

Effects of temperature and salinity stress on DNA methylation in a highly invasive marine invertebrate, the colonial ascidian *Didemnum vexillum*

Nicola A Hawes^{Corresp., 1, 2}, Louis A Tremblay^{2, 3}, Xavier Pochon^{1, 2}, Brendon Dunphy^{1, 3}, Andrew E Fidler¹, Kirsty F Smith²

¹ Institute of Marine Science, University of Auckland, Auckland, New Zealand

² Cawthron Institute, Nelson, New Zealand

³ School of Biological Sciences, University of Auckland, Auckland, New Zealand

Corresponding Author: Nicola A Hawes

Email address: Nicola.hawes@cawthron.org.nz

Environmentally induced epigenetic changes may contribute to phenotypic plasticity, increase adaptive potential in changing environments, and play a key role in the establishment and spread of invasive species in new habitats. In this study, we used Methylation Sensitive Amplified Length Polymorphism (MS-AFLP) to assess environmentally induced DNA methylation changes in a globally invasive clonal ascidian, *Didemnum vexillum*. We tested the effect of increasing temperature (19, 25 and 27°C) and decreasing salinity (34, 32, 30, 28 and 26 practical salinity units (PSU)) on global DNA methylation, growth and survival rates. Exposure to 27°C resulted in significant changes in DNA methylation over time, while there were no significant changes in non-methylated loci (representing genetic variation). Growth also decreased in colonies exposed to high temperatures, suggesting they were under thermal stress. In contrast, no differences in growth or DNA methylation patterns were observed in colonies exposed to a decreasing salinity gradient, potentially due to prior adaptation to conditions experienced at the site of collection. The results of this study show that environmental stressors can induce significant global DNA methylation changes in an invasive marine invertebrate on very rapid timescales, and that this response varies depending on the type, magnitude, and duration of the stressor. Changes in genomic DNA methylation and the rate of growth may act to 'buy survival time' until conditions improve, and determine the distribution limits of this globally invasive species.

Effects of temperature and salinity stress on DNA methylation in a highly invasive marine invertebrate, the colonial ascidian *Didemnum vexillum*.

Nicola A. Hawes^{1,2}, Louis A. Tremblay^{1,3}, Xavier Pochon^{1,2}, Brendon Dunphy^{2,3}, Andrew E. Fidler², Kirsty F. Smith¹

¹*Cawthron Institute, 98 Halifax Street East, Nelson 7010, New Zealand*

²*Institute of Marine Science, University of Auckland, Private Bag 92019, Auckland 1142, New Zealand*

³*School of Biological Sciences, University of Auckland, Private Bag 92019, Auckland 1142, New Zealand*

Corresponding Author:

Nicola Hawes, Email address: nicola.hawes@cawthron.org.nz

Abstract.

Environmentally induced epigenetic changes may contribute to phenotypic plasticity, increase adaptive potential in changing environments, and play a key role in the establishment and spread of invasive species in new habitats. In this study, we used Methylation-Sensitive Amplified Fragment Length Polymorphism (MS-AFLP) to assess environmentally induced DNA methylation changes in a globally invasive colonial ascidian, *Didemnum vexillum*. We tested the effect of increasing temperature (19, 25 and 27°C) and decreasing salinity (34, 32, 30, 28 and 26 practical salinity units (PSU)) on global DNA methylation, growth and survival rates. After three days of exposure to elevated temperature, significant DNA methylation differences were observed between treatments. Exposure to 27°C resulted in changes in DNA methylation over time, while there were no significant changes in non-methylated loci (representing genetic variation). Growth also decreased in colonies exposed to high temperatures, suggesting they were under thermal stress. In contrast, no differences in growth or DNA methylation patterns were observed in colonies exposed to a decreasing salinity gradient, potentially due to prior adaptation to conditions experienced at the site of collection. The results of this study show that environmental stressors can induce significant global DNA methylation changes in an invasive marine invertebrate on very rapid timescales, and that this response varies depending on the type, magnitude, and duration of the stressor. Changes in genomic DNA methylation and the rate of growth may act to ‘buy survival time’ until conditions improve, and determine the distribution limits of this globally invasive species.

Keywords: Marine Invertebrate, Rapid Adaptation, Invasive species, Epigenetics, Environmental Change, Environmental Stress

Introduction.

Species invasions, climate change, habitat fragmentation and environmental degradation are altering ecosystems and threatening biodiversity (Leadley 2010). A key question in evolutionary biology is whether species will be able to adapt in response to these human-driven environmental changes (Visser 2008). Biological invasions can provide a unique model to investigate adaptation and evolution within short timescales, as introduced species must rapidly adapt to new habitats (Allendorf & Lundquist 2003; Sakai et al. 2001). It has been suggested that epigenetic mechanisms could play a critical role in environmental adaptation, and may be particularly important for the success of invasive species (Estoup et al. 2016; Pérez et al. 2006; Prentis et al. 2008). Recently introduced populations frequently have reduced genetic diversity (e.g., genetic bottlenecks and founder effects) (Dlugosch & Parker 2008), which is thought to constrain the colonisation potential of a species (e.g., Crawford & Whitney 2010). Despite this, invasive species can still be highly successful in their new environments, and often outcompete locally adapted native species (Allendorf & Lundquist 2003). By increasing both phenotypic plasticity and heritable variation, epigenetic changes might allow invasive species to quickly respond to environmental challenges. However, the role that epigenetic mechanisms play during the process of invasion is only beginning to be understood (Hawes et al. 2018, in press) and, for many species, the effect of environmental stressors on DNA methylation is unknown.

Epigenetic modifications have been shown to respond to environmental cues and, in some cases, be associated with significant phenotypic change (Dias & Ressler 2014; Kucharski et al. 2008; Waterland & Jirtle 2003). Epigenetic mechanisms are diverse and interactive (e.g., DNA methylation, histone modifications, small RNAs), but all alter gene expression without the

requirement for changes in the underlying DNA nucleotide sequences (Bossdorf et al. 2008). Currently the most studied epigenetic mechanism is the methylation of cytosine nucleotides to form 5 methyl-cytosine (DNA methylation). DNA methylation is common in eukaryotes, and there are a range of methods for its detection and quantification (Plongthongkum et al. 2014). One such method, Methylation-Sensitive Amplified Fragment Length Polymorphism (MS-AFLP), allows for cost-effective screening of variation in global DNA methylation, without the requirement for a reference genome (Reyna-Lopez et al. 1997). The MS-AFLP technique enables epigenetic research in non-model organisms and can provide a first look at DNA methylation-environment interactions, which may underlie adaptive plasticity. Interest in the ecological relevance of DNA methylation in non-model organisms is growing, and marine invertebrates have been identified as an emerging taxonomic group of interest for studies of ecological epigenetics, particularly in the context of environmental change (Hofmann 2017). Despite this, few studies have investigated environmentally induced epigenetic changes in marine invertebrates (Marsh et al. 2016; Marsh & Pasqualone 2014; Putnam et al. 2016), and the role of epigenetic mechanisms in the success of marine invertebrate invaders is only beginning to be explored (Ardura et al. 2017; Huang et al. 2017; Pu & Zhan 2017).

