

Accumulation of di-2-ethylhexyl phthalate from polyvinyl chloride flooring into settled house dust and the effect on the bacterial community

Samantha Velazquez¹, Chenyang Bi^{2,3}, Jeff Kline^{1,4}, Susie Nunez¹, Rich Corsi^{3,5}, Ying Xu^{3,6}, Suzanne L Ishaq^{Corresp. 1,7}

¹ Biology and the Built Environment Center, University of Oregon, Eugene, Oregon, United States

² Department of Civil Engineering, Virginia Polytechnic Institute and State University (Virginia Tech), Blacksburg, Virginia, United States

³ Department of Civil, Architectural and Environmental Engineering, University of Texas at Austin, Austin, Texas, United States

⁴ Energy Studies and Buildings Laboratory, University of Oregon, Eugene, Oregon, United States

⁵ Fariborz Maseeh College of Engineering and Computer Science, Portland State University, Portland, Oregon, United States

⁶ Department of Building Science, Tsinghua University:(清华大学), Beijing, China

⁷ School of Food and Agriculture, University of Maine, Orono, Maine, United States

Corresponding Author: Suzanne L Ishaq

Email address: sue.ishaq@maine.edu

Di-2-ethylhexyl phthalate (DEHP) is a plasticizer used in consumer products and building materials, including polyvinyl chloride flooring material. DEHP adsorbs from material and leaches into soil, water, or dust, and presents an exposure risk to building occupants by inhalation, ingestion, or absorption. A number of bacterial isolates are demonstrated to degrade DEHP in culture, but bacteria may be susceptible to it as well, thus this study examined the relation of DEHP to bacterial communities in dust. Polyvinyl chloride flooring was seeded with homogenized house dust and incubated for up to 14 days, and bacterial communities in dust were identified at days 1, 7, and 14 using the V3-V4 regions of the bacterial 16S rRNA gene. DEHP concentration in dust increased over time, as expected, and bacterial richness and Shannon diversity were negatively correlated with DEHP concentration. Some sequence variants of *Bacillus*, *Corynebacterium jeddahense*, *Streptococcus*, and *Peptoniphilus* were relatively more abundant at low concentrations of DEHP, while some *Sphingomonas*, *Chryseobacterium*, and a member of the Enterobacteriaceae family were relatively more abundant at higher concentrations. The built environment is known to host lower microbial diversity and biomass than natural environments, and DEHP or other chemicals indoors may contribute to this paucity.

Accumulation of di-2-ethylhexyl phthalate from polyvinyl chloride flooring into settled house dust and the effect on the bacterial community.

Samantha Velazquez¹, Chenyang Bi^{2,3}, Jeff Kline^{1,4}, Susie Nunez¹, Richard Corsi^{3,5}, Ying Xu^{3,6}, Suzanne L. Ishaq^{1,7*}

Author emails:

SV: smv@uoregon.edu

CB: chenyangbi@vt.edu

JK: jkline@uoregon.edu

SN: snunez@uoregon.edu

RC: rcorsi@pdx.edu

YX: xuying@mail.utexas.edu

SLI: sue.ishaq@maine.edu

Affiliations:

¹ Biology and the Built Environment Center, University of Oregon, Eugene, OR, 97403

² Department of Civil Environmental Engineering, Virginia Polytechnic Institute and State University, Blacksburg, VA 24061 (current)

³ Department of Civil, Architectural and Environmental Engineering, University of Texas, Austin, TX 78712

⁴ Energy Studies and Buildings Laboratory, University of Oregon, Eugene, OR, 97403

⁵ Fariborz Maseeh College of Engineering and Computer Science, Portland State University, Portland, OR 97207 (current)

⁶ Department of Building Science, Tsinghua University, 100084, Beijing, P. R. China (current)

⁷ School of Food and Agriculture, University of Maine, Orono, ME 04469 (current)

*** Corresponding author**

Suzanne Ishaq, 130A Hitchner Hall, School of Food and Agriculture, University of Maine, Orono, ME 04469. sue.ishaq@maine.edu. 207-581-2770.

Keywords: DEHP; gas chromatography; Illumina MiSeq; indoor microbiome; vinyl flooring

Funding: This work was funded by grants from the Alfred P. Sloan Foundation to the Biology and the Built Environment Center (BioBE) at the University of Oregon.

Conflict of Interest: The authors declare no conflict of interest.

