

Experimental removal of introduced slider turtles offers new insight into competition with a native, threatened turtle

Corresponding Author: Max R Lambert Email address: mrl24@berkeley.edu

The red-eared slider turtle (Trachemys scripta elegans; RES) is often considered one of the world's most invasive species. Results from laboratory and mesocosm experiments suggest that introduced RES outcompete native turtles for key ecological resources, but such experiments can overestimate the strength of competition. We report on the first field experiment with a wild turtle community, involving introduced RES and a declining native species of conservation concern, the western pond turtle (*Emys marmorata*; WPT). Using a before/after experimental design, we show that after removing most of an introduced RES population, the remaining RES dramatically shifted their spatial basking distribution in a manner consistent with strong intraspecific competition. WPT also altered their spatial basking distribution after the RES removal, but in ways inconsistent with strong interspecific competition. However, we documented reduced levels of WPT basking post-removal, which may reflect a behavioral shift attributable to the lower density of the turtle community. WPT body condition also increased after we removed RES, consistent with either indirect or direct competition between WPT and RES and providing the first demonstration of RES competing with a native turtle in the wild. We conclude that the negative impacts on WPT basking by RES in natural contexts are more limited than suggested by experiments with captive turtles. However, native WPT appear to compete for food with introduced RES. Our results highlight the importance of manipulative field experiments when studying biological invasions.

¹ Department of Environmental Science, Policy, and Management, UC Berkeley, Berkeley, California, United States of America

² Department of Forestry and Natural Resources, University of Kentucky, Lexington, Kentucky, United States of America

³ Department of Biology, University of Hawaii at Manoa, Honolulu, Hawaii, United States

⁴ Department of Ecology and Evolutionary Biology & La Kretz Center for California Conservation Science, University of California, Los Angeles, Los Angeles, California, United States

⁵ Department of Ecology and Evolutionary Biology, Cornell University, Ithaca, New York, United States

⁶ Biology Department, Louisiana State University, Baton Rouge, Louisiana, United State of America

⁷ Urban Nature Research Center & Department of Herpetology, Natural History Musem of Los Angeles County, Los Angeles, California, United States of America



- 1 Full Title: Experimental removal of introduced slider turtles offers new insight into competition
- 2 with a native, threatened turtle

4 Short Title: Competition between introduced sliders and threatened turtles

5

- 6 Max R. Lambert^{1,2,*}, Jennifer M. McKenzie^{1,3}, Robyn M. Screen^{1,4}, Adam G. Clause⁵, Benjamin
- 7 J. Johnson⁶, Genevieve G. Mount⁷, H. Bradley Shaffer⁵, and Gregory B. Pauly⁸

8

- 9 ¹Authors contributed equally to the work and are listed alphabetically
- ²Department of Environmental Science, Policy, and Management, Mulford Hall, 130 Hilgard
- 11 Way, Berkeley, CA 94720 USA. *
- ³Department of Forestry and Natural Resources, University of Kentucky, 214 T.P. Cooper
- 13 Building, Lexington, KY 40546, USA.
- ⁴Department of Biology, University of Hawaii, Manoa, 2538 McCarthy Mall, Honolulu, HI
- 15 96822, USA.
- ⁵Department of Ecology and Evolutionary Biology, and La Kretz Center for California
- 17 Conservation Science, Institute of the Environment and Sustainability, University of California,
- Los Angeles, CA 90095, USA.





19	⁶ Department of Ecology and Evolutionary Biology, Cornell University, Corson Hall, Ithaca, NY
20	14853, USA.
21	⁷ Biology Department, Louisiana State University, 202 Life Sciences Building, Baton Rouge, La
22	70803, USA.
23	⁸ Urban Nature Research Center, and Department of Herpetology, Natural History Museum of
24	Los Angeles County, 900 Exposition Blvd, Los Angeles, CA 90007, USA.
25	*Correspondence author: email – <u>lambert.mrm@gmail.com</u> , Twitter – @MaxRLambert
26	
27	
28	
29	
30	
31	
32	
33	
34	
35	



37

38

39

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

Abstract: The red-eared slider turtle (*Trachemys scripta elegans*; RES) is often considered one of the world's most invasive species. Results from laboratory and mesocosm experiments suggest that introduced RES outcompete native turtles for key ecological resources, but such experiments can overestimate the strength of competition. We report on the first field experiment with a wild turtle community, involving introduced RES and a declining native species of conservation concern, the western pond turtle (*Emys marmorata*; WPT). Using a before/after experimental design, we show that after removing most of an introduced RES population, the remaining RES dramatically shifted their spatial basking distribution in a manner consistent with strong intraspecific competition. WPT also altered their spatial basking distribution after the RES removal, but in ways inconsistent with strong interspecific competition. However, we documented reduced levels of WPT basking post-removal, which may reflect a behavioral shift attributable to the lower density of the turtle community. WPT body condition also increased after we removed RES, consistent with either indirect or direct competition between WPT and RES and providing the first demonstration of RES competing with a native turtle in the wild. We conclude that the negative impacts on WPT basking by RES in natural contexts are more limited than suggested by experiments with captive turtles. However, native WPT appear to compete for food with introduced RES. Our results highlight the importance of manipulative field experiments when studying biological invasions.

