

# The effect of climate change on the distribution of a tropical zoanthid (*Palythoa caribaeorum*) and its ecological implications

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Palythoa caribaeorum, a sea mat coral, is a zoanthid often dominant in shallow rocky environments along the west coast of the Atlantic Ocean, from the tropics to the subtropics. This species has high environmental tolerance and is a good space competitor in reef environments. Considering current and future scenarios in the global climate regime, this study aimed to model and analyze the distribution of *P. caribaeorum*, generating maps of potential distribution for the present and year 2100. The distribution was modeled using maximum entropy (Maxent) based on 327 occurrence sites retrieved from the literature. Calcite concentration, maximum chlorophyll-a concentration, salinity, pH, and temperature range yielded a model with the smallest Akaike information criterion (2649.8), and were used in the present and future distribution model. Data from the HadGEM2-ES climate model were used to generate the projections for the year 2100. The present distribution of *P. caribaeorum* shows that parts of the Brazilian coast, Caribbean Sea, and Florida are suitable regions for the species, as they are characterized by high salinity and pH and small temperature variation. An expansion of the species' distribution was forecast northward under mild climate scenarios, while a decrease of suitable areas was forecast in the south. In the climate scenario with the most intense changes, P. caribaeorumwould lose one-half of its suitable habitats, including the northernmost and southernmost areas of its distribution. The Caribbean Sea and northeastern Brazil, as well as other places under the influence of coastal upwellings, may serve as potential havens for this species.

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#### **Abstract**

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Palythoa caribaeorum, a sea mat coral, is a zoanthid often dominant in shallow rocky environments along the west coast of the Atlantic Ocean, from the tropics to the subtropics. This species has high environmental tolerance and is a good space competitor in reef environments. Considering current and future scenarios in the global climate regime, this study aimed to model and analyze the distribution of P. caribaeorum, generating maps of potential distribution for the present and year 2100. The distribution was modeled using maximum entropy (Maxent) based on 327 occurrence sites retrieved from the literature. Calcite concentration, maximum chlorophyllconcentration, salinity, pH, and temperature range yielded a model with the smallest Akaike information criterion (2649.8), and were used in the present and future distribution model. Data from the HadGEM2-ES climate model were used to generate the projections for the year 2100. The present distribution of *P. caribaeorum* shows that parts of the Brazilian coast, Caribbean Sea, and Florida are suitable regions for the species, as they are characterized by high salinity and pH and small temperature variation. An expansion of the species' distribution was forecast northward under mild climate scenarios, while a decrease of suitable areas was forecast in the south. In the climate scenario with the most intense changes, P. caribaeorum would lose one-half of its suitable habitats, including the northernmost and southernmost areas of its distribution. The Caribbean Sea and northeastern Brazil, as well as other places under the influence of coastal upwellings, may serve as potential havens for this species.

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40	1. Introduction
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42	Palythoa caribaeorum (Duchassaing & Michelotti, 1860) is a sessile colonial anthozoan
43	common along the west coast of the Atlantic Ocean (Kemp et al., 2006). This sea mat species
14	can be found from Florida in the US (Kemp et al., 2006) to Arvoredo Island in Brazil (Bouzon,
45	Brandini & Rocha, 2012), as well as around many oceanic islands in the Atlantic, such as
46	Bermuda (Lesser et al., 1990) and the São Pedro e São Paulo archipelago (Edwards & Lubbock,
<b>1</b> 7	1983). Its distribution is not limited to the western side of the Atlantic Ocean, however; with
48	records in Cape Verde islands (Reimer, Hirose & Wirtz, 2010). This broad geographic
19	distribution is possible because P. caribaeorum has a great tolerance to a wide range of
50	environmental conditions (Sebens, 1982). Although studies on <i>P. caribaeorum</i> larvae are
51	lacking, it is expected that its biology would be similar to its sister species, Palythoa turbeculosa
52	(Esper 1791), which has zoanthellae larvae with planktonic habits and a larval stage up to 170
53	days long (Polak et al., 2011; Ryland, 1997).
54	It is found on hard bottom environments, from the intertidal zone to depths up to 12
55	meters (Sebens, 1982), where it can be dominant (Monteiro et al., 2008). Colonies have a mat
56	format usually occupy large areas along the substratum (Sebens, 1982; Acosta, 2001; Silva et al.,
57	2015). The animal's color varies from yellow and brown to bronze, easily identifiable in the field
58	(Fig. 1a).
59	P. caribaeorum is a mixotrophic organism, hosting symbiotic dinoflagellates

P. caribaeorum is a mixotrophic organism, hosting symbiotic dinoflagellates
 (Symbiodinium) inside its tissues (Sebens, 1977; Suchanek & Green, 1981) and feeding on



61	planktonic organisms, which contributes to the energy flux from plankton to benthic
62	environments (Sorokin, 1991; Santana et al., 2015). Few studies have highlighted the importance
63	of <i>P. caribaeorum</i> in association with other invertebrates (Pérez, Vila-Nova & Santos, 2005),
64	reef dynamics (Silva et al., 2015), or the enrichment of the bottom fish community (Mendonça-
65	Neto et al., 2008). However, a study conducted in the Todos os Santos Bay in Brazil (Cruz et al.,
66	2015b) showed that the increased abundance of <i>Palythoa</i> spp. decreased habitat heterogeneity,
67	and consequently the richness of fish and benthic organisms, changing the local trophic structure
68	in that region.
69	P. caribaeorum is capable of overgrowing and thus adversely affecting the growth and
70	recruitment of several organisms (Suchanek & Green, 1981; Castro et al., 2012), such as hard
71	corals, hydrocorals, and other zoanthids, showing great efficiency competing for space. P.
72	caribaeorum also has a fast growth rate estimated to vary between 0.04-0.15 mm per day
73	(Bastidas &Bone, 1996; Silva et al., 2015). In general, zoanthids can dominate in areas where the
74	environmental factors make it difficult for the settlement of scleractinian corals (Fautin, 1988;
75	Cruz et al., 2015a), often becoming important as the main benthic component (Mendonça-Neto
76	et al., 2008) and nursery habitat for other invertebrates (Pérez, Vila-Nova & Santos, 2005).
77	Suchanek and Green (1981) and Sebens (1982) claimed that predation does not seem to be an
78	ecological element controlling the distribution and abundance of <i>P. caribaeorum</i> , probably due
79	to defense strategies related to nematocysts and palytoxin that are found in the animal's tissue
80	(Hines & Pawlik, 2011). However, Stampar et al. (2007) report a case of a Hawksbill turtle,
81	Eretmochelys imbricata, preying colonies of P. caribaeorumn southeastern Brazil.
82	Regarding its ecological role, investigating the current and future distribution of <i>P</i> .
83	caribaeorum will yield a prognosis of areas likely to experience changes in their benthic



84 communities because of the presence or absence of P. caribaeorum. Assessing how species and 85 biological systems are affected by climate changes as well as the generation of forecast models 86 are increasingly needed by decision makers and environmental managers (Gutt et al., 2012). 87 Species distribution models (SDM) are largely used in ecology, biogeography, and 88 conservation biology to help solve practical problems in these fields (Guisan & Thuiller, 2005; 89 Elith et al., 2011; Acosta et al., 2016). One of the main assumptions of SDM is that the 90 environment is in equilibrium with its species, whose distribution is determined by 91 environmental variability (Araújo & Pearson, 2005), excluding biological interactions, dispersion 92 barriers, and fast or small-scale environmental disturbances (De Marco, Diniz-Filho & Bini, 93 2008). Temporal and climatic variations also interfere in the modeling, and environmental data 94 must be gathered from the same period in which the species was recorded, making it difficult to 95 acquire large amounts of data for mobile organisms. In this scenario, P. caribaeorum is a good 96 candidate for the use of SDM. 97 The accelerating rates of environmental changes pose a challenge to most organisms 98 living in shallow waters throughout the world's oceans. At a global scale, between 1971 and 99 2010, the warming rate of the oceans' surface waters reached 0.13°C per decade (IPCC, 2014). 100 Since 1950, coastal regions with high evaporation rates are becoming saltier; and for the same 101 period, the pH of surface waters has decreased by 0.1 (IPCC, 2014). The forecast for the year 102 2050 shows a redistribution of marine organisms as a function of the increase in temperature. 103 This redistribution is expected to be toward mid-latitude zones, increasing species richness in 104 these areas while decreasing richness in the tropics due to extinction, migration, and reduction in 105 genetic diversity (Fields et al., 1993). However, it is still unknown for sure how the effects of



climate change, at the population level, will affect the environment at large ecological scales (Harley et al., 2006; Thomas et al., 2017).

This study aimed to investigate and model the distribution of *P. caribaeorum* along the western coast of the Atlantic Ocean, generating potential distribution maps for the present and the year 2100 based on forecast climate models. The maps were used to characterize the species' habitat and identify areas where suitability was retained, lost, or gained for *P. caribaeorum* over time and under different climate scenarios.

