Effects of religiously motivated harvesting on the density and size composition of a sulfur cave-dwelling population of *Poecilia mexicana*

The poeciliid fish Poecilia mexicana, successfully colonized a sulfur cave in Tabasco, Mexico and gained increasing importance as a model organism to examine evolutionary adaptations to extreme environmental conditions. Closely linked to the ecology of the cave molly population are the local Zoque peoples every year, they harvest cave fish in a religious ceremony called La Pesca, which consists of the introduction of 'barbasco' roots containing the fish toxicant rotenone. Basic ecological data on the endemic cave molly population is still missing; therefore, our first aim was to provide data on population densities, size class compositions and use of different microhabitats. We tested for a signature of La Pesca on populations that are annually exposed to the toxicant, and populations not exposed to barbasco. We found overall densities to be strongly increased at sites annually exposed to the ritual, along with a marked shift in size/age compositions towards an overabundance of juvenile fish. We tentatively argue the annual harvest locally diminishes competition (and possibly, cannibalism by large adults), which is followed by a phase of overcompensation of fish densities. This line of interpretation, however, needs to be discussed with caution, as environmental gradients inside the Cueva del Azufre (like higher H2S concentrations in deeper portions) may also contribute to the observed differences among populations.

- Jonas Jourdan^{1,2*}, David Bierbach³, Rüdiger Riesch⁴, Angela Schießl¹, Adriana Wigh¹, Lenin
- 2 Arias-Rodriguez⁵, Jeane Rimber Indy⁵, Sebastian Klaus¹, Claudia Zimmer¹ and Martin Plath¹
- 3 ¹Evolutionary Ecology Group, Goethe University of Frankfurt, Max-von-Laue-Straße 13, D-60438 Frankfurt am
- 4 Main, Germany
- ² Biodiversity and Climate Research Centre (BiK^F), Senckenberganlage 25, 60325 Frankfurt am Main, Germany.
- 6 ³Leibniz-Institute of Freshwater Ecology and Inland Fisheries, Müggelseedamm 310, D-12587 Berlin, Germany
- 8 ⁵ División Académica de Ciencias Biológicas, Universidad Juárez Autónoma de Tabasco (UJAT), C.P. 86150
- 9 Villahermosa, Tabasco, México
- 10 *Corresponding author's email: JonasJourdan@googlemail.com

Introduction

12 Cave fishes are emerging as model systems to study regressive evolutionary processes	like	e t	the
---	------	-----	-----

- reduction of eyes and pigmentation typically accompanying the colonization of caves by
- previously surface-dwelling species (<u>Jeffery 2009</u>; <u>Romero & Green 2005</u>). Beside the well-
- for example

 15 known model organism for EvoDevo studies, the characid Astyanax mexicanus, another cave fish

 13 a model organism for EvoDevo studies
- 16 _ that has received comparable attention in the study of cave evolution, the cave form of a Mexican
- live-bearing fish, the so-called cave molly (Poecilia mexicana; Parzefall 1993; Parzefall 2001).
- 18 Poecilia mexicana has adapted to the vastly divergent ecological conditions inside a South
- 19 Mexican sulfide cave, the Cueva del Azufre (also referred to as Cueva Villa Luz or Cueva de las
- 20 Sardinas: Gordon & Rosen 1962). The cave molly differs from its surface-dwelling ancestors in a
- 21 distinct set of morphological, physiological, behavioral, and life-history traits; e.g., cave mollies
- 22 have reduced eye size and reduced pigmentation (Parzefall 2001; Riesch et al. 2010; Tobler et al.
- 23 2008a; Tobler et al. 2011b). Cave environments are assumed to be energy limited compared to
- 24 photosynthetically based epigean habitats (<u>Hüppop 2005</u>) and fish densities reported for several
- different cave systems are usually low, with often less than one individual per m² (Trajano 2001).
- 26 In contrast, the Cueva del Azufre is a sulfidic habitat and nutrient-rich due to chemoautotrophic
- 27 primary production (Hose & Pisarowicz 1999; Summers Engel 2005). Although the cave molly
- 28 has been established as a model to examine evolutionary adaptations to extreme environmental
- 29 conditions (namely hydrogen sulfide and absence of light), population densities have not yet been
- 30 quantified in the Cueva del Azufre system. Moreover, even less is known about the influence of
- the local indigenous Zoque people and an annual ceremony named 'La Pesca' on the cave molly's
- 32 ecology. The Cueva del Azufre is sacred to Zoque people, and once a year, at the first Sunday of
- 33 the Easter week, the Zoque enter the cave and introduce rotenone- and deguelin-containing
- 34 barbasco roots (Lonchocarpus sp., Fabaceae) into the water. Rotenone is an inhibitor of the

