

Dynamics of a methanol-fed marine denitrifying biofilm: 1- Impact of environmental changes on the denitrification and the co-occurrence of *Methylophaga nitratireducenticrescens* and *Hyphomicrobium nitrativorans*

Geneviève Payette¹, Valérie Geoffroy², Christine Martineau³, Richard Villemur^{Corresp. 1}

¹ INRS-Centre Armand-Frappier Santé et Biotechnologie, Laval, Québec, Canada

² Lallemand, Montréal, Québec, Canada

³ Laurentian Forestry Centre, Québec, Québec, Canada

Corresponding Author: Richard Villemur
Email address: richard.villemur@iaf.inrs.ca

Background. The biofilm of a methanol-fed denitrification system that treated a marine effluent is composed of multi-species microorganisms, among which *Hyphomicrobium nitrativorans* strain NL23 and *Methylophaga nitratireducenticrescens* strain JAM1 are the principal bacteria involved in the denitrifying activities. Here, we report the capacity of the denitrifying biofilm to sustain environmental changes, and the impact of these changes on the co-occurrence of *H. nitrativorans* and *M. nitratireducenticrescens*.

Methods. In a first set of assays, the original biofilm (OB) was cultivated in an artificial seawater (ASW) medium under anoxic conditions to colonize new carriers. The new formed biofilm was then subjected to short exposures (1-3 days) of a range of NaCl, methanol, nitrate (NO_3^-) and nitrite (NO_2^-) concentrations, and to different pHs and temperatures. In a second set of assays, the OB was cultivated in ASW medium for five weeks with i) a range of NaCl concentrations, ii) four combinations of NO_3^- /methanol concentrations and temperatures, iii) NO_2^- , and iv) under oxic conditions. Finally, the OB was cultivated for five weeks in the commercial Instant Ocean (IO) seawater. The growth of the biofilm and the dynamics of NO_3^- and NO_2^- were determined. The levels of *M. nitratireducenticrescens* and *H. nitrativorans* were measured by qPCR.

Results. In the first set of assays, the biofilm cultures had the capacity to sustain denitrifying activities in most of the tested conditions. Inhibition occurred when they were exposed to high pH (10) or to high methanol concentration (1.5%). In the second set of assays, the highest specific denitrification rates occurred with the biofilm cultures cultivated at 64.3 mM NO_3^- and 0.45% methanol, and at 30°C. Poor biofilm development occurred with the biofilm cultures cultivated at 5% and 8% NaCl. In all biofilm cultures cultivated in ASW at 2.75% NaCl, *H. nitrativorans* strain NL23 decreased by three orders of magnitude in concentrations compared to that found in OB. This decrease coincided with the increase of the same magnitude of a subpopulation of *M. nitratireducenticrescens* (strain GP59 as representative). In the biofilm cultures cultivated at low NaCl concentrations (0% to 1.0%), persistence of *H. nitrativorans* strain NL23 was observed, with the gradual increase in concentrations of *M. nitratireducenticrescens* strain GP59. High levels of *H. nitrativorans* strain NL23 were found in the IO biofilm cultures. The concentrations of *M. nitratireducenticrescens* strain JAM1 were lower in most of the biofilms cultures than in OB.

Conclusions. These results demonstrate the plasticity of the marine methylotrophic denitrifying biofilm in

adapting to different environmental changes. The NaCl concentration is a crucial factor in the dynamics of *H. nitrativorans* strain NL23, for which growth was impaired above 1% NaCl in the ASW-based biofilm cultures in favor of *M. nitratireducentescens* strain GP59.

1 Dynamics of a methanol-fed marine denitrifying biofilm: 1-Impact of
2 environmental changes on the denitrification and the co-occurrence of
3 *Methylophaga nitratreducenticrescens* and *Hyphomicrobium nitrativorans*

4

5 Geneviève Payette¹, Valérie Geoffroy^{1,2}, Christine Martineau^{1,3}, Richard Villemur¹

6

7 ¹INRS-Centre Armand-Frappier Santé et Biotechnologie, Laval, Québec, Canada

8 ²Present address: Lallemand, Montreal, Québec, Canada

9 ³Present address: Laurentian Forestry Centre, Québec, Québec, Canada

10

11 **Corresponding author:**

12 Richard Villemur

13 INRS-Institut Armand-Frappier

14 531 Boulevard des Prairies

15 Laval, QC, Canada

16 Laval, Québec, Canada, H7V 1B7

17 Email address: richard.villemur@iaf.inrs.ca

18 **Abstract**

19

20 **Background.** The biofilm of a methanol-fed denitrification system that treated a marine effluent
21 is composed of multi-species microorganisms, among which *Hyphomicrobium nitratorans*
22 strain NL23 and *Methylophaga nitratireducenticrescens* strain JAM1 are the principal bacteria
23 involved in the denitrifying activities. Here, we report the capacity of the denitrifying biofilm to
24 sustain environmental changes, and the impact of these changes on the co-occurrence of *H.*
25 *nitratorans* and *M. nitratireducenticrescens*.

26

27 **Methods.** In a first set of assays, the original biofilm (OB) was cultivated in an artificial
28 seawater (ASW) medium under anoxic conditions to colonize new carriers. The new formed
29 biofilm was then subjected to short exposures (1-3 days) of a range of NaCl, methanol, nitrate
30 (NO_3^-) and nitrite (NO_2^-) concentrations, and to different pHs and temperatures. In a second set
31 of assays, the OB was cultivated in ASW medium for five weeks with i) a range of NaCl
32 concentrations, ii) four combinations of NO_3^- /methanol concentrations and temperatures, iii)
33 NO_2^- , and iv) under oxic conditions. Finally, the OB was cultivated for five weeks in the
34 commercial Instant Ocean (IO) seawater. The growth of the biofilm and the dynamics of NO_3^-
35 and NO_2^- were determined. The levels of *M. nitratireducenticrescens* and *H. nitratorans* were
36 measured by qPCR.

37

38 **Results.** In the first set of assays, the biofilm cultures had the capacity to sustain denitrifying
39 activities in most of the tested conditions. Inhibition occurred when they were exposed to high
40 pH (10) or to high methanol concentration (1.5%). In the second set of assays, the highest
41 specific denitrification rates occurred with the biofilm cultures cultivated at 64.3 mM NO_3^- and
42 0.45% methanol, and at 30°C. Poor biofilm development occurred with the biofilm cultures
43 cultivated at 5% and 8% NaCl. In all biofilm cultures cultivated in ASW at 2.75% NaCl, *H.*
44 *nitratorans* strain NL23 decreased by three orders of magnitude in concentrations compared to
45 that found in OB. This decrease coincided with the increase of the same magnitude of a
46 subpopulation of *M. nitratireducenticrescens* (strain GP59 as representative). In the biofilm
47 cultures cultivated at low NaCl concentrations (0% to 1.0%), persistence of *H. nitratorans*
48 strain NL23 was observed, with the gradual increase in concentrations of *M.*

49 *nitratireducenticrescens* strain GP59. High levels of *H. nitrativorans* strain NL23 were found in
50 the IO biofilm cultures. The concentrations of *M. nitratireducenticrescens* strain JAM1 were
51 lower in most of the biofilms cultures than in OB.

52

53 **Conclusions.** These results demonstrate the plasticity of the marine methylotrophic denitrifying
54 biofilm in adapting to different environmental changes. The NaCl concentration is a crucial
55 factor in the dynamics of *H. nitrativorans* strain NL23, for which growth was impaired above
56 1% NaCl in the ASW-based biofilm cultures in favor of *M. nitratireducenticrescens* strain GP59.

57 Introduction

58 Most naturally-occurring microbial biofilms, such as those encountered in bioremediation
59 processes, are composed of multiple microbial species. Studying such complex biofilms is a
60 challenge, as each species can influence the biofilm development. The biofilm microbial
61 community inside a bioremediation process adapts to the prescribed operating conditions and
62 shapes the efficiency of the bioprocess to degrade the pollutant(s) (Cabrol & Malhautier 2011;
63 Roder et al. 2016; Salta et al. 2013; Tan et al. 2017). The mechanisms of how a microbial
64 population in a biofilm adapts to changes however are poorly understood.

65 Denitrification takes place in bacterial cells where N oxides serve as terminal electron
66 acceptor instead of oxygen (O₂) for energy production when oxygen depletion occurs, leading to
67 the production of gaseous nitrogen (N₂). Four sequential reactions are strictly required for the
68 reduction of NO₃⁻ to gaseous nitrogen, via nitrite (NO₂⁻), nitric oxide and nitrous oxide, and each
69 of these reactions is catalyzed by different enzymes, namely NO₃⁻ reductases (Nar and Nap),
70 NO₂⁻ reductases (Nir), nitric oxide reductases (Nor) and nitrous oxide reductases (Nos) (Kraft et
71 al. 2011; Philippot & Hojberg 1999; Richardson et al. 2001). These biological activities have
72 been used with success to remove NO₃⁻ in different wastewater treatment processes
73 (Tchobanoglous et al. 2003). The Montreal Biodome, a natural science museum, operated a
74 continuous, fluidized-bed methanol-fed denitrification reactor to remove NO₃⁻ that accumulated
75 (up to 200 mg NO₃⁻-N/L) in the 3 million-L seawater aquarium that contains fish, birds and
76 invertebrate. The fluidized carriers in the denitrification reactor were colonized by naturally
77 occurring multispecies bacteria to generate a marine methylotrophic denitrifying biofilm to be
78 composed of 15-20 bacterial species (Labbé et al. 2003). Fluorescence *in situ* hybridization
79 experiments on the biofilm showed that the methylotrophic bacteria *Methylophaga* spp. and
80 *Hyphomicrobium* spp. accounted for 60 to 80% of the bacterial community (Labbé et al. 2007).

81 *Hyphomicrobium* spp. are methylotrophic bacteria that are ubiquitous in the environment
82 (Gliesche et al. 2005). They have also been found in significant levels in several methanol-fed
83 denitrification systems treating municipal or industrial wastewaters or a seawater aquarium, and
84 they occurred often with other denitrifying bacteria such as *Paracoccus* spp., *Methylophilales* or
85 *Methyloversatilis* spp. Their presence correlates with optimal performance of bioprocess
86 denitrifying activities (Baytshok et al. 2008; CatalanSakairi et al. 1997; Ginige et al. 2004;
87 Layton et al. 2000; Lemmer et al. 1997; Rissanen et al. 2017; Wang et al. 2014). *Methylophaga*

88 spp. are methylotrophic bacteria isolated from saline environments (Boden 2012; Bowman
89 2005). They have been found in association with diatoms, phytoplankton blooms, marine algae,
90 which are known to generate C1 carbons (Bertrand et al. 2015; Landa et al. 2018; Li et al. 2007).
91 The importance of the co-occurrence of *Methylophaga* spp. and *Hyphomicrobium* spp. has been
92 shown in two other methanol-fed denitrification systems treating saline effluents (Osaka et al.
93 2008; Rissanen et al. 2016). Therefore, understanding how collaboration between these two taxa
94 could benefit in optimizing of denitrification systems.

