

Comparative study of Cd uptake and tolerance of two Italian ryegrass (*Lolium multiflorum*) cultivars

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Cadmium (Cd) is one of the most toxic heavy metals and is difficult to remove from contaminated soil and water. Italian ryegrass (*Lolium multiflorum*), as an energy crop, exhibits a valuable potential to develop Cd polluted sites due to its use as a biofuel rather than as food and forage. Previously, via a screening for Cd-tolerant ryegrass, the two most extreme cultivars (Idyll and Harukaze) with high and low Cd tolerance during seed germination, respectively, were selected. However, the underlying mechanism for Cd tolerance was not well investigated. In this study, we comparatively investigated the growth, physiological responses, and Cd uptake and translocation of Idyll and Harukaze when the seedlings were exposed to a Cd (0-100 μ M) solution for 12 days. As expected, excess Cd inhibited seedling growth and was accompanied by an accumulation of malondialdehyde (MDA) and reduced photosynthetic pigments in both cultivars. The effects of Cd on the uptake and translocation of other nutrient elements (Zn, Fe, Mn and Mg) were dependent on Cd concentrations, cultivars, plant tissues and elements. Compared with Harukaze, Idyll exhibited better performance with less MDA and higher pigment content. Furthermore, Idyll was less efficient in Cd uptake and translocation compared to Harukaze, which might be explained by the higher NPT (non-protein thiols) content in its roots. Taken together, our data indicate that Idyll is more tolerant than Harukaze, which partially resulted from the differences in Cd uptake and translocation.

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24 **ABSTRACT**

25 Cadmium (Cd) is one of the most toxic heavy metals and is difficult to remove from
 26 contaminated soil and water. Italian ryegrass (*Lolium multiflorum*), as an energy crop, exhibits a
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 37 Compared with Harukaze, IdyII exhibited better performance with less MDA and higher pigment
 38 content. Furthermore, IdyII was less efficient in Cd uptake and translocation compared to
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42 **Keywords** Cadmium, Italian ryegrass, physiological response, tolerance, uptake and
 43 translocation

44 **INTRODUCTION**

45 Heavy metal contamination in agricultural soil and water introduced by human activities poses a
 46 serious environment issue (Bonfranceschi, Flocco & Donati, 2009; Mwamba et al., 2016; Toth
 47 et al., 2016). Among heavy metals, cadmium (Cd), known as a highly toxic and non-
 48 biodegradable pollutant, is easily taken up by plant roots and translocated to other parts (Nocito

et al., 2011), thus threatening human health via food contamination and ecosystem safety (Agami & Mohamed 2013). In view of the difficulty of a clean up of Cd-contaminated soil by physical or chemical means, planting energy crops became a viable alternative for exploiting heavy-metal contaminated land (Shi & Cai 2009; Zhang *et al.*, 2013; Al Chami *et al.*, 2015; Pandey, Bajpai & Singh, 2016). Thus, increasing efforts have been devoted to isolating a tolerant cultivar and dissecting the mechanisms underlying their tolerance.

Several direct and indirect toxic effects caused by excess Cd have been well-documented. First, excess Cd generates free radicals and reactive oxygen species (ROS), which can oxidize proteins, DNA, lipids and carbohydrates, thus disturbing a number of physical and biological processes in plants (Belkhadi *et al.*, 2010; Fernandez *et al.*, 2013). For instance, excess Cd induced an accumulation of a lipid peroxidation product, MDA, in rice, which is an indicator of oxidative stress and disturbed cellular metabolism (Celekli, Kapi & Bozkurt, 2013; Mostofa, Seraj & Fujita, 2014; Xie *et al.*, 2015). Second, the uptake of nutrient elements (Fe, Mn, Cu and Zn) is disturbed by Cd stress, which can impair the transport of these elements from the roots to aerial parts, thus leading to a reduction of electron transport in photosystem I transport due to the lack of chlorophyll synthesis (Aravind & Prasad 2005; Lopes Júnior, Mazzafera & Arruda 2014).

Due to Cd toxicity in plants, it is not surprising that a set of strategies have been evolved to cope with exogenous Cd, which include active exclusion, vacuolar sequestration, retention in the roots, immobilization by cell walls and complexation by binding metal to low-molecular weight proteins (Ramos *et al.*, 2002; Wahid, Arshad & Farooq. 2010). It has been reported that reducing Cd accumulation by exclusion in the roots of *Thlaspi arvense* conferred enhanced tolerance in the Cd-tolerance ecotype (Martin *et al.*, 2012). In *Dittrichia viscosa* (L.) Greuter, the responses of Cd toxicity involved Cd retention in the cell wall of the roots and the upregulated contents of non-protein thiols and organic acids (Fernandez *et al.*, 2014). In wheat, Cd binds to the sulphur group of cysteine-forming Cd-PC complexes, reducing the free Cd²⁺ in the cytosol, and the Cd-PC complexes are in turn transported into the vacuole or out of the cell by ATP binding cassette transporters (Greger *et al.*, 2016).

Different approaches have been employed to unravel the mechanisms that address Cd toxicity, such as screening for cadmium sensitive mutants (*McHugh & Spanier 1994*) and dissecting the role of metal transporters by transgenic manipulations (*Uraguchi & Fujiwara 2012*). The cultivar screen is another way to find evolved mechanisms in plants due to different environments and natural variations. This has been conducted for Cd tolerance and accumulation in several species such as hemp (*Shi et al., 2012*), Indian mustard (*Gill, Khan & Tuteja, 2011*), barley (*Sghayar et al., 2014*) and castor (*Zhang et al., 2014*), revealing that Cd tolerance is related to the characteristics of plant morphology, the amounts of phytochrome synthesis, Cd uptake and thiol levels.

Italian ryegrass (*Lolium multiflorum*), also called annual ryegrass, is broadly grown in the south of China during the winter before the emergence of rice to relieve green fodder shortages (*Ye et al., 2015*). Recently, this species has been considered an appropriate material for bio-ethanol production due to its high ethanol conversion, rapid growth and low input costs (*Yasuda et al., 2015; Ye et al., 2015*). Two recent studies reported that Italian ryegrass had a high tolerance to Cd during seed germination and was able to be cultivated in sites polluted by mine tailings (*Liu et al., 2013; Mugica-Alvarez et al., 2015*). In regards to these properties, Italian ryegrass has been suggested as a new species for the bioremediation of heavy metal polluted soils (including Cd) (*Yamada et al., 2013*). However, compared with other species, little information is available concerning the capacity of Cd tolerance and uptake and tolerance mechanisms in Italian ryegrass. Here, through investigating the underlying causes for differential Cd tolerance in two ryegrass cultivars (IdyII and Harukaze), we found that IdyII is less efficient in Cd uptake and translocation than Harukaze. Furthermore, a high NPT content in IdyII might be one of causes for low Cd translocation. Our findings can provide a new tool for further dissecting the molecular mechanisms of Cd uptake and translocation in ryegrass cultivars and will be helpful for breeding Italian ryegrass as a bioenergy crop for heavy metal remediation.

MATERIALS AND METHODS

Plant cultivation

Two extreme ryegrass cultivars (Idyll and Harukaze) with high and low Cd tolerance during seed germination (Fang *et al.*, 2016), respectively, were selected for this study. Seeds were sterilized with 10% H₂O₂ for 10 min, rinsed thoroughly with distilled water, and germinated via immersion in distilled water at 25°C in the dark. After five days, uniform seedlings were transferred to 1-L plastic pots (14 plants per pot) filled with 1/4 Hoagland's solution. Seedlings were maintained for 10 days in a growth chamber at a 12 h light/dark cycle with 300 $\mu\text{mol m}^{-2} \text{s}^{-1}$ light intensity, a day/night temperature of 25/20°C and 65 \pm 5% relative humidity. Five Cd concentrations for the treatments were as follows: 0, 10 μM , 25 μM , 50 μM , and 100 μM Cd as CdCl₂·2.5H₂O (analytical reagent) was added to the nutrient solution. Each treatment had six replicates. The nutrient solution was renewed every 3 days, and the pH was adjusted to 6.5 with 2 M NaOH or 2.7 M HCl.