Author contributions: RC, SI, JK, and YX conceived of scope and experimental design, and contributed to writing and editing. CB and SN performed chemical and microbial laboratory protocols, respectively. SV performed DNA sequencing data analysis, literature review, writing, and editing, under the guidance of SI.

Abstract

Di-2-ethylhexyl phthalate (DEHP) is a plasticizer used in consumer products and building materials, including polyvinyl chloride flooring material. DEHP adsorbs from material and leaches into soil, water, or dust, and presents an exposure risk to building occupants by inhalation, ingestion, or absorption. A number of bacterial isolates are demonstrated to degrade DEHP in culture, but bacteria may be susceptible to it as well, thus this study examined the relation of DEHP to bacterial communities in dust. Polyvinyl chloride flooring was seeded with homogenized house dust and incubated for up to 14 days, and bacterial communities in dust were identified at days 1, 7, and 14 using the V3-V4 regions of the bacterial 16S rRNA gene. DEHP concentration in dust increased over time, as expected, and bacterial richness and Shannon diversity were negatively correlated with DEHP concentration. Some sequence variants of *Bacillus*, *Corynebacterium jeddahense*, *Streptococcus*, and *Peptoniphilus* were relatively more abundant at low concentrations of DEHP, while some *Sphingomonas*, *Chryseobacterium*, and a member of the Enterobacteriaceae family were relatively more abundant at higher concentrations. The built environment is known to host lower microbial diversity and biomass than natural environments, and DEHP or other chemicals indoors may contribute to this paucity.

Introduction

Di-2-ethylhexyl phthalate (DEHP), a known endocrine-disrupting chemical, is widely used in consumer goods and building materials, from where it is leached into dust, water, or soil and is readily inhaled, absorbed, or ingested by building occupants (Xu et al., 2009; Just et al., 2015; Rowdhwal & Chen, 2018; Bi et al., 2018). DEHP concentration is significantly higher in settled (i.e. surface) dust in homes with polyvinyl flooring (Just et al., 2015; Bi et al., 2018), as well as in HVAC filters (Bi et al., 2018), indicating regular aerial dispersion from floor surfaces. DEHP has been associated with increased reports of children's asthma during warmer seasons (Bi et al., 2018) and body mass index (Wang et al., 2013), and its byproducts or metabolites can likewise be hazardous to human health (Just et al., 2015; Rowdhwal & Chen, 2018; Bi et al., 2018; Bope et al., 2019).

Due to the proximity to human occupants and risk for exposure, indoor chemistry is relevant to health (Just et al., 2015; Wang et al., 2017; Rowdhwal & Chen, 2018). But our understanding of indoor chemistry is complicated by the diversity of chemical compounds present in the built environment, emission from materials, chemical interactions (e.g. adsorption, absorption) with materials, and local environmental conditions such as ventilation rate and temperature, all of which can alter chemical reactivity (Bi, Liang & Xu, 2015; Weschler & Carslaw, 2018). Moreover, some monocultures of bacteria may degrade chemical contaminants *in vitro* (Chao & Cheng, 2007; Vamsee-Krishna & Phale, 2008; Liang et al., 2008; Yang et al., 2018), including the phthalates that are emitted from vinyl flooring, as well as in soil or other environmental samples (Liang et al., 2008; Yang et al., 2018; Zhu et al., 2018). However, *in situ* microbial communities in indoor dust have only been demonstrated to degrade chemical contaminants under conditions of high relative humidity (80 -100 %), and this appeared to be

largely due to fungal degradation (Bope et al., 2019). It has been suggested that microorganisms naturally produce DEHP, particularly microorganisms in soils, (Ortiz & Sansinenea, 2018). Regardless, the presence of DEHP or other chemicals in dust may affect microbial survival (Brundrett et al., 2018; Lotfy et al., 2018; Zhu et al., 2018), which would alter the composition of microorganisms in dust and potential community dynamics.

Little research has been done to elucidate the relationships between phthalate composition and the effects on microbial communities in dust. This pilot study investigated whether dust which accumulated DEHP emitted from polyvinyl chloride flooring would affect the bacterial communities in dust gathered from homes.