95 Two bacterial strains representatives of *Methylophaga* spp. and *Hyphomicrobium* spp.
96 were isolated from the Biodome denitrifying biofilm. The first one, *Methylophaga*
97 *nitratireducenticrescens* strain JAM1, is capable of growing in pure cultures under anoxic
98 conditions by reducing NO_3^- to NO_2^- , which accumulates in the medium (Auclair et al. 2010;
99 Villeneuve et al. 2013). It was later shown to be able to reduce NO and N_2O to N_2 (Mauffrey et
100 al. 2017). The second bacterium, *Hyphomicrobium nitrativorans* strain NL23, is capable of
101 complete denitrification from NO_3^- to N_2 (Martineau et al. 2013b). These two strains are the
102 main bacteria responsible of the dynamics of denitrification in the biofilm. The genomes of *M.*
103 *nitratireducenticrescens* strain JAM1 and *H. nitrativorans* strain NL23 have been sequenced
104 previously (Martineau et al. 2013a; Villeneuve et al. 2012). Strain JAM1 contains two Nar-type
105 NO_3^- reductase gene clusters (Nar1 and Nar2). Both Nar systems contribute to the reduction of
106 NO_3^- to NO_2^- during growth of strain JAM1 (Mauffrey et al. 2015). Gene clusters encoding two
107 Nor and one Nos systems are present and expressed in strain JAM1. Their presence correlates
108 with the reduction of NO and N_2O by strain JAM1 (Mauffrey et al. 2017). A dissimilatory NO-
109 forming NO_2^- reductase gene (*nirS* or *nirK*) is absent, which correlates with accumulation of
110 NO_2^- in the culture medium during NO_3^- reduction. The genome of *H. nitrativorans* strain NL23
111 contains the operons that encode for the four different nitrogen oxide reductases, among which a
112 Nap-type NO_3^- reductase (Martineau et al. 2015).

113 We have initiated a study with the aim of assessing the performance of the Biodome
114 denitrifying biofilm when exposed to environmental changes. The original biofilm (OB) taken
115 from the Biodome denitrification system was cultivated in a homemade artificial seawater
116 (ASW) medium, instead of the commercial Instant Ocean (IO) used by the Biodome, under batch
117 mode and anoxic conditions at laboratory scale and exposed to a range of specific physico-
118 chemical parameters. Thus, the objectives of this study were to determine the impact of these

119 changes: 1) on the denitrification performance of the biofilm; 2) on the dynamics of the co-
120 occurrence of the *H. nitrativorans* and *M. nitratireducenticrescens* in the biofilm; and 3) on the
121 overall microbial community. The fourth objective of the study was to determine whether
122 denitrifying bacteria other than *H. nitrativorans* strain NL23 and *M. nitratireducenticrescens*
123 strain JAM1 are present in the biofilm.

124 In the first part of this study reported by Geoffroy *et al.* (2018), we showed that by
125 cultivating the OB in the ASW medium, important decreases in the concentration of *H.*
126 *nitrativorans* strain NL23 occurred without the loss of denitrifying activities. This decrease
127 concurred with the occurrence of the new denitrifying bacteria *M. nitratireducenticrescens* strain
128 GP59. The genome of strain GP59 has been sequenced and is highly similar to that of strain
129 JAM1. Both strains have the same denitrification genes except that the genome of strain GP59
130 contains a *nirK* gene, which is missing in strain JAM1 (Geoffroy *et al.* 2018).

131 Here, we report the results for the first two objectives of our study. First, the OB was
132 cultivated in ASW medium under anoxic conditions to derive the reference biofilm cultures.
133 These cultures were then exposed for short periods (1 to 3 days) to different physico-chemical
134 parameters. Such parameters included a range of NaCl, NO₃⁻, NO₂⁻ and methanol concentrations,
135 and varying pH and temperatures. These parameters were chosen as possible factors that could
136 affect a denitrification reactor. The OB was then cultivated for a long period (five weeks) under
137 specific conditions based on the short exposure assays to develop new biofilm cultures. The
138 impact of these changes on the denitrifying activities and on the population dynamics of *M.*
139 *nitratireducenticrescens* strain JAM1 and strain GP59, and *H. nitrativorans* strain NL23 were
140 measured. Results from the third and fourth objectives are presented by Villemur *et al.* (2019),
141 where we have looked at the overall microbial community of our biofilm cultures and identified
142 new denitrifying bacteria. To our knowledge, this is the first systematic study on the co-evolution
143 of *Methylophaga* and *Hyphomicrobium* strains in a marine biofilm system.

144

145 **Material and Methods**

146 **Seawater media**

147 The artificial seawater (ASW) medium was composed of (for 1 liter solution): 27.5 g NaCl,
148 10.68 g MgCl₂*6H₂O, 2 g MgSO₄*7H₂O, 1 g KCl, 0.5 g CaCl₂, 456 μL of FeSO₄*7H₂O 4 g/L, 5
149 mL of KH₂PO₄ 51.2 g/L, 5 mL of Na₂HPO₄ 34 g/L. The Instant Ocean (IO) seawater medium

150 was bought from Aquarium systems (Mentor, OH, USA) and dissolved at 30 g/L. Estimation of
151 its composition and comparison with the ASW medium is provided in Table S1. Both media
152 were supplemented with 1 mL of trace elements ($\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$ 0.9 g/L, $\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$ 0.03 g/L
153 and $\text{MnSO}_4 \cdot \text{H}_2\text{O}$ 0.2 g/L) and with NaNO_3 (Fisher Scientific Canada, Ottawa, ON, Canada) at
154 various concentrations, depending on the experiment (Tables 1 and 2). The pH was adjusted
155 (NaOH) at 8.0 before autoclaving. Filter-sterilized methanol (Fisher Scientific) was then added
156 as a carbon source to support bacterial growth, also at various concentrations depending on the
157 experiment (Tables 1 and 2). The sterile media were distributed (60 ml) in sterile serologic vials,
158 which included twenty « Bioflow 9 mm » carriers (Rauschert, Steinwiessen, Allemagne). Prior to
159 use, these carriers were washed with HCl 10% (v/v) for 3 h, rinsed with water and autoclaved.
160 The vials were purged of oxygen for 10 min with nitrogen gas (Praxair, Mississauga, ON,
161 Canada) and sealed with sterile septum caps.

162

163 **The reference biofilm cultures (Ref300N-23C)**

164 The carriers (Bioflow 9) containing the denitrifying biofilm (here named the Original
165 Biofilm; OB) were taken from the denitrification reactor of the Montreal Biodome and frozen at
166 -20°C in seawater with 20% glycerol (Laurin et al. 2006) until use. Biomass from several carriers
167 were thawed, scraped, weighted and dispersed in the ASW at 0.08 g (wet weight)/mL. The
168 biomass (5 mL/vial, 0.4 g of biofilm) was then distributed with a syringe and an 18G1½ needle
169 into vials containing twenty free carriers and 60 mL ASW and supplemented with 300 mg NO_3^- -
170 N/L (21.4 mM NO_3^-) and 0.15% (v/v) methanol (C/N=1.5). The vials were incubated at 23°C
171 and shaken at 100 rpm (orbital shaker).

172 On average once a week, the twenty carriers were taken out of the vial, gently washed to
173 remove the excess medium and the free bacteria, transferred into a new vial containing fresh
174 anaerobic medium, and then incubated again under the same conditions (Fig. S1). Samples were
175 taken each one or two days to measure the concentrations of NO_3^- and NO_2^- in the growth
176 medium. The residual suspended biomass was taken for DNA extraction. Methanol and NaNO_3
177 were added when needed if NO_3^- was completely depleted during the week. During the fifth
178 carrier-transfer cultures, multiple samples (500 μL) were collected at regular interval for 1-3
179 days to determine the NO_3^- and NO_2^- concentrations. After this monitoring, total biomass in a
180 whole vial was determined by protein quantification.

181

182 Short exposure of the reference biofilm cultures to physico-chemical changes

183 The carriers containing the reference biofilm cultures were used as starting material to test
184 the impact on the denitrification performance of the biofilm by varying specific physico-
185 chemical parameters (Table 1). Figure S2 describes in detail the protocol. To perform these
186 assays, fifteen carriers from the reference biofilm cultures were distributed into vials in the
187 prescribed conditions described in Table 1. For the assays that were performed with different
188 pHs and temperatures and different NaCl concentrations, the vials were incubated for two days
189 for the biofilm to adapt to the new culture conditions, and then the carriers were transferred into
190 their respective fresh medium (Fig. S2). NO_3^- and NO_2^- concentrations in the vials were then
191 monitored by collecting 500 μL medium samples at regular intervals for 1 to 3 days. Biomass
192 was taken at the end of this monitoring to extract DNA and for protein quantification. All assays
193 were performed in triplicates, except those with NO_2^- .

194

195 Cultivation of the original biofilm to different culture conditions

196 The OB was taken and processed as described in *The reference biofilm cultures* subsection.
197 The formulation of the ASW medium was adjusted with the prescribed NaCl, NO_3^- , NO_2^- and
198 methanol concentrations, and pHs, and the vials were incubated at prescribed temperatures
199 (Table 2; Fig. S3). Cultivation of the OB under oxic conditions were performed as for the
200 reference biofilm cultures but in 250-mL Erlenmeyer flasks with agitation (Table 2). Culturing
201 the OB in the IO medium was performed as for the reference biofilm cultures (Table 2). All
202 biofilm cultures were shaken at 100 rpm (orbital shaker), or 200 rpm for the oxic cultures. Five
203 carrier-transfer cultures were performed, after which the NO_3^- , NO_2^- and protein concentrations
204 were determined as described in *The reference biofilm cultures* subsection.

205

206 Determination of the denitrification rates

207 Measurements of NO_3^- and NO_2^- concentrations in the biofilm cultures were performed in
208 all assays using ion chromatography as described by Mauffrey *et al.* (2017). The total biomass in
209 a vial was measured by collecting the suspended biomass and the biomass attached to the carriers
210 (Fig. S1 to S3). The amount of protein in the biomass was determined by the Quick Start™
211 Bradford Protein Assay (BioRad, Mississauga, ON, Canada). The denitrification rates were

212 calculated by the linear portion of the NO_3^- plus NO_2^- concentrations (NO_x) over time for each
213 replicate (mM h^{-1}). The specific denitrification rates were reported as the denitrification rates
214 divided by the quantity of biomass (mg protein) in a vial ($\text{mM h}^{-1} \text{mg-protein}^{-1}$). Ammonium
215 concentration was measured using the colorimetric method described by Mulvaney (1996).

