

Response of arbuscular mycorrhizal fungal community in soil and roots to grazing differs in a wetland on the Qinghai-Tibet plateau

Zhong-Feng Li^{1,2}, Peng-Peng Lü^{1,2}, Yong-Long Wang^{1,2}, Hui Yao^{1,2}, Pulak Maitra^{1,2}, Xiang Sun¹, Yong Zheng¹, Liang-Dong Guo^{Corresp. 1,2}

¹ State Key Laboratory of Mycology, Institute of Microbiology, Chinese Academy of Sciences, Beijing, China

² College of Life Sciences, University of Chinese Academy of Sciences, Beijing, China

Corresponding Author: Liang-Dong Guo
Email address: guold@im.ac.cn

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 5 Yong Zheng¹, Liang-Dong Guo^{1,2,*}

6 ¹ State Key Laboratory of Mycology, Institute of Microbiology, Chinese Academy of Sciences,
 7 Beijing 100101, China

8 ² College of Life Sciences, University of Chinese Academy of Sciences, Beijing 100049, China

9 * Corresponding author: guold@im.ac.cn (L.D. Guo)

ABSTRACT

Grazing as one of the most important disturbances affects the abundance, diversity and community composition of arbuscular mycorrhizal (AM) fungi in ecosystems, but the AM fungi in response to grazing in wetland ecosystems remain poorly documented. Here, we examined AM fungi in roots and soil in grazing and non-grazing plots in Zoige wetland on the Qinghai-Tibet plateau. Grazing significantly increased AM fungal spore density and glomalin-related soil proteins, but had no significant effect on the extra radical hyphal density of AM fungi. AM fungal richness was significantly lower in roots than in soil, but not significantly influenced by grazing. AM fungal community composition was significantly different between roots and soil, and was significantly influenced by grazing in soil but not in roots. This finding may enhance our understanding of the AM fungi in response to grazing in the wetland on the Qinghai-Tibet plateau.

Keywords: Arbuscular mycorrhizal fungal abundance, Community, Diversity, Grazing, Wetland

INTRODUCTION

Wetlands cover about 6% of the land surface on the earth and have high species diversity, including many endemic species (*Junk et al., 2013*). In China, wetlands account for 7% of the wetland on the world (*Junk et al., 2013*) and have about 225 families, 815 genera and 2,276 species of higher plants (*Yan & Zhang, 2005*). Wetlands provide important ecological functions in water resource conservation and quality purification, climate regulation, substance circulation and regional ecological balance maintenance (*Barbier, Acreman & Knowler, 1997; Li, 2001; Chmura et al., 2003; Green et al., 2017*). Moreover, as an important carbon (C) pool, wetlands can reduce the impact of increased greenhouse gases on global climate change (*Frolking et al., 2011*). However, wetland ecosystems have suffered severe degradations in recent decades due to global warming, intense resource exploitation, changes in hydrology and human disturbance (*Xiang et al., 2009; Junk et al., 2013; Meng et al., 2016*). For example, in human activities overgrazing has affected biodiversity, productivity, community stability and soil C cycling in wetlands (*Wang et al., 2012; Hoffmann et al., 2016; Zhou et al., 2017*).

Arbuscular mycorrhizal (AM) fungi, as one of the key components of soil microorganisms, form symbiotic associations with most terrestrial plant species (*Smith & Read, 2008*). In the AM associations, plants provide C source for the growth and function of fungi, thereby affecting the community of AM fungi (*Bonfante & Genre, 2010*). On the contrary, AM fungi can increase the nutrient and water absorption of host plants, then affecting plant community and productivity (*van der Heijden, Bardgett & van Straalen, 2008*). Furthermore, glomalin-related soil protein (GRSP) produced by AM fungi can stably exist in soil and play an important role in soil C pool (*Rillig, Field & Allen, 2001; Godbold et al., 2006*). In addition, AM fungi can improve plants to

tolerate grazing and other stresses from the environment (Bennett & Bever, 2007; Sharifi, Ghorbanli & Ebrahimzadeh, 2007). Thus, revealing the AM fungi in response to grazing is of great importance for understanding the diversity maintenance and community stability of plants in ecosystems.

Previous studies have demonstrated that the effect of grazing on AM fungi is depended on grazing intensity (García & Mendoza, 2012; Shi et al., 2017; Kusakabe et al., 2018; Yang et al., 2019). For instance, the light and moderate grazing intensity positively influenced AM fungal spore density in grasslands in Jilin province, China (Ba et al., 2012) and in British Columbia, Canada (van der Heyde et al., 2017). In contrast, over-grazing negatively affected AM fungal spore density in a semi-arid grassland in China (Su & Guo, 2007). Moderate grazing intensity did not influence AM fungal extra radical hyphal density in an alpine meadow in China (Yang et al., 2013a), but high grazing pressure negatively affected the extra radical hyphal density of AM fungi in a semi-arid grassland in China (Ren et al., 2018). Besides, moderate grazing had a neutral effect on AM fungal richness in a meadow in China (Ba et al., 2012). Moderate grazing significantly affected the community composition of AM fungi in soil and roots in grassland ecosystems (Bai et al., 2013; Yang et al., 2013a; Kusakabe et al., 2018). In contrast, others found that moderate grazing did not influence the community composition of AM fungi in roots in alpine meadow (Jiang et al., 2018) and in soil in mountain grassland (van der Heyde et al., 2017) ecosystems. However, previous studies have mainly focused on semi-arid, arid, alpine and mountain grassland ecosystems (Epelde et al., 2017; Jiang et al., 2018; Kusakabe et al., 2018; Ren et al., 2018). So far, we know little about how grazing affects AM fungi in wetland ecosystems.

Zoige wetland is a typical representative of the alpine wetland ecosystem on the Qinghai-

Tibet plateau in China, and has high plant species diversity and an important C sink function (Guo *et al.*, 2013). However, Zoige wetland has suffered severe ecosystem degradations since the 1970s, due to global warming, low precipitation and human disturbance, such as ditching for grassland enlargement, peat exploitation and over-grazing (Xiang *et al.*, 2009; Guo *et al.*, 2013; Wang *et al.*, 2017). Previous studies have mainly focused on plant diversity, microbial community (archaea group), and ecosystem conservation and restoration in the Zoige Wetland (Wang, Bao & Yan, 2002; Zhang *et al.*, 2008; Xiang *et al.*, 2009). However, the grazing effect on AM fungi has never been studied.

In order to reveal the AM fungi in response to grazing in wetland ecosystem, we established non-grazing (natural) and moderate grazing plots in Zoige wetland on the Qinghai-Tibet plateau. AM fungal spore density, extra radical hyphal density and GRSP content were examined in grazing and non-grazing plots. We examined the communities of AM fungi in roots and soil by Illumina MiSeq sequencing of 18S rDNA region. We hypothesize that: (H1) moderate grazing increases AM fungal spore density and GRSP content, but does not change the extra radical hyphal density of AM fungi, and (H2) moderate grazing changes the community composition of AM fungi but not richness in roots and soil in the Zoige wetland.

MATERIALS AND METHODS

Study site and sampling

The study was carried out in the centre of Zoige Swamp in the Zoige National Nature Reserve on the Qinghai-Tibet plateau (33°25'-34°80'N, 102°29'-102°59'E, 16,671 ha, 3365 m above sea level). The site has a plateau cold temperate humid monsoon climate, with a mean

annual temperature (MAT) of 1.1 °C, and a mean annual precipitation (MAP) of 660 mm. The site begins to freeze in late September and is completely thawed in mid-May (Wang, Bao & Yan, 2002). The abundant plant species are *Blymus sinocompressus*, *Potentilla anserina*, *Carex enervis*, *Caltha scaposa*, *Elymus nutans* and *Leontopodium wilsonii* in the site (Wang, Bao & Yan, 2002).

We established 20 plots (each 1 m × 1 m), > 20 m away from each other, in non-grazing (natural grass) and grazing area, respectively (Supplementary Fig. S1). The average species number and height of vegetation were 7.8 ± 0.495 (mean ± SE) and *ca.* 31 cm in the non-grazing plots and 5.4 ± 0.255 and *ca.* 7 cm in the grazing plots. This site was mainly grazed by yak. The grazing intensity in this study site was described as moderate (He *et al.*, 2000). In July 2018, we randomly collected five soil cores (3 cm in diameter; 15 cm in depth; *ca.* 300 g) and mixed into one composite sample from each plot. A total of 40 samples were obtained, packed in an ice box and transported to our laboratory. Soil samples were sieved (1-mm sieve) to remove debris and roots. Subsoil samples were kept at −80 °C until the extraction of fungal hyphae and DNA, and the remaining subsoil samples were air dried and kept at 10 °C until the analysis of AM fungal spore density, GRSP content and soil properties. We manually collected the mixed roots (< 2 mm in diameter) from each sample, washed with sterilized deionized water and kept at −80 °C until DNA extraction.

Soil property analysis

Soil moisture was measured using oven-drying at 105 °C for 24 h. Soil pH was measured at a ratio of 1:2.5 (w/v, soil: water) with a glass electrode (Thermo Orion T20, Columbia, USA). Soil total nitrogen (N) and C were determined by CHNOS Elemental Analyser (Vario EL III

Elementar Analysensysteme GmbH, Germany). Soil total phosphorus (P) was extracted using the HClO₄-H₂SO₄ digestion method and determined with a spectrophotometer (UV-2550, Shimadzu, Japan). The soil properties analyzed are shown in Supplementary Table S1.