Materials & Methods

Experimental design was similar to previous studies (Wu et al., 2016; Bi et al., 2018). Polyvinyl chloride flooring was cut into 60 mm diameter specimens, placed into glass petri dishes, and sprayed evenly with freshly collected, homogenized house dust containing living microbial communities (Fahimipour et al., 2018) at a concentration of 5 g/m². The dishes were placed into a 10 L stainless steel container with one inlet and one outlet for airflow, with an air change rate of 0.5 hr⁻¹, ~0% relative humidity (RH), 25 ± 0.5°C, and no light. Experimental control samples of dust in glass petri dishes, one per time point, were subjected to identical conditions without the inclusion of vinyl flooring, however the homogenized house dust did contain some background DEHP (291 µg/g dust). All samples were randomly placed in the steel container and selected for removal at each time period. During the incubation periods, DEHP

was passively adsorbed from polyvinyl flooring and loaded into dust in the experimental samples.

Petri dishes were cleaned three times with hexane through ultrasonication and baked at 80 °C for 12 hours before experiments. Prior to sample measurements, five matrix spiking experiments were conducted for DEHP; and the recovery efficiencies were between 72% and 102%. Additionally, all samples were spiked with three surrogate chemicals (i.e., dibenzyl phthalate, diphenyl isophthalate, and diphenyl phthalate) to determine analytical recovery efficiencies. The average recovery for the surrogates in all samples was $85.3\% \pm 7.3\%$. Each sample was weighed three times using a five-digit microbalance and mean weight was calculated. To check drift of the GC system, calibration standards were run after every six samples, and results were accepted only when the standards varied less than 10%. A calibration curve, which has a $R^2 > 0.9999$, was constructed using six different concentrations in the range of 1-500 ng/ μ L for DEHP. Lab blanks (i.e., empty petri dish) were collected and treated identically to the samples (except that no dust was collected). None of the target compounds were detected in these laboratory blanks. To avoid contamination, all glassware and metal ware used for sample collection and analysis were ultrasonically cleaned in methanol for 20 minutes, rinsed with methanol three times, and then baked at 80 °C in an oven prior to use.

Following the incubation period, the dust placed on the source material within the petri dish was removed gently with a quantitative filter and transferred to an 8-ml glass vial. The vial was weighed before and after placing the dust to determine the mass of dust collected in each petri dish. The collected dust samples were ultrasonically extracted three times for 30 min each with n-hexane (Hexane (mixture of isomers) $\geq 98.5\%$ ACS, VWR Chemicals BDH). The volume of the extract was concentrated to approximately 500 μ L by gently bubbling the samples with

ultrapure nitrogen. DEHP concentrations in the extract was then analyzed using a GC-FID system (Agilent 7890A) equipped with a DB-5ht column (30 m, 0.25 mm, 0.1 μ m). The GC-FID system was operated using a splitless injection at 275°C. The temperature program was set at 120°C and held for 2 min, ramped up 12°C/min for 15 min and held 3 min, and then ramped up 20°C/min for 2 min and held 2 min. The detector was set at 320°C. Helium was used as the carrier gas with a constant column flow set at 1.2 ml/min. A calibration standard was run prior to each GC injection, and the variance was always below 10% for all injections.

Sterile cotton swabs were used to collect dust from remaining petri dishes for microbial analysis. A dust sample was collected at time point 0, and three experimental replicates were collected each at day 1, 7, and 14, using sterile swabs. A sterile swab and petri dish, and no-template extraction and PCR samples, were included as negative controls. Plates were stored at 20°C until DNA extraction was performed; swab tips were added directly to bead tubes for extraction. Nucleic acids were extracted using the MoBio PowerSoil DNA Extraction Kit (MoBio, Carlsbad, CA, USA) following kit instructions.

The V3 and V4 (319F-806R) regions of the 16S rRNA gene were PCR-amplified following a previously described protocol (Kembel et al., 2014), and amplicons were purified with a bead-based clean-up using Mag-Bind RxnPure Plus (Omega Bio-tek, Norcross, GA, USA). Cleaned DNA was quantified using Quant-iT dsDNA assay kit, and pooled with equal concentrations of amplicons for Illumina Miseq ver. 4 paired-end sequencing using a 250-cycle kit at the University of Oregon Genomics Core (Eugene, OR). Raw data are available through the National Center for Biotechnology Information (NCBI) Sequence Read Archive (SRA) under BioProject accession PRJNA541585.