216

217 **DNA-based analyses**

218 The DNA was extracted from the suspended biomass and the biofilm that was scraped
219 from the carriers as previously described (Geoffroy et al. 2018). The PCR amplification of the
220 V3 region of the 16S ribosomal RNA (rRNA) genes for denaturing gradient gel electrophoresis
221 (DGGE) experiments was performed as described (Lafortune et al. 2009) with the 341f and 534r
222 primers (Table S2). Total DNA extracted from the planktonic pure cultures of strains JAM1 and
223 NL23 was used to perform PCR amplifications of the same region of the 16S rRNA genes. The
224 resulting strain-derived amplicons were co-migrated with the biofilm cultures-derived amplicons
225 on DGGE gels to identify the DNA fragments associated to strain JAM1 and strain NL23 in
226 biofilm samples.

227 Quantitative PCR assays (SYBR green) to measure the concentrations of *M.*
228 *nitratireducenticrescens* strains JAM1/GP59 (*narG1*), *M. nitratireducenticrescens* strain GP59
229 (*nirK*), *M. nitratireducenticrescens* strain JAM1 (*tagH*) and *H. nitrativorans* strain NL23 (*napA*)
230 (Table S2) in the biofilm cultures were performed as previously described (Geoffroy et al. 2018).

231

232 **Statistics**

233 Statistical significance in the denitrification rates between different types of cultures was
234 performed with one-way ANOVA with Tukey's Multiple Comparison Test with the GraphPad
235 Prism 5.0 software.

236

237 **Results**

238 **Establishment of the reference biofilm cultures (Ref300N-23C)**

239 The original biofilm (OB) of the Biodome denitrification reactor was used to inoculate
240 vials containing new carriers to allow the development of a fresh denitrifying biofilm on the
241 carrier surface. The vials were incubated under batch-mode anoxic conditions in a homemade
242 ASW medium (2.75% NaCl) with 300 mg NO_3^- -N/L (21.4 mM NO_3^-) and 0.15% (v/v) methanol

243 (C/N=1.5) at 23°C. The choice of ASW medium was based on our ability to change the
244 composition of the medium, as opposed to the commercial seawater medium (Instant Ocean)
245 used by the Biodome. Each week for five weeks, only the carriers were transferred in fresh
246 medium (Fig. S1). The biofilm cultures cultivated under these conditions are referred as the
247 *reference biofilm cultures* and named from here as the Ref300N-23C biofilm cultures (for 300
248 mg NO₃⁻-N/L, 23°C). As its name implies, the conditions of these cultures were used as a
249 reference to measure changes in the denitrifying activities resulting from changes in physico-
250 chemical parameters in the culture medium.

251 The NO₃⁻ and NO₂⁻ concentrations were analyzed sporadically for the first four carrier-
252 transfer cultures, and showed increasing rates of denitrifying activities with subsequent transfers.
253 During the fifth carrier-transfer cultures, the concentrations of NO₃⁻ and NO₂⁻ were analyzed
254 more thoroughly. NO₃⁻ was consumed in 4 to 6 hours, with accumulation of NO₂⁻ that peaked at
255 10 mM after 4 hours. NO₂⁻ was completely consumed after 12 hours (Fig. 1A). The
256 denitrification rates were calculated at 1.69 (±0.14) mM-NO_x h⁻¹. Taking into account the
257 biomass, the specific denitrification rates corresponded to 0.0935 (±0.0036) mM-NO_x h⁻¹ mg-
258 protein⁻¹. The production of ammonia was negative in these conditions, which rules out
259 dissimilatory NO₃⁻ reduction to ammonium. Further carrier transfers did not improve the specific
260 denitrification rates.

261 The evolution of the bacterial community of the OB under these conditions was assessed
262 by PCR-DGGE profiles for each carrier-transfer cultures (Fig. 1B). These profiles were similar
263 after the third carrier-transfer cultures, suggesting stabilization of the bacterial community in
264 these cultures. The most substantial changes in these profiles compared to the OB profile is the
265 intensity of the DGGE band corresponding to *M. nitratreducentescens* strain JAM1 that was
266 much stronger in all carrier-transfer cultures, whereas the DGGE band corresponding to *H.*
267 *nitrativorans* strain NL23 was no longer visible after the second carrier-transfer cultures (Fig.
268 1B). qPCR analysis confirmed these results (Fig 1C). Strain JAM1 increased in concentration in
269 the Ref300N-23C biofilm cultures from 5.6 x 10³ *narG1* copies/ng DNA in the OB to 2.0 x 10⁵
270 *narG1* copies/ng DNA in the fifth carrier-transfer cultures. The concentrations of strain NL23
271 however dropped by three orders of magnitude from 8.7 x 10⁴ in the OB to 3.2 x 10¹ *napA*
272 copies/ng DNA in the fifth carrier-transfer cultures.

273 These results showed important changes in the bacterial community of OB occurred when

274 cultivated in the ASW medium under the batch-operating mode, with the important decrease in
275 concentration of the main denitrifying bacterium in the biofilm, *H. nitrativorans* strain NL23,
276 without the loss of denitrifying activities. We discovered later that a new *M.*
277 *nitratireducenticrescens* strain, strain GP59 with full denitrification capacity, was enriched in
278 these cultures (Geoffroy et al. 2018). qPCR assays targeting *narG1* cannot discriminate strain
279 JAM1 from strain GP59, as this gene is identical in nucleic sequences in both strains (Geoffroy
280 et al. 2018). Therefore, the concentrations of strain JAM1 in the biomass derived from *narG1*-
281 targeted qPCR assays and illustrated in Figure 1C reflects in fact the concentrations of both
282 strains.

283

284 **Impact of a short exposure to environmental changes on the Ref300N-23C biofilm cultures**

285 The colonized carriers from the Ref300N-23C biofilm cultures were exposed for a short
286 period (few hours to few days) to a range of methanol, NO_3^- , NO_2^- and NaCl concentrations, and
287 to different pHs and temperatures (Table 1, Fig. S2, Table S3). These conditions were chosen to
288 assess the capacity range of the denitrifying activities of the biofilm (methanol and NO_3^-
289 concentrations), but also its resilience to adverse conditions that could occur during the operation
290 of a bioprocess (e.g. abrupt change of pH, temperature or salt concentration).

291 Increasing the concentrations of NO_3^- and methanol (fixed C/N at 1.5) showed increases in
292 the denitrification rates compared to the original conditions of the Ref300N-23C biofilm cultures
293 (300 mg- NO_3^- -N/L with 0.15% methanol), with the highest increases observed at 600 mg- NO_3^- -
294 N/L with 0.3% methanol (58% increase), and at 900 mg- NO_3^- -N/L with 0.45% methanol (56%
295 increase) (Fig. 2). Inhibition of the denitrifying activities occurred at 3000 mg- NO_3^- -N/L. This
296 inhibition could have been caused by the methanol concentration (1.5%) in the medium.
297 Variations in methanol concentrations (0% to 0.5% with fixed concentration of NO_3^- at 300 mg-
298 NO_3^- -N/L) showed no improvement of the denitrification activities compared to the original
299 conditions (Fig. 2). As methanol is the only source of carbon in the medium, its absence (0%)
300 resulted in very weak denitrifying activities as expected. Variations in NO_3^- concentrations (fixed
301 concentration of methanol at 0.15%) showed the highest denitrification rates (74% increase) at
302 900 mg- NO_3^- -N/L (Fig. 2). Denitrifying activities were observed this time at 3000 mg- NO_3^- -N/L.
303 However, NO_3^- consumption stopped after 5 days and 50% consumption. Addition of methanol
304 (0.15%) allowed activities to resume for one day with a further 20% NO_3^- consumption.

305 Exposure of the Ref300N-23C biofilm cultures to pH 4 and 6 showed 2.7 and 2.6-fold
306 increases of the denitrification rates, respectively, compared to the original conditions (pH 8)
307 (Fig. 2). However, the pH at the end of the assays (10 to 12 hours) increased at around 8 in the
308 medium. At pH 10, the denitrifying activities were completely inhibited. For the temperature
309 assays, the highest denitrification rates were recorded at 30 and 36°C (Fig. 2), with 53% and
310 62% increases, respectively, compared to the original conditions (23°C). At 5°C, the denitrifying
311 activities were still occurring but at 4-time lower rates than at 23°C. At 15°C, the denitrification
312 rates were not significantly different from the original conditions. Exposure of the Ref300N-23C
313 biofilm cultures for 3 days in ASW with NaCl concentrations ranging from 0% to 8% showed
314 little effect on the denitrification rates (Fig. 2). Finally, presence of NO_2^- has a strong effect on
315 the denitrification rates with 11-fold and 6.6-fold decreases with 400 mg- NO_2^- -N/L and 200 mg-
316 NO_2^- -N/L, respectively in the medium, compared to the original conditions (Fig. 2). Mixture of
317 200 mg- NO_2^- -N/L and 200 mg- NO_3^- -N/L in the medium had a less pronounced effect with a 3-
318 fold decrease of the denitrification rates. Although these last three assays with NO_2^- were
319 performed with only one replicate, it revealed the tendency of NO_2^- to affect negatively the
320 denitrifying capacity of the biofilm.

321 No significant differences in the quantity of biomass (protein amount per vial) was
322 observed at the end of all these assays between vials. In all the tested conditions, the bacterial
323 diversity profiles (PCR-DGGE profiles) showed no changes (Fig. 1B, lane Tr5B). The levels of
324 *M. nitratireducens* strains JAM1/GP59 was around 10^5 *narG1* copies/ng DNA and of *H.*
325 *nitrativorans* strain NL23 between 10^1 and 10^{-1} *napA* copies/ng DNA.