#### 2. Materials and Methods

#### 2.1. Study area

The study area comprised the western Atlantic Ocean (from -42.5° to 42.5° latitude), between 2 meters of altitude and 100 meters of depth (Fig. 1b), hence extending beyond the current known distribution of *P. caribaeorum*. The oceanographic features of this region are influenced by the South Equatorial Current, from the east, which is divided into two components: northward (North Brazil Current) and southward along the coast (Brazil Current). The North Brazil Current flows toward the northwest, passing by the Caribbean Sea, Yucatan Channel, and Gulf of Mexico before merging with the Antilles Current in the Florida Strait to form the Gulf Stream, which flows toward Europe. Due to the strong tropical component, the region between 20°S and 30°N is strongly influenced by waters heated in equatorial regions (Ekman, 1953). In contrast, regions closer to the boundaries of this study are influenced by both tropical and colder currents, such as the Slopewater Current from the Labrador Current in the north (Churchill & Berger, 1998) and the Malvinas Current in the south (Talley et al., 2011). The study region is



129	also influenced by large river discharges, mainly from the La Plata, Amazon, Orinoco,
130	Mississippi, and Saint Lawrence rivers (Fig. 1b).
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132	2.2. Biological data mining and filtering
133	An extensive literature review was performed to obtain the largest possible number of
134	studies with occurrence data for P. caribaeorum (Supporting information 1). Occurrence data
135	recorded by the authors of this study and from global databases, such as WORMS, OBIS, and
136	GBIF, were also used, totaling 327 occurrence points for the species. The three identified
137	synonyms (Reimer, 2015)—Palythoa caribaea Duerden, 1898, Palythoa caribbaea Goreau,
138	1959, and Palythoa caribdea Carballeira et al., 1998—were included.
139	Occurrence points found on land or at depths deeper than 100 meters were automatically
140	excluded, as were those localized in freshwater environments. Because sampled areas and
141	methodologies varied among study sites, only one occurrence per pixel (5 arcminutes), based on
142	the environmental variables dataset, was kept for analysis (Elith & Leathwick, 2009). This point
143	was placed in the middle of each pixel, disregarding the exact occurrence point. After that, to
144	avoid bias originating from densely sampled sites, the geographic filter "OccurrenceThinner"
145	version 1.04 (Verbruggen, 2012), with thresholds of 0.5 and 1, was used, resulting in the final
146	dataset.
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148	2.3. Selecting environmental variables, the RM value, and the running Maxent
149	Environmental and bathymetric data, with resolutions of 5 and 1 arcminute, were
150	obtained from the Bio-Oracle (Tyberghein et al., 2012) dataset and the ETOPO1 Global Relief
151	Model, respectively (Amante & Eakins, 2009). Twenty-seven layers of environmental data and



152	their transformations available on the Bio-Oracle dataset were used in the analyses (Table 1).
153	Each layer was cropped using ArcGIS 10.2 to only include areas within 2 meters of altitude and
154	100 meters of depth.
155	The habitat modeling was done using the Maxent 3.3.3 program (Phillips, Anderson &
156	Schapire, 2006). Maxent has the advantage of using only occurrence points, instead of
157	occurrence and absence/pseudo-absence, considering the lack of this type of data, especially for
158	marine organisms. Maxent also demonstrates good performance identifying suitable habitats
159	(Phillips, Anderson & Schapire, 2006; Reiss et al., 2011; Meißner et al., 2014) and predicting
160	novel distributions for species (Costa et al., 2012; Acosta et al., 2016) when compared with other
161	software.
162	All the features available in Maxent were used during the modeling (standard set up). To
163	choose the environmental variables and the regulation multiplier $(Rm)$ value, the
164	"MaxentVariableSelection" package (Jueterbock, 2016) for R 3.3.1 (R Core Team, 2015) was
165	used, with a contribution threshold of 4% and Pearson's correlation coefficient of 0.9. A series of
166	runs were conducted, varying the Rm from 0.5 to 10 with increments of 0.5, and were then
167	assessed using the Akaike Information Criterion (AIC). The environmental variables and Rm of
168	the model with a lower AIC were used to run the final Maxent model. The AIC tends to choose
169	models that sort the most important variables, as well as models that better estimate
170	environmental suitability (Warren et al., 2014), and therefore behave better when projected onto
171	novel climate scenarios than models chosen using the Area Under the Receiver Operator Curve
172	(AUC) value (Warren & Seifert, 2011). The AIC selects the optimal and most parsimonious
173	model, taking into account its complexity (Bozdogan, 1987). Thirty percent of the data were
174	randomly sampled to test the model along each of the 100 replicates, resulting in an average for



all replicates. In all the analyses, 10,000 random points were used as the background along the whole study area.

#### 2.4. Climate model data

Data from the HadGEM2-ES model (Collins et al., 2011) retrieved from the Coupled Model Intercomparison Project (CMIP5) were used on the projections for the year 2100 under three different Representative Concentration Pathway (RCP) climate scenarios: low, middle, and high radiative forcing of 2.6, 4.5, and 8.5 W/m², respectively (Van Vuuren et al., 2011). The RCPs are related to retention of energy by the Earth's surface due to accumulation of greenhouse gases from human activities, therefore higher numbers represent a greater anthropogenic impact on our biosphere. The HadGEM2-ES climate model was freely available and provided outputs of the main interest variables and high resolution in tropical areas under the RCP climate scenarios. When there were no available data for the projections, the layers were kept unchanged. NetCDF files from the climate model were interpolated using the bilinear interpolation "interp2" of Matlab 7.14, and were then cropped and converted to ASC format.

#### Projecting the results onto maps

Maps were generated using the cumulative likelihood of species occurrence obtained from Maxent, which can be interpreted as the sum of the relative occurrence rate of the species in one pixel and other pixels with the same or lower values of the relative occurrence rate (ROR). This type of output is best for drawing species distribution boundaries for different climate scenarios (Merow, Smith & Silander, 2013) and also has a better visual presentation. For the binary maps, the threshold equal to the maximum sensitivity plus specificity (max SSS) was



chosen to detect suitable habitats for *P. caribaeorum*. This threshold improves the distinction between occurrence and absence sites, as well as between occurrence and background, also selecting the same threshold for both occurrence-only and occurrence-absence models (Liu, Newell & White, 2015).

Because *P. caribaeorum* is distributed along coastal zones, the final results were cropped to depths up to 30 meters, better defining its real habitat. The map of the Marine Ecoregions of the World (Spalding et al., 2007) was used as a reference for the results and discussion of specific areas (Fig. 2).

#### 3. Results

After the filtering, 135 occurrence points of P. caribaeorum were kept for the analyses (Fig. 2). The model with a lower AIC (2649.8) presented an Rm of 3.5 and showed good performance defining occurrence and absence sites for P. caribaeorum (AUC =  $0.860 \pm 0.023$ ) (Swets, 1988). The difference between the test AUC and the train AUC yielded a value of 0.015, showing that the model was not over-adjusted and is transferable to novel climate scenarios (Jueterbock et al., 2016; Warren & Seifert, 2011).

The environmental variables that contributed the most to the model were salinity (47.9%), concentration of calcite (21%), maximum concentration of chlorophyll-a (13.3%), pH (9.2%), and temperature range (8.5%), and were therefore kept in the model (Table 1). The habitat of *P. caribaeorum* is characterized by salinity values higher than 34, a maximum chlorophyll-poncentration from 0 to 28 mg/m³, a pH higher than 7.85, and temperature variation up to 10°C, although these intervals become much narrower when variable correlations



are considered. For calcite, the model pointed to a modulating influence, as no minimum or 222 maximum values were indicated. 223 The output of the HadGEM2-ES climate model for salinity, pH, and temperature (used to 224 calculate temperature range) was used to project the SDM model to novel scenarios for the year 225 2100, while calcite and chlorophyll-a were kept constant. A max SSS equal to 25.97 was used as 226 a threshold to infer suitable areas, thus identifying the potential distribution of the species. 227 Five hundred forty eight square kilometers were characterized as currently suitable for P. caribaeorum (Fig. 3). Under the RCP 2.6 scenario (Fig. 4), the southern range of P. caribaeorum 228 229 distribution would lose suitability, as would the Caribbean, Bermudan, and Floridian regions; 230 however, the total suitable area is close to the one observed for the present: 492,000 km<sup>2</sup>. An 231 increase in suitable areas in the Guianan, Amazonian, Bahamian, and Caribbean regions was also 232 observed, but the biggest change occurred in the northern part of its distribution, in the northern 233 Gulf of Mexico and along the Carolinian and Virginian regions. In these areas, the potential 234 distribution of the species reached almost as far north as Chesapeake Bay. The southernmost area 235 capable of retaining suitability for the species was near Cape Frio, in southeastern Brazil. 236 Not much differently, the projection under RCP 4.5 (Fig. 5) also shows that part of the 237 Brazilian coast would lose habitat suitability, only remaining suitable around the Abrolhos Bank 238 and in northeastern Brazil. In the north, the species' potential distribution is similar to the 2.6 239 scenario, only presenting smaller suitable areas in those regions, 470,000 km<sup>2</sup> in total. 240 Under the scenario with the higher anthropogenic impact, RCP 8.5 (Fig. 6), almost the entire potential distribution for P. caribaeorum along the Brazilian coast would be lost, including 241 242 the northeast region. Under RCP 8.5, the anthozoan would also lose suitable habitats in regions 243 around Florida, the northern Gulf of Mexico, and the eastern and western Caribbean.



Overall, for this climate scenario, only one-half of the suitable areas seen in the present, around 275,000 km², was characterized as a potential habitat for the species, comprising the southern, western, and southwestern Caribbean, the southern Gulf of Mexico, and the Bahamian, Greater Antilles, and eastern Brazil regions. Nevertheless, under the RCP 8.5 scenario, there would be an increase in suitable areas in the southwestern Caribbean, along Nicaragua's coast, similar to the RCP 4.5 forecast (Fig. 5).

To demonstrate the most important regions for the species, a map of the intersection of all suitable areas under any climate scenarios for the year 2100 was made (Fig. 7). Figure 7 shows that the northern part of the map (from 0° northward) would have much more significance for the potential distribution of *P. caribaeorum* than the southern part. The suitable areas comprise six ecoregions in the north (southern, southwestern, and western Caribbean, southern Gulf of Mexico, and Bahamian and Greater Antilles regions) against only two in the south (northeastern and eastern Brazil).