Introduction

The International Union for Conservation of Nature (IUCN) has labeled the red-eared slider turtle (*Trachemys scripta elegans*; RES) one of the "world's worst invasive species" (Lowe et al. 2000). RES are native to the central United States but are now established on every continent except Antarctica, predominantly because of unwanted pet turtle releases (Kraus 2009; Rhodin et



al. 2017). Results from laboratory and mesocosm experiments suggest that RES can outcompete 59 native European and eastern United States freshwater turtles for food and basking sites (Cadi and 60 Joly 2003, 2004; Polo-Cavia et al. 2010, 2011; Pearson et al. 2015). While such controlled 61 experiments are informative, they can also inflate the effects of competition compared to in situ 62 field manipulations (Skelly 2002; Winkler and Van Buskirk 2012). Comparing laboratory and 63 64 mesocosm experiments with field manipulations is a critical step to a more complete understanding of the strength and mechanisms underlying species interactions in nature. 65 However, to our knowledge, no study has experimentally tested for competition between non-66 native RES and any native turtle species in the wild. 67 Basking sites are a key resource for thermoregulation, disease control, and reproduction 68 in freshwater turtles (Ernst and Lovich 2009), and previous ex situ experiments suggest that 69 basking sites are an important axis of competition between native turtles and introduced RES 70 (Cadi and Joly 2003; Polo-Cavia et al. 2010). Prior work in the University of California, Davis 71 72 Arboretum waterway (hereafter, UCD Arboretum) found that introduced RES and native western pond turtles (*Emys marmorata*; WPT) sometimes bask at the same sites (Fig 1), although they 73 tend to use basking sites that differ physically and spatially (Lambert et al. 2013). Whether these 74 75 basking site differences are the result of species-specific habitat choices or competition has never been resolved and requires an experimental approach. 76 77 Freshwater turtles, including both WPT and RES, are dietary generalists as adults and 78 consume a broad array of food items (Ernst and Lovich 2009). Even so, laboratory and mesocosm experiments suggest RES might directly interfere with native turtle food consumption 79 80 through aggressive behaviors or higher food consumption rates (Cadi and Joly 2004; Polo-Cavia 81 et al. 2011; Pearson et al. 2015). Additionally, if turtle densities are high for a given habitat,



83

84

85

86

87

88

89

90

91

92

93

94

95

96

97

98

99

100

101

102

103

104

exploitative competition could limit food availability, both intra- and interspecifically, and therefore decrease growth rates and / or body condition of native species.

Here, we present the results of an *in situ* field experiment in which we substantially reduced the introduced RES population at the UCD Arboretum to test for competition with WPT. For both WPT and RES, the UCD Arboretum represents a closed system that is well suited for experimental manipulations; natural immigration/emigration is not possible for freshwater turtles in this system, although occasional human-assisted transport does occur, particularly with RES released into the waterway. Our experiment is the first to explicitly test whether invasive species removal, a commonly-advocated management practice for invasive species including RES (Gaeta et al. 2015; García-Díaz et al. 2017), influences the basking behavior and body condition of a native turtle in the wild. If the distribution of WPT basking is a result of direct, competitive exclusion by RES from optimal basking sites, then RES removal should result in an increase of post-removal WPT basking at sites previously dominated by RES. Alternatively, if WPT basking activity does not significantly change in this manner after RES removal, then existing behavioral basking differences between the two species likely reflect species-specific habitat preferences, competitive superiority of WPT, or both. We also assessed WPT body condition pre- and postremoval as a proxy for whether removing RES improves WPT access to food resources. If introduced RES compete with WPT for food, then removing RES should result in an increase in WPT body condition. Given the broad overlap of these two species across California (Thomson et al. 2010, 2016; Fisher unpubl.), the range-wide imperilment of WPT (Spinks et al. 2003; Thomson et al. 2016), and the current Status Review for possible WPT listing under the U.S. Endangered Species Act (USFWS 2015), this experiment is directly relevant to ongoing WPT management actions.

105	
106	Methods
107	UC Davis IACUC Protocols #15263 and #16227 and California Department of Fish and Wildlife
108	Scientific Collecting Permits #2480, #4307, and #11663 approved this work.
109	
110	Turtle trapping and RES removal
111	Across the UCD Arcoretum, we deployed baited submersible traps in optimal habitat for both
112	RES and WPT over approximately 900 trap-nights from 10 July-1 August, 2011 and again from
113	13-29 September, 2011. We supplemented this trapping with dip netting, opportunistic hand
114	captures, a fyke net, and a basking trap. Dip netting and hand captures were targeted at RES but
115	other trapping was not. We removed and euthanized all RES, depositing most specimens at the
116	UC Davis School of Veterinary Medicine, the Natural History Museum of Los Angeles County,
117	or the UC Davis Museum of Wildlife and Fish Biology. We used linear regression to test
118	whether our trapping depleted the RES population over time by regressing cumulative RES
119	captures against trapping day for adult RES. Using likelihood ratio tests to assess model fits, we
120	compared a quadratic model, which would indicate population depletion, to a linear model,
121	which would indicate that the RES population was not leveling off with our removal effort.
122	
123	WPT Body Condition
124	To estimate changes in WPT body condition, we trapped for one week the following year from
125	27 May-2 June, 2012. In both 2011 and 2012 we measured WPT plastron length (notch-to-
126	notch; mm) and body mass (g) (Iverson and Lewis 2018). We used a linear mixed-effects model

(function 'lmer', R package "lme4") to test whether WPT body condition (i.e., differences in