326

327 **Cultivation of the original biofilm under optimal conditions**

328 As described above, the Ref300N-23C biofilm cultures exposed for a short period to 900
329 mg- NO_3^- -N/L (0.45% methanol; C/N 1.5) or to 300 mg- NO_3^- -N/L (0.15% methanol; C/N 1.5) at
330 30 and 36°C showed between 53% to 74% increases in the denitrification rates compared to the
331 original conditions (Fig. 2). Based on these results, we cultivated the OB under these optimal
332 conditions in the attempt to derive a biofilm with higher denitrifying capacity. As before, the
333 dispersed biomass of the OB was used as inoculums to colonize new carriers. Five carrier-
334 transfer cultures were carried out in ASW medium supplemented with 900 mg- NO_3^- -N/L and
335 0.45% methanol (C/N = 1.5), and incubated at either 23°C or 30°C (here named 900N-23C and

336 900N-30C, respectively), or at 300 mg-NO₃⁻-N/L and 0.15% methanol (C/N = 1.5) at 30°C (here
337 named 300N-30C). In parallel, the Ref300N-23C biofilm cultures were derived by the same
338 protocol (Table 2, Fig. S3). Compared to the Ref300N-23C biofilm cultures, the specific
339 denitrification rates of these biofilm cultures were 20% to 85% higher than those found in the
340 Ref300N-23C biofilm cultures (Table 3). Assuming methanol was not a limiting factor in these
341 cultures (not all methanol was consumed after NO₃⁻ reduction), two-way ANOVA showed that
342 temperature had the main factor affecting the specific denitrification rates ($p < 0.001$), not the
343 concentrations of NO₃⁻ ($p > 0.05$).

344 The bacterial diversity profiles (derived by PCR-DGGE) of these four biofilm cultures
345 were all similar (same profile illustrated in Fig. 1B, lane Tr5B). qPCR analyses showed the same
346 trend in the dynamics of the populations of *M. nitratireducenticrescens* strains JAM1/GP59 and
347 *H. nitrativorans* strain NL23. The concentrations of strains JAM1/GP59 increased by two orders
348 of magnitude (1.9 to 2.3×10^5 *narG1* copies/ng DNA), whereas the concentrations of strain
349 NL23 decreased by 2-3 orders of magnitude (6.1 to 18×10^1 *napA* copies/ng DNA) in the four
350 biofilm cultures compared to the OB (Fig. 3A). After the isolation of strain GP59 and the
351 sequencing of its genome, qPCR primers specific to strain JAM1 and to strain GP59 were
352 developed (Geoffroy et al. 2018). The concentrations of these two strains were then measured in
353 the OB and in the four biofilm cultures. Between 1.6 and 2.9×10^5 *nirK* copies/ng DNA of strain
354 GP59 were measured in these biofilm cultures, which is three orders of magnitude higher than
355 what was measured in the OB (1.6×10^2 *nirK* copies/ng DNA). The concentrations of strain
356 JAM1 in the four biofilm cultures ranged from 2.4 to 8.5×10^2 *tagH* copies/ng DNA, which is 3
357 to 10 times lower than its concentration in the OB (2.5×10^3 *tagH* copies/ng DNA) (Fig. 3A).

358

359 **Cultivation of the original biofilm with different NaCl concentrations**

360 In the first set of assays, we showed that the Ref300N-23C biofilm cultures could sustain
361 high variations of NaCl concentrations for a short period without affecting the denitrification
362 capacity. Based on these results, we tested the capacity of the OB to acclimate to a range of NaCl
363 concentrations. The dispersed biomass of the OB was cultivated during five carrier-transfers
364 under anoxic conditions in modified ASW medium with NaCl concentrations ranging from 0%
365 to 8% (Table 2, Fig. S3). The ASW used for the Ref300N-23C biofilm cultures contains 2.75%
366 NaCl.

367 In biofilm cultures cultivated with low NaCl concentrations (0, 0.5 and 1.0%), the amounts
368 of biomass developed were 1.3 to 2 times lower than in the Ref300N-23C biofilm cultures (Table
369 3). This lower development did not impair the specific denitrification rates of the 0% NaCl and
370 0.5% NaCl biofilm cultures, which increased by 72 and 34%, respectively, compared to the
371 Ref300N-23C biofilm cultures. However, in the 1% NaCl biofilm cultures, these rates decreased
372 by one third compared to the Ref300N-23C biofilm cultures (Table 3). The 5% NaCl and 8%
373 NaCl biofilm cultures were strongly affected by poor biofilm development on the carriers, which
374 was 50-100 less abundant in biomass than in the Ref300N-23C biofilm cultures (Table 3). In
375 these biofilm cultures, NO_3^- took more than 24 hours to be consumed and NO_2^- consumption was
376 minimal. This result contradicts what was obtained in the first set of assays where the Ref300N-
377 23C biofilm cultures subjected for 3 days to 5% and 8% NaCl did not affect the denitrification
378 performance. To investigate this further, another Ref300N-23C biofilm cultures were derived,
379 but after the fifth transfer cultures, the carriers were transferred in 5% NaCl ASW medium (here
380 named 2.75-5% NaCl) and cultivated for another three carrier transfers (20 days). These biofilm
381 cultures were able to sustain this high salt concentration. Compared to the Ref300N-23C biofilm
382 cultures, the 2.75-5% NaCl biofilm cultures had a 74% increase in the denitrification rates.
383 Because the biomass was twice the amount in these cultures, the specific denitrification rates
384 were similar to those found in the Ref300N-23C biofilm cultures (Table 3).

385 The PCR-DGGE migrating profiles showed the persistence of strain NL23 in the 0% NaCl
386 and 0.5% NaCl biofilm cultures. In the 1% NaCl biofilm cultures, however, strain NL23 was
387 barely visible in DGGE profiles (Fig. S4). Persistence of strain NL23 was confirmed by qPCR
388 where its concentrations in the 0% NaCl, 0.5% NaCl and 1.0% NaCl biofilm cultures (1.1 to 5.3
389 $\times 10^4$ *napA* cp/ng DNA) were at the same order of magnitude than in the OB (8.7×10^4 *napA*
390 cp/ng DNA) (Fig. 3B). As observed before, the concentrations of strain NL23 decreased by two
391 to three orders of magnitude in the Ref300N-23C, 5% NaCl, 8% NaCl and 2.75-5% NaCl
392 biofilm cultures (4.3 to 18×10^1 *napA* cp/ng DNA) compared to the OB (Fig. 3B). The levels of
393 *M. nitratreducens* strains JAM1/GP59 increased by one order of magnitude in the 0%
394 NaCl and 0.5% NaCl biofilm cultures (*ca.* 4×10^4 *narG1* cp/ng DNA) compared to the OB ($5.6 \times$
395 10^3 *narG1* cp/ng DNA), at similar levels than strain NL23 (Fig. 3B). However, in the 1% NaCl
396 biofilm cultures, the concentrations of *M. nitratreducens* strains JAM1/GP59 increased
397 by another order of magnitude (1.5×10^5 *narG1* cp/ng DNA), which is similar to that found in

398 the Ref300N-23C biofilm cultures (2.3×10^5 *narG1* cp/ng DNA) (Fig. 3A and 3B). As observed
399 before, the concentrations of strain JAM1 decreased by approximately one order of magnitude in
400 these cultures (6 to 23×10^1 *tagH* cp/ng DNA) when compared to the OB (2.5×10^3 *tagH* cp/ng
401 DNA) (Fig. 3B). The concentrations of strain GP59 in the 0% NaCl biofilm cultures (4.6×10^2
402 *nirK* cp/ng DNA) were at similar level as in the OB (1.6×10^2 *nirK* cp/ng DNA). It increased by
403 one order of magnitude in the 0.5% NaCl biofilm cultures (1.2×10^3 *nirK* cp/ng DNA) and by
404 two orders of magnitude in the 1.0% NaCl biofilm cultures (2.5×10^5 *nirK* cp/ng DNA), similar
405 to the levels reached in the Ref300N-23C biofilm cultures (2.9×10^5 *nirK* cp/ng DNA) (Fig. 3A
406 and 3B). All these results showed that strain GP59 was favored in the ASW medium with
407 increasing concentrations of NaCl, which was the opposite for strain NL23 in these cultures.

408

409 **Cultivation of the original biofilm in the IO medium**

410 The important decrease of strain NL23 and the enrichment of strain GP59 in the Ref300N-
411 23C biofilm cultures suggest that the switch from the continuous-operating mode that prevailed
412 in the Biodome reactor to the batch mode in our biofilm cultures could have generated these
413 specific changes in the *H. nitrativorans* and *M. nitratireducenticrescens* populations. Another
414 factor that could have influenced these changes is the ASW medium that we used instead of the
415 Instant Ocean (IO) from Aquatic systems Inc. used by the Biodome (Table S1). The dispersed
416 cells of the OB were therefore cultivated in the IO medium, under the same conditions than the
417 Ref300N-23C biofilm cultures (performed in parallel). The NO_3^- and NO_2^- dynamics in the
418 Ref300N-23C biofilm cultures were similar to those reported in Figure 1A. The denitrification
419 rates were calculated at $1.89 \text{ mM-NO}_x \text{ h}^{-1}$, and the specific denitrification rates corresponded to
420 $0.0684 \text{ mM-NO}_x \text{ h}^{-1} \text{ mg-protein}^{-1}$ (Table 3). In the IO biofilm cultures, the denitrification rates
421 ($0.76 \text{ mM-NO}_x \text{ h}^{-1}$) were 2.5 lower than in the Ref300N-23C biofilm cultures (Table 3).
422 However, because less biomass developed in the IO biofilm cultures (Table 3), its specific
423 denitrification rates ($0.0661 \text{ mM-NO}_x \text{ h}^{-1} \text{ mg-protein}^{-1}$) were the same as in the Ref300N-23C
424 biofilm cultures.

425 As before, PCR-DGGE migration profiles (Fig. S5) and qPCR assays showed the decrease
426 in strain NL23 concentrations during the five carrier transfers in the Ref300N-23C biofilm
427 cultures. Surprisingly, the DGGE band corresponding to strain NL23 was still present in the IO
428 biofilm cultures (Fig. S5). This result was confirmed by qPCR assays where the concentrations

429 of strain NL23 in the IO biofilm cultures (Fig. 3A) were at similar levels (7.0×10^4 *napA* cp/ng
430 DNA) as in the OB (8.7×10^4 *napA* copies/ng DNA). In the IO biofilm cultures, the
431 concentrations of strain JAM1 and strain GP59 were both approximately one order of magnitude
432 lower than those of strain NL23 (Fig. 3A).

433

434 **Cultivation of the original biofilm under oxic conditions and with NO_2^-**

435 The dispersed cells of the OB were cultivated under the same conditions of the Ref300N-
436 23C biofilm cultures except that the cultures were incubated under oxic conditions. Growth was
437 not impaired in these cultures, but the specific denitrification rates were 5 times lower than in the
438 Ref300N-23C anoxic biofilm cultures (Table 3). Culturing the OB with a mix of 200 mg- NO_3^- -
439 N/L and 200 mg- NO_2^- -N/L (here named 200-200N) generated biofilm cultures with
440 denitrification rates 2.5-times lower than those in the Ref300N-23C biofilm cultures (Table 3).
441 The proportions of *M. nitratireducens* strains JAM1/GP50 and strain NL23 in these two
442 types of cultures were similar to the ones in the Ref300N-23C biofilm cultures (Fig. 3A and 3B).

