1	rreparea as an original research article for reers
2	
3	Snapshot recordings provide a first description of the acoustic signatures of deeper
4	habitats adjacent to coral reefs of Moorea.
5	
6	Frédéric Bertucci ^{1,2,*} , Eric Parmentier ² , Cécile Berthe ¹ , Marc Besson ^{1,3} , Anthony D.
7	Hawkins ⁴ , Thierry Aubin ⁵ , David Lecchini ^{1,6}
8	
9	¹ PSL Research University: EPHE-UPVD-CNRS, USR 3278 CRIOBE, Moorea, French
10	Polynesia.
11	² Laboratoire de Morphologie Fonctionnelle et Evolutive, AFFISH-RC, University of Liège,
12	Belgium.
13	³ Observatoire Océanologique de Banyuls-sur-Mer, UMR 7232, CNRS-UPMC, Banyuls-sur-
14	Mer, France.
15	⁴ Loughine Marine Research, Aberdeen, UK
16	⁵ Neuro-PSI, UMR 9197, CNRS, University Paris-Saclay, France
17	⁶ Laboratoire d'Excellence "CORAIL".
18	
19	*Corresponding author: fred.bertucci@gmail.com

ABSTRACT

Acoustic recording has been recognized as a valuable tool for non-intrusive monitoring of the marine environment, complementing traditional visual surveys. Acoustic surveys conducted on coral ecosystems have so far been restricted to barrier reefs and to shallow depths (10-30m). Since they may provide refuge for coral reef organisms, the monitoring of outer reef slopes and describing of the soundscapes of deeper environment could provide insights into the characteristics of different biotopes of coral ecosystems. In this study, the acoustic features of four different habitats, with different topographies and substrates, located at different depths from 10 to 100m, were recorded during day-time on the outer reef slope of the north Coast of Moorea Island (French Polynesia). Barrier reefs appeared to be the noisiest habitats whereas the average sound levels at other habitats decreased with their distance from the reef and with increasing depth. However, sound levels were higher than expected by propagation models; supporting that these habitats possess their own sound sources. While reef sounds are known to attract marine larvae, sounds from deeper habitats may then also have a non-negligible attractive potential, coming into play before the reef itself.

INTRODUCTION

35

The existence of coral reefs at depths of more than 150 m in tropical regions has been 36 known for decades (Fricke & Schumacher, 1983; Maragos & Jokiel, 1985; Kahng & 37 38 Maragos, 2006). Recently, the conservation and management of these so-called mesophotic coral ecosystems (MCEs) has been considered a priority, although the reefs themselves 39 40 remain largely unexplored (Pyle et al., 2016). More generally, deeper zones and habitats close to coral reefs may serve as refuges and be the origin of recruits that contribute to the recovery 41 of reefs located in shallow waters (Bongaerts et al., 2010; van Oppen et al., 2011). Deeper 42 habitats were, at first, thought to be more protected from temperature increases and coral 43 bleaching events, since the impacts of human and natural perturbations typically diminish 44 with depth and distance from shore (Glynn, 1996; Feingold, 2001; Glynn et al., 2001; Bak, 45 Nieuwland & Meesters, 2005). However, the current degree of global climate change may 46 also have an impact upon deeper habitats (Appeldoorn et al., 2016). Very little is known about 47 48 MCEs, and the deeper habitats adjacent to coral reefs. Hence there is a wide range of possible topics to be investigated. As an example, being able to provide an acoustic description of such 49 habitats may appear crucial for extending our current knowledge on marine soundscapes. In 50 particular, it may provide insights into the qualities and characteristics of the deeper habitats 51 associated with coral ecosystems (Staaterman et al., 2014; Bertucci et al., 2015; Nedelec et 52 al., 2015; Bobryk et al., 2016). 53 A soundscape is defined as the collection of all sounds that are present in a landscape, 54 55 which vary over space and time (Southworth, 1969; Schafer, 1977; Krause, 1987; Pijanowski 56 et al., 2011). Within soundscapes, sound sources are divided into three main components: the biophony (corresponding to biologically produced sounds), the geophony (the geophysically 57 produced sounds) and the anthropophony (the sounds produced by human activities). The 58 collection of data regarding the nature and qualities of marine soundscapes is growing 59

