

Evidence for high inter-generational individual quality in yellow-eyed penguins

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Longitudinal studies focusing on lifetime reproductive success (LRS) have been used to measure individual breeding performance and identify commonalities among successful breeders. By extending the focus to subsequent generations we identify a proportion of high-quality individuals that contribute disproportionately to the population over multiple generations. We used 23 years of yellow-eyed penguin (*Megadyptes antipodes*) breeding data from one breeding area to identify the proportion of individual birds that raised successful breeders, which in turn raised offspring. We explored which life-history components influenced lifetime reproductive success, as this knowledge would enable conservation resources to be focused on high-performing individuals in this endangered population. From 2147 birds marked as chicks, 370 (17.2%) survived to adulthood and recruited to their natal location, of which 219 (10.2%) fledged offspring: 124 (56.6%) of the 219 birds produced offspring that recruited as breeders. Only 102 birds (4.8% of 2147) fledged first-generation offspring that in turn fledged offspring (second-generation offspring, or grand-offspring). We found that c. 25% of the birds that survived to breed had above-average LRS as well as above-average numbers of grand-offspring, and were more likely to have produced first-generation chicks that recruited and also produced above-average numbers of second-generation chicks. Our findings suggest that there is a core of “super-breeders” that contribute disproportionately to the population over successive generations. Lifespan and age-at-first-breeding were correlated with LRS. We suggest that traits of birds relating to longevity, health (e.g. immunocompetence) and fitness could be examined to identify potential links with high LRS and inter-generational fecundity. “Super-breeders” appear to consistently achieve high LRS and long lifespans in a stochastic environment, demonstrating greater resilience in the face of extreme events.

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9 ABSTRACT

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 26 “super-breeders” that contribute disproportionately to the population over successive
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32

34 INTRODUCTION

35 Unlike cross-sectional studies across one or two breeding seasons, longitudinal studies based on
 36 lifetime reproductive success (LRS) average out occasional breeding failures, and increase the
 37 accuracy of measurement of individual success (Krüger & Lindström, 2001). The collection of
 38 long-term life-history data from a population of marked individuals makes it possible to identify
 39 the proportion of animals that produce recruits, enabling conservation efforts to be efficiently
 40 focused on individuals with successful traits (Moreno, 2003). Moreover, the overall contribution
 41 of individuals with different lifespans or reproductive strategies to subsequent generations can be
 42 compared (Clutton-Brock, 1988; Newton, 1989; Wooller, Bradley & Croxall, 1992; Brommer,
 43 Pietiäinen & Kolunen 1998).

44 Studies of LRS have revealed commonalities across bird species: a significant proportion
 45 of fledglings from a given population will die before sexual maturity (Bryant, 1979; Newton,
 46 1989); not all individuals that attempt to breed will be successful; and successful individuals vary
 47 in their productivity (Newton, 1989). The LRS distribution of a population is typically highly
 48 skewed, with large numbers of individuals producing small numbers of young, and only a small
 49 proportion of adults producing large numbers of young (Clutton-Brock, 1988, Newton, 1989).
 50 Specific life-history and reproductive traits can be indicative of LRS. Lifespan is the strongest
 51 correlate, with longer-lived individuals commonly achieving a higher LRS (Gustafsson, 1986;
 52 Clutton-Brock, 1988; Newton, 1989). In seabirds, where individuals can start breeding at various
 53 ages, variance in LRS is largely related to variation in breeding lifespan (Moreno, 2003), because
 54 an increased number of breeding seasons allows individuals more opportunities to successfully
 55 fledge offspring.

If only a small proportion of individuals maintain most of the population, the identification of traits affecting lifetime reproductive success of an individual are relevant for conservation efforts. Resources can be diverted towards protecting particularly productive individuals at times when the population is assailed by environmental challenges such as adverse climate conditions, reduced food availability, disease outbreaks, or catastrophic events (e.g. oil spills, Gartrell et al., 2013). In the case of a pest species, culling efforts could be focused on these highly productive individuals (Moreno, 2003).

Factors known to affect reproductive performance in seabirds and which can interact with each other include age, experience, pair bond duration, health condition, sex, number of mates, mate fidelity and site fidelity (Ryder, 1980; Clutton-Brock, 1988; Gavin & Bollinger, 1988; Bradley et al., 1990; Wooller et al., 1990; Chastel, Weimerskirch & Jouventin, 1995). For example, in long-lived seabird species, a period of poor reproductive success at a young age or at a lower level of experience may be superseded by a period at which the individual performs at their peak reproductive output (Forslund & Pärt, 1995). At an older age, senescence may begin to reduce reproductive output, followed by terminal illness and death (Fowler, 1995; Nisbet & Dann, 2009; Froy et al., 2013). With increasing age, maternal efficiency might allow for control of the timing, size, volume, composition and pore density of eggs, and high levels of pair synchrony through maintenance of long-term pair bonds may reduce incubation periods and increase nesting success (Massaro et al., 2002; Massaro et al., 2004). Breeding skills may therefore improve with both age and experience, as well as with improved synchrony between mated pairs (Forslund & Pärt, 1995). The relative importance of each of these factors in estimating LRS is difficult to assess, and can differ dramatically between species and geographic locations.

Breeder quality might not necessarily be age related. Some birds might just be better than others because they have better skills. State-based assessment of individual breeder quality often requires the assumption or prediction that a component of the individual's health or skill within that system is a driving factor determining its reproductive success (Wendeln & Becker, 1999; Moreno, 2003), and that this superior skill is independent of age-related performance. These state-based qualities and their relationship with breeding success and LRS can be highly variable within a population of seabirds, but might have low variation for individual birds over time (e.g. common terns *Sterna hirundo*, Wendeln & Becker, 1999). Breeding seabirds have a range of responses to catastrophic climatic or weather events, e.g. El Niño Southern Oscillation (ENSO) (Boersma, 1978), and vary in the rate at which food is delivered to chicks (Ens et al., 1992). Territoriality also can be associated with: individual quality in relation to nesting density (Stokes & Boersma, 2000); nest site characteristics (Stokes & Boersma, 1998); and rank dominance (Schubert et al., 2007).

LRS measures the number of offspring produced over a lifetime, however it does not consider the viability of those offspring. There may be variability in the quality of offspring produced by different individuals that further reduces the proportion of individuals contributing to subsequent generations. By using a 23-year dataset of individually-tracked yellow-eyed penguins we were able to follow reproductive success over more than one generation and identify the proportion of a penguin population that produces grand-offspring. Yellow-eyed penguins (hōiho, *Megadyptes antipodes*) are endemic to New Zealand and listed as 'endangered' on the IUCN Red List (Birdlife International, 2015). Some mainland populations are intensively managed to mitigate threats posed by introduced predators, disturbance and habitat destruction (McKinlay, 2001). Because yellow-eyed penguins are sedentary (Seddon, van Heezik &

Ellenberg, 2013), long-lived, have high natal philopatry, high breeding site fidelity once breeding, and are monogamous (Richdale, 1957), they are an ideal species to study LRS. We investigated: (1) the proportion of birds that survive to adulthood; (2) the proportion of adults that breed; (3) the proportion of breeders that produce young that recruit to the breeding population (first-generation); and (4) the proportion of adults producing grand-offspring (second-generation). We also explored the characteristics of highly successful breeders, and the relationship of this trait between generations. We predicted that in line with other seabird species, only a small proportion of yellow-eyed penguins would survive to adulthood, breed, and produce young, and that differences in LRS between males and females exist due to the difference in age-at-first-breeding, with females known to begin breeding earlier than males (Richdale, 1957; Darby & Seddon, 1990). We predicted that lifespan would have the greatest influence on the number of offspring produced, and be positively correlated with LRS. Yellow-eyed penguins with earlier age-at-first-breeding and fewer overall mates were predicted to have greater LRS.

METHODS

Yellow-eyed penguins are solitary breeders, in contrast with most other penguin species and seabirds that breed colonially (Richdale, 1957; Darby & Seddon, 1990). In September-October clutches of up to two eggs are laid, and chicks fledge at c. 106 days from late January to late February (Richdale, 1957; Seddon & Davis, 1989; Darby & Seddon, 1990).

Yellow-eyed Penguin Database

We obtained breeding data from the Yellow-eyed Penguin Database administered by the New Zealand Department of Conservation (DOC) and accessed through a Memorandum of Understanding between DOC, representing the contributors to the database, and the University of

Otago. We analysed only data from yellow-eyed penguins breeding at the Boulder Beach complex on the Otago Peninsula, New Zealand, (45°500 S and 170°300 E; Fig. 1) because it supports a relatively large population of yellow-eyed penguins, has an inter-decadal history of intensive monitoring and it has been trapped for introduced predators over time. This site has the longest history of chick marking, with the majority of chicks fledged at this site marked with a stainless steel flipper band issued by the New Zealand National Bird Banding Scheme (NZBBS). We acknowledge that the use of flipper bands might present bias (*sensu* Petersen et al., 2005), however in contrast to the foraging ranges of penguin species for which negative impacts of flipper bands have been identified, yellow-eyed penguins are inshore foragers and have much shorter foraging trips (van Heezik & Seddon, 1990; Mattern et al., 2007). Consequently impacts of flipper bands are likely to be minimal. While negative impacts have been documented for some penguin species (e.g. king penguins, *Aptenodytes patagonicus*, Gauthier-Clerc et al. (2004); Adélie penguins, *Pygoscelis adeliae*, Dugger et al. 2006; little penguins, *Eudyptula minor*, Hoskins et al. 2008), impacts have been negligible for others (e.g. African penguins, *Spheniscus demersus*, Barham et al. 2008; Hampton et al. 2009; magellanic penguins, *S. magellanicus*, Boersma and Rebstock 2009). A separate study analysing the impact of research manipulations found that even a double banding study on yellow-eyed penguins in one season had no effect on productivity or subsequent survival (Seddon, van Heezik & Ellenberg, 2013). Before commencing this study, we completed a comprehensive error check, which involved checking the original notebook records against electronic database records to ensure a high level of accuracy and consistency.