443

444 **Discussion**

445 The Montreal Biodome operated a continuous fluidized methanol-fed denitrification
446 system between 1998 and 2006. After the termination of the system, our group took the
447 opportunity to preserve carriers with the denitrifying biofilm in glycerol at -20°C (Laurin et al.
448 2006). Our study showed that the bacterial community of the OB taken from the carriers could
449 be revived by the colonization of new carriers, and could adapt to environment changes, which
450 included the batch-operating mode, instead of the continuous mode, and the use of a homemade
451 ASW medium, instead of the commercial IO medium used in the Biodome system. The resulting
452 reference biofilm cultures could sustain denitrifying activities after a short exposure to various
453 environmental changes such as a range of concentrations of NaCl (0% to 8%), NO_3^- (6.4 to 214
454 mM), NO_2^- (14.3 to 28.6 mM) and methanol (0% to 0.75%), and a range of pH (4 to 8) and
455 temperature (4°C to 36°C). Denitrifying activities were inhibited when the biofilm cultures were
456 exposed at pH 10 or with 1.5% methanol, and were strongly impacted by the presence of NO_2^- .
457 Low pHs had a strong effect on the denitrification rates, with a 2.7-fold increase at pH 4.0 than at
458 pH 8.0 in the original conditions. At the end of the assays (10-12 hours), however, the pH in the
459 medium increased to reach pH 8.0. Denitrification is an alkalization process that leads to the

460 increase of pH in the medium (Lee & Rittmann 2003). Phosphate used in ASW as buffering
461 system explains the stabilization of the pH at around 8.0.

462 Cultivating the OB in ASW had a strong impact on the population of *H. nitrativorans*
463 strain NL23. A decrease of three orders of magnitude was noticed in the levels of strain NL23
464 after 5 weeks in the Ref300N-23C biofilm cultures. However, this important decrease did not
465 translate into impaired denitrification rates. On the opposite, the concentrations of *M.*
466 *nitratireducenticrescens* in these biofilm cultures increased by two orders of magnitude during
467 that time. We already reported in Geoffroy *et al.* (2018) that *H. nitrativorans* strain NL23 was
468 displaced in ASW biofilm cultures by a new strain of *M. nitratireducenticrescens*, strain GP59,
469 that can perform complete denitrification. The increase of concentrations of *M.*
470 *nitratireducenticrescens* involved mainly strain GP59, which was at a very low level in the OB,
471 but increased by three orders of magnitude in the Ref300N-23C biofilm cultures. Strain JAM1
472 stayed at the same levels in the OB and the Ref300N-23C biofilm cultures. The reasons why
473 strain GP59 did not appear in first instance in the denitrification system of the Biodome are
474 obscure.

475 In the second set of assays, a series of biofilm cultures were performed, in which the
476 bacterial community of the OB was cultivated to different conditions for five weeks. The best
477 denitrification performance occurred in the 900N-30C biofilm cultures (64.3 mM NO₃⁻, 0.45%
478 methanol, 30°C), which allowed a 85% increase in the specific denitrification rates compared to
479 the Ref300N-23C biofilm cultures. We did not try to cultivate the OB in low pHs because of the
480 rapid fluctuations of the pH during the first set of assays. The NaCl concentration was a critical
481 factor for the persistence of *H. nitrativorans* strain NL23 in the biofilm cultures. In the ASW-
482 modified media with low NaCl concentrations (0%, 0.5% and 1.0%), high levels of *H.*
483 *nitrativorans* strain NL23 were found. Martineau *et al.* (2015) showed that strain NL23 can
484 sustain good growth and denitrifying activities up to 1% NaCl in planktonic pure cultures, but
485 underwent substantial decrease in these features at 2% NaCl. In contrast, growth of
486 *Methylophaga* spp. requires Na⁺ (Bowman 2005). In the 0% NaCl biofilm cultures, low
487 concentration of Na⁺ (0.06%; originating from NaNO₃, Na₂HPO₄ and NaOH, Table S1) may
488 have impaired the growth of *M. nitratireducenticrescens*. Consequently, strain *H. nitrativorans*
489 NL23 may have better competed against the *Methylophaga* strains in these cultures. On the
490 opposite, the rapid processing of NO₃⁻ by *M. nitratireducenticrescens* strain JAM1 (Mauffrey et

491 al. 2015) during the acclimation of the bacterial community to new culture conditions in the
492 Ref300N-23C biofilm cultures (containing 2.75% NaCl) with transient accumulation of NO_2^-
493 may have impaired the growth of *H. nitrativorans* strain NL23. The adverse environment for
494 strain NL23 would have favored the emergence of strain GP59 that took over strain NL23 in
495 completing the full denitrification pathway.

496 An interesting result was to find the persistence of strain NL23 in the biofilm cultivated
497 under the same conditions than the Ref300N-23C biofilm cultures but with the commercial
498 medium (IO) that was used in the Biodome denitrification system. The concentrations of *M.*
499 *nitratireducenticrescens* strains JAM1/GP59 and *H. nitrativorans* strain NL23 were similar in
500 the IO biofilm cultures. Although the IO biofilm cultures generated less biomass than the
501 Ref300N-23C biofilm cultures, the specific denitrification rates were the same between the two
502 types of cultures, suggesting similar dynamics of NO_3^- processing by both bacterial communities.
503 The reason of this fundamental difference has to be found in the composition of the seawater
504 medium. Based on limited information on the composition of the IO medium, a difference that
505 might impair the growth of strain NL23 is the lack of bicarbonate and carbonate in the ASW
506 medium. *Hyphomicrobium* spp. use the serine pathway for the C1 assimilation, which requires
507 one molecule of CO_2 and two molecules of formaldehyde in each cycle forming a three-carbon
508 intermediate, while in *Methylophaga* spp. the ribulose monophosphate cycle requires three
509 molecules of formaldehyde (Anthony 1982).

510 The co-occurrence of *Methylophaga* spp. and *Hyphomicrobium* spp. has also been reported
511 in two other denitrification systems. Osaka et al. (2008) studied a laboratory-scale continuously
512 stirred tank reactor with synthetic wastewater and sludge adapted to denitrification conditions.
513 Among the tested assays, the reactor was fed with methanol and acclimated with increasing
514 concentrations of NaCl. This reactor could perform high levels of NO_3^- removal at up to 3%
515 NaCl, but a drastic decrease in the removal efficiency occurred at 4% NaCl. Further increases in
516 NaCl concentrations failed to get denitrifying activities in the reactor. Co-occurrence of
517 *Methylophaga* spp. and *Hyphomicrobium* spp. was revealed in cloned 16S rRNA gene libraries
518 derived from the 0% and 4% NaCl synthetic wastewater media, with respectively 35% and 20%
519 of the 16S clones affiliated to *Hyphomicrobium* spp., and 8% and 11% affiliated to
520 *Methylophaga* spp. The other denitrification system is a methanol-fed fluidized-bed continuous
521 denitrification system treating a marine aquarium in Helsinki, Finland (Rissanen et al. 2016). In

522 this system, the biofilm developed on oolitic sand. *Methylophaga* spp. and *Hyphomicrobium* spp.
523 were found in high proportions although high variations occurred between samples taken at
524 different years. All these studies including ours showed the importance of these two taxa in
525 marine denitrification processes. Our study demonstrated how these two taxa co-evolved in a
526 biofilm cultures subjected to environmental changes.

527

528 **Conclusions**

529 The denitrifying biofilm taken from the Biodome denitrification system, which was frozen
530 for many years, can be revived and cultivated at laboratory scale without the loss of denitrifying
531 activities. The resulting reference biofilm cultures can sustain for a short period of time (2-5
532 days) changes in NaCl, NO_3^- , NO_2^- and methanol concentrations, and in pH and temperature.
533 Inhibition of the denitrifying activities occurred at high pH (10) and high methanol
534 concentrations (1.5%). The biofilm cultures that resulted in the highest specific denitrification
535 rates were those cultivated for 5 weeks with 64.3 mM NO_3^- , 0.45% methanol at 30°C. The OB
536 can sustain growth and denitrifying activity in ASW with low concentrations of NaCl, but
537 showed poor developmental growth in ASW with 5% and 8% NaCl. Although lower biomass
538 developed in the biofilm cultures cultivated in the commercial IO medium, specific
539 denitrification rates similar to the one of the reference biofilm cultures (ASW medium) were
540 observed. Important changes in the co-occurrence of the populations of *M.*
541 *nitratreducenticrescens* and *H. nitratorans* were observed in these biofilm cultures when
542 compared to the OB. In the ASW biofilm cultures, substantial growth of a subpopulation of *M.*
543 *nitratreducenticrescens* (strain GP59 as representative) with full denitrification capacity
544 displaced *H. nitratorans* strain NL23. Strain NL23 however persisted in biofilm cultures
545 cultivated in ASW with low concentration of NaCl, but also in the ones cultivated in the IO
546 medium. This study shows the plasticity of the denitrifying biofilm to adapt to different
547 conditions. Knowing the dynamics of the principal actors in the denitrifying biofilm in response
548 to environmental changes will allow to better predict the behavior of a denitrification system in
549 response to changes in the operating mode. For instance, as the mature biofilm cultures were able
550 to denitrify under low or high NaCl concentrations, this suggests that the Biodome denitrification
551 system could have sustained high variations in salt concentrations.

552

553 **Acknowledgements**

554 We thank Karla Vasquez for her technical assistance.

555

556 **References**

557 Anthony C. 1982. *Biochemistry of methylotrophs. Chapters 3 and 4. Academic Press, London:*
558 Academic Press.

559 Auclair J, Lepine F, Parent S, and Villemur R. 2010. Dissimilatory reduction of nitrate in
560 seawater by a *Methylophaga* strain containing two highly divergent *narG* sequences.
561 *ISME J* 4:1302-1313. doi: 10.1038/ismej.2010.47

562 Baytshtok V, Kim S, Yu R, Park H, and Chandran K. 2008. Molecular and biokinetic
563 characterization of methylotrophic denitrification using nitrate and nitrite as terminal
564 electron acceptors. *Water Science and Technology* 58:359-365. doi:
565 10.2166/wst.2008.391

566 Bertrand EM, McCrow JP, Moustafa A, Zheng H, McQuaid JB, Delmont TO, Post AF, Sipler
567 RE, Spackeen JL, Xu K, Bronk DA, Hutchins DA, and Allen AE. 2015. Phytoplankton-
568 bacterial interactions mediate micronutrient colimitation at the coastal Antarctic sea ice
569 edge. *Proceedings of the National Academy of Sciences of the United States of America*
570 112:9938-9943. doi: 10.1073/pnas.1501615112

571 Boden R. 2012. Emended description of the genus *Methylophaga* Janvier et al. 1985.
572 *International Journal of Systematic and Evolutionary Microbiology* 62:1644-1646. doi:
573 10.1099/ijs.0.033639-0

574 Bowman JP. 2005. Genus IV. *Methylophaga*. *Bergey's manual of systematic bacteriology:*
575 Springer, Ney York, NY, USA.