worldwide (Cato & McCauley, 2002; Chapman & Price, 2011; Bertucci et al., 2016) but despite their potential, investigations of temporal and spatial variations in the soundscapes of 62 coral reefs have mainly been concentrated on comparing neighbouring sites consisting of different habitat types, e.g. mangrove, fringing reef and barrier reef, and they have been 63 restricted to the first 10-20 m of the water column (Piercy et al., 2014; Radford, Stanley & 64 Jeffs, 2014; Bertucci et al., 2015). For instance, Staaterman et al. (2014) described marine soundscapes of Florida reefs in 7m of water while only the 0-5m range was recorded in a French Polynesian reef by Nedelec et al. (2015), and in a temperate coastal marine environment by Rossi, Connell & Nagelkerken (2016). Overall, the studies highlighted that different habitat types are characterised by peculiar acoustic features that constitute their acoustic signatures. When assessing the relationship between biodiversity and soundscape 70 features of similar reefs habitats in Virgin Islands and French Polynesia, recordings performed by Kaplan et al. (2015) and Bertucci et al. (2016) were carried out at 18 m and 10 m depth, 72 73 respectively. 74 These studies performed in shallow waters demonstrate that acoustical differences between reef habitats are due to variations in the sonic activity of marine organisms, i.e. soniferous 75 fishes, snapping shrimps (Radford, Stanley & Jeffs, 2014) and the geo-morphology of recorded sites. Acoustic cues within habitats close to coral reefs are known to influence the behaviour and orientation of many fish and invertebrate larvae at settlement (Simpson et al., 78 2004; Vermeij et al., 2010; Nedelec et al., 2015; Parmentier et al., 2015). Describing 79 80 soundscapes from deeper habitats' could further highlight the importance of acoustic cues in 81 the distribution of marine organisms, and the attractiveness of deeper habitats associated with reefs. Many coral reef-associated fish species have highly specialized habitat requirements. 82 Some species are typically found in sandy patches while some other will use different types of coral as shelters, which will lead to differing species assemblages (Bacchet et al., 2006).

60

61

65

66

67

68

69

71

76

77

83

Vocal species from these assemblages should create differential acoustic signatures in the frequency range in which they produce sounds. From this perspective, the objective of the present study was to investigate the variations of sound pressure levels in the low frequency range between underexplored habitats adjacent to coral reefs on the north coast of Moorea Island, French Polynesia (17°30'S, 149°5'W) and thereby provide a first insight into the acoustic features of these biotopes. Positive results in this pioneer study should motivate additional works including more depths and additional time scales.

Comment [JB1]: Id cut this sentence

MATERIAL & METHODS

Study sites

 The study was carried out at the end of the warm season, from June to July 2015, along three North-orientated seaward transects characterized by increasing depths, extending from the barrier reef (BR, <20 m), to the sandy plain (SP, 30-50 m), the reef slope (RS, 50-65 m) and to the more distant reef drop-off (DO, 75-100 m). The barrier reef of transect 1 was located in a Marine Protected Area (MPA) while barrier reefs of transects 2 and 3 were located in non-protected areas. Transects 1 and 2 were separated by the pass of Tiahura, a coral-free area located in front of an opening in the reef crest that canalizes the water flow to the ocean (Fig. 1). Bertucci et al. (2015) showed that this habitat had a mean sound intensity of ca. 90 dB re 1μ Pa (20 – 5000 Hz range). Transects 1 and 2 were 1.0 km apart, transects 2 and 3 were 1.3 km apart. For each transect, 4 different habitats corresponding to different depths were explored from the barrier towards the ocean: the barrier reef (BR, characterized by a water depth of 1–20 m and a substratum comprised of up to 40% live coral and a wide range of fish and invertebrate species; the sandy plain (SP), constituting the base of the reef with a declivity of 30-45° and characterised by vast expanses of patchy rocks covered by coral, with a high species diversity and located at 30-50 m depth; the reef slope (RS),

