

# Impact of exposure to environmental tobacco smoke, arsenic and phthalates on locally advanced cervical cancer treatment - preliminary results

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**Background.** Cancer research is a national and international priority, with the efficiency and effectiveness of current anti-tumor therapies being one of the major challenges with which physicians are faced. **Objective.** To assess the impact of exposure to environmental tobacco smoke (ETS), arsenic, and phthalates on cervical cancer treatment. **Methods.** We enrolled 43 patients with locally advanced cervical carcinoma who underwent chemotherapy and radiotherapy. We determined cotinine and five phthalate metabolites in urine samples collected prior to cancer treatment, by gas chromatography coupled to mass spectrometry, and urinary total arsenic by atomic absorption spectrometry with hydride generation. We used linear regression to evaluate the effects of cotinine, arsenic, and phthalates on the change in tumor size after treatment, adjusted for confounding variables. **Results.** We detected no significant associations between urinary cotinine, arsenic, or phthalate monoesters on change in tumor size after treatment, adjusted for urine creatinine, age, baseline tumor size, and cotinine (for arsenic and phthalates). However, higher %mono-ethylhexyl phthalate (%MEHP), a putative indicator of phthalate diester metabolism, was associated with a larger change in tumor size ( $\beta=0.015$ , 95%CI=0.003-0.03,  $P=0.019$ ). **Conclusion.** We found no statistically significant association between the urinary levels of arsenic, cotinine, and phthalates metabolites and the response to cervical cancer treatment as measured by the change in tumor size. Still, our results suggested that phthalates metabolism may be associated with response to treatment for locally advanced cervical cancer. However, these observations are preliminary and will require confirmation in a larger, more definitive investigation.

1 **Impact of exposure to environmental tobacco smoke, arsenic, and phthalates on locally**  
2 **advanced cervical cancer treatment – preliminary results**

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24

25 **Abstract**

26 **Background.** Cancer research is a national and international priority, with the efficiency and  
27 effectiveness of current anti-tumor therapies being one of the major challenges with which  
28 physicians are faced.

29 **Objective.** To assess the impact of exposure to environmental tobacco smoke (ETS), arsenic,  
30 and phthalates on cervical cancer treatment.

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32 chemotherapy and radiotherapy. We determined cotinine and five phthalate metabolites in urine  
33 samples collected prior to cancer treatment, by gas chromatography coupled to mass  
34 spectrometry, and urinary total arsenic by atomic absorption spectrometry with hydride  
35 generation. We used linear regression to evaluate the effects of cotinine, arsenic, and phthalates  
36 on the change in tumor size after treatment, adjusted for confounding variables.

37 **Results.** We detected no significant associations between urinary cotinine, arsenic, or phthalate  
38 monoesters on change in tumor size after treatment, adjusted for urine creatinine, age, baseline  
39 tumor size, and cotinine (for arsenic and phthalates). However, higher %mono-ethylhexyl  
40 phthalate (%MEHP), a putative indicator of phthalate diester metabolism, was associated with a  
41 larger change in tumor size ( $\beta=0.015$ , 95%CI=0.003-0.03, P=0.019)

42 **Conclusion.** We found no statistically significant association between the urinary levels of  
43 arsenic, cotinine, and phthalates metabolites and the response to cervical cancer treatment as  
44 measured by the change in tumor size. Still, our results suggested that phthalates metabolism  
45 may be associated with response to treatment for locally advanced cervical cancer. However,

46 these observations are preliminary and will require confirmation in a larger, more definitive  
47 investigation.

48 **Keywords:** Environmental tobacco smoke, arsenic, phthalates, exposure, locally advanced  
49 cervical cancer, treatment response.

## 50 1. Introduction

51 Cervical cancer is among the leading causes of cancer-related morbidity in women worldwide; in  
52 2012 there were approximately 4,343 new cases in Romania alone, accounting for 1,909 deaths  
53 (Ferlay et al., 2013). Response to treatment consisting of radiotherapy combined with cisplatin-  
54 based chemotherapy in late stages, in most cases of locally advanced cervical cancers, is good,  
55 however, 5-year free survival rates are unsatisfactory (Eifel et al., 2009).