Colonial ascidians (phylum Chordata) are common marine invertebrate invaders worldwide, particularly in habitats perturbed by human activities, e.g., marinas, ports, aquaculture structures (Lambert 2001). Ascidians can thrive in a variety of environmental conditions, and display unique biological characteristics, including a broad tolerance to common environmental stressors such as temperature and salinity (Rocha et al. 2017). Additionally, the germ-cell lineages of colonial ascidians originate from somatic-cell lineages, contrasting with the germ-cell lineage

sequestration found in vertebrate Chordates (Rosner et al. 2009). Having no true germ-cell lineage sequestration increases the likelihood that stress-induced epigenetic modifications induced in somatic cells can be passed on to gametes (Verhoeven & Preite 2014). Finally, asexual reproduction (by budding) leads to colonies of genetically identical individual animals (termed zooids) that all share the same DNA nucleotide sequences (genotype). Clonal reproduction allows for genetically identical replicates across environment stress treatments, and repeated sampling of the same individual at multiple time-points. This reduces the confounding effects of genetic variation that frequently complicate epigenetic studies of non-clonal organisms (Douhovnikoff & Dodd 2015; Verhoeven & Preite 2014).

Invasive populations of the colonial ascidian, *Didemnum vexillum* Kott, 2002 have extremely low levels of genetic diversity compared to populations within its native range (Stefaniak et al. 2012). Despite this, *D. vexillum* is extremely successful where it has invaded (Beveridge et al. 2011; Cohen et al. 2011; Griffith et al. 2009; Hitchin 2012; Lambert et al. 2009; Tagliapietra et al. 2012), often forming large colonies that smother other marine invertebrates, including commercial aquaculture species (Fletcher et al. 2013b). We used MS-AFLP to determine whether a) DNA methylation is present in the genome of *D. vexillum*, and b) if genome-wide DNA methylation patterns in *D. vexillum* change in response to two prominent types of environmental stress: temperature and salinity. Temperature and salinity are often reported as the most important environmental determinants controlling the distribution of marine species, and these two parameters have been used repeatedly when studying the tolerance of ascidians to environmental stress (Dybern 1967; Gröner et al. 2011; Renborg et al. 2014; Serafini et al. 2011; Zerebecki & Sorte 2011). Furthermore, extreme climatic events such as precipitation events and

heatwaves are expected to increase in frequency in the near future (IPCC 2014), making it increasingly important to evaluate the response and resilience of marine invertebrates to thermal and osmotic stress.

Materials and Methods.

Sample collection and establishment of experimental colonies

Colonies of *D. vexillum* were collected from the Nelson Marina (South Island, New Zealand; 41°15'38"S, 173°16'54"E) in April 2016. At the time of collection, the water temperature was 19°C and the salinity was 33 PSU. Colonies were gently removed from wharf pilings and immediately placed in labelled 2 L plastic containers filled with ambient seawater for transport to the Cawthron Institute (less than 5 minutes commute). Attempts were made to remove whole colonies that were free from debris, but as the colonies were mostly small and growing flat over piles covered in other fouling, colonies often broke apart during removal. We treated these fragments as one colony for the following experiments but it is possible that it was two or more colonies growing in close proximity. After arrival at the laboratory, the colonies were gently cleaned with seawater to remove mud, silt and other organisms, and approximately equal sized fragments (c. 1.5 cm x 1.5 cm) were cut with a razor blade. Colony fragments were then placed on glass slides and gently wrapped with cotton thread to encourage attachment (Rinkevich & Fidler 2014). Glass slides were inserted into slide holders and placed in pre-conditioned 40 L glass aquaria in ambient, control conditions (19°C, 34 PSU) for one week to allow attachment to occur. During this acclimation period, colonies were fed 1.6×10^8 cells.L⁻¹ of cultured algae (*Isochrysis galbana*) every second day. Following attachment, the cotton was removed and the colony fragments were randomly allocated (Temperature: n = 9, Salinity: n = 15) to pre-

conditioned treatment tanks (n = three tanks per treatment, one colony fragment per treatment tank). All treatment tanks were maintained at 19°C and 34 PSU for a further two weeks acclimation time prior to beginning the experiments.

Experimental system

Seawater for the experiments was collected from Tasman Bay, Nelson (41°11'29.2"S 173°21'01.9"E), passed through three filters (pore size 50, 5, and 0.35 µm) and ultraviolet light treated. Each tank was filled with freshly collected seawater at the start of the experiment, and five L water exchanges were done daily throughout the duration of the experiment with pre-heated or reduced salinity seawater. Each day following water exchange, colonies were fed a diet of 1.6×10^8 cells.L⁻¹ of *I. galbana*. To prevent stratification, mixing in experimental tanks was ensured by gentle aeration using air stones. Experimental tanks were exposed to a 14:10 hour light:dark cycle to mimic summer conditions. Water temperatures were maintained using thermostatically regulated aquarium heaters (EHEIM JAIGER 100W, Deizisau, Germany) and salinity treatments were achieved and maintained by the addition of reverse osmosis (RO) water to seawater. Water temperature and salinity were measured twice daily using hand-held probes to ensure stable treatment conditions were maintained $\pm 0.5^\circ\text{C}$ or 0.5 PSU (YSI Professional Plus, YSI Incorporated, Yellow Springs Ohio, USA). After two weeks of acclimation, tissue samples were collected from all colonies for MS-AFLP analyses (Time 0; T0). Temperatures were then increased by 1°C per day for temperature treatments, and salinity was reduced by 1 PSU per day for the salinity treatments until, after eight days, all treatment conditions were reached. Treatments were as follows: temperature = 19 (Control), 25, and 27°C and salinity= 34 (Control), 32, 30, 28, and 26 PSU. Temperature and salinity experiments were run in parallel.

Tissue samples were taken at day eight for MS-AFLP analyses (Time 1; T1). Colony fragments were then maintained in experimental treatments for a further three days (day 11), at which time final tissue samples were taken for MS-AFLP analyses (Time 2; T2).

Sampling protocol

Prior to tissue collection, colonies were not fed for 16 hours to minimise contamination by feed microalgae. To collect tissue for MS-AFLP analyses, small (c. 5 mm x 5 mm) samples were taken from each colony using a sterile razorblade. Tissue samples were preserved in 95% ethanol, which was refreshed once before storage at -20°C until processing. Pre- and post-tissue sampling, photos were taken for growth rate calculations. Colony growth rates (quantified by changes in colony surface-area over time) were measured using Image J 1.48v software (Schneider et al. 2012). The survival of colonies was also monitored at the time of sampling by assessments of colony health, including zooid integrity, colour and texture of the colony, build-up of detritus and dead tissue. Throughout the experiment, all colonies were transported, sampled and photographed while submerged in trays of temperature and salinity adjusted seawater to minimise handling stress.