129

130

131

132

133

134

135

136

137

138

139

140

141

142

143

144

145

146

147

148

mass controlling for body length) changed after the RES removal (Cadi and Joly 2003; Schulte-Hostedde et al. 2005; Litzgus et al. 2008). Our model of WPT mass controlled for plastron length and included treatment (pre- or post-removal) and sex as fixed effects and individual WPT (which were uniquely marked with scute notches) as a random effect to control for repeated measures. We used likelihood ratio tests to assess the significance ($\alpha < 0.05$) of fixed effects and removed non-significant variables from our model. We obtained full model conditional R² (cR²) for fixed and random effects combined and a marginal R² (mR²) for the model's fixed effects alone (function 'r.squaredGLMM', package "MuMIn"). Basking Site Monitoring We conducted binocular surveys of 24 pre-selected basking sites (Fig 1) for 34 total days, including 16 days pre-removal from 18 March—22 April 2010 (Lambert et al. 2013) and 18 days post-removal from 18 March—22 April 2012. Following Lambert et al. (2013), we performed all surveys between 1000 and 1500 hr to coincide with the expected maximum turtle basking activity during this time of year. We surveyed all sites once daily in rapid succession to avoid counting the same turtle at multiple sites. During each survey we recorded the number of species basking at each basking site as well as water temperature because we previously found that basking activity of both species increases with warmer water temperatures, more so than air temperatures (Lambert et al. 2013). We also obtained air temperature data from the UC Davis Russell Ranch Weather Station which is located ~4 km NW of the UCD Arboretum.

149 Modeling the Effects of RES Removal on Turtle Basking





151

152

153

154

155

156

157

158

159

160

161

162

163

164

165



We tested for changes in the proportion of WPT:RES basking across the UCD Arboretum preand post-RES removal using a generalized linear mixed effects model (GLMM) with a binomial family for proportion data (function 'glmer', R package "lme4"). We modeled WPT:RES basking as a function of treatment (pre- or post-removal) and the distance of each basking site from the west end of the UCD Arboretum because turtle basking distributions were previously shown to vary west-east (Lambert et al. 2013). We accounted for repeated measures by treating survey date as a random effect (Lambert et al. 2013). To explore site-specific changes in the proportion of the two species, we also used individual binomial GLMMs for each basking site. In addition, we modeled the absolute basking abundance of both species pre- and postremoval using Poisson GLMMs for count data. Our approach here was the same as with the binomial GLMM and, if an interaction was significant, we used individual GLMMs for each year to test the pattern and strength of turtle basking distributions across the UCD Arboretum in each year. To test whether certain basking sites made up larger or smaller proportions of total WPT basking observations pre- or post-removal, we used contingency tables, focusing on the five most heavily-used turtle basking sites (combined for both species) pre-removal (P, O, E, Q, and R) and site X, the most heavily-used turtle basking site post-removal.

166

167

169

170

171

Results:

168 Trapping and RES Removal

We removed and euthanized 177 RES (100.6 kg total biomass), including 28 adult males (16.3)

kg), 72 adult females (79.4 kg), and 77 juveniles (4.9 kg, defined as \leq 100 mm carapace length;

Ernst and Lovich 2009). A quadratic (rather than linear) model fit our data best (likelihood ratio





test p < 0.0001, full model $R^2 = 0.95$) and showed RES captures leveling off, signifying we had 172 removed a substantial fraction of the RES population. 173 We also captured, marked (or re-marked), and released 115 unique WPT (62.7 kg total 174 biomass) comprising 36 females (24.1 kg), 51 males (36.1 kg), and 28 juveniles (2.5 kg, defined 175 as ≤ 11 mm plastron length; Holland, 1991). The number of adult RES captures was ca 1.5 times 176 greater than that of WPT. Given the high population size and density of REstand given we 177 likely removed the majority of the RES population, our RES removal effort was likely 178 substantial enough to exert an effect on WPT if the two species compete for food or basking. 179 180 WPT Body Condition 181 While we trapped a large number of WPT in each year, we trapped 25 unique adult WPT in both 182 2011 and 2012; we used these 25 WPT for the body condition analysis. The body condition 183 linear model showed no interaction between treatment and sex (p = 0.92) and so we removed this 184 interaction from the model. Sex (p = 0.009), treatment (p < 0.001), and plastron length (p < 0.001) 185 0.0001) were significant (full model $cR^2 = 0.95$, $mR^2 = 0.86$). For a given plastron length, males 186 were on average 61.54 g ($\pm 23.55 \text{ g}$) heavier than females. Individual WPT varied in the degree 187 188 of body condition change post-removal (Fig 2A) and, on average, the 25 WPT measured before and after RES removal were 39.80 g (\pm 9.92 g) heavier for a given plastron length post-removal 189 (Fig 2B). The absence of an interaction between sex and treatment indicates that male and female 190 191 WPT responded similarly to RES removal. 192

193 Basking Site Monitoring



195

196

197

198

199

200

201

202

203

204

205

206

207

208

209

210

211

212

213

214

215

216

In 2010, we recorded 283 WPT and 645 RES observations, but only 43 WPT and 61 RES observations in 2012. Although the reduction in numbers of observed WPT was unexpected, we do not believe this reflects a decline in the WPT population. From 27 May-2 June 2012, we trapped 54 unique WPT over seven days and a Schnabel population estimate derived from trapping data suggest that ~109 WPT were present in the UCD Arboretum immediately after our post-removal surveys (McKenzie, Screen, and Pauly unpubl.); this estimate is similar to preremoval estimates of WPT population size. Given these estimates, we are confident that the WPT population was essentially unchanged during our experiment, and thus our focus on the relative proportions of turtles at monitored basking sites meaningfully reflects the impact of our removal experiment and not a catastrophic decline in WPT. The basking sites most commonly used by WPT pre-removal were generally the same sites used post-removal (Fig 3A, B). We recorded WPT basking at 15 of 24 basking sites preremoval, but at only 8 of 24 sites post-removal. WPT were absent from 8 sites they used preremoval (although of these, only 2 were frequently used pre-removal: sites A and N, and were present at one additional site they were not recorded using pre-removal (site B). We recorded RES basking at 17 of 24 basking sites pre-removal, and only 8 of 24 sites post-removal (Fig 3A, B). RES were absent from 9 sites they used pre-removal and were not recorded using new sites post-removal. Water temperatures were warmer (two-tailed t-test, p < 0.0001) in 2010 (17.0 C \pm 0.24 SE) than in 2012 (15.4 C \pm 0.36 SE). However, maximum daily air temperatures (averaged across all days of each survey period) were not different between years (two-tailed t-test, p = 0.74; 2010, 19.2 C ± 0.69 SE; 2012, 18.8 C ± 0.88 SE). Furthermore, in the two weeks prior to our surveys, maximum daily air temperatures were warmer in 2012 (18.8 C \pm 1.08 SE) than in





