

Gene flow between subpopulations of gray snapper (*Lutjanus griseus*) from the Caribbean and Gulf of Mexico

Oscar de Jesús Rosado-Nic¹, J. Derek Hogan², José Héctor Lara-Arenas¹, Rigoberto Rosas-Luis³, Laura Carrillo⁴, Carmen Amelia Villegas-Sánchez^{Corresp. 1}

¹ División de Estudios de Posgrado e Investigación, Tecnológico Nacional de México / Instituto Tecnológico de Chetumal, Chetumal, Quintana Roo, México

² Department of Life Sciences, Texas A&M University - Corpus Christi, Corpus Christi, Texas, United States

³ Conacyt-División de Estudios de Posgrado e Investigación, Tecnológico Nacional de México / Instituto Tecnológico de Chetumal, Chetumal, Quintana Roo, México

⁴ Departamento de Sistemática y Ecología Acuática, El Colegio de la Frontera Sur, Chetumal, Quintana Roo, México

Corresponding Author: Carmen Amelia Villegas-Sánchez
Email address: cvillegas@itchetumal.edu.mx

Background. The gray snapper (*Lutjanus griseus*) has a tropical and subtropical distribution. In much of its range this species represents one of the most important fishery resources because of its high quality meat and market value. Due to this, the species is vulnerable to overfishing, and population declines have been observed in parts of its range. In recent decades, it has been established that knowing the level of genetic connectivity is useful for establishing appropriate management and conservation strategies given that genetic isolation can drive towards genetic loss. Presently the level of genetic connectivity between subpopulations of *L. griseus* of the southern region of the Gulf of Mexico and the Caribbean Sea remains unknown. **Methods.** In the present study we analyse genetic structure and diversity for seven subpopulations in the southern Gulf of Mexico and the Mexican Caribbean Sea. Eight microsatellite primers of phylogenetically closely related species to *L. griseus* were selected. **Results.** Total heterozygosity was 0.628 and 0.647 in the southern Gulf of Mexico and the Mexican Caribbean Sea, however, results obtained from AMOVA and R_{ST} indicated a lack of genetic difference between the major basins. We also found no association between genetic difference and geographic distance, and moderately high migration rates ($N_m = > 4.1$) suggesting ongoing gene flow among the subpopulations. Gene flow within the southern Gulf of Mexico appears to be stronger going from east-to-west. **Conclusions.** Migration rates tended to be higher between subpopulations within the same basin compared to those across basins indicating some regionalization. High levels of genetic diversity and genetic flow suggest that the population is quite large; apparently, the fishing pressure has not caused a bottleneck effect.

Gene flow between subpopulations of gray snapper (*Lutjanus griseus*) from the Caribbean and Gulf of Mexico

Oscar de Jesús Rosado-Nic¹, J. Derek Hogan², José Héctor Lara-Arenas³, Rigoberto Rosas-Luis⁴, Laura Carrillo⁵, Carmen Amelia Villegas-Sánchez⁶

¹ División de Estudios de Posgrado e Investigación, Tecnológico Nacional de México / Instituto Tecnológico de Chetumal, Chetumal, Quintana Roo, México.

² Department of Life Sciences, Texas A&M University-Corpus Christi, Corpus Christi, Texas, USA.

³ División de Estudios de Posgrado e Investigación, Tecnológico Nacional de México / Instituto Tecnológico de Chetumal, Chetumal, Quintana Roo, México.

⁴ Conacyt-División de Estudios de Posgrado e Investigación, Tecnológico Nacional de México / Instituto Tecnológico de Chetumal, Chetumal, Quintana Roo, México.

⁵ El Colegio de la Frontera Sur, Unidad Chetumal, Departamento de Sistemática y Ecología Acuática, Chetumal, Quintana Roo, México.

⁶ División de Estudios de Posgrado e Investigación, Tecnológico Nacional de México / Instituto Tecnológico de Chetumal, Chetumal, Quintana Roo, México.

Corresponding Author:

Carmen Amelia Villegas-Sánchez⁶

Av Insurgentes, Chetumal, Quintana Roo, 77013, México

Email address: cavs005@gmail.com

Abstract

Background

The gray snapper (*Lutjanus griseus*) has a tropical and subtropical distribution. In much of its range this species represents one of the most important fishery resources because of its high quality meat and market value. Due to this, this species is vulnerable to overfishing, and population declines have been observed in parts of its range. In recent decades, it has been established that knowing the level of genetic connectivity is useful for establishing appropriate management and conservation strategies given that genetic isolation can drive towards genetic loss. Presently the level of genetic connectivity between subpopulations of *L. griseus* of the southern region of the Gulf of Mexico and the Caribbean Sea remains unknown.

Methods

In the present study we analyse genetic structure and diversity for seven subpopulations in the southern Gulf of Mexico and the Mexican Caribbean Sea. Eight microsatellite primers of phylogenetically closely related species to *L. griseus* were selected.

Results

Total heterozygosity was 0.628 and 0.647 in the southern Gulf of Mexico and the Mexican Caribbean Sea, however, results obtained from AMOVA and R_{ST} indicated a lack of genetic difference between the major basins. We also found no association between genetic difference and geographic distance, and moderately high migration rates ($N_m = > 4.1$) suggesting ongoing gene flow among the subpopulations. Gene flow within the southern Gulf of Mexico appears to be stronger going from east-to-west.

Conclusions

Migration rates tended to be higher between subpopulations within the same basin compared to those across basins indicating some regionalization. High levels of genetic diversity and genetic flow suggest that the population is quite large; apparently, the fishing pressure has not caused a bottleneck effect.

Introduction

Establishing effective fisheries regulations is a complex and multidisciplinary task (INAPESCA, 2012). In recent decades, it has been determined that delineating stock boundaries and knowing the level of genetic connectivity among stocks is useful for establishing appropriate management and conservation strategies (Villegas Sánchez et al., 2014). Improper management can lead to genetic diversity loss and increased inbreeding within genetically isolated populations with negative effects for the survival of the populations (Urbiola-Rangel & Chassin-Noria, 2013; Villegas Sánchez et al., 2014). With the use of genetic markers, diversity and the level of genetic connectivity can be estimated between populations at different geographic scales.

The Gray snapper (*Lutjanus griseus*) is an economic and ecologically important fishery species and can be highly abundant throughout its range yet there is a paucity of information for its fishery management (Lindeman et al., 2016). Gray snapper fisheries are subjected to fishery regulations in some countries, for example: in the Everglades National Park, United States, a catch limit of 10 individuals per person was established in the 1970's (Claro & Lindeman, 2008). In Cuba the fishery is closed in June, during the breeding season (Claro & Lindeman, 2008). Nevertheless, in Mexico, although it's being captured by fishers, it has not been classified as

overexploited or subjected to fishery overexploitation, thus no closed season or other regulation has been established (INAPESCA, 2018); this lack of regulation could place populations at risk in the future (Costello et al., 2012).

The gray snapper presents a wide distribution from North Carolina, United States to southern Brazil, in the countries where it is found it represents an important reef fishery resource because of its high quality meat and market value (Claro & Lindeman, 2008). The gray snapper is a predator that feeds on a wide variety of organisms in different habitats including estuaries, mangroves and seagrass beds, whereof changes in their populations have great impacts in other elements of the community (Claro & Lindeman, 2008; Rocha & Molina, 2008). Migratory movements of adult individuals are mainly local and sexual maturity occurs when they are one year old reaching a total length between 260 and 280 mm. The pelagic larval period of this species is 25 – 26 days and larval settlement has been recorded between 30 to 40 days after hatching (Claro & Lindeman, 2008). Its large geographic range and month-long larval duration indicate that connectivity among populations may be widespread.

Despite its importance as a fisheries species, there is relatively little knowledge about the stock structure of gray snapper, however stocks appear to be declining in some places. Commercial landings in the US South Atlantic region (North Carolina to Florida) have been declining since the 1950's and populations in the Florida Keys are thought to be potentially over-fished (Ault, Bohnsack & Meester, 1998). In Cuba, gray snapper is abundant; however, fisheries landings have declined over the past 30 years as have spawning aggregations (Claro et al., 2009). In Mexico, there is currently no information about stock structure and gray snappers are caught in a mixed stock fishery (FAO, 2011; SAGARPA, 2012), the health of which is not well known. There are few population genetics studies for the gray snapper and only one study related to genetic connectivity exists (Gold et al., 2009). The authors looked at genetic structure at a scale of 2,400 km in the north of the Gulf of Mexico and the U.S. Atlantic using microsatellites, reporting the existence of genetically different populations and a decrease in their effective size from east to west presumably driven by ocean currents. They recommended that these populations be treated as independent stocks for effective management of the fishery. However, there have been no studies of genetic structure or connectivity of gray snapper in the southern Gulf of Mexico or Mexican Caribbean.