Sequence data were processed using the DADA2 pipeline (“DADA2 Pipeline Tutorial (1.4)”); Callahan et al., 2016) in R statistical program (R Core Team, 2018). During filtering, sequences were trimmed by 10 bases at the start position, and 10 and 20 bases from the end position for forward and reverse reads, respectively. No ambiguous bases were permitted, and the maximum expected errors (maxEE) were 2 and 3 for forward and reverse, respectively. DNA matching the phiX sequencing controls was removed. Error rates were learned on 1×10^6 randomly-selected dereplicated reads, and used to identify sequence variants (SVs), after which forward and reverse reads were merged into contiguous sequences. Bimeras (two-parent chimeras) were identified and removed, and taxonomy assigned using the Silva NR ver 132 database (Yilmaz et al., 2014). Laboratory negative controls were used to remove identical sequence variants from experimental samples, using adapted code (Ishaq, 2017). Two experimental controls did not successfully generate sequences, and were included in certain figures but removed from the analysis prior to statistical comparison. Samples were rarefied to 6,732 sequences per sample, and statistical comparison was performed with R packages phyloseq (McMurdie & Holmes, 2013), vegan (Oksanen, 2018), and rfPermute (Breiman et al., 2015; Archer, 2019).

Results and Discussion

Home dust was incubated on polyvinyl chloride flooring, from which the dust was seeded with di-2-ethylhexyl phthalate (DEHP) over 14 days, averaging 1,041 $\mu\text{g/g}$ on day 1, 4,547 $\mu\text{g/g}$ on day 7, and 10,066 $\mu\text{g/g}$ on day 14 (Figure S1).

Bacterial communities in dust samples exhibited reduced richness on day 14 (Figure 1A; $\text{lm}, t = -2.560, p = 0.043$, adjusted $R^2 = 0.366$) and was negatively correlated with DEHP concentration, although significance was only trending ($\text{lm}, t = -1.939, p = 0.094$, adjusted $R^2 = 0.256$). Shannon diversity was not affected by Day or DEHP concentration (Figure 1B; Mann-Whitney, $p > 0.05$). Control dust, which was incubated on glass plates, exhibited variable richness and diversity, but days 0 and 7 had such low richness (Figure S2) as to be removed during rarefaction. As no replicates were included in microbial analysis it is unclear whether sample days 0 and 7 failed sequencing procedures or accurately had lower bacterial richness.

Dust samples clustered significantly by Day (Figure 2; PerMANOVA, unweighted Jaccard, $f = 1.147, R^2 = 0.141, p = 0.016$). A number of bacterial taxa were associated with DEHP concentration (Figure 3). In particular, two *Bacillus* sp., *Corynebacterium jeddahense*, a *Streptococcus* sp., and a *Peptoniphilus* sp. were relatively more abundant at low concentrations, while multiple *Sphingomonas* sp, a *Chryseobacterium* sp., and a member of the Enterobacteriaceae family were relatively more abundant at higher concentrations of DEHP. Several bacterial taxa were significantly associated with control samples with no DEHP (Figure 3B), only some of which showed a decline in relative abundance between days 1 and 14. Phylum-level taxonomy for all samples is provided in Figure S3.

Bacterial community richness declined over the 14 day study period, and while only trending towards being affected by DEHP, this may reflect low statistical power. And while DEHP concentration is inherently linked to Day, we speculate that the marginal effect observed here is due to SVOC concentration. Bacteria are known to survive in dust even after 90 days with no input of resources, and in particular, when incubated in the dark (Fahimipour et al., 2018). In the present study, only two controls had sufficient data for analysis, but the day 14

sample did have higher richness than the control day 1. Additionally, the bacterial taxa which were significantly associated with the controls showed similar relative abundance between the two time points. We posit that the bacterial community in our control samples remained stable during the experiment.

However, a more thorough investigation is needed to understand the dynamic here, including additional time points, replication in the controls, and culture assays to determine if bacteria which changed in relative abundance indeed increase or decrease when incubated with DEHP. It is also possible that DEHP itself had no direct effect, but that the degradation or utilization of DEHP by certain taxa created byproducts or competitive inhibitors which affected the microbial community. While preliminary, these results represent the first investigation into the effect of DEHP on an entire bacterial community in dust. Despite the recent scientific advances in indoor chemistry and indoor microbiology, respectively, very little work has been done to elucidate the interaction between chemicals in the built environment and the microorganisms found there.

