

Effects of temperature and salinity stress on DNA methylation in a highly invasive marine invertebrate, the colonial ascidian *Didemnum vexillum*

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Environmentally induced epigenetic changes may contribute to phenotypic plasticity, increase adaptive potential in changing environments, and play a key role in the establishment and spread of invasive species in new habitats. In this study, we used Methylation Sensitive Amplified Length Polymorphism (MS-AFLP) to assess environmentally induced DNA methylation changes in a globally invasive clonal ascidian, *Didemnum vexillum*. We tested the effect of increasing temperature (19, 25 and 27°C) and decreasing salinity (34, 32, 30, 28 and 26 practical salinity units (PSU)) on global DNA methylation, growth and survival rates. Exposure to 27°C resulted in significant changes in DNA methylation over time, while there were no significant changes in non-methylated loci (representing genetic variation). Growth also decreased in colonies exposed to high temperatures, suggesting they were under thermal stress. In contrast, no differences in growth or DNA methylation patterns were observed in colonies exposed to a decreasing salinity gradient, potentially due to prior adaptation to conditions experienced at the site of collection. The results of this study show that environmental stress can induce significant global DNA methylation changes in an invasive marine invertebrate on very rapid timescales, and that this response varies depending on the type, magnitude, and duration of the stressor. Changes in genomic DNA methylation and the rate of growth may act to 'buy survival time' until conditions improve, and determine the distribution limits of this globally invasive species.

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23 Abstract

24 Environmentally induced epigenetic changes may contribute to phenotypic plasticity, increase
25 adaptive potential in changing environments, and play a key role in the establishment and spread
26 of invasive species in new habitats. In this study, we used Methylation-Sensitive Amplified
27 Fragment Length Polymorphism (MS-AFLP) to assess environmentally induced DNA
28 methylation changes in a globally invasive colonial ascidian, *Didemnum vexillum*. We tested the
29 effect of increasing temperature (19, 25 and 27°C) and decreasing salinity (34, 32, 30, 28 and 26
30 practical salinity units (PSU)) on global DNA methylation, growth and survival rates. After three
31 days of exposure to elevated temperature, significant DNA methylation differences were
32 observed between treatments. Exposure to 27°C resulted in changes in DNA methylation over
33 time, while there were no significant changes in non-methylated loci (representing genetic
34 variation). Growth also decreased in colonies exposed to high temperatures, suggesting they
35 were under thermal stress. In contrast, no differences in growth or DNA methylation patterns
36 were observed in colonies exposed to a decreasing salinity gradient, potentially due to prior
37 adaptation to conditions experienced at the site of collection. The results of this study show that
38 environmental stress can induce significant global DNA methylation changes in an invasive
39 marine invertebrate on very rapid timescales, and that this response varies depending on the type,
40 magnitude, and duration of the stressor. Changes in genomic DNA methylation and the rate of
41 growth may act to ‘buy survival time’ until conditions improve, and determine the distribution
42 limits of this globally invasive species.

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46 Introduction

47 Species invasions, climate change, habitat fragmentation and environmental degradation are
48 altering ecosystems and threatening biodiversity (Leadley 2010). A key question in evolutionary
49 biology is whether species will be able to adapt in response to these human-driven environmental
50 changes (Visser 2008). Biological invasions can provide a unique model to investigate
51 adaptation and evolution within short timescales, as introduced species must rapidly adapt to new
52 habitats (Allendorf & Lundquist 2003; Sakai et al. 2001). It has been suggested that epigenetic
53 mechanisms could play a critical role in environmental adaptation, and may be particularly
54 important for the success of invasive species (Estoup et al. 2016; Pérez et al. 2006; Prentis et al.
55 2008). Recently introduced populations frequently have reduced genetic diversity (e.g., genetic
56 bottlenecks and founder effects) (Dlugosch & Parker 2008), which is thought to constrain the
57 colonisation potential of a species (e.g., Crawford & Whitney 2010). Despite this, invasive
58 species can still be highly successful in their new environments, and often outcompete locally
59 adapted native species (Allendorf & Lundquist 2003). By increasing both phenotypic plasticity
60 and heritable variation, epigenetic changes might allow invasive species to quickly respond to
61 environmental challenges. However, the role that epigenetic mechanisms play during the process
62 of invasion is only beginning to be understood (Hawes et al. 2018) and, for many species, the
63 effect of environmental stressors on DNA methylation is unknown.

64

65 Epigenetic modifications have been shown to respond to environmental cues and, in some cases,
66 be associated with significant phenotypic change (Dias & Ressler 2014; Kucharski et al. 2008;
67 Waterland & Jirtle 2003). Epigenetic mechanisms are diverse and interactive (e.g., DNA
68 methylation, histone modifications, small RNAs), but all alter gene expression without the

69 requirement for changes in the underlying DNA nucleotide sequences (Bossdorf et al. 2008).
70 Currently the most studied epigenetic mechanism is the methylation of cytosine nucleotides to
71 form 5 methyl-cytosine (DNA methylation). DNA methylation is common in eukaryotes, and
72 there are a range of methods for its detection and quantification (Plongthongkum et al. 2014).
73 One such method, Methylation-Sensitive Amplified Fragment Length Polymorphism (MS-
74 AFLP), allows for cost-effective screening of variation in global DNA methylation, without the
75 requirement for a reference genome (Reyna-Lopez et al. 1997). The MS-AFLP technique enables
76 epigenetic research in non-model organisms and can provide a first look at DNA methylation-
77 environment interactions, which may underlie adaptive plasticity. Interest in the ecological
78 relevance of DNA methylation in non-model organisms is growing, and marine invertebrates
79 have been identified as an emerging taxonomic group for studies of ecological epigenetics,
80 particularly in the context of environmental change (Hofmann 2017). Despite this, few studies
81 have investigated environmentally induced epigenetic changes in marine invertebrates (Marsh et
82 al. 2016; Marsh & Pasqualone 2014; Putnam et al. 2016), and the role of epigenetic mechanisms
83 in the success of marine invertebrate invaders is only beginning to be explored (Ardura et al.
84 2017; Huang et al. 2017; Pu & Zhan 2017).