Deleted: presenting

Deleted:

Deleted:

characterized by a change of slope with an increased declivity located at 50-65 m depth, this zone of sedimentary accumulation is remarkable for its low specific richness and the density of benthic communities; and the reef drop-off (DO), characterized by a cliff located at 75-100 m depth and numerous fish species. Depth and topography were measured with multi-beam sonar (Lowrance LMS 527) installed on an 18-feet boat. The positions of the different recording sites were localized with a GPS in order to replicate measurements.

Recordings

Recordings were conducted between 09:00 and 16:00. This period of time shows little variations in acoustic activity, in contrast to early morning and late afternoon where drastic changes in sound intensity and complexity may happen rapidly (Bertucci et al., 2015; Bertucci et al., 2016). Recordings were made when wind speed was lower than 5 knots, with no swell, so as to prevent the boat from drifting and bobbing, and to reduce noise from the wind and sea surface turbulence against the boat's hull. Recordings were only made when no other boats were observed in the recording area.

An underwater Remora acoustic recorder (Loggerhead Instruments, Sarasota, FL, USA) connected to a HTI96-min hydrophone (sensitivity: −211 dB re: 1V for a sound pressure of 1μPa, frequency response: 2 Hz − 30 kHz; High Tech Inc., Long Beach, MS, USA) was attached to a block of lead placed at the end of a 100m rope. This minimized vibration of the rope and current-driven movement of the device. The measurements sequence consisted of recording the 4 different habitats of a single northward transect, i.e. BR, SP, RS, DO, in a random order. For each habitat, the recorder was suspended from the boat and lowered by an experimenter into the water until it was 5m above the sea floor. The depth of the device (recorder and hydrophone) was determined by means of marks positioned every 5m along the rope. Water depth was measured every 2 minutes with the sonar system to ensure that

Deleted:

Deleted: housing

Deleted: by means of a

Deleted: Recordings started and the recorder was deployed when the boat was positioned above the desired habitat and after the boat engine was turned off to reduce acoustic and physical disturbances.

recordings were made within the appropriate habitat, at a constant depth, and that the hydrophone did not risk hitting the sea floor. Recordings lasted 10 minutes (sampling rate of 44.1 kHz, 16-bit resolution with a 33 dB gain) before the recorder was pulled back on board and switched off. Completing one transect took approximately 90 min. For each transect, 3 replicates were obtained for each habitat type with a one week time interval between them (Table 1). A total of 360 minutes of recordings were collected.

151

152 153

154

155

156

157

158

159

160

161

162

163

164

165

166

167

168

169

145

146

147

148

149

150

Data Analysis

A 20 Hz high-pass frequency filter was applied to all recordings to eliminate very low frequencies. The start and end sections corresponding respectively to the positioning and withdrawal of the recorder were deleted. Recordings were further cleansed by visually inspecting sound spectrograms using Avisoft SASLab Pro 5.2.07 software (Avisoft Bioacoustics, Glienicke, Germany) in order to cut-out anthropogenic sound sources and other artefactual sounds (e.g. animals probing the recording device or movement of the rope). This cleansing-step shortened some of the recordings down to 4 min. For each habitat, a set of 12 subsamples of 60s were used, which were randomly extracted from the 3 replicates in order to produce spectra based on recordings of the same duration (Fast Fourier Transform FFT, 1024 points Hamming window, providing a 21.53 Hz resolution). The sound pressure levels measured for each 21.53 Hz frequency band (SPL in dB re: 1µPa, logarithmic scale) were transformed into µPa (linear scale) and averaged. Averaged sound pressure levels were then converted back into dB re: 1µPa to present the average spectrum of each habitat. The characteristics of each habitat were described on the basis of variations of the average spectral profiles. For each habitat, the root mean square (RMS) of the sound pressure level was measured on the 20Hz – 2.5 kHz frequency band using Avisoft SASLab Pro 5.2.07 software. This low frequency band is dominated by fish vocalizations (Lobel, Kaatz & Rice, 2010;