In Mexico, we hypothesize that there may be genetic differences between the populations of gray snapper in the Mexican Caribbean and the southern Gulf of Mexico. Prior studies have shown that there is genetic differentiation between these major basins in other taxa. Blacktip sharks (*Carcharhinus limbatus*), a low dispersal species, show strong genetic differentiation between the Mesoamerican Barrier Reef System and the southern Gulf of Mexico (Keeney et al., 2005). The bicolor damselfish (*Stegastes partitus*), a high dispersal reef fish (Hogan et al., 2012), has shown evidence of a weak restriction in gene flow between the Mexican Caribbean and southern Gulf of Mexico (Villegas Sánchez et al., 2014). Similarly, the lionfish (*Pterois volitans*), the most studied invasive species, has been reported as having significant genetic differentiation between both regions, which suggests a phylogeographic break (Labastida-Estrada et al., 2019). The objective of this study is to determine the diversity and genetic connectivity among seven subpopulations of gray snapper in the southern Gulf of Mexico and Mexican Caribbean Sea.

Materials & Methods

Study area. Two regions were studied: 1) Gulf of Mexico, with the subpopulations Campeche (C), Puerto de Veracruz (PV), and Tuxpan (TX); 2) Mexican Caribbean Sea, with the

subpopulations Bahia de Chetumal (*BC*), Xahuayxol (*X*), Punta Herrero (*PH*) and Chiquilá (*CH*) (Fig. 1). Our seven sampling sites are distributed along approximately 1,950 km of coast between the Caribbean Sea and southern Gulf of Mexico. The Gulf of Mexico region possesses an extensive continental shelf with a diversity of ecosystems like wetlands, the largest area of mangroves in Mexico, coastal dunes and coral reefs with a tropical and subtropical climate (Lara-Lara et al., 2008). The Mexican Caribbean region with a narrow continental shelf is located in the Southeast of Mexico, with a warm subhumid climate and an annual average temperature of 26°C and a mean annual precipitation of 1,300 mm (Lara-Lara et al., 2008).

Fish Sampling and sample storage. Muscular tissue was taken from the base of the caudal fin of 348 organisms captured by fishermen (50 samples per site, except for the *X* site, with 48 samples; SEMARNAT approved the field study with the number: 23/K4-0002/05/18). The sampling was carried out during November and December 2016 for the Mexican Caribbean Sea region, and during the same months 2017 for the Southern Gulf of Mexico region. Such periods were chosen in order to only catch resident individuals so as not to bias the estimates of gene flow, given that the migration of the gray snapper occurs from May to October (Claro & Lindeman, 2008; Espinoza Ávalos, 2009). All samples were preserved in 95% ethanol and stored at -5° C for transportation and laboratory storage prior to DNA extraction.

DNA Isolation. DNA was isolated with the Qiagen DNeasy Blood & Tissue Kit following manufacturer's protocols. DNA concentration and quality was verified using a spectrophotometer (Thermo Scientific NanoDrop™ 2000).

Molecular markers. Microsatellites are commonly used DNA markers in genetic diversity studies because they present greater number of polymorphs (allelic variation) per locus than other markers. This feature makes them more sensitive to changes in size, structure and dispersion rate of populations (Goldstein & Schlötterer, 1999). Additionally, microsatellites are co-dominant markers which allow to distinguish between homozygous and heterozygous individuals (Estoup et al., 1998; Goldstein & Schlötterer, 1999). Eight microsatellite primers were selected from published literature of phylogenetically closely related species to *L. griseus*, which were selected for their high levels of polymorphism and their ability to amplify in *L. griseus*. Seven loci were developed from the red snapper (*Lutjanus campechanus*) while one locus (Ra1) was developed from the vermillion snapper (*Rhombloplites aurorubens*) (Gold, Pak & Richardson, 2001; Renshaw et al., 2007).

Genotyping. Genomic DNA was amplified by polymerase chain reactions carried out in a total volume of 10 µl, the solution contained 5.40 µl of H₂O, 2.0 µl of green buffer, 0.8 µl of MgCl₂, 0.125 µl dNTP, 0.25 µl of primer forward, 0.30 µl of primer reverse, 0.04 of Taq-polymerase and 1 µl (5 ng) of DNA extract. The Eppendorf thermocycler was programmed with an initial denaturing cycle at 95° C for 2 minutes, followed by 45 cycles of a denaturing step at 95° C for 30 seconds, followed by an annealing step (temperature varied depending on the microsatellite; Table 1) for a period of 30 seconds, and a final elongation step at 72° C for 40 seconds.

Allele sizes were estimated using a DNA fragment analyser (ABI 3730xl DNA). We used ABI DS-33 dye set with G5 filter set. Forward primers were dye labelled with either 6-FAM, VIC, NED, or PET dye labels; GS-600 standard set was used with LIZ dye. This protocol allowed the detection of several PCR products at the same time by fluorescence emission. Fragment sizes were estimated using GeneMarker® software (SoftGenetics).

Polymorphism analysis and genetic diversity. Using the Microsatellite Toolkit software (Park, 2008), the Polymorphic information content (PIC) was calculated, which is an indicator of the marker quality and the degree of polymorphism in genetic cartography studies. Values of PIC

higher than 0.5 indicate that the marker is highly informative, values from 0.25 to 0.5 are related to markers moderately informative and values below 0.25 indicate that the marker is slightly informative (Botstein et al., 1980). Genetic diversity was calculated as the number of alleles (N_a) and effective number of alleles (A_E); the latter defined as the alleles with the capacity of passing to the next generation (Kimura & Crow, 1964). Observed heterozygosity (H_o), expected heterozygosity (H_e) and total heterozygosity (H_t) were also calculated. Heterozygosity is a measure used to know the diversity of a locus and is defined as the probability that upon selecting two loci, both will be different (Cabrero & Camacho, 2002). These analyses were carried out with the GenAlex 6.5 software (Peakall & Smouse, 2012). The frequency of null alleles (F_a) was also estimated, considering that markers exceeding a 0.2 value should be excluded from further analyses (Dakin & Avise, 2004). This analysis was carried out using the MicroChecker 2.2 software (Van Oosterhout et al., 2004).

Hardy-Weinberg equilibrium (HWE) and linkage disequilibrium (LD). Wright's Fixation Index (F_{IS}) was used to test if the allelic frequencies conformed to the Hardy-Weinberg equilibrium (Cockerham & Weir, 1984). The F_{IS} and p values were calculated in the Arlequin 3.5 software (Excoffier & Lischer, 2010). LD between loci pairs was evaluated considering all individuals as a single subpopulation. LD indicates the association between loci pairs given that some alleles don't segregate in an independent manner. This was carried out using the Arlequin 3.5 software (Excoffier & Lischer, 2010) and the false discovery rate (Benjamini & Yekutieli, 2001) was applied to multiple tests of HWE and LD .

Genetic structure. Genetic differences between subpopulation pairs were evaluated using the Wright Index (F_{ST}) (Cockerham & Weir, 1984). Given that in the present study no subpopulation had a mean of H_e higher than 0.9, it was not necessary to standardize F_{ST} , as the range of this index tends to become very small when H_e is large (Meirmans & Hedrick, 2011). The R_{ST} index was estimated given that it is analogous to the F_{ST} index, nevertheless, this index is the one that best reflects the mutation pattern of the microsatellites (Slatkin, 1995). This index was estimated between subpopulation pairs. The R_{ST} were calculated using the Arlequin 3.5 software (Excoffier & Lischer, 2010).

The effective population size (N_e) and the number of effective migrants per generation (N_m) were estimated using the population parameter theta (Θ) and the migration parameter (M) was calculated between subpopulation pairs. The analyses were carried out using the Maximum Likelihood Estimation with ten short chains and one long chain with 50,000 and 500,000 genealogies respectively at a constant mutation rate using the Migrate 3.6 software (Beerli, 2016).

To measure the degree of genetic difference between subpopulations at different hierarchic levels (among regions, among subpopulations and among individuals and within individuals) an Analysis of Molecular Variance (AMOVA) was carried out using Arlequin 3.5 software (Excoffier & Lischer, 2010).

Isolation-by-distance. An association between genetic difference (F_{ST}) and the geographic distance can indicate restricted gene flow. We tested for this association with the Mantel test (Aguirre-Planter, 2007) using the GenAlex 6.5 software (Peakall & Smouse, 2012).