EE-GRSP and T-GRSP

Total GRSP (T-GRSP) and easily extracted GRSP (EE-GRSP) were measured according to Wright & Upadhyaya (1998) and the modification method of Janos *et al.* (2008). We extracted EE-GRSP from 0.1 g air dried soil using sodium citrate buffer (8 mL, 0.02 M, pH 7.0) at 121 °C for 90 min in an autoclave (Yamato SQ810C, China). We repeatedly extracted T-GRSP from 0.1 g air dried soil using sodium citrate buffer (8 mL, 0.05 M, pH 8.0) at 121 °C for 90 min until no obvious color in the supernatant was observed. Supernatants were separated by centrifugation at 6000 g for 15 min to remove the soil particles and saved in a plastic tube (4 °C). Then 0.5 mL of supernatant of EE-GRSP and T-GRSP was stained with 5 mL of Coomassie Brilliant Blue G-250 and was read in a micro-plate reader (Biotek Synergy H4, Winooski, VT, USA) at 595 nm. The bovine serum albumin was used as a standard solution with Coomassie Brilliant Blue method and a standard curve was drawn to determine the content of EE-GRSP and T-GRSP.

AM fungal extra radical hyphal density and spore density

We extracted fungal hyphae from soil according to the membrane filter method (Rillig, Field & Allen, 1999). In total, 4.0 g of frozen soil from each sample was mixed with 12 mL sodium hexametaphosphate (35 g L⁻¹) and 100 mL distilled deionized water in a flask, and then blended for 30 s, settled for 30 min and sieved (38-µm sieve). The fungal hyphae on the sieve were washed into a flask with 200 mL distilled water, and then 2 mL aliquot was filtered through

a 25- μ m Millipore filter. The fungal hyphae on the filter were stained with 1% acid fuchsin and distinguished into AM and non-AM fungi on the basis of morphological characteristics and staining color (Miller, Jastrow & Reinhardt, 1995). We measured the hyphal length of AM fungi according to the grid-line intersect method (Tennant, 1975). We extracted AM fungal spores from 20 g air dried soil from each sample according to the wet-sieving and decanting method (Daniels & Skipper, 1982) and counted the spore numbers under 40 \times magnification (Nikon 80i, Japan).

DNA extraction, PCR and Illumina Miseq sequencing

We extracted DNA from 0.2 g frozen roots and soil using the PowerSoil[®] DNA isolation kit (MOBIO Laboratories, Inc., Carlsbad, USA) in accordance with the manufacturer's instructions, and measured the DNA concentration using a NanoDrop 1000 Spectrophotometer (Thermo Scientific, Wilmington, USA). We amplified the fungal 18S rDNA region using a two-step PCR procedure. The first PCR using primers AML2 (Lee, Lee & Young, 2008) and GeoA2 (Schwarzott & Schüßler, 2001) was conducted in a final 25 μ L reaction mixture, including *ca.* 10 ng of template DNA, 0.75 μ M of each primer, 250 μ M of each dNTP, 0.5 U KOD-plus-Neo polymerase (Toyobo, Tokyo, Japan), 1.5 mM MgSO₄, and 2.5 μ L 10 \times buffer. The thermal cycling conditions were performed as follows: an initial denaturation at 95 $^{\circ}$ C for 5 min, 30 cycles for denaturation at 94 $^{\circ}$ C for 1 min, annealing at 58 $^{\circ}$ C for 50 s and extension at 68 $^{\circ}$ C for 1 min, and a final extension at 68 $^{\circ}$ C for 10 min. The products of the first amplification were diluted 100 times, and 1 μ L of the diluted DNA template was used for the second amplification. The thermal cycling conditions for the second amplification were the same as first amplification, except that the primers NS31 (Simon, Lalonde & Bruns, 1992) and AMDGR (Sato *et al.*, 2005)

linked with 12-base barcode sequences were used. The size of amplified fragment was about 300 base pairs (bp). We purified the PCR products using a PCR Product Gel Purification Kit (Omega Bio-Tek, USA), and pooled the purified PCR products with the same amount (100 ng) from each sample and adjusted the concentration to 10 ng μL^{-1} . We constructed a sequencing library by addition of an Illumina sequencing adaptor (5'-GATCGGAAGAGCACACGTCTGAACTCCAGTCACATCACGATCTCGTATGCCGTCTTCTGCTTG-3') to the products using the Illumina TruSeq DNA PCR-Free LT Library Prep Kit (Illumina, CA, USA) according to the manufacturer's instructions. We sequenced the library by an Illumina MiSeq PE 250 platform using the paired-end (2×250 bp) option in the Chengdu Institute of Biology, Chinese Academy of Sciences, China.

Bioinformatics analysis

We filtered the raw sequences using Quantitative Insights into Microbial Ecology (QIIME) v.1.7.0 (*Caporaso et al., 2010*) to eliminate low-quality sequences, such as read length < 200 bp, no valid primer sequence or barcode sequence, containing ambiguous bases, or an average quality score < 20. We checked and deleted the potential chimeras against the MaarjAM database (*Öpik et al., 2010*) using the 'chimera.uchime' command in Mothur version 1.31.2 (*Schloss et al., 2009*). High quality sequences were subjected to de-replication and de-singleton, and then clustered into operational taxonomic units (OTUs) at a 97% sequence similarity level using the cluster_otus command in USEARCH v8.0 (*Edgar, 2013*). Using a basic local alignment search tool (BLAST) (*Altschul et al., 1990*), we selected the most abundant sequence of each OTU and searched against the MaarjAM database and National Center for Biotechnology Information (NCBI) nt database. We identified OTUs as the AM fungi based on the closest BLAST hit

annotated as ‘Glomeromycotina’ and E values $< e^{-50}$. Furthermore, we normalized the sequence number of each sample to the smallest sample size using the ‘sub.sample’ command in Mothur. We have submitted the representative sequence of each AM fungal OTU to the European Molecular Biology Laboratory (EMBL) database (accession no. LR736402-LR736557). The identified AM fungi are shown in Supplementary Table S2.

Statistical analysis

We conducted all statistical analyses in R version 3.3.2 (*R Development Core Team, 2017*). Tukey's honestly significant difference (HSD) test or Conover's test was used to examine the significant difference of soil moisture, pH, total N, total C and total P in the grazing and non-grazing plots at $P < 0.05$. One-way analysis of variance (ANOVA) was conducted to evaluate the effect of grazing on AM fungal spore density, extra radical hyphal density, T-GRSP and EE-GRSP, and then Tukey's HSD test was used to examine the significant difference between treatments at $P < 0.05$. Two-way ANOVA was conducted to evaluate the effect of grazing, sample type (soil and root) and their interaction on the OTU richness and the relative abundance of abundant OTUs (relative abundance $> 1\%$) and orders of AM fungi, and then Tukey's HSD test was used to examine the significant difference between treatments at $P < 0.05$. As these data (except for the relative abundance of OTU141) did not satisfy homogeneity of variance after log and square root transformation, nonparametric Kruskal–Wallis test was carried out, and then Conover's test was conducted for comparisons between grazing and non-grazing treatments in soil and roots using the post-hoc.kruskal.conover.test function in the PMCMR package (*Pohlert, 2014*).

The distance matrices of AM fungal community composition (Hellinger-transformed OTU

read data) in roots and soil were established by the Bray-Curtis method (*Clarke, Somerfield & Chapman, 2006*). Permutational multivariate analysis of variance (PerMANOVA) was conducted to examine the effect of sample type, grazing and their interaction on AM fungal community composition, using the ‘adonis’ function in the vegan with 999 permutations (*Oksanen et al., 2013*). Redundancy analysis (RDA) was conducted to reveal the significant correlation of AM fungal community composition and soil variables using the Monte Carlo permutation test with 999 permutations.

RESULTS

EE-GRSP and T-GRSP contents

The EE-GRSP content was $28.96 \pm 3.73 \mu\text{g g}^{-1}$ (mean \pm SE) and $25.71 \pm 2.26 \mu\text{g g}^{-1}$ in grazing and non-grazing treatments, respectively. The T-GRSP content was $96.7 \pm 18.82 \mu\text{g g}^{-1}$ and $71.87 \pm 12.87 \mu\text{g g}^{-1}$ in grazing and non-grazing treatments, respectively. One-way ANOVA showed that grazing significantly influenced EE-GRSP ($F_{1,38} = 9.907$, $P = 0.003$) and T-GRSP ($F_{1,38} = 23.57$, $P < 0.001$). For example, EE-GRSP and T-GRSP contents were significantly lower in non-grazing than in grazing treatments (Fig. 1a and b).

AM fungal spore density and extra radical hyphal density

The spore density of AM fungi was $25.89 \pm 12.17 \text{ g}^{-1}$ (mean \pm SE) and $15.03 \pm 5.88 \text{ g}^{-1}$ in grazing and non-grazing treatments, respectively. The extra radical hyphal density of AM fungi was $4.00 \pm 2.51 \text{ m g}^{-1}$ and $3.10 \pm 1.56 \text{ m g}^{-1}$ in grazing and non-grazing treatments, respectively. One-way ANOVA revealed that grazing significantly affected AM fungal spore density ($F_{1,38} =$

10.71, $P = 0.002$) but not extra radical hyphal density ($F_{1,38} = 2.000$, $P = 0.165$). For example, the spore density of AM fungi was significantly lower in non-grazing than in grazing treatments (Fig. 1c). No significant difference of the extra radical hyphal density of AM fungi in non-grazing and grazing treatments was observed (Fig. 1d).