MS-AFLP analysis

To assess whole genome DNA methylation patterns, DNA was extracted using G-spin Total DNA extraction kits (animal tissue protocol; Intron, Gyeonggi-do, South Korea). Following DNA extraction, in parallel reactions (25 µL final reaction volume), 500 ng of DNA was digested with 10 U of each restriction enzyme (MspI and EcoRI or HpaII and EcoRI; New England BioLabs, Ipswich, MA, USA) and 10X CutSmart Buffer (England BioLabs), and

incubated at 37°C for 2 hours followed by 80°C for 20 min to inactivate the enzymes. The digested DNA were ligated in a final volume of 20 µL containing 1 U of T4 DNA ligase (New England BioLabs), 10X ligase buffer, 250 nM of EcoRI adaptor, and 2.5 µM of MspI or HpaII adaptor for 3 hours at 37°C. Table 1 presents a list of all adaptor and primer sequences used for the MS-AFLP protocol. Pre-selective polymerase chain reaction (PCR) was performed in a total volume of 20 µL using 8 µL of ligated DNA, MyTaq™ 2X PCR master mix (Bioline, MA, USA), and 500 nM of each pre-selective primer. Thermocycling conditions were 20 cycles of: 94°C for 30 s, 56°C for 60 s, and 72°C for 60 s. Selective PCR was performed using four EcoRI and MspI/HpaII primer combinations in a final volume of 20 µL using 1 µL of pre-selective PCR product, MyTaq™ 2X PCR master mix (Bioline, MA, USA), and 500 nM of each combination of forward and reverse selective primers. Thermocycling conditions were 1 cycle of: 94°C for 2 min; 10 cycles of 94°C for 30 s, 65°C for 30 s (decreasing 1°C per cycle), and 72°C for 60 s; and 30 cycles of 94°C for 30 s, 56°C for 30 s, and 72°C for 60 s; and a hold cycle of 72°C for 30 min. The resulting selective PCR product was diluted 1:5 with sterile distilled water and analysed using an ABI 3130 capillary sequencer (Applied Biosystems, Foster City, CA, USA) with internal size standards (GS600LIZ) by an external contractor (Genetic Analysis Services, University of Otago, Dunedin, New Zealand).

Data analysis

PEAKSCANNER software (Applied Biosystems, Foster City, CA, USA) was used to assign the MS-AFLP fragments peak height and size. To determine the parameters for subsequent analysis, the MS-AFLP procedure described above was first repeated three times for one individual sample (data not shown). The following settings were associated with the lowest error rate

between replicates and were applied to the *msap* analysis (Pérez-Figueroa 2013) described below. Error rate per primer, 0.07; analysis range, 50 – 500 base pairs (bp); minimum peak height, 1200 relative fluorescence units. Peak presence/absence data corresponding to HpaII and MspI fragments was then converted to a binary matrix (presence = 1, absence = 0), so that the methylation state of each restriction site could be identified. MS-AFLP profiles were assessed using the R package *msap* v. 1.1.8 (Pérez-Figueroa 2013). The *msap* package determines whether individual fragments (loci) are methylated (MSL) or not methylated (NML) by analysing the contents of the binary matrix, and comparing differences representing the differential sensitivities of HpaII and MspI to cytosine methylation (Table 2). For NML, differences between samples can represent genetic variation. From this, DNA methylation profiles of control and experimental samples were assessed by means of principal coordinate analyses (PCoA) followed by analyses of molecular variance (AMOVA) (Excoffier et al. 1992). Colony growth rates (mean growth per day, mm \pm 1 s.e.) were assessed using standard ANOVA and post-hoc pairwise comparisons were made using Tukey's honest significance difference (HSD) test.

Results.

Temperature stress experiments

Growth rates

All colonies survived elevated temperature exposure, and colony growth rates (Figure 1) were significantly different between treatments (single level ANOVA, $F(2, 6) = 7.82$, $p = 0.0213$; Table S1). Post hoc comparisons using Tukey HSD indicated that growth per day was significantly reduced when colonies were exposed to 27°C compared to colonies grown at 19°C

($Diff = -61.43$, $p = 0.01$; Table S2, Figure 1). Growth was not significantly different between colonies exposed to 25°C and 27°C or 19°C and at 25°C.

Whole genome DNA methylation patterns (MS-AFLP)

Using four primer combinations, 1157 fragments (loci) were produced and analysed. At sampling time zero (T0), prior to temperature treatment exposure, 586 of the 1157 loci were MSL, of which 178 were polymorphic (30%). There were no significant differences in DNA methylation (Figure 2A) (MSL, $\phi_{ST} = -0.04981$, $p = 0.7093$, single level AMOVA; Table S3), or differences in non-methylated loci (NML, $\phi_{ST} = -0.03527$, $p = 0.9885$; Table S4), between treatment groups. Following eight days of gradual temperature increase, (T1), of the 1157 loci analysed 613 were MSL, of which 172 were polymorphic (28%). There were still no significant differences between DNA methylated (MSL) (Figure 2B) ($\phi_{ST} = 0.05606$, $p = 0.2176$, single level AMOVA; Table S3) or non-methylated loci (NML) ($\phi_{ST} = -0.013546$, $p = 0.7362$, single level AMOVA; Table S4), between the three temperature treatments. However, after three days of exposure to elevated temperature, of 1157 loci 600 were MSL, of which 201 were polymorphic (34%) and statistically significant differences in DNA methylation were evident between treatment groups (Figure 2C) (MSL, $\phi_{ST} = 0.1585$, $p = 0.0215$, single level AMOVA; Table S3). There were no significant global methylation changes between sampling time points in colonies held at 19°C (Figure 2D) (MSL, $\phi_{ST} = -0.0212$, $p = 0.6019$, single level AMOVA; Table S5) or 25°C (Figure 2E) (MSL, $\phi_{ST} = -0.01373$, $p = 0.5637$, single level AMOVA; Table S5). In contrast, there were significant DNA methylation changes following exposure to 27°C (Figure 2F) (MSL, $\phi_{ST} = 0.1727$, $p = 0.0223$, single-level AMOVA; Table S5), while non-methylated loci were not-significantly different over time (NML, $\phi_{ST} = -0.01769$, $p = 0.9425$,

single-level AMOVA, Table S6). Additionally, a mantel test showed there was no correlation between epigenetic (MSL) and genetic (NML) variation ($r = 0.1498322$, $p = 0.7303$; $nperm = 1000$). This suggests that that observed changes in epigenetic variation cannot be explained by the genetic background. Variation between individuals was also reduced in the 27°C treatment group, with DNA methylation patterns becoming more similar, as visualised by the reduced spread of samples around the centroid in the PCoA for each temperature (Figure 2C).