After a 12-day Cd exposure, plants were divided into two groups and collected. One group was for biomass and Cd concentration determination, and the other group was for physiological index measurements, including chlorophyll content, lipid peroxidation and non-protein thiols (NPT). Each group had three replicates.

Estimation of plant growth and Cd accumulation

The harvested plants were soaked in 20 mM Na₂-EDTA for 15 min, rinsed with distilled water to remove metals on the root surfaces, and separated into roots and shoots. Subsequently, samples were oven dried at 70°C to the constant weight. The dried samples were weighed and digested with mixed acid [HNO₃ + HClO₄ (85:15, v/v)]. The concentrations of Cd, zinc (Zn), iron (Fe), manganese (Mn) and magnesium (Mg) were determined by an inductively coupled plasma optical emission spectrometer (ICP-OES, Optima 2100DV, PerkinElmer, Inc. US).

The TIs (tolerance index), translocation factors (TFs), bioconcentration factors (BCF), and Cd accumulation were determined according to the method of Chen *et al.* (2011):

$$\text{TIs} = \text{biomass}_{\text{Cd}} / \text{biomass}_{\text{control}}$$

$$\text{TFs} = \text{Cd}_{\text{concentration in shoot}} / \text{Cd}_{\text{concentration in root}}$$

$$\text{BCF} = \text{Cd}_{\text{concentration in root}} / \text{Cd}_{\text{concentration in the nutrient solution}}$$

Cd accumulation = [biomass]_{dry weight} × [Cd]_{concentration in plant tissues}

Total Cd accumulation = Cd accumulation in root + Cd accumulation in shoot

Cadmium distribution proportion of root = Cd accumulation in root/ Total Cd accumulation

Estimation of photosynthetic pigment contents

The middle part of 100 mg of fresh leaves was extracted in 10 mL 95% ethanol for 24 h in the dark. Absorbance at 665, 649 and 470 nm were determined with spectrophotometry (SHIMADZU UV-2450, Kyoto, Japan). Chlorophyll a, b and carotenoids (Car) were calculated according to the method of *Knudson, Tibbitts & Edwards* (1977).

Estimation of lipid peroxidation

Plant fresh tissues (0.1-0.3 g) were homogenized and extracted in 5 mL of 0.25% TBA made in 10% trichloroacetic acid (TCA). The sample extractions were heated at 95°C for 30 min and then quickly cooled on ice. After centrifugation at 10000 × g for 10 min, the absorbance of the supernatant was measured at 532 and 600 nm. Lipid peroxidation and MDA content was estimated with the concentration of thiobarbituric acid-reacting substances (TBARS) as described by *Ali et al.* (2014).

Determination of non-protein thiols

Non-protein thiols (NPT) were assayed following *Tian et al.* (2011) with minor modifications. Fresh tissues (approximately 0.3 g) were homogenized in 3 mL ice-cold 5% sulfosalicylic acid solution and centrifuged at 12000 × g (4°C) for 15 min. The resulting supernatant was used for NPT assays. First, 0.3 mL of the supernatant was mixed with 1.2 mL 0.1 M K-phosphate buffer (pH 7.6) and 50 µL 6 mM 5,5'-dithiobis-2- nitrobenzoic acid (DTNB) (dissolved in 5 mM EDTA and 0.1 M phosphate buffer solution at pH 7.6). The mixture was incubated at room temperature for 20 min and then measured for absorbance at 412 nm with spectrophotometry. The NPT content was estimated with a standard curve of reduced glutathione in the range of 0-100 µg/mL.

Statistical analysis

Statistical analyses were performed using a two-way analysis of variance (ANOVA) with SPSS Version 20.0 (SPSS Inc., USA). Duncan's multiple range test was employed to compare the

changes among the different treatments at $P<0.05$. The relationships among TIs, Cd concentration, accumulation, and some physiology parameters in plant roots were determined by Pearson's correlation analysis.

RESULTS

Plant biomass, TIs and root/shoot ratio response to Cd stress

Increasing the Cd supply in the medium posed variable effects on plant biomass, TIs and the root/shoot ratio (Table 1). Cd treatments tended to reduce the biomass of both cultivars. The biomass of IdyII was significantly higher than that of Harukaze in the same treatment ($P<0.01$), and biomass reductions with increasing Cd dose from 5 μM to 100 μM were more distinct in Harukaze ($P<0.01$). Similar alterations were also observed with the TIs. For example, when exposed to 25 μM Cd, root biomass was reduced by 37% in Harukaze and 22% in IdyII. A clear decline in the roots/shoot ratio was demonstrated in both cultivars with increasing Cd concentrations.

Ecotoxicological response based on the plant biomass inhibition rate

As shown in Table 2, positive correlations were observed between the inhibition of plant biomass (root and shoot) and Cd concentrations in the solution ($P<0.01$), which was represented by the quadratic equation. To evaluate toxicity, the inhibitory concentration (EC50; Cd concentration when the root or shoot biomass decreased by 50% compared with the control) and lethal concentration (IC90; Cd concentration when the root or shoot biomass decreased by 90% compared with the control) were determined by the fitting equation. The EC50 values of the shoots and roots of IdyII were 1.9-fold and 3.2-fold higher than that of Harukaze, respectively. The lethal concentration of the roots and shoots (IC90) in IdyII were also higher than that of Harukaze, implying that IdyII was tolerant to Cd.

Cd concentration, accumulation in plant tissues, and BCFs and TFs responses to Cd stress

With elevating Cd concentrations in the treatment solutions, root Cd concentration in both cultivars increased, ranging from 2.09 to 9.89 $\text{mg}\cdot\text{g}^{-1}$ in Harukaze and from 1.26 to 6.89 $\text{mg}\cdot\text{g}^{-1}$ in IdyII. Obviously, Cd concentrations in Harukaze roots were higher than that of IdyII roots,

especially at the 50 to 100 μM Cd treatments (Fig. 1A, $P < 0.01$). Similar trends were also observed in the shoots (Fig. 1B). A gradual increase of Cd TFs in Harukaze was correlated with the increasing Cd concentration in the treatment solutions, whereas no significant Cd TFs changes in IdyII were observed. The Cd TFs value in Harukaze was significantly higher than that in IdyII after exposure to the highest Cd concentration (Fig. 1C, $P < 0.01$). Cd accumulation in the roots of Harukaze remained constant, while there was a dramatic increase in IdyII with the application of 25-100 μM Cd in the treatment solutions (Fig. 1D). An increasing trend of Cd accumulation in the shoots and total accumulation was observed with an increasing Cd supply, and their accumulation amounts in IdyII were significantly higher than that in Harukaze in the presence of higher Cd dosages (Figs. 1E and 1F, $P < 0.05$).

The proportion of cadmium distribution in the roots was 78.5-45.4% in Harukaze and 67.2-54.7% in IdyII under Cd stress. The distribution proportion in both Harukaze and IdyII significantly decreased with the 25 μM and 100 μM Cd treatments, respectively (Fig. 2A, $P < 0.01$). The increasing Cd supply reduced the root BCFs of the two cultivars (Fig. 2B), and the reduction was especially obvious in Harukaze (range from 3715 to 880). At low Cd treatments (5-10 μM), the BCFs of Harukaze were markedly higher than that of IdyII ($P < 0.01$).