Identification of AM fungi

In total, 3,205,557 high-quality sequences were filtered from 3,335,816 raw sequences and clustered into 882 OTUs at a 97% sequence similarity level. Among 882 OTUs, 156 (2,919,706 sequences) belonged to AM fungi. As the sequence number of AM fungi varied from 20,408 to 48,572 in the 80 samples, the number of sequence was normalized to 20,408. The normalized dataset contained 156 AM fungal OTUs (1,632,640 sequences). Of the 156 AM fungal OTUs obtained, 154 were from soil, 152 from roots, and 150 shared both soil and roots. Among 156 AM fungal OTUs, 153 were detected from more than three samples (frequency $\geq 3.75\%$) (Supplementary Fig. S2a). Furthermore, the 21 abundant AM fungal OTUs (relative abundance $> 1\%$) occupied 83.85% of the total sequences (Supplementary Fig. S2b). Among 156 AM fungal OTUs, 109 were identified to Glomerales (79.52% of sequences), 22 to Diversisporales (10.84%), 21 to Archaeosporales (8.75%), and 4 to Paraglomerales (0.89%). In addition, the rarefaction curves indicated that the sample numbers were sufficient to detect the most AM fungi in this study (Supplementary Fig. S3).

AM fungal OTU richness

AM fungal OTU richness in grazing and non-grazing treatments was 123.70 ± 2.96 (mean \pm SE) and 122.85 ± 3.01 in soil, and 117.55 ± 2.26 and 118.00 ± 4.38 in roots, respectively.

Kruskal–Wallis test revealed that AM fungal OTU richness was influenced by sample type (root and soil; $\chi^2 = 35.01$, $P < 0.001$), but not by grazing ($\chi^2 = 0.045$, $P = 0.832$). For example, AM fungal OTU richness was significantly lower in roots than in soil in both grazing and non-grazing treatments (Fig. 2). However, no significant difference of AM fungal OTU richness between grazing and non-grazing treatments in roots and soil was observed (Fig. 2).

AM fungal community

Two-way ANOVA and Kruskal–Wallis tests revealed that grazing had significant effect on the relative abundance of abundant AM fungal OTU13 and OTU141 (Glomerales), and sample type had significant effect on the relative abundance of abundant AM fungal OTU4, OTU5, OTU7, OTU12, OTU14, OTU15, OTU18, OTU25 and OTU141 (Glomerales), OTU8 and OTU17 (Diversisporales) and OTU23 (Archaeosporales) (Fig. 3). For example, the relative abundance of OTU5, OTU8, OTU15, OTU17, OTU23, OTU25 and OTU141 was significantly lower in roots than in soil, and that of OTU141 was significantly lower in non-grazing treatment than in grazing treatment (Fig. 3; Supplementary Table S3). By contrast, the relative abundance of OTU4, OTU7, OTU12, OTU14 and OTU18 was significantly lower in soil than in roots (Fig. 3; Supplementary Table S3). In the grazing treatment, the relative abundance of OTU5, OTU8, OTU15, OTU17, OTU23 and OTU141 was significantly higher in soil than in roots, while that of OTU4, OTU7, OTU14 and OTU18 was significantly lower in soil than in roots (Fig. 3; Supplementary Table S3). In the non-grazing treatment, the relative abundance of OTU5, OTU8, OTU17, OTU23 and OTU141 was significantly higher in soil than in roots, while that of OTU4, OTU7, OTU12, OTU14 and OTU18 was significantly lower in soil than in roots (Fig. 3; Supplementary Table S3). Besides, the relative abundance of soil OTU13 and OTU141 was

significantly lower in non-grazing treatment than in grazing treatment (Fig. 3; Supplementary Table S3).

Kruskal–Wallis test revealed that sample type but not grazing significantly influenced the relative abundance of Glomerales, Diversisporales and Archaeosporales, but not on Paraglomerales (Fig. 4). The relative abundance of Glomerales was significantly lower in soil than in roots; by contrast, the relative abundance of Diversisporales and Archaeosporales was significantly lower in roots than in soil, regardless of non-grazing and grazing treatments (Fig. 4).

The PerMANOVA demonstrated that the community composition of AM fungi was significantly influenced by sample type (soil and root; $F = 15.49$, $R^2 = 0.149$, $P = 0.001$) and grazing ($F = 2.617$, $R^2 = 0.025$, $P = 0.008$). Furthermore, the community composition of AM fungi was significantly influenced by grazing in soil ($F = 2.639$, $R^2 = 0.055$, $P = 0.001$), but not in roots ($F = 0.998$, $R^2 = 0.025$, $P = 0.419$). Furthermore, RDA showed that the community composition of AM fungi in soil and roots was significantly correlated with soil pH, moisture, total C, total N and total P (Fig. 5).

DISCUSSION

We found that grazing had positive effect on AM fungal spore density, EE-GRSP and T-GRSP, in consistent with some previous studies (Hammer & Rillig, 2011; Yang *et al.*, 2013a; van der Heyde *et al.*, 2017). Previous findings suggest that moderate removal of aboveground biomass may increase the allocation of C to the roots and exudation from roots to soil (Eom, Wilson & Hartnett, 2001; Hamilton *et al.*, 2008; Soka & Ritchie, 2018), which could be

beneficial for the sporulation of AM fungi (*Ba et al., 2012; van der Heyde et al., 2017*). Furthermore, since about 80% of GRSP is produced by the AM fungi, grazing increased AM fungal spore density, resulting in increasing GRSP content in soil (*Driver, Holben & Rillig, 2005*). However, grazing did not significantly influence AM fungal extra radical hyphal density, as reported in a previous study (*García & Mendoza, 2012*). Although moderate grazing may increase C allocation to the roots, this increase may be ephemeral (*van der Heyde et al., 2019*) and not be sufficient to promote the growth of AM fungal hyphae.

AM fungal richness was significantly lower in roots than in soil, as previous studies reported in alpine and meadow ecosystems (*Hempel, Renker & Buscot, 2007; Liu et al., 2012; Yang et al., 2013a*). This may be due to the seasonal nature of AM fungal communities (*Clark, Rillig & Nowak, 2009; Liu et al., 2009; Martínez-García et al., 2011*). Furthermore, the currently and formerly active propagules of AM fungi could remain in soil, but only currently active AM fungi could occur in the roots (*Liu et al., 2009; Martínez-García et al., 2011*). However, we found that grazing did not significantly influence AM fungal richness in roots and soil. Similarly, a previous study showed that moderate grazing could maintain the AM fungal diversity (*Dudinszky et al., 2019*). In general, AM fungi have low specificity (*Smith & Read, 2008*), thus AM fungal richness may not be influenced by the low plant species diversity caused by moderate grazing, as some studies found that AM fungal richness was not related to plant species diversity (*Wolf et al., 2003; Xiang et al., 2014*).

The community composition of AM fungi significantly differed between roots and soil in this study, as previous studies reported in grassland (*Yang et al., 2013a*), farmland (*Liu et al., 2016*) and temperate (*Saks et al., 2014*) and subtropical forest (*Maitra et al., 2019*) ecosystems. This may be explained by the difference in AM fungal abundance in roots and soil (*Hempel,*

Renker & Buscot, 2007; Varela-Cervero et al., 2015; Maitra et al., 2019). Indeed, our result found that some AM fungi were abundant in roots and soil, respectively. In addition, AM fungal phenology may produce different communities in soil and roots (*Pringle & Bever, 2002; Liu et al., 2012*).

Grazing significantly affected the AM fungal community composition in soil, in consistent with some previous studies reported in desert steppe and grassland ecosystems (*Murray, Frank & Gehring, 2010; Bai et al., 2013*). Grazing may influence the AM fungal community composition by changing soil properties through animal trampling and fecal deposition (*McNaughton, Banyikwa & McNaughton, 1997; Yang et al., 2013b; Liu et al., 2015; Yang et al., 2019*). For example, animal trampling may make the soil tight and alter soil bulk density (*Kobayashi, Hori & Nomoto, 1997; Kauffman, Thorpe & Brookshire, 2004; Byrnes et al., 2018*), thereby influencing AM fungal community (*Yang et al., 2018*). Moreover, dung and urine produced by animals, as soil fertilization, may decrease soil pH and increase soil nutrients as shown in this and previous studies (*McNaughton, Banyikwa & McNaughton, 1997; Kohler et al., 2005*), thus altering AM fungal community composition. Indeed, our result showed that the community composition of AM fungi was significantly related to soil pH, moisture, total C, total N and total P, as previous studies reported in semi-arid, alpine and temperate grassland and subtropical forest ecosystems (*Zheng et al., 2014; Gao et al., 2016; Zhang et al., 2016; Goldmann et al., 2019; Maitra et al., 2019*). However, the AM fungal community composition in roots was not significantly influenced by grazing, as previous studies reported in semi-arid and alpine grassland ecosystems (*González et al., 2018; Jiang et al., 2018*). It is possible that moderate grazing does not much change the allocation of carbohydrates to roots, thereby without altering AM fungal community. Furthermore, although grazing may alter the AM fungal function,

it does not necessarily alter the community in roots (*González et al., 2018*).

CONCLUSIONS

In conclusion, we examined the AM fungi in response to grazing in the Zoige wetland on the Qinghai-Tibet plateau for the first time. AM fungal spore density and GRSP content positively responded to grazing. The extra radical hyphal density and OTU richness of AM fungi had neutral response to grazing. The community composition of AM fungi significantly differed between roots and soil, and was significantly influenced by grazing in soil but not in roots. This finding may enhance our understanding of the AM fungi in response to grazing in the wetland ecosystem on the Qinghai-Tibet Plateau.