576 Cabrol L, and Malhautier L. 2011. Integrating microbial ecology in bioprocess understanding:
577 the case of gas biofiltration. *Applied Microbiology and Biotechnology* 90:837-849. doi:
578 10.1007/s00253-011-3191-9

579 CatalanSakairi MA, Wang PC, and Matsumura M. 1997. High-rate seawater denitrification
580 utilizing a macro-porous cellulose carrier. *Journal of Fermentation and Bioengineering*
581 83:102-108. doi: 10.1016/S0922-338x(97)87335-7

582 Geoffroy V, Payette G, Mauffrey F, Lestin L, Constant P, and Villemur R. 2018. Strain-level
583 genetic diversity of *Methylophaga nitratireducentis* confers plasticity to

- 584 denitrification capacity in a methylotrophic marine denitrifying biofilm. *Peerj* 6:e4679.
585 doi: 10.7717/peerj.4679
- 586 Ginige MP, Hugenholtz P, Daims H, Wagner M, Keller J, and Blackall LL. 2004. Use of stable-
587 isotope probing, full-cycle rRNA analysis, and fluorescence in situ hybridization-
588 microautoradiography to study a methanol-fed denitrifying microbial community.
589 *Applied and Environmental Microbiology* 70:588-596. doi: 10.1128/Aem.70.1.588-
590 596.2004
- 591 Gliesche C, Fesefeld A, and Hirsch P. 2005. Genus I. *Hyphomicrobium*. In: Garrity GM, Brenner
592 DJ, Krieg NR, and Staley JT, eds. *Bergey's manual of systematic bacteriology*. 2nd ed:
593 Springer, New York, NY, USA, 476-494.
- 594 Kraft B, Strous M, and Tegetmeyer HE. 2011. Microbial nitrate respiration - Genes, enzymes
595 and environmental distribution. *Journal of Biotechnology* 155:104-117. doi:
596 10.1016/j.jbiotec.2011.12.025
- 597 Labbé N, Juteau P, Parent S, and Villemur R. 2003. Bacterial diversity in a marine methanol-fed
598 denitrification reactor at the Montreal Biodome, Canada. *Microbial Ecology* 46:12-21.
599 doi: 10.1007/s00248-002-1056-6
- 600 Labbé N, Laurin V, Juteau P, Parent S, and Villemur R. 2007. Microbiological community
601 structure of the biofilm of a methanol-fed, marine denitrification system, and
602 identification of the methanol-utilizing microorganisms. *Microbial Ecology* 53:621-630.
603 doi: 10.1007/s00248-006-9168-z
- 604 Lafortune I, Juteau P, Deziel E, Lepine F, Beaudet R, and Villemur R. 2009. Bacterial diversity
605 of a consortium degrading high-molecular-weight polycyclic aromatic hydrocarbons in a
606 two-liquid phase biosystem. *Microbial Ecology* 57:455-468. doi: 10.1007/s00248-008-
607 9417-4
- 608 Landa M, Blain S, Harmand J, Monchy S, Rapaport A, and Obernosterer I. 2018. Major changes
609 in the composition of a Southern Ocean bacterial community in response to diatom-
610 derived dissolved organic matter. *Fems Microbiology Ecology* 94. doi:
611 10.1093/femsec/fiy034
- 612 Laurin V, Labbe V, Juteau P, Parent S, and Villemur R. 2006. Long-term storage conditions for
613 carriers with denitrifying biomass of the fluidized, methanol-fed denitrification reactor of

- 614 the Montreal Biodome, and the impact on denitrifying activity and bacterial population.
615 *Water Research* 40:1836-1840. doi: 10.1016/j.watres.2006.03.002
- 616 Layton AC, Karanth PN, Lajoie CA, Meyers AJ, Gregory IR, Stapleton RD, Taylor DE, and
617 Sayler GS. 2000. Quantification of *Hyphomicrobium* populations in activated sludge
618 from an industrial wastewater treatment system as determined by 16S rRNA analysis.
619 *Applied and Environmental Microbiology* 66:1167-1174. doi: 10.1128/Aem.66.3.1167-
620 1174.2000
- 621 Lee KC, and Rittmann BE. 2003. Effects of pH and precipitation on autohydrogenotrophic
622 denitrification using the hollow-fiber membrane-biofilm reactor. *Water Research*
623 37:1551-1556. doi: 10.1016/S0043-1354(02)00519-5
- 624 Lemmer H, Zaglauer A, Neef A, Meier H, and Amann R. 1997. Denitrification in a methanol-fed
625 fixed-bed reactor .2. Composition and ecology of the bacterial community in the
626 biofilms. *Water Research* 31:1903-1908. doi: 10.1016/S0043-1354(97)00027-4
- 627 Li TD, Doronina NV, Ivanova EG, and Trotsenko YA. 2007. Vitamin B-12-independent strains
628 of *Methylophaga marina* isolated from red sea algae. *Microbiology* 76:75-81. doi:
629 10.1134/S0026261707010110
- 630 Martineau C, Mauffrey F, and Villemur R. 2015. Comparative analysis of denitrifying activities
631 of *Hyphomicrobium nitratorans*, *Hyphomicrobium denitrificans*, and *Hyphomicrobium*
632 *zavarzinii*. *Applied and Environmental Microbiology* 81:5003-5014. doi:
633 10.1128/Aem.00848-15
- 634 Martineau C, Villeneuve C, Mauffrey F, and Villemur R. 2013a. Complete genome sequence of
635 *Hyphomicrobium nitratorans* strain NL23, a denitrifying bacterium isolated from
636 biofilm of a methanol-fed denitrification system treating seawater at the Montreal
637 Biodome. *Genome Announc* 2: e01165-13. doi: 10.1128/genomeA.01165-13
- 638 Martineau C, Villeneuve C, Mauffrey F, and Villemur R. 2013b. *Hyphomicrobium nitratorans*
639 sp. nov., isolated from the biofilm of a methanol-fed denitrification system treating
640 seawater at the Montreal Biodome. *International Journal of Systematic and Evolutionary*
641 *Microbiology* 63:3777-3781. doi: 10.1099/ijs.0.048124-0
- 642 Mauffrey F, Cucaita A, Constant P, and Villemur R. 2017. Denitrifying metabolism of the
643 methylotrophic marine bacterium *Methylophaga nitratorans* strain JAM1.
644 *PeerJ* 5. doi: 10.7717/Peerj.4098

- 645 Mauffrey F, Martineau C, and Villemur R. 2015. Importance of the two dissimilatory (Nar)
646 nitrate reductases in the growth and nitrate reduction of the methylotrophic marine
647 bacterium *Methylophaga nitratreducenticrescens* JAM1. *Frontiers in Microbiology*
648 6:1475. doi: 10.3389/fmicb.2015.01475.
- 649 Mulvaney RL. 1996. Nitrogen—inorganic forms. In: Sparks DL, ed. *Methods of Soil Analysis*
650 *Part 3 Book Ser 5*: Soil Science Society of America: Madison, WI, USA, pp. 1123–1184.
- 651 Osaka T, Shirotani K, Yoshie S, and Tsuneda S. 2008. Effects of carbon source on denitrification
652 efficiency and microbial community structure in a saline wastewater treatment process.
653 *Water Research* 42:3709-3718. doi: 10.1016/j.watres.2008.06.007
- 654 Philippot L, and Hojberg O. 1999. Dissimilatory nitrate reductases in bacteria. *Biochimica Et*
655 *Biophysica Acta-Gene Structure and Expression* 1446:1-23. doi: 10.1016/S0167-
656 4781(99)00072-X
- 657 Richardson D, Berks B, Russell D, Spiro S, and Taylor C. 2001. Functional, biochemical and
658 genetic diversity of prokaryotic nitrate reductases. *Cellular and Molecular Life Sciences*
659 *CMLS* 58:165-178.
- 660 Rissanen AJ, Ojala A, Dernjatin M, Jaakkola J, and Tirola M. 2016. *Methylophaga* and
661 *Hyphomicrobium* can be used as target genera in monitoring saline water methanol-
662 utilizing denitrification. *Journal of Industrial Microbiology & Biotechnology* 43:1647-
663 1657. doi: 10.1007/s10295-016-1839-2
- 664 Rissanen AJ, Ojala A, Fred T, Toivonen J, and Tirola M. 2017. *Methylophilaceae* and
665 *Hyphomicrobium* as target taxonomic groups in monitoring the function of methanol-fed
666 denitrification biofilters in municipal wastewater treatment plants. *Journal of Industrial*
667 *Microbiology & Biotechnology* 44:35-47. doi: 10.1007/s10295-016-1860-5
- 668 Roder HL, Sorensen SJ, and Burmolle M. 2016. Studying bacterial multispecies biofilms: Where
669 to start? *Trends in microbiology* 24:503-513. doi: 10.1016/j.tim.2016.02.019
- 670 Salta M, Wharton JA, Blache Y, Stokes KR, and Briand JF. 2013. Marine biofilms on artificial
671 surfaces: structure and dynamics. *Environmental Microbiology* 15:2879-2893. doi:
672 10.1111/1462-2920.12186
- 673 Tan CH, Lee KW, Burmolle M, Kjelleberg S, and Rice SA. 2017. All together now:
674 experimental multispecies biofilm model systems. *Environmental Microbiology* 19:42-
675 53. doi: 10.1111/1462-2920.13594

- 676 Tchobanoglous G, Burton FL, and Stensel HD. 2003. *Wastewater Engineering: Treatment and*
677 *Reuse, 4th.*
- 678 Villemur R, Payette G, Geoffroy V, Mauffrey F, and Martineau C. 2019. Dynamics of a
679 methanol-fed marine denitrifying biofilm: 2- Impact of environmental changes on the
680 microbial community. *Submitted in PeerJ.*
- 681 Villeneuve C, Martineau C, Mauffrey F, and Villemur R. 2012. Complete genome sequences of
682 *Methylophaga* sp. strain JAM1 and *Methylophaga* sp. strain JAM7. *Journal of*
683 *Bacteriology* 194:4126-4127. doi: 10.1128/Jb.00726-12
- 684 Villeneuve C, Martineau C, Mauffrey F, and Villemur R. 2013. *Methylophaga*
685 *nitratireducentirescens* sp. nov. and *Methylophaga frappieri* sp. nov., isolated from the
686 biofilm of the methanol-fed denitrification system treating the seawater at the Montreal
687 Biodome. *International Journal of Systematic and Evolutionary Microbiology* 63:2216-
688 2222. doi: 10.1099/Ijs.0.044545-0
- 689 Wang Z, Zhang XX, Lu X, Liu B, Li Y, Long C, and Li AM. 2014. Abundance and diversity of
690 bacterial nitrifiers and denitrifiers and their functional genes in tannery wastewater
691 treatment plants revealed by high-throughput sequencing. *Plos One* 9. doi:
692 10.1371/journal.pone.0113603
- 693

Figure 1(on next page)

Dynamics of NO_3^- and NO_2^- concentrations and of the bacterial community in the Ref300N-23C biofilm cultures

Panel A: NO_3^- and NO_2^- concentrations were measured during the fifth carrier-transfer cultures. Results from triplicate biofilm cultures.