2010 (15.2 C \pm 0.65 SE); two-tailed t-test, p < 0.001). Air temperatures were also warmer in the 217 winter (beginning of December to end of February) preceding the post-removal survey than the 218 winter preceding the pre-removal survey (two-tailed t-test, p < 0.0001; 2009-2010, 13.4 C \pm 0.36 219 SE; 2011-2012, 15.5 C \pm 0.30 SE). Colder water temperatures may thus have contributed to the 220 lower overall turtle basking we observed in 2012, but this effect might have been modulated by 221 222 warmer air temperatures prior to our 2012 surveys. 223 Effects of RES Removal on Turtle Basking 224 The interaction between removal treatment and distance from the west end of the UCD 225 Arboretum was not significant (p = 0.18) and was removed from the model. Both treatment (p <226 0.0001) and distance from the west end (p < 0.0001) were retained (cR² = 0.31, mR² = 0.31). 227 The non-significant interaction indicates the removal did not change the ratio of 228 WPT:RES basking distributions across the UCD Arboretum. Both pre- and post-removal, the 229 basking distribution of turtles was WPT-biased in the west end and RES-biased in the east end 230 (Fig 4). However, the proportion of basking individuals that were WPT increased from 30.5% 231 pre-removal to 41.3% post-removal (p < 0.0001; Tukey's post-hoc test, function 'glht', package 232 233 "multcomp"). Individual binomial GLMMs for each basking site showed removal treatment effects on the WPT:RES basking ratio for site Q (p = 0.002, 9% WPT to 55% WPT) and a 234 235 marginal effect for site O (p = 0.09, 30% WPT to 75% WPT). All other individual basking sites 236 showed no differences (all p > 0.1). Pre-removal, the WPT basking distribution declined from west to east, and post-removal 237 238 WPT basking had a relatively flat distribution (Fig 5). We detected a shift in the absolute basking 239 distribution of WPT with a significant interaction between removal treatment and distance from



the west end (Poisson GLMM, p = 0.012, $cR^2 = 0.23$, $mR^2 = 0.06$). Individual GLMMs for each year indicated that distance from the west end was significantly associated with WPT basking abundance in the pre-removal year (p < 0.012, $cR^2 = 0.27$, $mR^2 = 0.03$) but not in the post-removal year (p = 0.55).

WPT predominantly used the same basking sites post-removal but showed a more even distribution across basking sites, with more basking activity at two center-east sites (Q and X) compared to before the RES removal (Fig 3A, B). Contingency table analyses showed that sites Q (p = 0.01) and X (p = 0.001) encompassed larger proportions of total WPT basking observations post-removal than pre-removal (Fig 3A, B). All other sites made up similar proportions pre- and post-removal (all p > 0.1), though some sites had generally low basking activity (Fig 3A, B), possibly limiting our power to detect shifts.

Our experimental removal of RES was associated with flatter observed distribution of WPT. Even so, if more eastern sites that were dominated by RES pre-removal (e.g., sites O, P, Q, and R) are also preferred WPT basking locations, then WPT should have increased basking at these sites post-removal. We did not see this shift. While our experiment suggests that WPT basking is influenced by RES densities, we did not observe a dramatic shift toward previously RES-dominated sites; thus, we did not find evidence of strong interspecific competition for those sites. Interspecific competition is greatest at higher densities and the effects of an introduced competitor can similarly manifest or become most pronounced when the introduced species is at high densities (Gurnell et al. 2004). Therefore, competition is presumably greatest at high densities of RES (or turtles generally) and perhaps influenced by the relative densities of both species.



After removal, remaining RES were sparse throughout much of the UCD Arboretum and concentrated in the east end (Fig 3B, 5). For RES, a Poisson GLMM indicated a significant interaction between treatment and distance to the west end (p < 0.0001, cR 2 = 0.30, mR 2 = 0.16). Individual GLMMs for each year showed a positive relationship between RES basking and the distance to the west end pre-removal (p < 0.0001, cR 2 = 0.27, mR 2 = 0.02) and post-removal (p < 0.0001, cR 2 = 0.14, mR 2 = 0.14). The distance of each basking site from the west end explained substantially more of the variation in RES basking abundance post-removal than pre-removal, indicating that remaining RES concentrated in the east end more strongly after we removed most of the RES population (Fig 3B, 5). Contingency table analyses indicated that sites E, O, P, and R comprised lower proportions of total RES basking observations after the removal and site X comprised a higher proportion (all p < 0.05). Site Q made up similar proportions of total RES observations in both years (p = 0.14).