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REFERENCES

- Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ. 1990. Basic local alignment search tool. *Journal of Molecular Biology* 215: 403–410 DOI: 10.1016/S0022-2836(05)80360-2.
- Ba L, Ning JX, Wang DL, Facelli E, Facelli JM, Yang YN, Zhang LC. 2012. The relationship

between the diversity of arbuscular mycorrhizal fungi and grazing in a meadow steppe.

Plant and Soil 352: 143–156 DOI: 10.1007/s11104-011-0985-6.

Bai G, Bao YY, Du GX, Qi YL. 2013. Arbuscular mycorrhizal fungi associated with vegetation

and soil parameters under rest grazing management in a desert steppe ecosystem.

Mycorrhiza 23: 289–301 DOI: 10.1007/s00572-012-0468-5.

Barbier EB, Acreman M, Knowler D. 1997. Economic valuation of wetlands: a guide for policy

makers and planners. Gland, Switzerland: Ramsar Convention Bureau.

Bennett AE, Bever JD. 2007. Mycorrhizal species differentially alter plant growth and response

to herbivory. *Ecology* 88: 210–218 DOI: 10.1890/0012-

9658(2007)88[210:MSDAPG]2.0.CO;2.

Bonfante P, Genre A. 2010. Mechanisms underlying beneficial plant-fungus interactions in

mycorrhizal symbiosis. *Nature Communications* 1: 48 DOI: 10.1038/ncomms1046.

Byrnes RC, Eastburn DJ, Tate KW, Roche LM. 2018. A global meta-analysis of grazing impacts

on soil health indicators. *Journal of Environmental Quality* 47: 758–765 DOI:

10.2134/jeq2017.08.0313.

Caporaso JG, Kuczynski J, Stombaugh J, Bittinger K, Bushman FD, Costello EK, Fierer N, Pena

AG, Goodrich JK, Gordon JI, Huttley GA, Kelley ST, Knight D, Koenig JE, Ley RE,

Lozupone CA, McDonald D, Muegge BD, Pirrung M, Reeder J, Sevinsky JR, Tumbaugh

PJ, Walters WA, Widmann J, Yatsunenko T, Zaneveld J, Knight R. 2010. QIIME allows

analysis of high-throughput community sequencing data. *Nature Methods* 7: 335–336 DOI:

10.1038/nmeth.f.303.

Chmura GL, Anisfeld SC, Cahoon DR, Lynch JC. 2003. Global carbon sequestration in tidal,

saline wetland soils. *Global Biogeochemical Cycles* 17: 1111 DOI:

10.1029/2002GB001917.

Clark NM, Rillig MC, Nowak RS. 2009. Arbuscular mycorrhizal fungal abundance in the Mojave Desert: seasonal dynamics and impacts of elevated CO₂. *Journal of Arid Environments* 73: 834–843 DOI: 10.1016/j.jaridenv.2009.03.004.

Clarke KR, Somerfield PJ, Chapman MG. 2006. On resemblance measures for ecological studies, including taxonomic dissimilarities and a zero-adjusted Bray-Curtis coefficient for denuded assemblages. *Journal of Experimental Marine Biology and Ecology* 330: 55–80 DOI: 10.1016/j.jembe.2005.12.017.

Daniels B, Skipper H. 1982. Methods for the recovery and quantitative estimation of propagules from Soil. In: Schenck, N. (Ed.), *Methods and Principles of Mycorrhizal Research*, vol. 29. American Phytopathological Society, Minn, pp. 29–35.

Driver JD, Holben WE, Rillig MC. 2005. Characterization of glomalin as a hyphal wall component of arbuscular mycorrhizal fungi. *Soil Biology and Biochemistry* 37: 101–106 DOI: 10.1016/j.soilbio.2004.06.011.

Dudinszky N, Cabello MN, Grimoldi AA, Schalamuk S, Golluscio RA. 2019. Role of grazing intensity on shaping arbuscular mycorrhizal fungi communities in Patagonian semiarid steppes. *Rangeland Ecology & Management* 72: 692–699 DOI: 10.1016/j.rama.2019.02.007.

Edgar RC. 2013. UPARSE: highly accurate OTU sequences from microbial amplicon reads. *Nature Methods* 10: 996–998 DOI: 10.1038/nmeth.2604.

Eom AH, Wilson GW, Hartnett DC. 2001. Effects of ungulate grazers on arbuscular mycorrhizal symbiosis and fungal community structure in tallgrass prairie. *Mycologia* 93: 233–242 DOI: 10.1080/00275514.2001.12063153.

- Epelde L, Lanzén A, Mijangos I, Sarrionandia E, Anza M, Garbisu C. 2017. Short-term effects of non-grazing on plants, soil biota and aboveground-belowground links in Atlantic mountain grasslands. *Scientific Reports* 7: 15097. DOI: 10.1038/s41598-017-15345-1.
- Frolking S, Talbot J, Jones MC, Treat CC, Kauffman JB, Tuittila ES, Roulet N. 2011. Peatlands in the Earth's 21st century climate system. *Environmental Reviews* 19: 371–396 DOI: 10.1139/a11-014.
- Gao C, Kim YC, Zheng Y, Yang W, Cheng L, Ji NN, Wan SQ, Guo LD. 2016. Increased precipitation, rather than warming, exerts a strong influence on arbuscular mycorrhizal fungal community in a semiarid steppe ecosystem. *Botany* 94: 459–469 DOI: 10.1139/cjb-2015-0210.
- García I, Mendoza R. 2012. Impact of defoliation intensities on plant biomass, nutrient uptake and arbuscular mycorrhizal symbiosis in *Lotus tenuis* growing in a saline-sodic soil. *Plant Biology* 14: 964–971 DOI: 10.1111/j.1438-8677.2012.00581.x.
- Godbold DL, Hoosbeek MR, Lukac M, Cotrufo MF, Janssens LA, Ceulemans R, Polle A, Velthorst EJ, Scarascia-Mugnozza G, Angelis P, Miglietta F, Peressotti A. 2006. Mycorrhizal hyphal turnover as a dominant process for carbon into soil organic matter. *Plant and Soil* 281: 15–24 DOI: 10.1007/s11104-005-3701-6.
- Goldmann K, Boeddinghaus RS, Klemmer S, Regan KM, Heintz-Buschart A, Fischer M, Prati D, Piepho HP, Berner D, Marhan S, Kandeler E, Buscot F, Wubet E. 2019. Unraveling spatiotemporal variability of arbuscular mycorrhiza fungi in a temperate grassland plot. *Environmental Microbiology* DOI: 10.1111/1462-2920.14653.
- González JB, Petipas RH, Franken O, Kiers ET, Veblen KE, Brody AK. 2018. Herbivore removal reduces influence of arbuscular mycorrhizal fungi on plant growth and tolerance in

an east African savanna. *Oecologia* 187: 123–133 DOI: 10.1007/s00442-018-4124-4.

Green AJ, Alcorlo P, Peeters ET, Morris EP, Espinar JL, Bravo-Utrera M A, Bustamante J, Díaz-Delgado R, Koelmans AA, Mateo R, Mooij WM, Rodríguez-Rodríguez M, van Nes EH, Scheffer M. 2017. Creating a safe operating space for wetlands in a changing climate. *Frontiers in Ecology and the Environment* 15, 99–107 DOI: 10.1002/fee.1459.

Guo XJ, Du W, Wang X, Yang ZF. 2013. Degradation and structure change of humic acids corresponding to water decline in Zoige peatland, Qinghai-Tibet Plateau. *Science of the Total Environment* 445: 231–236 DOI: 10.1016/j.scitotenv.2012.12.048.

Hamilton EW, Frank DA, Hinchey PM, Murray TR. 2008. Defoliation induces root exudation and triggers positive rhizospheric feedbacks in a temperate grassland. *Soil Biology and Biochemistry* 40: 2865–2873 DOI: 10.1016/j.soilbio.2008.08.007.

Hammer EC, Rillig MC. 2011. The influence of different stresses on glomalin levels in an arbuscular mycorrhizal fungus-salinity increases glomalin content. *PLoS ONE* 6: e28426. DOI: 10.1371/journal.pone.0028426.

He CQ, Zhao KY, Zhao ZC. 2000. Wetlands pasture degeneration mechanism and its sustainable utilization countermeasure in Zoige plateau. *Grassland of China* 6: 11–16 (in Chinese).

Hempel S, Renker C, Buscot F. 2007. Differences in the species composition of arbuscular mycorrhizal fungi in spore, root and soil communities in a grassland ecosystem. *Environmental Microbiology* 9: 1930–1938 DOI: 10.1111/j.1462-2920.2007.01309.x.

Hoffmann C, Giese M, Dickhoefer U, Wan HW, Bai YF, Steffens M, Liu CY, Butterbach-Bahl K, Han XG. 2016. Effects of grazing and climate variability on grassland ecosystem functions in Inner Mongolia: synthesis of a 6-year grazing experiment. *Journal of Arid Environments* 135: 50–63 DOI: 10.1016/j.jaridenv.2016.08.003.