Panel B: PCR-DGGE migration profiles representing the bacterial diversity during the first five carrier-transfer cultures. DNA extraction was performed on samples from the suspended biomass in the first four carrier-transfer cultures because the biofilm was not enough abundant on the carriers in these cultures. In the fifth carrier-transfer cultures, samples were taken from the suspended biomass (Tr5s) and the biofilm (Tr5B). OB: Original biofilm. Approximately 300 ng of the PCR product was loaded per sample.

Panel C. Quantification of *M. nitratireducenticrescens* strains JAM1/GP59 (*narG1*) and *H. nitrativorans* strain NL23 (*napA*) by qPCR in the five carrier-transfer cultures (Tr1 to Tr5). Results from 3 to 9 replicate cultures, of one to three different inoculums.

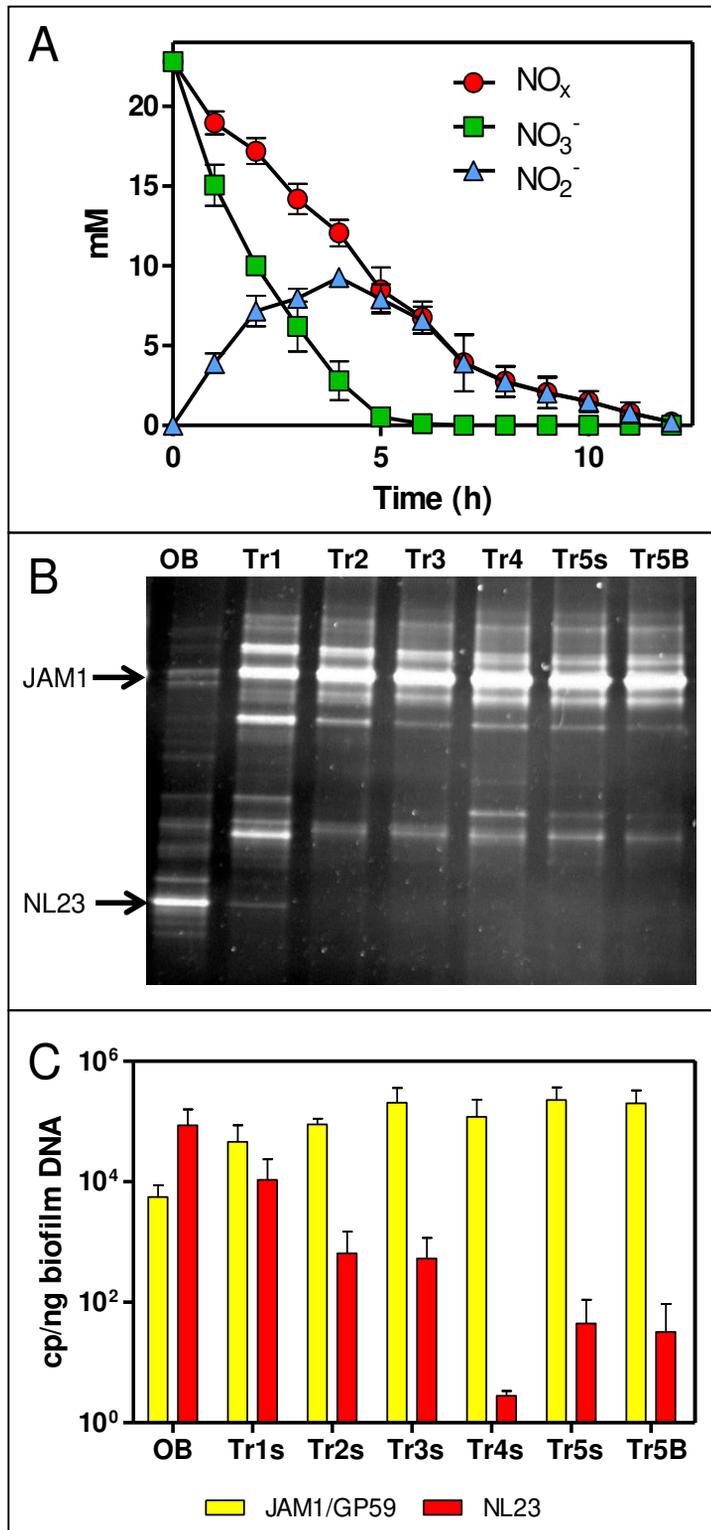


Figure 2 (on next page)

Denitrifying activities of the Ref300N-23C biofilm cultures impacted by different physico-chemical parameters

The Ref300N-23C biofilm cultures were exposed for few hours or few days in modified ASW medium as described in Table 1. Denitrification rates are expressed relative to the denitrification rates of the original conditions of the Ref300N-23C biofilm cultures (set to one) identified by *. Significance between the denitrification rates were determined by one-way ANOVA. Rates with the same letter are not significantly different. X: no activities recorded. Average of triplicates cultures except for the 400N-NO₂, 200N-NO₂ and 200N-NO₃/200N-NO₂ conditions with one replicate. N: Nitrogen NO₃⁻. M: methanol.

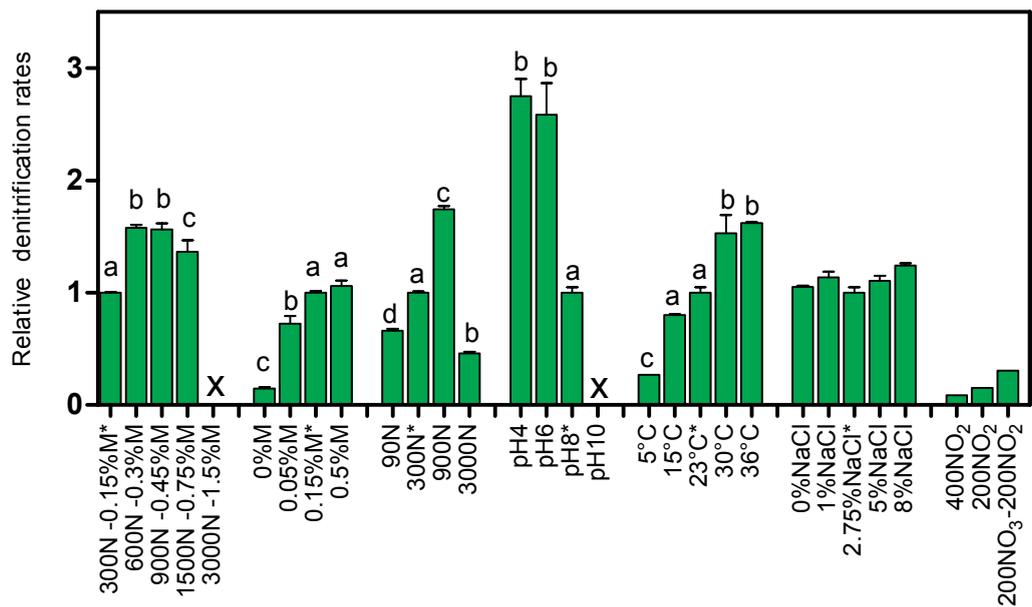


Figure 3(on next page)

Denitrifying activities and dynamics of *H. nitrativorans* strain NL23 and *M. nitratireducenticrescens* in the biofilm cultures.

Panels A and B. Quantification of *M. nitratireducenticrescens* strains JAM1/GP59 (*narG1*), *M. nitratireducenticrescens* strain JAM1 (*tagH*), *M. nitratireducenticrescens* strain GP59 (*nirK*) and *Hyphomicrobium nitrativorans* strain NL23 (*napA*) by qPCR in the corresponding biofilm cultures. Results from triplicate cultures. OB: Original biofilm. Significance of the qPCR levels (\log_{10} transformed) of the respective genes between the biofilm cultures were determined by one-way ANOVA with Tukey's Multiple Comparison Test. Rates with the same letter (within the bar) are not significantly different.

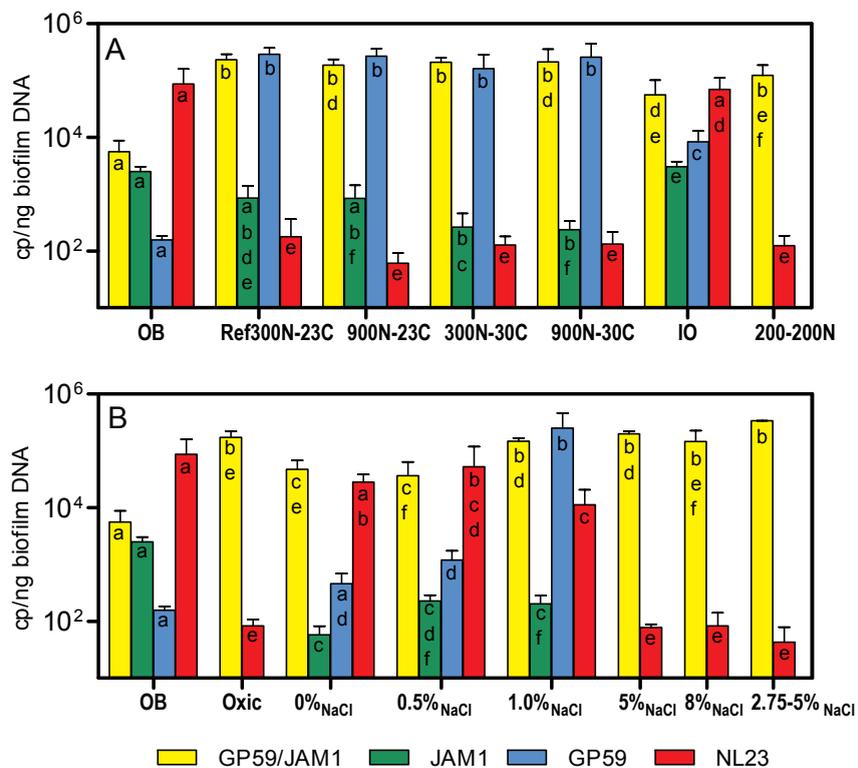


Table 1 (on next page)