mitochondrial complex-I of the respiratory chain, thus depressing cellular respiration (Singer & 35 Ramsay 1994). Barbasco is introduced into the water in the middle portion of the cave, so only 36 downstream cave chambers are affected, while (sub-)populations upstream of the release site— 37 like the population inhabiting cave chamber X (nomenclature after Gordon & Rosen 1962)—are 38 not affected (Fig. 1). Capture of poisoned cave fish is facilitated by the anesthetic effect of 39 barbasco, as narcotized fish are flushed out of the cave, where they are harvested using wooden 40 baskets, and afterwards cooked and eaten as part of a religious ceremony honoring the Rain Gods 41 (Tobler et al. 2011a). The yield of the annual harvest is considered to be indicative of the quality 42 of the subsequent crop harvest (Hose & Pisarowicz 1999; Tobler et al. 2011a). 43 Population genetic analyses using nuclear microsatellites not only detected strong 44 population genetic differentiation between cave mollies and adjacent surface populations, but 45 also some degree of genetic differentiation among populations from different cave chambers 46 inside the Cueva del Azufre (Plath et al. 2007; Tobler et al. 2008a), and so it appears as if cave 47 mollies do not migrate freely between different cave chambers. Intriguingly, Tobler et al. (2011a) 48 found (sub-)populations living downstream of the barbasco-release site to show an increased 49 rotenone tolerance (Tobler et al. 2011a). Annual harvests are enormous and amount to several 50 thousand individuals, and the ceremony is likely to have taken place for centuries (Hose & 51 Pisarowicz 1999), so it ought to have left further footprints on P. mexicana populations annually 52 53 exposed to it. From a conservational point of view, knowledge about whether and how human activities 54 affect teleost populations is especially pertinent in the case of locally adapted populations that are 55 endemic to a small distribution area. Furthermore, the annually recurring anthropogenic 56 disturbance may act as a constant selection factor in the system, removing a large proportion of 57 certain age classes

small (and thus young) individuals from the affected populations.

59	Evidence from a broad range of taxa exemplifies how man-made, novel selection factors
60	can profoundly alter the evolutionary trajectories of populations (Smith & Bernatchez 2008;
61	Tuomainen & Candolin 2011). Well-known examples include industrial melanism in butterflies
62	(Kettlewell 1961), pesticide- and heavy metal-resistance in insects (Antonovics et al. 1971;
63	McKenzie & Batterham 1994), and convergent morphological shifts in different stream-dwelling
64	fishes as a consequence of altered flow regimes following the construction of impoundments
65	(Franssen et al. 2013). Today, teleost fishes are increasingly under selection resulting from
66	commercial or recreational fisheries (Jørgensen et al. 2007), both of which are known to have
67	various effects on teleost populations, including the loss of genetic variability and (plastic or
68	heritable) alterations of the morphological phenotype (Hutchings & Fraser 2008). Moreover,
69	several studies demonstrated heritable changes in behavioral, physiological and life-history traits
70	in fish populations that had been harvested experimentally (Biro & Post 2008; Walsh et al. 2006).
71	In the present study, we provide first data of local densities within different chambers of
72	the Cueva del Azufre and further tested the hypothesis that La Pesca affects local
73	(sub-)population densities as well as age/size class distributions. We predicted to find generally
74	lower population densities, and especially fewer large-bodied individuals downstream of the
75	barbasco-release site. We used a non-invasive technique to repeatedly assess fish densities and
76	size-distribution patterns (as a proxy for age) inside the Cueva del Azufre (up- and downstream of
77	the barbasco-release site) and in the sulfidic creek leaving the cave (El Azufre). Moreover, given
78	the high structural heterogeneity of the water course inside the Cueva del Azufre with respect to
79	water depth and flow velocity (Hose & Pisarowicz 1999), and because Croft et al. (2004)
80	reported on size-specific preferences regarding water depth in another poeciliid, the Trinidadian
81	guppy (P. reticulata), we combined our assessment of fish densities with an investigation of
82	microhabitat use of different size classes of cave mollies.

Material and Methods

84	Study	system
04	Dinuy	system

- 85 Locally adapted subterranean populations of *P. mexicana* can be found in at least two different
- 86 limestone caves in the vicinity of the southern Mexican city of Tapijulapa (state of Tabasco,
- 87 México): the Cueva del Azufre (Gordon & Rosen 1962) and the much smaller, non-sulfidic
- 88 Cueva Luna Azufre (Tobler et al. 2008c). La Pesca, however, affects only the Cueva del Azufre,
- which is about 500 600 m deep and divided into 13 different cave chambers (I-XIII), with the
- 90 innermost chamber being XIII (Gordon & Rosen 1962). Several springs discharge water with
- 91 high concentrations of hydrogen sulfide (H₂S) into the creek draining the cave (Tobler et al.
- 92 2006). The cave creek forms a highly heterogeneous mosaic of shallow pools and backwaters that
- 93 are partially divided by swift flowing riffle passages (Gordon & Rosen 1962; Hose & Pisarowicz
- 94 1999). While the front cave chambers receive some dim light through cracks in the ceiling, the
- 95 inner parts of the cave are lightless. Consequently, (sub-)populations experience divergent
- 96 selection regimes regarding light exposure, with populations from innermost chambers living
- 97 under perpetually dark conditions, while those from front chambers are exposed to dim sunlight
- 98 through a number of sky lights (Fontanier & Tobler 2009).
- Upon leaving the underground, the sulfidic creek draining the Cueva del Azufre is
- 100 henceforth called 'El Azufre'. After meandering for approximately 1.5 km, it eventually drains
- into the Río Oxolotán, which is part of the Río Grijalva drainage system. Despite the gradual
- oxidation of H₂S to sulfate and elemental sulfur with increasing distance from the sulfide sources
- 103—increasingly rendering the water turbid—and despite the influx of some smaller clear water
- affluents, El Azufre still has a remarkably high H_2S concentration of up to $\sim 40 \mu Mol$ (Schlupp et
- 105 <u>al. 2013; Tobler et al. 2006</u>).