Salinity stress experiments

Growth rates

Akin to the temperature treatment, all colonies survived decreased salinity exposure. Colony growth rates were not significantly different between treatment groups (Figure 3) (single level ANOVA, $F(4, 10) = 2.066$, $p = 0.161$; Table S7).

Whole genome DNA methylation patterns (MS-AFLP)

Using four primer combinations, 1050 loci were produced and analysed. At sampling time zero (T0), prior to differential salinity exposure, of these, 626 were methylation sensitive loci (MSL), of which 169 were polymorphic (27%). There were no significant differences in DNA methylation between treatment groups (Figure 4A) (MSL, $\Phi_{ST} = -0.05023$, $p = 0.686$, single level AMOVA; Table S8), but non-methylated loci were significantly different (NML, $\Phi_{ST} = 0.01602$, $p = 0.0224$; Table S9). Following the gradual salinity decrease, at time one (T1), of 1050 loci 654 were MSL, of which 173 were polymorphic (26%). There were no significant differences in DNA methylation (Figure 4B) (MSL, $\Phi_{ST} = 0.02674$, $p = 0.3518$; single level AMOVA; Table S8), and in non-methylated loci (NML, $\Phi_{ST} = -0.0065$, $p = 0.575$, single level AMOVA; Table

S9), between salinity treatments. After three days of exposure to elevated salinity, of 1050 loci, 682 were MSL, of which 95 were polymorphic (14%). Non-significant differences in DNA methylation and non-methylated loci remained (Figure 4C) (MSL, $\phi_{ST} = -0.06389$, $p = 0.8328$, single level AMOVA; NML, $\phi_{ST} = 0.00471$, $p = 0.25$, single level AMOVA). There were no significant methylation differences (Table S10) or NML differences (Table 11) in any of the salinity treatments over time (Figure 4D, E, F, G H).

Discussion.

Changes in DNA methylation may be one of the mechanisms by which invasive species can rapidly adapt to new environments. However, for many species, the responsiveness of DNA methylation to environmental challenges has not yet been tested. Our results indicate that environmental stressors can induce significant global DNA methylation changes in an invasive marine invertebrate on very rapid timescales, and that this response varies depending on the type, magnitude, and duration of the stressor. After three days of exposure to elevated temperature, significant changes in whole-genome patterns of DNA methylation had occurred in *D. vexillum* colonies held at 27°C. This difference in DNA methylation could not be explained by the genetic background of the experimental colonies. In contrast, DNA methylation patterns in colonies exposed to 25°C and 19°C did not change significantly over time. It is yet to be tested if significant changes would be observed at 25°C if the duration of exposure was extended, but our results provide the first indication of methylation divergence with increasing temperature, and this effect may increase with time. In contrast, we did not find any significant DNA methylation changes in response to the salinity treatments used in this study.

Didemnum vexillum is a subtidal species and can tolerate severe, short-term declines in salinity, but extended periods of low-salinity stress lead to mortality (Gröner et al. 2011; Rocha et al. 2017) and ascidians are rarely found in salinities lower than 25 PSU (Lambert 2005). Based on the above, we selected five salinity treatments: 34 PSU (within the upper range at collection site), 32 PSU (within the middle range at collection site), 30 PSU (within the lower range at the collection site), 28 PSU (within the lower global range) and 26 PSU (colonies are rarely found below this level globally). However, at the colony collection site for this study (the Nelson marina), salinity frequently drops below 26 PSU (Atalah 2017). The Nelson marina is located within 1.5 km of a river mouth, and salinity drops are likely associated with rain events (Fletcher et al. 2013a). The lack of response of genomic DNA methylation to changes in salinity suggests that *D. vexillum* in the Nelson marina may already be adapted to a lower salinity environment than was assessed in this experiment. This result is supported by the positive growth of colonies in all salinity treatments and a lack of negative health indicators, an indication that colonies were not experiencing significant stress. However, non-significant global methylation changes do not necessarily demonstrate that important DNA methylation changes are not occurring, as very small methylation differences have been associated with environmental differences in salinity and temperature (Pu & Zhan 2017), and MS-AFLP results can be difficult to interpret due to changes in many genes at once. One other study has experimentally investigated the effect of environmental stress on DNA methylation in an invasive marine invertebrate, and found very rapid (<3 hours) global DNA methylation differences in response to low salinity stress, but these differences had disappeared within 48 hours (Huang et al. 2017). In this study, solitary ascidians (*Ciona savignyi*), were exposed to a lower salinity than the present study (20 PSU), with no gradual decrease allowing acclimation time, and all individuals died prior to the final sampling

time point (120 hours) (Huang et al. 2017). Studies of solitary ascidians have shown salinity exposure can induce a strong behavioural and physiological response in the first 48 hours of severe osmotic stress, such as siphon closure and the excretion of intracellular osmolytes (Toop & Wheatly 1993). This may be a good strategy for surviving short-term non-optimal salinity events, such severe storms or transport through low salinity waters (Rocha et al. 2017), but prolonged stress will likely lead to mortality.

Didemnum vexillum is a cool water temperate species found in a wide range of temperatures, from $<0 - >24^{\circ}\text{C}$ (Bullard et al. 2007). Optimal temperature for growth appears to be between 14°C and 20°C . Based on this, we selected three temperature treatments: 19°C (within the temperature range of *D. vexillum* colonies at the collection site, and the optimal range for *D. vexillum* globally), 25°C (just outside the temperature range at the collection site, and near the upper limit globally), and 27°C ($+ 3^{\circ}\text{C}$ of the upper limit at the collection site, but within global climate change predictions for the year 2100; IPCC 2014). Temperatures at the collection site in the Nelson marina typically range from $9^{\circ}\text{C} - 10^{\circ}\text{C}$ (winter minima) and $22^{\circ}\text{C} - 23^{\circ}\text{C}$ (summer maxima) (Fletcher et al. 2013a). The *D. vexillum* colonies used in this study would rarely experience temperatures of 25°C or greater. The significant global DNA methylation changes observed in colonies held at 27°C , after just three days exposure, is indicative of a dramatic response to thermal stress. A conclusion that is further supported by the significant negative growth of colonies held at these elevated temperatures compared to controls. Alterations to energetic balance (e.g., decreased growth/reproduction, switching to anaerobic metabolism) and protein expression profiles (e.g., upregulation of heat shock chaperones), are well known and energetically costly processes undertaken by invertebrates in response to thermal stress

(Sokolova et al. 2012). Such responses can act to ‘buy survival time’ until conditions improve, and determine species distribution limits. This strategy may allow colonial ascidians to invade new areas. For example, the recent expansion of *D. vexillum* from temperate regions into the warmer, subtropical waters of the Mediterranean Sea (8 – 28°C) provides evidence of the remarkable capacity of *D. vexillum* to adapt to increasing temperatures (Ordóñez et al. 2015). In temperate regions, maximum growth and reproduction in *D. vexillum* occurs during the warmer, summer months, with regression of colony growth and size occurring during winter (Fletcher et al. 2013a). In the Mediterranean, this cycle is reversed (Ordóñez et al. 2015). By growing and reproducing in the winter months and regressing during the summer months, *D. vexillum* is able to extend its introduced range towards warmer waters (Ordóñez et al. 2015).