The RES remaining post-removal abandoned several basking sites that they previously used heavily (particularly sites O and P) and shifted towards the east end of the UCD Arboretum (e.g., site X). This result suggests that RES prefer habitat at this end of the waterway and that, prior to our experiment, RES densities were high enough for intraspecific competition to force many RES into less preferred areas of the waterway. Our previous work showed that RES basking activity was highest at sites with shallow slopes, deeper water adjacent to the site, a steel mesh (rather than concrete or dirt) substrate, and high human activity (Lambert et al. 2013). Post-removal, RES basking activity was highest at the two sites (V and X) that maximized this combination of variables based on 2010 surveys (Fig 4B from Lambert et al. 2013).

Discussion



Our experimental removal dramatically altered both RES and total turtle density in the UCD Arboretum by eliminating over half of the turtles in the waterway, and offers new insights into competition for basking habitats and food between introduced RES and native WPT. Our removal experiment produced three important results.

First, the prevalence of basking turtles at our survey sites post-removal was about 15% of that pre-removal, and this reduction in basking observations was measured in both species. We have no evidence that the removal of RES negatively affected the WPT population size, and a follow-up trapping survey confirmed that the number of WPT present remained roughly constant. Rather, it appears that the overall lower density of turtles in the UCD Arboretum allowed many WPT to either shift their basking activity patterns, redistribute themselves to sites that we were not monitoring, or both. Environmental differences, including cooler water temperatures during our post-removal monitoring, may also explain the lower WPT basking numbers, although our previous results from the same site suggest that the water was warm enough for maximal basking activity in WPT (Lambert et al. 2013).

Second, after removing RES, we found that WPT basking activity at our monitoring sites shifted but did not increase at sites previously dominated by RES. Remaining RES concentrated their basking at sites (V and X) consistent with their previously identified preferred habitat characteristics (Lambert et al. 2013), suggesting that high RES densities prior to our experimental removal produced strong intraspecific competition, forcing many RES to use less-preferred basking habitat. While earlier laboratory and mesocosm experiments suggest introduced RES outcompete native turtles for basking sites and other resources, (Cadi and Joly 2003; Polo-Cavia et al. 2010; Pearson et al. 2015), our results suggest more subtle effects found





308

309

310

311

312

313

314

315

316

317

318

319

320

321

322

323

324

325

326

327

328

in complex, natural communities that are not predicted by simplified mesocosm experiments (Skelly 2002, Winkler and Van Buskirk 2012).

Third, we found that removing RES led to an increase in WPT body condition, suggesting that these turtle species compete for food. Whether this reflects interference competition (direct interactions between the two species), exploitation competition (both species indirectly competing for overlapping food resources), or a combination of the two is unclear. Experimental work on RES and other native turtles suggests RES may behaviorally prevent native turtles from obtaining sufficient food (Cadi and Joly 2004, Polo-Cavia et al. 2011, Pearson et al. 2015), and our experimental removal may have reduced such interference if it does exist in this population. However, we also removed a substantial portion of the overall turtle community thereby reducing the overall pressure on food resources in the system. Regardless of the mechanism, the ~40 g average increase in body condition we detected is substantial given that all WPT in our analysis pre-removal weighed under 1,100 g. To our knowledge, this result represents the first evidence from wild populations that introduced RES might compete with native turtles for food and that RES removal might improve the body condition of native turtles. Should We Remove RES to Benefit Declining Native Turtles? A recent summary of research goals for effective conservation of WPT (Thomson et al. 2016) identified the need for a clearer quantitative understanding of the impact of introduced RES. Controlling invasive species is a substantial commitment that rarely eliminates the entire population, particularly in situations with continual introductions (Kikillus et al. 2012, Gaeta et al. 2015, Garcia-Diaz et al. 2017), and removing 177 RES from the UCD Arboretum was an intensive effort requiring hundreds of person-hours of field work across 40 days. While our study



330

331

332

333

334

335

336

337

338

339

340

341

342

343

344

345

346

347

348

349

350

351

suggests that removing RES does influence native turtle basking ecology and feeding, the potential benefits with respect to short-term basking-site usage appear to be quantitatively modest. However, the substantial increase in WPT body condition during the year following the RES removal suggests that removing RES meaningfully increased resource availability for WPT. Whether these returns justify the effort may well depend on several variables, including RES abundance / density, attitudes of local human residents to introduced RES, disease risk (Héritier et al. 2017), other potential axes of competition (e.g., nesting sites), and other aspects of ecosystem health.

Our results also provide evidence that RES introductions may affect native turtles simply by inflating turtle densities in general (regardless of species identity). Therefore, removing RES may not necessarily relieve native turtles from a dominant competitor but, rather, may relieve ecological or behavioral pressures associated with high turtle densities and could conceivably result in unexpected responses by native species. The very low number of WPT basking observations post-removal here points to the unexpected consequence of removing over half of the turtle community, leading to changes in WPT behavior and habitat use that our experimental design, with fixed monitoring sites, failed to capture. Unlike many other freshwater turtles, WPT are aggressive baskers and prefer to bask alone or in low numbers (Bury and Wolfheim 1973). As such, reducing turtle densities may have allowed WPT to occupy other basking habitats in lower numbers as is their preference. Additionally, higher WPT body condition post-removal was likely influenced by there simply being fewer turtles overall competing for food in the UCD Arboretum. Improved body condition may also be the result of WPT adopting preferred basking behaviors, thereby improving digestive efficiency and mass gain. Future studies that include unmanipulated control sites, pre-removal surveys that span multiple years and account for year-



to-year variation, as well as a design that tracks the behavior of native turtles pre- and post-RES removal (e.g., using GPS-enabled radio transmitters) may better elucidate these unexpected outcomes on native turtles. Overall ar analyses suggest WPT responded to removing RES in a manner consistent with interspecific competition for food but inconsistent with strong interspecific competition for basking habitats, implying that removing RES may well be an important management strategy in some situations.