130

106 Study sites and data collection

We compared the abundance and distribution of different size classes of P. mexicana among 107 different sampling sites along a transect starting at chamber X in the Cueva del Azufre, following 108 the water flow outside the cave to the confluence of El Azufre with the first freshwater influx, 109 110 __namely the Clear Creek (Fig. 1A). This transect, therefore, covers sample sites located upstream of the release point of barbasco (sample point in chamber X), several sites directly downstream of 111 the release point of barbasco that are strongly affected by the annual ceremony (three sites in 112 chamber V; CV-1, CV-2, and CV-3; Fig. 2B-C), and sites outside of the cave that are not under 113 the influence of barbasco [clear creek (CC) and its confluence with EA (EA-con; Fig. 2A)], as 114 well as two surface sites of El Azufre that are annually exposed to barbasco (EA-ex, EA-mf), 115 even though concentrations are probably considerably lower than inside the cave (Table 1). 116 Field work was conducted in January 2010, i.e., about 9 month after the last La Pesca 117 ceremony in 2009 (L. Arias-Rodriguez, pers. obs.). At each of the eight sample sites, we defined 118 sampling grids consisting of 50 × 50 cm quadrants with wooden sticks fixed in the ground (or 119 stones where a grid angle fell on the shore). The number of quadrants was mostly 25 per 120 sampling grid (i.e., 5 × 5 quadrants). In the narrow non-sulfidic surface creek (CC), however, the 121 arrangement of quadrants was more longitudinal (4 × 7 = 28 quadrants), and in chamber V, where 122 a particularly high degree of structural heterogeneity precluded defining larger grids, one 123 sampling site of 5 × 5 quadrants and two smaller ones (15 and 8 quadrants, respectively) could be 124 defined (Table 1). The grids reflected the natural variation in water depth, flow velocity, and 125 substrate types, thus covering the range of different microhabitats inhabited by mollies (an 126 example is shown in Fig. 3). 127 Daily measurements took place between 11:00 a.m. and 4:30 p.m. Each site was visited at 128

least 5 times (mean \pm SD = 6.25 ± 1.16) on consecutive days. During the counts, we slowly approached a site while trying to avoid any movements that would cause the resident fish to flee,

We and thus counted juveniles (< 10 mm standard length (SL)), sub-adults (10-30 mm) and adults (> 131 30 mm) in each quadrant. The observer was standing motionless at least 1.5 m downstream from 132 the respective quadrant. Sizes were estimated qualitatively, aided by a prior training session that 133 used wooden sticks of known size as a reference. Our definition of adults roughly followed 134 Riesch et al. (2010), who determined the mean (± SD) standard length of reproducing females to 135 be 31.44 ± 4.40 mm (El Azufre) and 36.97 ± 4.59 mm (Cueva del Azufre, chambers V and X). 136 Habitat parameters were assessed after the last fish count. For each quadrant, we 137 determined water depth using a makeshift wooden ruler stuck vertically into the water on five 138 random locations and calculating the mean from those five measurements. Flow velocity was 139 measured on the water surface by scoring the time a small wooden stick of about 3 cm length 140 took to float through the whole length of a quadrant (measurement was repeated five times per 141 quadrant). Mean surface flow velocity was then expressed as m*s-1. Research follows the 142 authorizations from CONAPESCA-DGOPA.09004.041111.3088 and Tacotalpa, Tabasco municipality. 143

- 144 Statistical analysis
- Our first question was whether population densities differed between sampling sites. We used
- data for the different quadrants per site (averaged from the repeated measurements) and
- expressed density as total numbers of individuals per square meter. Density estimates per
- quadrant were used as the dependent variable in a univariate general linear model (GLM) with
- 'sampling site' as a fixed factor. We initially entered 'mean water depth' $(F_{1,162} = 0.12, P = 0.98)$
- and 'mean flow rate' ($F_{1,162} = 0.22$, P = 0.64) as covariates, but removed them from the final
- 151 model since neither had a significant effect (also none of the interaction terms were significant).
- We used Fisher's LSD tests for pairwise post hoc comparisons between sites.
- A further question was whether size-class compositions differed among sample sites and
- whether distribution patterns would be stable between repeated sampling days. We used the Bray-

Curtis index (Bray & Curtis 1957) to estimate pairwise similarities between each sampling point

(calculated with the R-package 'ecodist 1.2.7'; Goslee and Urban 2012; R Development Core

Team 2008), and used these for non-metric multidimensional scaling ('NMDS PROXSCAL'

function in SPSS 21). To detect clusters, we used the 'two step cluster analysis' function based on

Euclidian distances and the Bayesian information criterion.

Our first analysis detected pronounced variation in population densities and size 160 distributions (see results), so we decided to analyze potential effects of water depth and flow 161 velocity (i.e., microhabitat choice) in a site-wise fashion. We concentrated on sites inside the 162 Cueva del Azufre (CV-1, CV-3 and CX) where (a) fish densities were sufficiently high and (b) 163 sufficient variability of those environmental variables was found to allow for a meaningful 164 analysis. For each site, fish density per quadrant was entered as the dependent variable in 165 repeated measures (rm) GLMs with 'size class' (three levels) as the repeated measurement. We 166 grouped water depth (< 5 cm, 5 - 10 cm, > 10 cm) and flow velocity (< 10 cm*s⁻¹, 10 - 50 cm*s⁻¹ 167 ¹, > 50 cm*s⁻¹) into three classes each and used these habitat parameters as fixed factors. 168 However, neither the main factor 'flow velocity' nor any interaction term involving 'flow 169 velocity' had a significant effect in any of the three site-specific models (CV-1: $F_{4,38} = 1.27$, P =170 0.30; CV-3: $F_{2,4} = 3.28$, P = 0.14; CX: $F_{2,20} = 0.44$, P = 0.65), and so we subsequently removed 171 172 this term from all models.