- 462 Janos DP, Garamszegi S, Beltran B. 2008. Glomalin extraction and measurement. *Soil Biology*
463 *and Biochemistry* 40: 728–739 DOI: 10.1016/j.soilbio.2007.10.007.
- 464 Jiang SJ, Pan JB, Shi GX, Dorji T, Hopping KA, Klein JA, Liu YJ, Feng H. 2018. Identification
465 of root-colonizing AM fungal communities and their responses to short-term climate
466 change and grazing on Tibetan plateau. *Symbiosis* 74: 159–166 DOI: 10.1007/s13199-017-
467 0497-0.
- 468 Junk WJ, An SQ, Finlayson CM, Gopal B, Květ J, Mitchell SA, Mitsch WJ, Robarts R D. 2013.
469 Current state of knowledge regarding the world’s wetlands and their future under global
470 climate change: a synthesis. *Aquatic Sciences* 75: 151–167 DOI: 10.1007/s00027-012-
471 0278-z.
- 472 Kauffman JB, Thorpe AS, Brookshire ENJ. 2004. Livestock exclusion and belowground
473 ecosystem responses in riparian meadows of eastern Oregon. *Ecological Applications* 14:
474 1671–1679 DOI: 10.1890/03-5083.
- 475 Kobayashi T, Hori Y, Nomoto N. 1997. Effects of trampling and vegetation removal on species
476 diversity and micro-environment under different shade conditions. *Journal of Vegetation*
477 *Science* 8: 873–880 DOI: 10.2307/3237032.
- 478 Kohler F, Hamelin J, Gillet F, Gobat JM, Buttler A. 2005. Soil microbial community changes in
479 wooded mountain pastures due to simulated effects of cattle grazing. *Plant and Soil* 278:
480 327–340 DOI: 10.1007/s11104-005-8809-1.
- 481 Kusakabe R, Taniguchi T, Goomaral A, Undarmaa J, Yamanaka N, Yamato M. 2018.
482 Arbuscular mycorrhizal fungal communities under gradients of grazing in Mongolian
483 grasslands of different aridity. *Mycorrhiza* 28: 621–634. DOI: 10.1007/s00572-018-0855-7.
- 484 Lee J, Lee S, Young JPW. 2008. Improved PCR primers for the detection and identification of

- arbuscular mycorrhizal fungi. *FEMS Microbiology Ecology* 65: 339–349 DOI: 10.1111/j.1574-6941.2008.00531.x.
- Li LK. 2001. Wetland and Ramsar convention. *World Forestry Research* 14: 1–7 (in Chinese).
- Liu N, Kan HM, Yang GW, Zhang YJ. 2015. Changes in plant, soil, and microbes in a typical steppe from simulated grazing: explaining potential change in soil C. *Ecological Monographs* 85: 269–286 DOI: 10.1890/14-1368.1.
- Liu W, Zhang YL, Jiang SS, Deng Y, Christie P, Murray PJ, Li XL, Zhang JL. 2016. Arbuscular mycorrhizal fungi in soil and roots respond differently to phosphorus inputs in an intensively managed calcareous agricultural soil. *Scientific Reports* 6: 24902 DOI: 10.1038/srep24902.
- Liu YJ, He L, An LZ, Helgason T, Feng HY. 2009. Arbuscular mycorrhizal dynamics in a chronosequence of *Caragana korshinskii* plantations. *FEMS Microbiology Ecology* 67: 81–92 DOI: 10.1111/j.1574-6941.2008.00597.x.
- Liu YJ, Shi GX, Mao L, Cheng G, Jiang SJ, Ma XJ, An LZ, Du ZG, Johnson NC, Feng H. 2012. Direct and indirect influences of 8 yr of nitrogen and phosphorus fertilization on Glomeromycota in an alpine meadow ecosystem. *New Phytologist* 194: 523–535 DOI: 10.1111/j.1469-8137.2012.04050.x.
- Maitra P, Zheng Y, Chen L, Wang YL, Ji NN, Lü PP, Gan HY, Li XC, Sun X, Zhou XH, Guo LD. 2019. Effect of drought and season on arbuscular mycorrhizal fungi in a subtropical secondary forest. *Fungal Ecology* 41: 107–115 DOI: 10.1016/j.funeco.2019.04.005.
- Martínez-García LB, Armas C, de Dios Miranda J, Padilla FM, Pugnaire FI. 2011. Shrubs influence arbuscular mycorrhizal fungi communities in a semi-arid environment. *Soil Biology and Biochemistry* 43: 682–689 DOI: 10.1016/j.soilbio.2010.12.006.

- McNaughton SJ, Banyikwa FF, McNaughton MM. 1997. Promotion of the cycling of diet-enhancing nutrients by African grazers. *Science* 278: 1798–1800 DOI: 10.1126/science.278.5344.1798.
- Meng L, Roulet N, Zhuang QL, Christensen TR, Froking S. 2016. Focus on the impact of climate change on wetland ecosystems and carbon dynamics. *Environmental Research Letters* 11: 100201 DOI: 10.1088/1748-9326/11/10/100201.
- Miller RM, Jastrow JD, Reinhardt DR. 1995. External hyphal production of vesicular-arbuscular mycorrhizal fungi in pasture and tallgrass prairie communities. *Oecologia* 103: 17–23 DOI: 10.1007/BF00328420.
- Murray TR, Frank DA, Gehring CA. 2010. Ungulate and topographic control of arbuscular mycorrhizal fungal spore community composition in a temperate grassland. *Ecology* 91: 815–827 DOI: 10.1890/09-0209.1.
- Oksanen J, Blanchet FG, Kindt R, Legendre P, Minchin PR, O'Hara RB, Simpson GL, Solymos P, Stevens MHH, Wagner H. 2013. Vegan: community ecology package. R package version 2.0-7 Available at <http://CRAN.R-project.org/package=vegan>.
- Öpik M, Vanatoa A, Vanatoa E, Moora M, Davison J, Kalwij JM, Reier Ü, Zobel M. 2010. The online database MaarjAM reveals global and ecosystemic distribution patterns in arbuscular mycorrhizal fungi (Glomeromycota). *New Phytologist* 188: 223–241 DOI: 10.1111/j.1469-8137.2010.03334.x.
- Pohlert T. 2014. The pairwise multiple comparison of mean ranks package (PMCMR). R package version 4.1. Available at <http://cran.r-project.org/package=PMCMR>.
- Pringle A, Bever JD. 2002. Divergent phenologies may facilitate the coexistence of arbuscular mycorrhizal fungi in a north Carolina grassland. *American Journal of Botany* 89:

- 1439–1446 DOI: 10.3732/ajb.89.9.1439.
- R Development Core Team. 2017. R: A language and environment for statistical computing. Vienna, Austria: R Foundation for Statistical Computing. Available at: <http://www.R-project.org>.
- Ren HY, Gui WY, Bai YF, Stein C, Rodrigues JLM, Wilson GWT, Cobb AB, Zhang YJ, Yang GW. 2018. Long-term effects of grazing and topography on extra-radical hyphae of arbuscular mycorrhizal fungi in semi-arid grasslands. *Mycorrhiza* 28: 117–127 DOI: 10.1007/s00572-017-0812-x.
- Rillig MC, Field CB, Allen MF. 1999. Soil biota responses to long-term atmospheric CO₂ enrichment in two California annual grasslands. *Oecologia* 119: 572–577 DOI: 10.1007/s004420050821.
- Rillig MC, Wright S, Nichols K, Schmidt W, Torn M. 2001. Large contribution of arbuscular mycorrhizal fungi to soil carbon pools in tropical forest soils. *Plant and Soil* 233: 167–177 DOI: 10.1023/A:1010364221169.
- Saks Ü, Davison J, Öpik M, Vasar M, Moora M, Zobel M. 2014. Root-colonizing and soil-born communities of arbuscular mycorrhizal fungi in a temperate forest understory. *Botany* 92: 277–285 DOI: 10.1139/cjb-2013-0058.
- Sato K, Suyama Y, Saito M, Sugawara K. 2005. A new primer for discrimination of arbuscular mycorrhizal fungi with polymerase chain reaction-denature gradient gel electrophoresis. *Grassland Science* 51: 179–181 DOI: 10.1111/j.1744-697X.2005.00023.x.
- Schloss PD, Westcott SL, Ryabin T, Hall JR, Hartmann M, Hollister EB, Lesniewski RA, Oakley BB, Parks DH, Robinson CJ, Sahl JW, Stres B, Thallinger GG, Van Horn DJ, Weber CF. 2009. Introducing mothur: open-source, platform-independent, community-

- supported software for describing and comparing microbial communities. *Applied and Environmental Microbiology* 75: 7537–7541 DOI: 10.1128/AEM.01541-09.
- Schwarzott D, Schüßler A. 2001. A simple and reliable method for SSU rRNA gene DNA extraction, amplification, and cloning from single AM fungal spores. *Mycorrhiza* 10: 203–207 DOI: 10.1007/PL00009996.
- Sharifi M, Ghorbanli M, Ebrahimzadeh H. 2007. Improved growth of salinity-stressed soybean after inoculation with salt pre-treated mycorrhizal fungi. *Journal of Plant Physiology* 164: 1144–1151 DOI: 10.1016/j.jplph.2006.06.016.
- Shi GX, Yao BQ, Liu YJ, Jiang SJ, Wang WY, Pan JB, Zhao XQ, Feng HY, Zhou HK. 2017. The phylogenetic structure of AMF communities shifts in response to gradient warming with and without winter grazing on the Qinghai-Tibet plateau. *Applied Soil Ecology* 121: 31–40 DOI: 10.1016/j.apsoil.2017.09.010.
- Simon L, Lalonde M, Bruns TD. 1992. Specific amplification of 18S fungal ribosomal genes from vesicular-arbuscular endomycorrhizal fungi colonizing roots. *Applied and Environmental Microbiology* 58: 291–295 Available at <https://aem.asm.org/content/58/1/291>.
- Smith SE, Read DJ. 2008. Mycorrhizal Symbiosis, third ed. Academic Press, New York; USA.
- Soka GE, Ritchie ME. 2018. Arbuscular mycorrhizal spore composition and diversity associated with different land uses in a tropical savanna landscape, Tanzania. *Applied Soil Ecology* 125: 222–232 DOI: 10.1016/j.apsoil.2018.01.013.
- Su YY, Guo LD. 2007. Arbuscular mycorrhizal fungi in non-grazed, restored and over-grazed grassland in the Inner Mongolia steppe. *Mycorrhiza* 17: 689–693 DOI: 10.1007/s00572-007-0151-4.