Directly managing basking habitat may be a tractable conservation activity for WPT in addition to non-native RES removal (Spinks et al. 2003, Thomson et al. 2016). In human-modified waterways, removal of floating basking sites for flood control and aesthetics (Spinks et al., 2003) could exacerbate competition for basking sites. Emerging research suggests that experimentally-added floating logs are preferred by WPT compared to bank-side basking sites and are more heavily used by WPT than RES, especially when they are isolated from human activities (Cossman et al. unpubl.). Adding artificial basking sites that favor WPT, alone or in combination with RES population reduction, is a simple, comparatively inexpensive manipulation that should be explored in future field experiments.

Study Limitations

The primary limitations of our study center on interpreting our basking results. We expected to observe fewer basking RES in the second year of study due to our intense removal effort but did not expect a concomitant decline in WPT observations. It is possible that water temperature, other environmental variation, or unforeseen consequences of our manipulation resulted in reduced overall turtle basking activity, or (more likely to us) radical shifts in basking to new and unmonitored locations, after the RES removal. Unfortunately, we cannot confidently identify which factor(s) resulted in fewer WPT basking observations. Although we studied both basking



and feeding, we also recognize that our experiment did not address other potentially important axes of competition that are important for the continued recruitment and persistence of WPT populations. While we employed a before-after comparative design, the use of unmanipulated control sites would have improved our ability to make stronger inferences in this study. We do not believe that the lower number of WPT basking observations confounds our results because our analyses of relative basking distribution differences between species and years across the waterway can accommodate sample size differences. Additionally, our analyses found that residual RES shifted their basking in intuitive ways (i.e., towards sites with preferred characteristics), increasing confidence in our results. While field experiments offer more biological realism than experiments in captivity, that added complexity may also yield unexpected results, such as changing a focal species' behaviors or habitat use.

Conclusions

We present the first large-scale field manipulation testing for basking site and food competition between non-native RES and native turtles. Consistent with expectations based on laboratory and mesocosm studies, RES removal increased WPT body condition and altered WPT basking activity. However, contrary to expectations, this change in basking was not consistent with intense competition between RES and WPT for individual basking sites in the UCD Arboretum. Our results offer evidence for intraspecific competition for food and basking sites at high RES densities, and suggest that reducing introduced RES densities allows WPT to access more food and occupy a broader range and distribution of basking sites. We encourage other researchers to replicate our field-based experiment, perhaps using control sites or multiple years of pre-removal observations. These modifications to our protocol would improve the ability to interpret



398	competition between RES and native turtles and the magnitude of behavioral shifts that occur
399	when removals lead to changes in both relative and absolute turtle densities.
400	
401	Data Availability: Basking and body size data are available in the electronic supplementary
402	material.
403	Acknowledgments: This work was started while all authors were at UC Davis. The UCD
404	Arboretum staff, members of the Peter Moyle and Janet Foley labs, Nicholas Buckmaster,
405	Lauren Cassidy, Jillian Howard, Anna Jordan, Brian Mahardja, Hilary Rollins, Bryce Sullivan,
406	and Matthew Young provided invaluable assistance.
407	
408	
409	References
410	Bury, R. B. and J. H. Wolfheim. 1973. Aggression in free-living pond turtles (<i>Clemmys</i>
411	marmorata). BioScience 23: 659–662.
412	Cadi, A. and P. Joly. 2003. Competition for basking places between the endangered European
413	pond turtle (Emys orbicularis galloitalica) and the introduced red-eared slider
414	(Trachemys scripta elegans). Canadian Journal of Zoology 81: 1392–1398.
415	Cadi, A. and P. Joly. 2004. Impact of the introduction of the red-eared slider (<i>Trachemys scripta</i>)
416	elegans) on survival rates of the European pond turtle (Emys orbicularis). Biodiversity
417	and Conservation 13: 2511–2518.
418	Ernst, C. H. and J. E. Lovich. 2009. Turtles of the United States and Canada. Johns Hopkins
419	University Press, 2 nd Ed. 840 pages.



Gaeta, J. W., Hrabik, T. R., Sass, G. G., Roth, B. M., Gilber, S. J., and M. J. Vander Zanden. 420 2015. A whole-lake experiment to control invasive rainbow smelt (Actinopterygii, 421 Osmeridae) via overharvest and a food web manipulation. Hydrobiologia 746: 433–444. 422 García-Díaz, P., Ramsey, D. S. L., Woolnough, A. P., Franch, M., Llorente, G. A., Montori, A., 423 Buenetxea, X., Larrinaga, A. R., Lasceve, M., Álvarez, A., Traverso, J. M., Valdeón, A., 424 Crespo, A., Rada, V., Ayllón, E., Sancho, V., Lacomba, J. I., Bataller, J. V., and M. 425 426 Lizana. 2017. Challenges in confirming eradication success of invasive red-eared sliders. Biological Invasions 19: 2739–2750. 427 Gurnell, J., Wauters, L. A., Lurz, P. W. W., and G. Tosi. 2004. Alien species and interspecific 428 competition: effects of introduced eastern grey squirrels on red squirrel population 429 430 dynamics. Journal of Animal Ecology 73: 26–35. 431 Héritier, L., A. Valdeón, A. Sadaoui, T. Gendre, S. Ficheux, S. Bouamer, N. Kechemir-Issad, L. Du Preez, C. Palacios, and O. Verneau. 2017. Introduction and invasion of the red-eared 432 slider and its parasites in freshwater ecosystems of southern Europe: Risk assessment for 433 the European pond turtle in wild environments. Biodiversity and Conservation 26: 1817– 434 1843. 435 Holland, D. C. 1991. A Synopsis of the Ecology and Status of the Western Pond Turtle Clemmys 436 marmorata in 1991. Report to National Ecological Research Center. United States Fish 437 and Wildlife Service, San Simeon, CA. 438 Iverson, J. B. and E. L. Lewis. 2018. How to measure a turtle. Herpetological Review 49: 453-439 460. 440