Semi-volatile organic compounds (SVOCs) are commonly incorporated into manufactured products for a variety of purposes, but in building materials they may leach and accumulate in dust on surfaces, adsorb to other materials, or react to form a variety of chemical byproducts (Lucattini et al., 2018; Bi et al., 2018). DEHP readily adsorbs and desorbs from suspended particles, i.e. dust, and can be readily transported this way (Xu et al., 2009). Moreover, in addition to facilitating particle transport, increasing air exchange only increases DEHP emission rates from vinyl flooring (Xu et al., 2009). Thus, DEHP could potentially affect bacterial communities not adjacent to vinyl sources. The built environment is known to host

lower microbial diversity and biomass than natural environments (Gibbons, 2016; NASEM, 2017), and the presence of DEHP or other SVOCs may be contributing to this paucity.

References

- Archer E. 2019. *rfPermute*.
- Bi C, Liang Y, Xu Y. 2015. Fate and Transport of Phthalates in Indoor Environments and the Influence of Temperature: A Case Study in a Test House. *Environmental science & technology* 49:9674–9681.
- Bi C, Maestre JP, Li H, Zhang G, Givehchi R, Mahdavi A, Kinney KA, Siegel J, Horner SD, Xu Y. 2018. Phthalates and organophosphates in settled dust and HVAC filter dust of U.S. low-income homes: Association with season, building characteristics, and childhood asthma. *Environment international* 121:916–930.
- Bope A, Haines SR, Hegarty B, Weschler CJ, Peccia J, Dannemiller KC. 2019. Degradation of phthalate esters in floor dust at elevated relative humidity. *Environmental science. Processes & impacts*:1268–1279.
- Breiman L, Cutler A, Liaw A, Wiener M. 2015. *Breiman and Cutler’s Random Forests for Classification and Regression*.
- Brundrett J, Orach J, Sonderegger L, Yan T. 2018. Investigating the effects of bis (2-ethylhexyl) phthalate (DEHP) on the growth of *Escherichia coli* strain BL21. *Journal of Experimental Microbiology and Immunology* 22:1–9.
- Callahan BJ, Sankaran K, Fukuyama JA, McMurdie PJ, Holmes SP. 2016. Bioconductor Workflow for Microbiome Data Analysis: from raw reads to community analyses. *F1000Research* 5:1492.

254 Chao WL, Cheng CY. 2007. Effect of introduced phthalate-degrading bacteria on the diversity of
255 indigenous bacterial communities during di-(2-ethylhexyl) phthalate (DEHP) degradation in
256 a soil microcosm. *Chemosphere* 67:482–488.

257 DADA2 Pipeline Tutorial (1.4). Available at <https://benjjneb.github.io/dada2/tutorial.html>
258 (accessed August 24, 2017).

259 Fahimipour AK, Hartmann EM, Siemens A, Kline J, Levin DA, Wilson H, Betancourt-Roman
260 CM, Brown GZ, Fretz M, Northcutt D, Siemens KN, Huttenhower C, Green JL, Van Den
261 Wymelenberg K. 2018. Daylight exposure modulates bacterial communities associated with
262 household dust. *Microbiome* 6:175.

263 Gibbons SM. 2016. The Built Environment Is a Microbial Wasteland. *mSystems* 11:e00033–16.

264 Ishaq SL. 2017. *Phyloseq_dealing_with_neg_controls_Ishaq_example.R*. Github.

265 Just AC, Miller RL, Perzanowski MS, Rundle AG, Chen Q, Jung KH, Hoepner L, Camann DE,
266 Calafat AM, Perera FP, Whyatt RM. 2015. Vinyl flooring in the home is associated with
267 children’s airborne butylbenzyl phthalate and urinary metabolite concentrations. *Journal of*
268 *exposure science & environmental epidemiology* 25:574–579.

269 Kembel SW, Meadow JF, O’Connor TK, Mhuireach G, Northcutt D, Kline J, Moriyama M,
270 Brown GZ, Bohannan BJM, Green JL. 2014. Architectural design drives the biogeography
271 of indoor bacterial communities. *PloS one* 9:e87093.

272 Liang D-W, Zhang T, Fang HHP, He J. 2008. Phthalates biodegradation in the environment.
273 *Applied microbiology and biotechnology* 80:183–198.

274 Lotfy MM, Hassan HM, Hetta MH, El-Gendy AO, Mohammed R. 2018. Di-(2-ethylhexyl)
275 Phthalate, a major bioactive metabolite with antimicrobial and cytotoxic activity isolated
276 from River Nile derived fungus *Aspergillus awamori*. *Beni-Suef University Journal of Basic*

277 *and Applied Sciences* 7:263–269.