Effects of Cd on plant mineral concentrations and TFs

Cadmium treatments altered the uptake and TFs of several nutrient elements (Fig. 3). Compared with the control, the 25-100 μM Cd supply markedly increased the Zn and Fe concentrations in the roots of Harukaze (Figs. 3A and 3B), whereas the Mn concentration in the roots of Harukaze was significantly decreased with lower Cd concentrations (5 and 10 μM) (Fig. 3C). Additionally, Mg concentration was greatly increased at the highest Cd concentrations (Fig. 3D). In contrast, compared with the control, 100 μM Cd significantly promoted Zn uptake in the roots of IdyII (Fig. 3A), while the uptakes of Fe, Mn, and Mg exhibited no change in the roots of IdyII (Figs. 3B, 3C and 3D). In the shoot, Cd supply did not affect Zn and Mg concentrations in both cultivars, but severely decreased Mn concentrations (Figs. 3E, 3H and 3G). Compared with the control, shoot Fe concentrations in IdyII exhibited a gentle decrease with increasing Cd

concentration, while a reduction occurred in Harukaze at the highest Cd treatment (Fig. 3F). In both cultivars, Zn TFs were significantly inhibited at the 25-100 μ M Cd treatments (Fig. 3I), and the amounts in IdyII were significantly higher than that of Harukaze. Fe TFs reached a maximum in both cultivars under 5 μ M Cd and then showed a decrease with increasing Cd (Fig. 3J). Cd treatments in IdyII significantly reduced Mn TFs and had no change in Mg TFs. In contrast, the TFs of Mn and Mg in Harukaze decreased considerably only at the highest concentration of Cd (Figs. 3K and 3L). According to a two-way ANOVA analysis, significant differences in the Zn, Fe, Mn ($P<0.01$) and Mg ($P<0.05$) concentrations of the roots were found between the two cultivars, as well as the Zn concentration of the shoots ($P<0.05$) and the TF of Zn ($P<0.01$).

Effects of Cd on pigment content, lipid peroxidation and NPT content

The cadmium supply tended to reduce chlorophyll a, chlorophyll b, chlorophyll (a + b) and Car contents in both cultivars (Table 3). For example, the chlorophyll a, chlorophyll b, and Car content decreased by 42.5%, 44.7% and 44.4% in Harukaze and by 11.8%, 5.9%, and 22.2% in IdyII under 25 μ M Cd stress, respectively. The chlorophyll a, chlorophyll b and chlorophyll (a+b) content of IdyII was significantly higher than that of Harukaze at 25-100 μ M Cd concentrations ($P<0.01$); a similar trend occurred in Car under 50-100 μ M Cd stress. The Cd treatments did not affect the Chl a/b ratio of IdyII but significantly inhibited that of Harukaze at 100 μ M Cd.

The MDA content in plant tissues was increased with elevated Cd concentrations, and the amounts in the leaves were higher than in the roots (Fig. 4A $P<0.01$). Compared with the control, when treated with high Cd concentrations (50 and 100 μ M), the MDA content of the seedling roots increased by 252.5% and 610.2% in Harukaze but only by 140.7% and 291.7% in IdyII, respectively. Similarly, the MDA content of the leaves in Harukaze increased sharply with the 25-100 μ M Cd treatments and was considerably higher than that of IdyII (Fig. 4B, $P<0.01$).

As shown in Fig. 5, compared with the control, the NPT content in the roots of IdyII increased at least 3 times with the application of 5 μ M Cd and reached its maximum under 50 μ M Cd. No significant change occurred in Harukaze with the addition of 5-50 μ M Cd in the medium (Fig. 5A). Furthermore, the NPT content in the roots of IdyII were significantly higher than that of

Harukaze with the treatments of 10-100 μM Cd ($P<0.01$). The NPT content in the shoots of the two cultivars were also enhanced under Cd stress. The values in IdyII were significantly higher than that in Harukaze at the 25-100 μM Cd treatments (Fig. 5B, $P<0.05$).

Correlation analysis

Pearson's correlation analysis was carried out to investigate the correlations among TIs, Cd uptake, Cd TFs, Cd accumulation, MDA content and NPT content of the two cultivars in the roots (Table 4). TIs were negatively correlated to Cd concentration, MDA content and Cd TFs. Cd uptake and TFs were positively correlated to MDA content. Additionally, a positive correlation was observed between Cd accumulation and NPT content.

DISCUSSION

In this study, our data demonstrated that Italian ryegrass possesses a stronger capacity in Cd uptake than common crops. After 12 days of exposure at 50 μM Cd, the Cd concentration in Italian ryegrass reached at least 4.9 $\text{mg}\cdot\text{g}^{-1}$ (DW) in the roots and 0.4 $\text{mg}\cdot\text{g}^{-1}$ (DW) in the shoots (Fig. 1A and B). These concentrations are higher than that in rice after 15 days of exposure at 50 μM Cd (Lin *et al.*, 2012), as well as that in maize after 15 days of exposure at 100 μM Cd (Wang *et al.*, 2007).

The biomass reduction in the roots was more visible than that in the shoots when Cd levels were over 10 μM (Table 1). Similar results were reported in barley (Tiryakioglu *et al.*, 2006), and the reason may be the fact that the roots are directly exposed to Cd (Hegedüs, Erdei & Horváth 2001). As plant biomass and TIs are two important parameters to evaluate the Cd tolerance in plants (Metwally *et al.*, 2005; Shi *et al.*, 2012), the biomass of the roots and shoot in Harukaze were reduced over 50% under 50 and 100 μM Cd, and the TIs were less than 0.5, which was not observed in IdyII (Table 1), thus demonstrating that IdyII was more tolerant to Cd than Harukaze. This was further supported by the higher EC50 and IC90 of Cd toxicity in IdyII (Table 2); two parameters commonly represent phytotoxin under a threshold and acute toxicity, respectively (Paschke, Valdecantos & Redente, 2005; An, 2006; Pannacci, Pettorossi & Tei 2013).

Although Cd is a non-redox metal unable to produce reactive oxygen species (ROS) through single electron transfer, Cd interferes with the antioxidant defence system and diminishes the capacity for ROS removal (*Wahid, Arshad & Farooq, 2010*). Cd also affects the functions of two important organelles, the mitochondria and chloroplasts, which in turn disturb their electron transfers and generate free radicals and ROS in the cell (*Celekli, Kapi & Bozkurt, 2013; Mostofa, Seraj & Fujita, 2014*). The accumulated ROS can interact with proteins, lipids, carbohydrates, and DNA, perturbing a number of physiological processes (*Gallego et al., 2012*). In IdyII and Harukaze, the Cd supply enhanced the MDA content, indicating Cd induced oxidative damage (Fig. 4). The oxidative damage might partially be attributed to the reduction in photosynthetic pigments and plant biomass under Cd stress in both cultivars (Table 1 and Table 3). Relatively high MDA content and low photosynthetic pigments demonstrated that Cd-induced toxicity in Harukaze was more severe than in IdyII, which was consistent with their Cd tolerance. Similar correlations between Cd tolerance and MDA content were observed in oilseed cultivars (*Wu et al. 2015*), the leaves of Indian mustard cultivars (*Gill, Khan & Tuteja, 2011*) and Artichoke cultivars (*Chen et al., 2011*).

Apart from oxidative damage, the uptake, transport, and subsequent distribution of nutrient elements in IdyII and Harukaze were affected by the presence of Cd (Fig. 3). An elevated Cd dosage increased Zn, Fe, Mg and Mn concentrations in Harukaze roots, whereas it did not significantly increase that in the IdyII roots, except for Zn (Fig. 3A-D). Possibly, the metal transportation systems in the roots are different between Harukaze and IdyII, and Harukaze may have a high-dose Cd activated transportation system. This is further supported by the higher Cd concentrations in Harukaze. Likewise, Cd promoting the uptake of Mg, Ca, and Fe were reported in tomato (*Kisa, Ozturk & Tekin 2016*). No significant differences in metal concentration were observed between the shoots of Harukaze and IdyII, indicating that cultivar differences in metal uptake are mainly in the roots rather than in the shoots. *Goncalves et al. (2009)* also suggested that microelement uptake, such as Fe^{2+} , Mn^{2+} and Zn^{2+} , was determined by the level of Cd in the substrate, cultivar and plant tissue specificity in potato (*Solanum tuberosum*). Several metal

transporters have been identified that translocate nutrient elements from the roots to the shoot, such as NRAMP families and ZIP families (*Choppala et al., 2014*). With exposure to Cd, the Cd TF remained constant or increased with Cd treatments (Fig. 1C), whereas the TFs of Zn, Fe and Mn exhibited a decline (Fig. 3I, 3J and 3K), indicating that there may be possible competition with the metal transporters for translocation between Cd and other micronutrients in Italian ryegrass. It was reported that there were antagonistic effects from Cd and microelement elements (Zn, Fe, Mn) using the same transporters and/or cation channels as Ca and Mg (*Sarwar et al., 2010; Kisa, Ozturk & Tekin 2016*).