Results

Genetic diversity. A total of 73 different alleles were recorded. The markers that resulted with the highest N_a values were Prs260, Ra1, Lca107 and Prs137 (with 15, 13, 10 and 11 alleles, respectively). Based on the PIC results, six microsatellites were considered to be highly

informative given that they exceed the 0.5 value, locus Prs320 (0.469) was medium level and locus Lca43 (0.239) was poorly informative (Table 2). Mean number of effective alleles (A_E) varied among subpopulations, between 3.115 (TX) and 3.384 (X). Higher values were observed in the Mexican Caribbean Sea subpopulations. H_e mean values for the Gulf of Mexico varied between 0.626 and 0.631 (Table 1), with the highest mean values in subpopulations C and PV ($H_e = 0.631$ both). H_e mean values observed in the Mexican Caribbean Sea region ranged between 0.641 and 0.655 (Table 2), with the highest mean values in subpopulations X and PH (0.650 and 0.655, respectively). H_t values were similar for the Gulf of Mexico and the Mexican Caribbean Sea, with mean values of 0.628 and 0.647 respectively.

Genetic structure. Microsatellites that showed a significant deviation to the HWE in at least one subpopulation were Lca107, Prs137, Prs260 and Prs275 (Table 1, 2). In the LD results, no marker showed a significant LD after applying the false discovery rate from a total of 28 comparisons, meaning that the eight microsatellites were inherited in an independent manner. The mean frequency of null alleles for each subpopulation was between 0.045 and 0.104 (Table 1, 2) with no marker exceeding 0.2 of the null alleles. Given this, no marker or subpopulation was excluded from the diversity and genetic structure analyses. Pairwise values of F_{ST} ranged between 0.003 and 0.008 (Table 3), which indicates that the difference in the allelic frequencies between subpopulations are minimal. The highest values were between CH and both BC and TX (0.008). In subpopulation pairwise comparison, the F_{ST} and R_{ST} (Table 3) showed no significant difference. In the genetic flow estimation using N_m values a variation from 4.1 to 25.2 was observed. High levels of connectivity were observed in all sites (Table 3). N_e values varied from 1,922 and 3,799 individuals where the highest values were observed in X and PH with 3799 and 3686 respectively (Table 4). Per generation migration rates (M) ranged between 1.3 and 11.2. Highest values were observed in the pairs $CH-C$ (8.5), $C-PV$ (10.0), and $PV-TX$ (11.2) (Table 4). The highest variation registered by AMOVA was within individuals ($PV\% = 98.22$). The F_{IS} value was 0.017, which suggests that the subpopulations present in the studied regions constitute a panmictic population. The F_{CT} and F_{SC} values indicated no genetic structure (Table 5).

Isolation-by-distance. The coefficient of determination from the Mantel test was 0.0221 and the coefficient of correlations was 0.149 ($p = 0.517$), therefore, we found no relationship between the genetic and geographic distances.

Discussion

The main objective of this study was to evaluate stock structure, connectivity, and to estimate genetic diversity between seven subpopulations of gray snapper (*Lutjanus griseus*) in two regions of the Mexican Atlantic: the Gulf of Mexico and the Mexican Caribbean Sea. We found no significant genetic differences between basins or among subpopulations (AMOVA: among regions and among subpopulations). Similarly, estimations of F_{ST} and R_{ST} between subpopulation pairs were all not significant. This lack of genetic difference indicates high genetic connectivity between all subpopulations of *L. griseus* along this ~1950 km stretch of coastline. This result is also supported by the Mantel Test ($p > 0.05$), which didn't show an association between genetic differences and geographic distances. In all cases, the number of migrants per generation (N_m) entering any given subpopulation was greater than 4, this indicates that there is unrestricted gene flow and that the populations behave like a panmictic population, as theoretically with the migration of a single organism ($N_m = 1$) allele fixation is avoided (Slatkin, 1995). However, our

estimates of M appear to be greater within basins than among basins, and in the Gulf of Mexico region there appears to be a distinct directionality to migration, with greater M -values between neighbouring subpopulation pairs going west compared to east (Table 4; Table S1). This suggests that despite the fact that F_{ST} based estimates of gene flow show no regionalization, there may be some subtle patterns of reduced connectivity between basins and perhaps a distinct directionality of gene flow in the Gulf of Mexico.

The fact that we found that connectivity may be slightly restricted between the Mexican Caribbean and southern Gulf of Mexico is aligned with findings from previous studies of connectivity between the regions. Blacktip sharks (*Carcharhinus limbatus*) are known to show strong genetic differentiation between the Mesoamerican Barrier Reef System (Caribbean) and the southern Gulf of Mexico (Keeney et al., 2005). Additionally, bicolor damselfish (*Stegastes partitus*) also shows evidence of a weak restriction in gene flow between the Mexican Caribbean and southern Gulf of Mexico (Villegas Sánchez et al., 2014). Our finding here of generally greater migration rates among subpopulations within regions than across regions further supports these studies.

Ocean currents and the biology of *L. griseus* may play an important role in connectivity between the subpopulations. A particle tracking model of a closely related species *Lutjanus analis* virtual larvae carried out by Martinez et al., (2019) suggests that the marine protected areas of the Mesoamerican Reef network are all highly connected through ocean current. This species has a similar life cycle as *L. griseus*, therefore the results of this project support the findings by these authors. The Yucatan current (average speed of 1.5 m/s), which later becomes the Loop Current (Athié et al., 2011), can export fish larvae from the Caribbean up to the Gulf of Mexico (Carrillo et al., 2015). Possibly these ocean dynamics favour the dispersion of *L. griseus* larvae. In adults, migrations are present but these are mainly local, with registered distances between 35 and 122 km (Claro & Lindeman, 2008), and these movements are typically between inshore habitats and shelf habitats rather than among reefs. So it is unlikely that adult movements alone are resulting in high levels of gene flow observed in this study.

In terms of the apparent directionality of gene flow in the southern Gulf of Mexico, studies of oceanographic connectivity in the southern Gulf of Mexico indicate that east-gene flow from Campeche Bank to Veracruz and Tuxpan reefs is low and that connectivity is stronger going west to east (Sanvicente-Añorve et al., 2014). The pattern of migration observed in this study appears to contradict these previous findings (Fig. 2). However, ocean currents in this region are complex with the presence of eddies and a seasonal shifts in direction within the inner shelf with summer months showing flux from east to west and winter months showing flux from west to east (Salas-Monreal et al., 2017). The summer period of greater east to west connectivity also coincides generally with gray snapper spawning season (Domeier, Koenig & Coleman, 1996). The pattern of east-to-west migration observed here is a feasible explanation.

The highest migration rates were between *Ch-C* and *C-PV*. According to this result, Campeche can be an important point for the connectivity between the two regions. It is important to mention that this Mexican state has the largest extent of mangroves (Lara-Lara et al., 2008), and it has been shown that mangroves are the principal habitats for larvae and juveniles of *L. griseus*, even though they also use seagrass beds at times (Claro & Lindeman, 2008). For successful sea dispersion, it is essential that larvae find an adequate habitat for recruitment (Cowen, 2006). Thus, if the habitat is fragmented or destroyed, the connectivity can be limited (Jones, Srinivasan & Almany, 2007). We also found high migration rates between subpopulations that were apparently counter to the flow of main currents (e.g., *C-CH*, *CH-BC* and *PH-X*). This could be

due to displacing reproductive aggregations that occur between the months of June and August around the new moon on the shelf border with a duration period of 8 to 10 days (Claro & Lindeman, 2008). Moreover, it has been reported periods of a coastal countercurrent over the shelf (Carrillo et al., 2017) that could promote migrations as it was found in our results. The effective population sizes estimated from Migrate (θ) were large, however, they must be considered with some caution, especially when interpreting for management and conservation issues. The mutation rate that is considered for the N_e calculation can create a bias with different values with several orders of magnitude. It is suggested that the values of N_e should only be taken as comparative ones between sampled sites rather than a true point estimate (Beerli, 2016). Genetic diversity in these populations was high (H_t general average = 0.640; PIC general average = 0.597). According to the PIC, the microsatellites used are highly informative and polymorphic (Botstein et al., 1980). The number of alleles per microsatellite varied between 5 (Lca43) and 15 (Prs260), with a general mean of 6.17 ($n = 8$); previous studies of *L. griseus* have reported means of 5.43 ($n = 14$) (Renshaw et al., 2007) and 7.0 ($n = 14$) (Gold et al., 2009), similar to the values reported in this study. These subpopulations appear to be similar in diversity to those of the U.S. Gulf of Mexico and southern Florida. In this study, deviations to the HWE were observed for several loci in some subpopulations with mean values of F_{IS} varying from 0.066 (X) to 0.188 (TX), which suggests a heterozygote deficit not previously reported for *L. griseus*. MicroChecker indicated that the most possible cause for the significant values in the F_{IS} index could be the presence of null alleles. Null alleles are produced when there is a mutation in one of the primer binding sites of the microsatellites; this prevents annealing with the designed primer preventing the amplification of one of the alleles, resulting in a false homozygote (Estoup et al., 1998; Chapuis & Estoup, 2007). Such condition is frequent in microsatellites given their high levels of polymorphism. This has been reported in various species especially in populations with high effective size (Neff & Gross, 2001; Chapuis & Estoup, 2007). Additionally, fish have a high mutation rate in comparison with other classes (reptiles, birds, amphibians and mammals), because they have larger microsatellites and length is an important factor that influences mutation rate (Neff & Gross, 2001). There are ecological explanations for deviations from HWE including inbreeding and genetic drift. The deficiency in this case is unlikely to be explained by inbreeding due to the reproductive habits of *L. griseus*, given that they form aggregations and in these, gametes are simultaneously released to be fertilized (Claro & Lindeman, 2008). However, heterozygote excesses have been explained by sweepstakes recruitment in marine organisms with small pelagic larvae period and highly variable adult reproductive success (Hedgecock, 1994) which can lead to instantaneous genetic drift.