278 Lucattini L, Poma G, Covaci A, de Boer J, Lamoree MH, Leonards PEG. 2018. A review of

279 semi-volatile organic compounds (SVOCs) in the indoor environment: occurrence in

280 consumer products, indoor air and dust. *Chemosphere* 201:466–482.

281 McMurdie PJ, Holmes S. 2013. phyloseq: an R package for reproducible interactive analysis and

282 graphics of microbiome census data. *PloS one* 8:e61217.

283 NASEM. 2017. *Microbiomes of the Built Environment*. Washington, D.C.: National Academies

284 Press.

285 Oksanen J. 2018. *Vegan: Ecological Diversity*.

286 Ortiz A, Sansinenea E. 2018. Di-2-ethylhexylphthalate May Be a Natural Product, Rather than a

287 Pollutant. *Journal of chemistry and chemical engineering* 2018. DOI:

288 10.1155/2018/6040814.

289 R Core Team. 2018. *R: A language and environment for statistical computing*. R Foundation for

290 Statistical Computing.

291 Rowdhwal SSS, Chen J. 2018. Toxic Effects of Di-2-ethylhexyl Phthalate: An Overview.

292 *BioMed research international* 2018:1750368.

293 Vamsee-Krishna C, Phale PS. 2008. Bacterial degradation of phthalate isomers and their esters.

294 *Indian journal of microbiology* 48:19–34.

295 Wang L, Gong M, Xu Y, Zhang Y. 2017. Phthalates in dust collected from various indoor

296 environments in Beijing, China and resulting non-dietary human exposure. *Building and*

297 *environment* 124:315–322.

298 Wang H, Zhou Y, Tang C, He Y, Wu J, Chen Y, Jiang Q. 2013. Urinary phthalate metabolites

299 are associated with body mass index and waist circumference in Chinese school children.

PloS one 8:e56800.

Weschler CJ, Carslaw N. 2018. Indoor Chemistry. *Environmental science & technology* 52:2419–2428.

Wu Y, Cox SS, Xu Y, Liang Y, Won D, Liu X, Clausen PA, Rosell L, Benning JL, Zhang Y, Little JC. 2016. A reference method for measuring emissions of SVOCs in small chambers. *Building and environment* 95:126–132.

Xu Y, Hubal EAC, Clausen PA, Little JC. 2009. Predicting residential exposure to phthalate plasticizer emitted from vinyl flooring: a mechanistic analysis. *Environmental science & technology* 43:2374–2380.

Yang T, Ren L, Jia Y, Fan S, Wang J, Wang J, Nahurira R, Wang H, Yan Y. 2018. Biodegradation of Di-(2-ethylhexyl) Phthalate by *Rhodococcus ruber* YC-YT1 in Contaminated Water and Soil. *International journal of environmental research and public health* 15. DOI: 10.3390/ijerph15050964.

Yilmaz P, Parfrey LW, Yarza P, Gerken J, Pruesse E, Quast C, Schweer T, Peplies J, Ludwig W, Glöckner FO. 2014. The SILVA and “All-species Living Tree Project (LTP)” taxonomic frameworks. *Nucleic acids research* 42:D643–8.

Zhu F, Zhu C, Doyle E, Liu H, Zhou D, Gao J. 2018. Fate of di (2-ethylhexyl) phthalate in different soils and associated bacterial community changes. *The Science of the total environment* 637-638:460–469.

Figure Legends

Figure 1 Observed richness (SVs) and Shannon Diversity for bacteria in dust, which was incubated on polyvinyl chloride flooring and thereby accumulated di-2-ethylhexyl phthalate, over 14 days. Data are rarefied and quality-filtered.

Figure 2 Non-metric multidimensional scaling plot of bacterial community membership in dust, which was incubated on polyvinyl chloride flooring and thereby accumulated di-2-ethylhexyl phthalate, over 14 days. Lowest stress = 0.067.

Figure 3 Bacterial taxa significantly associated with di-2-ethylhexyl phthalate in dust (A), compared to controls (B). Panels are faceted by Day (1, 7, 14). Only significant features are shown (rfpermute, $p < 0.05$).

Supplemental Material

Figure S1 Di-2-ethylhexyl phthalate concentration in experimental samples of dust incubated on polyvinyl chloride floor over 14 days.

Figure S2 Observed richness (SVs) and Shannon Diversity for all rarefied, quality-filtered sequence data, for bacterial communities in dust, which was incubated on polyvinyl chloride flooring and thereby accumulated di-2-ethylhexyl phthalate, over 14 days.