Non-protein thiols (NPT), including glutathione, thiol-rich peptides and other SH groups, play an important role in defence response against the detoxification of heavy metals in plants (*Ozdener & Aydin 2009; Nadgorska-Socha et al., 2013*). In our study, Cd concentration in the roots and shoots of Harukaze were significantly higher than that of IdyII, whose tendency was the opposite of NPT content (Figs. 1A, 1B and Fig. 5). NPT are essential for the synthesis of Cd-binding peptides such as phytochelatins, which inactivate and sequester Cd by forming stable Cd-complexes in the vacuole (*Cobbett 2000*). The high NPT in IdyII may promote Cd sequestration into the vacuole and block its translocation, thus leading to the decline of Cd concentration in the shoots and the Cd TF. A similar phenomenon was observed in the variation of Cd tolerance among cultivars of cabbage and barley, suggesting that NPT content may be an important indicator for Cd tolerance (*Tiryakioglu et al., 2006; Sun et al., 2013*). NPT triggering sequestration was also for other metal elements, except for Cd, and the higher NPT content in IdyII might be contributed to the decreased translocation of Zn, Fe, Mg and Mn from the roots to the shoots.

CONCLUSIONS

In the present study, the biomass, Cd uptake, translocation, accumulation, and physiology parameters of two Italian ryegrass cultivars were significantly affected by Cd treatments. Compared with Harukaze, IdyII is a Cd-tolerant cultivar, exhibiting a low Cd uptake and a high NPT content. These two distinct capacities may be the major physiological changes that

contributed to the difference of Cd tolerance between the two cultivars. Taken together, our data demonstrates that IdyII is more tolerant than Harukaze, which is correlated with low Cd uptake and high NPT content. This will be helpful in investigating the molecular mechanisms of Cd uptake and translocation in Italian ryegrass.

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REFERENCES

- Agami RA, Mohamed GF. 2013.** Exogenous treatment with indole-3-acetic acid and salicylic acid alleviates cadmium toxicity in wheat seedlings. *Ecotoxicology and Environmental Safety* **94**:164-171 DOI 10.1016/j.ecoenv.2013.04.013.
- Al Chami Z, Amer N, Al Bitar L, Cavoški I. 2015.** Potential use of Sorghum bicolor and Carthamus tinctorius in phytoremediation of nickel, lead and zinc. *International Journal of Environmental Science and Technology* **12**:3957-3970 DOI 10.1007/s13762-015-0823-0.
- Ali B, Gill RA, Yang S, Gill MB, Ali S, Rafiq MT, Zhou W. 2014.** Hydrogen sulfide alleviates cadmium-induced morpho-physiological and ultrastructural changes in Brassica napus. *Ecotoxicology and Environmental Safety* **110**:197-207 DOI 10.1016/j.ecoenv.2014.08.027.
- An YJ. 2006.** Assessment of comparative toxicities of lead and copper using plant assay. *Chemosphere* **62**:1359-1365 DOI 10.1016/j.chemosphere.2005.07.044.
- Aravind P, Prasad MNV. 2005.** Cadmium-Zinc interactions in a hydroponic system using *Ceratophyllum demersum* L: adaptive ecophysiology, biochemistry and molecular toxicology. *Brazilian Journal of Plant Physiology* **17**:3-20 DOI 10.1590/S1677-

04202005000100002.

- Belkhadi A, Hediji H, Abbes Z, Nouairi I, Barhoumi Z, Zarrouk M, Chaibi W, Djebali W. 2010.** Effects of exogenous salicylic acid pre-treatment on cadmium toxicity and leaf lipid content in *Linum usitatissimum* L. *Ecotoxicology and Environmental Safety* **73**:1004-1011 DOI 10.1016/j.ecoenv.2010.03.009.
- Bonfranceschi BA, Flocco CG, Donati ER. 2009.** Study of the heavy metal phytoextraction capacity of two forage species growing in an hydroponic environment. *Journal of Hazardous Materials* **165**:366-371 DOI 10.1016/j.jhazmat.2008.10.024.
- Celekli A, Kapi M, Bozkurt H. 2013.** Effect of cadmium on biomass, pigmentation, malondialdehyde, and proline of *Scenedesmus quadricauda* var. *longispina*. *Bulletin of Environmental Contamination and Toxicology* **91**:571-576 DOI 10.1007/s00128-013-1100-x.
- Chen L, Long X-H, Zhang Z-H, Zheng X-T, Rengel Z, Liu Z-P. 2011.** Cadmium Accumulation and Translocation in Two Jerusalem Artichoke (*Helianthus tuberosus* L.) Cultivars. *Pedosphere* **21**:573-580 DOI 10.1016/s1002-0160(11)60159-8.
- Choppala G, Saifullah, Bolan N, Bibi S, Iqbal M, Rengel Z, Kunhikrishnan A, Ashwath N, Ok YS. 2014.** Cellular Mechanisms in Higher Plants Governing Tolerance to Cadmium Toxicity. *Critical Reviews in Plant Sciences* **33**:374-391 DOI 10.1080/07352689.2014.903747.
- Cobbett CS. 2000.** Phytochelatins and their roles in heavy metal detoxification. *Plant Physiology* **123**:825-832 DOI <http://dx.doi.org/10.1104/pp.123.3.825>
- Fang ZG, Hu ZY, Zhao HH, Yang L, Ding CL, Lou LQ, Cai QS. 2016.** Screening for cadmium tolerance of 21 cultivars from Italian ryegrass (*Lolium multiflorum* Lam) during germination. *Grassland Science* (First published: 24 January 2017) Available at <http://onlinelibrary.wiley.com/doi/10.1111/grs.12138/full>
- Fernandez R, Bertrand A, Reis R, Mourato MP, Martins LL, Gonzalez A. 2013.** Growth and physiological responses to cadmium stress of two populations of *Dittrichia viscosa*

- (L.) Greuter. *Journal of Hazardous Materials* **244**:555-562
DOI 10.1016/j.jhazmat.2012.10.044.
- Fernandez R, Fernandez-Fuego D, Bertrand A, Gonzalez A. 2014.** Strategies for Cd accumulation in *Dittrichia viscosa* (L.) Greuter: role of the cell wall, non-protein thiols and organic acids. *Plant Physiol Biochem* **78**:63-70 DOI 10.1016/j.plaphy.2014.02.021.
- Gallego SM, Pena LB, Barcia RA, Azpilicueta CE, Iannone MF, Rosales EP, Zawoznik MS, Groppa MD, Benavides MP. 2012.** Unravelling cadmium toxicity and tolerance in plants: Insight into regulatory mechanisms. *Environmental and Experimental Botany* **83**:33-46 DOI 10.1016/j.envexpbot.2012.04.006.
- Gill SS, Khan NA, Tuteja N. 2011.** Differential cadmium stress tolerance in five Indian mustard (*Brassica juncea* L.) cultivars. *Plant Signalling and Behaviour* **6**:293-300 DOI 10.4161/psb.6.2.15049.
- Goncalves JF, Antes FG, Maldaner J, Pereira LB, Tabaldi LA, Rauber R, Rossato LV, Bisognin DA, Dressler VL, Flores EM, Nicoloso FT. 2009.** Cadmium and mineral nutrient accumulation in potato plantlets grown under cadmium stress in two different experimental culture conditions. *Plant Physiology and Biochemistry* **47**:814-821 DOI 10.1016/j.plaphy.2009.04.002.
- Greger M, Kabir AH, Landberg T, Maity PJ, Lindberg S. 2016.** Silicate reduces cadmium uptake into cells of wheat. *Environmental Pollution* **211**:90-97 DOI 10.1016/j.envpol.2015.12.027.
- Hegedüs A, Erdei S, Horváth G. 2001.** Comparative studies of H₂O₂ detoxifying enzymes in green and greening barley seedlings under cadmium stress. *Plant Science* **160**:1085-1093 DOI 10.1016/S0168-9452(01)00330-2.
- Kisa D, Ozturk L, Tekin S. 2016.** Gene expression analysis of metallothionein and mineral elements uptake in tomato (*Solanum lycopersicum*) exposed to cadmium. *Journal of Plant Research* **129**:989-995 DOI 10.1007/s10265-016-0847-7.
- Knudson LL, Tibbitts T W, Edwards GE 1977.** Measurement of Ozone Injury by