The MS-AFLP technique does not provide any insight into the identification of genes which are differentially methylated, so we are unable to demonstrate that any environmentally induced changes to DNA methylation are associated with functional traits that could lead to adaptive outcomes. However, correlative experiments have previously suggested a role for DNA methylation in adaptation to thermal stress, with natural populations of fish having higher levels of methylation in polar and sub-Antarctic species compared to temperate/tropical species (Varriale & Bernardi 2006). Some natural populations of the solitary ascidian, *Ciona robusta*, display significant DNA methylation differences in genes that can be correlated with environmental differences in temperature and salinity (Pu & Zhan 2017). Furthermore, experimental evidence for an adaptive response to temperature stress has been shown using an Antarctic marine polychaete worm, *Spiophanes tcherniai*. In this species, large DNA methylation shifts were observed after exposure to a 5.5°C temperature increase, and this shift was

accompanied by physiological adaptation, with respiration and metabolic rates returning to control levels in less than four weeks (Marsh & Pasqualone 2014). Future studies utilising techniques with base pair resolution (e.g., bisulphite sequencing) will provide detailed insights into the location of methylation changes in specific genes associated with functional outcomes. This type of technique would also benefit from the analysis of downstream biological pathways, such as the analysis of gene expression and metabolomic profiling.

Conclusions.

In this study, we demonstrate the responsiveness of DNA methylation following exposure to an environmental gradient (temperature), which was correlated with phenotypic change (growth). Furthermore, DNA methylation changes did not occur in colonies exposed to an environmental gradient to which they may already be adapted (salinity). This is the first study to investigate DNA methylation patterns in a colonial ascidian, specifically the highly invasive *D. vexillum*, and adds to a growing body of evidence that DNA methylation plays a key role in the plasticity of adaptive traits. Epigenetic changes may contribute not only to the success of invasive species, but also to the adaptability of native species to changes within their environmental range. *D. vexillum* is an excellent model organism for future research into epigenetic responses to environmental stress. The responsiveness of DNA methylation to changes in the environment in this species lends itself to future studies testing the stability and longevity of these changes, and whether these changes can be associated with adaptive outcomes. Evidently, many key questions remain unanswered, including whether differences in methylation persist over time? Does tolerance for elevated temperature increase following exposure and, is tolerance associated with a specific epigenetic modification? However, our study establishes a baseline understanding of

the role of DNA methylation in a globally invasive species. This unique study system provides a powerful framework for ecological epigenetic studies that could enhance our understanding of adaptation to rapid environmental change.

Acknowledgements.

We are grateful to Eric Goodwin (Cawthron) for help with R coding. Thank you to Achira Amadoru (Nelson Marlborough Institute of Technology) for assistance with experiment maintenance.

References.

- Allendorf FW, and Lundquist LL. 2003. Introduction: population biology, evolution, and control of invasive species. *Conservation Biology* 17:24-30.
- Ardura A, Zaiko A, Morán P, Planes S, and Garcia-Vazquez E. 2017. Epigenetic signatures of invasive status in populations of marine invertebrates. *Scientific Reports* 7.
- Atalah, Javier (2017): Nelson Marina, New Zealand: SST and salinity data. figshare. <https://doi.org/10.6084/m9.figshare.5277025.v1>.
- Beveridge C, Cook E, Brunner L, MacLeod A, Black K, Brown C, and Manson F. 2011. Initial response to the invasive carpet sea squirt, *Didemnum vexillum*. *Scotland Scottish Natural Heritage Commissioned Report* 413:24.
- Bossdorf O, Richards CL, and Pigliucci M. 2008. Epigenetics for ecologists. *Ecology Letters* 11:106-115.
- Bullard SG, Lambert G, Carman MR, Byrnes J, Whitlatch RB, Ruiz G, Miller R, Harris L, Valentine PC, and Collie JS. 2007. The colonial ascidian *Didemnum* sp. A: current

distribution, basic biology and potential threat to marine communities of the northeast and west coasts of North America. *Journal of Experimental Marine Biology and Ecology* 342:99-108.

Cohen CS, McCann L, Davis T, Shaw L, and Ruiz G. 2011. Discovery and significance of the colonial tunicate *Didemnum vexillum* in Alaska. *Aquatic Invasions* 6:263-271.

Crawford K, and Whitney K. 2010. Population genetic diversity influences colonization success. *Molecular Ecology* 19:1253-1263.

Douhovnikoff V, and Dodd RS. 2015. Epigenetics: a potential mechanism for clonal plant success. *Plant Ecology* 216:227-233.

Dias BG, and Ressler KJ. 2014. Parental olfactory experience influences behavior and neural structure in subsequent generations. *Nature Neuroscience* 17:89-96.

Dlugosch KM, and Parker IM. 2008. Founding events in species invasions: genetic variation, adaptive evolution, and the role of multiple introductions. *Molecular Ecology* 17:431-449.

Dybern BI. 1967. The distribution and salinity tolerance of *Ciona intestinalis* (L.) f. typica with special reference to the waters around southern Scandinavia. *Ophelia* 4:207-226.

Estoup A, Ravigné V, Hufbauer R, Vitalis R, Gautier M, and Facon B. 2016. Is there a genetic paradox of biological invasion? *Annual Review of Ecology Evolution and Systematics* 47:51-72.

Excoffier L, Smouse PE, and Quattro JM. 1992. Analysis of molecular variance inferred from metric distances among DNA haplotypes: application to human mitochondrial DNA restriction data. *Genetics* 131:479-491.