Conclusions

The high levels of genetic diversity, similar to those observed in other gray snapper populations from the northern Gulf of Mexico, as well as the high levels of gene flow in general, suggest that *L. griseus* constitutes a single genetic population in the Mexican Caribbean and the southern Gulf of Mexico. The absence of population bottlenecks or disturbances on its connectivity across this large region should be taken into account to define a proper management for the stock. More work is needed to verify the connectivity and the apparent unidirectionality in gene flow in the southern Gulf of Mexico (east-to-west). The next priority for understanding gray snapper populations in the Western Atlantic is to determine the degree of connectivity between Mexican populations and the northern Gulf of Mexico and between Mexico and Cuba.

Acknowledgements

We thank Mauricio Iván Espadas Alcocer and Roberto Zamora Bustillos for providing laboratory and field assistance with the samples obtained in the Mexican Caribbean Sea. Jason Selwyn for his support in the laboratory at TAMUCC. Chloe Rosas and Dariel Correa for the translation of the document.

References

- Aguirre-Planter E. 2007. Flujo génico: métodos para estimarlo y marcadores moleculares. In: Eguiarte LE, Souza V, Aguirre X eds. *Ecología molecular*. Ciudad de México, México: Instituto Nacional de Ecología (Mexico), Universidad Nacional de México, & Comisión Nacional para el Conocimiento y Uso de la Biodiversidad (México), 40–61.
- Athié G, Candela J, Sheinbaum J, Badan A, Ochoa J. 2011. Yucatan Current variability through the Cozumel and Yucatan channels estructura de la corriente de Yucatán en los canales de Cozumel y Yucatán. *Ciencias Marinas* 37:471–492.
- Ault JS, Bohnsack JA, Meester GA. 1998. A retrospective (1979-1996) multispecies assessment of coral reef fish stocks in the Florida Keys. *Fishery Bulletin* 96:395–414.
- Beerli P. 2016.Migrate: documentation and program. Version 3x. Available at <https://peterbeerli.com/migrate-html5/index.html>. (accessed June 26, 2019).
- Benjamini Y, Yekutieli D. 2001. The control of the false discovery rate in multiple testing under dependency. *The Annals of Statistics* 29:1165–1188.
- Botstein D, White RL, Skolnick M, Davis RW. 1980. Construction of a genetic linkage map in man using restriction fragment length polymorphisms. *American Journal of Human Genetics* 32:314–331.
- Cabrero J, Camacho JP. 2002. Fundamentos de genética de poblaciones. In: Soler M ed. *Evolución: La base de la Biología*. Granada, Es: Proyecto Sur de Ediciones, S.L., 83–126.

397 Carrillo L, Johns EM, Smith RH, Lamkin JT, Largier JL. 2015. Pathways and hydrography in the
398 mesoamerican barrier reef system part 1: circulation. *Continental Shelf Research*
399 109:164–176. DOI: 10.1016/j.csr.2015.09.014.

400 Carrillo L, Lamkin JT, Johns EM, Vásquez-Yeomans L, Sosa-Cordero F, Malca E, Smith RH,
401 Gerard T. 2017. Linking oceanographic processes and marine resources in the western
402 Caribbean Sea Large Marine Ecosystem Subarea. *Environmental Development* 22:84–96.
403 DOI: 10.1016/j.envdev.2017.01.004.

404 Chapuis M-P, Estoup A. 2007. Microsatellite null alleles and estimation of population
405 differentiation. *Molecular Biology and Evolution* 24:621–631. DOI:
406 10.1093/molbev/msl191.

407 Claro R, Lindeman KC. 2008. *Biología y manejo de los pargos (Lutjanidae) en el Atlántico*
408 *occidental*. Habana, Cuba: Instituto de oceanología.

409 Claro R, Mitcheson YS de, Lindeman KC, García-Cagide AR. 2009. Historical analysis of
410 Cuban commercial fishing effort and the effects of management interventions on
411 important reef fishes from 1960–2005. *Fisheries Research* 99:7–16. DOI:
412 10.1016/j.fishres.2009.04.004.

413 Cockerham CC, Weir BS. 1984. Covariances of relatives stemming from a population
414 undergoing mixed self and random mating. *Biometrics* 40:157–164. DOI:
415 10.2307/2530754.

416 Costello C, Ovando D, Hilborn R, Gaines DG, Deschenes O, Lester SE. 2012. Status and
417 solutions for the world's unassessed fisheries. *Science* 338:517–520.

418 Cowen RK. 2006. Scaling of connectivity in marine populations. *Science* 311:522–527. DOI:
419 10.1126/science.1122039.

420 Dakin EE, Avise JC. 2004. Microsatellite null alleles in parentage analysis. *Heredity* 93:504–
421 509. DOI: 10.1038/sj.hdy.6800545.

422 Domeier ML, Koenig C, Coleman F. 1996. Reproductive biology of the gray snapper (*Lutjanus*
423 *griseus*), with notes on spawning for other western Atlantic snappers (*Lutjanidae*). In:
424 Arreguin-Sánchez F, Munro JL, Balgos MC, Pauly D eds. *Biology and culture of tropical*
425 *groupers and snappers*. Campeche, México: International center for living aquatic
426 resources management, 189–201.

427 Espinoza Ávalos J. 2009. *El sistema ecológico de la bahía de Chetumal / Corozal: costa*
428 *occidental del Mar Caribe*. Quintana Roo, México: El Colegio de la Frontera Sur.

429 Estoup A, Fousset F, Michalakakis Y, Cornuet M, Adriamanga, Guyomard R. 1998. Comparative
430 analysis of microsatellite and allozyme markers: a case study investigating
431 microgeographic differentiation in brown trout (*salmo trutta*). *Molecular Ecology* 7:339–
432 353.

433 Excoffier L, Lischer HEL. 2010. Arlequin suite ver 3.5: a new series of programs to perform
434 population genetics analyses under Linux and Windows. *Molecular Ecology Resources*
435 10:564–567. DOI: 10.1111/j.1755-0998.2010.02847.x.

436 FAO. 2011. *Fishery and aquaculture statistics year book*. Rome: FAO.

437 Gold JR, Pak E, Richardson LR. 2001. Microsatellite variation among red snapper (*Lutjanus*
438 *campechanus*) from the Gulf of Mexico. *Marine Biotechnology* 3:293–304. DOI:
439 10.1007/s10126-001-0004-7.

440 Gold JR, Saillant E, Ebelt ND, Lem S. 2009. Conservation genetics of gray snapper (*Lutjanus*
441 *griseus*) in U.S. Waters of the northern Gulf of Mexico and Western Atlantic ocean.
442 *Copeia* 2009:277–286. DOI: 10.1643/CI-08-071.

443 Goldstein DB, Schlötterer C. 1999. *Microsatellites, evolution and applications*. Nueva York:
444 Oxford.

445 Hedgecock D. 1994. Temporal and spatial genetic structure of marine animal populations in the
446 California current. *California Cooperative Oceanic Fisheries Investigations Reports*
447 35:73–81.

448 Hogan JD, Thiessen RJ, Sale PF, Heath DD. 2012. Local retention, dispersal and fluctuating
449 connectivity among populations of a coral reef fish. *Oecologia* 168:61–71. DOI:
450 10.1007/s00442-011-2058-1.

451 INAPESCA. 2012. Carta Nacional Pesquera. *Diario oficial de la federacion* 2:21–128.

452 INAPESCA. 2018. Carta Nacional Pesquera. *Diario oficial de la federacion* 2:96–99.

453 Jones G, Srinivasan M, Almany G. 2007. Population connectivity and conservation of marine
454 biodiversity. *Oceanography* 20:100–111. DOI: 10.5670/oceanog.2007.33.

455 Keeney DB, Heupel MR, Hueter RE, Heist EJ. 2005. Microsatellite and mitochondrial DNA
456 analyses of the genetic structure of blacktip shark (*Carcharhinus limbatus*) nurseries in
457 the northwestern Atlantic, Gulf of Mexico, and Caribbean Sea: BLACKTIP SHARK
458 POPULATION STRUCTURE. *Molecular Ecology* 14:1911–1923. DOI: 10.1111/j.1365-
459 294X.2005.02549.x.