- 577 Tennant D. 1975. A test of a modified line intersect method of estimating root length. *Journal of*
578 *Ecology* 995–1001 DOI: 10.2307/2258617.
- 579 van der Heijden MGA, Bardgett RD, van Straalen NM. 2008. The unseen majority: soil microbes
580 as drivers of plant diversity and productivity in terrestrial ecosystems. *Ecology Letters* 11:
581 296–310 DOI: 10.1111/j.1461-0248.2007. 01139.x.
- 582 van der Heyde M, Abbott LK, Gehring C, Kokkoris V, Hart MM. 2019. Reconciling disparate
583 responses to grazing in the arbuscular mycorrhizal symbiosis. *Rhizosphere* 11: 100167 DOI:
584 10.1016/j.rhisph.2019.100167.
- 585 van der Heyde M, Bennett JA, Pither J, Hart M. 2017. Longterm effects of grazing on arbuscular
586 mycorrhizal fungi. *Agriculture, Ecosystems & Environment* 243: 27–33 DOI:
587 10.1016/j.agee.2017.04.003.
- 588 Varela-Cervero S, Vasar M, Davison JM, Öpik M, Azcón-Aguilar C. 2015. The composition of
589 arbuscular mycorrhizal fungal communities differ among the roots, spores and extraradical
590 mycelia associated with five Mediterranean plant species. *Environmental Microbiology* 17:
591 2882–2895 DOI: 10.1111/1462-2920. 12810.
- 592 Wang H, Yu LF, Zhang ZH, Liu W, Chen LT, Cao GM, Yue HW, Zhou JZ, Yang YF, Tang YH,
593 He JS. 2017. Molecular mechanisms of water table lowering and nitrogen deposition in
594 affecting greenhouse gas emissions from a Tibetan alpine wetland. *Global Change Biology*
595 23: 815–829 DOI: 10.1111/gcb.13467.
- 596 Wang Q, Bao WK, Yan ZL. 2002. Basic types and characters of the western Zoige meadows and
597 their changes in recent decades. *Chinese Journal of Applied and Environmental Biology* 8:
598 133–141 (in Chinese).
- 599 Wang SP, Duan JC, Xu GP, Wang YF, Zhang ZH, Rui YC, Luo CY, Xu B, Zhu XX, Chang XF,

- Cui XY, Niu HS, Zhao XQ, Wang WY. 2012. Effects of warming and grazing on soil N availability, species composition, and ANPP in an alpine meadow. *Ecology* 93: 2365–2376 DOI: 10.1890/11-1408.1.
- Wolf J, Johnson NC, Rowland DL, Reich PB. 2003. Elevated CO₂ and plant species richness impact arbuscular mycorrhizal fungal spore communities. *New Phytologist* 157: 579–588 DOI: 10.1046/j.1469-8137.2003.00696.x.
- Wright SF, Upadhyaya A. 1998. A survey of soils for aggregate stability and glomalin, a glycoprotein produced by hyphae of arbuscular mycorrhizal fungi. *Plant and Soil* 198: 97–107 DOI: 10.1023/A:1004347701584.
- Xiang D, Verbruggen E, Hu YJ, Veresoglou SD, Rillig MC, Zhou WP, Xu TL, Li H, Hao ZP, Chen YL, Chen BD. 2014. Land use influences arbuscular mycorrhizal fungal communities in the farming-pastoral ecotone of northern China. *New Phytologist* 204: 968–978 DOI: 10.1111/nph.12961.
- Xiang S, Guo RQ, Wu N, Sun SC. 2009. Current status and future prospects of Zoige marsh in eastern Qinghai-Tibet plateau. *Ecological Engineering* 35, 553–562 DOI: 10.1016/j.ecoleng.2008.02.016.
- Yan CG, Zhang MX. 2005. Wetland vegetation in China and its conservation advices. *Wetland Science* 03: 210–215 (in Chinese).
- Yang F, Niu KC, Collins CG, Yan XB, Ji YG, Ling N, Zhou XH, Du GZ, Guo H, Hu SJ. 2019. Grazing practices affect the soil microbial community composition in a Tibetan alpine meadow. *Land Degradation & Development* 30: 49–59 DOI: 10.1002/ldr.3189.
- Yang W, Gu SY, Xin Y, Bello A, Sun WP, Xu XH. 2018. Compost addition enhanced hyphal growth and sporulation of arbuscular mycorrhizal fungi without affecting their community

composition in the soil. *Frontiers in Microbiology* 9: 169 DOI: 10.3389/fmicb.2018.00169.

Yang W, Zheng Y, Gao C, He XH, Ding Q, Kim YC, Rui YC, Wang SP, Guo LD. 2013a. The arbuscular mycorrhizal fungal community response to warming and grazing differs between soil and roots on the Qinghai-Tibetan plateau. *PLoS ONE* 8: e76447 DOI: 10.1371/journal.pone.0076447.

Yang YF, Wu LW, Lin QY, Yuan MT, Xu DP, Yu H, Hu YG, Duan JC, Li XZ, He ZL, Xue K, van Nostrand J, Wang SP, Zhou JZ. 2013b. Responses of the functional structure of soil microbial community to livestock grazing in the Tibetan alpine grassland. *Global Change Biology* 19: 637–648 DOI: 10.1111/gcb.12065.

Zhang GS, Tian JQ, Jiang N, Guo XP, Wang YF, Dong XZ. 2008. Methanogen community in Zoige wetland of Tibetan plateau and phenotypic characterization of a dominant uncultured methanogen cluster ZC-I. *Environmental Microbiology* 10: 1850–1860 DOI: 10.1111/j.1462-2920.2008.01606.x.

Zhang J, Wang F, Che RX, Wang P, Liu HK, Ji BM, Cui XY. 2016. Precipitation shapes communities of arbuscular mycorrhizal fungi in Tibetan alpine steppe. *Scientific Reports* 6: 23488 DOI: 10.1038/srep23488.

Zheng Y, Kim YC, Tian XF, Chen L, Yang W, Gao C, Song MH, Xu XL, Guo LD. 2014. Differential responses of arbuscular mycorrhizal fungi to nitrogen addition in a near pristine Tibetan alpine meadow. *FEMS Microbiology Ecology* 89: 594–605 DOI: 10.1111/1574-6941.12361.

Zhou GY, Zhou XH, He YH, Shao JJ, Hu ZH, Liu RQ, Zhou HM, Hosseinibai S. 2017. Grazing intensity significantly affects belowground carbon and nitrogen cycling in grassland ecosystems: a meta-analysis. *Global Change Biology* 23: 1167–1179 DOI:

646 10.1111/gcb.13431.

Figure legend

Fig. 1. Easily extracted glomalin-related soil protein (EE-GRSP, a), total extracted GRSP (T-GRSP, b), spore density (c) and extra radical hyphal (ERH) density (d) of arbuscular mycorrhizal (AM) fungi in grazing and non-grazing treatments. One-way ANOVA showed the effect of grazing on AM fungal variables. Data are means \pm SE (n = 20). Bars with different letters denote significant difference in grazing and non-grazing treatments according to Tukey's HSD test at $P < 0.05$.

Fig. 2. The operational taxonomic unit (OTU) richness of arbuscular mycorrhizal (AM) fungi in soil and roots in grazing and non-grazing treatments. Kruskal-Wallis test showed the effect of grazing and sample type (soil and root) on the OTU richness. Data are means \pm SE (n = 20). Bars with different letters denote significant difference in grazing and non-grazing treatments according to Conover's test at $P < 0.05$.

Fig. 3. Relative abundance of arbuscular mycorrhizal (AM) fungal operational taxonomic units (OTUs) in soil and roots in grazing and non-grazing treatments. Two-way ANOVA and Kruskal-Wallis tests showed the effect of grazing and sample type (soil and root) on the relative abundance of AM fungal OTUs (ns; $P \geq 0.05$, * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$). The rare AM fungal OTUs ($< 1\%$ of total AM fungal reads) and abundant AM fungal OTUs ($> 1\%$ of total AM fungal reads) that was not significantly affected by grazing and sample type were all assigned to "Others". SN, soil non-grazing; SG, soil grazing; RN, root non-grazing; RG, root grazing.

670

671 **Fig. 4.** Relative abundance of arbuscular mycorrhizal (AM) fungi at the order level in soil and
 672 roots in grazing and non-grazing treatments. Kruskal-Wallis test showed the effect of grazing
 673 and sample type (soil and root) on the relative abundance of AM fungal orders (ns; $P \geq 0.05$, ***
 674 $P < 0.001$). Different letters are significantly different at $P < 0.05$, as indicated by Conover's test.
 675 SN, soil non-grazing; SG, soil grazing; RN, root non-grazing; RG, root grazing.