85

86 Of the marine invertebrates, colonial ascidians stand out as model species to study both
87 environmentally induced DNA methylation changes (Hawes et al. 2018) and invasion success
88 (Zhan et al. 2015). Colonial ascidians (phylum Chordata) are common marine invaders
89 worldwide, particularly in habitats perturbed by human activities, e.g., marinas, ports,
90 aquaculture structures (Lambert 2001). Due to the considerable ecological and economic damage
91 caused by ascidian invasions, they have become a prominent study species in the field of

92 invasion biology (Zhan et al. 2015). Ascidians can thrive in a variety of environmental
93 conditions, and display unique biological characteristics, including a broad tolerance to common
94 environmental stressors such as temperature and salinity (Rocha et al. 2017). Additionally, the
95 germ-cell lineages of colonial ascidians originate from somatic-cell lineages, contrasting with the
96 germ-cell lineage sequestration found in vertebrate Chordates (Rosner et al. 2009). Having no
97 true germ-cell lineage sequestration increases the likelihood that stress-induced epigenetic
98 modifications induced in somatic cells can be passed on to gametes (Verhoeven & Preite 2014).
99 Finally, asexual reproduction (by budding) leads to colonies of genetically identical individual
100 animals (termed zooids) that all share the same DNA nucleotide sequences (genotype). Clonal
101 reproduction allows for genetically identical replicates across environment stress treatments, and
102 repeated sampling of the same individual at multiple time-points. This reduces the confounding
103 effects of genetic variation that frequently complicate epigenetic studies of non-clonal organisms
104 (Douhovnikoff & Dodd 2015; Verhoeven & Preite 2014).

105

106 Invasive populations of the colonial ascidian, *Didemnum vexillum* Kott, 2002 have extremely
107 low levels of genetic diversity compared to populations within its native range (Stefaniak et al.
108 2012). Despite this, *D. vexillum* is extremely successful where it has invaded (Beveridge et al.
109 2011; Cohen et al. 2011; Griffith et al. 2009; Hitchin 2012; Lambert et al. 2009; Tagliapietra et
110 al. 2012), often forming large colonies that smother other marine invertebrates, including
111 commercial aquaculture species (Fletcher et al. 2013b). We used MS-AFLP to determine
112 whether a) DNA methylation is present in the genome of *D. vexillum*, and b) if genome-wide
113 DNA methylation patterns in *D. vexillum* change in response to two prominent types of
114 environmental stress: temperature and salinity. Temperature and salinity are often reported as the

115 most important environmental determinants controlling the distribution of marine species, and
116 these two parameters have been used repeatedly when studying the tolerance of ascidians to
117 environmental stress (Dybern 1967; Gröner et al. 2011; Renborg et al. 2014; Serafini et al. 2011;
118 Zerebecki & Sorte 2011). Furthermore, extreme climatic events such as precipitation events and
119 heatwaves are expected to increase in frequency in the near future (IPCC 2014), making it
120 increasingly important to evaluate the response and resilience of marine invertebrates to thermal
121 and osmotic stress.

122

123 **Materials and Methods**

124 *Sample collection and establishment of experimental colonies*

125 Colonies of *D. vexillum* were collected from the Nelson Marina (South Island, New Zealand;
126 41°15'38"S, 173°16'54"E) in April 2016. At the time of collection, the water temperature was
127 19°C and the salinity was 33 PSU. Colonies were gently removed from wharf pilings and
128 immediately placed in labelled 2 L plastic containers filled with ambient seawater for transport to
129 the Cawthron Institute (less than 5 minutes commute). Attempts were made to remove whole
130 colonies that were free from debris, but as the colonies were mostly small and growing flat over
131 piles covered in other fouling, colonies often broke apart during removal. We treated these
132 fragments as one colony for the following experiments but it is possible that it was two or more
133 colonies growing in close proximity. After arrival at the laboratory, the colonies were gently
134 cleaned with seawater to remove mud, silt and other organisms, and approximately equal sized
135 fragments (c. 1.5 cm x 1.5 cm) were cut with a razor blade. Colony fragments were then placed
136 on glass slides and gently wrapped with cotton thread to encourage attachment (Rinkevich &
137 Fidler 2014). Glass slides were inserted into slide holders and placed in pre-conditioned 40 L

138 glass aquaria in ambient, control conditions (19°C, 34 PSU) for one week to allow attachment to
139 occur. During this acclimation period, colonies were fed 1.6×10^8 cells.L⁻¹ of cultured algae
140 (*Isochrysis galbana*) every second day. Following attachment, the cotton was removed and the
141 colony fragments were randomly allocated (Temperature: n = 9, Salinity: n = 15) to pre-
142 conditioned treatment tanks (n = three tanks per treatment, one colony fragment per treatment
143 tank). All treatment tanks were maintained at 19°C and 34 PSU for a further two weeks
144 acclimation time prior to beginning the experiments.