#### 4. Discussion

#### 4.1. Environmental variables and P. caribaeorum's habitat

The variables selected to model *P. caribaeorum's* habitat suitability are in line with those used in studies about environmental controls for other anthozoans. For instance, temperature and salinity are two of the most important factors shaping the growth of modern reefs (Wood, 1993). Temperature is also related to coral (Hoegh-Guldberg, 1999; Hughes et al., 2003; Villamizar et al., 2008) and *P. caribaeorum* bleaching (Kemp et al., 2006). Variation in salinity may induce osmotic stress and can alter the zonation of zoanthids, including *P. caribaeorum* (Sebens, 1982). Changes in chlorophyll-a concentrations can also play a role in the distribution of *P*.



caribaeorum due to its heterotrophic nutrition and dependency on the water column productivity, unlike most coral reef species dependent solely on autotrophy and therefore, clear oligotrophic waters. This species also seems to be affected by lower pH (Kroeker et al., 2013), which is another consequence of global environmental changes (Okazaki et al., 2017).

When the projection of the habitat suitability of *P. caribaeorum* for the present (Fig. 3) is compared with the map of the biggest river discharges in the study area (Fig. 1), it becomes evident that salinity has a major influence on the distribution of this species. Salinity had a higher contribution to the model, sometimes reducing habitat suitability even when the values of other variables, like temperature and chlorophyll-a, were adequate for the species. Freshwater discharges from rivers are also an important source of calcite for oceans; however, calcite is 100 times more soluble in seawater than in freshwater (Libes, 2009), so high concentrations of this mineral are not limited to river-influenced regions. In fact, regarding only calcite, habitat suitability is positively correlated with its concentration, possibly because calcite concentration also indicates other important biogeochemical controls.

The maximum concentration of chlorophyll-a presented a distribution pattern similar to calcite and was also strongly affected by coastal processes, mainly big river discharges, which fertilize the sea and enhance primary production. From satellite data, chlorophyll-a cannot be distinguished from dissolved organic matter (DOM) without an algorithm specific to each region (Ali et al., 2016; Liew et al., 2001; Vazyulya et al., 2014). Away from river discharges, chlorophyll-a can be re-suspended back into the water column by wind, currents, and waves, and is therefore correlated with suspended sediment (Walsh, Dieterle & Esaias, 1987). The values of chlorophyll-a data used in this study are relative to its maximum concentration over several years. Hence, it is likely that these values are the sum of all processes listed above: river



discharge, new production, and re-suspension. *P. caribaeorum* inhabits coastal regions with high chlorophyll-a concentrations, but when these concentrations are related to river discharge, the regions lose habitat suitability due to lower salinity and pH values. This can indicate that despite the low salinity, chlorophyll-a and turbidity can be important factors for the species' niche. This information corroborates previous studies (Segal & Castro, 2011; Steiner et al., 2015), which suggested that low salinity was an exclusion factor for *P. caribaeorum*. However, it is difficult to decouple the influence of variables related to water quality from demands related more to depth and substrate in coastal areas, as both are intrinsically linked in continental shores. For instance, *P. caribaeorum* is quite abundant around remote oceanic islands where eutrophication, turbidity, and other processes typical of continental shores are not present. However, *P. caribaeorum* seems to be more common in regions influenced by coastal processes with high turbidity along the Brazilian coast (Cruz, personal observation).

Similar to calcite and chlorophyll-a, pH has a strong coastal component related to salinity, especially around the Amazonian river mouth, where pH drops below 7, and in the Virginian and Carolinian regions, where values exceed 8.25—both regions are devoid of the anthozoan. Regarding pH alone, a habitat with a pH over 7.85 is considered suitable, but due to correlations with the other variables, the habitat of *P. caribaeorum* is constrained to a pH between 8.05 and 8.25.

Lastly, temperature range seems to influence the species in a typical latitudinal fashion, presenting higher values in the northernmost and southernmost regions of the study area due to the intrusion of colder waters onto the continental shelf at higher latitudes. Although *P. caribaeorum* exhibits high tolerance to changes in temperature and water properties in tide pool environments (Bastidas & Bone, 1996) and in southeastern Brazil (Bouzon, Brandini & Rocha,



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2012), extreme variations in temperature and decreases in salinity caused by rain can cause mortality (Sebens, 1977). For current environmental conditions, the latitudes 30° North and 30° South can be recognized as limits to the distribution of this organism, mainly due to the temperature range: between 9°C and 13°C in the south and 13°C and 24°C in the north. In these regions, salinity reaches a value of 34, with high seasonal variation (Piola et al., 2000; Rasmussen, Gawarkiewicz & Owens, 2005). 4.2. Present distribution of P. caribaeorum There are regions characterized as potential habitats for P. caribaeorum that have no record of the species (Fig. 2 and 3). This indicates either a lack of adequate sampling effort in the Guianan, Greater Antilles, Bahamian, and western and southwestern Caribbean regions, the absence of proper substrata, or other conditions not considered in the present modeling. In general, the region between Florida and the Yucatan Peninsula, with the exception of Flower Garden Banks off the coast of Texas (Tester et al., 2013), is characterized by the absence of coral reefs due to low temperatures during winter and high rainfall (Longhurst, 2007). In shallow areas in the northern Gulf of Mexico previously dominated by corals and sponges, changes in temperature and an increase in river discharge were followed by a community shift around 2005, and these areas are now dominated by algae (DeBose et al., 2013). This information corroborates our model predictions for *P. caribaeorum* in that region, strongly influenced by the Mississippi River Delta. Using Briggs and Bowen's (2012) classification, it is evident that all of *P. caribaeorum's* current potential habitat is within warm marine provinces, characterizing the species' distribution

as tropical. However, the limits of its distribution are close to warm-temperate and cold



provinces, which are characterized as marginal parts of the distribution. Studying reef communities located at higher latitudes can improve our knowledge of how this species will be affected during environmental and climatic changes (Thomas et al., 2017); thus, examining *P. caribaeorum's* distribution in these regions is extremely important.

4.3 Palythoa caribaeorum's response to climate change in the year 2100

According to the RCP 2.6 climate scenario for the year 2100, there would be an increase in salinity of 2.5 in the northern Gulf of Mexico, Virginian, and Carolinian regions, in turn increasing habitat suitability for *P. caribaeorum* (Fig. 4). Along the southeastern Brazilian coast, salinity would drop by 0.3, sufficient to become an unsuitable area for the species. Changes would also be seen in the plume of the Amazon river, influencing adjacent coastal regions to a lesser degree.

Values of pH in the year 2100 would drop for all regions, which is one of the factors that could exclude *P. caribaeorum*. Since the anthozoan response to changes in pH is related to salinity, its northernmost distribution should increase due to a weaker influence of rivers and a decline in the temperature range under the RCP 2.6 scenario. On the other hand, decreases in salinity and pH and more intense temperature fluctuations in the southeast and east of Brazil would result in a loss of suitable areas, except along the coast of Cape Frio in southeastern Brazil. This result may be associated with upwelling in this region (Castro et al., 2006), which modifies water characteristics and reduces the frequency of osmotic and thermal stress on the species (Hu & Guillemin, 2016), making that region more suitable for the animal. The same process seems to occur along the coast of Colombia, as seen for coral species in that region



(Chollet, Mumby & Cortés, 2010). The Colombian coast is very important for *P. caribaeorum* since it would remain a suitable habitat under any climate scenario (Fig. 7).

The same trend observed under the RCP 2.6 scenario occurs in the RCP 4.5 scenario, with an increase in potential habitat in the northernmost range of its distribution (Fig. 5).

However, for this scenario, *P. caribaeorum* would also lose habitat in the south, including around Cape Frio. Differently, other areas, such as the continental shelf of the Yucatan Peninsula, would gain suitable zones in the RCP 4.5 scenario, due to higher salinity values and lower temperature variations. For regions experiencing a loss of suitable areas, such as the northeastern Brazil and Amazonian regions, the drop in pH values, salinity, and an increase in the temperature range would be responsible.

As expected, the climate scenario with the strongest anthropogenic impact, RCP 8.5, presented the smallest suitable area for the species (Fig. 6). Although it presented smaller temperature fluctuations compared to other scenarios, in the RCP 8.5 scenario *P. caribaeorum* would lose potential habitats in both the northernmost and southernmost ranges of its distribution, without any gain in suitable areas due to a drastic decrease in salinity and pH values. Freshwater intrusions from the La Plata and Amazon rivers would influence the species' habitat along the Brazilian shelf, restricting its distribution to northeastern Brazil. Similarly, the Mississippi River plume would decrease the adequacy for the species in the north. In general, when looking at changes in salinity, pH, and temperature for any climate scenario, the variations are more intense in the north and south via the intrusion of polar and temperate currents into the continental shelf. These differences alter the habitat of *P. caribaeorum* and are more prominent in the Southern Hemisphere.



Regions such as the southern, southwestern, and western Caribbean, the southern Gulf of Mexico, the Bahamas, the Greater Antilles, and northeastern and eastern Brazil would retain suitable areas for the species under any climate scenario (Fig. 7), identifying them as refuges for the species under adverse environmental conditions (Haffer, 1969). Except for the Brazilian coast, all of these regions are within the Caribbean Province, which is considered the center of biodivesity, in the Atlantic Ocean (Rocha et al., 2008; Briggs & Bowen, 2013). Under intense climate stress, it is likely that both *P. caribaeorum* and other invertebrate organisms will have their distributions restricted toward this diversity-gathering hotspot in the Atlantic Ocean.

The difference in suitable areas between the Northern and Southern Hemispheres suggests asymmetrical species distribution, with the Northern Hemisphere containing larger suitable habitats. This may result from differences in the oceanographic conditions of both basins. In contrast to other oceans, the Atlantic Ocean presents a net flux of superficial water transport to the north, with the Labrador Current being much weaker than the Malvinas Current; therefore, its principal component is toward the north (Onken, 1994). This occurs mainly because of thermohaline circulation and wind, both of which are likely to be altered by climate change (IPCC, 2014), in turn helping to remodel the distribution of marine organisms, as predicted in this study.