676

677 **Fig. 5.** Redundancy analysis (RDA) biplots showing arbuscular mycorrhizal (AM) fungal
 678 community composition in soil and roots (a), soil (b) and roots (c). Significant soil variables
 679 were presented as vectors on the RDA biplot graphs using the 'envfit' (based on 999
 680 permutations) at $P < 0.05$. SN, soil non-grazing; SG, soil grazing; RN, root non-grazing; RG,
 681 root grazing; N, soil total nitrogen; C, soil total carbon; P, soil total phosphorus.

Figure 1

Grazing significantly affected AM fungal biomass.

Fig. 1. Easily extracted glomalin-related soil protein (EE-GRSP, a), total extracted GRSP (T-GRSP, b), spore density (c) and extra radical hyphal (ERH) density (d) of arbuscular mycorrhizal (AM) fungi in grazing and non-grazing treatments. One-way ANOVA showed the effect of grazing on AM fungal variables. Data are means \pm SE (n = 20). Bars with different letters denote significant difference in grazing and non-grazing treatments according to Tukey's HSD test at $P < 0.05$.

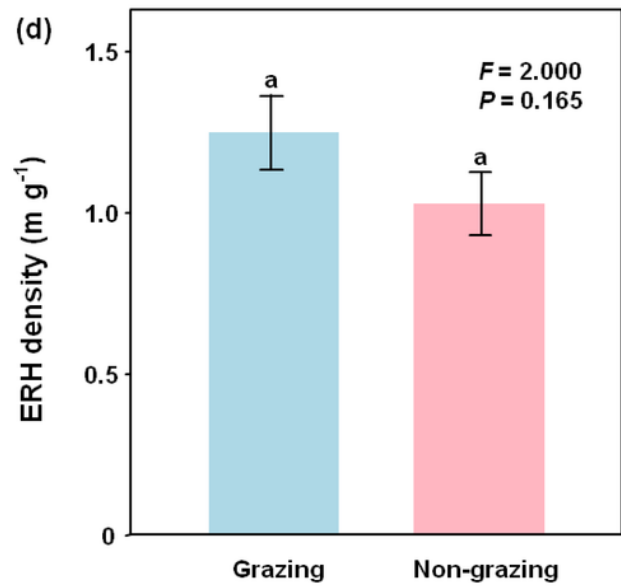
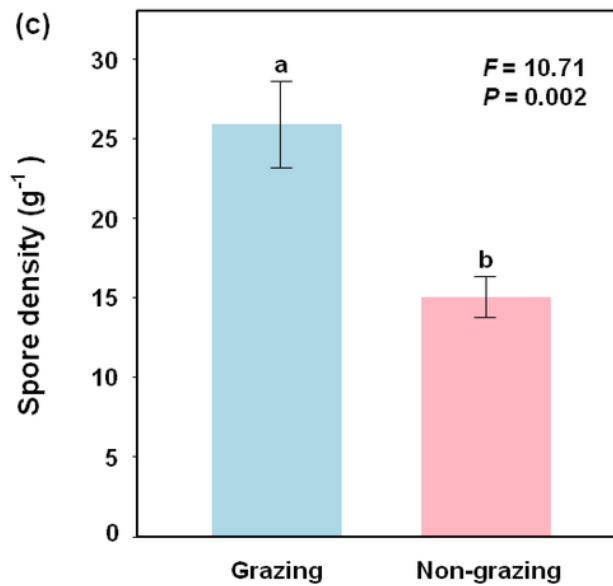
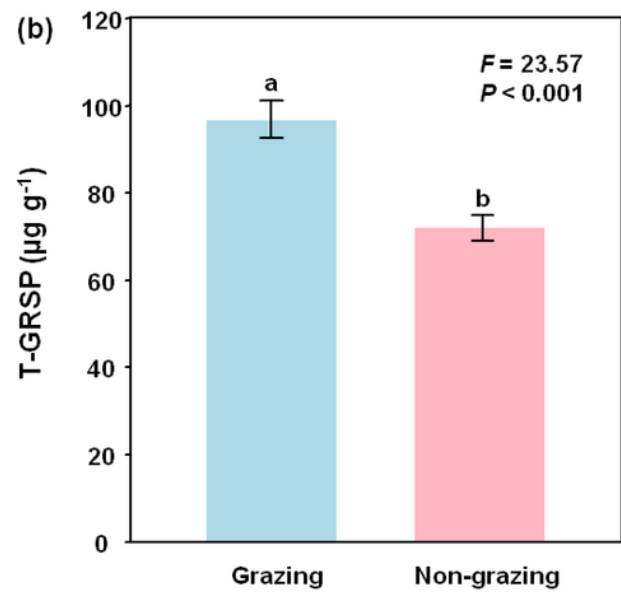
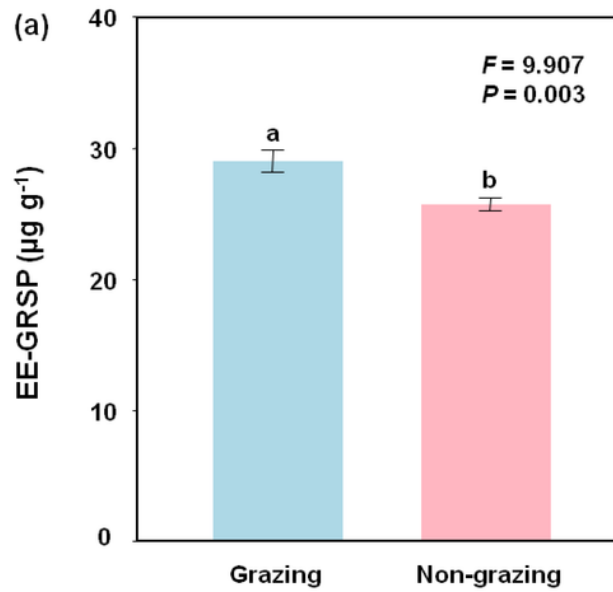


Figure 2

Grazing did not significantly affected AM fungal richness.

Fig. 2. The operational taxonomic unit (OTU) richness of arbuscular mycorrhizal (AM) fungi in soil and roots in grazing and non-grazing treatments. Kruskal-Wallis test showed the effect of grazing and sample type (soil and root) on the OTU richness. Data are means \pm SE (n = 20). Bars with different letters denote significant difference in grazing and non-grazing treatments according to Conover's test at $P < 0.05$.

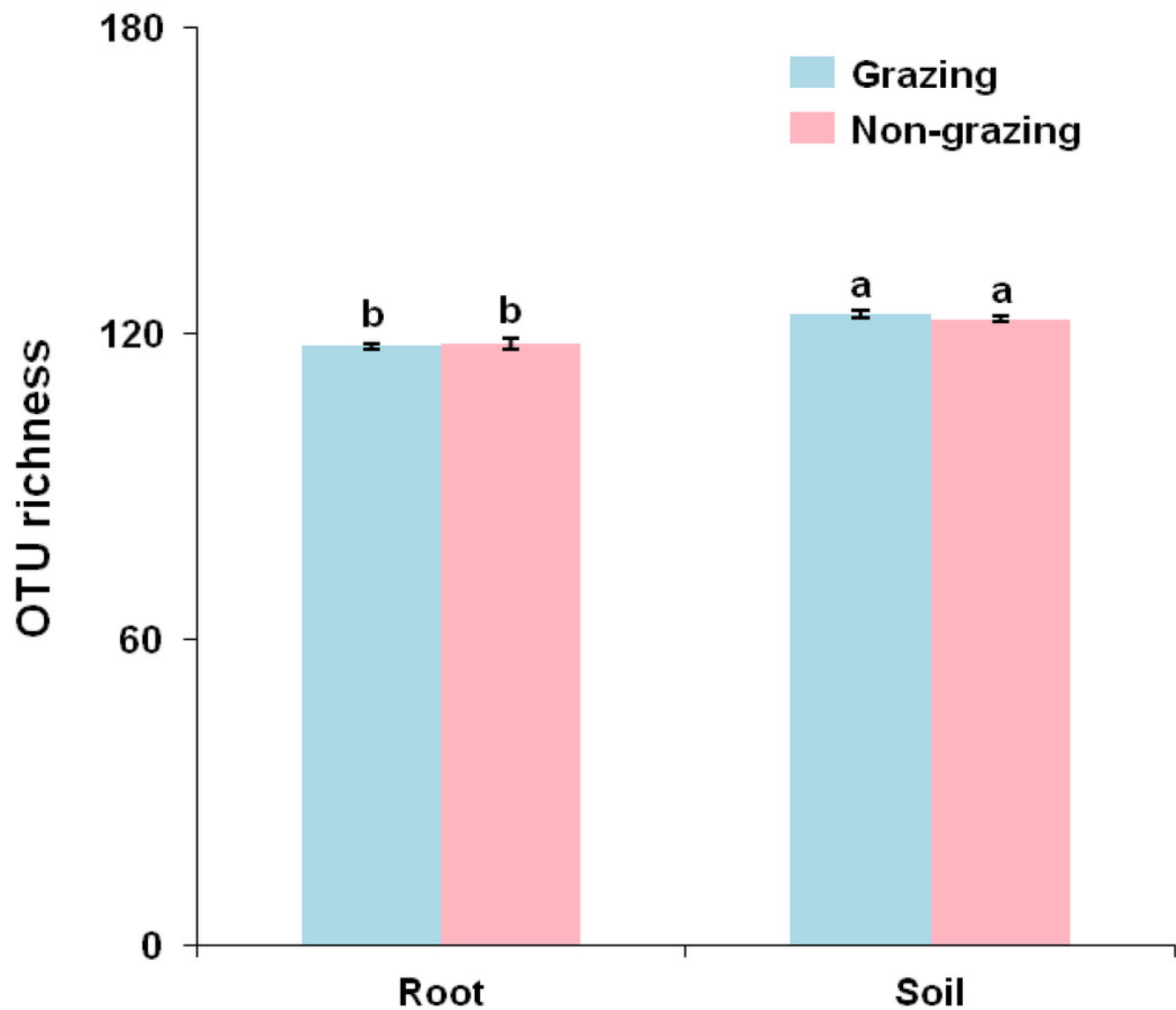


Figure 3

Relative abundance of arbuscular mycorrhizal (AM) fungal operational taxonomic units (OTUs) in soil and roots in grazing and non-grazing treatments.