- Fletcher LM, Forrest BM, Atalah J, and Bell JJ. 2013a. Reproductive seasonality of the invasive ascidian *Didemnum vexillum* in New Zealand and implications for shellfish aquaculture. *Aquaculture Environment Interactions* 3:197-211.
- Fletcher LM, Forrest BM, and Bell JJ. 2013b. Impacts of the invasive ascidian *Didemnum vexillum* on green-lipped mussel *Perna canaliculus* aquaculture in New Zealand. *Aquaculture Environment Interactions* 4:17-30.
- Griffith K, Mowat S, Holt RH, Ramsay K, Bishop JD, Lambert G, and Jenkins SR. 2009. First records in Great Britain of the invasive colonial ascidian *Didemnum vexillum* Kott, 2002. *Aquatic Invasions* 4:581-590.
- Gröner F, Lenz M, Wahl M, and Jenkins SR. 2011. Stress resistance in two colonial ascidians from the Irish Sea: The recent invader *Didemnum vexillum* is more tolerant to low salinity than the cosmopolitan *Diplosoma listerianum*. *Journal of Experimental Marine Biology and Ecology* 409:48-52.
- Hawes NA, Fidler AE, Tremblay LA, Pochon X, Dunphy B and Smith KF. 2018. Understanding the role of DNA methylation in successful biological invasions - A review. *Biological Invasions*. In Press.
- Hitchin B. 2012. New outbreak of *Didemnum vexillum* in North Kent: on stranger shores. *Porcupine Marine Natural History Society Newsletter* 31:43-48.
- Hofmann G. 2017. Ecological Epigenetics in Marine Metazoans. *Front. Mar Sci* 4.
- Huang X, Li S, Ni P, Gao Y, Jiang B, Zhou Z, and Zhan A. 2017. Rapid response to changing environments during biological invasions: DNA methylation perspectives. *Molecular Ecology* 26(23):6621-6633.

- IPCC, 2014: Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and
III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change
[Core Writing Team, R.K. Pachauri and L.A. Meyer (eds.)]. IPCC, Geneva, Switzerland,
151pp.
- Kucharski R, Maleszka J, Foret S, and Maleszka R. 2008. Nutritional control of reproductive
status in honeybees via DNA methylation. *Science* 319:1827-1830.
- Lambert G. 2001. A global overview of ascidian introductions and their possible impact on the
endemic fauna. *The biology of ascidians*: Springer, 249-257.
- Lambert G. 2005. Ecology and natural history of the protochordates. *Canadian Journal of
Zoology* 83:34-50.
- Lambert G, Locke A, and Carman M. 2009. Adventures of a sea squirt sleuth: unraveling the
identity of *Didemnum vexillum*, a global ascidian invader. *Aquatic Invasions* 1:5-28.
- Leadley P. 2010. *Biodiversity scenarios: projections of 21st century change in biodiversity, and
associated ecosystem services: a technical report for the global biodiversity outlook 3*:
UNEP/Earthprint.
- Marsh AG, Hoadley KD, and Warner ME. 2016. Distribution of CpG Motifs in Upstream Gene
Domains in a Reef Coral and Sea Anemone: Implications for Epigenetics in Cnidarians.
PLOS ONE 11:e0150840.
- Marsh AG, and Pasqualone AA. 2014. DNA methylation and temperature stress in an Antarctic
polychaete, *Spiophanes tcherniai*. *Frontiers in Physiology* 5:173-181.
- Ordóñez V, Pascual M, Fernández-Tejedor M, Pineda M, Tagliapietra D, and Turon X. 2015.
Ongoing expansion of the worldwide invader *Didemnum vexillum* (Ascidacea) in the

Mediterranean Sea: high plasticity of its biological cycle promotes establishment in warm waters. *Biological Invasions* 17:2075-2085.

Pérez-Figueroa A. 2013. msap: a tool for the statistical analysis of methylation-sensitive amplified polymorphism data. *Molecular ecology resources* 13:522-527.

Pérez JE, Nirchio M, Alfonsi C, and Muñoz C. 2006. The biology of invasions: the genetic adaptation paradox. *Biological Invasions* 8:1115-1121.

Plongthongkum N, Diep DH, and Zhang K. 2014. Advances in the profiling of DNA modifications: cytosine methylation and beyond. *Nature Reviews Genetics* 15:647.

Prentis PJ, Wilson JR, Dormontt EE, Richardson DM, and Lowe AJ. 2008. Adaptive evolution in invasive species. *Trends in Plant Science* 13:288-294.

Pu C, and Zhan A. 2017. Epigenetic divergence of key genes associated with water temperature and salinity in a highly invasive model ascidian. *Biological Invasions* 19:2015-2028.

Putnam HM, Davidson JM, and Gates RD. 2016. Ocean acidification influences host DNA methylation and phenotypic plasticity in environmentally susceptible corals. *Evolutionary Applications* 9:1165-1178.

Renborg E, Johannesson K, and Havenhand J. 2014. Variable salinity tolerance in ascidian larvae is primarily a plastic response to the parental environment. *Evolutionary Ecology* 28:561-572.

Reyna-Lopez G, Simpson J, and Ruiz-Herrera J. 1997. Differences in DNA methylation patterns are detectable during the dimorphic transition of fungi by amplification of restriction polymorphisms. *Molecular and General Genetics MGG* 253:703-710.

Rinkevich B, and Fidler AE. 2014. Initiating laboratory culturing of the invasive ascidian *Didemnum vexillum*. *Management of Biological Invasions* 5:55-62.

Rocha RM, Castellano GC, and Freire CA. 2017. Physiological tolerance as a tool to support invasion risk assessment of tropical ascidians. *Marine Ecology Progress Series* 577:105-119.

Rosner A, Moiseeva E, Rinkevich Y, Lapidot Z, and Rinkevich B. 2009. Vasa and the germ line lineage in a colonial urochordate. *Developmental Biology* 331:113-128.

Sakai AK, Allendorf FW, Holt JS, Lodge DM, Molofsky J, With KA, Baughman S, Cabin RJ, Cohen JE, and Ellstrand NC. 2001. The population biology of invasive species. *Annual Review of Ecology and Systematics* 32:305-332.

Schneider CA, Rasband WS, and Eliceiri KW. 2012. NIH Image to ImageJ: 25 years of image analysis. *Nature Methods* 9:671.

Serafini L, Hann JB, Kültz D, and Tomanek L. 2011. The proteomic response of sea squirts (genus *Ciona*) to acute heat stress: a global perspective on the thermal stability of proteins. *Comparative Biochemistry and Physiology Part D: Genomics and Proteomics* 6:322-334.

Sokolova IM, Frederick M, Bagwe R, Lannig G, and Sukhotin AA. 2012. Energy homeostasis as an integrative tool for assessing limits of environmental stress tolerance in aquatic invertebrates. *Marine Environmental Research* 79:1-15.

Stefaniak L, Zhang H, Gittenberger A, Smith K, Holsinger K, Lin S, and Whitlatch RB. 2012. Determining the native region of the putatively invasive ascidian *Didemnum vexillum* Kott, 2002. *Journal of Experimental Marine Biology and Ecology* 422:64-71.

Tagliapietra D, Keppel E, Sigovini M, and Lambert G. 2012. First record of the colonial ascidian *Didemnum vexillum* Kott, 2002 in the Mediterranean: Lagoon of Venice (Italy). *BioInvasions Records* 1:247-254.