Data for survival to adulthood and breeding

A total of 2147 birds were marked at Boulder Beach as chicks or juveniles between 1981 and 2003. We used this sample to calculate the proportion of birds that survived post-fledging, to adulthood (defined as reaching two years of age, when sexually mature), the proportion that recruited to Boulder Beach and attempted to breed, fledged offspring, fledged offspring that survived to adulthood and recruited to the breeding population, and fledged offspring (first-generation) that in turn successfully bred and produced offspring (second-generation, or grand-chicks). Birds from cohorts 1981-2003 that were still alive or had been sighted after 2007 were excluded from the sample as their breeding lifetime had not ended ($n = 73$ birds).

Data for LRS analyses

For the detailed analysis of life-history and lifetime reproductive success we used a subset of the data which included information on 130 “founding generation” birds of known sex that survived to breed at least once, from the 1981 to 2003 cohorts, as we considered sex to be an important factor, but sexing information was not always available. We excluded birds whose parents were included in the data subset to avoid pseudoreplication ($n = 87$ birds). We chose the year 2003 as the cut-off cohort, since mean age at first breeding is between three and four years (Richdale, 1957): this allowed for birds to have complete lifespans by age 4 in 2007, breed, have offspring that survived to breed (up to 2011) and grand offspring that survived up until the 2014/15 breeding season, when these data were last updated. We did not include birds beyond the 2003 cohort, or birds that were alive after 2007, so that we were sure that all of their surviving offspring had recruited to breed by 2014.. The latest year a bird with a complete life history could be recorded as breeding at least once, and be presumed dead was 2007 (i.e. birds from the 2003 cohort would have had a complete lifespan of 4 years to meet the study’s parameters). We excluded birds if they were marked before the 1981 cohort when intensive monitoring began.

We defined LRS as the number of offspring produced by an individual over their lifetime. Only data recorded from birds marked as chicks or as juveniles (one-year olds) could be used to ensure complete life-histories, as they were of known age due to plumage differences (duller colouration and lack of the distinctive yellow eye and crown of the adults). Yellow-eyed penguins are typically marked at *c.* 80 - 90 days, before fledging at *c.* 106 days (Seddon et al., 2013). If a bird was marked as a juvenile, we estimated the cohort year by subtracting one year. A bird was considered to have attempted to breed if it or its mate laid an egg.

Analysis of LRS requires complete detectability of the focal population: this would be compromised if individuals bred elsewhere, or skipped breeding years and were consequently recorded as having died. We are confident that we have full records of birds due to the intensive monitoring at the study site and annual monitoring at adjacent sites along the coast, and the high level of natal (*c.* 81%) and breeding philopatry (*c.* 98%), and monogamy exhibited by this species (Richdale, 1957; Ratz et al., 2004). Birds that skip breeding remain largely undetectable during the breeding season, with only *c.* 8% of skipped birds in our sample being resighted as a non-breeder, however detection of breeders is close to 100% (Hegg et al., 2012). In our sample, 53 yellow-eyed penguins that survived to breed skipped at least one breeding season once they had established breeding, which is not uncommon, particularly in the year following a poor season, death of a mate or a divorce (Moore, 1994; Ratz et al., 2004; Setiawan et al., 2005). Due to the small, discrete size of nesting areas, the intensity of monitoring at this and in surrounding sites, and the high degree of breeding site fidelity, we assumed that if a bird or breeding pair were not seen during multiple visits to the breeding area and to other surrounding areas from early incubation to the end of the guard period they were undertaking a breeding skip. None of

the birds we assumed to be undertaking a breeding skip were re-sighted at adjacent monitored breeding areas during their skipped year.

Sample parameters

The sample parameters for modelling life-history included sex, total number of breeding attempts, recruitment of first-generation offspring, recruitment of successful first-generation offspring, age-at-first-breeding, breeding lifespan, total number of mates and lifespan. We did not include ‘Cohort’ as a factor because the longer-lived birds in later cohorts had not completed their breeding lifetimes, unlike shorter-lived birds. Therefore including cohort would give a false impression that the super breeder phenomenon ceased at 1994. To ensure that cohort was not an important variable we ran a modified model on a subset of the total data set - the birds from cohorts 1981 – 1994, which included both short-lived and long-lived birds for which we had full LRS ($n = 161$), and found no strong effects (Suppl. Material S2, T2).

Birds were sexed by adult head and foot measurements according to Setiawan, Darby & Lambert (2004). In instances where birds had never been measured or when fledgling measurements were analysed, we inferred the sex from mates where possible, on the assumption that pairings were between males and females only and that the mate had been correctly identified. If there were no measurements or sex recorded for mates, we removed these birds from any lifetime data analysis. We limited our sample for the analysis of life-history traits affecting LRS to birds that survived to breed from the original sample of 2147 birds, that we were able to sex, and we excluded their offspring (“founding generation”, females $n = 62$, and males $n = 68$), so as not to pseudoreplicate breeding pairs, or parents and their offspring.

We measured lifespan in whole years at the time of marking as chicks (*c.* 3 months old), to the time of either being found dead or ‘missing’ after three consecutive years. Age-at-first-

breeding was recorded as the age of the bird during its first recorded breeding attempt. We calculated the number of mates as the minimum possible number of mates, due to 60 of 130 birds in the sample having unidentified mates in some years. We assumed that if a bird's mate was not recorded but it was breeding with a particular bird in the previous and subsequent years that it was the same mate in all three seasons.

Statistical analysis

We carried out all statistical analyses using R (Version 3.3.1, R Core Team 2016). We used two-sample Wilcoxon rank-sum tests to test for statistical significance between males and females, for parameters including LRS, age-at-first-breeding, recruitment of first-generation offspring, recruitment of successful first-generation offspring, lifespan, total number of mates, total number of breeding attempts and breeding lifespan (see Supplementary Material S1).

The relationships between recruitment of first-generation breeders, successful first-generation breeders, sex, and the effect of life-history traits (age-at-first-breeding, lifespan, total number of mates, total number of breeding attempts, breeding lifespan) on LRS were analysed using a generalised linear mixed models (GLMM) with a Poisson distribution and a random factor, mate code, to account for pseudoreplication associated with mated pairs being included in the analysis, using the *lme4* package (Bates et al., 2015). Fit of the maximal model was assessed using R^2GLMM from the AICcmodavg package (Mazerolle 2016). We included only uncorrelated variables within the same model ($r < 0.6$; Hosmer & Lemeshow 2005) to avoid multicollinearity. We used an information-theoretic approach to model selection, by constructing a maximal model containing all probable input variables (based on *a priori* reasoning), and then ranking this model against all of its derivatives using QAICc. To account for model selection

237 uncertainty, model-averaging was conducted for the best models ($2\Delta\text{QAICc}$, see Supplementary
238 Material S3), using the *MuMIn* and *AICcmodavg* packages in R (Bartoń, 2016; Mazerolle, 2016).

239 In order to compare specifically the life-history characteristics (age-at-first-breeding,
240 lifespan, total number of mates, total number of breeding attempts, breeding lifespan) between
241 birds which proved over the two generations to be highly successful breeders (“high quality”),
242 and the remainder of the birds (“ordinary”), we defined the highly successful individuals as
243 follows: those birds with above-average LRS relative to their sex (females ≥ 7 , males ≥ 6) and
244 those that had an above-average number of grand-offspring (second-generation chicks; females \geq
245 10, males ≥ 6). Using the same GLMM approach, we analysed the effect of life-history
246 parameters on the LRS of these two groups of birds to determine differences in breeder quality.

247

248 RESULTS

249 Of the total sample of 2147 birds marked as chicks or juveniles from 1981 to 2003, 1546
250 (72.0%) were thought to have died before reaching adulthood, whereas 441 birds survived to be
251 seen at least once as an adult: 71 of these birds were sighted away from Boulder Beach at other
252 monitoring locations where they subsequently bred. Of the 370 birds that were re-sighted at
253 Boulder Beach as an adult at least once (17.2%), 264 attempted to breed at least once (12.3%),
254 and 219 bred successfully at least once (10.2%). Only 124 birds produced at least one first-
255 generation chick that recruited to the breeding population and attempted to breed at least once
256 (5.8%), and 102 had first-generation offspring that not only recruited but bred successfully at
257 least once (4.8%). Overall figures are presented in Table 1.

258 LRS of male and female yellow-eyed penguins

There was high individual variance in LRS calculated for both males and females ($n = 130$), with this variance being higher for females (Table 2; Figure 2). The maximum number of total offspring a female yellow-eyed penguin produced was 24, compared to 23 for males. The only significant differences between males and females were for recruitment of breeders ($W = 2554$, $p\text{-value} = 0.03212$) and recruitment of successful breeders ($W = 2615$, $p\text{-value} = 0.01198$). There was no difference between males and females for LRS ($W = 2452$, $p\text{-value} = 0.1077$), age-at-first-breeding ($W = 1750.5$, $p\text{-value} = 0.0791$), lifespan ($W = 2169.5$, $p\text{-value} = 0.7755$), total number of mates, ($W = 2187.5$, $p\text{-value} = 0.6948$), total number of breeding attempts ($W = 2370$, $p\text{-value} = 0.2197$), and breeding lifespan ($W = 2359.5$, $p\text{-value} = 0.2392$).

Relationships between fledging and recruitment

There was a strong positive relationship between number of chicks fledged per parent (LRS) and number that recruited for females ($\lambda(\text{Female}_i) = \exp(-0.27) * \exp(0.12 * \text{LRS}_i)$) and males ($\lambda(\text{Male}_i) = \exp(-0.60) * \exp(0.13 * \text{LRS})$). There was also a significant positive relationship between the number of chicks fledged (LRS) and number of successful recruits (i.e. recruits that in turn successfully fledged offspring during at least one breeding attempt) for females ($\lambda(\text{Female}_i) = \exp(-0.65) * \exp(0.12 * \text{LRS}_i)$) and males ($\lambda(\text{Male}_i) = \exp(-1.26) * \exp(0.14 * \text{LRS})$).