56

57 Human papilloma virus (HPV) infection (particularly types 16 and 18) is present in more than  
58 90% of cervical cancer cases (Bosch et al., 1995), however, not all HPV infections result in  
59 cervical cancer (Burd, 2003). In addition to HPV infection, evidence from epidemiologic studies  
60 also supports an association between active cigarette smoking and cervical neoplasia, suggesting  
61 that exposure to environmental tobacco smoke (ETS) may increase the risk for cervical cancer  
62 (Haverkos et al., 2003; Trimble et al., 2005). Even in non-smokers, lifetime exposure to ETS  
63 may contribute to the development of pre-cancerous cervical intraepithelial neoplasia (Wu et al.,  
64 2003).

65

66 Exposure to arsenic via drinking water has also raised concern worldwide, as it has been  
67 associated with the development of lung, bladder, and skin cancer among others (Smith et al.,  
68 1992). Arsenic induces higher levels of reactive oxygen species (Flora, 2011), an excess of  
69 which peroxidizes lipids and oxidizes proteins and nucleic acids, leading to abnormal function  
70 and DNA lesions (Ercal, Gurer-Orhan & Aykin-Burns, 2001).

71 A growing body of evidence indicates that arsenic may impair the immune system function and  
72 so play a role in carcinogenesis (Acharya et al., 2010). Exposure to arsenic suppressed T and B

73 lymphocyte maturation, induced apoptosis of macrophages and lymphocytes, and impeded T cell  
74 specific cytokine expression (Vega et al., 1999; Cheng et al., 2004). Immune system suppression  
75 promotes cancer development (Acharya et al., 2010). Oncogenesis triggered by arsenic exposure  
76 also leads to immune suppression, further facilitating tumor growth and generating a vicious  
77 circle of positive feedback between cancer and immune function (Acharya et al., 2010), that may  
78 also impact the effectiveness of treatment.

79

80 Phthalates are widely used in many consumer goods, facilitating frequent human exposures  
81 (Kamrin, 2009). Sources of exposure include personal care products that come into contact with  
82 the skin, home improvement products and building materials from which phthalates may be  
83 released into the air, and contamination during the production of food and beverages (Kamrin,  
84 2009). Importantly, phthalates appear to play a role in inflammation, which at a chronic level  
85 precedes tumorigenesis (Anand et al., 2008). Phthalates-mediated increases in chronic  
86 inflammation have been demonstrated in the prostate, uterus, ovary, and breast, all common  
87 locations for neoplastic proliferation (Singh & Li, 2012).

88

89 While environmental factors are likely to contribute as component causes in the etiology of  
90 cancer (Wu et al., 2016), there appears to be little if any data available to assess effects of  
91 environmental pollutants on cancer treatment. As part of a larger investigation into the factors  
92 that modify response to treatment among women with invasive cervical cancer, we conducted an  
93 interim, exploratory analysis to assess the impact of exposures to ETS, arsenic and phthalates. To  
94 the best of our knowledge, no prior investigations have assessed the impact of these widely

95 distributed environmental pollutants on the effectiveness of treatment for locally advanced  
96 cervical cancer.

97

## 98 **2. Materials and Methods**

### 99 *Study population*

100 We recruited 43 women receiving treatment for locally advanced cervical cancer at the Oncology  
101 Institute “Prof Dr Ion Chiricuta” (Cluj-Napoca, Romania), between 2013 and 2014. Patients  
102 were eligible for participation if: 1) diagnosed with histologically-confirmed squamous cell  
103 carcinoma of the cervix at clinical stages IIB, IIIA, or IIIB according to the MD Anderson  
104 Cancer Center modification of the International Federation of Gynecology and Obstetrics  
105 (FIGO) diagnostic criteria (Benedet et al., 2000); 2) aged 20-79 years at the time of diagnosis; 3)  
106 diagnosed with Zubrod score  $\leq 2$  (an overall well-being index scored as 0-5, for which 0 =  
107 asymptomatic and 5 = death) (Oken et al., 1982); 4) blood hemoglobin  $\geq 9$  g/dL, leucocytes  $\geq$   
108  $3000/\text{mm}^3$ , and platelets  $\geq 100\,000/\text{mm}^3$ ; 5) urine creatinine  $< 1.2$  mg/L and urine nitrogen  $< 80$   
109 mg/L; and 6) normal transaminases. We excluded patients with: 1) a history of a prior  
110 malignancy, including previous cervical cancer; 2) interrupted treatment (women who stopped  
111 therapy for any reason); or 3) cardiovascular, kidney, or liver function that was deemed too poor  
112 to initiate treatment. Of the eligible patients contacted, 97.7% enrolled in our study. All  
113 participants provided written informed consent prior to study participation and the research  
114 protocol was approved by the Ethics Committee in Research and Development and Quality  
115 Assurance for Clinical Studies at the Oncology Institute “Prof. Dr. Ion Chiricuta” (approval  
116 stated in the Ethics Committee Evaluation Report no. 6490/2013).