145

146 *Experimental system*

147 Seawater for the experiments was collected from Tasman Bay, Nelson (41°11'29.2"S
148 173°21'01.9"E), passed through three filters (pore size 50, 5, and 0.35 µm) and ultraviolet light
149 treated. Each tank was filled with freshly collected seawater at the start of the experiment, and
150 five L water exchanges were done daily throughout the duration of the experiment with pre-
151 heated or reduced salinity seawater. Each day following water exchange, colonies were fed a diet
152 of 1.6×10^8 cells.L⁻¹ of *I. galbana*. To prevent stratification, mixing in experimental tanks was
153 ensured by gentle aeration using air stones. Experimental tanks were exposed to a 14:10 hour
154 light:dark cycle to mimic summer conditions. Water temperatures were maintained using
155 thermostatically regulated aquarium heaters (EHEIM JAIGER 100W, Deizisau, Germany) and
156 salinity treatments were achieved and maintained by the addition of reverse osmosis (RO) water
157 to seawater. Water temperature and salinity were measured twice daily using hand-held probes to
158 ensure stable treatment conditions were maintained $\pm 0.5^\circ\text{C}$ or 0.5 PSU (YSI Professional Plus,
159 YSI Incorporated, Yellow Springs Ohio, USA). After two weeks of acclimation, tissue samples
160 were collected from all colonies for MS-AFLP analyses (Time 0; T0). Temperatures were then

161 increased by 1°C per day for temperature treatments, and salinity was reduced by 1 PSU per day
162 for the salinity treatments until, after eight days, all treatment conditions were reached.
163 Treatments were as follows: temperature = 19 (Control), 25, and 27°C and salinity= 34
164 (Control), 32, 30, 28, and 26 PSU. Temperature and salinity experiments were run in parallel.
165 Tissue samples were taken at day eight for MS-AFLP analyses (Time 1; T1). Colony fragments
166 were then maintained in experimental treatments for a further three days (day 11), at which time
167 final tissue samples were taken for MS-AFLP analyses (Time 2; T2).

168

169 *Sampling protocol*

170 Prior to tissue collection, colonies were not fed for 16 hours to minimise contamination by feed
171 microalgae. To collect tissue for MS-AFLP analyses, small (c. 5 mm x 5 mm) samples were
172 taken from each colony using a sterile razorblade. Tissue samples were preserved in 95%
173 ethanol, which was refreshed once before storage at -20°C until processing. Pre- and post-tissue
174 sampling, photos were taken for growth rate calculations. Colony growth rates (quantified by
175 changes in colony surface-area over time) were measured using Image J 1.48v software
176 (Schneider et al. 2012). The survival of colonies was also monitored at the time of sampling by
177 assessments of colony health, including zooid integrity, colour and texture of the colony, build-
178 up of detritus and dead tissue. Throughout the experiment, all colonies were transported, sampled
179 and photographed while submerged in trays of temperature and salinity adjusted seawater to
180 minimise handling stress.

181

182 *MS-AFLP analysis*

183 To assess whole genome DNA methylation patterns, DNA was extracted using G-spin Total
184 DNA extraction kits (animal tissue protocol; Intron, Gyeonggi-do, South Korea). Following
185 DNA extraction, in parallel reactions (25 μ L final reaction volume), 500 ng of DNA was
186 digested with 10 U of each restriction enzyme (MspI and EcoRI or HpaII and EcoRI; New
187 England BioLabs, Ipswich, MA, USA) and 10X CutSmart Buffer (England BioLabs), and
188 incubated at 37°C for 2 hours followed by 80°C for 20 min to inactivate the enzymes. The
189 digested DNA were ligated in a final volume of 20 μ L containing 1 U of T4 DNA ligase (New
190 England BioLabs), 10X ligase buffer, 250 nM of EcoRI adaptor, and 2.5 μ M of MspI or HpaII
191 adaptor for 3 hours at 37°C. Table 1 presents a list of all adaptor and primer sequences used for
192 the MS-AFLP protocol. Pre-selective polymerase chain reaction (PCR) was performed in a total
193 volume of 20 μ L using 8 μ L of ligated DNA, MyTaqTM 2X PCR master mix (Bioline, MA,
194 USA), and 500 nM of each pre-selective primer. Thermocycling conditions were 20 cycles of:
195 94°C for 30 s, 56°C for 60 s, and 72°C for 60 s. Selective PCR was performed using four EcoRI
196 and MspI/HpaII primer combinations in a final volume of 20 μ L using 1 μ L of pre-selective PCR
197 product, MyTaqTM 2X PCR master mix (Bioline, MA, USA), and 500 nM of each combination
198 of forward and reverse selective primers. Thermocycling conditions were 1 cycle of: 94°C for 2
199 min; 10 cycles of 94°C for 30 s, 65°C for 30 s (decreasing 1°C per cycle), and 72°C for 60 s; and
200 30 cycles of 94°C for 30 s, 56°C for 30 s, and 72°C for 60 s; and a hold cycle of 72°C for 30
201 min. The resulting selective PCR product was diluted 1:5 with sterile distilled water and
202 analysed using an ABI 3130 capillary sequencer (Applied Biosystems, Foster City, CA, USA)
203 with internal size standards (GS600LIZ) by an external contractor (Genetic Analysis Services,
204 University of Otago, Dunedin, New Zealand).

205

206 *Data analysis*

207 PEAKSCANNER software (Applied Biosystems, Foster City, CA, USA) was used to assign the
208 MS-AFLP fragments peak height and size. To determine the parameters for subsequent analysis,
209 the MS-AFLP procedure described above was first repeated three times for one individual
210 sample. The following settings were associated with the lowest error rate between replicates and
211 were applied to the *msap* analysis (Pérez-Figueroa 2013) described below. Error rate per primer,
212 0.07; analysis range, 50 – 500 base pairs (bp); minimum peak height, 1200 relative fluorescence
213 units. Peak presence/absence data corresponding to HpaII and MspI fragments was then
214 converted to a binary matrix (presence = 1, absence = 0), so that the methylation state of each
215 restriction site could be identified. MS-AFLP profiles were assessed using the R package *msap* v.
216 1.1.8 (Pérez-Figueroa 2013). The *msap* package determines whether individual fragments (loci)
217 are methylated (MSL) by analysing the contents of the binary matrix, and comparing differences
218 representing the differential sensitivities of HpaII and MspI to cytosine methylation (Table 2).
219 From this, DNA methylation profiles of control and experimental samples were assessed by
220 means of principal coordinate analyses (PCoA) followed by analyses of molecular variance
221 (AMOVA) (Excoffier et al. 1992). Colony growth rates (mean growth per day, mm \pm 1 s.e.)
222 were assessed using standard ANOVA and post-hoc pairwise comparisons were made using
223 Tukey's honest significance difference (HSD) test.