- 400 Determination of Leaf chlorophyll concentration. *Plant Physiology* **60**:606-608 DOI 10.
401 1104/pp.60.4.606.
- 402 **Lin L, Zhou W, Dai H, Cao F, Zhang G, Wu F. 2012.** Selenium reduces cadmium uptake and
403 mitigates cadmium toxicity in rice. *Journal of Hazardous Materials* 235-236:343-351
404 DOI 10.1016/j.jhazmat.2012.08.012.
- 405 **Liu Z, He X, Chen W, Zhao M. 2013.** Ecotoxicological responses of three ornamental herb
406 species to cadmium. *Environmental Toxicology and Chemistry* **32**:1746-1751 DOI
407 10.1002/etc.2237.
- 408 **Lopes Júnior CA, Mazzafera P, Arruda MAZ. 2014.** A comparative ionomic approach
409 focusing on cadmium effects in sunflowers (*Helianthus annuus* L.). *Environmental and*
410 *Experimental Botany* **107**:180-186 DOI 10.1016/j.envexpbot.2014.06.002.
- 411 **Martin SR, Llugany M, Barceló J, Poschenrieder C. 2012.** Cadmium exclusion a key factor in
412 differential Cd-resistance in *Thlaspi arvense* ecotypes. *Biologia Plantarum* **56**:729-734
413 DOI 10.1007/s10535-012-0056-8
- 414 **McHugh JP, Spanier JG. 1994.** Isolation of cadmium sensitive mutants in *Chlamydomonas*
415 *reinhardtii* by transformation/insertional mutagenesis. *FEMS microbiology letters*
416 **124**:239-244 DOI <https://doi.org/10.1111/j.1574-6968.1994.tb07290.x>.
- 417 **Metwally A, Safronova VI, Belimov AA, Dietz KJ. 2005.** Genotypic variation of the response
418 to cadmium toxicity in *Pisum sativum* L. *Journal of Experimental Botany* **56**:167-178
419 DOI 10.1093/jxb/eri017.
- 420 **Mostofa MG, Seraj ZI, Fujita M. 2014.** Exogenous sodium nitroprusside and glutathione
421 alleviate copper toxicity by reducing copper uptake and oxidative damage in rice (*Oryza*
422 *sativa* L.) seedlings. *Protoplasma* **251**:1373-1386
423 DOI 10.1007/s00709-014-0639-7.
- 424 **Mugica-Alvarez V, Cortés-Jiménez V, Vaca-Mier M, Domínguez-Soria V. 2015.**
425 Phytoremediation of Mine Tailings Using *Lolium Multiflorum*. *International Journal of*
426 *Environmental Science and Development* **6**:246-251 DOI 10.7763/ijesd.2015.v6.599.

Mwamba TM, Li L, Gill RA, Islam F, Nawaz A, Ali B, Farooq MA, Lwalaba JL, Zhou W.

2016. Differential subcellular distribution and chemical forms of cadmium and copper in *Brassica napus*. *Ecotoxicology and Environmental Safety* **134**:239-249 DOI 10.1016/j.ecoenv.2016.08.021.

Nadgorska-Socha A, Kafel A, Kandziora-Ciupa M, Gospodarek J, Zawisza-Raszka A. 2013.

Accumulation of heavy metals and antioxidant responses in *Vicia faba* plants grown on monometallic contaminated soil. *Environmental science and pollution research international* **20**:1124-1134 DOI 10.1007/s11356-012-1191-7.

Nocito FF, Lancilli C, Dendena B, Lucchini G, Sacchi GA. 2011. Cadmium retention in rice

roots is influenced by cadmium availability, chelation and translocation. *Plant Cell Environment* **34**:994-1008 DOI 10.1111/j.1365-3040.2011.02299.x.

Ozdener Y, Aydin BK. 2009. The effect of zinc on the growth and physiological and

biochemical parameters in seedlings of *Eruca sativa* (L.) (Rocket). *Acta Physiologiae Plantarum* **32**:469-476 DOI 10.1007/s11738-009-0423-z.

Pannacci E, Pettorossi D, Tei F. 2013. Phytotoxic effects of aqueous extracts of sunflower on

seed germination and growth of *Sinapis alba* L., *Triticum aestivum* L. and *Lolium multiflorum* Lam. *Allelopathy Journal* **32**:23 Available at

<http://search.proquest.com/docview/1458282973?accountid=43630>

Pandey VC, Bajpai O, Singh N. 2016. Energy crops in sustainable phytoremediation.

Renewable and Sustainable Energy Reviews **54**:58-73

DOI 10.1016/j.rser.2015.09.078.

Paschke MW, Valdecantos A, Redente EF. 2005. Manganese toxicity thresholds for

restoration grass species. *Environmental Pollution* **135**:313-322

DOI 10.1016/j.envpol.2004.08.006.

Ramos I, Esteban E, Lucena JJ, Gárate An. 2002. Cadmium uptake and subcellular

distribution in plants of *Lactuca sp.* Cd–Mn interaction. *Plant Science* **162**:761-767 DOI

10.1016/S0168-9452(02)00017-1.

- 454 **Sarwar N, Malhi SS, Zia MH, Naeem A, Bibi S, Farid G. 2010.** Role of mineral nutrition in
455 minimizing cadmium accumulation by plants. *Journal of the Science of Food and*
456 *Agriculture* **90**:925-937 DOI 10.1002/jsfa.3916
- 457 **Sghayar S, Ferri A, Lancilli C, Lucchini G, Abruzzese A, Porrini M, Ghnaya T, Nocito FF,**
458 **Abdelly C, Sacchi GA. 2014.** Analysis of cadmium translocation, partitioning and
459 tolerance in six barley (*Hordeum vulgare L.*) cultivars as a function of thiol metabolism.
460 **3**:311-320 *Biology and Fertility of Soils*
461 DOI 10.1007/s00374-014-0977-9.
- 462 **Shi G, Cai Q. 2009.** Cadmium tolerance and accumulation in eight potential energy crops.
463 *Biotechnology Advance* **27**:555-561 DOI 10.1016/j.biotechadv.2009.04.006.
- 464 **Shi G, Liu C, Cui M, Ma Y, Cai Q. 2012.** Cadmium tolerance and bioaccumulation of 18 hemp
465 accessions. *Applied Biochemistry Biotechnology* **168**:163-173
466 DOI 10.1007/s12010-011-9382-0.
- 467 **Sun J, Cui J, Luo C, Gao L, Chen Y, Shen Z. 2013.** Contribution of cell walls, nonprotein
468 thiols, and organic acids to cadmium resistance in two *cabbage* varieties. *Archives of*
469 *Environment Contamination and Toxicology* **64**:243-252
470 DOI 10.1007/s00244-012-9824-x.
- 471 **Tian S, Lu L, Zhang J, Wang K, Brown P, He Z, Liang J, Yang X. 2011.** Calcium protects
472 roots of *Sedum alfredii* H. against cadmium-induced oxidative stress. *Chemosphere*
473 **84**:63-69 DOI 10.1016/j.chemosphere.2011.02.054.
- 474 **Tiryakioglu M, Eker S, Ozkutlu F, Husted S, Cakmak I. 2006.** Antioxidant defence system
475 and cadmium uptake in *barley* genotypes differing in cadmium tolerance. *Journal of*
476 *Trace Elements in Medicine and Biology* **20**:181-189
477 DOI 10.1016/j.jtemb.2005.12.004.
- 478 **Toth G, Hermann T, Da Silva MR, and Montanarella L. 2016.** Heavy metals in agricultural
479 soils of the European Union with implications for food safety. *Environment International*
480 **88**:299-309 DOI 10.1016/j.envint.2015.12.017.