#### 4.4 Likely changes in reef communities

Studies have shown that *P. caribaeorum* coexists with many other organisms, especially algae such as *Caulerpa racemosa* var. *peltata*, *Bryopsis* spp. (Magalhães et al., 2015), *Sargassum*, *Halimeda*, *Laurencia*, *Dictyota*, *Dictyopteris*, *Ulva*, and coralline algae (Steiner et al., 2015). Macroalgae are common in coral reefs, sometimes dominating the community and



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competing for space with corals (Tanner, 1995; Bruno et al., 2009). The disappearance of P. caribaeorum from many Brazilian reefs in the Abrolhos Bank would imply a change in the reef communities. It is possible that with the vanishing of P. caribaeorum, colonization of the available space would first be by algae, most likely cyanobacteria, due to their fast growth rate, especially in nutrient-poor environments (Villaça & Pitombo, 1997). However, the composition of the new community will depend on multiple factors, from changes in recruitment rates, growth, and mortality to changes in nutrient concentrations and herbivory (Littler, Littler & Taylor, 1983; Sandin & McNamara, 2012); community resilience is another factor (Roff & Mumby, 2012). Regarding the colonization of new areas by P. caribaeorum, the temperate region of the northwestern Atlantic would see its suitable area increased under RCP 2.6 and 4.5, around Cape Hatteras and Chesapeake Bay. Zoanthids have the potential to disperse to great distances due to their long larval stage (Ryland et al., 2000; Polak et al., 2011) and asexual reproduction (Acosta, Sammarco & Duarte, 2005), alongside their potential to raft on floating objects and boats (Santos et al., 2016). Thus, it is possible that the colonization of *P. caribaeorum* in the northwestern Atlantic would occur under these climate scenarios for hard bottom environments. Nicaragua's coast, which is part of the southwestern Caribbean region, was particularly important for the species due to the increase in potential habitat area under RCP 4.5 and 8.5 (Fig. 5 and 6). Few studies were conducted about the benthic community in that region, which presents high coverage of algae of the genera Dictyota and Padina, corals such as Orbicella annularis, Agaricia spp., Porites spp., Acropora palmata, and Millepora alcicornis, and lower coverage of sponge and soft corals (Kramer et al., 2000). The same study discusses the likelihood of a phase shift in the coming years due to natural disturbances and climate change.



Phase shifts from reef-building corals to zoanthids were previously reported in the western Atlantic (Cruz et al., 2015a), where they reduced fish diversity associated with the reefs (Cruz et al., 2015b), and have the potential to adversely affect regional fisheries and tourism, which are main income sources for the local populatio.

Although it was assumed in this work, it is not expected that all individuals will be affected in the same way by climate change. It is known that, at smaller scales, differences between symbiont algae (Kemp et al., 2006), as well as the anthozoan's plasticity (Costa et al., 2011), can result in distinct outcomes for different populations under climate change. Average and maximum temperatures had small contributions to the model and were not considered here; however, it is known that surface waters are likely to experience an increase in temperature, especially in tropical areas (IPCC, 2014). The ability to endure such changes will also play a role in *P. caribaeorum's* distribution. The authors highlight here the importance of future experimental work on temperature stress on key organisms, such as *P. caribaeorum*, to better understand the future of tropical reef communities.

It is possible that, due to the scale of the problem, some local variations were not accounted for, such as the discharge of small rivers around the Gulf of Mexico region. Smaller scales and regional studies should use higher resolution data for this matter. Sea level rise can modify wave stress atop coral reefs, favoring the colonization of *P. caribaeorum* over other organisms, including sea mat eorals (Rabelo et al., 2015). This effect was not considered in this study because the IPCC forecast for sea level rise did not surpass 0.8 meters, which is smaller than the error of the bathymetric data used in this work.

#### 5. Conclusions



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The present habitat of P. caribaeorum can be characterized by high salinity values, marine pH, and a small to moderate range of temperature in coastal and shallow environments. The species' current distribution in the western Atlantic Ocean is restricted to warm provinces, identifying it as a coastal and warm-water species, together with shallow water corals. According to this study, this organism's potential distribution is likely to change by the year 2100. It is possible that, under low to mild anthropogenic impacts (RCP 2.6 and 4.5), the species' distribution will be extended northward and will face a decrease of suitable habitat in the southernmost range of its distribution under any climate scenario. Under the scenario with the highest anthropogenic impacts (RCP 8.5), novel suitable areas poleward were not predicted, and only one-half of its potential habitat area would be retained when compared to the present. Under any climate scenario, the Caribbean Sea remained suitable for the species, characterizing this region as a possible refuge. Upwelling regions, such as Cape Frio and northern Colombia, could provide a refuge for this species as well, diminishing the local impacts of climate changes. Although community changes were discussed according to the available literature, biological and oceanographic processes not considered in this study may play an important role in the outcome of these changes.

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#### Acknowledgments

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We thank our colleagues Paulo Polito, André Acosta, Romina Barbosa, Bruno Ferrero, and Alexander Jueterbock, each of whom provided advice and expertise that contributed to this



.73	work. We would also like to thank Ana Flora Sarti for kindly providing part of the data used in
74	this work, as well as Erica Donlon for the English review. This work is also a product of the
75	SISBIOTA-Mar research network.
76	
77	References
78	Acosta A. 2001. Disease in Zoanthids: Dynamics in space and time. <i>Hydrobiologia</i> 460:113–
79	130. DOI: 10.1023/A:1013135702430.
80	Acosta AL., Giannini TC., Imperatriz-Fonseca VL., Saraiva AM. 2016. Worldwide Alien
81	Invasion: A Methodological Approach to Forecast the Potential Spread of a Highly Invasive
82	Pollinator. Plos One 11:e0148295. DOI: 10.1371/journal.pone.0148295.
83	Acosta A., Sammarco PW., Duarte LF. 2005. New Fission Processes in the Zoanthid Palythoa
84	caribaeorum: Description and Quantitative Aspects. Bulletin of Marine Science 76:1–26.
85	Ali KA., Ortiz J., Bonini N., Shuman M., Sydow C. 2016. Application of Aqua MODIS sensor
86	data for estimating chlorophyll a in the turbid Case 2 waters of Lake Erie using bio-optical
87	models. GIScience & Remote Sensing 53:483–505. DOI: 10.1080/15481603.2016.1177248.
88	Amante C., Eakins BW. 2009. ETOPO1 1 Arc-Minute Global Relief Model: Procedures, Data
89	Sources and Analysis. NOAA Technical Memorandum NESDIS NGDC-24:19. DOI:
90	10.1594/PANGAEA.769615.
91	Araújo MB., Pearson RG. 2005. Equilibrium of species' distribution with climate. <i>Ecography</i>
92	28:693–695.
.93	Bastidas C., Bone D. 1996. Competitive Strategies Between Palythoa Caribaeorum and Zoanthus
94	Sociatus (Cnidaria: Anthozoa) at a Reef Flat Environment in Venezuela. Bulletin of Marine
95	Science 59:543–555.



96	Bouzon JL., Brandini FP., Rocha RM. 2012. Biodiversity of Sessile Fauna on Rocky Shores of
97	Coastal Islands in Santa Catarina, Southern Brazil. Marine Science 2:39-47. DOI:
98	10.5923/j.ms.20120205.01.
.99	Bozdogan H. 1987. Model selection and Akaike's information criterion (AIC): the general theory
00	and its analytical extensions. 52:345–370.
01	Briggs JC., Bowen BW. 2012. A realignment of marine biogeographic provinces with particular
02	reference to fish distributions. <i>Journal of Biogeography</i> 39:12–30. DOI: 10.1111/j.1365-
03	2699.2011.02613.x.
04	Briggs JC., Bowen BW. 2013. Marine shelf habitat: biogeography and evolution. <i>Journal of</i>
05	Biogeography 40:1023–1035. DOI: 10.1111/jbi.12082.
06	Bruno JF., Sweatman H., Precht WF., Selig ER., Schutte VGW. 2009. Assessing evidence of
07	phase shifts from coral to macroalgal dominance on coral reefs. <i>Ecology</i> 90:1478–1484.
80	DOI: 10.1890/08-1781.1.
09	Castro BM., Brandini FP., Pires-Vanin AMS., Miranda LB. 2006. Multidisciplinary
10	Oceanography Processes on the Western Atlantic Continental Shelf Between 4°N and 34°S.
11	In: Robinson AR, Brink KH eds. The Sea: The Global Coastal Ocean. Harvard University,
12	259–295.
13	Castro CB., Segal B., Negrão F., Calderon EN. 2012. Four-year monthly sediment deposition on
14	turbid southwestern Atlantic coral reefs, with a comparison of benthic assemblages.
15	Brazilian Journal of Oceanography 60:49–63. DOI: 10.1590/S1679-87592012000100006.
16	Chollet I., Mumby PJ., Cortés J. 2010. Upwelling areas do not guarantee refuge for coral reefs in
17	a warming ocean. Marine Ecology Progress Series 416:47–56. DOI: 10.3354/meps08775.
18	Churchill JH., Berger TJ. 1998. Transport of Middle Atlantic Bight shelf water to the Gulf