224

225 **Results**226 *Temperature stress experiments*227 *Growth rates*

228 All colonies survived elevated temperature exposure, and colony growth rates (Figure 1) were
229 significantly different between treatments (single level ANOVA, $F(2, 6) = 7.82$, $p = 0.0213$;
230 Table S1). Post hoc comparisons using Tukey HSD indicated that growth per day was
231 significantly reduced when colonies were exposed to 27°C compared to colonies grown at 19°C
232 ($Diff = -61.43$, $p = 0.01$; Table S2, Figure 1). Growth was not significantly different between
233 colonies exposed to 25°C and 27°C or 19°C and at 25°C.

234

235 *Whole genome DNA methylation patterns (MS-AFLP)*

236 Using four primer combinations, 1157 fragments (loci) were produced and analysed. At
237 sampling time zero (T0), prior to temperature treatment exposure, 586 of the 1157 loci were
238 MSL, of which 178 were polymorphic (30%). There were no significant differences in DNA
239 methylation (Figure 2A) (MSL, $\phi_{ST} = -0.04981$, $p = 0.7093$, single level AMOVA; Table S3)
240 between treatment groups. Following eight days of gradual temperature increase, (T1), of the
241 1157 loci analysed 613 were MSL, of which 172 were polymorphic (28%). There were still no
242 significant differences between DNA methylated (MSL) loci (Figure 2B) ($\phi_{ST} = 0.05606$, $p =$
243 0.2176 , single level AMOVA; Table S3) between the three temperature treatments. However,
244 after three days of exposure to elevated temperature, of 1157 loci 600 were MSL, of which 201
245 were polymorphic (34%) and statistically significant differences in DNA methylation were
246 evident between treatment groups (Figure 2C) (MSL, $\phi_{ST} = 0.1585$, $p = 0.0215$, single level
247 AMOVA; Table S3). There were no significant global methylation changes between sampling
248 time points in colonies held at 19°C (Figure 2D) (MSL, $\phi_{ST} = -0.0212$, $p = 0.6019$, single level
249 AMOVA; Table S5) or 25°C (Figure 2E) (MSL, $\phi_{ST} = -0.01373$, $p = 0.5637$, single level
250 AMOVA; Table S4). In contrast, there were significant DNA methylation changes following

251 exposure to 27°C (Figure 2F) (MSL, $\phi_{ST} = 0.1727$, $p = 0.0223$, single-level AMOVA; Table
252 S4). Variation between individuals was also reduced in the 27°C treatment group, with DNA
253 methylation patterns becoming more similar, as visualised by the reduced spread of samples
254 around the centroid in the PCoA for each temperature (Figure 2C).

255

256 *Salinity stress experiments*

257 *Growth rates*

258 Akin to the temperature treatment, all colonies survived decreased salinity exposure. Colony
259 growth rates were not significantly different between treatment groups (Figure 3) (single level
260 ANOVA, $F(4, 10) = 2.066$, $p = 0.161$; Table S5).

261

262 *Whole genome DNA methylation patterns (MS-AFLP)*

263 Using four primer combinations, 1050 loci were produced and analysed. At sampling time zero
264 (T0), prior to differential salinity exposure, of these, 626 were methylation sensitive loci (MSL),
265 of which 169 were polymorphic (27%). There were no significant differences in DNA
266 methylation between treatment groups (Figure 4A) (MSL, $\phi_{ST} = -0.05023$, $p = 0.686$, single level
267 AMOVA; Table S6). Following the gradual salinity decrease, at time one (T1), of 1050 loci 654
268 were MSL, of which 173 were polymorphic (26%). There were no significant differences in
269 DNA methylation (Figure 4B) (MSL, $\phi_{ST} = 0.02674$, $p = 0.3518$; single level AMOVA; Table
270 S6), between salinity treatments. After three days of exposure to elevated salinity, of 1050 loci,
271 682 were MSL, of which 95 were polymorphic (14%). Non-significant differences in DNA
272 methylation remained (Figure 4C) (MSL, $\phi_{ST} = -0.06389$, $p = 0.8328$, single level AMOVA).

273 There were no significant methylation differences (Table S7) in any of the salinity treatments
274 over time (Figure 4D, E, F, G H).

275

276 **Discussion**

277 Changes in DNA methylation may be one of the mechanisms by which invasive species can
278 rapidly adapt to new environments. However, for many species, the responsiveness of DNA
279 methylation to environmental challenges has not yet been tested. Our results indicate that
280 environmental stressors can induce significant global DNA methylation changes in an invasive
281 marine invertebrate on very rapid timescales, and that this response varies depending on the type,
282 magnitude, and duration of the stressor. After three days of exposure to elevated temperature,
283 significant changes in whole-genome patterns of DNA methylation had occurred in *D. vexillum*
284 colonies held at 27°C. In contrast, DNA methylation patterns in colonies exposed to 25°C and
285 19°C did not change significantly over time. It is yet to be tested if significant changes would be
286 observed at 25°C if the duration of exposure was extended, but our results provide the first
287 indication of methylation divergence with increasing temperature, and this effect may increase
288 with time. In contrast, we did not find any significant DNA methylation changes in response to
289 the salinity treatments used in this study.