- Toop T, and Wheatly MG. 1993. Some behavioural and physiological responses of the ascidian *Styela plicata* (Lesueur) during acclimation to low salinity. *Marine & Freshwater Behaviour & Phy* 24:33-44.
- Varriale A, and Bernardi G. 2006. DNA methylation and body temperature in fishes. *Gene* 385:111-121.
- Verhoeven KJF, and Preite V. 2014. Epigenetic variation in asexually reproducing organisms. *Evolution* 68:644-655.
- Visser ME. 2008. Keeping up with a warming world; assessing the rate of adaptation to climate change. *Proceedings of the Royal Society of London B: Biological Sciences* 275:649-659.
- Waterland RA, and Jirtle RL. 2003. Transposable elements: targets for early nutritional effects on epigenetic gene regulation. *Molecular and Cellular Biology* 23:5293-5300.
- Zerebecki RA, and Sorte CJB. 2011. Temperature tolerance and stress proteins as mechanisms of invasive species success. *PLOS ONE* 6:e14806.

547

548

Table 1(on next page)

Adapter and primer sequences

Adapter and primer sequences used for MS-AFLP protocol

Adapters	Sequence 5'- 3'
<i>Eco</i> RI-adapter F	5'-CTC GTA GAC TGC GTA CC-3'
<i>Eco</i> RI-adapter R	5'-AAT TGG TAC GCA GTC TAC-3'
<i>Hpa</i> II and <i>Msp</i> I-Adapter F	5'-GAC GAT GAG TCT AGA A-3'
<i>Hpa</i> II and <i>Msp</i> I-Adapter R	5'-CGT TCT AGA CTC ATC-3'
Pre-selective primers	
<i>Eco</i> RI-A	5'-GAC TGC GTA CCA ATT CA-3'
<i>Hpa</i> II and <i>Msp</i> I -T	5'-GAT GAG TCT AGA ACG GT-3'
Selective primers	
<i>Eco</i> RI + AAG	5'-GAC TGC GTA CCA ATT CAA G-3'
<i>Eco</i> RI + ACT	5'-GAC TGC GTA CCA ATT CAC T-3'
<i>Hpa</i> II / <i>Msp</i> I + TAC	5'-6-FAM-GAT GAG TCT AGA ACG GTA C-3'
<i>Hpa</i> II / <i>Msp</i> I + TCC	5'-6-FAM-GAT/ GAG TCT AGA ACG GTC C-3'

Table 2 (on next page)

CCGG sites where methylation sensitive restriction enzymes cleave to generate methylation dependent fragment patterns.

CCGG sites where methylation sensitive restriction enzymes (HpaII and MspI) cleave (Yes) or do not cleave (No) to generate methylation dependent fragment patterns. Both MspI and HpaII recognise CCGG sites and cleave unmethylated CCGG sites (1/1), but MspI cannot cleave when the outer cytosine is fully or hemimethylated (^m), and HpaII cannot cleave when the inner or outer cytosine is methylated on both strands. Cleaving by both enzymes is blocked when both cytosines are methylated. From this the methylation state of restriction sites can be scored (e.g., methylated (1/0 or 0/1) unmethylated (1/1) and uninformative (0/0)).

	Restriction sites		HpaII	Mspl	Fragment classification
Type I	5'-CCGG GGCC-5'		Yes	Yes	Unmethylated 1/1
Type II	5'- ^m CCGG GGCC-5'	5'- ^m C ^m CGG GGCC-5'	Yes	No	Hemimethylated 1/0
Type III	5'-C ^m CGG GG ^m CC-5'		No	Yes	Internal cytosine methylation 0/1
Type IV	5'- ^m C ^m CGG GG ^m C ^m C-5'	5'- ^m CCGG GGC ^m C-5'	No	No	Uninformative 0/0

Figure 1(on next page)

Colony growth rates at 19°C, 25°C and 27°C.

Colony growth rates (mean growth per day, mm \pm 1 s.e.) at 19°C, 25°C and 27°C. Significant differences between treatments denoted by different letters (P < 0.05, Tukey's test HSD).

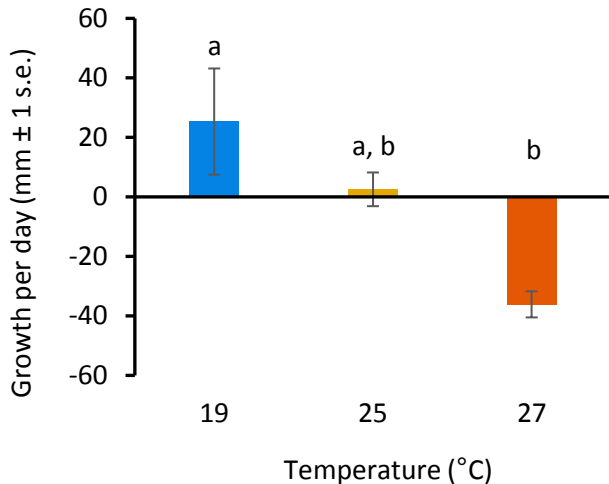


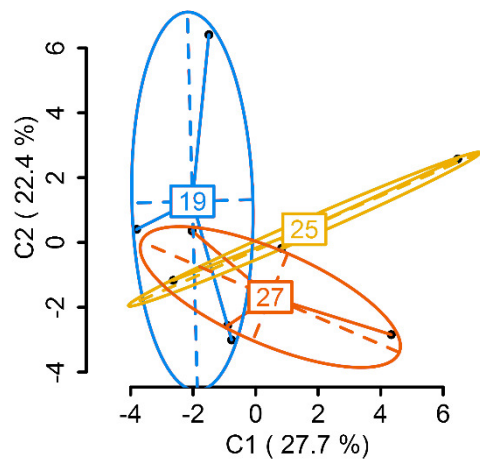
Figure 2 (on next page)

Principal Coordinate Analysis (PCoA) of methylation (MSL) differences between colonies exposed 19°C, 25°C and 27°C.

Principal Coordinate Analysis (PCoA) of methylation (MSL) differences: **A)** between colonies at Time 0 (T0), prior to elevated temperature exposure (baseline methylation); **B)** between colonies at Time 1 (T1) following a gradual temperature increase of 1°C per day until all treatment temperatures were reached: 19°C (control) (n = 3), 25°C (n = 3) and 27°C (n = 3); **C)** at Time 2 (T2)* after 3 days of elevated temperature exposure **D)** between sampling time points (T0, T1, T2) in colonies held at 19°C; **E)** 25°C; **F)** 27°C*. The first two coordinates (C1 and C2) are shown with the percentage of variance explained by them. Points in each group cloud represent individuals from different groups. Temperature labels show the centroid for the points cloud in each group. Ellipses represent average dispersion of those points around their centre (Pérez-Figueroa 2013) . AMOVA tests for significant differences in methylation (MSL) are shown in supplementary material, Tables S3 and S5. * represents significant difference (p < 0.05).