Life-history traits

Lifespan was the strongest positive correlate of LRS (0.61 ± 0.03), followed by a negative correlation with age-at-first-breeding (-0.14 ± 0.04 ; GLMM $R^2_{\text{m}} = 0.56$, $R^2_{\text{c}} = 0.79$; Table 3). There was a trend associated with sex, with males having slightly lower LRS compared to females (Fig. 3). There was no association with the number of mates and LRS; or interactions between sex, age-at-first-breeding or lifespan (Table 3).

282 Determining the traits of “high quality” breeders

283 “High-quality” birds (i.e., those with above-average LRS and higher than average numbers of
 284 successful breeding offspring; females = 16, males = 16) produced 389 fledged chicks, of which
 285 133 first-generation offspring recruited, and 94 were successful, producing 713 grand-offspring
 286 (second-generation chicks) (Fig. 4). These higher-quality breeding birds produced 26 first-
 287 generation offspring with an above-average $LRS \geq 6$. In contrast, the remaining “ordinary” birds
 288 (females = 46, males = 52) produced 379 chicks, of which 100 first-generation offspring
 289 recruited, and 55 were successful in producing 252 grand-offspring (second-generation chicks)
 290 (Fig. 4). The ordinary breeding birds ($n = 98$) produced 10 above-average first-generation
 291 offspring.

292 An interaction effect between breeder type and lifespan was detected, but there was no
 293 interaction between age-at-first breeding and breeder type (Table 4), despite the ordinary female
 294 birds beginning breeding at least one year younger than the high-quality female birds (Table 5).
 295 The ordinary birds had shorter lifespans and therefore lower LRS, with high-quality birds having
 296 on average double the longevity of their short-lived conspecifics (Table 5).

297

298 DISCUSSION

299 By tracking reproductive success in yellow-eyed penguins over more than one generation we
 300 show that only a small proportion of fledglings survive, recruit, and attempt to breed, however
 301 lifetime reproductive success and survival appears to be an inter-generational trait, with above-
 302 average breeders more likely to produce chicks that will be highly successful breeders. Fewer
 303 than 1.5% (32 of 2147) of these breeding birds are what we call “super-breeders”. These “super-

breeders” appear to be successful in producing offspring that will themselves survive and go on to contribute disproportionately to the next generation.

Only 10.2% of the sample population of 2147 fledgling yellow-eyed penguins eventually recruited and produced offspring at all, meaning that 89.8% of young fledged did not contribute to the next generation at Boulder Beach. Low juvenile survival is likely to be the principal reason for the low number of penguins recruiting to breeding populations. Only 20.5% of yellow-eyed penguin fledglings survived to sexual maturity (two years of age, including birds that did not recruit), a similar proportion to the 20.8% yellow-eyed penguins resighted as sexually-mature adults at the Boulder Beach complex between 1981 to 1990 (Efford, Darby & Spencer, 1996), although our reported result includes 3.3% of birds that were resighted away from Boulder Beach. Our survival to adulthood rate was higher than the 10.4% of Adélie penguins (*Pygoscelis adeliae*) that survived to age two (Ainley & DeMaster, 1980), but low compared to the range of values for survivorship from fledging to sexual maturity for 19 species of passerines and seabirds (42 - 86%; Newton, 1989). Survivorship to two years was significantly lower than reported in several other studies of seabirds, including 57.6% for common guillemots (*Uria aalge*; Crespin et al., 2006), 41 to 54% for sooty shearwaters (*Puffinus griseus*; Fletcher et al., 2013), and c. 77% for king penguins after one year (*Aptenodytes patagonicus*; Saraux et al., 2011). It was even lower than the c. 32% of yellow-eyed penguins that survived to age two between 1936 and 1952 (Richdale, 1957).

The probability that birds survive the period between parental care and adulthood has a large influence on population dynamics, but is highly variable (Maness & Anderson, 2013). The most common hypothesis for high rates of mortality in young birds is their lack of experience, poor foraging skills and physical immaturity (Lack, 1954; Ashmole, 1963; Orians, 1969; Dunn,

1972). Positive correlations are predicted between body mass and juvenile survival, based on the assumption that heavier juveniles have fat reserves that buffer the food limitation associated with inexperience (Lack, 1966; McClung et al., 2004; Maness & Anderson 2013). Yellow-eyed penguins are sedentary foragers that lack a long-distance migratory phase in their life history, however juveniles undergo a pelagic phase lasting for up to two years, during which time they are sighted only erratically along the coast. No information exists on where juveniles disperse to (Darby & Seddon, 1990), and most mortality occurs during this post-fledging pelagic phase. It is unclear whether the low survival of juvenile penguins in this study is normal or depressed by changing environmental conditions.

This study indicates that of the chicks seen again at Boulder Beach after fledging (24.7% of the original sample), 69.8% were seen at two years of age (i.e. up to the onset of sexual maturity, 69.8%, Table 1), but only 49.8% survived to breed at least once. These figures suggest that juvenile mortality occurs in two or more stages: as high post-fledging mortality due to inexperience, immaturity and lack of skill, and possibly due to seasonal fluctuations in prey availability later in the breeding season when juveniles must prepare for their first annual moult. The difference in juvenile survival between Richdale's (1957) study and ours may be indicative of an adverse change in foraging conditions (Browne et al., 2011; Mattern et al., 2014), entanglement in recreational or commercial fishing gear (Darby & Dawson, 2000), increasing frequency of poor seasons (van Heezik 1990), and competition with or predation by recovering otariid populations (Bradshaw, Lalas & Thompson, 2000; Lalas et al., 2007), all of which are documented to affect adult yellow-eyed penguins. Marine pollution that results in disease outbreaks and mass mortality events, have also been hypothesised (e.g. 1990 mass mortality

event, Gill & Darby 1993; e.g. diphtheritic stomatitis, Alley et al., 2004, 2017; Trathan et al., 2015).

From the sample of 2147 fledglings, the proportions that survived and attempted breeding (12.3%), fledged offspring (10.2%) and fledged offspring that recruited (5.8%) seem low. However, the proportion of birds that attempted breeding and were successful is relatively high (219 of 264, 82.9%). In other words, if a bird was successful in surviving to make a breeding attempt, there was an 83% probability that it would be successful in fledging at least one chick in its lifetime, a 47% (124/264) probability it would fledge at least one chick that would recruit to the breeding population, and a 39% (102/264) probability that the bird would fledge chicks that would recruit and subsequently fledge offspring. The proportion of yellow-eyed penguins surviving to attempt to breed at least once was comparatively lower than in red-billed gulls (*Larus novaehollandiae*, 18 - 22%), little penguins (*Eudyptula minor*, 28 - 35%), kittiwakes (*Rissa tridactyla*, 34 - 42%) and short-tailed shearwaters (*P. tenuirostris*, 69 - 73%). However the proportion of breeding yellow-eyed penguins that produced recruits (47%) is one of the highest, with only kittiwakes having similar recruitment rates (*c.* 41 - 50%) (Coulson, 1988; Wooller et al., 1988; Mills, 1989; Dann & Cullen, 1990; Moreno, 2003). In yellow-eyed penguins, recruitment into the breeding population appears to be driven in part by the higher survival rate of the offspring of a subset of breeders, with high-quality birds, labelled here as “super-breeders”, producing more recruits (133 first-generation recruits from 32 birds, 57.1%) than the ordinary breeders (100 first-generation recruits from 98 birds, 42.9%).

Both male and female penguins that survived to breed varied considerably in the total number of offspring they fledged. For a long-lived species, the average number of fledged young seems relatively small (female mean = 6.82, male mean = 5.07), but falls within the range of

values reported from the few studies that have estimated mean LRS in seabirds, passerines and birds of prey, demonstrating that LRS for many species of birds remain similar as a result of life-history trade-offs (Table 6). Females had slightly longer lifespans and longer breeding lifespans than males, because females started breeding earlier than males. Females may have more opportunities to breed than males, due to an apparent sex-skew, with males outnumbering females (Richdale, 1957). The maximum number of fledged offspring for both male (23) and female yellow-eyed penguins (24) was much higher than mean values, reflecting the highly negatively skewed distribution of LRS (Fig.2). This is consistent with the observation that most individuals produce small numbers of young, and only a few produce many (Newton 1989). Nevertheless, there was a wide range in the number of young fledged by individual birds regardless of sex, despite the greater cost of reproduction incurred by breeding females. Newton (1989) concluded that LRS is generally similar for males and females in species that lack high levels of sexual dimorphism, which is the case for yellow-eyed penguins (Seddon et al., 2013).

LRS predictors

Lifespan was the strongest correlate of LRS, with the number of offspring produced increasing significantly with increased lifespan. This trend is very common for seabirds (Clutton-Brock, 1988; Newton, 1989; Newton, 1995), and is attributed to a number of factors: increased breeding opportunities, and increasing parental experience with lifespan, as has been demonstrated in other seabird species (Limmer & Becker, 2009; Saraux et al., 2012), and the general fitness required for a long lifespan. Long-lived birds are the primary contributors to the gene pool in many species, meaning there is likely to be selection for viability (Moreno, 2003; Mauck, Huntington & Grubb, 2004). We found that high-quality breeders had lifespans that were on average double that of ordinary breeders, but they produced three to four times more offspring in

their lifetimes than ordinary birds (Table 5). In other species lifespan explains less of the variance when the number of recruits is examined, as opposed to number of offspring produced (Newton, 1989). In contrast, we found a highly significant relationship between LRS and the number of recruits and successful recruits produced for yellow-eyed penguins, meaning that the characteristics of birds with longer lifespans are likely to be reliable predictors of parental quality for this species.