Results

- 174 Local population densities
- When comparing mean densities per quadrant across sites we detected a significant difference
- between sampling sites (GLM; $F_{7,164} = 32.49$, P < 0.001; Fig. 1B). Post hoc pairwise LSD tests
- 177 found most pairwise comparisons to be statistically significant; qualitatively, densities increased

- from surface sites (mean \pm SE across sites = 21.0 \pm 5.0 individuals*m⁻²) towards the cave (119.5)
- 179 ± 12.7 individuals*m⁻²). Also, sites downstream of the barbasco release-site (chamber V; 162.3 ±
- 180 16.1 individuals*m⁻²) had considerably higher densities than the site in chamber X that lies
- upstream of the release-site (37.4 \pm 4.8 individuals*m⁻²).
- 182 Differences in size-class compositions
- 183 The NMDS based on Bray-Curtis similarities found data from repeated sampling days to cluster
- together, suggesting that the observed size-class compositions were stable over time (Fig. 4).
- 185 There were three distinct clusters, and in only two cases were single sampling days of a given
- sampling site assigned to the 'wrong' cluster. The first cluster comprised the three sample sites in
- 187 cave chamber V and EA-con. Samples had high overall densities and were composed mostly of
- small individuals. Cluster two comprised the rearmost cave site CX. Samples in this cluster were
- characterized by intermediate densities but a particularly high proportion of large individuals.
- 190 Cluster three comprised all surface sites except EA-con and was characterized by overall low
- densities and mostly intermediate-sized fish (Fig. 1C).
- 192 Microhabitat use of different size classes
- 193 In the rmGLMs treating the different size-classes as the repeated measurement, the interaction
- 194 term 'size-class × water-depth' had a significant effect for two of the three sampling sites
- included in this analysis—notably, those sites with the most variation in water depth (Fig. 5). This
- result is indicative of differences in microhabitat use between different size classes of cave
- 197 mollies: generally, larger fish were found in deeper areas, while smaller fish resided in shallow
- 198 parts. A significant main effect of the repeated measurement ('size class') in all three analyses
- 199 confirms the overabundance of small-sized fish in cave chamber V, and of large-bodied fish in
- 200 chamber X (Fig. 5).

Discussion

be consistent maily

Here, we provide detailed information on population densities of cave-adapted *P. mexicana* in the Cueva del Azufre. Repeated measurement in different cave chambers uncovered very stable patterns of high densities, confirming qualitative estimates provided by Parzefall (1993). Density estimates of *P. mexicana* in the cave are extraordinarily high and exceed those of other cave fishes which are usually low, with often less than one individual per m² (Trajano 2001). Furthermore, densities are higher inside the cave compared to adjacent surface populations.

The annual 'La Pesca' ceremony leads to a strong temporary reduction of local fish densities in those cave chambers that are situated downstream of the barbasco release site (Tobler et al. 2011a). Our study was conducted approximately nine months after the last ceremony, but given rather long generation times in *P. mexicana* (roughly 3–6 months for males and 7–10 months for females from birth until reaching maturation under common-garden rearing conditions; Riesch, Reznick, Plath & Schlupp, submitted), we predicted to find lower (sub-)population densities and especially fewer large-bodied individuals downstream of the site in the Cueva del Azufre where barbasco is annually released. Instead, while fish densities were generally high in the cave, they were highest downstream of the barbasco release site. However, in line with our prediction, sample sites affected by the release of barbasco had population structures that were strongly shifted towards an overabundance of the smallest size classes (i.e., juveniles). These patterns were stable when repeated samplings from subsequent days were compared.

Migration within the Cueva del Azufre is unidirectional, out of the cave, and migration among different cave chambers occurs just to a small extent, which results in population genetic differentiation as shown based on nuclear microsatellites (<u>Plath et al. 2007</u>), and is also reflected

omony fish from

in morphological differences between different cave chambers (Fontanier & Tobler 2009). Hence, 224 re-colonization of the affected sites from other parts of the cave (i.e., source-sink dynamics) is 225 unlikely, and the observed recovery of the respective populations likely represents an 226 autochthonous effect. After the temporal decline in population density, the few surviving animals 227 benefit from reduced intraspecific resource competition. While detritus and green algae are the 228 dominant food sources of surface-dwelling P. mexicana from non-sulfidic streams, diets of 229 conspecifics in the sulfidic surface and cave streams are dominated by chemoautotrophic (sulfur) 230 bacteria and aquatic invertebrates (like larvae of the dipteran Goeldichironomus fulvipilus and 231 232 small snails; Roach et al. 2011). In particular, access to invertebrate prey could be favored not only by the absence of competing fish species, but also by relaxed competition among the 233 surviving adult P. mexicana. Generally, relaxed competition translates into higher growth rates, 234 235 faster maturation, and increased adult fecundity (Stearns 1976) which may lead to stage specific 236 biomass overcompensation thereby compensating for the removal of individuals from the population (de Roos et al. 2007; Schröder et al. 2009; Werner & Gilliam 1984). This idea 237 received support from empirical harvesting experiments that found the negative relationship 238 between adult mortality and abundance/density to be reversed if mortality does not affect a 239 certain portion of the population. Experimental studies on laboratory populations of the poeciliid 240 fish Heterandria formosa showed that biomass of the juvenile size class increased in response to 241 intermediate adult mortality rates (Schröder et al. 2009). Another study showed that a pathogen 242 outbreak in a wild perch population (Perca fluviatilis) was followed by a biomass 243 overcompensation of the juvenile stage in response to increased adult mortality. Age-specific 244 245 adult fecundity and body mass of one- and two year old perch increased after the disease outbreak, suggesting that mortality released adult perch from competition and cannibalism, 246 thereby increasing somatic and reproductive growth (Ohlberger et al. 2011). In the Cueva del 247 Azufre, the stage-specific biomass overcompensation leads to increasing population densities, 248