Fig. 3. Relative abundance of arbuscular mycorrhizal (AM) fungal operational taxonomic units (OTUs) in soil and roots in grazing and non-grazing treatments. Two-way ANOVA and Kruskal-Wallis tests showed the effect of grazing and sample type (soil and root) on the relative abundance of AM fungal OTUs (ns; $P \geq 0.05$, * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$). The rare AM fungal OTUs (< 1% of total AM fungal reads) and abundant AM fungal OTUs (> 1% of total AM fungal reads) that was not significantly affected by grazing and sample type were all assigned to “Others”. SN, soil non-grazing; SG, soil grazing; RN, root non-grazing; RG, root grazing.

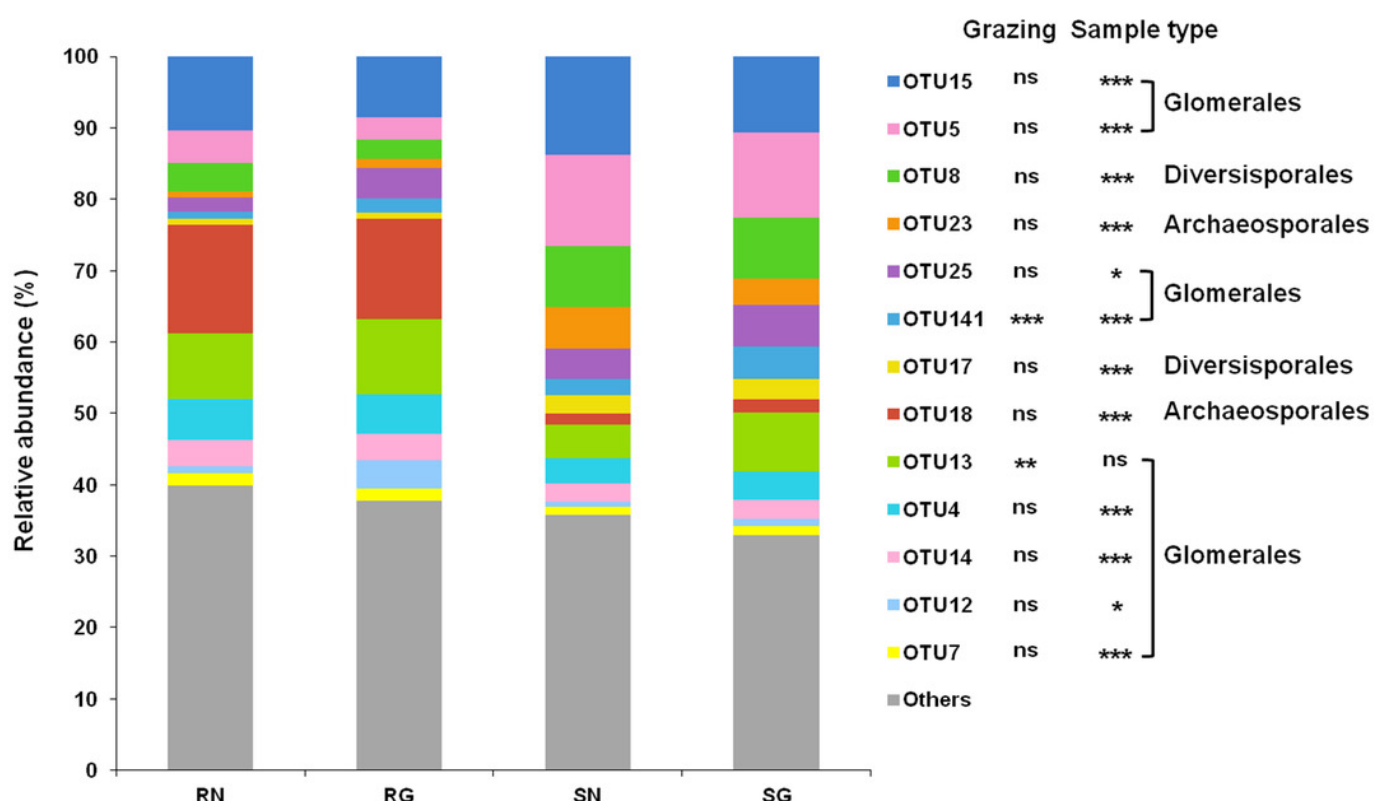


Figure 4

Relative abundance of arbuscular mycorrhizal (AM) fungi at the order level in soil and roots in grazing and non-grazing treatments.

Fig. 4. Relative abundance of arbuscular mycorrhizal (AM) fungi at the order level in soil and roots in grazing and non-grazing treatments. Kruskal-Wallis test showed the effect of grazing and sample type (soil and root) on the relative abundance of AM fungal orders (ns; $P \geq 0.05$, *** $P < 0.001$). Different letters are significantly different at $P < 0.05$, as indicated by Conover's test. SN, soil non-grazing; SG, soil grazing; RN, root non-grazing; RG, root grazing.

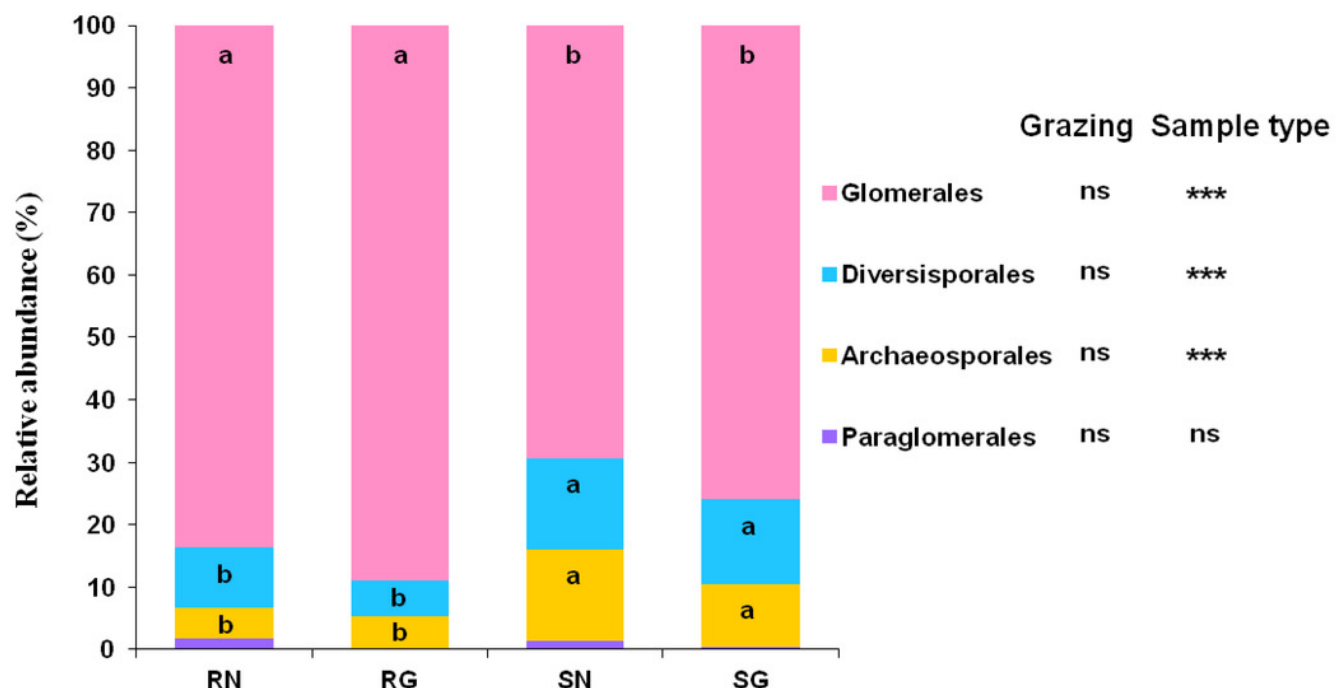


Figure 5

Redundancy analysis (RDA) biplots showing arbuscular mycorrhizal (AM) fungal community composition.

Fig. 5. Redundancy analysis (RDA) biplots showing arbuscular mycorrhizal (AM) fungal community composition in soil and roots (a), soil (b) and roots (c). Significant soil variables were presented as vectors on the RDA biplot graphs using the 'envfit' (based on 999 permutations) at $P < 0.05$. SN, soil non-grazing; SG, soil grazing; RN, root non-grazing; RG, root grazing; N, soil total nitrogen; C, soil total carbon; P, soil total phosphorus.