Age-at-first-breeding was the second strongest predictor of LRS in yellow-eyed penguins, with birds that began breeding later having lower lifetime totals of offspring, due to a decrease in total breeding opportunities (Newton, 1989). The theory of antagonistic pleiotropy suggests that high early-life reproductive output is at the expense of later-life productivity, and can be selected for if selection is stronger at early stages of life, so that early benefits outweigh later costs (Williams, 1957). While there was a difference between the LRS of males and females in this study, no interaction effect could be detected, despite earlier reproduction in females, which may potentially result in accelerated reproductive senescence (Partridge, 1992, Reed et al., 2008). This trend has been observed in several long-lived bird species, which all showed a positive correlation between age-at-first-breeding and survival in females, suggesting a trade-off between early recruitment and lifespan (Ollason & Dunnet, 1978; Ainley & DeMaster, 1980; Pyle et al., 1997; Tavecchia et al., 2001). Individual variation in LRS for yellow-eyed penguins therefore appears to be due to variation in lifespan (1-24 years) and age-at-first-breeding (2-12 years), together determining the length of the breeding lifespan (1-18 years).

It is common for many species of seabirds to show reduced breeding success after changing mates, most likely due to a trade-off in time and energy expenditure for finding a new mate and foraging, and also due to lack of familiarity with the new mate (Ollason & Dunnet,

1978; Coulson & Thomas, 1985; Newton, 1989). In short-tailed shearwaters a mate change results in a temporary decrease in breeding success, but this effect is smaller in individuals that are more experienced breeders (Wooller et al., 1989). Breeding success of male common guillemots decreased with an increasing number of mates (Lewis et al. 2006). Yellow-eyed penguins that change mates are more likely to experience breeding failure the subsequent year (Setiawan et al., 2005). We did not detect a significant negative effect of number of mates on lifetime number of offspring produced, possibly due to the tendency for longer-lived birds to outlive their mates, resulting in higher overall numbers of mates.

Conservation implications

Chronic and acute stress as a result of climate change, marine pollution, disturbance at terrestrial breeding sites and extreme nutritional stress may decrease LRS, as the cumulative effects of increasing types of stressors force individuals to reduce their investment in productivity, increase breeding skip behaviours (e.g. red-footed boobies *Sula sula*, Cubayanes et al., 2011) or result in breeder mortality (Kitaysky et al., 2010). In black-legged kittiwakes, breeding behaviour is mediated by increased corticosterone production during periods of poor food supply (Kitaysky et al., 2010; Schultner et al. 2013). Clarifying the factors that separate the success of the few that produce many offspring from the many that do not may therefore need to take into account the role of chronic or acute stress on the parameters that may be used to measure their fitness. Likewise, birds that contribute disproportionately to successive generations may have higher thresholds for anthropogenic and environmental stressors than average birds. The impact of extreme events on different phenotypes of conspecifics may differ as a consequence of the "super-breeder" phenomenon, since these birds tend to consistently achieve high LRS and long-lifespans in a stochastic environment.

It appears that the Boulder Beach population of yellow-eyed penguins is sustained by a small proportion of high-quality, long-lived birds, the “super-breeders”. High levels of philopatry may drive high-quality and ordinary breeding recruits to return to their natal area, and once they begin breeding they are likely to remain at these breeding sites for life. This behaviour may be hazardous for population stability should either one of their marine or terrestrial habitats become threatened. If circumstances require it necessary to protect specific individuals in a population from catastrophe, or to differentially allocate resources due to budget constraints, it would be important to distinguish between potentially very successful breeders and the evolutionary “living dead” (Moreno, 2003). Oiling is the greatest anthropogenic threat to penguins (Trathan et al., 2015), requiring triage of breeders for temporary captive management. The unexplained mass mortalities of adult and juvenile yellow-eyed penguins on the Otago Peninsula in 1990, 1996 and 2013 due to exposure to an unknown toxic agent have presented conservation managers with opportunities to safeguard specific individuals from harm (Gill & Darby, 1993; DOC, unpublished data). While effort should be placed on safeguarding all individuals in a threatened population during a period of catastrophe, only a small proportion of individuals will contribute to the recovery of the population following such an event.

Although it seems sensible to focus conservation resources on “super-breeders”, the challenge lies in identifying them. The positive relationship between LRS and the number of successful recruits indicates that birds demonstrating relatively high LRS are also those that produce high-quality offspring. Lifespan is the main predictor of LRS, but unfortunately it cannot be calculated until the death of an individual. Age-at-first-breeding can be identified before death, although its association with potential LRS is much weaker, however high-quality female birds tended to recruit a year later than ordinary birds. Fledging weight of the individual

could indicate that the individual comes from a high quality lineage and will be more likely to survive (McClung et al. 2004) and go on to breed and produce high quality offspring that also fledge at a higher than average weight. We could not explore this relationship with our dataset as fledging weights were not reliably recorded for historical records. The value of other potential indicators of living super-breeders include oxidative stress, white blood cell counts, hue and size of the coloured eye and eye stripe; these are being explored in ongoing research.

It may be possible to single out birds on the basis of life-history traits that relate to state-based quality. The importance of state-based assessments for yellow-eyed penguins has yet to be fully explored, especially with regard to analysing the immunocompetence of individuals. Disease prevalence has increased in recent years (Alley et al., 2004, 2017; Hill et al., 2010; Argilla et al., 2013). Moreno et al. (1998) measured variables related to health state and cell-mediated immunity between early and late breeders for chinstrap penguins (*P. antarctica*), finding that early breeders experienced better health than later breeders. Female chinstrap penguins with leukocytosis laid smaller eggs, had slower chick growth rates, and were more prone to failure (Moreno et al., 1998). Information about foraging ecology, particularly in young birds is also necessary. Foraging strategies in high quality Adélie penguins have been linked to better provisioning of chicks, suggesting that some birds may be physiologically more capable by virtue of genetic superiority (Lescroël et al., 2010).

Around half of all seabird species globally are thought to be in decline, many of which have restricted numbers and ranges, and demographic characteristics that severely limit their rate of recovery (Croxall et al. 2012). As long-lived species, and with demographic characteristics similar to those of yellow-eyed penguins, it is likely that population persistence in other species is also dependent to some extent on a subset of successful breeders. Understanding what causes

some birds to display enhanced resilience and identifying and protecting these individuals may be vital in the face of growing threats.

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REFERENCES

- Ainley DG, DeMaster DP. 1980. Survival and mortality in a population of Adélie penguins. *Ecology* 61: 522-530.
- Alley MR, Morgan KJ, Gill JM, Hocken AG. 2004. Diseases and causes of mortality in yellow-eyed penguins, *Megadyptes antipodes*. *Kokako* 11(2): 18-23.
- Alley MR, Suepaul RB, McKinlay B, Young MJ, Wang J, Morgan KJ, Hunter SA, and Gartrell BD, 2017. Diphtheritic stomatitis in yellow-eyed penguins (*Megadyptes antipodes*) in New Zealand. *Journal of Wildlife Diseases* 53 (1): DOI: 10.7589/2015-07-195
- Argilla LS, Howe L, Gartrell BD, Alley MR. 2013. High prevalence of *Leucocytozoon spp.* in the endangered yellow-eyed penguin (*Megadyptes antipodes*) in the sub-Antarctic regions of New Zealand. *Parasitology* 140(5): 672-682.
- Ashmole N P. 1963. The regulation of numbers of tropical oceanic birds. *Ibis* 103(3): 458-473.
- Bartoń K. 2016. MuMIn: Multi-Model Inference. R package version 1.15.6. <https://CRAN.R-project.org/package=MuMIn>.

511 Bates D, Maechler M, Bolker B, Walker S. 2015. Fitting Linear Mixed-Effects Models Using
512 lme4. *Journal of Statistical Software* 67(1), 1-48. doi:10.18637/jss.v067.i01.

513 BirdLife International, 2015. Species factsheet: *Megadyptes antipodes*. Downloaded from
514 <http://www.birdlife.org> on 14/08/2015.

515 Boersma PD. 1978. Breeding patterns of Galapagos penguins as an indicator of oceanographic
516 conditions. *Science* 200:1481-1483.

517 Bradley J, Wooller R, Skira I, Serventy D. 1990. The influence of mate retention and divorce
518 upon reproductive success in Short-tailed Shearwaters *Puffinus tenuirostris*. *Journal of*
519 *Animal Ecology* 58: 487-496.

520 Bradshaw CJA, Lalas C, Thompson CM. 2000. Clustering of colonies in an expanding
521 population of New Zealand fur seals (*Arctocephalus forsteri*). *Journal of Zoology* 250(1):
522 105-112.

523 Brommer JE, Pietiäinen H, Kolunen H. 1998. The effect of age at first breeding on Ural Owl
524 lifetime reproductive success and fitness under cyclic food conditions. *Journal of Animal*
525 *Ecology* 67: 359-369.

526 Brommer J E, Ahola K, Karstinen T. 2005. The colour of fitness: plumage coloration and
527 lifetime reproductive success in the tawny owl. *Proceedings of the Royal Society of London*
528 *B* 272: 935-940.

529 Browne T, Lalas C, Mattern T, van Heezik Y. 2011. Chick starvation in yellow-eyed penguins:
530 Evidence for poor diet quality and selective provisioning of chicks from conventional diet
531 analysis and stable isotopes. *Austral Ecology* 36: 99-108.

532 Bryant D. 1979. Reproductive costs in the House Martin *Delichon urbica*. *Journal of Animal*
533 *Ecology* 48: 655-675.

534 Chastel O, Weimerskirch H, Jouventin P. 1995. Influence of body condition on reproductive
535 decision and reproductive success in the Blue Petrel. *The Auk* 112: 964-972.

536 Clutton-Brock TH. 1988. *Reproductive success: Studies of Individual Variation in Contrasting*
537 *Breeding Systems*. Chicago: University of Chicago Press. 548p.

538 Costantini D, Meillère A, Carravieri A, Lecomte V, Sorci G, Faivre B, Weimerskirch H,
539 Bustamante P, Labadie P, Budzinski H, Chastel O. 2014. Oxidative stress in relation to
540 reproduction, contaminants, gender and age in a long-lived seabird. *Oecologia* 175: 1107-
541 1116.

542 Coulson Thomas C. 1985. Differences in the breeding performance of individual kittiwake gulls,
543 *Rissa tridactyla*. In: R.M. Sibley and Smith, R.H. (eds.). *Behavioural Ecology*: p. 489-503.
544 Blackwell Scientific Publications, London.