Figure S3 Phylum-level taxonomy of bacterial communities in dust, which was incubated on polyvinyl chloride flooring and thereby accumulated di-2-ethylhexyl phthalate, over 14 days.

Figure 1

Observed richness (SVs) and Shannon Diversity for bacteria in dust, which was incubated on polyvinyl chloride flooring and thereby accumulated di-2-ethylhexyl phthalate, over 14 days.

Data are rarefied and quality-filtered.

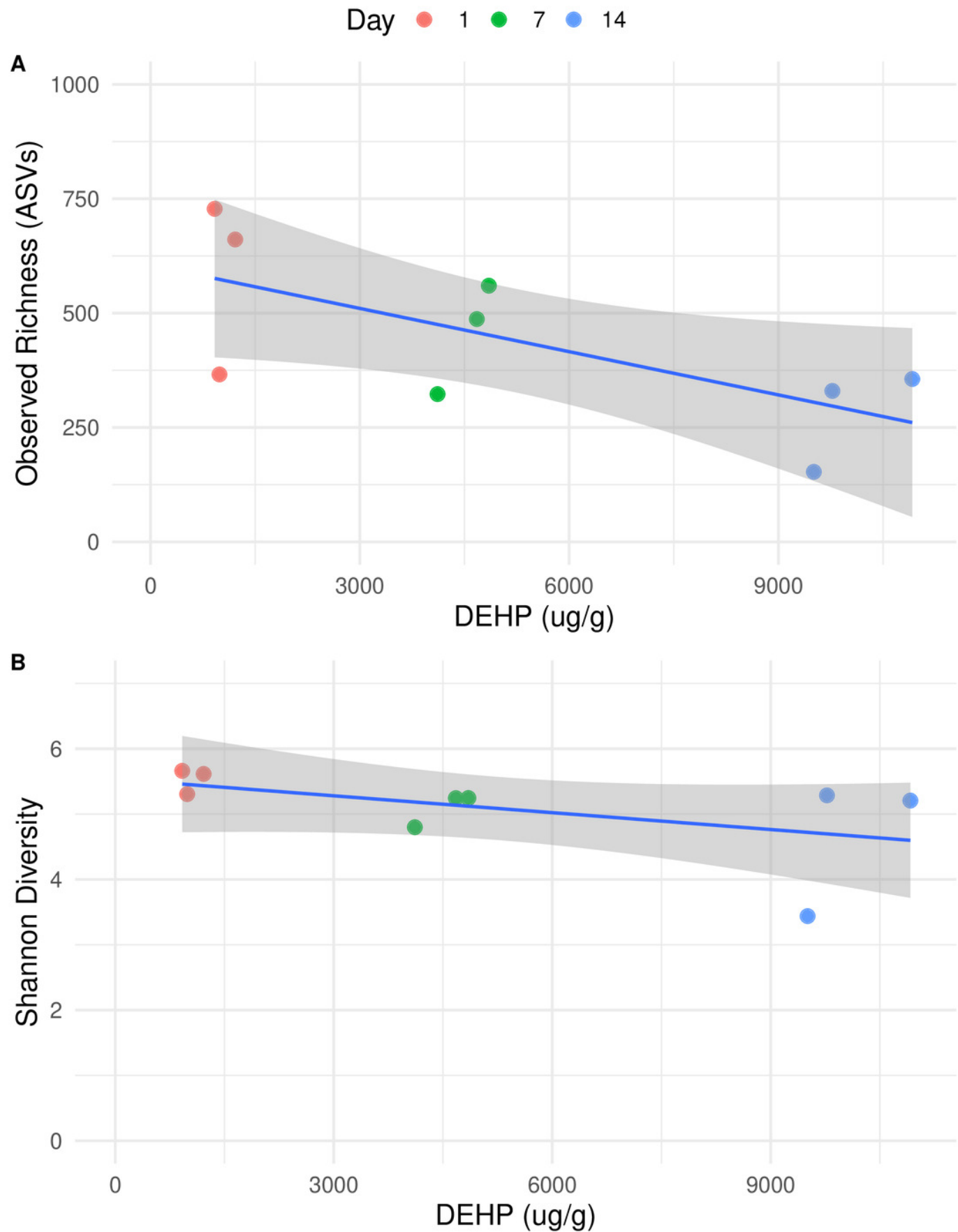


Figure 2

Non-metric multidimensional scaling plot of bacterial community membership in dust, which was incubated on polyvinyl chloride flooring and thereby accumulated di-2-ethylhexyl phthalate, over 14 days.