Formatted: Space After: 8 pt, Adjust space between Latin and Asian text, Adjust space between Asian text and numbers

Tavolga, Popper & Fay, 2012). Due to a low sample size (3 replicates for each sampling point), average sound pressure levels were compared between the 3 transects and also between the 4 different habitats with a Kruskal-Wallis tests, followed by Tukey's post-hoc tests for pairwise comparisons. Intensity values of each frequency bins (N = 116) of power spectra were normally distributed (Shapiro-Wilks tests, W = 0.98-0.99, all P > 0.05) and were compared between the 3 transects and the 4 habitat types with two-way ANOVAs followed by Tukey's post-hoc tests for pairwise comparisons. All analyses were two-tailed, at $\alpha = 0.05$ and carried out with R 3.1.2 (R Core Team 2014) using customized scripts.

Deleted: by means of

Deleted: by means of

Deleted:

179 RESULTS

Spectral signatures of deeper habitats are mainly characterized by lower sound

intensities

For each transect, BR locations presented spectra with significantly higher average sound pressure levels, followed by SP, RS and DO. Average sound pressure levels decreased from BR towards the most distant and deepest habitat, *i.e.* DO (Kruskal-Wallis, $\chi^2 = 19.24 - 30.62$, df = 3, all P-values < 10^{-3} ; Tukey's post-hoc tests for pairwise comparisons) (Fig. 2). For transect 1, BR showed significantly higher average sound pressure levels than the 3 other habitats for all frequencies above 100 Hz. SP showed significantly higher average sound pressure levels than RS and DO from 150 Hz to 2500 Hz (only a narrow 2200 Hz - 2300 Hz range did not differ significantly between SP and RS). RS and DO spectra were not significantly different for any frequencies but a narrow frequency band from 1250 Hz to 1400 Hz (Two-way ANOVA, $F_{3,116} = 2.00 - 2.70$; P < 0.05) (Fig. 3A). For transect 2, BR and SP differed significantly only from 400 Hz to 1500 Hz. As for transect 1, RS and DO showed significantly lower intensities than BR for most frequencies, and were significantly lower in their intensities than SP for frequencies below 2000 Hz. In contrast to transect 1, RS showed

significantly higher average sound pressure levels than DO for frequencies below 750 Hz, between 1400 Hz and 1500 Hz, and from 2000 Hz to 2300 Hz (Two-way ANOVA, $F_{3,116}$ = 2.50; P < 0.05) (Fig. 3B). For transect 3, all spectra but RS and DO were significantly different for all frequencies from 400 Hz to 2000 Hz with decreasing intensities from BR to DO. BR showed significantly higher intensities at lower intensities too and significantly higher intensities than RS and DO above 2000 Hz. RS and DO did not differ significantly in their intensities except for a very narrow 1100 Hz – 1200 Hz band caused by the absence of a peak present in the other spectra (Two-way ANOVA, $F_{3,116}$ = 2.30; P < 0.05) (Fig. 3C).

Similar habitats show differences in their spectral signatures

At the habitat type level, BRs and RSs showed the greatest difference in power spectra between the three transects. For BR, transect 1 showed significantly higher intensities for all frequencies of the spectrum compared to transect 2. The spectrum of BR of transect 2 was characterised by two intensity peaks around 600 Hz and 1200 Hz. The spectra of transects 1 and 3 differed significantly for frequencies above 1250 Hz. BR spectra of transects 2 and 3 differed significantly only for a narrow frequency range between 1250 Hz and 1500 Hz (Twoway ANOVA, $F_{2,116} = 2.80$; P < 0.05) (Fig. 4). For RS, transect 1 showed the significantly highest intensities for most frequencies above 500 Hz while transects 2 and 3 were similar (Two-way ANOVA, $F_{2,116} = 2.45$; P < 0.05) (Fig. 4). Spectra of SP and DO showed little variation in their intensities between the three transects. DO of transect 1 showed higher intensities below 400 Hz than DO of transect 2; and DO of transect 2 and 3 differed for higher frequencies between 1600 Hz and 2300 Hz (Two-way ANOVA, $F_{2,116} = 2.35 - 3.00$; P < 0.05) (Fig. 4).