290

291 *Didemnum vexillum* is a subtidal species and can tolerate severe, short-term declines in salinity,
292 but extended periods of low-salinity stress lead to mortality (Gröner et al. 2011; Rocha et al.
293 2017) and ascidians are rarely found in salinities lower than 25 PSU (Lambert 2005). Based on
294 the above, we selected five salinity treatments: 34 PSU (within the upper range at collection
295 site), 32 PSU (within the middle range at collection site), 30 PSU (within the lower range at the

296 collection site), 28 PSU (within the lower global range) and 26 PSU (colonies are rarely found
297 below this level globally). However, at the colony collection site for this study (the Nelson
298 marina), salinity frequently drops below 26 PSU (Atalah 2017). The Nelson marina is located
299 within 1.5 km of a river mouth, and salinity drops are likely associated with rain events (Fletcher
300 et al. 2013a). The lack of response of genomic DNA methylation to changes in salinity suggests
301 that *D. vexillum* in the Nelson marina may already be adapted to a lower salinity environment
302 than was assessed in this experiment. This result is supported by the positive growth of colonies
303 in all salinity treatments and a lack of negative health indicators, an indication that colonies were
304 not experiencing significant stress. However, non-significant global methylation changes do not
305 necessarily demonstrate that important DNA methylation changes are not occurring. Locus
306 specific methylation differences have been associated with environmental differences in
307 temperature and salinity in solitary ascidians (Pu & Zhan 2017), and MS-AFLP results can be
308 difficult to interpret due to changes in many genes at once.

309

310 One other study has experimentally investigated the effect of environmental stress on DNA
311 methylation in an invasive marine invertebrate, and found very rapid (<3 hours) global DNA
312 methylation differences in response to low salinity stress, but these differences had disappeared
313 within 48 hours (Huang et al. 2017). In this study, solitary ascidians (*Ciona savignyi*), were
314 exposed to a lower salinity than the present study (20 PSU), with no gradual decrease allowing
315 acclimation time, and all individuals died prior to the final sampling time point (120 hours)
316 (Huang et al. 2017). Studies of solitary ascidians have shown salinity exposure can induce a
317 strong behavioural and physiological response in the first 48 hours of severe osmotic stress, such
318 as siphon closure and the excretion of intracellular osmolytes (Toop & Wheatly 1993), which

319 might be associated with such rapid, global DNA methylation changes. This may be a good
320 strategy for surviving short-term non-optimal salinity events, such severe storms or transport
321 through low salinity waters (Rocha et al. 2017), but prolonged stress will likely lead to mortality.
322 Nonetheless, this response demonstrates the potential for genome-wide methylation changes in
323 response to salinity stress.

324

325 *Didemnum vexillum* is a cool water temperate species found in a wide range of temperatures,
326 from <0 - >24°C (Bullard et al. 2007). Optimal temperature for growth appears to be between
327 14°C and 20°C. Based on this, we selected three temperature treatments: 19°C (within the
328 temperature range of *D. vexillum* colonies at the collection site, and the optimal range for *D.*
329 *vexillum* globally), 25°C (just outside the temperature range at the collection site, and near the
330 upper limit globally), and 27°C (+ 3°C of the upper limit at the collection site, but within global
331 climate change predictions for the year 2100; IPCC 2014). Temperatures at the collection site in
332 the Nelson marina typically range from 9°C – 10°C (winter minima) and 22°C – 23°C (summer
333 maxima) (Fletcher et al. 2013a). The *D. vexillum* colonies used in this study would rarely
334 experience temperatures of 25°C or greater. The significant global DNA methylation changes
335 observed in colonies held at 27°C, after just three days exposure, is indicative of a dramatic
336 response to thermal stress. A conclusion that is further supported by the significant negative
337 growth of colonies held at these elevated temperatures compared to controls. Alterations to
338 energetic balance (e.g., decreased growth/reproduction, switching to anaerobic metabolism) and
339 protein expression profiles (e.g., upregulation of heat shock chaperones), are well known and
340 energetically costly processes undertaken by invertebrates in response to thermal stress
341 (Sokolova et al. 2012). Such responses can act to ‘buy survival time’ until conditions improve,

342 and determine species distribution limits. This strategy may allow colonial ascidians to invade
343 new areas. For example, the recent expansion of *D. vexillum* from temperate regions into the
344 warmer, subtropical waters of the Mediterranean Sea (8 – 28°C) provides evidence of the
345 remarkable capacity of *D. vexillum* to adapt to increasing temperatures (Ordóñez et al. 2015). In
346 temperate regions, maximum growth and reproduction in *D. vexillum* occurs during the warmer,
347 summer months, with regression of colony growth and size occurring during winter (Fletcher et
348 al. 2013a). In the Mediterranean, this cycle is reversed (Ordóñez et al. 2015). By growing and
349 reproducing in the winter months and regressing during the summer months, *D. vexillum* is able
350 to extend its introduced range towards warmer waters (Ordóñez et al. 2015).

351

352 The MS-AFLP technique does not provide any insight into the identification of genes which are
353 differentially methylated, so we are unable to demonstrate that any environmentally induced
354 changes to DNA methylation are associated with functional traits that could lead to adaptive
355 outcomes. However, correlative experiments have previously suggested a role for DNA
356 methylation in adaptation to thermal stress, with natural populations of fish having higher levels
357 of methylation in polar and sub-Antarctic species compared to temperate/tropical species
358 (Varriale & Bernardi 2006). Some natural populations of the solitary ascidian, *Ciona robusta*,
359 display significant DNA methylation differences in genes that can be correlated with
360 environmental differences in temperature and salinity (Pu & Zhan 2017). Furthermore,
361 experimental evidence for an adaptive response to temperature stress has been shown using an
362 Antarctic marine polychaete worm, *Spiophanes tcherniai*. In this species, large DNA methylation
363 shifts were observed after exposure to a 5.5°C temperature increase, and this shift was
364 accompanied by physiological adaptation, with respiration and metabolic rates returning to

365 control levels in less than four weeks (Marsh & Pasqualone 2014). Future studies utilising
366 techniques with base pair resolution (e.g., bisulphite sequencing) will provide detailed insights
367 into the location of methylation changes in specific genes associated with functional outcomes.
368 This type of technique would also benefit from the analysis of downstream biological pathways,
369 such as the analysis of gene expression and metabolomic profiling.