Lowest stress = 0.067.

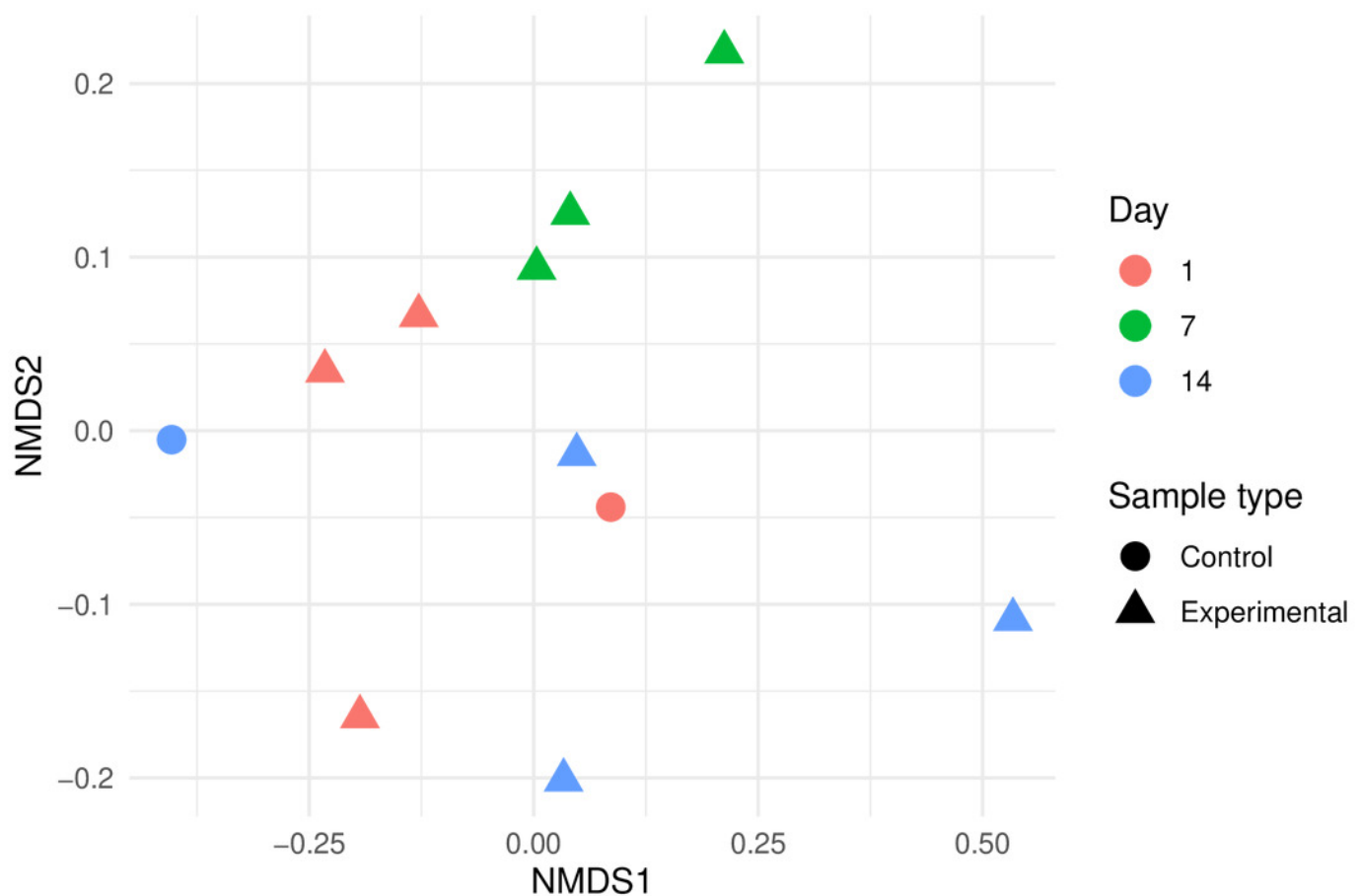


Figure 3

Bacterial taxa significantly associated with di-2-ethylhexyl phthalate in dust (A), compared to controls (B).

Panels are faceted by Day (1, 7, 14). Only significant features are shown (rfpermute, $p < 0.05$).