- 481 **Uraguchi S, Fujiwara T. 2012.** Cadmium transport and tolerance in rice: perspectives for
482 reducing grain cadmium accumulation. *Rice (N Y)* 5:5 DOI 10.1186/1939-8433-5-5.
- 483 **Wahid A, Arshad M, Farooq M. 2010.** Cadmium phytotoxicity: responses, mechanisms and
484 mitigation strategies: a review. *Organic Farming, Pest Control and Remediation of Soil*
485 *Pollutants*: Springer, 371-403.
- 486 **Wang M, Zou J, Duan X, Jiang W, Liu D. 2007.** Cadmium accumulation and its effects on
487 metal uptake in maize (*Zea mays* L.). *Bioresource Technology* **98**:82-88 DOI
488 10.1016/j.biortech.2005.11.028.
- 489 **Wu Z, Zhao X, Sun X, Tan Q, Tang Y, Nie Z, Qu C, Chen Z, Hu C. 2015.** Antioxidant
490 enzyme systems and the ascorbate-glutathione cycle as contributing factors to cadmium
491 accumulation and tolerance in two oilseed rape cultivars (*Brassica napus* L.) under
492 moderate cadmium stress. *Chemosphere* **138**:526-536
493 DOI 10.1016/j.chemosphere.2015.06.080.
- 494 **Xie P-p, Deng J-w, Zhang H-m, Ma Y-h, Cao D-j, Ma R-x, Liu R-j, Liu C, Liang Y-g. 2015.**
495 Effects of cadmium on bioaccumulation and biochemical stress response in rice (*Oryza*
496 *sativa* L.). *Ecotoxicology and Environmental Safety* **122**:392-398 DOI.
- 497 **Yamada T, Inoue M, Stewart AV, Cai H, Saha MC, Rao KM, Rognli OA, Ellison NW,**
498 **Bushman BS, Amundsen K. 2013.** Genetics, genomics and breeding of forage crops.
499 CRC Press, Boca Rato. 53-54.
- 500 **Yasuda M, Takenouchi Y, Nitta Y, Ishii Y, Ohta K. 2015.** Italian ryegrass (*Lolium*
501 *multiflorum* Lam) as a High-Potential Bio-Ethanol Resource. *Bioenergy Research*
502 **8**:1303-1309. DOI 10.1007/s12155-015-9582-5.
- 503 **Ye S, Yang Y, Xin G, Wang Y, Ruan L, Ye G. 2015.** Studies of the Italian ryegrass–rice
504 rotation system in southern China: Arbuscular mycorrhizal symbiosis affects soil
505 microorganisms and enzyme activities in the *Lolium mutiflorum* L. rhizosphere. *Applied*
506 *Soil Ecology* **90**:26-34 DOI 10.1016/j.apsoil.2015.01.017.
- 507 **Zhang H, Guo Q, Yang J, Chen T, Zhu G, Peters M, Wei R, Tian L, Wang C, Tan D, Ma J,**

- 508 **Wang G, Wan Y. 2014.** Cadmium accumulation and tolerance of two castor cultivars in
509 relation to antioxidant systems. *Journal of Environment Science* **26**:2048-2055 DOI
510 10.1016/j.jes.2014.08.005.
- 511 **Zhang H, Zhang X, Li T, Fu H. 2013.** Variation of cadmium uptake, translocation among *rice*
512 lines and detecting for potential cadmium-safe cultivars. *Environmental Earth Sciences*
513 **71**:277-286 DOI 10.1007/s12665-013-2431-y.

Figure 1

Cadmium concentration, translocation factors (TFs) and cadmium accumulation in two cultivars of Italian ryegrass

Data are means \pm SE. (n = 3) of three replicates. * $P < 0.05$. ** $P < 0.01$. ns, not significant. Different letters indicate significant differences at $P < 0.05$ according to the Duncan's test multiple range.

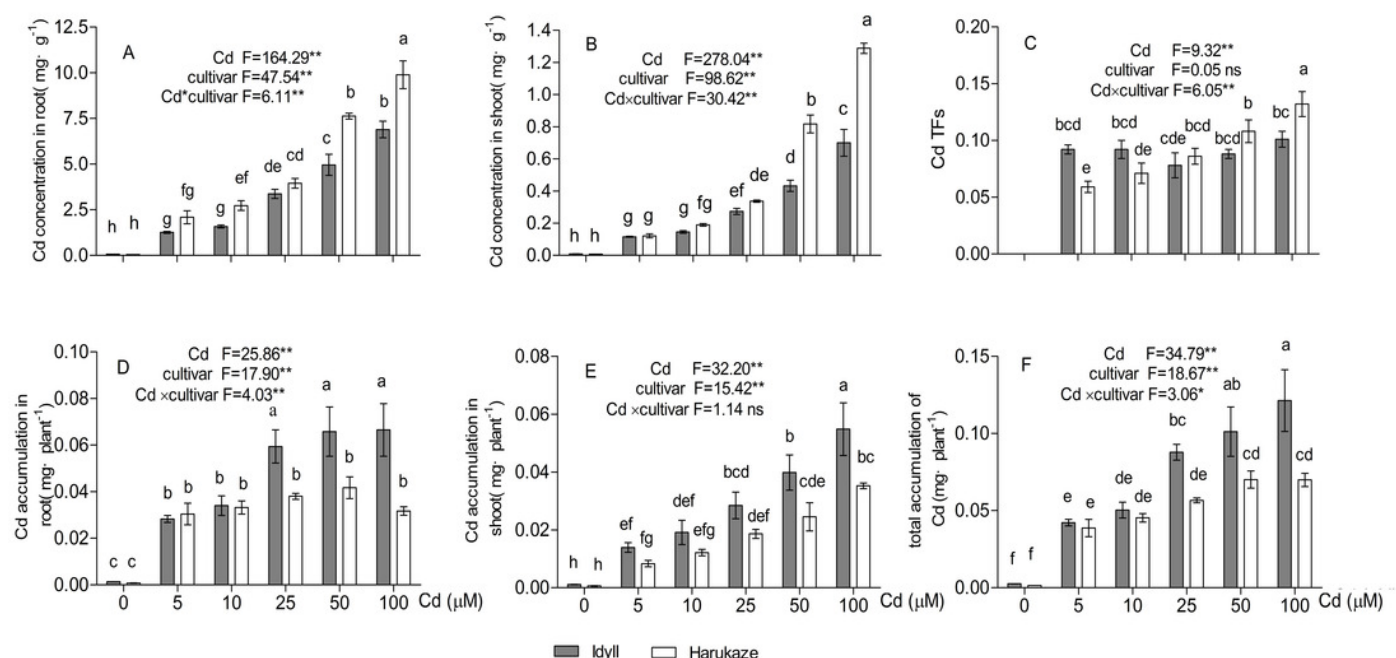


Figure 2

The distribution proportion of Cd (A) and Cd bioconcentration factors (B) in roots under Cd stress

Data are means \pm SE. (n = 3) of three replicates. * $P < 0.05$. ** $P < 0.01$. ns, not significant. Different letters indicate significant differences at $P < 0.05$ according to the Duncan's test multiple range.

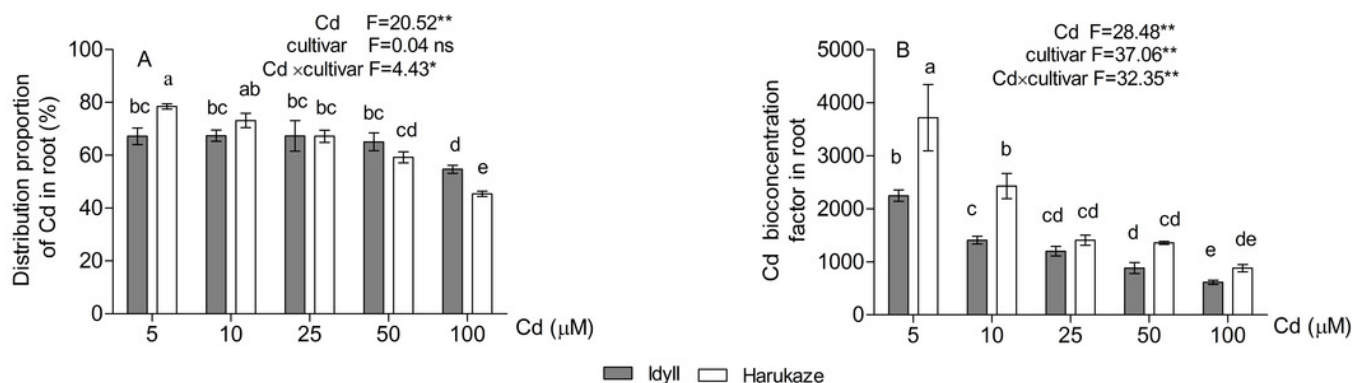


Figure 3

Nutrient element concentrations in the roots, shoot and TFs in the presence of Cd

Data are means \pm SE. (n = 3) of three replicates. * $P < 0.05$. ** $P < 0.01$. ns, not significant. Different letters indicate significant differences at $P < 0.05$ according to the Duncan's test multiple range.