545 Coulson J C. 1988. Lifetime reproductive success in the Black-legged Kittiwake (*Rissa*
546 *tridactyla*). In: Ouellet H, ed. *Acta XIX Congressus Internationalis Ornithologici* Vol. 2., p.
547 2140-2147. Ottawa: University of Ottawa Press.

548 Crawford RJM, Davis SA, Harding RT, Jackson LF, Leshoro TM, Meyer M., Randall RM,
549 Underhill LG, Upfold L, van Dalsen AP, van der Merwe E, Whittington PA, Williams AJ,
550 Wolfhaardt AC. 2000. Initial impact of the *Treasure* oil spill on seabirds off western South
551 Africa. *South African Journal of Marine Science* 22: 157-176.

552 Crespin L, Harris MP, Lebreton JD, Frederiksen M, Wanless S. 2006. Recruitment to a seabird
553 population depends on environmental factors and on population size. *Journal of Animal*
554 *Ecology* 75:228-238.

555 Croxall JP, Butchart, SHM, Lascelles B, Stattersfield AJ, Sullivan B, Symes A, Taylor P. 2012.
556 Seabird conservation status, threats and priority actions: a global assessment. Bird
557 Conservation International 22(1): 1-34. Doi: <https://doi.org/10.1017/s0959270912000020>

558 Cubayanes S, Doherty PF, Schieber EA, and Gimenez O. 2011. To breed or not to breed: a
559 seabird's response to extreme climatic events. *Biology Letters* 7: 303-306.

560 Dann P, Cullen J. 1990. Survival, patterns of reproduction, and lifetime reproductive output in
561 Little Blue Penguins *Eudyptula minor* on Phillip Island, Victoria, Australia. In: Davis LS,
562 Darby JT, eds. *Penguin Biology*. p. 63-84. San Diego: Academic Press.

563 Darby JT, Dawson SM. 2000. Bycatch of yellow-eyed penguins (*Megadyptes antipodes*) in
564 gillnets in New Zealand waters 1979 - 1997. *Biological Conservation* 93: 327 - 332.

- 565 Darby JT, Seddon PJ. 1990. Breeding biology of Yellow-eyed Penguins *Megadyptes antipodes*.
566 In: Davis LS, Darby JT, eds. *Penguin Biology*: 45–62. San Diego: Academic Press.
- 567 Daunt F, Wanless S, Harris MP, Money L, Monaghan P. 2007. Older and wiser: improvements
568 in breeding success are linked to better foraging performance in European shags. *Functional*
569 *Ecology* 21: 561 - 567.
- 570 Dunn EK. 1972. The effect of age on the fishing ability of sandwich terns, (*Sterna sandvicensis*).
571 *Ibis* 114: 360-366.
- 572 Efford M, Darby J, Spencer N. 1996. Population studies of Yellow-eyed Penguins. *Science for*
573 *Conservation* 22. Department of Conservation, Wellington. 29p.
- 574 Ellenberg U, Setiawan AN, Cree A, Houston DM, Seddon PJ. 2007. Elevated hormonal stress
575 response and reduced reproductive output in yellow-eyed penguins exposed to unregulated
576 tourism. *General and Comparative Endocrinology* 152: 54-63.
- 577 Ens BJ, Kersten M, Brenninkmeijer A, Hulscher JB. 1992. Territory quality, parental effort and
578 reproductive success of oystercatchers *Haematopus ostralegus*. *Journal of Animal Ecology*
579 61: 703-715.
- 580 Fletcher D, Moller H, Clucas R, Bragg C, Scott D, Scofield P, Hunter CM, Win I W, Newman J,
581 McKechnie S, De Cruz J, Lyver P. 2013. Age at First Return to the Breeding Colony and
582 Juvenile Survival of Sooty Shearwaters. *Condor* 115: 465-476.
- 583 Forslund P, Pärt T. 1995. Age and reproduction in birds - hypotheses and tests. *Trends in*
584 *Ecology and Evolution* 10(9): 375-378.
- 585 Fowler GS. 1995. Stages of age-related reproductive success in birds: simultaneous effects of
586 age, pair-bond duration and reproductive experience. *American Zoology* 35: 318-328.
- 587 Froy H, Phillips RA, Wood AG, Nussey DH, Lewis S. 2013. Age-related variation in
588 reproductive traits in the wandering albatross: evidence for terminal improvement following
589 senescence. *Ecology Letters* 16: 642-649.
- 590 Garamszegi LZ, Török J, Michl G, Møller AP. 2004. Female survival, lifetime reproductive
591 success and mating status in a passerine bird. *Oecologia* 138: 48-56.
- 592 Gartrell BD, Collen R, Dowding JE, Gummer H, Hunter S, King EJ, Laurenson L, Lilley CD,
593 Morgan KJ, McConnell HM, Simpson K, Ward JM. 2013. Captive husbandry and veterinary

594 care of northern New Zealand dotterels (*Charadrius obscurus aquilonius*) during the CV
595 Rena oil-spill response. *Wildlife Research* 40: 624-632.

596 Gavin TA, Bollinger EK. 1988. Reproductive correlates of breeding-site fidelity in Bobolinks
597 *Dolichonyx oryzivorus*. *Ecology* 69: 96-103.

598 Gill JM, Darby JT. 1993. Deaths in yellow-eyed penguins (*Megadyptes antipodes*) on the Otago
599 Peninsula during the summer of 1990. *New Zealand Veterinary Journal* 41: 39-42.

600 Goldsworthy SD, Gales RP, Giese M, Brothers N. 2000. Effects of the Iron Baron oil spill on
601 little penguins (*Eudyptula minor*). I. Estimates of mortality. *Wildlife Research* 27: 559-571.

602 Gustafsson L. 1986. Lifetime reproductive success and heritability: empirical support for Fisher's
603 fundamental theorem. *The American Naturalist*, 128: 761-764.

604 Hegg D, Giroir T, Ellenberg U, Seddon PJ. 2012. Yellow-eyed penguin (*Megadyptes antipodes*)
605 as a case study to assess the reliability of nest counts. *Journal of Ornithology* 153(2): 457-
606 466.

607 Hill A, Howe L, Gartrell B, Alley M. 2010. Prevalence of *Leucocytozoon spp.*, in the endangered
608 Yellow-eyed Penguin *Megadyptes antipodes*. *Parasitology* 137: 1477-1485.

609 Hosmer DW, Lemeshow S. 2005. Introduction to the Logistic Regression Model. *Applied*
610 *Logistic Regression* (John Wiley & Sons, Inc.), pp 1–30.

611 Jensen H, Sæther BE, Ringsby TH, Tufto J, Griffith SC, Ellegren H. 2004. Lifetime reproductive
612 success in relation to morphology in the House Sparrow *Passer domesticus*. *Journal of*
613 *Animal Ecology* 73: 599-611.

614 Kim S.-Y, Velando A, Torres R, Drummond H. 2011. Effects of recruiting age on senescence,
615 lifespan and lifetime reproductive success in a long-lived seabird. *Oecologia* 166: 615-626.

616 King SD, Harper GA, Wright JB, McInnes JC, van der Lubbe JE, Dobbins M, Murray SJ. 2012.
617 Site-specific reproductive failure and decline of a population of the endangered yellow-eyed
618 penguin: a case for foraging habitat quality. *Marine Ecology Progress Series* 467: 233–244.

619 Kitaysky AS, Piatt JF, Hatch SA, Kitaikaia EV, Benowitz-Frdericks ZM, Shultz MT, Wingfield
620 JC. 2010. Food availability and population processes: severity of nutritional stress during
621 reproduction predicts survival of long-lived seabirds. *Functional Ecology* 24: 625-637.

622 Korpimäki E. 1992. Fluctuating food abundance determines the lifetime reproductive success of
623 male Tengmalm's Owls. *Journal of Animal Ecology* 61: 103-111.

624 Krüger O, Lindström J. 2001. Lifetime reproductive success in Common Buzzard, *Buteo buteo*:
625 from individual variation to population demography. *Oikos* 93: 260-273.

626 Lack DL. 1954. *The Natural Regulation of Animal Numbers*. Oxford: Clarendon Press, 343p.

627 Lack D L. 1966. *Population studies of birds*. Oxford: Clarendon Press. 341p.

628 Lalas C, Ratz H, McEwan K, McConkey SD. 2007. Predation by New Zealand sea lions
629 (*Phocarctos hookeri*) as a threat to the viability of yellow-eyed penguins (*Megadyptes*
630 *antipodes*) at Otago Peninsula, New Zealand. *Biological Conservation* 135: 235-246.

631 Lescroël A, Ballard G, Toniolo V, Barton KJ, Wilson PR, Lyver PO'B, Ainley DG. 2010.
632 Working less to gain more: when breeding quality relates to foraging efficiency. *Ecology*
633 91(7): 2044-2055.

634 Lewis S, Wanless S, Elston DA, Schultz MD, Mackley E, Du Toit M, Underhill JG, Harris MP.
635 2006. Determinants of quality in a long-lived colonial species. *Journal of Animal Ecology*
636 75: 1304-1312.

637 Limmer B, Becker PH. 2009. Improvement in chick provisioning with parental experience in a
638 seabird. *Animal Behaviour* 77: 1095-1101.

639 Maness TJ, Anderson DJ. 2013. Predictors of juvenile survival in birds. *Ornithological*
640 *Monographs* 78: 1-55.

641 Massaro M, Darby JT, Davis LS, Edge K-A, Hazel MJ. 2002. Investigation of interacting effects
642 of female age, laying dates, and egg size in yellow-eyed penguins (*Megadyptes antipodes*).
643 *The Auk* 119(4): 1137-1141.

644 Massaro M, Davis LS, Darby JT. 2003. Carotenoid-derived ornaments reflect parental quality in
645 male and female yellow-eyed penguins (*Megadyptes antipodes*). *Behavioural Ecology and*
646 *Sociobiology* 55: 169-175.

647 Massaro M, Davis LS, Darby JT, Robertson GJ, Setiawan AN. 2004. Intraspecific variation of
648 incubation periods in yellow-eyed penguins *Megadyptes antipodes*: testing the influence of
649 age, laying date and egg size. *Ibis* 146: 526-530.