370

371 **Conclusions**

372 In this study, we demonstrate the responsiveness of DNA methylation following exposure to an
373 environmental gradient (temperature), which was correlated with phenotypic change (growth).
374 Furthermore, DNA methylation changes did not occur in colonies exposed to an environmental
375 gradient to which they may already be adapted (salinity). This is the first study to investigate
376 DNA methylation patterns in a colonial ascidian, specifically the highly invasive *D. vexillum*,
377 and adds to a growing body of evidence that DNA methylation plays a key role in the plasticity
378 of adaptive traits. Epigenetic changes may contribute not only to the success of invasive species,
379 but also to the adaptability of native species to changes within their environmental range. *D.*
380 *vexillum* is an excellent model organism for future research into epigenetic responses to
381 environmental stress. The responsiveness of DNA methylation to changes in the environment in
382 this species lends itself to future studies testing the stability and longevity of these changes, and
383 whether these changes can be associated with adaptive outcomes. Evidently, many key questions
384 remain unanswered, including whether differences in methylation persist over time? Does
385 tolerance for elevated temperature increase following exposure and, is tolerance associated with
386 a specific epigenetic modification? However, our study establishes a baseline understanding of
387 the role of DNA methylation in a globally invasive species. This unique study system provides a

388 powerful framework for ecological epigenetic studies that could enhance our understanding of
389 adaptation to rapid environmental change.

390

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395

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Table 1 (on next page)**Adapter and primer sequences.**

Adapter and primer sequences used for MS-AFLP protocol.

Adapters	Sequence 5'- 3'
<i>EcoRI</i> -adapter F	5'-CTC GTA GAC TGC GTA CC-3'
<i>EcoRI</i> -adapter R	5'-AAT TGG TAC GCA GTC TAC-3'
<i>HpaII</i> and <i>MspI</i> -Adapter F	5'-GAC GAT GAG TCT AGA A-3'
<i>HpaII</i> and <i>MspI</i> -Adapter R	5'-CGT TCT AGA CTC ATC-3'
Pre-selective primers	
<i>EcoRI</i> -A	5'-GAC TGC GTA CCA ATT CA-3'
<i>HpaII</i> and <i>MspI</i> -T	5'-GAT GAG TCT AGA ACG GT-3'
Selective primers	
<i>EcoRI</i> + AAG	5'-GAC TGC GTA CCA ATT CAA G-3'
<i>EcoRI</i> + ACT	5'-GAC TGC GTA CCA ATT CAC T-3'
<i>HpaII MspI</i> + TAC	5'-6-FAM-GAT GAG TCT AGA ACG GTA C-3'
<i>HpaII MspI</i> + TCC	5'-6-FAM-GAT GAG TCT AGA ACG GTC C-3'

Table 2 (on next page)**CCGG sites where methylation sensitive restriction enzymes (HpaII and MspI) cleave (Yes) or do not cleave (No) to generate methylation dependent fragment patterns.**

Both MspI and HpaII recognise CCGG sites and cleave unmethylated CCGG sites (1/1), but MspI cannot cleave when the outer cytosine is fully or hemimethylated (m), and HpaII cannot cleave when the inner or outer cytosine is methylated on both strands. Cleaving by both enzymes is blocked when both cytosines are methylated. From this the methylation state of restriction sites can be scored (e.g., methylated (1/0 or 0/1) unmethylated (1/1) and uninformative (0/0)).

	Restriction sites		HpaII	MspI	Fragment classification
Type I	5'-CCGG GGCC-5'		Yes	Yes	Unmethylated 1/1
Type II	5'- ^m CCGG GGCC-5'	5'- ^m C ^m CGG GGCC-5'	Yes	No	Hemimethylated 1/0
Type III	5'-C ^m CGG GG ^m CC-5'		No	Yes	Internal cytosine methylation 0/1
Type IV	5'- ^m C ^m CGG GG ^m C ^m C-5'	5'- ^m CCGG GGC ^m C-5'	No	No	Uninformative 0/0

Figure 1

Colony growth rates with increasing temperature.

Colony growth rates (mean growth per day, mm \pm 1 s.e.) at 19, 25 and 27°C. Significant differences between treatments are denoted by different letters ($p < 0.05$, Tukey's test HSD).