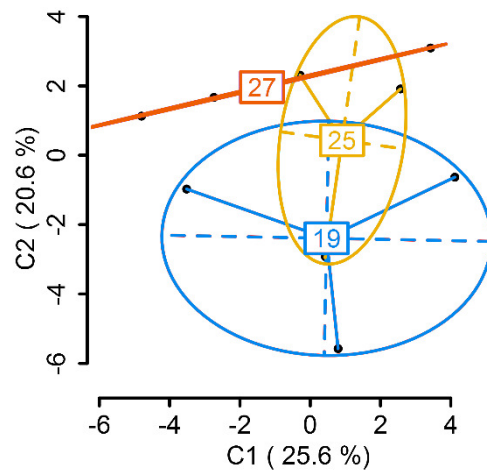
A:

Time 0



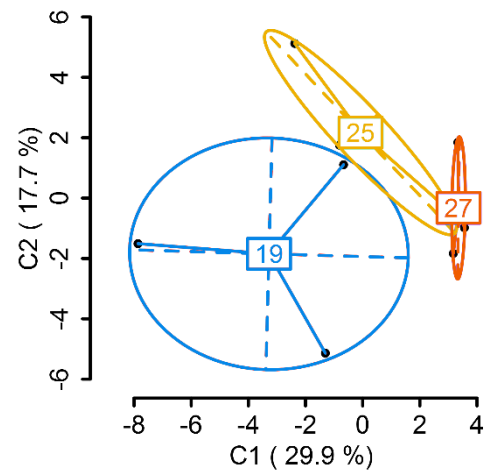
B:

Time 1



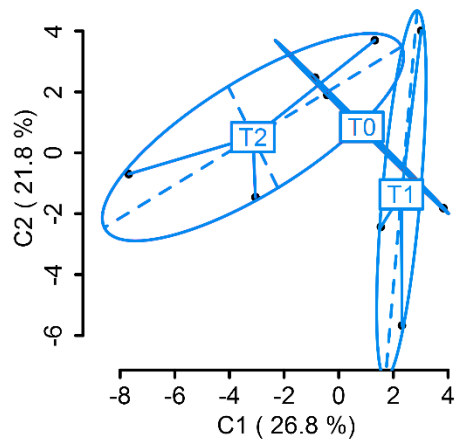
C:

Time 2



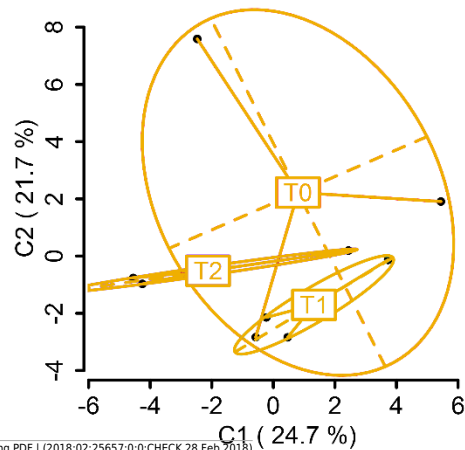
D:

19°C



E:

25°C



F:

27°C

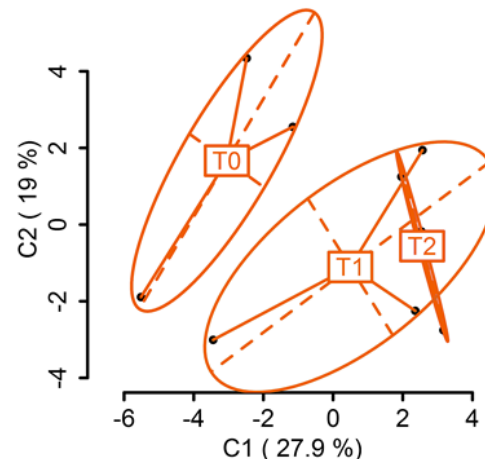


Figure 3(on next page)

Colony growth rates at 26, 28, 30, 32, and 34 PSU.

Colony growth rates (mean growth per day, mm \pm 1 s.e.) at 26, 28, 30, 32, and 34 PSU.

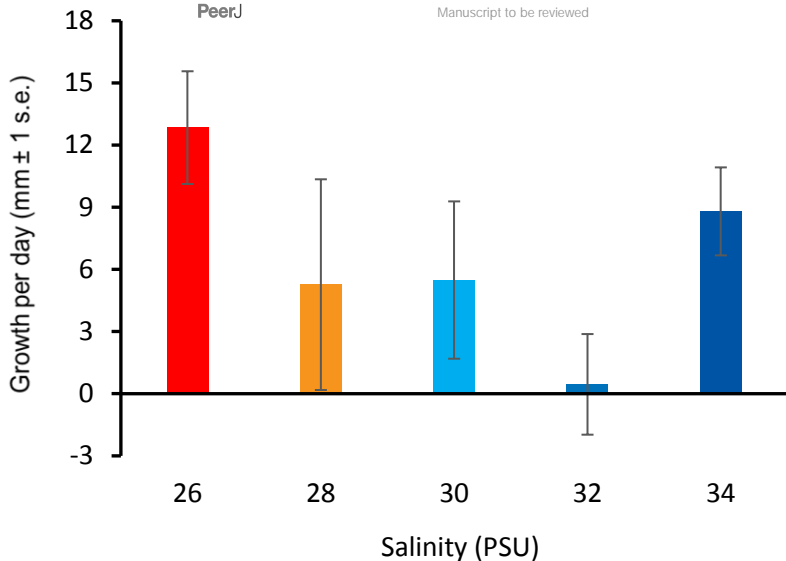


Figure 4(on next page)

Principal Coordinate Analysis (PCoA) of methylation (MSL) differences between colonies exposed to 34, 32, 30, 28 and 26 PSU.

Principal Coordinate Analysis (PCoA) of methylation (MSL) differences: **A)** between colonies at Time 0 (T0), prior to salinity treatment exposure (baseline methylation); **B)** between colonies at Time 1 (T1) following a gradual salinity decrease of 1 PSU per day until all salinity treatments were reached: 34 PSU (control) (n = 3), 32 PSU (n = 3), 30 PSU (n = 3), 28 PSU (n = 3) and 26 PSU (n = 3); **C)** at Time 2 (T2) after 3 days of elevated temperature exposure **D)** between sampling time points (T0, T1, T2) in colonies held at 26 PSU **E)** 28 PSU; **F)** 30 PSU; **G)** 32 PSU and **H)** 34 PSU. The first two coordinates (C1 and C2) are shown with the percentage of variance explained by them. Points in each group cloud represent individuals from different groups. Labels show the centroid for the points cloud in each group. Ellipses represent average dispersion of those points around their centre (Pérez-Figueroa 2013) . AMOVA tests for significant differences in methylation (MSL) are shown in supplementary material, Tables S7 and S8. No significant differences between groups were found. 1