460 Kimura M, Crow JF. 1964. The number of alleles that can be maintained in a finite population.
461 *Genetics* 49:725–738.

462 Labastida-Estrada E, Machkour-M’Rabet S, Carrillo L, Hénaut Y, Castelblanco-Martínez DN.
463 2019. Genetic structure of Mexican lionfish populations in the southwest Gulf of Mexico
464 and the Caribbean Sea. *PLOS ONE* 14:e0222997. DOI: 10.1371/journal.pone.0222997.

465 Lara-Lara JR, Arreola Lizárraga JA, Calderón Aguilera LE, Camacho Ibar VF, de la Lanza
 466 Espino G, Escofet Giansone A, Espejel Carbajal MI, Guzmán Arroyo M, Ladah LB,
 467 López Hernández M, Meling López EA, Moreno Casa sola Barceló P, Reyes-Bonilla H,
 468 Ríos Jara E, Zertuche-González JA. 2008. *Capital natural de México*. Ciudad de México,
 469 México: Conocimiento actual de la biodiversidad, México.

470 Lindeman K, Anderson W, Carpenter KE, Claro R, Cowan J, Padovani-Ferreira B, Rocha LA,
 471 Sedberry G, Zapp-Sluis M. 2016. *Lutjanus griseus*. *The IUCN Red List of Threatened*
 472 *Species 2016* e.T12416A506350. DOI: [http://dx.doi.org/10.2305/IUCN.UK.2016-](http://dx.doi.org/10.2305/IUCN.UK.2016-1.RLTS.T12416A506350.en)
 473 [1.RLTS.T12416A506350.en](http://dx.doi.org/10.2305/IUCN.UK.2016-1.RLTS.T12416A506350.en).

474 Martínez S, Carrillo L, Marinone SG. 2019. Potential connectivity between marine protected
 475 areas in the mesoamerican Reef for two species of virtual fish larvae: *Lutjanus analis* and
 476 *Epinephelus striatus*. *Ecological Indicators* 102:10–20. DOI:
 477 [10.1016/j.ecolind.2019.02.027](https://doi.org/10.1016/j.ecolind.2019.02.027).

478 Meirmans PG, Hedrick PW. 2011. Assessing population structure: FST and related measures.
 479 *Molecular Ecology Resources* 11:5–18. DOI: [10.1111/j.1755-0998.2010.02927.x](https://doi.org/10.1111/j.1755-0998.2010.02927.x).

480 Neff BD, Gross MR. 2001. Microsatellite evolution in vertebrates: inference from AC
 481 dinucleotide repeats. *The Society for the Study of Evolution* 55:1717–1733.

482 Park S. 2008. Excel Microsatellite Toolkit. Available at
 483 <http://animalgenomics.ucd.ie/sdepark/ms-toolkit/> (accessed March 16, 2019).

484 Peakall R, Smouse PE. 2012. GenAlEx 6.5: genetic analysis in Excel. Population genetic
 485 software for teaching and research-an update. *Bioinformatics* 28:2537–2539. DOI:
 486 [10.1093/bioinformatics/bts460](https://doi.org/10.1093/bioinformatics/bts460).

- Renshaw MA, Saillant E, Lem S, Berry P, Gold JR. 2007. Microsatellite multiplex panels for genetic studies of gray snapper (*Lutjanus griseus*) and lane snapper (*Lutjanus synagris*). *Fishery Bulletin* 105:436–439.
- Rocha ÉC, Molina WF. 2008. Cytogenetic analysis in western Atlantic snappers (Perciformes, Lutjanidae). *Genetics and Molecular Biology* 31:461–467.
- SAGARPA. 2012. Status of the Fisheries of Mexico. *Diario oficial de la federacion* 2:104–105.
- Salas-Monreal D, Salas Pérez J de J, Gómez MAM, España HP, Lozano LDO, Barba AG, Enzástiga MLR, Sánchez CAV. 2017. Corrientes superficiales dentro del corredor arrecifal del Suroeste del Golfo de México. *UVserva*:32–36.
- Sanvicente-Añorve L, Zavala-Hidalgo J, Allende-Arandía ME, Hermoso-Salazar M. 2014. Connectivity patterns among coral reef systems in the southern Gulf of Mexico. *Marine Ecology Progress Series* 498:27–41.
- Slatkin M. 1995. A measure of population subdivision based on microsatellite allele frequencies. *Genetics Society of America* 139:457–462.
- Urbiola-Rangel E, Chassin-Noria O. 2013. Conectividad genética de *Stegastes acapulcoensis* (Pomacentridae) en el Pacífico central de México. *Hidrobiológica* 23:415–419.
- Van Oosterhout C, Hutchinson WF, Wills DPM, Shipley P. 2004. micro-checker: software for identifying and correcting genotyping errors in microsatellite data. *Molecular Ecology Notes* 4:535–538. DOI: 10.1111/j.1471-8286.2004.00684.x.
- Villegas Sánchez CA, Pérez España H, Rivera Madrid R, Salas Monreal D, Arias González JE. 2014. Subtle genetic connectivity between Mexican Caribbean and south-western Gulf of Mexico reefs: the case of the bicolor damselfish, *Stegastes partitus*. *Coral Reefs* 33:241–251. DOI: 10.1007/s00338-013-1083-4.

511 Athié G, Candela J, Sheinbaum J, Badan A, Ochoa J. 2011. Yucatan Current variability through
512 the Cozumel and Yucatan channels estructura de la corriente de Yucatán en los canales
513 de Cozumel y Yucatán. *Ciencias Marinas* 37:471–492.

514 Ault JS, Bohnsack JA, Meester GA. 1998. A retrospective (1979-1996) multispecies assessment
515 of coral reef fish stocks in the Florida Keys. *Fishery Bulletin* 96:395–414.

516 Beerli P. 2016. Migrate: documentation and program. Version 3x. Available at
517 <https://peterbeerli.com/migrate-html5/index.html>. (accessed June 26, 2019).

518 Benjamini Y, Yekutieli D. 2001. The control of the false discovery rate in multiple testing under
519 dependency. *The Annals of Statistics* 29:1165–1188.

520 Botstein D, White RL, Skolnick M, Davis RW. 1980. Construction of a genetic linkage map in
521 man using restriction fragment length polymorphisms. *American Journal of Human*
522 *Genetics* 32:314–331.

523 Cabrero J, Camacho JP. 2002. Fundamentos de genética de poblaciones. In: Soler M ed.
524 *Evolución: La base de la Biología*. Granada, Es: Proyecto Sur de Ediciones, S.L., 83–
525 126.

526 Carrillo L, Johns EM, Smith RH, Lamkin JT, Largier JL. 2015. Pathways and hydrography in the
527 mesoamerican barrier reef system part 1: circulation. *Continental Shelf Research*
528 109:164–176. DOI: 10.1016/j.csr.2015.09.014.

529 Chapuis M-P, Estoup A. 2007. Microsatellite null alleles and estimation of population
530 differentiation. *Molecular Biology and Evolution* 24:621–631. DOI:
531 10.1093/molbev/msl191.

532 Claro R, Lindeman KC. 2008. *Biología y manejo de los pargos (Lutjanidae) en el Atlántico*
533 *occidental*. Habana, Cuba: Instituto de oceanología.

534 Claro R, Mitcheson YS de, Lindeman KC, García-Cagide AR. 2009. Historical analysis of
535 Cuban commercial fishing effort and the effects of management interventions on
536 important reef fishes from 1960–2005. *Fisheries Research* 99:7–16. DOI:
537 10.1016/j.fishres.2009.04.004.

538 Cockerham CC, Weir BS. 1984. Covariances of relatives stemming from a population
539 undergoing mixed self and random mating. *Biometrics* 40:157–164. DOI:
540 10.2307/2530754.

541 Cowen RK. 2006. Scaling of connectivity in marine populations. *Science* 311:522–527. DOI:
542 10.1126/science.1122039.

543 Dakin EE, Avise JC. 2004. Microsatellite null alleles in parentage analysis. *Heredity* 93:504–
544 509. DOI: 10.1038/sj.hdy.6800545.

545 Domeier ML, Koenig C, Coleman F. 1996. Reproductive biology of the gray snapper (*Lutjanus*
546 *griseus*), with notes on spawning for other western Atlantic snappers (*Lutjanidae*). In:
547 Arreguin-Sánchez F, Munro JL, Balgos MC, Pauly D eds. *Biology and culture of tropical*
548 *groupers and snappers*. Campeche, México: International center for living aquatic
549 resources management, 189–201.

550 Espinoza Ávalos J. 2009. *El sistema ecológico de la bahía de Chetumal / Corozal: costa*
551 *occidental del Mar Caribe*. Quintana Roo, México: El Colegio de la Frontera Sur.