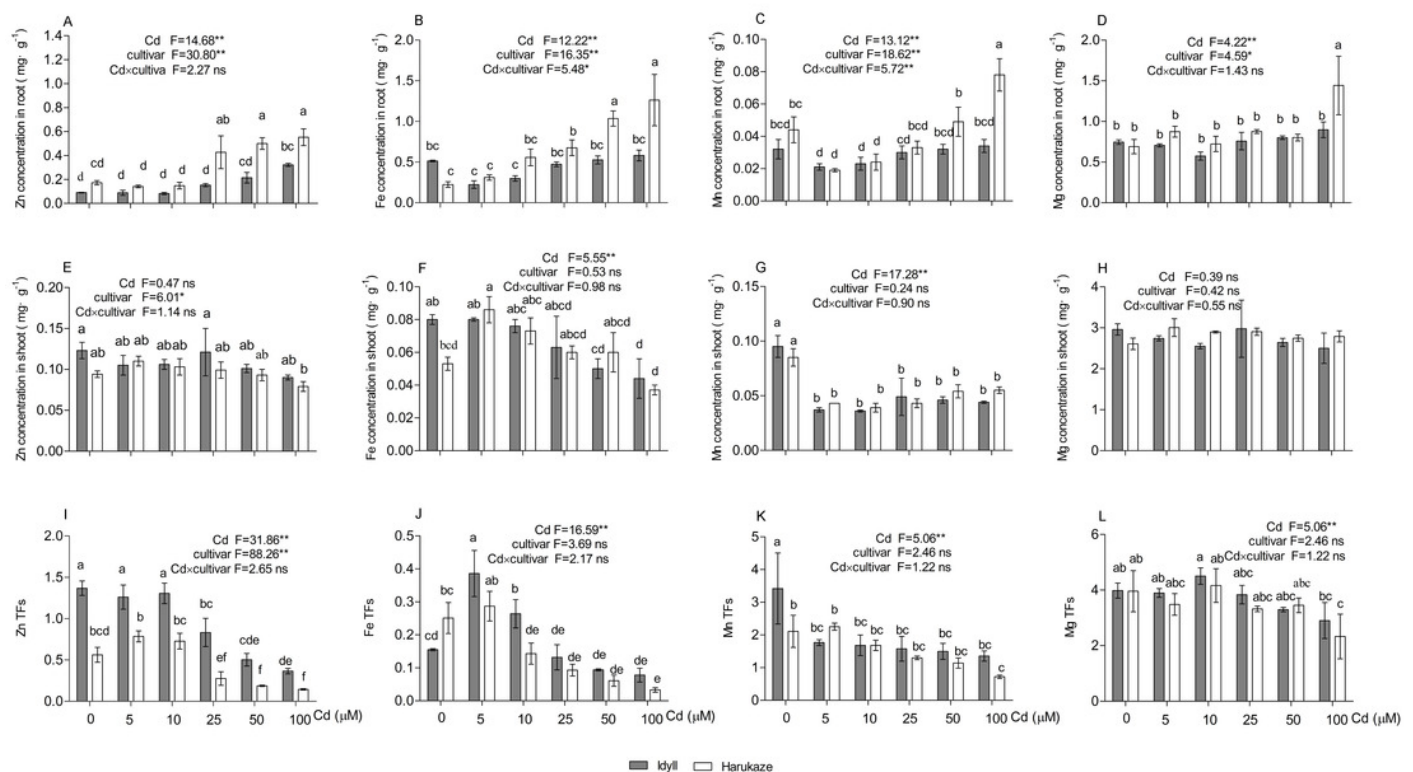


Figure 4

Effects of Cd on malondialdehyde (MDA) content in the roots and shoots of two Italian ryegrass cultivars

Data are means \pm SE. (n = 3) of three replicates. * $P < 0.05$. ** $P < 0.01$. ns, not significant. Different letters indicate significant differences at $P < 0.05$ according to the Duncan's test multiple range.

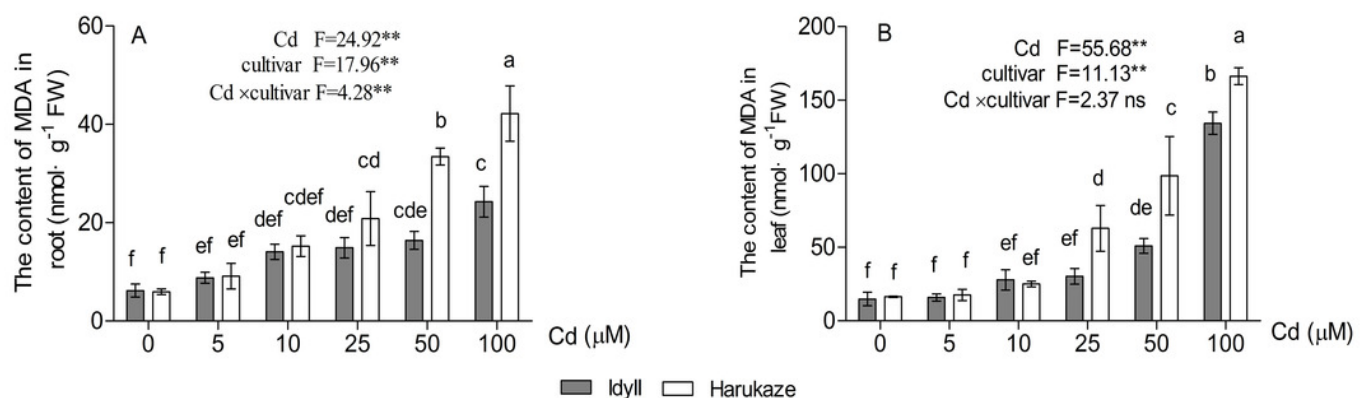


Figure 5

Effects of Cd on non-protein thiols (NPT) content in the roots and shoots of two Italian ryegrass cultivars

Data are means \pm SE.(n = 3) of three replicates. * P <0.05.** P <0.01. ns, not significant. Different letters indicate significant differences at P <0.05 according to the Duncan's test multiple range.

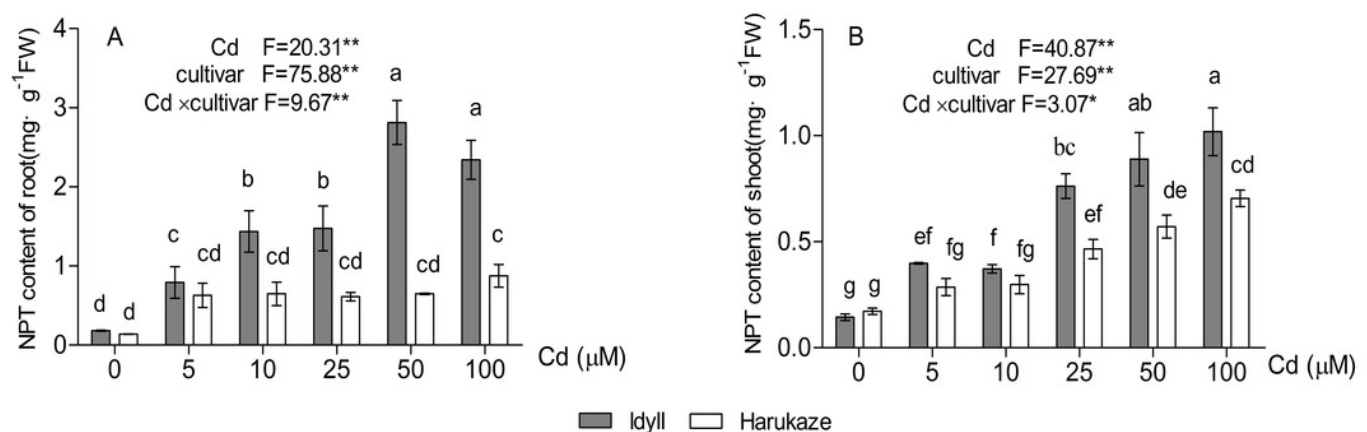


Table 1(on next page)

Effects of Cd on plant biomass, Cd tolerance, and the root/shoot ratio in two cultivars of Italian ryegrass

DW indicates dry weight; TI, tolerance index; R, root; S, shoot. Values (means \pm S.E., n=3) followed by different letters in the same columns are significantly different according to Duncan's test. * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$, n.s., not significant.