Kikillus, K. H., Hare, K. M., and S. Hartley. 2012. Online trading tools as a method of estimating 441 propagule pressure via the pet-release pathway. Biological Invasions 14: 2657–2664. 442 Kraus, F. 2009. Alien Reptiles and Amphibians, a Scientific Compendium and Analysis. 443 Springer, Netherland. 563 pp. 444 Lambert, M. R., Nielsen, S. N., Wright, A. N., Thomson, R. C., and H. B. Shaffer. 2013. Habitat 445 features determine the basking distribution of introduced red-eared sliders and native 446 western pond turtles. Chelonian Conservation and Biology 12: 192–199. 447 Litzgus, J. D., Bolton, F., and A. I. Schulte-Hostedde. 2008. Reproductive output depends on 448 449 body composition in spotted turtles (*Clemmys guttata*), Copeia 2008: 86–92. 450 Lowe, S., Browne, M., Boudjelas, S., and M. De Poorter. 2000. 100 of the World's Worst Invasive Alien Species. A Selection from the Global Invasive Species Database. The 451 Invasive Species Specialist Group (ISSG) of the Species Survival Commission (SSC) of 452 the World Conservation Union (IUCN), 12 pp. 453 Pearson, S. H., Avery, H. W., and J. R. Spotila. 2015. Juvenile invasive red-eared slider turtles 454 negatively impact the growth of native turtles: implications for global freshwater turtle 455 456 populations. Biological Conservation 186: 115–121. Polo-Cavia, N., Lopez, P., and J. Martin. 2008. Interspecific differences in responses to predation 457 risk may confer competitive advantages to invasive freshwater turtle species. Ethology 458 114: 115–123. 459 Polo-Cavia, N., Lopez, P., and J. Martin. 2010. Competitive interactions during basking between 460 native and invasive freshwater turtle species. Biological Invasions 12: 2141–2152. 461



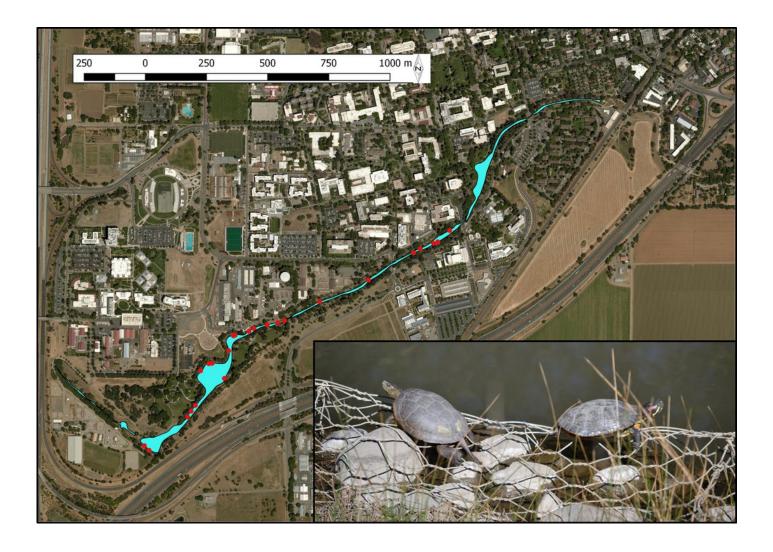
Polo-Cavia, N., Lopez, P., and J. Martin. 2011. Aggressive interactions during feeding between 462 native and invasive freshwater turtles. Biological Invasions 13: 1387–1396. 463 Rhodin, A. G. J., Iverson, J. B., Bour, R., Fritz, U., Georges, A., Shaffer, H. B., and P. P. van 464 Dijk. 2017. Turtles of the World. Annotated checklist and atlas of taxonomy, synonymy, 465 distribution, and conservation status (8th Ed.). Chelonian Research Monographs 7: 1–292. 466 Schulte-Hostedde, A. I., Zinner, B., Millar, J. S., and G. J. Hickling. 2005. Restitution of mass-467 size residuals: validating body condition indices. Ecology 86: 155–163. 468 Skelly, D. K. 2002. Experimental venue and estimation of interaction strength. Ecology 83: 469 470 2097-2101. 471 Spinks, P. Q., Pauly, G. B., Crayon, J. J., and H. B. Shaffer. 2003. Survival of the western pond turtle (*Emys marmorata*) in an urban California environment. Biological Conservation 472 113: 257–267. 473 Thomson, R. C., Spinks, P. Q., and H. B. Shaffer. 2010. Distribution and abundance of invasive 474 red-eared sliders (Trachemys scripta elegans) in California's Sacramento River basin and 475 possible impacts on native western pond turtles (*Emys marmorata*). Chelonian 476 Conservation and Biology 9: 297–302. 477 Thomson, R. C., Wright, A. N., and H. B. Shaffer. 2016. California Amphibian and Reptile 478 Species of Special Concern. University of California Press, Oakland, CA. 408 pp. 479 United States Fish & Wildlife Service. 2015. Endangered and threatened wildlife and plants; 90-480 day findings on 10 petitions. Federal Register 80: 19259–19263. 481



- Winkler, J. D. and J. Van Buskirk. 2012. Influence of experimental venue on phenotype:
- multiple traits reveal multiple answers. Functional Ecology 26: 513–521.

