻¹)	Salinity (‰)	Temperature (°C)
1. Control	0	34	27
2. Medium suspended sediment	100	34	27
3. High suspended sediment	200	34	27
4. Low salinity	0	28	27
5. High temperature	0	34	31

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Figure 1

Diagram of early life stages and exposure timing

Diagram illustrating the 24 h exposure of stressors to the early stages of *M. capitata*.

Exposure began prior to egg-sperm bundle breakup and lasted through the non-motile round stage. Fertilization success was measured at 8 h and 24 h. Then at 24 h, the embryos were transferred to clean vials (labeled treatment stressors) but now with filtered seawater with ambient salinity and temperature. These same treatments were assessed for larval survival at 48 h and were finally placed into petri dishes for settlement. Settlement was counted 14 days after spawning. Figure adapted from Jones et al. 2015.

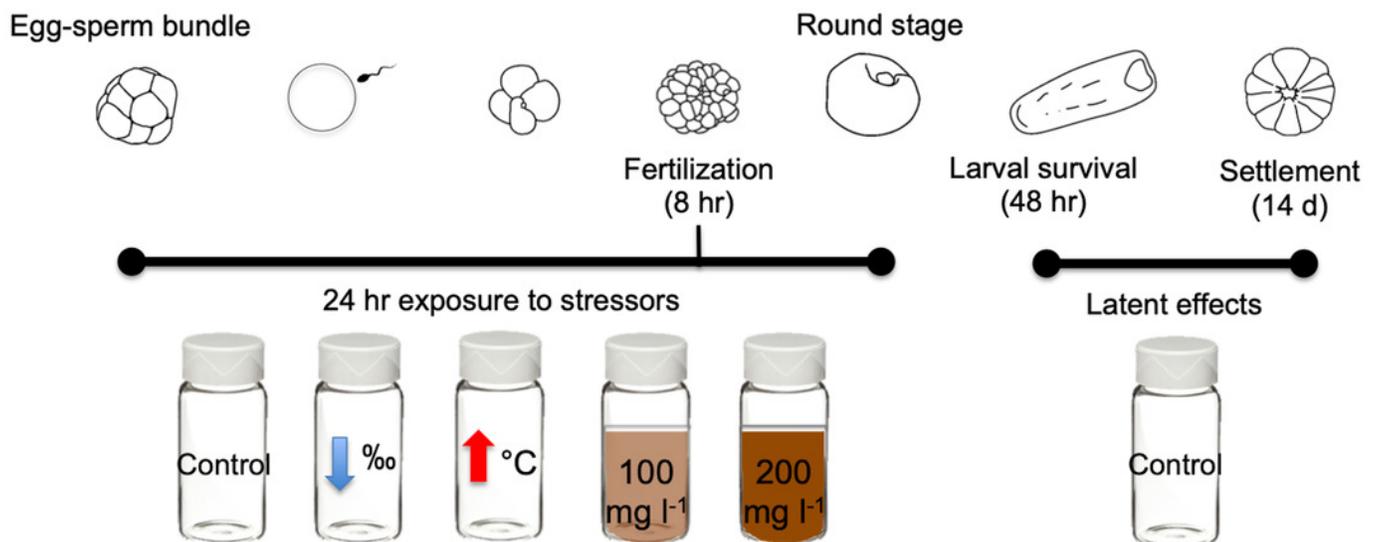


Figure 2

Effects of near shore stressors on early life stages

The early life stages of *M. capitata* were exposed to increased suspended sediment, low salinity, and increased temperature. The percent fertilization, larval survival, and settlement for each treatment and control is graphed.