	would ansittent
249	with high numbers of juvenile fish. This will soon result in cave molly populations regaining the
250	high densities seen before La Pesca, again leading to increased competition. This is consistent
251	with earlier observations of cave mollies showing reduced body condition (measured, e.g., as fat
252	content) compared to fish from benign surface sites (Riesch et al. 2010; Riesch et al. 2011).
253	We are aware of potential caveats of our discussion, as not all differences reported here

may be due to the annual La Pesca ceremony; e.g., environmental heterogeneity may also contribute to population differences. The highest H₂S concentrations (> 300 μM) are found in parts of chamber X, where most large sulfide sources are located, while concentrations in chamber V are lower (2 – 32 μM), as H₂S is increasingly oxygenated with increasing distance from the sulfide sources (Tobler et al. 2006). However, ecotoxicological experiments repeatedly found small adults to have higher H₂S-resistance than large-bodied adults, possibly reflecting senescence effects (Plath et al. 2013; Tobler et al. 2011b). Hence, we would expect fish in chamber X to actually be smaller than fish from chamber V if different H₂S concentrations—but west

Furthermore, cave chambers may differ in predation regimes. Inside the Cueva del Azufre, aquatic water bugs of the genus *Belostoma* prey upon cave mollies and *Belostoma* prefer large over small cave mollies as prey (<u>Plath et al. 2011</u>; <u>Tobler et al. 2008b</u>; <u>Tobler et al. 2007</u>). Mark-recapture analysis found individual densities of water bugs to be approximately one individual per m² in chamber V (<u>Tobler et al. 2007</u>), and while empirical data are as yet lacking, observational evidence over several years of field work suggests that densities are much lower in the innermost chambers. Hence, human-induced selection and predation by *Belostoma* ought to have very similar effects on the populations exposed to them. We are inclined to argue, however, that the relative influence of *Belostoma* predation is considerably lower than the effects of the massive annual fish harvest. Previous reports of increased rotenone-resistance in fish from

chamber V, but not chamber X (<u>Tobler et al. 2011a</u>), confirm that La Pesca undoubtedly has a strong selective influence on populations annually exposed to it.

Belostoma predation might also explain microhabitat use of different size classes of cave mollies. Belostoma are typically found on rocks at the water's edge (Tobler et al. 2008b), and so large cave mollies—being preferred by the water bugs (Plath et al. 2011)—could use deeper parts of the water column to avoid predation risk. The preference for large size-classes was confirmed for another belostomatid preying on mosquitofish (Schumann et al. 2012). On the other hand, small fish could avoid filial cannibalism, which is known from other poeciliids (Loekle et al. 1982; Nesbit & Meffe 1993), by using shallow parts of the water column.

In summary, we found remarkable fish densities of more than 200 individuals per m² in some parts of the cave. The annual La Pesca appears to have multiple effects on the population ecology and evolutionary trajectory of cave mollies. While human influences on highly endemic, locally adapted populations (or, in terms of conservation biology, evolutionary significant units; Moritz 1994) generally are to be evaluated as highly problematic, management plans for cave mollies ought to consider the important role La Pesca plays in the religion of the local human population. Carried out in the traditional way, fish populations in downstream cave chambers can obviously recover after the ceremony. However, we wish to highlight the necessity to critically review that those practices do not affect deeper parts of the cave and that no commercially available, more efficient fish toxins are being employed in the future.

Acknowledgements

We thank Justina Dzienko, Nora Karau, Alexandra Oranth and Stefan Stadler for their help in the field. Artwork (drawing of a cave molly female; Fig. 1A) was prepared by M. Ziege.