519	Stream near Cape Hatteras. <i>Journal of Geophysical Research</i> 103:30605–30621. DOI:
520	10.1029/98JC01628.
521	Collins WJ., Bellouin N., Doutriaux-Boucher M., Gedney N., Halloran P., Hinton T., Hughes J.,
522	Jones CD., Joshi M., Liddicoat S., Martin G., O 'connor F., Rae J., Senior C., Sitch S.,
523	Totterdell I., Wiltshire A., Woodward S. 2011. Geoscientific Model Development
524	Development and evaluation of an Earth-System model – HadGEM2. Geosci. Model Dev
525	4:1051–1075. DOI: 10.5194/gmd-4-1051-2011.
526	Costa DL., Gomes PB., Santos AM., Valenca NS., Vieira N a., Perez CD. 2011. Morphological
527	plasticity in the reef zoanthid Palythoa caribaeorum as an adaptive strategy. Annales
528	Zoologici Fennici 48:349–358. DOI: 10.5735/086.048.0602.
529	Costa BM., Kendall MS., Parrish FA., Boland RC., Lecky J., Spalding H. 2012. Prediction of
530	Mesophotic Coral Distributions in the Au'au Channel, Hawaii.
531	Cruz ICS., de Kikuchi RKP., Longo LL., Creed JC. 2015a. Evidence of a phase shift to
532	Epizoanthus gabrieli Carlgreen, 1951 (Order Zoanthidea) and loss of coral cover on reefs in
533	the Southwest Atlantic. Marine Ecology 36:318–325. DOI: 10.1111/maec.12141.
534	Cruz ICS., Loiola M., Albuquerque T., Reis R., de Anchieta C. C. Nunes J., Reimer JD.,
535	Mizuyama M., Kikuchi RKP., Creed JC. 2015b. Effect of Phase Shift from Corals to
536	Zoantharia on Reef Fish Assemblages. Plos One 10:e0116944. DOI:
537	10.1371/journal.pone.0116944.
538	DeBose JL., Nuttall MF., Hickerson EL., Schmahl GP. 2013. A high-latitude coral community
539	with an uncertain future: Stetson Bank, northwestern Gulf of Mexico. Coral Reefs 32:255-
540	267. DOI: 10.1007/s00338-012-0971-3.
541	Edwards A., Lubbock R. 1983. The ecology of Saint Paul's Rocks (Equatorial Atlantic). Journal



542	of Zoology 200:51–69. DOI: 10.1111/j.1469-7998.1983.tb06108.x.
543	Ekman S. 1953. Zoogeography of the sea. London: Sidgwick & Jackson.
544	Elith J., Leathwick JR. 2009. Species Distribution Models: Ecological Explanation and
545	Prediction Across Space and Time. Annu Rev Ecol Syst 40:415–436. DOI:
546	10.1146/annurev.ecolsys.l.
547	Elith J., Phillips SJ., Hastie T., Dudík M., Chee YE., Yates CJ. 2011. A statistical explanation of
548	MaxEnt for ecologists. Diversity and Distributions 17:43-57. DOI: 10.1111/j.1472-
549	4642.2010.00725.x.
550	Fautin DG. 1988. Anthozoan dominated benthic environments. In: Proceedings of the 6th
551	International Coral Reef Symposium, Australia. 1–6.
552	Fields PA., Graham JB., Rosenblatt RH., Somero GN. 1993. Effects of expected global climate
553	change on marine faunas. Trends in ecology & evolution (Personal edition) 8:361–367.
554	DOI: 10.1016/0169-5347(93)90220-J.
555	Guisan A., Thuiller W. 2005. Predicting species distribution: Offering more than simple habitat
556	models. <i>Ecology Letters</i> 8:993–1009. DOI: 10.1111/j.1461-0248.2005.00792.x.
557	Gutt J., Zurell D., Bracegridle TJ., Cheung W., Clark MS., Convey P., Danis B., David B., De
558	Broyer C., di Prisco G., Griffiths H., Laffont R., Peck LS., Pierrat B., Riddle MJ., Saucède
559	T., Turner J., Verde C., Wang Z., Grimm V. 2012. Correlative and dynamic species
560	distribution modelling for ecological predictions in the Antarctic: A cross-disciplinary
561	concept. Polar Research 31:1–23. DOI: 10.3402/polar.v31i0.11091.
562	Haffer J. 1969. Speciation in amazonian forest birds. <i>Science (New York, N.Y.)</i> 165:131–7. DOI:
563	10.1126/science.165.3889.131.

Harley CDG., Hughes AR., Kristin M., Miner BG., Sorte CJB., Carol S., Randall Hughes a.,



363	Hultgren KM., Thornber CS., Rodriguez LF., Tomanek L., Williams SL. 2006. The impacts
566	of climate change in coastal marine systems. <i>Ecology letters</i> 9:228–41. DOI:
567	10.1111/j.1461-0248.2005.00871.x.
568	Hines DE., Pawlik JR. 2011. Assessing the antipredatory defensive strategies of Caribbean non-
569	scleractinian zoantharians (Cnidaria): is the sting the only thing? Marine Biology 159:389–
570	398. DOI: 10.1007/s00227-011-1816-2.
571	Hoegh-Guldberg O. 1999. Climate change, coral bleaching and the future of the world's coral
572	reefs. Marine Freshwater Research 50:839–866. DOI: 10.1071/MF99078.
573	Hu Z-M., Guillemin M-L. 2016. Coastal upwelling areas as safe havens during climate warming
574	Journal of Biogeography: 1–2. DOI: 10.1111/jbi.12887.
575	Hughes TP., Baird AH., Bellwood DR., Card M., Connolly SR., Folke C., Grosberg R., Hoegh-
576	Guldberg O., Jackson JBC., Kleypas J., Lough JM., Marshall P., Nystrom M., Palumbi SR.
577	Pandolfi JM., Rosen B., Roughgarden J. 2003. Climate Change, Human Impacts, and the
578	Resilience of Coral Reefs. <i>Science</i> 301:929–934. DOI: 10.1126/science.1085046.
579	IPCC. 2014. Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and
580	III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change.
581	Jueterbock A. 2016. Package "MaxentVariableSelection": Selecting the Best Set of Relevant
582	Environmental Variables along with the Optimal Regularization Multiplier for Maxent
583	Niche Modeling.
584	Kemp DW., Cook CB., LaJeunesse TC., Brooks WR. 2006. A comparison of the thermal
585	bleaching responses of the zoanthid Palythoa caribaeorum from three geographically
586	different regions in south Florida. Journal of Experimental Marine Biology and Ecology
587	335:266–276. DOI: 10.1016/j.jembe.2006.03.017.



588	Kramer P., Kramer PR., Arias-Gonzalez E., McField M. 2000. Status of coral reefs of Northern
589	Central America: Mexico, Belize, Guatemala, Honduras, Nicaragua and El Salvador.
590	Kroeker KJ., Kordas RL., Crim R., Hendriks IE., Ramajo L., Singh GS., Duarte CM., Gattuso
591	JP. 2013. Impacts of ocean acidification on marine organisms: Quantifying sensitivities and
592	interaction with warming. Global Change Biology 19:1884–1896. DOI: 10.1111/gcb.12179
593	Lesser MP., Stochaj WR., Tapley DW., Shick JM. 1990. Bleaching in coral reef anthozoans:
594	effects of irradiance, ultraviolet radiation, and temperature on the activities of protective
595	enzymes against active oxygen. Coral Reefs 8:225–232. DOI: 10.1007/BF00265015.
596	Libes SM. 2009. Calcite, Alkalinity, and the pH of Seawater. In: Introduction to Marine
597	Biogeochemistry. Academic Press, 909.
598	Liew SC., Chia AS., Kwoh LK. 2001. Evaluating the validity of SeaWiFS chlorophyll algorithm
599	for coastal waters. In: Proc. 22nd Asian Conference on Remote Sensing. 1–5.
600	Littler M., Littler D., Taylor P. 1983. Evolutionary stragegies in a tropical barrier reef system:
601	functional-form groups of marine macroalgae. Deep Sea Research Part B. Oceanographic
602	Literature Review 30:935. DOI: 10.1016/0198-0254(83)96598-6.
603	Liu C., Newell G., White M. 2015. On the selection of thresholds for predicting species
604	occurrence with presence-only data. Ecology and Evolution:337–348. DOI:
605	10.1002/ece3.1878.
606	Longhurst AR. 2007. Ecological Gepgraphy of the Sea. Academic Press.
607	Magalhães GM., Amado-Filho GM., Rosa MR., de Moura RL., Brasileiro PS., De Moraes FC.,
608	Francini-Filho RB., Pereira-Filho GH. 2015. Changes in benthic communities along a 0-60
609	m depth gradient in the remote St. Peter and St. Paul Archipelago (Mid-Atlantic Ridge,
610	Brazil). Bulletin of Marine Science 91:377–396. DOI: 10.5343/bms.2014.1044.



611	De Marco P., Diniz-Filho JA., Bini LM. 2008. Spatial analysis improves species distribution
612	modelling during range expansion. Biol Lett 4:577–580. DOI: 10.1098/rsbl.2008.0210.
613	Meißner K., Fiorentino D., Schnurr S., Martinez Arbizu P., Huettmann F., Holst S., Brix S.,
614	Svavarsson J. 2014. Distribution of benthic marine invertebrates at northern latitudes - An
615	evaluation applying multi-algorithm species distribution models. Journal of Sea Research
616	85:241–254. DOI: 10.1016/j.seares.2013.05.007.
617	Mendonça-Neto JP., Ferreira CEL., Chaves LCT., Pereira RC. 2008. Influence of Palythoa
618	caribaeorum (Anthozoa, Cnidaria) zonation on site-attached reef fishes. Anais da Academic
619	Brasileira de Ciencias 80:494–513. DOI: 10.1590/S0001-37652008000300010.
620	Merow C., Smith MJ., Silander JA. 2013. A practical guide to MaxEnt for modeling species'
621	distributions: what it does, and why inputs and settings matter. <i>Ecography</i> 36:1058–1069.
622	DOI: 10.1111/j.1600-0587.2013.07872.x.
623	Monteiro J., Almeida C., Freitas R., Delgado A., Porteiro F., Santos R. 2008. Coral assemblages
624	of Cabo Verde: preliminary assessment and description. Proceedings of the 11th
625	International Coral Reef Symposium-2008:1416–1419.
626	Okazaki RR., Towle EK., van Hooidonk R., Mor C., Winter RN., Piggot AM., Cunning R.,
627	Baker AC., Klaus JS., Swart PK., Langdon C. 2017. Species-specific responses to climate
628	change and community composition determine future calcification rates of Florida Keys
629	reefs. Global Change Biology 23:1023–1035. DOI: 10.1111/gcb.13481.
630	Onken R. 1994. The Asymmetry of Western Boundary Currents in the Upper Atlantic Ocean.
631	Journal of Physical Oceanography 24:928–948. DOI: 10.1175/1520-
632	0485(1994)024<0928:TAOWBC>2.0.CO;2.
633	Pérez CD., Vila-Nova D a., Santos a. M. 2005. Associated community with the zoanthid