650 Mattern T, Ellenberg U, Houston DM, Davis LS. 2007. Consistent foraging routes and benthic
651 foraging behaviour in Yellow-eyed Penguins. *Marine Ecology Progress Series* 343: 295-
652 306.

653 Mattern T, Ellenberg U, Houston DM, Lamare M, Davis LS, van Heezik Y, Seddon PJ. 2014.
654 Straight line foraging in yellow-eyed penguins: new insights into cascading fisheries effects
655 and orientation capabilities of marine predators. *PLoS ONE* 8(12): e84381.
656 doi:10.1371/journal.pone.0084381

657 Mauck RA, Huntington CE, Grubb TC. 2004. Age-specific reproductive success: evidence for
658 the selection hypothesis. *Evolution* 58(4): 880-885.

659 Mazerolle MJ, 2016. AICcmodavg: Model selection and multimodel inference based on
660 (Q)AIC(c). R package version 2.0-4. <http://CRAN.R-project.org/package=AICcmodavg>.

661 McClung MR, Seddon PJ, Massaro M, Setiawan A. 2004. Nature-based tourism impacts on
662 Yellow-eyed Penguins *Megadyptes antipodes*: does unregulated visitor access affect
663 fledging weight and juvenile survival? *Biological Conservation* 119: 279-285.

664 McKinlay B. 2001. Hoiho *Megadyptes antipodes* recovery plan, 2000-2025. Threatened Species
665 Recovery Plan 35. Department of Conservation, Wellington. 27p.

666 Mills JA. 1989. Red-billed Gull. In: Newton I, ed. *Lifetime Reproduction in Birds*: p. 387-404.
667 London: Academic Press.

668 Moore PJ. 1994. What is a Bad Season for Yellow-eyed Penguins? Conservation Advisory
669 Science Notes No. 103. Department of Conservation, Wellington. 7p.

670 Moore PJ. 1999. Foraging range of the Yellow-eyed Penguin *Megadyptes antipodes*. *Marine*
671 *Ornithology* 27: 49–58.

672 Moreno J, De Leon A, Fargallo JA, Moreno E. 1998. Breeding time, health and immune
673 response in the chinstrap penguin (*Pygoscelis antarcticus*). *Oecologia* 115: 312-319.

674 Moreno J. 2003. Lifetime reproductive success in seabirds: interindividual differences and
675 implications for conservation. *Scientia Marina* 67: 7-12.

676 Newton I. 1988. Age and reproduction in the Sparrowhawk. *In*: Clutton Brock, T.H. (ed.)
677 *Reproductive Success: Studies of Individual Variation in Contrasting Breeding Systems*, p.
678 201-219. Chicago: University of Chicago Press.

679 Newton I. 1989. *Lifetime reproduction in birds*. Academic Press, London. 479p.

680 Newton I. 1995. The contribution of some recent research on birds to ecological understanding.
681 *Journal of Animal Ecology* 64: 675-695.

682 Nisbet ICT, Dann P. 2009. Reproductive performance of little penguins *Eudyptula minor* in
683 relation to year, age, pair-bond duration, breeding date and individual quality. *Journal of*
684 *Avian Biology* 40: 296-308.

685 Ollason JC, Dunnet G. 1978. Age, experience and other factors affecting the breeding success of
686 the Fulmar, *Fulmarus glacialis*, in Orkney. *Journal of Animal Ecology* 47: 961-976.

687 Orians GH. 1969. On the evolution of mating systems in birds and mammals. *American*
688 *Naturalist* 103: 589-603.

689 Partridge L. 1992. Measuring reproductive costs. *Trends in Ecology and Evolution* 7(3): 99-100.

690 Petersen SL, Branch GM, Ainley DG, Boersma PD, Cooper J, Woehler EJ. 2005. Is flipper
691 banding of penguins a problem? *Marine Ornithology*,33: 75-79.

692 Potti J, Canal D, Serrano D. 2013. Lifetime fitness and age-related female ornament signalling:
693 evidence for survival and fecundity selection in the pied flycatcher. *Journal of Evolutionary*
694 *Biology* 26: 1445-1457.

695 Pyle P, Nur N, Sydeman WJ, Emslie SD. 1997. Cost of reproduction and the evolution of
696 deferred breeding in the Western Gull. *Behavioural Ecology* 8: 140-147.

697 R Core Team, 2016. R: A language and environment for statistical computing. R Foundation for
698 Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.

699 Ratz H, Darby JT, Edge K-A, Thompson C. 2004. Survival and breeding of yellow-eyed
700 penguins (*Megadyptes antipodes*), at two locations on Otago Peninsula, South Island,
701 New Zealand, 1991–96. *New Zealand Journal of Zoology* 31: 133-147.

702 Reed TE, Kruuk LE, Wanless S, Frederiksen M, Cunningham EJ, Harris MP. 2008.
 703 Reproductive senescence in a long-lived seabird: rates of decline in late-life performance are
 704 associated with varying costs of early reproduction. *American Naturalist* 171: E89-E101.

705 Richdale LE. 1957. *A Population Study of Penguins*. Oxford University Press, Oxford. 195p.

706 Ryder J. 1980. The influence of age on the breeding biology of colonial nesting seabirds.
 707 Behavior of Marine Animals: *Current Perspectives in Research* 4: 153-168.

708 Saraux C, Le Bohec C, Durant JM, Viblanc VA, Gauthier-Clerc M, Deaune D, Park YH, Yoccoz
 709 NG, Stenseth C, Le Maho Y. 2011. Reliability of flipper-banded penguins as indicators of
 710 climate change. *Nature* 469: 203-206.

711 Saraux C, Friess B, Le Maho Y, Le Bohec C. 2012. Chick-provisioning strategies used by king
 712 penguins to adapt to a multiseasonal breeding cycle. *Animal Behaviour* 84: 675-683.

713 Schubert KA, Mennill DJ, Ramsay SM, Otter KA, Boag PT, Ratcliffe LM. 2007. Variation in
 714 social rank acquisition influences lifetime reproductive success in Black-capped Chickadees.
 715 *Biological Journal of the Linnean Society* 90: 85-95.

716 Schultner J, Kitaysky AS, Gabrielsen GW, Hatch SA, Bech C. 2013. Differential reproductive
 717 responses to stress reveal the role of life-history strategies within a species. *Proc. R. Soc. B*.
 718 280: 20132090. <http://dx.doi.org/10.1098/rspb.2013.2090>

719 Seddon PJ, Davis LS. 1989. Nest-site selection by Yellow-eyed Penguins. *The Condor* 91: 653-
 720 659.

721 Seddon PJ, van Heezik Y, Ellenberg U. 2013. Yellow-eyed penguin. In: Borboroglu PG,
 722 Boersma PD, eds. *Penguins: Natural History and Conservation*. Seattle: University of
 723 Washington Press. 360p.

724 Setiawan AN, Darby JT, Lambert DM. 2004. The use of morphometrics to sex yellow-eyed
 725 penguins. *Waterbirds* 27: 96-27.

726 Setiawan AN, Massaro M, Darby JT, Davis LS. 2005. Mate and territory retention in Yellow-
 727 eyed Penguins. *The Condor* 107: 703-709.

728 Stokes DL, Boersma PD. 1998. Nest-site characteristics and reproductive success in Magellanic
 729 Penguins (*Spheniscus magellanicus*). *The Auk* 115: 34-49.

730 Stokes DL, Boersma PD. 2000. Nesting density and reproductive success in a colonial seabird,
731 the Magellanic Penguin. *Ecology* 81: 2878-2891.

732 Tavecchia G, Pradel R, Boy V, Johnson AR., Cézilly F. 2001. Sex-and age-related variation in
733 survival and cost of first reproduction in Greater Flamingos. *Ecology* 82: 165-174.

734 Trathan PN, García-Borboroglu P, Boersma PD, Bost CA, Crawford RJM, Crossin GT, Cuthbert
735 RJ, Dann P, Davis LS, de la Puente S, Ellenberg U, Lynch HJ, Mattern T, Pütz K, Seddon
736 PJ, Trivelpiece W, Wienecke B. 2015. Pollution, habitat loss, fishing and climate change as
737 critical threats to penguins. *Conservation Biology* 29(1): 31-41.

738 van Heezik Y. 1990. Seasonal, geographical, and age-related variations in the diet of the yellow-
739 eyed penguin (*Megadyptes antipodes*). *New Zealand Journal of Zoology* 17: 201-212.

740 van Heezik Y, Seddon PJ. 1989. Stomach sampling in the yellow-eyed penguin: Erosion of
741 otoliths and squid beaks. *Journal of Field Ornithology* 60: 451-458.

742 Wauters L, Dhont AA. 1989. Body weight, longevity, and reproductive success in red squirrels
743 (*Sciurus vulgaris*). *Journal of Animal Ecology* 58: 637-651.

744 Wendeln H, Becker PH. 1999. Effects of parental quality and effort on the reproduction of
745 common terns. *Journal of Animal Ecology* 68: 205-214.

746 Williams GC. 1957. Pleiotropy, natural selection, and the evolution of senescence. *Evolution* 11:
747 398-411.

748 Wooller R, Bradley J, Serventy D, Skira I. 1988. Factors contributing to reproductive success in
749 Short-tailed Shearwaters *Puffinus tenuirostris*. In: *Proceedings of the International*
750 *Ornithological Congress* 19: 848-856.

751 Wooller RD, Bradley JS, Skira IJ, Serventy DL. 1989. Short-tailed Shearwater. In: Newton, I.
752 (ed.) *Lifetime Reproduction in Birds*. p. 405-417. London: Academic Press.

753 Wooller R, Bradley J, Skira I, Serventy D. 1990. Reproductive success of Short-tailed
754 Shearwaters *Puffinus tenuirostris* in relation to their age and breeding experience. *Journal of*
755 *Animal Ecology* 59: 161-170.

756 Wooller RD, Bradley JS, Croxall JP. 1992. Long-term population studies of seabirds. *Trends in*
757 *Ecology and Evolution* 7(4): 111-114.