References

295

303 304

305

306

307

308 309

310

311

312

313

314

315

316

317318

- Antonovics J, Bradshaw AD, and Turner R. 1971. Heavy metal tolerance in plants. *Advances in ecological research* 7:1-85.
- Biro PA, and Post JR. 2008. Rapid depletion of genotypes with fast growth and bold personality traits from harvested fish populations. *Proceedings of the National Academy of Sciences* 105:2919-2922.
- Bray JR, and Curtis JT. 1957. An ordination of the upland forest communities of southern Wisconsin. *Ecological monographs* 27:325-349.
 - Croft DP, Botham MS, and Krause J. 2004. Is sexual segregation in the guppy, *Poecilia* reticulata, consistent with the predation risk hypothesis? *Environmental Biology of Fishes* 71:127-133.
 - de Roos AM, Schellekens T, van Kooten T, van de Wolfshaar KE, Claessen D, and Persson L. 2007. Food-dependent growth leads to overcompensation in stage-specific Biomass when mortality increases: The influence of maturation versus reproduction regulation. *American Naturalist* 170:E59-E76.
 - Fontanier ME, and Tobler M. 2009. A morphological gradient revisited: cave mollies vary not only in eye size. *Environmental Biology of Fishes* 86:285-292.
 - Franssen NR, Harris J, Clark SR, Schaefer JF, and Stewart LK. 2013. Shared and unique morphological responses of stream fishes to anthropogenic habitat alteration. *Proceedings of the Royal Society B-Biological Sciences* 280.
 - Gordon MS, and Rosen DE. 1962. A cavernicolous form of the poeciliid fish *Poecilia sphenops* from Tabasco, Mexico. *Copeia*:360-368.
 - Hose LD, and Pisarowicz JA. 1999. Cueva de Villa Luz, Tabasco, Mexico: Reconnaissance study of an active sulfur spring cave and ecosystem. *Journal of Cave and Karst Studies* 61:13-21.
- Hüppop K. 2005. Adaptation to low food. In: Culver DC, and White WB, eds. *Encyclopedia of caves*. Amsterdam: Elsevier, 4-10.
- Hutchings JA, and Fraser DJ. 2008. The nature of fisheries- and farming-induced evolution. *Molecular Ecology* 17:294-313.
- Jeffery WR. 2009. Regressive evolution in Astyanax cavefish. Annual review of genetics 43:25.
- Jørgensen C, Enberg K, Dunlop ES, Arlinghaus R, Boukal DS, Brander K, Ernande B, Gardmark
 A, Johnston F, and Matsumura S. 2007. Ecology-Managing evolving fish stocks. *Science* 318:1247-1248.
- Kettlewell HBD. 1961. The phenomenon of industrial melanism in Lepidoptera. *Annual review of entomology* 6:245-262.
- Loekle DM, Madison DM, and Christian JJ. 1982. Time dependency and kin recognition of cannibalistic behavior among poeciliid fishes. *Behavioral and Neural Biology* 35:315-332 318.
- McKenzie JA, and Batterham P. 1994. The genetic, molecular, and phenotypic consequences of selection for insecticide resistance. *Trends in Ecology & Evolution* 9:166-169.
- Moritz C. 1994. Defining'evolutionarily significant units' for conservation. *Trends in Ecology & Evolution* 9:373-374.
- Nesbit DH, and Meffe GK. 1993. Cannibalism frequencies in wild populations of the eastern mosquitofish (*Gambusia holbrooki*: Poeciliidae) in South Carolina. *Copeia* 3:867-870.
- Ohlberger J, Langangen Ø, Edeline E, Claessen D, Winfield IJ, Stenseth NC, and Vøllestad LA.

 2011. Stage-specific biomass overcompensation by juveniles in response to increased
- adult mortality in a wild fish population. *Ecology* 92:2175-2182.

351

352

353

354

355

356

357358

359

360

361

362

363

364

365

366

367 368

369 370

371

372

373

- Parzefall J. 1993. Behavioral ecology of cave-dwelling fishes. In: Pitcher TJ, ed. *Behavior of Teleost Fishes*. London: Chapman & Hall, 573 606.
- Parzefall J. 2001. A review of morphological and behavioural changes in the cave molly, *Poecilia mexicana*, from Tabasco, Mexico. *Environmental Biology of Fishes* 62:263-275.
- Plath M, Hauswaldt JS, Moll K, Tobler M, De Leon FJG, Schlupp I, and Tiedemann R. 2007.

 Local adaptation and pronounced genetic differentiation in an extremophile fish, *Poecilia mexicana*, inhabiting a Mexican cave with toxic hydrogen sulphide. *Molecular Ecology* 16:967-976.
 - Plath M, Pfenninger M, Lerp H, Riesch R, Eschenbrenner C, Slattery PA, Bierbach D, Herrmann N, Schulte M, Arias–Rodriguez L, Indy JR, Passow C, and Tobler M. 2013. Genetic differentiation and selection against migrants in evolutionarily replicated extreme environments. *Evolution*:n/a-n/a.
 - Plath M, Riesch R, Culumber Z, Streit B, and Tobler M. 2011. Giant water bug (*Belostoma* sp.) predation on a cave fish (*Poecilia mexicana*): effects of female body size and gestational state. *Evolutionary Ecology Research* 13:133-144.
 - Riesch R, Plath M, and Schlupp I. 2010. Toxic hydrogen sulfide and dark caves: life-history adaptations in a livebearing fish (*Poecilia mexicana*, Poeciliidae). *Ecology* 91:1494-1505.
 - Riesch R, Plath M, and Schlupp I. 2011. Toxic hydrogen sulphide and dark caves: pronounced male life-history divergence among locally adapted *Poecilia mexicana* (Poeciliidae). *Journal of Evolutionary Biology* 24:596-606.
 - Roach KA, Tobler M, and Winemiller KO. 2011. Hydrogen sulfide, bacteria, and fish: a unique, subterranean food chain. *Ecology* 92:2056-2062.
 - Romero A, and Green S. 2005. The end of regressive evolution: examining and interpreting the evidence from cave fishes. *Journal of Fish Biology* 67:3-32.
 - Schlupp I, Colston TJ, Joachim BL, and Riesch R. 2013. Translocation of cave fish (*Poecilia mexicana*) within and between natural habitats along a toxicity gradient. *Ecology of Freshwater Fish* 22:228-233.
 - Schröder A, Persson L, and de Roos AM. 2009. Culling experiments demonstrate size-class specific biomass increases with mortality. *Proceedings of the National Academy of Sciences of the United States of America* 106:2671-2676.
 - Schumann DA, Cavallaro MC, and Hoback WW. 2012. Size Selective Predation of Fish by *Hydrophilis triangularis* (Coleoptera: Hydrophilidae) and *Lethocerus americanus* (Hemiptera: Belostomatidae). *Journal of the Kansas Entomological Society* 85:155-159.
- Singer TP, and Ramsay RR. 1994. The reaction sites of rotenone and ubiquinone with mitochondrial NADH dehydrogenase. *Biochimica Et Biophysica Acta-Bioenergetics* 1187:198-202.
- Smith TB, and Bernatchez L. 2008. Evolutionary change in human-altered environments.
 Molecular Ecology 17:1-8.
- 380 Stearns SC. 1976. Life-history tactics: a review of the ideas. Quarterly review of biology 51:3-47.
- Summers Engel A. 2005. Chemoautotrophy. In: Culver DC, and White WB, eds. *Encyclopedia of caves*. Amsterdam: Elsevier, 90-102.
- Tobler M, Culumber Z, Plath M, Winemiller K, and Rosenthal G. 2011a. An indigenous religious ritual selects for resistance to a toxicant in a livebearing fish. *Biology letters* 7:229-232.
- Tobler M, DeWitt TJ, Schlupp I, García de León FJ, Herrmann R, Feulner PG, Tiedemann R, and Plath M. 2008a. Toxic hydrogen sulfide and dark caves: phenotypic and genetic divergence across two abiotic environmental gradients in *Poecilia mexicana Evolution*
- divergence across two abiotic environmental gradients in *Poecilia mexicana*. *Evolution* 62:2643-2659.
- Tobler M, Franssen CM, and Plath M. 2008b. Male-biased predation of a cave fish by a giant water bug. *Naturwissenschaften* 95:775-779.