634	Palythoa caribaeorum (Duchassaing & Michelotti, 1860) (Chidaria, Anthozoa) from littoral
635	of Pernambuco, Brazil. <i>Hydrobiologia</i> 548:207–215. DOI: 10.1007/s10750-005-5441-2.
636	Phillips SJ., Anderson RP., Schapire RE. 2006. Maximum entropy modeling of species
637	geographic distributions. Ecological Modelling 190:231–259. DOI:
638	10.1016/j.ecolmodel.2005.03.026.
639	Piola AR., Campos EJD., MÖller Jr OO., Charo M., Martinez C. 2000. Subtropical Shelf Front
640	off eastern South America. Jornal of Geouphysical Research 105:6565-6578.
641	Polak O., Loya Y., Brickner I., Kramarski-Winter E., Benayahu Y. 2011. The widely-distributed
642	Indo-Pacific Zoanthid Palythoa tuberculosa: a sexually conservative strategist. Bulletin of
643	marine science 87:605–621. DOI: 10.5343/bms.2010.1088.
644	Rabelo EF., Soares M de O., Bezerra LEA., Matthews-Cascon H. 2015. Distribution pattern of
645	zoanthids (Cnidaria: Zoantharia) on a tropical reef. Marine Biology Research 11:584–592.
646	DOI: 10.1080/17451000.2014.962542.
647	Rasmussen LL., Gawarkiewicz G., Owens WB. 2005. Slope water, Gulf Stream, and seasonal
648	influences on southern Mid-Atlantic Bight circulation during the fall-winter transition.
649	Journal of Geophysical Research 110:1–16. DOI: 10.1029/2004JC002311.
650	Reimer J. 2015. Palythoa caribaeorum (Duchassaing & Michelotti, 1860)
651	Reimer JD., Hirose M., Wirtz P. 2010. Zoanthids of the Cape Verde Islands and their symbionts:
652	previously unexamined diversity in the Northeastern Atlantic. Contributions to Zoology
653	79:147–163.
654	Reiss H., Cunze S., König K., Neumann H., Kröncke I. 2011. Species distribution modelling of
655	marine benthos: a North Sea case study. Marine Ecology Progress Series 442:71–86. DOI:
656	10.3354/meps09391.



657	Rocha LA., Rocha CR., Robertson DR., Bowen BW.2008. Comparative phylogeography of
658	Atlantic reef fishes indicates both origin and accumulation of diversity in the Caribbean.
659	BMC Evolutionary Biology 8:157. DOI: 10.1186/1471-2148-8-157.
660	Roff G., Mumby PJ. 2012. Global disparity in the resilience of coral reefs. <i>Trends in Ecology</i>
661	and Evolution 27:404–413. DOI: 10.1016/j.tree.2012.04.007.
662	Ryland JS. 1997. Reproduction in Zoanthidea (Anthozoa: Hexacorallia). <i>Invertebrate</i>
663	Reproduction & Development 31:177–188. DOI: 10.1080/07924259.1997.9672575.
664	Ryland JS., De Putron S., Scheltema RS., Chimonides PJ., Zhadan DG. 2000. Semper's
665	(zoanthid) larvae: pelagic life, parentage and other problems. <i>Hydrobiologia</i> 440:191–198.
666	Sandin SA., McNamara DE. 2012. Spatial dynamics of benthic competition on coral reefs.
667	Oecologia 168:1079–90. DOI: 10.1007/s00442-011-2156-0.
668	Santana EFC., Alves AL., Santos ADM., Cunha MDGGS., Perez CD., Gomes APB. 2015.
669	Trophic ecology of the zoanthid Palythoa caribaeorum (Cnidaria: Anthozoa) on tropical
670	reefs. Marine Biological Association of the United Kingdom 95:301–309. DOI:
671	10.1017/S0025315414001726.
672	Santos MEA., Kitahara MV., Lindner A., Davis Reimer J. 2016. Overview of the order
673	Zoantharia (Cnidaria: Anthozoa) in Brazil. Marine Biodiversity 46:547–559. DOI:
674	10.1007/s12526-015-0396-7.
675	Sebens K. 1977. Autotrophic and heterotrophic nutrition of coral reef zoanthids. In: <i>Proceedings</i>
676	of the 3rd International Coral Reef Symposium. 397–404.
677	Sebens KP. 1982. Intertidal distribuition of Zoanthids on the Caribbean Coast of Panamá: Effects
678	of predation and desiccation. Bulletin of Marine Science 32:316–335.
679	Segal B., Castro CB. 2011. Coral community structure and sedimentation at different distances



680	from the coast of the Abrolhos Bank, Brazil. Brazilian Journal of Oceanography 59:119-
681	129. DOI: 10.1590/S1679-87592011000200001.
682	Silva JF., Gomes PB., Santana EC., Silva JM., Lima ÉP., Santos AMM., Pérez CD. 2015.
683	Growth of the tropical zoanthid Palythoa caribaeorum (Cnidaria: Anthozoa) on reefs in
684	northeastern Brazil. Anais da Academia Brasileira de Ciências 87:00-00. DOI:
685	10.1590/0001-3765201520140475.
686	Sorokin YI. 1991. Biomass, metabolic rates and feeding of some common reef zoantharians and
687	octocorals. Australian Journal of Marine and Freshwater Research 42:729–741. DOI:
688	10.1071/MF9910729.
689	Spalding MD., Fox HE., Allen GR., Davidson N., Ferdaña Z a., Finlayson M., Halpern BS.,
690	Jorge M A., Lombana A., Lourie S A., Martin KD., Mcmanus E., Molnar J., Recchia C A.,
691	Robertson J. 2007. Marine Ecoregions of the World: A Bioregionalization of Coastal and
692	Shelf Areas. <i>BioScience</i> 57:573. DOI: 10.1641/B570707.
693	Stampar SN., Silva PF., Luiz Jr. OJ. 2007. 2007. Predation on the Zoanthid Palythoa
694	caribaeorum (Anthozoa, Cnidaria) by a Hawksbill Turtle (Eretmochelys imbricata) in
695	Southeastern Brazil. Marine Turtle Newsletter 117: 3.
696	Steiner AQ., Amaral FMD., Amaral JRBC., Sassi R., Barradas JI. 2015. Zonação de recifes
697	emersos da Área de Proteção Ambiental Costa dos Corais, Nordeste do Brasil. <i>Iheringia</i> .
698	Série Zoologia 105:184–192. DOI: 10.1590/1678-476620151052184192.
699	Suchanek TH., Green DJ. 1981. Interspecific competition between Palythoa caribaeorum and
700	other sessile invertebrates on St. Croix Reefs, U.S Virgin Islands. Proceedings of the 4th
701	International Coral Reef Symposium 2:679–684.
702	Swets JA. 1988. Measuring the accuracy of diagnostic systems. Science (New York, N.Y.)



- 703 240:1285–1293. DOI: 10.1126/science.3287615.
- Talley LD., Pickard GL., Emery WJ., Swif JH. 2011. Atlantic Ocean. In: Descriptive Physical
- 705 Oceanography. Elsevier, 555.
- 706 Tanner JE. 1995. Competition between scleractinian corals and macroalgae: An experimental
- investigation of coral growth, survival and reproduction. *Journal of Experimental Marine*
- 708 *Biology and Ecology* 190:151–168. DOI: 10.1016/0022-0981(95)00027-o.
- 709 Tester PA., Vandersea MW., Buckel CA., Kibler SR., Holland WC., Davenport ED., Clark RD.,
- Edwards KF., Taylor JC., Pluym JL Vander., Hickerson EL., Litaker RW. 2013.
- Gambierdiscus (Dinophyceae) species diversity in the Flower Garden Banks National
- 712 Marine Sanctuary, Northern Gulf of Mexico, USA. *Harmful Algae* 29:1–9. DOI:
- 713 10.1016/j.hal.2013.07.001.
- 714 Thomas L., Kennington WJ., Evans RD., Kendrick GA., Stat M. 2017. Restricted gene flow and
- local adaptation highlight the vulnerability of high-latitude reefs to rapid environmental
- 716 change. *Global Change Biology* 23:2197–2205. DOI: 10.1111/gcb.13639.
- 717 Tyberghein L., Verbruggen H., Pauly K., Troupin C., Mineur F., De Clerck O. 2012. Bio-
- ORACLE: a global environmental dataset for marine species distribution modelling. *Global*
- 719 *Ecology and Biogeography* 21:272–281. DOI: 10.1111/j.1466-8238.2011.00656.x.
- 720 Vazyulya S., Khrapko A., Kopelevich O., Burenkov V., Eremina T., Isaev A. 2014. Regional
- algorithms for the estimation of chlorophyll and suspended matter concentration in the Gulf
- of Finland from MODIS-Aqua satellite data. *Oceanologia* 56:737–756. DOI:
- 723 10.5697/oc.56-4.737.
- 724 Verbruggen H. 2012. Occurrence Thinner version 1.04.
- Villaça R., Pitombo FB. 1997. Benthic communities of shallow-water reefs of abrolhos, brazil.