Incubation conditions to measure the impact of specific parameters on the reference biofilm cultures (Ref 300N-23C)

Table 1. Incubation conditions of the reference biofilm cultures (Ref 300N-23C) exposed for a short period to specific environmental conditions

Name	NO ₃ ⁻ /NO ₂ ⁻ mg-N/L (mM)	Methanol % (v/v)	C/N	NaCl % (w/v)	pH	Temp °C
Tested conditions						
NO ₃ ⁻ and methanol (C/N=1.5). Exposure: 12-48 h						
300 _{N-NO3} -0.15% _{MeOH} *	300 (21.4)	0.15	1.5	2.75	8.0	23
600 _{N-NO3} -0.3% _{MeOH}	600 (42.8)	0.3	1.5	2.75	8.0	23
900 _{N-NO3} -0.45% _{MeOH}	900 (64.3)	0.45	1.5	2.75	8.0	23
1500 _{N-NO3} -0.75% _{MeOH}	1500 (107)	0.75	1.5	2.75	8.0	23
3000 _{N-NO3} -1.5% _{MeOH}	3000 (214)	1.5	1.5	2.75	8.0	23
NO ₂ ⁻ . Exposure: 12-48 h						
400 _{N-NO2}	400 (28.6)	0.15	1.5	2.75	8.0	23
200 _{N-NO2}	200 (14.3)	0.15	1.5	2.75	8.0	23
200 _{N-NO3} /200 _{N-NO2}	200+200 (28.6)	0.15	1.5	2.75	8.0	23
Methanol (C/N variable). Exposure: 12-48 h						
0% _{MeOH}	300 (21.4)	0	0	2.75	8.0	23
0.05% _{MeOH}	300 (21.4)	0.05	0.5	2.75	8.0	23
0.15% _{MeOH} *	300 (21.4)	0.15	1.5	2.75	8.0	23
0.5% _{MeOH}	300 (21.4)	0.5	5	2.75	8.0	23
NO ₃ ⁻ . (C/N variable): Exposure: 12-48 h						
90 _{N-NO3}	90 (6.4)	0.15	5	2.75	8.0	23
300 _{N-NO3} *	300 (21.4)	0.15	1.5	2.75	8.0	23
900 _{N-NO3}	900 (64.3)	0.15	0.5	2.75	8.0	23
3000 _{N-NO3}	3000 (214)	0.15	0.15	2.75	8.0	23
pH**. Exposure 48-120 h						
pH4	300 (21.4)	0.15	1.5	2.75	4.0	23
pH6	300 (21.4)	0.15	1.5	2.75	6.0	23
pH8*	300 (21.4)	0.15	1.5	2.75	8.0	23
pH10	300 (21.4)	0.15	1.5	2.75	10.0	23
Temperature**. Exposure 48-120 h						
5°C	300 (21.4)	0.15	1.5	2.75	8.0	5
15°C	300 (21.4)	0.15	1.5	2.75	8.0	15
23°C*	300 (21.4)	0.15	1.5	2.75	8.0	23
30°C	300 (21.4)	0.15	1.5	2.75	8.0	30
36°C	300 (21.4)	0.15	1.5	2.75	8.0	36
NaCl**. Exposure 48-120 h						
0% _{NaCl}	300 (21.4)	0.15	1.5	0	8.0	23
1% _{NaCl}	300 (21.4)	0.15	1.5	1.0	8.0	23
2.75% _{NaCl} *	300 (21.4)	0.15	1.5	2.75	8.0	23
5% _{NaCl}	300 (21.4)	0.15	1.5	5.0	8.0	23
8% _{NaCl}	300 (21.4)	0.15	1.5	8.0	8.0	23

45 The ASW medium was adjusted for specific concentrations of methanol (MeOH), NO_3^- , NO_2^-
46 and NaCl, and pH. The biofilm carriers from the reference biofilm cultures were transferred in
47 these solutions and incubated under the prescribed temperature. The $200_{\text{N-NO}_3}$ - $200_{\text{N-NO}_2}$ assays
48 were done with a mix of 200 mg NO_3^- -N/L and 200 mg NO_2^- -N/L. In gray are parameters
49 different from those of the reference biofilm cultures. C/N: Carbon Nitrogen ratio. A 1.5 C/N
50 ratio corresponds, for instance, to 300 mg NaNO_3 -N/L (21.4 mM) and 450 mg CH_3OH -C/L (37.5
51 mM or 0.15% v/v). All assays were carried out in triplicates, except the NO_2^- assays (one
52 replicate). See Figure S2 for more details.

53 *Reference biofilm cultures.

54 **The reference biofilm cultures were incubated in the prescribed conditions for two days, and
55 then transferred in fresh medium for 12 to 48 h.

Table 2 (on next page)

Culture conditions used for the cultivation of the original biofilm

1 **Table 2. Culture conditions used for the cultivation of the original biofilm**

2

3

4

Name	Medium	NO ₃ ⁻ /NO ₂ ⁻ mg-N/L (mM)	Methanol % (v/v)	NaCl % (w/v)	Temp °C
5 Ref300N-23C ^a	ASW	300 (21.4)	0.15	2.75	23
6 Oxic^b	ASW	300 (21.4)	0.15	2.75	23
7 300N-30C	ASW	300 (21.4)	0.15	2.75	30
8 900N-23C	ASW	900 (64.3)	0.45	2.75	23
9 900N-30C	ASW	900 (64.3)	0.45	2.75	30
10 0%NaCl	ASW	300 (21.4)	0.15	0	23
11 0.5%NaCl	ASW	300 (21.4)	0.15	0.5	23
12 1%NaCl	ASW	300 (21.4)	0.15	1.0	23
13 5%NaCl	ASW	300 (21.4)	0.15	5.0	23
14 8%NaCl	ASW	300 (21.4)	0.15	8.0	23
15 2.75-5%NaCl ^c	ASW	300 (21.4)	0.15	2.75/5	23
16 IO	IO	300 (21.4)	0.15	3.0 ^d	23
17 200-200N	ASW	200 NO₃⁻/	0.15	2.75	23
18		200 NO₂⁻ (28.6)			

19 The original biofilm was cultivated in triplicates in these conditions at pH 8.0. The carriers were
 20 transferred five times in fresh medium (eight times for the 200-200N biofilm cultures). In gray are
 21 changed parameters from the Ref300N-23C biofilm cultures. IO: Instant Ocean medium.

22 ^a Reference biofilm cultures.

23 ^b Cultures were performed under oxic conditions.

24 ^c The Ref300N-23C biofilm cultures were further transferred three times in ASW composed of
 25 5% NaCl.

26 ^d Based on the amount of Na⁺ and Cl⁻ in Table S1. The exact amount of NaCl added in the IO
 27 medium is not known. For comparison, the amount of Na⁺ and Cl⁻ in the ASW medium is 3.2%.

Table 3 (on next page)

Denitrifying activities in the biofilm cultures

Table 3. Denitrifying activities in the biofilm cultures

Biofilm cultures	Denitrifying rates	Relative activities	Protein concentration	Specific denitrification rates	Relative specific activities
	mM-NO _x h ⁻¹ flask ⁻¹		mg/vial	mM-NO _x h ⁻¹ mg-protein ⁻¹	
0%NaCl	1.28 (0.01)	0.87 (0.01)	14.1 (1.2)	0.0911 (0.0080)	1.72 (0.15) ^b
0.5%NaCl	1.57 (0.02)	1.06 (0.01)	22.2 (2.6)	0.0712 (0.0084)	1.34 (0.16) ^{ab}
1%NaCl	0.66 (0.01)	0.45 (0.01)	18.5 (0.2)	0.0357 (0.0011)	0.67 (0.02) ^{cd}
Ref300N-23C*	1.47 (0.05)	1.00 (0.03)	28.1 (0.4)	0.0530 (0.0064)	1.00 (0.12) ^{ad}
300N-30C	1.94 (0.04)	1.32 (0.03)	20.5 (0.9)	0.0946 (0.0027)	1.79 (0.05) ^b
900N-23C	1.95 (0.04)	1.32 (0.02)	30.7 (2.9)	0.0637 (0.0060)	1.20 (0.11) ^{abd}
900N-30C	2.37 (0.09)	1.61 (0.06)	24.5 (2.8)	0.0979 (0.0130)	1.85 (0.24) ^b
Ref300N-23C*	1.42 (0.01)	1.00 (0.01)	31.0 (1.8)	0.0458 (0.0027)	1.00 (0.06) ^{ad}
Oxic	0.18 (0.08)	0.12 (0.06)	19.8 (1.1)	0.0090 (0.0046)	0.20 (0.10) ^c
5%NaCl	0.06 (0.01)	0.04 (0.01)	0.38 (0.12)	0.1660 (0.0322)	3.63 (0.70) ^e
2.75-5%NaCl	2.46 (0.08)	1.74 (0.06)	62.7 (3.5)	0.0394 (0.0032)	0.86 (0.07) ^{ad}
8%NaCl	0.07 (0.01)	0.05 (0.01)	0.84 (0.04)	0.0845 (0.0071)	1.84 (0.15) ^b
Ref300N-23C*	1.89 (0.10)	1.00 (0.05)	27.6 (1.5)	0.0684 (0.0005)	1.00 (0.01) ^{ad}
IO	0.76 (0.03)	0.40 (0.01)	11.5 (0.8)	0.0661 (0.0063)	0.97 (0.10) ^{ad}
200/200N	0.39 (0.13)	0.41 (0.01)	nd	nd	nd

Except for the 200/200N biofilm cultures, three sets of cultures assays were performed with the Ref300N-23C biofilm cultures in each set for comparison. Relative activities are to the denitrification rates of the Ref300N-23C biofilm cultures set to one. Protein concentrations reflect the biomass content in vials at the end of the assays. Specific denitrification rates are the denitrification rates divided by protein amount in vials. Relative specific activities are to the specific denitrification rates of the Ref300N-23C biofilm cultures set to one. Significance between the relative specific denitrification rates were determined by one-way ANOVA with Tukey's Multiple Comparison Test. The relative specific denitrification rates with the same letter are not significantly different. See Table 2 and Material & Methods for nomenclature and culture conditions. Results are average of triplicate biofilm cultures. Values between parentheses: standard deviation. nd: not done.

34
35