399

400

401

402

403

404 405

406 407

408

409

410

- Tobler M, Palacios M, Chapman LJ, Mitrofanov I, Bierbach D, Plath M, Arias-Rodriguez L, de Leon FJG, and Mateos M. 2011b. Evolution in extreme environments: replicated phenotypic differentiation in livebearing fish inhabiting sulfidic springs. *Evolution* 65:2213-2228.
- Tobler M, Riesch R, de Leon FJG, Schupp I, and Plath M. 2008c. New and morphologically distinct population of cavernicolous *Poecilia mexicana* (Poeciliidae : Teleostei). *Environmental Biology of Fishes* 82:101-108.
 - Tobler M, Schlupp I, Heubel KU, Riesch R, de Leon FJG, Giere O, and Plath M. 2006. Life on the edge: hydrogen sulfide and the fish communities of a Mexican cave and surrounding waters. *Extremophiles* 10:577-585.
 - Tobler M, Schlupp I, and Plath M. 2007. Predation of a cave fish (*Poecilia mexicana*, Poeciliidae) by a giant water-bug (*Belostoma*, Belostomatidae) in a Mexican sulphur cave. *Ecological Entomology* 32:492-495.
 - Trajano E. 2001. Ecology of subterranean fishes: an overview. *Environmental Biology of Fishes* 62:133-160.
 - Tuomainen U, and Candolin U. 2011. Behavioural responses to human-induced environmental change. *Biological Reviews* 86:640-657.
 - Walsh MR, Munch SB, Chiba S, and Conover DO. 2006. Maladaptive changes in multiple traits caused by fishing: impediments to population recovery. *Ecology Letters* 9:142-148.
 - Werner EE, and Gilliam JF. 1984. The ontogenetic niche and species interactions in size-structured populations. *Annual Review of Ecology and Systematics* 15:393-425.

Table 1(on next page)

List of samling sites

Sampling sites, their abbreviation code as used throughout the article, numbers of quadrants examined, and details regarding barbasco release.

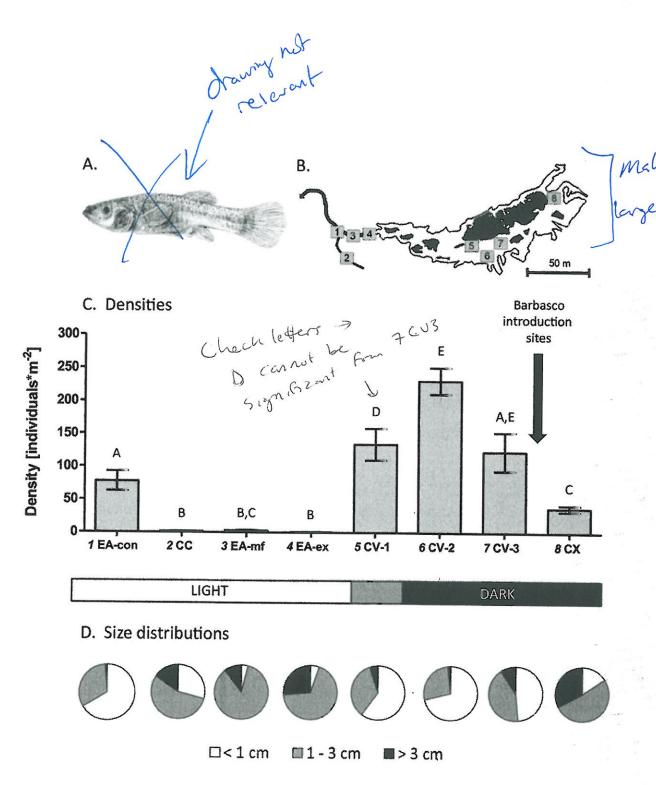
Tables

Table 1Sampling sites, their abbreviation code as used throughout the article, numbers of quadrants examined, and details regarding barbasco release.