726	Revista brasileira de oceanografia 45:35–43.
727	Villamizar E., Camisotti H., Rodríguez B., Pérez J., Romero M. 2008. Impacts of the 2005
728	Caribbean bleaching event at Archipiélago de Los Roques National Park, Venezuela.
729	Revista de biología tropical 56:255–270.
730	van Vuuren DP., Edmonds J., Kainuma M., Riahi K., Thomson A., Hibbard K., Hurtt GC., Kram
731	T., Krey V., Lamarque J-F., Masui T., Meinshausen M., Nakicenovic N., Smith SJ., Rose
732	SK.2011. The representative concentration pathways: an overview. Climatic Change 109:5-
733	31. DOI: 10.1007/s10584-011-0148-z.
734	Walsh JJ., Dieterle DA., Esaias WE. 1987. Satellite detection of phytoplankton export from the
735	mid-Atlantic Bight during the 1979 spring bloom. Deep-Sea Research 34:675-703.
736	Warren DL., Seifert SN. 2011. Ecological niche modeling in Maxent: the importance of model
737	complexity and the performance of model selection criteria. 21:335–342.
738	Warren DL., Wright AN., Seifert SN., Shaffer HB. 2014. Incorporating model complexity and
739	spatial sampling bias into ecological niche models of climate change risks faced by 90
740	California vertebrate species of concern. <i>Diversity and Distributions</i> 20:334–343. DOI:
741	10.1111/ddi.12160.
742	Wood R. 1993. Nutrients, predation and the History of Reef-Building. Society of Sedimentary
743	Geology 8:526–543.
744	

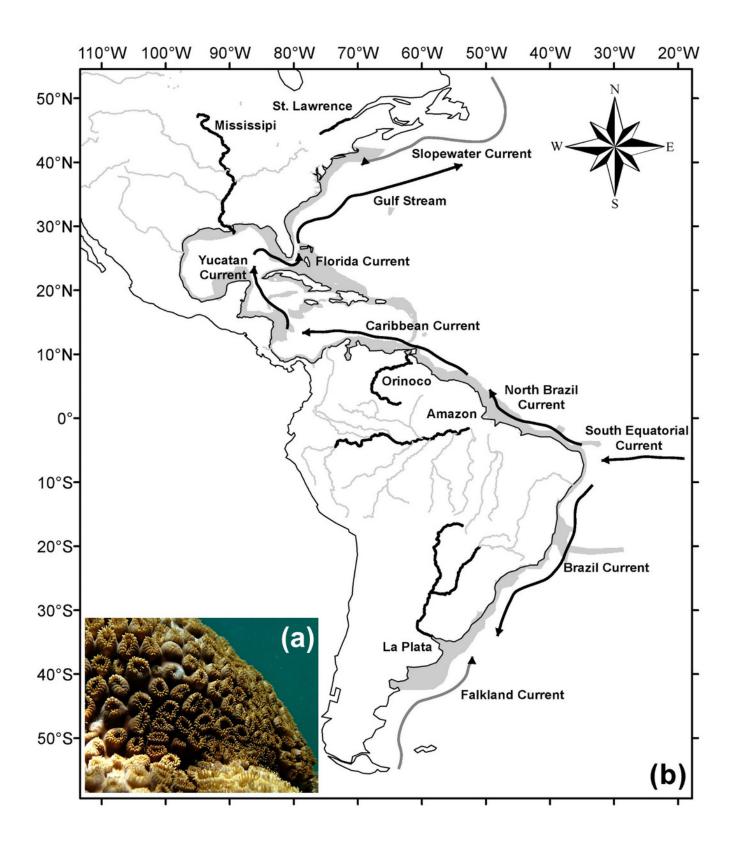


## Figure 1

Study area and major currents and rivers.

Underwater picture of (a) a *Palythoa caribaeorum* colony and (b) the study area (gray) with main rivers (black), adapted from the map "World Major Rivers" from ESRI; the schema of the main cold surface currents (gray) and warm surface currents (black) is adapted from Talley et al. (2011). Photo credit: Leonardo Durante.



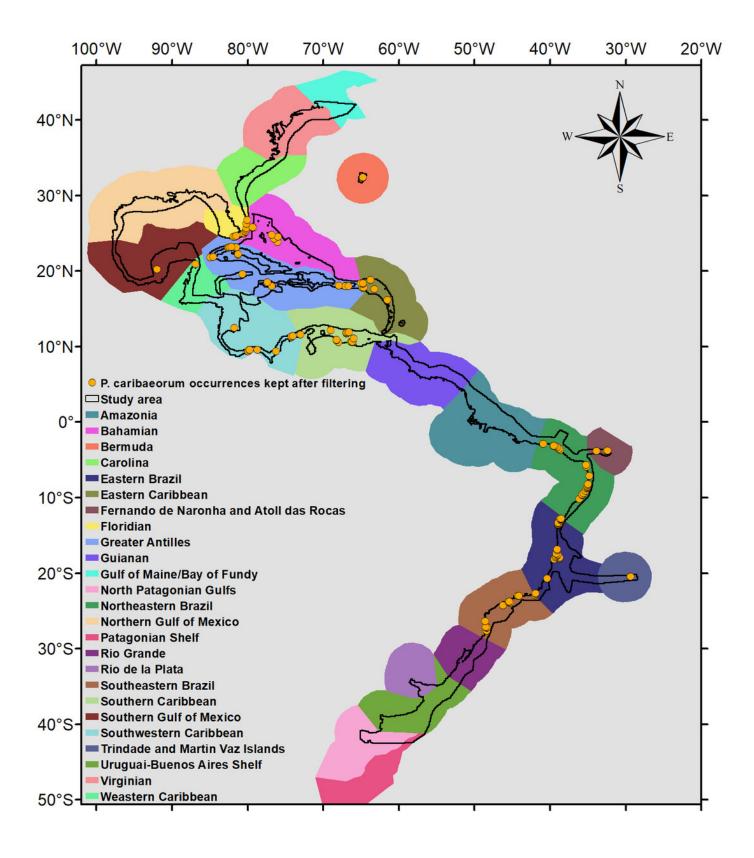




Occurrence points and Marine Ecoregions included in the study area

Study area and occurrence points of *P. caribaeorum* after filtering, with the Marine Ecoregions of the World as a reference (Spalding et al., 2007).

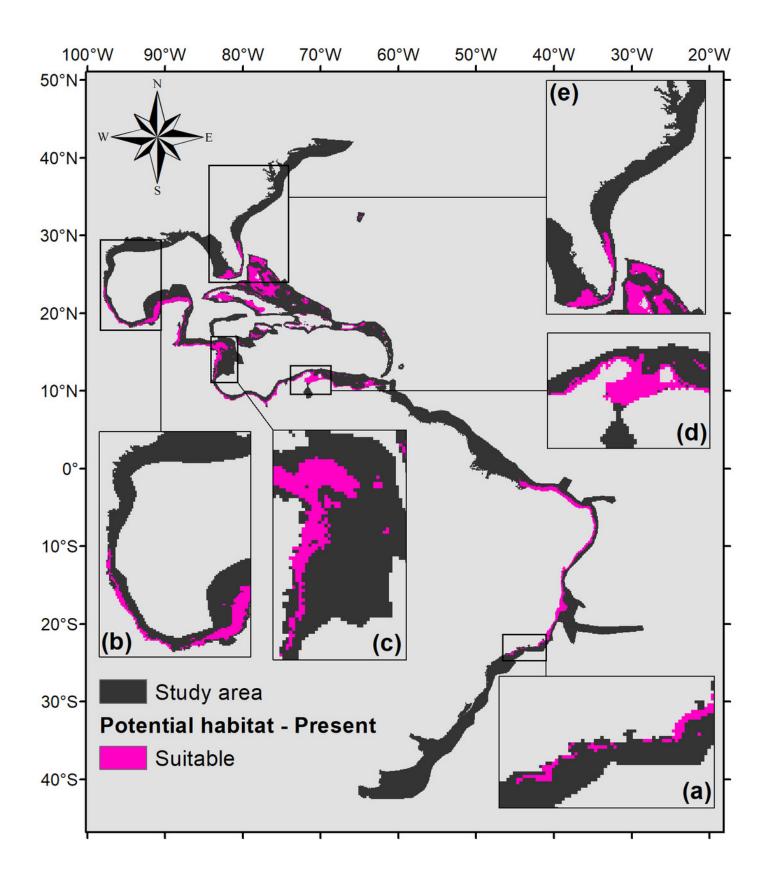






Suitable habitat for *P. caribaeorum* in the present

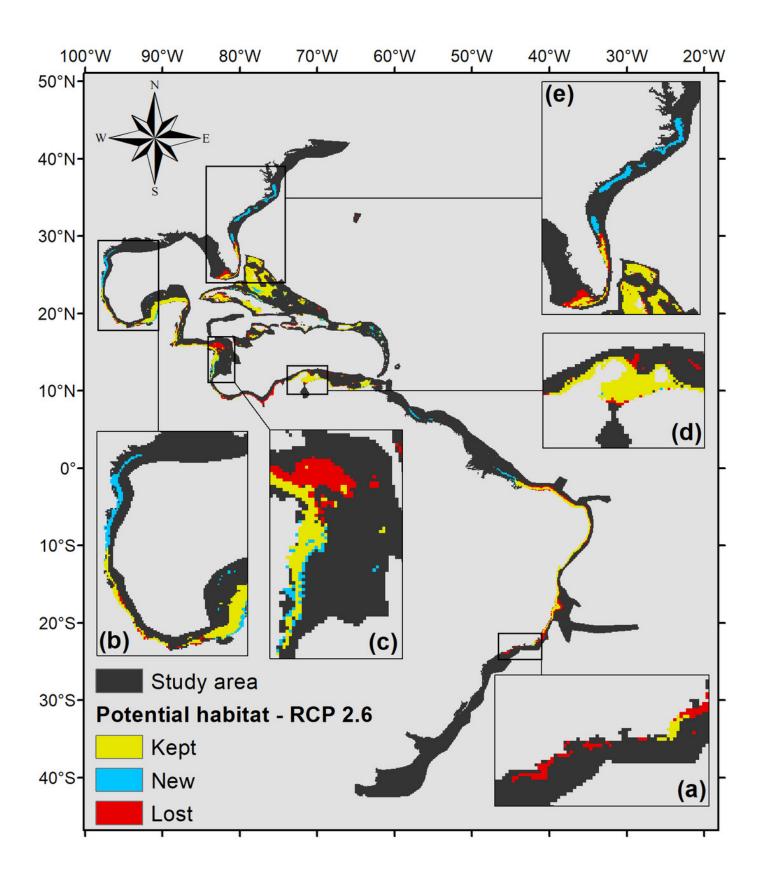
Details for (a) southeastern Brazil, (b) southern Gulf of Mexico, (c) southwestern Caribbean, (d) southern Caribbean, and (e) Floridian regions.