Table 1 Effects of Cd on biomass, Cd tolerance, and root/shoot in two cultivars of Italian ryegrass

cultivar	Cd supply μM	Plant biomass ($\text{mg} \cdot \text{plant}^{-1} \text{ DW}$)		TIs		Root/Shoot ratio
		R	S	R	S	
Idyll	0	22.51 \pm 2.77 a	123.98 \pm 10.40 a			0.18 \pm 0.012 a
	5	22.60 \pm 1.94 a	120.19 \pm 12.04 a	1.00 \pm 0.09 a	0.97 \pm 0.10 a	0.19 \pm 0.031 a
	10	21.52 \pm 2.33 ab	113.92 \pm 14.28 a	0.96 \pm 0.10 ab	0.92 \pm 0.12 ab	0.19 \pm 0.005 a
	25	17.54 \pm 1.09 bc	103.17 \pm 12.63 ab	0.78 \pm 0.05 bc	0.83 \pm 0.10 abc	0.18 \pm 0.036 a
	50	13.30 \pm 1.57 cde	80.48 \pm 8.41 bcd	0.59 \pm 0.07 d	0.65 \pm 0.07 c	0.16 \pm 0.003 ab
	100	9.26 \pm 0.99 ef	78.08 \pm 7.86 bcd	0.41 \pm 0.04 e	0.63 \pm 0.13 cd	0.12 \pm 0.014 b
Harukaze	0	15.29 \pm 0.99 cd	84.00 \pm 5.05 bc			0.18 \pm 0.023 a
	5	14.64 \pm 0.76 cd	67.99 \pm 2.92 cd	0.96 \pm 0.05 ab	0.81 \pm 0.03 bc	0.22 \pm 0.006 a
	10	12.26 \pm 0.90 de	63.36 \pm 3.64 cd	0.80 \pm 0.06 bc	0.75 \pm 0.04 bc	0.19 \pm 0.011 a
	25	9.69 \pm 0.70 ef	55.06 \pm 3.74 de	0.63 \pm 0.05 cd	0.65 \pm 0.04 c	0.18 \pm 0.010 a
	50	5.46 \pm 0.58 fg	35.20 \pm 3.62 ef	0.36 \pm 0.04 ef	0.42 \pm 0.04 de	0.16 \pm 0.017 ab
ANOVA	100	3.22 \pm 0.13 g	29.55 \pm 1.43 f	0.21 \pm 0.01 f	0.35 \pm 0.02 e	0.11 \pm 0.007 b
	Cd	153.72**	11.04**	38.32**	11.87**	5.97*
	cultivar	25.79**	92.04**	15.15**	18.93**	0.017n.s
	Cd \times cultivar	0.268n.s	0.13n.s	0.63n.s	0.24n.s	0.194n.s

Table 2 (on next page)

Fitted equations of Cd concentration and the inhibition rate of root or shoot biomass

x is Cd concentration, and y is inhabitation of root or shoot biomass. EC50 indicates an effective Cd concentration (when the root or shoot biomass decreased by 50% compared with the control), and IC90 indicates a lethal concentration (when the root or shoot biomass decreased by 90% compared with the control). ** indicates $P < 0.01$.

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Table 2 fitted equations of Cd concentration and inhibition rate of root or shoot biomass

cultivar	Fit equations between Cd	EC ₅₀	IC ₉₀	R ²	Fit equations between Cd	EC ₅₀	IC ₉₀	R ²
	concentration and inhibition rate of root biomass	(μM)	(μM)		concentration and inhibition rate of shoot biomass	(μM)	(μM)	
Idyll	$y = -0.0049x^2 + 1.1198x - 3.612$	68.27	192.03	0.991**	$y = -0.006x^2 + 0.9837x - 1.2897$	124.73	174.14	0.989**
Harukaze	$y = -0.01x^2 + 1.7983x - 0.8778$	35.17	121.58	0.994**	$y = -0.009x^2 + 1.4841x + 6.4125$	38.23	132.34	0.967**

Table 3(on next page)

Effects of Cd on the photosynthetic pigments in the leaves of two Italian ryegrass cultivars

Chl a, Chl b, and Car, indicate chlorophyll a, chlorophyll b, and carotenoids, respectively. Values (means \pm SE., n=3) followed by different letters in the same columns are significantly different according to the Duncan's test. * $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$, n.s., not significant.

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Table 3 Effects of Cd on photosynthetic pigments in leaves of two Italian ryegrass cultivars						
cultivar	Cd supply μM	chl a mg·g ⁻¹ FW	chl b mg·g ⁻¹ FW	car mg·g ⁻¹ FW	chl (a+b) mg·g ⁻¹ FW	chl a/b mg·g ⁻¹ FW
Idyll	0	1.10±0.07 abc	0.34±0.01 ab	0.18±0.01 abc	1.44±0.05 ab	3.25±0.02 ab
	5	1.03±0.09 bcd	0.32±0.02 abc	0.16±0.02 abcd	1.36±0.07 bc	3.21±0.05 abc
	10	0.97±0.22 cd	0.31±0.04 bc	0.16±0.02 bcde	1.28±0.16 bc	3.08±0.08 abcd
	25	0.97±0.05 cd	0.32±0.02 abc	0.14±0.03 cde	1.33±0.04 bc	3.05±0.22 abcd
	50	0.86±0.01 d	0.27±0.01 c	0.13±0.01 cde	1.13±0.02 c	3.13±0.12 abc
	100	0.58±0.03 ef	0.20±0.01 d	0.10±0.01 ef	0.79±0.01 de	2.86±0.20 bcd
Harukaze	0	1.20±0.12 ab	0.38±0.02 a	0.19±0.01 ab	1.60±0.10 a	3.17±0.08 abc
	5	1.22±0.06 a	0.38±0.01 a	0.20±0.01 a	1.60±0.05 a	3.18±0.04 abc
	10	0.95±0.12 cd	0.29±0.02 bc	0.14±0.01 cde	1.24±0.09 bc	3.34±0.04 a
	25	0.69±0.06 e	0.21±0.01 d	0.12±0.01 def	0.89±0.04 d	3.30±0.03 a
	50	0.48±0.11 f	0.17±0.02 d	0.08±0.01 f	0.65±0.08 e	2.83±0.03 cd
ANOVA	100	0.21±0.08 g	0.07±0.02 e	0.03±0.01 g	0.29±0.06 f	2.71±0.21 d
	Cd	51.73***	42.84***	20.56***	51.17***	4.43**
	cultivar	15.04*	19.39***	4.92*	16.73***	0.01 ns
	Cd×cultivar	9.584***	10.24***	5.39**	10.18***	1.86 ns

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Table 4(on next page)

Correlation coefficients (n=30) among Cd tolerance, uptake, translation, MDA, NPT in Italian ryegrass roots.

TIs, TFs, MDA, and NPT indicate tolerance indexes, translocation factors, malondialdehyde, and non-protein thiols, respectively. * $P < 0.05$; ** $P < 0.01$.

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Table 4 Correlation coefficients (n=30) among Cd tolerance, uptake, translation, MDA, NPT in Italian ryegrass root

index	TIs	Cd concentration	TFs	Cd accumulation	MDA content	NPT content
TIs	1					
Cd concentration	-0.922**	1				
TFs	-0.612*	0.635*	1			
Cd accumulation	-0.185	0.293	-0.003	1		
MDA content	-0.796**	0.837**	0.722**	0.054	1	
NPTcontent	-0.143	0.147	0.04	0.738**	-0.072	1