552 Estoup A, Fousset F, Michalakis Y, Cornuet M, Adriamanga, Guyomard R. 1998. Comparative
553 analysis of microsatellite and allozyme markers: a case study investigating
554 microgeographic differentiation in brown trout (*salmo trutta*). *Molecular Ecology* 7:339–
555 353.

Excoffier L, Lischer HEL. 2010. Arlequin suite ver 3.5: a new series of programs to perform population genetics analyses under Linux and Windows. *Molecular Ecology Resources* 10:564–567. DOI: 10.1111/j.1755-0998.2010.02847.x.

FAO. 2011. *Fishery and aquaculture statistics year book*. Rome: FAO.

Gold JR, Pak E, Richardson LR. 2001. Microsatellite variation among red snapper (*Lutjanus campechanus*) from the Gulf of Mexico. *Marine Biotechnology* 3:293–304. DOI: 10.1007/s10126-001-0004-7.

Gold JR, Saillant E, Ebel ND, Lem S. 2009. Conservation genetics of gray snapper (*Lutjanus griseus*) in U.S. Waters of the northern Gulf of Mexico and Western Atlantic ocean. *Copeia* 2009:277–286. DOI: 10.1643/CI-08-071.

Goldstein DB, Schlötterer C. 1999. *Microsatellites, evolution and applications*. Nueva York: Oxford.

Hedgecock D. 1994. Temporal and spatial genetic structure of marine animal populations in the California current. *California Cooperative Oceanic Fisheries Investigations Reports* 35:73–81.

Hogan JD, Thiessen RJ, Sale PF, Heath DD. 2012. Local retention, dispersal and fluctuating connectivity among populations of a coral reef fish. *Oecologia* 168:61–71. DOI: 10.1007/s00442-011-2058-1.

INAPESCA. 2012. Carta Nacional Pesquera. *Diario oficial de la federacion* 2:21–128.

Jones G, Srinivasan M, Almany G. 2007. Population connectivity and conservation of marine biodiversity. *Oceanography* 20:100–111. DOI: 10.5670/oceanog.2007.33.

Keeney DB, Heupel MR, Hueter RE, Heist EJ. 2005. Microsatellite and mitochondrial DNA analyses of the genetic structure of blacktip shark (*Carcharhinus limbatus*) nurseries in

the northwestern Atlantic, Gulf of Mexico, and Caribbean Sea: BLACKTIP SHARK
POPULATION STRUCTURE. *Molecular Ecology* 14:1911–1923. DOI: 10.1111/j.1365-
294X.2005.02549.x.

Kimura M, Crow JF. 1964. The number of alleles that can be maintained in a finite population.
Genetics 49:725–738.

Labastida-Estrada E, Machkour-M'Rabet S, Carrillo L, Hénaut Y, Castelblanco-Martínez DN.
2019. Genetic structure of Mexican lionfish populations in the southwest Gulf of Mexico
and the Caribbean Sea. *PLOS ONE* 14:e0222997. DOI: 10.1371/journal.pone.0222997.

Lara-Lara JR, Arreola Lizárraga JA, Calderón Aguilera LE, Camacho Ibar VF, de la Lanza
Espino G, Escofet Giansone A, Espejel Carbajal MI, Guzmán Arroyo M, Ladah LB,
López Hernández M, Meling López EA, Moreno Casa sola Barceló P, Reyes-Bonilla H,
Ríos Jara E, Zertuche-González JA. 2008. *Capital natural de México*. Ciudad de México,
México: Conocimiento actual de la biodiversidad, México.

Lindeman K, Anderson W, Carpenter KE, Claro R, Cowan J, Padovani-Ferreira B, Rocha LA,
Sedberry G, Zapp-Sluis M. 2016. *Lutjanus griseus*. *The IUCN Red List of Threatened
Species 2016* e.T12416A506350. DOI: <http://dx.doi.org/10.2305/IUCN.UK.2016-1.RLTS.T12416A506350.en>.

Martínez S, Carrillo L, Marinone SG. 2019. Potential connectivity between marine protected
areas in the mesoamerican Reef for two species of virtual fish larvae: *Lutjanus analis* and
Epinephelus striatus. *Ecological Indicators* 102:10–20. DOI:
10.1016/j.ecolind.2019.02.027.

Meirmans PG, Hedrick PW. 2011. Assessing population structure: FST and related measures.
Molecular Ecology Resources 11:5–18. DOI: 10.1111/j.1755-0998.2010.02927.x.

602 Neff BD, Gross MR. 2001. Microsatellite evolution in vertebrates: inference from AC
603 dinucleotide repeats. *The Society for the Study of Evolution* 55:1717–1733.

604 Park S. 2008. Excel Microsatellite Toolkit. Available at
605 <http://animalgenomics.ucd.ie/sdepark/ms-toolkit/> (accessed March 16, 2019).

606 Peakall R, Smouse PE. 2012. GenAlEx 6.5: genetic analysis in Excel. Population genetic
607 software for teaching and research-an update. *Bioinformatics* 28:2537–2539. DOI:
608 10.1093/bioinformatics/bts460.

609 Renshaw MA, Saillant E, Lem S, Berry P, Gold JR. 2007. Microsatellite multiplex panels for
610 genetic studies of gray snapper (*Lutjanus griseus*) and lane snapper (*Lutjanus synagris*).
611 *Fishery Bulletin* 105:436–439.

612 Rocha ÉC, Molina WF. 2008. Cytogenetic analysis in western Atlantic snappers (Perciformes,
613 Lutjanidae). *Genetics and Molecular Biology* 31:461–467.

614 SAGARPA. 2012. Status of the Fisheries of Mexico. *Diario oficial de la federacion* 2:104–105.

615 Salas-Monreal D, Salas Pérez J de J, Gómez MAM, España HP, Lozano LDO, Barba AG,
616 Enzástiga MLR, Sánchez CAV. 2017. Corrientes superficiales dentro del corredor
617 arrecifal del Suroeste del Golfo de México. *UVserva*:32–36.

618 Sanvicente-Añorve L, Zavala-Hidalgo J, Allende-Arandía ME, Hermoso-Salazar M. 2014.
619 Connectivity patterns among coral reef systems in the southern Gulf of Mexico. *Marine*
620 *Ecology Progress Series* 498:27–41.

621 Slatkin M. 1995. A measure of population subdivision based on microsatellite allele frequencies.
622 *Genetics Society of America* 139:457–462.

623 Urbiola-Rangel E, Chassin-Noria O. 2013. Conectividad genética de *Stegastes acapulcoensis*
624 (*Pomacentridae*) en el Pacífico central de México. *Hidrobiológica* 23:415–419.

625 Van Oosterhout C, Hutchinson WF, Wills DPM, Shipley P. 2004. micro-checker: software for
 626 identifying and correcting genotyping errors in microsatellite data. *Molecular Ecology*
 627 *Notes* 4:535–538. DOI: 10.1111/j.1471-8286.2004.00684.x.

628 Villegas Sánchez CA, Pérez España H, Rivera Madrid R, Salas Monreal D, Arias González JE.
 629 2014. Subtle genetic connectivity between Mexican Caribbean and south-western Gulf of
 630 Mexico reefs: the case of the bicolor damselfish, *Stegastes partitus*. *Coral Reefs* 33:241–
 631 251. DOI: 10.1007/s00338-013-1083-4.

632

Figure 1

Geographic location of the seven sampling sites.