Suitable habitat for *P. caribaeorum* for the year 2100 under the RCP 2.6 climate scenario

Suitable habitat for *P. caribaeorum* for the year 2100 under the RCP 2.6 climate scenario, which includes regions with retained, new, and lost suitability compared with the present. Details for (a) southeastern Brazil, (b) southern Gulf of Mexico, (c) southwestern Caribbean, (d) southern Caribbean, and (e) Floridian regions.





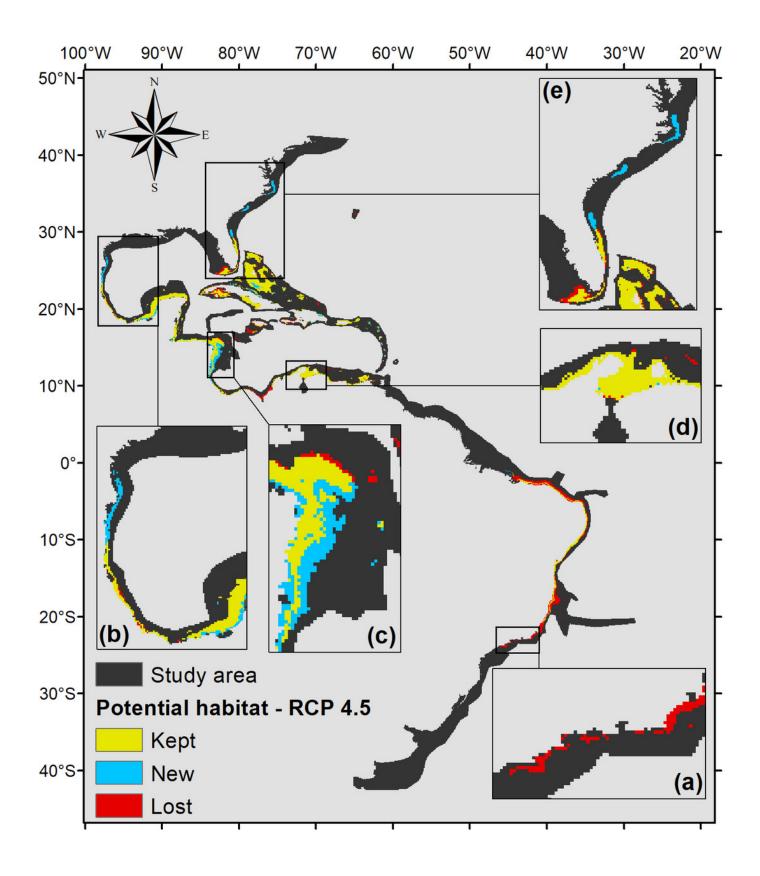
Habitat suitability map for *P. caribaeorum* in the year 2100 under the RCP 4.5 climate scenario.

The map includes regions with retained, new, and lost suitability compared with the present.

Details for (a) southeastern Brazil, (b) southern Gulf of Mexico, (c) southwestern Caribbean,

(d) southern Caribbean, and (e) Floridian regions.





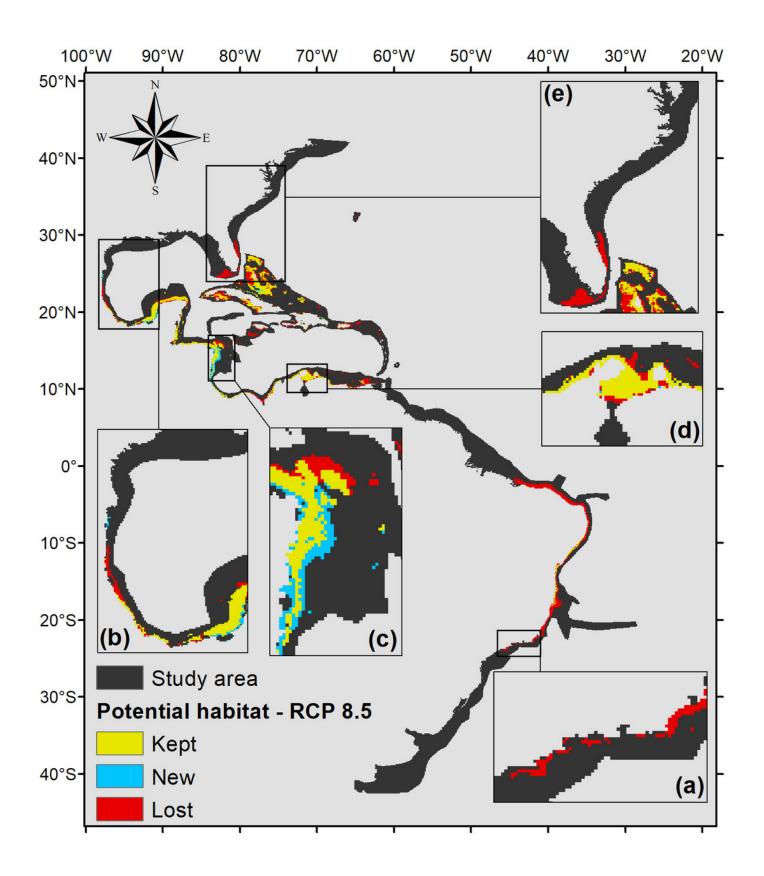


Habitat suitability map for *P. caribaeorum* in the year 2100 under the RCP 8.5 climate scenario.

The map includes regions with retained, new, and lost suitability compared with the present.

Details for (a) southeastern Brazil, (b) southern Gulf of Mexico, (c) southwestern Caribbean,

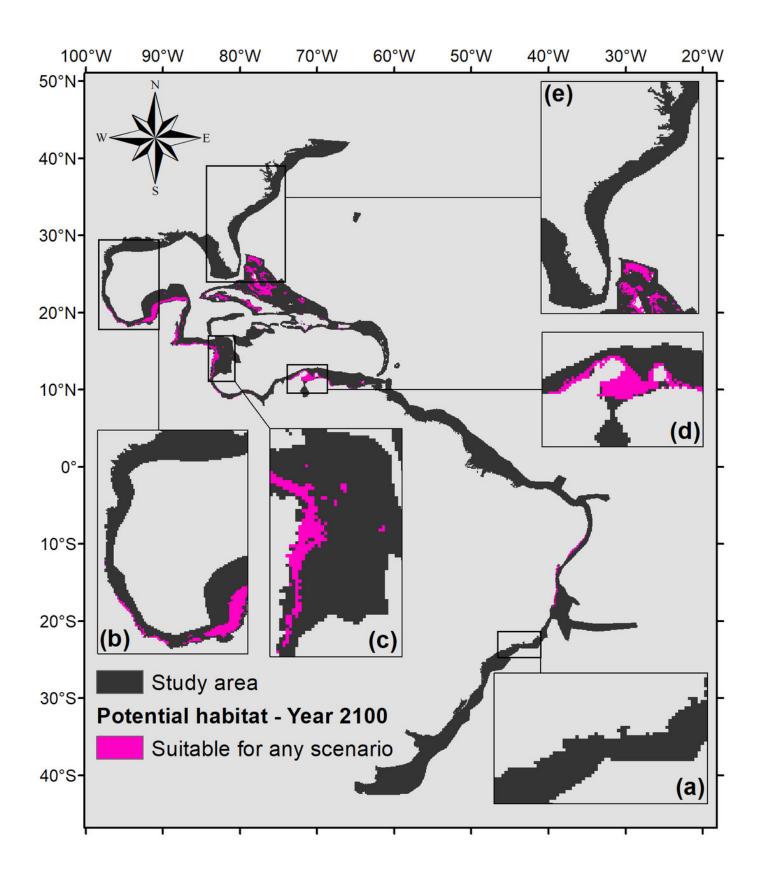
(d) southern Caribbean, and (e) Floridian regions.





Map of potential habitat for *P. caribaeorum* in the year 2100 under any climate scenario (pink)

Details for (a) southeastern Brazil, (b) southern Gulf of Mexico, (c) southwestern Caribbean, (d) southern Caribbean, and (e) Floridian regions.





#### Table 1(on next page)

Details about the variables retrieve from Bio-Oracle dataset and the layers kept in the final distribution model for *P. caribaeorum*.

Variables	Unit	Measurement	Measurement	Number of	Kept in the final
			transformations	layers	model
					Average
Calcite concentration	mol/m³	Remote sensor	Average	1	concentration
			Average, minimum, maximum		Maximum
Chlorophyll-a concentration	mg/m³	Remote sensor	and range	4	concentration
			Average, minimum and		
Cloud coverage	%	Remote sensor	maximum	3	-
			Average, minimum and		
Difusive atenuation coeficient (490 nm)	1/m	Remote sensor	maximum	3	-
Photosynthetic available radiation	Einstein/m²/day	Remote sensor	Average and maximum	2	-
			Average, minimum, maximum		
Air temperature above sea surface	°C	Remote sensor	and range	4	-
			Average, minimum, maximum		
Sea surface temperature	°C	Remote sensor	and range	4	Range
Dissolved oxygen	ml/l	In situ	Diva interpolation	1	-
Nitrate	μmol/l	In situ	Diva interpolation	1	-
рН	-	In situ	Diva interpolation	1	Diva interpolation
Phosphate	μmol/l	In situ	Diva interpolation	1	-
Salinity	PSU	In situ	Diva interpolation	1	Diva interpolation
Silicate	μmol/l	In situ	Diva interpolation	1	-

<sup>1</sup> 

Table 1: Details about the variables retrieve from Bio-Oracle dataset and the layers kept in the final distribution model for *P*.

<sup>3</sup> caribaeorum.