758

Table 1. Breeding and recruitment overview of numbers and percentages of individual yellow-eyed penguins marked between 1981 and 2003 at Boulder Beach, Otago Peninsula, New Zealand (n = 2147).

	Number	Percent
Marked as chick or juvenile at Boulder Beach	2147	
Marked chicks that were never resighted	1546	72.0
Marked chicks that were resighted under 2 years	601	28.0
Survived to adulthood (2 years)	441	20.5
<i>Sighted at Boulder Beach</i>	<i>370</i>	<i>17.2</i>
<i>Sighted elsewhere</i>	<i>71</i>	<i>3.3</i>
Attempted breeding at Boulder Beach	264	12.3
Fledged offspring at Boulder Beach	219	10.2
Fledged first-generation offspring that recruited	124	5.8
Fledged successful first-generation offspring	102	4.8

764 **Table 2.** Mean LRS, number of recruits, number of recruits that bred successfully, lifespan, age-
 765 at-first-breeding, breeding lifespan, number of breeding attempts and number of mates of
 766 founding generation female (n = 62) and male (n = 68) yellow-eyed penguins breeding at
 767 Boulder Beach, New Zealand. (Var = variance; se = standard error; min = minimum; med =
 768 median; max = maximum).

Variable	mean	var	se	min	med	max
<i>FEMALES</i>						
LRS	6.82	33.8	0.74	0	6	24
Recruits	2.24	5.01	0.28	0	2	9
Successful recruits	1.52	2.71	0.21	0	1	7
Lifespan (years)	9.44	32.77	0.73	2	7.5	24
Age at first breeding	3.60	3.16	0.23	2	3	12
Breeding lifespan (years)	5.84	26.69	0.66	0	4	17
Breeding attempts	5.79	17.74	0.53	1	4.5	16
Total mates	2.03	1.57	0.16	1	2	7
<i>MALES</i>						
LRS	5.07	22.07	0.57	0	4	23
Recruits	1.38	2.81	0.20	0	1	6
Successful recruits	0.81	1.29	0.14	0	0	5
Lifespan (years)	8.87	25.3	0.61	2	8	21
Age at first breeding	4.09	3.48	0.23	2	3	11
Breeding lifespan (years)	4.78	21.7	0.56	0	3	18
Breeding attempts	4.78	11.60	0.41	1	4	14
Total mates	1.93	1.32	0.14	1	2	6

Table 3. Model-averaged generalised linear mixed effects model of lifetime reproductive success (LRS) and life-history parameters of 130 founding generation yellow-eyed penguins (n = 809 breeding attempts) that were marked at Boulder Beach, New Zealand between 1981 and 2003 (All non-binary data are standardised to have mean = 0 and SD = 1).

Coefficients	Estimate	SE	95% Confidence Interval	Relative Importance
(Intercept)*	2.02	0.04	(1.94, 2.09)	-
Sex (Male) [†]	-0.16	0.04	(-0.25, -0.08)	1.00
z (Age at first breeding) [†]	-0.14	0.04	(-0.21, -0.07)	1.00
z (Lifespan) [†]	0.61	0.03	(0.55, 0.68)	1.00
Sex (Male): z (Lifespan)	0.01	0.03	(-0.07, 0.13)	0.24
z Total Mates	-0.004	0.02	(-0.08, 0.04)	0.23

*Sex (Female) is the reference category

[†] Significant results

Model statement: glmer(LRS ~ zLifespan + zAgeatfirstbreeding + zTotalmates + factor(Sex) + factor(Sex):zAgeatfirstbreeding + factor(Sex):zLifespan + (1|Matecode)

Table 4. Model-averaged generalised linear mixed-effects model of lifetime reproductive success in yellow-eyed penguins marked at Boulder Beach between 1981 and 2003, including breeder quality as well as (All non-binary data are standardised to have mean = 0 and SD = 1).

Coefficients	Estimate	SE	95% Confidence Interval	Relative Importance
(Intercept)*	1.84	0.04	(1.77, 1.91)	-
Breeder quality (higher-quality) [†]	0.40	0.06	(0.29, 0.51)	1.00
z(Age at first breeding) [†]	-0.10	0.03	(-0.15, -0.05)	1.00
z(Lifespan) [†]	0.60	0.04	(0.52, 0.67)	1.00
Breeder quality (higher-quality):z(Lifespan) [†]	-0.21	0.05	(-0.31, -0.11)	1.00
z(Total mates)	0.01	0.02	(-0.03, 0.08)	0.34

*Breeder quality (lower-quality) is the reference category

[†] Significant results

Model statement: `glmer(LRS ~ z(Lifespan) + z(Ageatfirstbreeding) + z(Totalmates) + factor(BreederType) + factor(BreederType):z(Ageatfirstbreeding) + factor(BreederType):z(Lifespan) + (1|Matecode)`

Table 5. Mean LRS, number of first-generation offspring, number of first-generation offspring that bred successfully, lifespan, age-at-first-breeding, breeding lifespan, number of breeding attempts and number of mates of ordinary and high-quality yellow-eyed penguins breeding at Boulder Beach, New Zealand. (Var = variance; se = standard error; min = minimum; med = median; max = maximum). Continued overleaf.

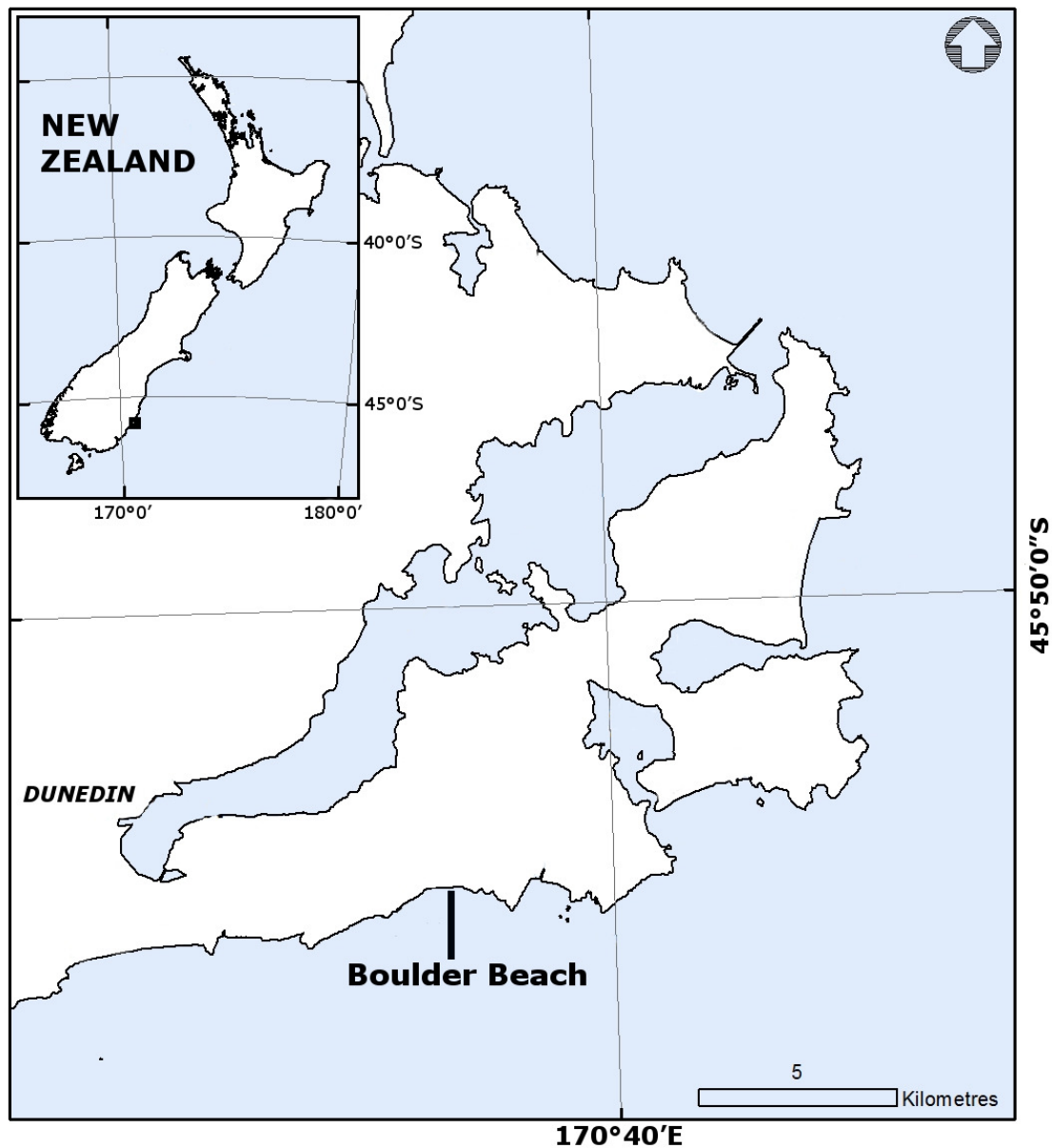
Variable	mean	var	se	min	med	max
<i>Females (ordinary breeders, n = 46)</i>						
LRS	4.37	14.06	0.55	0	4	17
Recruits	1.33	2.27	0.22	0	1	5
Successful recruits	0.83	1.12	0.16	0	0	4
Lifespan (years)	7.48	23.63	0.72	2	6	22
Age at first breeding	3.30	2.57	0.24	2	3	12
Breeding lifespan (years)	4.17	20.46	0.67	0	3	17
Breeding attempts	4.15	9.51	0.45	1	3.5	13
Total mates	1.80	1.36	0.17	1	1	7
<i>Females (high-quality breeders, n = 16)</i>						
LRS	13.88	23.85	1.22	7	12	24
Recruits	4.88	3.58	0.47	2	4.5	9
Successful recruits	3.5	2	0.35	1	3	7
Lifespan (years)	15.06	16.86	1.03	7	16	24
Age at first breeding	4.44	4.13	0.51	3	4	10
Breeding lifespan (years)	10.63	14.25	0.94	4	9.5	17
Breeding attempts	10.5	11.73	0.86	5	9	16
Total mates	2.69	1.70	0.33	1	3	5
<i>Males (ordinary breeders, n = 52)</i>						
LRS	3.42	9.82	0.43	0	2	14
Recruits	0.75	1.09	0.14	0	0	4
Successful recruits	0.33	0.30	0.08	0	0	2
Lifespan (years)	7.58	19.31	0.61	2	6	19
Age at first breeding	4.04	3.45	0.26	2	3	11
Breeding lifespan (years)	3.54	15.43	0.54	0	2.5	14
Breeding attempts	3.87	9.06	0.42	1	3	12
Total mates	1.83	1.52	0.17	1	1	6
<i>Males (high-quality breeders, n = 16)</i>						
LRS	10.44	25.06	1.25	5	8.5	23