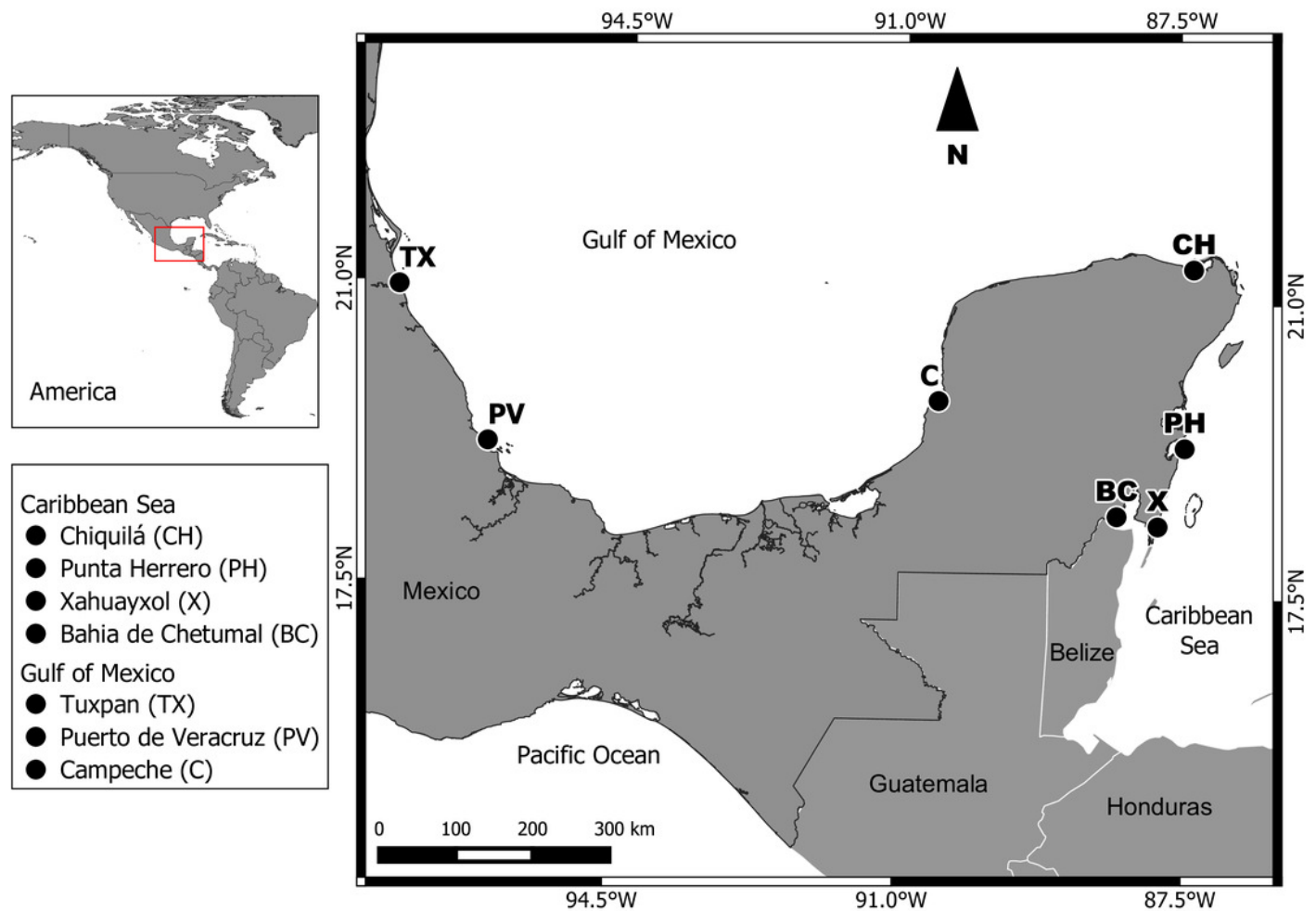


Figure 2

Directionality of migration in the Caribbean and the Gulf of Mexico between adjacent sites.

Based on the results obtained from Migrate, lines represent migration between points; thicker lines represent stronger levels of genetic flow and the values correspond to *MLE*. Abbreviations: Campeche (C), Puerto de Veracruz (PV), Tuxpan (TX), Bahia de Chetumal (BC), Chiquilá (CH), Punta Herrero (PH) and Xahuaxol (X).

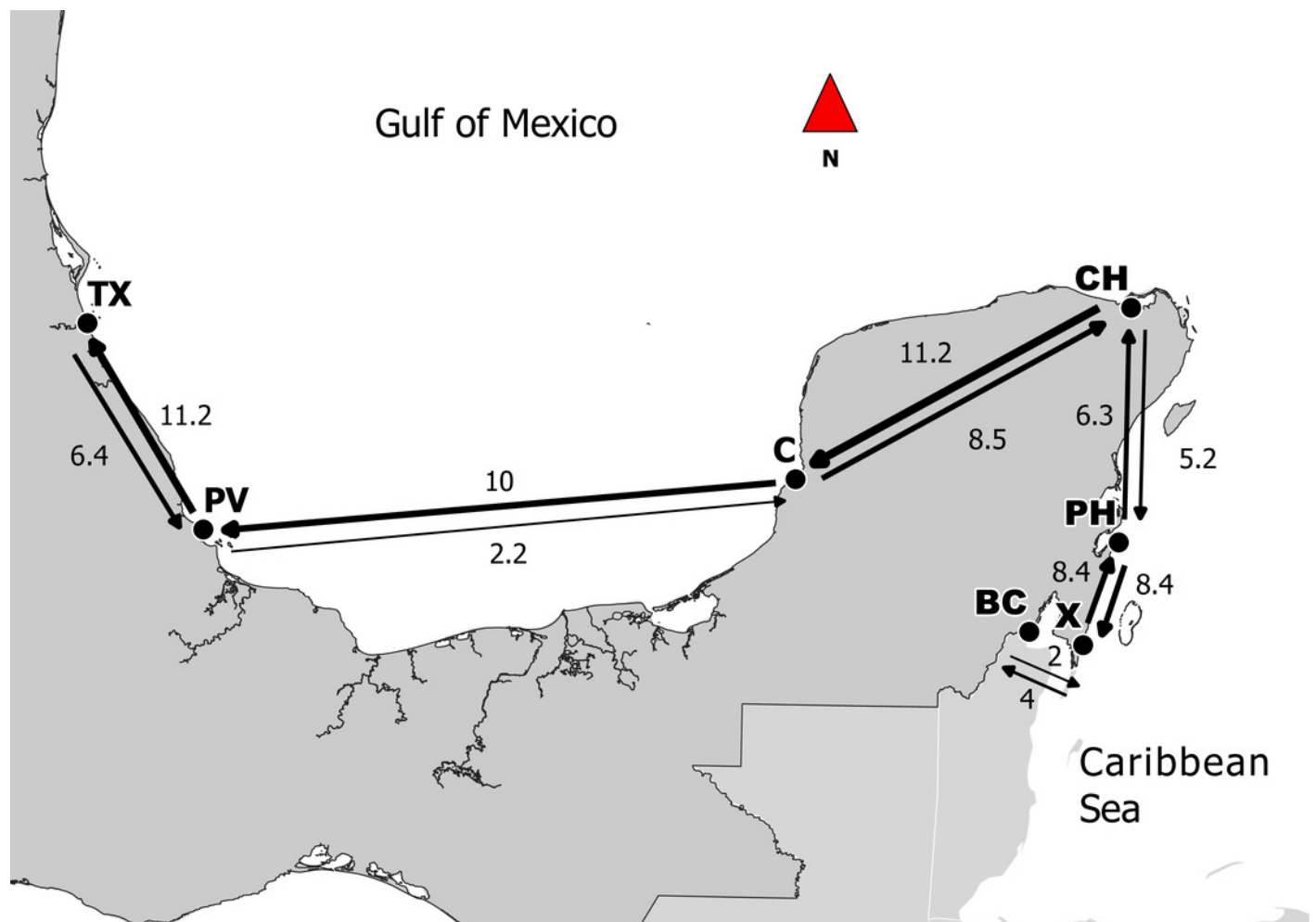


Table 1 (on next page)

Genetic diversity of the gray snapper (*Lutjanus griseus*) in the southern Gulf of Mexico.

Values in bold indicate significant deviations with respect to the Hardy Weinberg Equilibrium after applying the false discovery rate. Numbers below primer names are the annealing temperatures. Abbreviations: number of alleles (N_a), number of effective alleles (A_E), observed heterozygosity (H_o), expected heterozygosity (H_e), fixation index (F_{IS}), frequency of null alleles (F_a), total heterozygosity (H_t), Campeche (C), Puerto de Veracruz (PV), Tuxpan (TX).

Gulf of Mexico										
Sites		Lca20 58° C	Lca43 56° C	Prs260 56° C	Ra1 58° C	Lca107 48° C	Prs137 54° C	Prs275 54° C	Prs328 54° C	Mean
<i>C</i>	<i>Na</i>	5	2	7	9	8	9	4	3	5.875
	<i>A_E</i>	3.814	1.259	5.300	2.312	2.970	5.545	2.784	2.220	3.275
	<i>Ho</i>	0.720	0.186	0.837	0.542	0.327	0.531	0.500	0.600	0.530
	<i>He</i>	0.745	0.208	0.820	0.573	0.670	0.828	0.647	0.555	0.631
	<i>F_{IS}</i>	0.034	0.106	-0.021	0.056	0.515	0.362	0.229	-0.082	0.150
	<i>Fa</i>	0.012	0.042	0.022	0.019	0.236	0.176	0.112	0.045	0.083
<i>PV</i>	<i>Na</i>	5	2	11	8	6	11	5	4	6.500
	<i>A_E</i>	3.295	1.350	5.938	2.140	2.828	5.598	2.860	2.261	3.284
	<i>Ho</i>	0.574	0.265	0.660	0.449	0.378	0.522	0.740	0.500	0.511
	<i>He</i>	0.704	0.262	0.840	0.538	0.654	0.830	0.657	0.563	0.631
	<i>F_{IS}</i>	0.186	-0.013	0.216	0.167	0.425	0.374	-0.128	0.114	0.168
	<i>Fa</i>	0.087	0.012	0.104	0.073	0.190	0.179	0.074	0.050	0.096
<i>TX</i>	<i>Na</i>	4	2	9	9	8	8	6	4	6.250
	<i>A_E</i>	3.465	1.227	5.128	2.211	3.952	3.468	3.149	2.318	3.115
	<i>Ho</i>	0.796	0.147	0.700	0.531	0.479	0.362	0.520	0.520	0.507
	<i>He</i>	0.719	0.187	0.813	0.553	0.755	0.719	0.689	0.574	0.626
	<i>F_{IS}</i>	-0.109	0.218	0.140	0.041	0.368	0.500	0.247	0.095	0.188
	<i>Fa</i>	0.062	0.079	0.067	0.043	0.177	0.242	0.117	0.047	0.104
	<i>Ht</i>	0.720	0.217	0.819	0.552	0.693	0.796	0.667	0.559	0.628
	<i>F_{IS}</i>	0.026	0.079	0.103	0.077	0.425	0.399	0.108	0.033	0.156

Table 2 (on next page)

Genetic diversity of the gray snapper (*Lutjanus griseus*) in the Mexican Caribbean Sea.