The UC Davis Arboretum waterway, turtle basking sites, and basking turtles.

The UC Davis Arboretum waterway is in blue and turtle basking sites are displayed as red circles. Inset shows a native western pond turtle (left) and an introduced red-eared slider (right) basking side-by-side in the Arboretum. Map data © 2019 Bing. Photo credit Max Lambert.

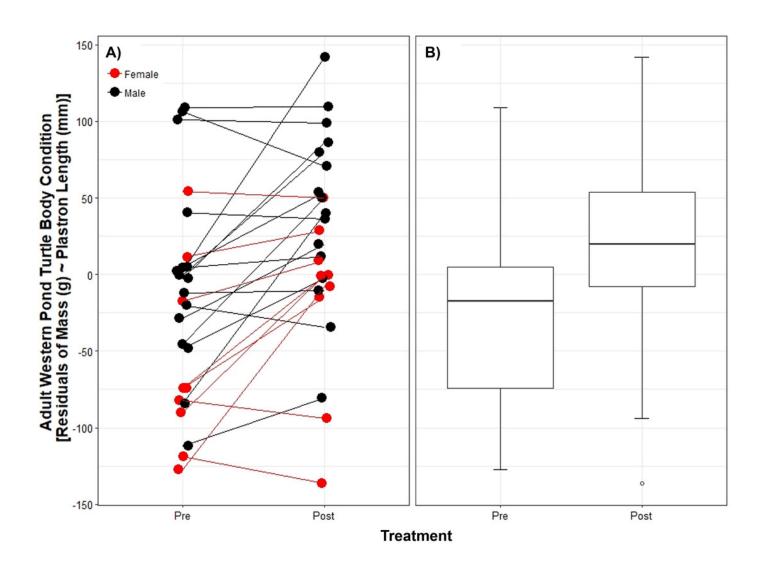




Native western pond turtle (WPT) body condition before and after introduced red-eared slider (RES) removal.

Body condition is shown as the residuals of body mass regressed against plastron length. Individual WPT varied in their body condition response to introduced RES removal (A) but body condition generally improved. On average (B) WPT are 39.80 g heavier after RES removal. Boxplot hinges show the 25th and 75th body condition percentiles, whiskers show the extent of data within 1.5 times the interquartile range, and the center line is the median for each treatment year pre- and post-removal.



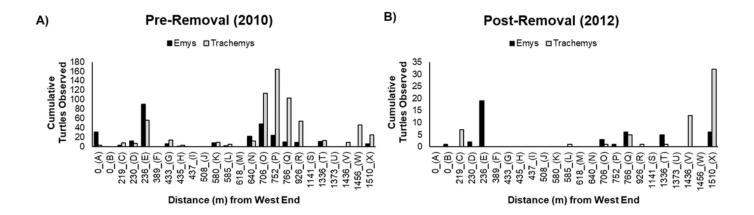






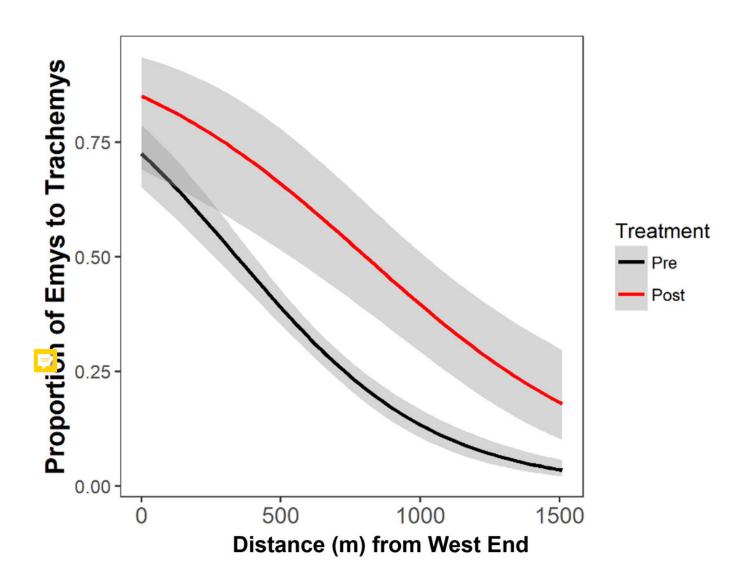
Cumulative basking observations of native WPT (Emys) and introduced (RES).

Basking observations before (A) and after (B) the RES removal are arrayed along a west-east gradient in the UCD Arboretum. Letters under the x-axis are basking site identifiers.



The proportion of native WPT (Emys) to introduced RES (Trachemys) basking across the UCD Arboretum waterway pre- and post-removal.

Curves are modeled proportions of the WPT to RES basking along a west-east gradient in the UCD Arboretum pre- and post-removal (black and red curves, respectively). The proportion of WPT to RES basking along the waterway was similarly WPT-biased in the west and RES-biased in the east in both years. WPT basking observations were higher after the RES removal.



The total number of native WPT (*Emys*) and introduced RES (*Trachemys*) basking along the UCD Arboretum.

Curves are the modeled daily number of WPT and RES along a west-east gradient in the UCD Arboretum pre- and post-removal (black and red curves, respectively). WPT displayed a more even basking distribution after the RES removal and RES concentrated basking activity towards the east end of the Arboretum after most of their population was removed.