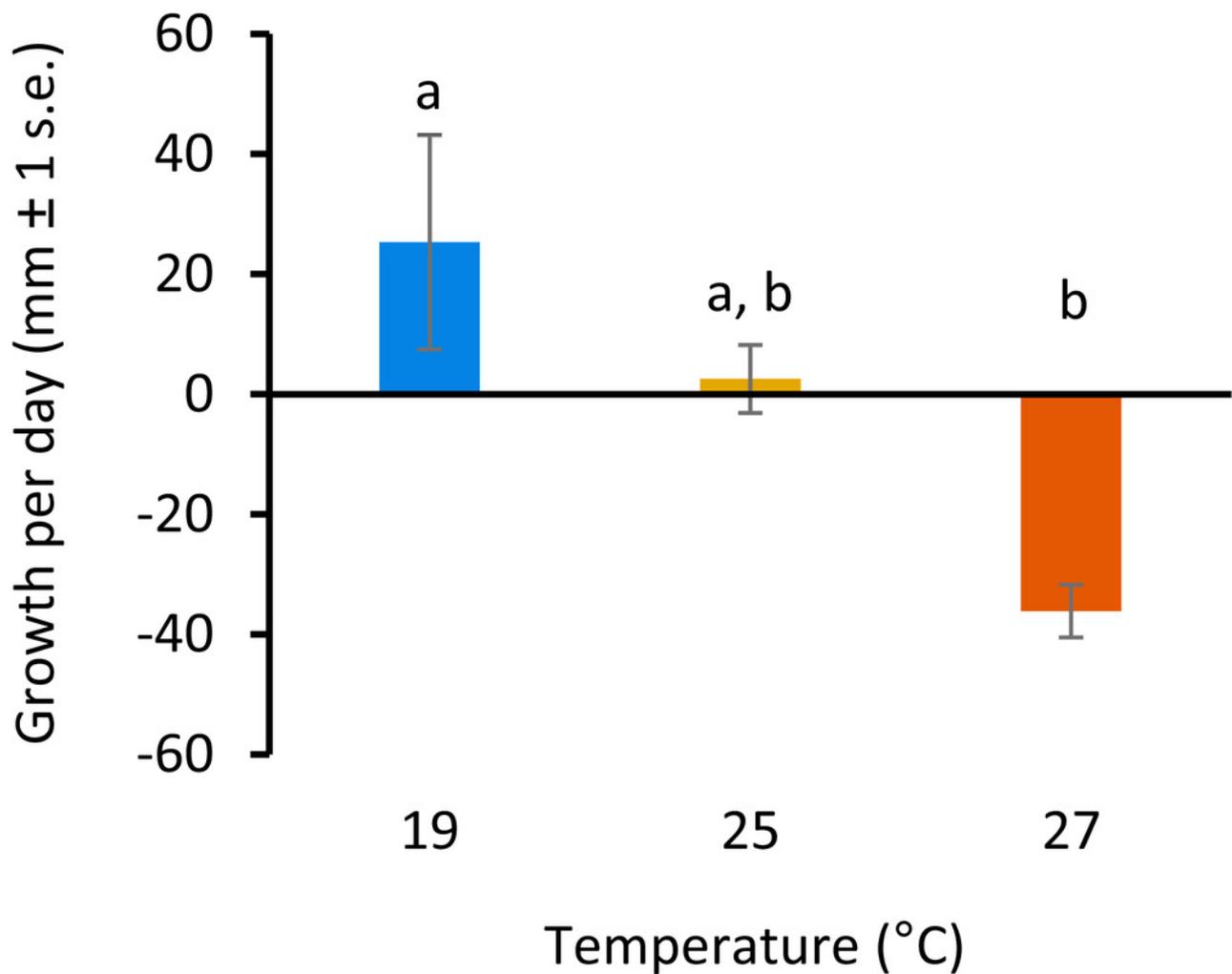


Figure 2

Principal Coordinate Analysis (PCoA) of methylation (MSL) differences between colonies exposed 19, 25 and 27°C.

(A) between colonies at Time 0 (T0), prior to elevated temperature exposure (baseline methylation); **(B)** between colonies at Time 1 (T1) following a gradual temperature increase of 1°C per day until all treatment temperatures were reached: 19°C (control) (n = 3), 25°C (n = 3) and 27°C (n = 3); **(C)** at Time 2 (T2)* after 3 days of elevated temperature exposure; **(D)** between sampling time points (T0, T1, T2) in colonies held at 19°C; **(E)** 25°C; (F) 27°C*. The first two coordinates (C1 and C2) are shown with the percentage of variance explained by them. Points in each group cloud represent individuals from different groups. Temperature labels show the centroid for the points cloud in each group. Ellipses represent average dispersion of those points around their centre (Pérez-Figueroa 2013). AMOVA tests for significant differences in methylation (MSL) are shown in supplementary material, Tables S3 and S4. * represents significant difference ($p < 0.05$).

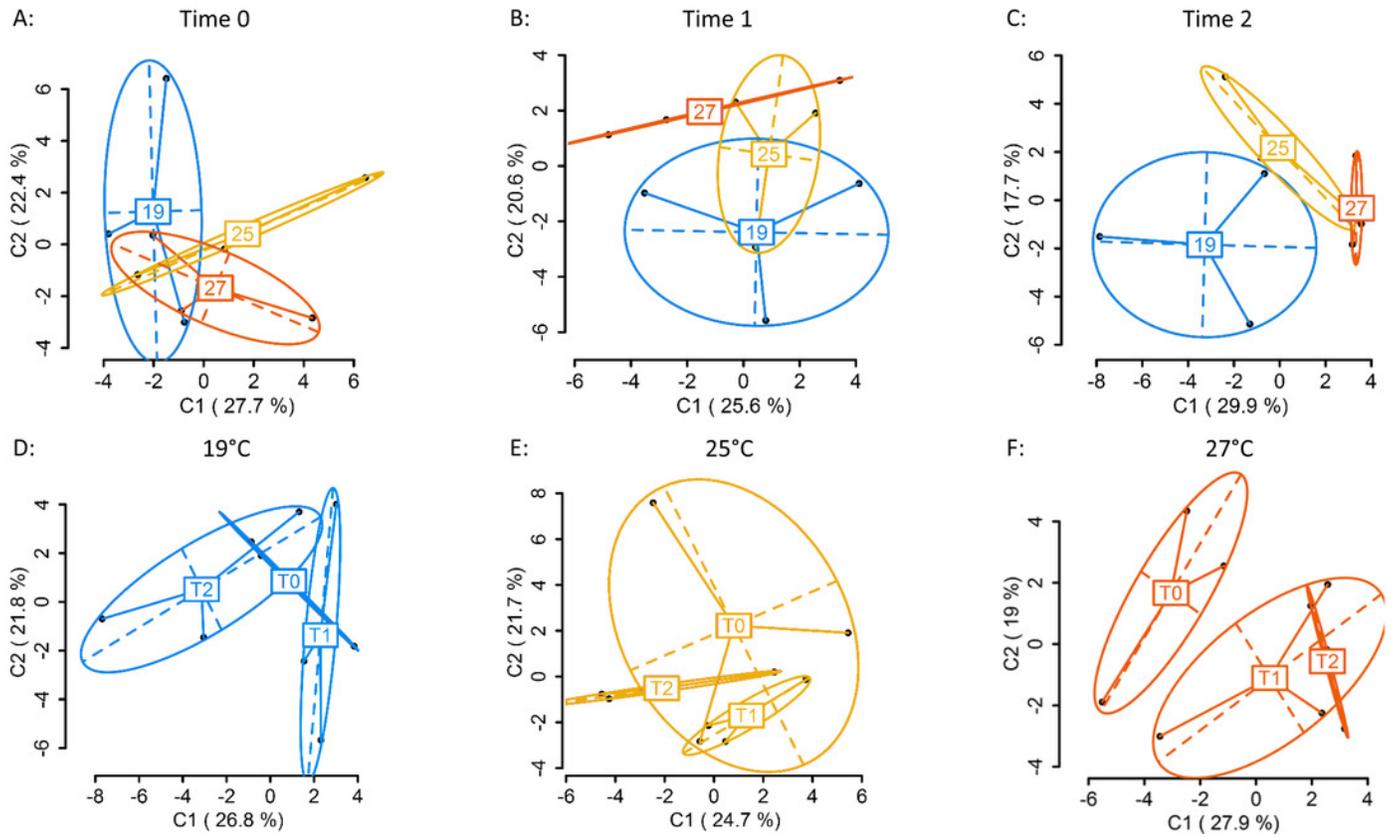


Figure 3

Colony growth rates with decreasing salinity.