Values in bold indicate significant deviations with respect to the Hardy Weinberg Equilibrium after applying the false discovery rate. Numbers below primer names are Polymorphic information content (PIC). Abbreviations: number of alleles (N_a), number of effective alleles (A_E), observed heterozygosity (H_o), expected heterozygosity (H_e), fixation index (F_{IS}), frequency of null alleles (F_a), total heterozygosity (H_t), Bahia de Chetumal (BC), Chiquilá (CH), Punta Herrero (PH) and Xahuaxol (X).

1

Mexican Caribbean Sea										Mean
Sites	PIC	Lca20	Lca43	Prs260	Ra1	Lca107	Prs137	Prs275	Prs328	0.597
<i>BC</i>	<i>Na</i>	4	3	10	11	7	8	5	3	6.375
	<i>A_E</i>	3.208	1.446	5.654	2.407	3.480	5.013	2.939	2.224	3.296
	<i>Ho</i>	0.766	0.326	0.833	0.612	0.837	0.776	0.708	0.510	0.671
	<i>He</i>	0.696	0.312	0.832	0.591	0.720	0.809	0.667	0.556	0.648
	<i>F_{IS}</i>	-0.102	-0.047	-0.002	-0.037	-0.164	0.042	-0.063	0.083	-0.036
	<i>Fa</i>	0.057	0.041	0.007	0.037	0.126	0.010	0.040	0.041	0.045
<i>CH</i>	<i>Na</i>	4	2	8	10	6	7	5	4	5.750
	<i>A_E</i>	3.476	1.403	5.020	2.666	2.807	4.754	2.944	2.262	3.167
	<i>Ho</i>	0.766	0.348	0.813	0.592	0.548	0.689	0.708	0.500	0.620
	<i>He</i>	0.720	0.290	0.809	0.631	0.651	0.799	0.667	0.564	0.641
	<i>F_{IS}</i>	-0.065	-0.200	-0.004	0.063	0.161	0.139	-0.062	0.114	0.018
	<i>Fa</i>	0.037	0.192	0.007	0.028	0.093	0.064	0.036	0.057	0.064
<i>PH</i>	<i>Na</i>	4	5	11	9	7	8	5	4	6.625
	<i>A_E</i>	3.309	1.491	6.394	2.900	3.103	4.469	2.744	2.328	3.342
	<i>Ho</i>	0.705	0.370	0.796	0.580	0.619	0.600	0.694	0.571	0.617
	<i>He</i>	0.706	0.333	0.852	0.662	0.686	0.786	0.642	0.576	0.655
	<i>F_{IS}</i>	0.002	-0.112	0.067	0.125	0.099	0.239	-0.082	0.008	0.043
	<i>Fa</i>	0.006	0.062	0.024	0.059	0.045	0.118	0.047	0.005	0.046
<i>X</i>	<i>Na</i>	5	2	10	9	6	8	4	3	5.875
	<i>A_E</i>	3.700	1.402	6.011	2.515	2.926	5.193	3.137	2.192	3.384
	<i>Ho</i>	0.766	0.265	0.851	0.673	0.500	0.804	0.531	0.500	0.611
	<i>He</i>	0.738	0.290	0.843	0.609	0.671	0.816	0.688	0.550	0.650
	<i>F_{IS}</i>	-0.039	0.085	-0.010	-0.108	0.259	0.015	0.231	0.091	0.066
	<i>Fa</i>	0.029	0.035	0.013	0.052	0.121	0.005	0.107	0.041	0.050
	<i>Ht</i>	0.712	0.303	0.833	0.619	0.679	0.799	0.666	0.564	0.647
	<i>F_{IS}</i>	-0.062	-0.080	0.003	0.004	0.070	0.096	-0.002	0.063	0.012

2

Table 3 (on next page)

Pairwise values of F_{ST} , R_{ST} indexes between subpopulations and number of effective migrants per generations (N_m).

Abbreviations: subpopulations (*Spop*), *p*-value (*p*), the standard error (SE), Campeche (C), Puerto de Veracruz (PV), Tuxpan (TX), Bahia de Chetumal (BC), Chiquilá (CH), Punta Herrero (PH) and Xahuaxol (X).

<i>Spop</i>		F_{st}	R_{st}		N_m
			p	SE±	
<i>C</i>	<i>PV</i>	0.006	0.432	0.006	11.5
<i>C</i>	<i>TX</i>	0.006	0.648	0.005	5.6
<i>C</i>	<i>BC</i>	0.005	0.620	0.005	5.7
<i>C</i>	<i>CH</i>	0.007	0.630	0.005	16.9
<i>C</i>	<i>PH</i>	0.006	0.736	0.005	6.1
<i>C</i>	<i>X</i>	0.006	0.375	0.004	6.8
<i>PV</i>	<i>TX</i>	0.005	0.804	0.004	16.7
<i>PV</i>	<i>BC</i>	0.005	0.410	0.005	6.3
<i>PV</i>	<i>CH</i>	0.005	0.398	0.005	8.9
<i>PV</i>	<i>PH</i>	0.006	0.162	0.003	17.0
<i>PV</i>	<i>X</i>	0.003	0.759	0.004	7.8
<i>TX</i>	<i>BC</i>	0.006	0.881	0.003	4.5
<i>TX</i>	<i>CH</i>	0.008	0.315	0.004	5.2
<i>TX</i>	<i>PH</i>	0.007	0.407	0.005	4.1
<i>TX</i>	<i>X</i>	0.007	0.328	0.004	6.6
<i>BC</i>	<i>CH</i>	0.008	0.200	0.004	14.9
<i>BC</i>	<i>PH</i>	0.005	0.664	0.005	13.2
<i>BC</i>	<i>X</i>	0.006	0.121	0.003	8.1
<i>CH</i>	<i>PH</i>	0.005	0.333	0.005	13.3
<i>CH</i>	<i>X</i>	0.003	0.671	0.005	10.1
<i>PH</i>	<i>X</i>	0.005	0.099	0.003	25.2

Table 4(on next page)

Values of the migration parameter M . In the diagonal cross section appear estimations of effective population sizes (N_e).

Abbreviations: receiving subpopulation (+), Campeche (C), Puerto de Veracruz (PV), Tuxpan (TX), Bahia de Chetumal (BC), Chiquilá (CH), Punta Herrero (PH) and Xahuaxol (X).

1

	<i>C+</i>	<i>PV+</i>	<i>TX+</i>	<i>BC+</i>	<i>CH+</i>	<i>PH+</i>	<i>X+</i>
<i>C</i>	2439	10.0	2.2	1.9	8.5	3.2	1.7
<i>PV</i>	2.2	1975	11.2	2.9	3.1	3.4	3.0
<i>TX</i>	2.8	6.4	3058	1.9	3.0	1.7	2.8
<i>BC</i>	3.7	4.0	2.1	2549	7.3	6.8	2.0
<i>CH</i>	11.2	8.4	1.9	9.7	1922	5.2	3.0
<i>PH</i>	2.0	9.7	1.3	4.3	6.3	3686	8.4
<i>X</i>	3.4	3.5	2.1	4.0	5.1	8.4	3799

2

Table 5(on next page)

AMOVA results.

Abbreviations: degrees of freedom (*df*), sum of squares (*SS*), percent variance (*PV%*).

	<i>df</i>	<i>SS</i>	<i>PV%</i>	<i>F-statistics</i>
Among regions	1	0.722	0.17	$F_{ct} = 0$
Among subpopulations	5	8.248	0	$F_{sc} = 0$
Among individuals	341	560.754	1.94	$F_{is} = 0.019$
Within individuals	348	550.500	98.22	$F_{it} = 0.017$
	695	1120.224		

1
2