Recruits	3.44	2.93	0.43	1	3	6
Successful recruits	2.38	1.32	0.29	1	2	5
Lifespan (years)	13.06	22.6	1.19	8	12	21
Age at first breeding	4.25	3.8	0.49	2	3.5	9
Breeding lifespan (years)	8.81	21.63	1.16	3	6.5	18
Breeding attempts	7.75	8.73	0.74	4	7	14
Total mates	2.25	0.6	0.19	1	2	4

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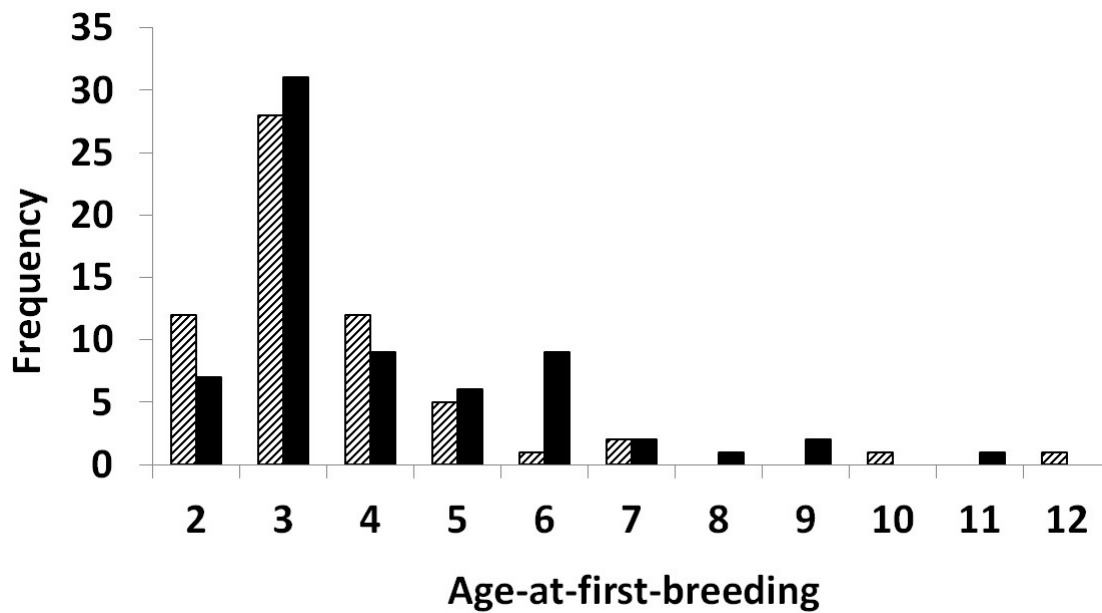
Table 6. Comparison of LRS and maximum number of young fledged by individuals of five different bird species for males and females (where data were available from Coulson 1988; Mills 1989; Dann and Cullen 1990; Korpimäki 1992; Krüger and Lindström 2001; Garamszegi *et al.* 2004).

Species	LRS (Female)	LRS (Male)	Max. fledged (Female)	Max. fledged (Male)
Yellow-eyed penguin (<i>Megadyptes antipodes</i>)	6.82	5.07	24	23
Black-legged kittiwake (<i>Rissa tridactyla</i>)	6.93	7.41	-	-
Red-billed gull (<i>Larus novaehollandiae</i>)	3.4	3	26	28
Little penguin (<i>Eudyptula minor</i>)	2.28	2.13	35	44
Tengmalm's owl (<i>Aegolius funereus</i>)	-	5.2	-	26
Common buzzard (<i>Buteo buteo</i>)	3.48	2.72	20	20
Collared flycatcher (<i>Ficedula albicollis</i>)	5.18	-	-	-



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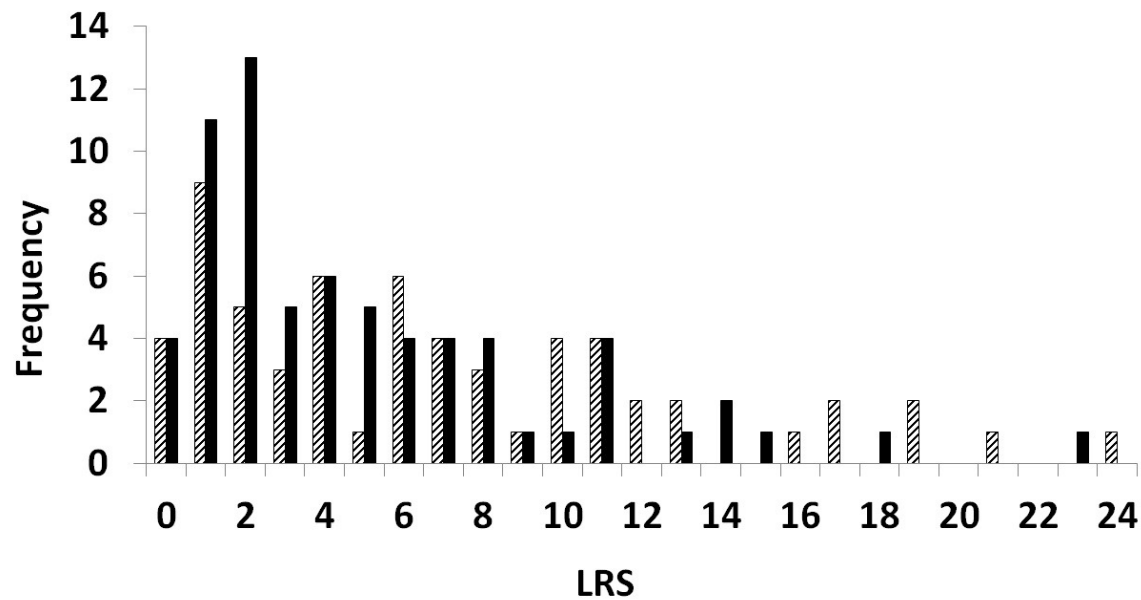
809 **Figure 1.** Map showing the location of Boulder Beach on the Otago Peninsula, Dunedin, New
 810 Zealand.



811

812 **Figure 2.** Age-at-first-breeding of female (striped, $n = 62$) and male (black, $n = 68$) yellow-eyed

813 penguins with complete life histories that bred at Boulder Beach, Dunedin, New Zealand.

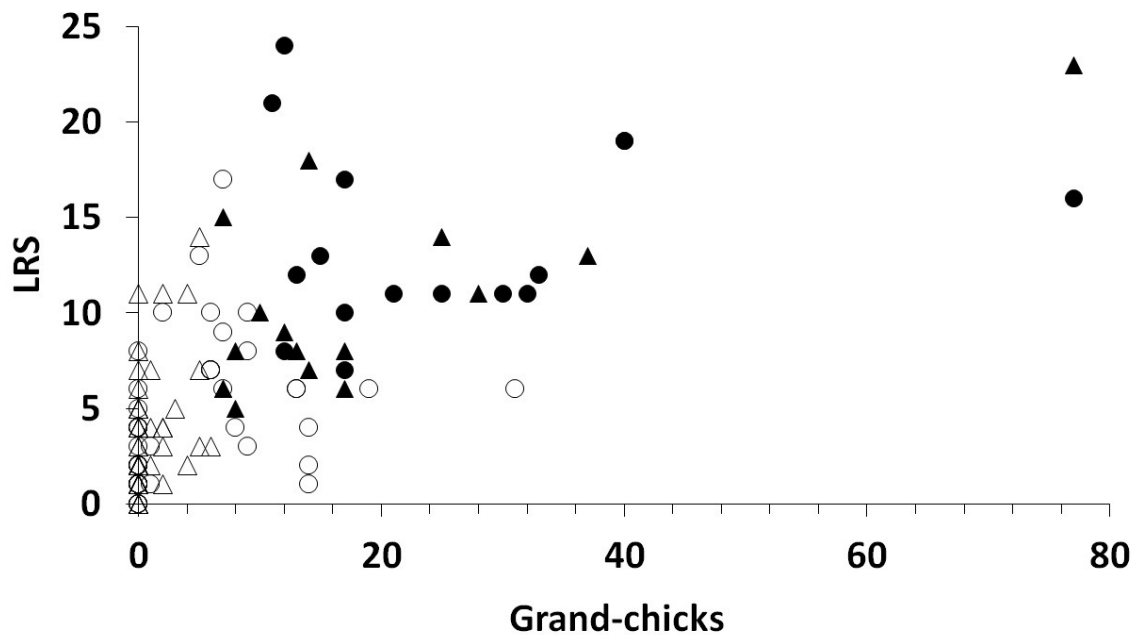


Figure

815

816 **3.** Frequencies of the total number of chicks fledged (lifetime reproductive success, LRS) by
 817 female (striped, $n = 62$) and male (black, $n = 68$) yellow-eyed penguins with complete life
 818 histories that bred at Boulder Beach, New Zealand.

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821

822 **Figure 4.** The relationship between lifetime reproductive success (LRS) and the production of
 823 second-generation offspring (grand-offspring) for female (filled dots = “high-quality breeder”,
 824 empty dots = “low-quality breeder”) and male (filled triangles = “high-quality breeder”, open
 825 triangles = “low-quality breeder”) yellow-eyed penguins with complete life histories that bred at
 826 Boulder Beach, New Zealand.