The BR of transect 1 displayed the highest average sound pressure level (Kruskal-Wallis,

 $\chi^2 = 11.24$, df = 2, P-value = 0.004, Tukey's post-hoc tests for pairwise comparisons, P-values

Deleted: significant

Deleted: s

< 0.05) (Fig. 5). No difference was observed between BRs of transect 2 and 3. The power spectra of SPs showed an inverted pattern with SP of transect 1 showing the lowest average sound pressure level (Kruskal-Wallis, $\chi^2 = 10.27$, df = 2, P-value = 0.006, Tukey's post-hoc tests for pairwise comparisons, P-values < 0.05) (Fig. 5). No difference was observed between SPs of transect 2 and 3. RS and DO showed the significantly lowest sound pressure level for all transects (Kruskal-Wallis, $\chi^2 = 19.24 - 30.62$, df = 3, all P-values < 10^{-3} ; Tukey's post-hoc tests for pairwise comparisons) with all values below 92 dB re: 1µPa. RS of transect 2 showed significantly lower average sound pressure level (Kruskal-Wallis, $\chi^2 = 9.08$, df = 2, P-value = 0.010, Tukey's post-hoc tests for pairwise comparisons, P-values < 0.05). The DO of transect 2 showed a significantly higher average sound pressure level of 95 dB re: 1µPa (Kruskal-Wallis, $\chi^2 = 13.14$, df = 2, P-value = 0.001, Tukey's post-hoc tests for pairwise comparisons, P-values < 0.05) (Fig. 5).

DISCUSSION

Recent studies on coral reefs have highlighted the positive relationships between sound signatures and coral cover, density of fishes or increased number of biotic sound sources in shallow waters (Kaplan et al., 2015; Nedelec et al., 2015; Bertucci et al., 2016). In this study, comparison of the spectra of four types of habitats adjacent to coral reefs, characterized by different topographies and substrates, by increasing depths and distances from the reef, revealed different spectral profiles most probably related to their different physical and biological properties.

Recordings made at the barrier reef presented higher sound pressure levels in the low frequency range despite the fact that low frequencies transmit poorly in shallow waters (Rogers & Cox, 1988). The higher level of the biophony would suggest this habitat type is occupied by more vocal organisms than the three others. We cannot exclude, however, the

Deleted: housing

possibility that low frequencies are related to a greater contribution of the geophony at the barrier reefs, with sounds produced by crashing waves or by moving substrate (sand) leading to increased sound pressure levels. The short distance between the sea surface and the bottom can also produce more reverberation in shallow waters than in deeper waters and increase sound levels at higher frequencies. Hence, lower sound levels in deeper habitats do not necessarily mean that mesophotic reefs would be less acoustically rich than barrier reefs. A description of potential vocal species present in these habitats together with the description of their physical environments may help to distinguish between the respective contributions from different sources, *i.e.* biophony and geophony.