Site code	Site	Number of quadrants	Affected by deposition of rotenone?	Approximate distance to upstream rotenone release site [m]
1 EA-con	El Azufre, confluence with Clear Creek	25	No (only partly)	150
2 CC	Clear Creek	28	No	-
3 EA-mf	El Azufre, mudflat	25	Yes	120
4 EA-ex	El Azufre, exit of the Cueva del Azufre	25	Yes	110
5 CV-1	Cueva del Azufre, Chamber V, site 1	25	Yes	0
6 CV-2	Cueva del Azufre, Chamber V, site 2	15	Yes	0
7 CV-3	Cueva del AzufreChamber V, site 3	8	Yes	0
8 CX	Cueva del Azufre, Chamber X	25	No	-

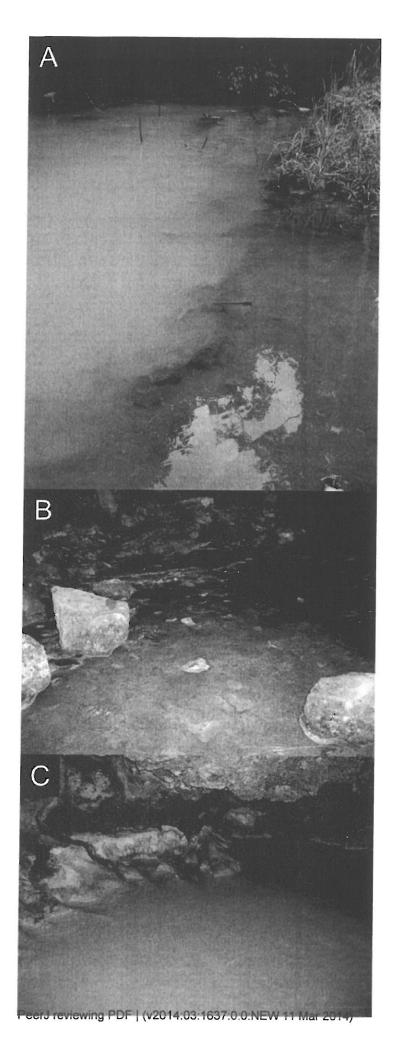
Study system and population densities.

(A) Drawing of a female cave molly. (B) Map of the study area showing the different sampling sites (*numbers*). 1 EA-con, 2 CC, 3 EA-mf, 4 EA-ex, 5 CV-1, 6 CV-2, 7 CV-3, 8 CX. With the exception of sampling site CV-1 all sampling sites inside the Cueva del Azufre are completely dark. Barbasco is released annually between chamber V (CV) and chamber X (CX). Three sampling sites inside chamber V were defined (CV-1 to CV-3). Downstream of the exit of the Cueva del Azufre (EA-ex), a rather homogeneous mudflat (EA-mf) was sampled. Further sampling sites were a small non-sulfidic creek (Clear Creek; CC) and its confluence with El Azufre (EA-con). (C) Mean (± SE) densities of mollies at each sampling site. Letters above the error bars signify statistically different groups (Fisher's LSD tests). (D) Size class compositions of mollies at the different sampling sites.



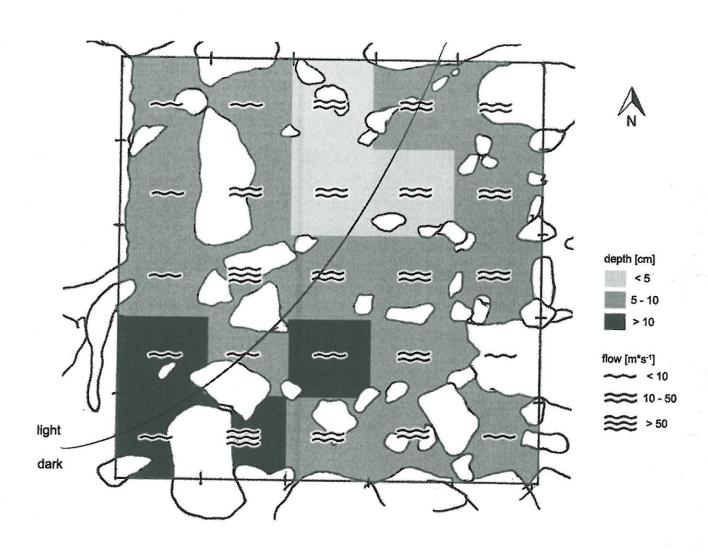
Pictures of probing sites

(A) El Azufre confluence with Clear Creek (1 EA-con), (B) Cueva del Azufre chamber V (6 CV-2) and (C) site 7 CV-3.



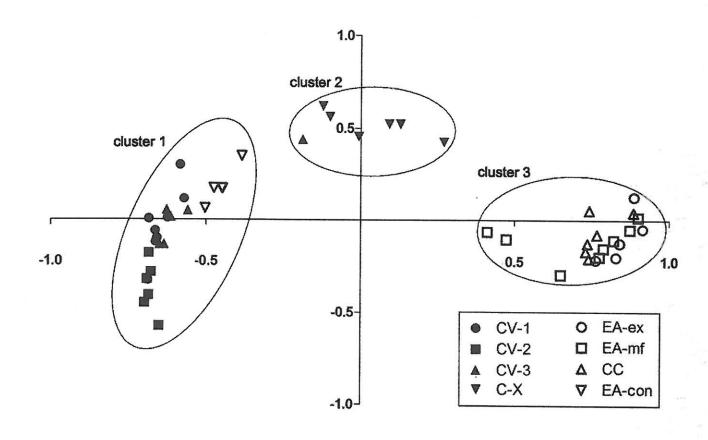
Examplary sketch of site 5 CV-1

showing the high degree of heterogeneity in flow regimes, water depth, substrate types, and (in this case) light regime.



Differences in size-class compositions of *Poecilia mexicana* in the Cueva del Azufre system.

Non-metric Multi-Dimensional Scaling (NMDS) plots based on Bray-Curtis similarities for each sampling site and day.



Population densities of cave mollies in the Cueva del Azufre.

Mean (\pm SE) densities of mollies, categorized in three size classes (< 1 cm, white, 1 – 3 cm, gray, and > 3 cm, black) in three water depths (< 5 cm, 5 – 10 cm, and > 10 cm). Results of rmGLMs are inserted. Note the different y-axis scales.