Colony growth rates (mean growth per day, mm \pm 1 s.e.) at 26, 28, 30, 32, and 34 PSU.

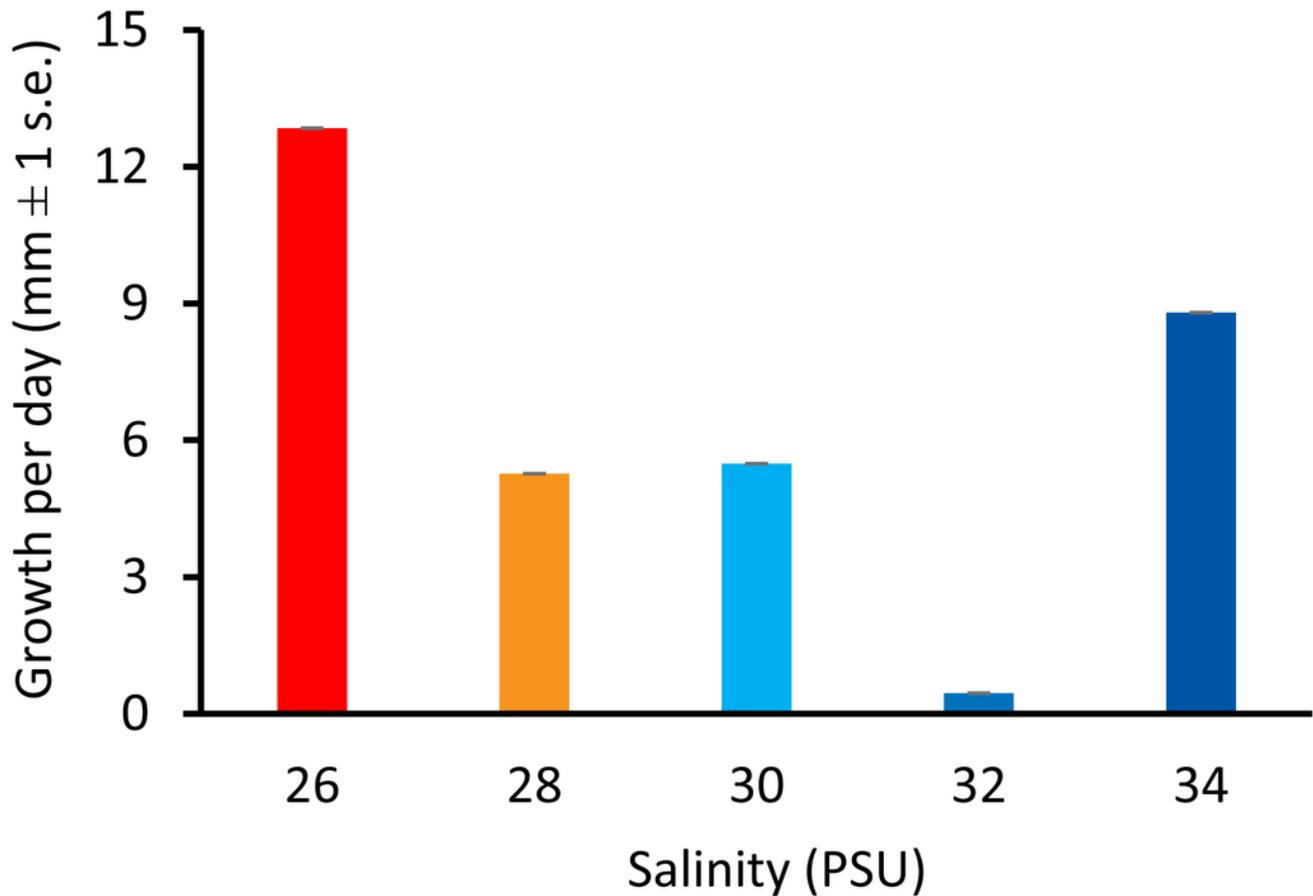


Figure 4

Principal Coordinate Analysis (PCoA) of methylation (MSL) differences between colonies exposed to 34, 32, 30, 28 and 26 PSU.

(A) between colonies at Time 0 (T0), prior to salinity treatment exposure (baseline methylation); **(B)** between colonies at Time 1 (T1) following a gradual salinity decrease of 1 PSU per day until all salinity treatments were reached: 34 PSU (control) (n = 3), 32 PSU (n = 3), 30 PSU (n = 3), 28 PSU (n = 3) and 26 PSU (n = 3); **(C)** at Time 2 (T2) after 3 days of decreased salinity exposure; **(D)** between sampling time points (T0, T1, T2) in colonies held at 26 PSU; **(E)** 28 PSU; **(F)** 30 PSU; **(G)** 32 PSU; **(H)** 34 PSU. The first two coordinates (C1 and C2) are shown with the percentage of variance explained by them. Points in each group cloud represent individuals from different groups. Labels show the centroid for the points cloud in each group. Ellipses represent average dispersion of those points around their centre (Pérez-Figueroa 2013). AMOVA tests for significant differences in methylation (MSL) are shown in supplementary material, Tables S6 and S7. No significant differences between groups were found.