251

252

253

254

255256

257

258

259

260

261

262

263

264

265

266

267

268

269

270

271

272

273

274

275

Recorded sound pressure levels in deeper habitats highlight that sounds do not result from the sole propagation and degradation of sounds produced at the level of the noisy barrier reef, as predicted by propagation models (Mann et al., 2007) (Table 2). Moreover, sound levels of sandy plains of transects 2 and 3 showed higher average sound levels than their counterpart of transect 1 despite being more distant from their respective barrier reefs. This clearly supports that barrier reef is not the only sound source and that deeper habitats possess their own sound sources. The weak variations between the sandy plain spectra may result from their topography and physical characteristics with large patchy rocky habitats and similar communities between sites. At reef slopes, more distinct signatures appear again with reef slope of transect 3 being as noisy as reef slope of transect 1 despite being located further away from the barrier. Finally, while the spectra of drop-offs of transects 1 and 3 show no significant variations with their respective reef slopes, drop-off at transect 2 was characterized by a higher sound pressure level compared to its reef slope, especially at the low frequencies. So, the decrease of sound intensities along transects appears to be variable and the spectral profiles of the different habitats show significantly increased intensities despite their distance to the barrier reef. These observations would reinforce the idea that the barrier reef is not the

single sound source responsible of observed spectra and that additional sources such as soniferous species – that are not present at the level of the barrier or may differ in abundance at deeper habitats – may actively play a role in the sonic signature of deeper environments. In particular, as the transitions between the reef slopes and the reef drop-offs are very short for all transects, the observed differences (especially for transect 2) between spectra and the potential link with differential biological activity deserve to be investigated. Indeed, while the reef slope is a zone of sedimentary accumulation with a low species richness, the reef drop-off houses numerous fish species. This might explain the increased intensities between these 2 types of habitats.

Overall, as in shallow waters, differences in spectral patterns seem to exist in deeper habitats and may therefore reflect different characteristics of the habitats, *i.e.* physical and/or biological. Acoustic cues produced by deeper habitats may therefore also be used in the orientation of marine larvae and may come into play earlier in the recruitment of coral reefs organism. However, the short period of time sampled in this study remains insufficient to reliably characterize the different habitats and only provide initial information on the different acoustic signatures of deeper habitats. Fish vocal activity can change drastically at dusk or dawn and be more sustained at night with specific species vocalizing at in specific time windows (Pieretti et al. 2017). Monitoring for longer time periods might highlight further differences between habitats at the diel scale. Moreover, sound production in fishes is often linked to social activities, such as courtship interactions and spawning events that will vary on a longer, seasonal scale. Long term recordings of deeper environments would then be necessary to capture the complete picture and identify acoustic differences (Pieretti et al. 2017).

CONCLUSION

We still know very little about the acoustic ecology of deeper habitats adjacent to coral reefs. It has been assumed that deeper environments are less likely to be impacted by anthropogenic activities or by global change, and therefore may provide refuge areas for shallow reef species (Bongaerts et al. 2010). This hypothesis has gained a growing interest in the scientific community, but has been only tested at few locations for few species. In the future, the opportunity to have proxies of the ecological state of these refuge areas by means of soundscape analysis and linking their acoustic characteristics to their refuge potential may help to better judge the impact of global change and the influence of these adjacent ecosystems on coral reefs. Several studies have demonstrated that barrier reef sound attracts fish and crustacean larvae during settlement onto the reef (Barth et al., 2015). The present study represents a first step towards the acoustic investigation of deeper environments and suggests that deeper habitats could also play a role in the orientation of larval marine organisms.

Deleted:

Research in practically unexplored depths will undoubtedly bring new knowledge and tools that will be extremely valuable for the creation of Marine Protected Areas (MPAs), watershed management plans and the development of conservation plans for coral reefs as a whole, from shallow to deep water habitats.

ACKNOWLEDGEMENTS

Authors would like to thank Franck Lerouvreur for his technical support. We also thank 2 anonymous reviewers and Dr. Nadia Pieretti for their comments on a previous version of this manuscript.

324	REFERENCES
325	
326	Appeldoorn R, Ballantine D, Bejarano I, Carlo M, Nemeth M, Otero E, Pagan F, Ruiz H,
327	Schizas N, Sherman C, Weil E. 2016. Mesophotic coral ecosystems under anthropogenic
328	stress: a case study at Ponce, Puerto Rico. Coral Reefs 35(1):63-75.
329	Bacchet P, Zysman T, Lefèvre Y. 2006. Guide des poissons de Tahiti et ses îles, 4th edition.
330	Tahiti, French Polynesia: Au vent des îles editions.
331	Bak RP, Nieuwland G, Meesters EH. 2005. Coral reef crisis in deep and shallow reefs: 30
332	years of constancy and change in reefs of Curacao and Bonaire. Coral Reefs 24(3):475-
333	479.
334	Barth P, Berenshtein I, Besson M, Roux N, Parmentier E, Banaigs B, Lecchini D. 2015. From
335	the ocean to a reef habitat: how do the larvae of coral reef fishes find their way home? A
336	state of art on the latest advances. Life and Environment 65:91-100.
337	Bertucci F, Parmentier E, Berten L, Brooker RM, Lecchini D. 2015. Temporal and spatial
338	comparisons of underwater sound signatures of different reef habitats in Moorea Island,
339	French Polynesia. PLoS One 10(9):e0135733.
340	Bertucci F, Parmentier E, Lecellier G, Hawkins AD, Lecchini D. 2016. Acoustic indices
341	provide information on the status of coral reefs: an example from Moorea Island in the
342	South Pacific. Scientific Reports 6:33326.
343	Bobryk CW, Rega-Brodsky CC, Bardhan S, Farina A, He HS, Jose S. 2016. A rapid
344	soundscape analysis to quantify conservation benefits of temperate agroforestry systems
345	using low-cost technology. Agroforestry Systems 90(6):997-1008.
346	Bongaerts P, Riginos C, Ridgway T, Sampayo EM, van Oppen MJH, Englebert N, Vermeulen
347	F, Hoegh-Guldberg O. 2010. Genetic divergence across habitats in the widespread coral

Seriatopora hystrix and its associated Symbiodinium. PLoS One 5:e10871.

349	Cato DH, McCauley RD. 2002. Australian research in ambient sea noise. Acoustics
350	Australia 30:13-20.
351	Chapman NR, Price A. 2011. Low frequency deep ocean ambient noise trend in the Northeast
352	Pacific Ocean. Journal of the Acoustical Society of America 129:161-165.
353	Feingold JS. 2001. Responses of three coral communities to the 1997-98 El Nino-Southern
354	Oscillation: Galapagos Islands, Ecuador. Bulletin of Marine Science 69:61-77.
355	Glynn PW, Mate JL, Baker AC, Calderon MO. 2001. Coral bleaching and mortality in Panama
356	and Ecuador during the 1997–1998 El Nino-Southern Oscillation event: spatial/temporal
357	patterns and comparisons with the 1982-1983 event. Bulletin of Marine Science 69:79-
358	109.
359	Fricke H, Schuhmacher H. 1983. The depth limits of Red Sea stony corals: an
360	ecophysiological problem (a deep diving survey by submersible). PSZNI Marine Ecology
361	4:163-194.
362	Glynn PW. 1996. Coral reef bleaching: facts, hypotheses and implications. Global Change
363	Biology 2:495-509.
364	Kahng SE., Maragos JE. 2006. The deepest zooxanthellate scleractinian corals in the world?
365	Coral Reefs 25:254.
366	Kaplan MB, Mooney TA, Partan J, Solow AR. 2015. Coral reef species assemblages are
367	associated with ambient soundscapes. Marine Ecology Progress Series 533:93-107.
368	Krause B. 1987. Bioacoustics, habitat ambience in ecological balance. Whole Earth Review
369	57 :14-18.
370	Lobel PS, Kaatz IM, Rice AN. 2010. Acoustical behavior of coral reef fishes. In: Cole KS, ed.

 $Reproduction\ and\ Sexuality\ in\ Marine\ Fishes:\ Patterns\ and\ Processes.\ Berkeley:$

University of California Press, 307-386.

371

- 373 Mann DA, Casper BM, Boyle KS, Tricas TC. 2007. On the attraction of larval fishes to reef
- 374 sounds. Marine Ecology Progress Series 338:307-310.
- 375 Maragos JE, Jokiel PL. 1985. Reef corals of Johnston Atoll: one of the world's most isolated
- 376 reefs. Coral Reefs 4:141-150.
- 377 Nedelec SL, Simpson SD, Holderied M, Radford AN, Lecellier G, Radford C, Lecchini D.
- 2015. Soundscapes and living communities in coral reefs: temporal and spatial variation.
- 379 Marine Ecology Progress Series 525:125-135.
- Parmentier E, Berten L, Rigo P, Aubrun F, Nedelec SL, Simpson SD, Lecchini D. 2015. The
- influence of various reef sounds on coral-fish larvae behaviour. *Journal of Fish Biology*
- 382 86:1507-1518.
- 383 Piercy JJB, Codling EA, Hill AJ, Smith DJ, Simpson SD. 2014. Habitat quality affects sound
- production and likely distance of detection on coral reefs. Marine Ecology Progress
- 385 Series 516:35-47.
- 386 Pieretti N, Lo Martire M, Farina A, Danovaro R. 2017. Marine soundscape as an additional
- biodiversity monitoring tool: A case study from the Adriatic Sea (Mediterranean Sea).
- 388 *Ecological Indicators* 83:13-20.
- 389 Pijanowski BC, Villanueva-Rivera LJ, Dumyahn SL, Farina A, Krause BL, Napoletano BM,
- Gage SH, Pieretti N. 2011. Soundscape ecology: the science of sound in the landscape.
- 391 *BioScience* 61(3): 203-216.
- 392 Pyle RL, Boland R, Bolick H, Bowen BW, Bradley CJ, Kane C, Kosaki RK, Langston R,
- Longenecker K, Montgomery A, Parrish FA, Popp BN, Rooney J, Smith CM, Wagner D,
- 394 Spalding HL. 2016. A comprehensive investigation of mesophotic coral ecosystems in
- the Hawaiian Archipelago. *PeerJ* 4:e2475.
- 396 Radford CA, Stanley JA, Jeffs AG. 2014. Adjacent coral reef habitats produce different
- underwater sound signatures. *Marine Ecology Progress Series* 505:19-28.

399	Popper AN, Tavolga WN, eds. Sensory Biology of Aquatic Animals. New-York:
400	Springer-Verlag, 131-149.
401	Rossi T, Connell SD, Nagelkerken I. 2016. Silent oceans: ocean acidification impoverishes
402	natural soundscapes by altering sound production of the world's noisiest marine
403	invertebrate. Proceedings of the Royal Society B 283(1826):20153046.
404	Schafer RM. 1977. The tuning of the world: Toward a theory of soundscape design.
405	Pennsylvania: University of Pennsylvania Press.
406	Simpson SD, Meekan MG, McCauley RD, Jeffs A. 2004. Attraction of settlement-stage coral
407	reefs fishes to ambient reef noise. Marine Ecology Progress Series 276:263–268.
408	Staaterman E, Paris CB, DeFerrari HA, Mann DA, Rice AN, D'Alessandro EK. 2014.
409	Celestial patterns in marine soundscapes. Marine Ecology Progress Series 508:17-32.
410	Southworth M. 1969. The sonic environment of cities. <i>Environment and Behaviour</i> 1:49-70.
411	Tavolga WN, Popper AN, Fay RR. 2012. Hearing and Sound Communication in Fishes. New
412	York: Springer Verlag.
413	Van Oppen MJH, Bongaerts P, Underwood JN, Peplow LM, Cooper TF. 2011. The role of
414	deep reefs in shallow reef recovery: an assessment of vertical connectivity in a brooding
415	coral from west and east Australia. Molecular Ecology 20:1647-1660.
416	Vermeij MJA, Marhaver KL, Huijbers CM, Nagelkerken I, Simpson SD. 2010. Coral larvae
417	move toward reef sounds. PLoS One 5(5):e10660.

Rogers PH, Cox M. 1988. Underwater sound as a biological stimulus. In: Atema J, Fay RR,