117

118 All study participants underwent three cycles of chemotherapy with Taxol and Carboplatin  
119 (AUC5) followed by radiotherapy, according to the usual clinical cervical cancer treatment  
120 protocol, as previously described in detail (Balacescu et al., 2014). For some patients, additional  
121 surgical treatment included hysterectomy and the removal of positive for malignant cells pelvic  
122 lymph nodes. We assessed treatment response as the difference in tumor size measured before  
123 and after therapy, using a computerized tomography scan. Treatment outcome data were  
124 available for 37 (86%) enrolled participants.

125

#### 126 *Urine samples collection and analysis*

127 We analyzed one urine sample, collected prior to cancer treatment, for each study  
128 participant. Study nurses collected urine specimens at the time of the cancer diagnosis, into 50  
129 mL polyethylene containers previously decontaminated with nitric acid and then rinsed with  
130 water. Within 15 minutes of urine collection, samples were frozen at  $-20^{\circ}\text{C}$  and then transferred  
131 to the Environmental Health Center (Cluj-Napoca, Romania), where they were stored until  
132 analysis for cotinine, arsenic, and five phthalate monoesters. Urinary creatinine was measured  
133 according to a previously described procedure, (Neamtii, Dumitrascu & Roba, 2014), for which  
134 the intra-assay coefficient of variation (CV) was 6.5%.

135

136 *Urinary cotinine* – The analytic method for determination of urinary cotinine was  
137 previously described in detail (Neamtii, Dumitrascu & Roba, 2014). In brief, cotinine was  
138 extracted with dichloromethane, dissolved in toluene and analyzed using a QP 2010 Plus NCI  
139 gas chromatograph (Shimadzu, Japan) coupled to a mass spectrometer (GC-MS) operated in the

140 selective ion monitoring mode. The method limit of detection (LOD) was 10 µg/L and the intra-  
141 assay CV was 3.47%.

142

143 *Urinary arsenic* – The total arsenic concentration in urine samples was analyzed using a  
144 Zeenit 700P atomic absorption spectrometer with hydride generation system (Analytik Jena,  
145 Germany). To determine arsenic, 5 mL of urine was mixed with 5 mL HNO<sub>3</sub> and 2 mL H<sub>2</sub>O<sub>2</sub> and  
146 mineralized using a Mars 6 microwave digester (CEM Corporation, USA). The mineralized  
147 sample was then diluted to 25 mL with ultrapure water. The mineralized sample reacts with  
148 sodium borohydride in an acid environment and forms volatile metal hydrides, which were  
149 atomized in a quartz cell heated at 960°C. After plotting the calibration curve (arsenic specific  
150 wavelength,  $\lambda=193.7\text{nm}$ ), the processed samples were atomized and their absorbencies were  
151 measured. The method LOD was 0.5 µg/L and the intra-assay CV was 2.36%.

152

153 *Urinary phthalates* – We determined urinary phthalate monoester metabolites of dibutyl  
154 phthalate (DBP), including mono-butyl phthalate (MBP), benzyl-butyl phthalate (BzBP),  
155 including mono-benzyl phthalate (MBzP), and di-ethylhexyl phthalate (DEHP), including mono-  
156 (2-ethyl hexyl) phthalate (MEHP), mono-(2-ethyl-5-hydroxyhexyl) phthalate (MEHHP), and  
157 mono-(2-ethyl-5-oxo-hexyl) phthalate (MEOHP), based on a recently published method (Kim et  
158 al., 2014). Briefly, after enzymatic ( $\beta$ -glucuronidase) hydrolysis of urine samples, phthalate  
159 metabolites were extracted with a solvent mixture (hexane, acetone) by sonication. The organic  
160 phase was dried and evaporated to dryness. A derivatization agent, N,O-bis(trimethylsilyl)  
161 trifluoroacetamide with trimethylchlorosilane 1% (BSTFA with TMCS 1%), was added, and the  
162 samples were kept in a thermoreactor (Techne, UK) at 65°C for 1 hour. For the analysis, we used

163 a GC-MS QP 2010 Plus NCI (Shimadzu Corporation) in the single ion monitoring mode. The  
164 method LOD was 2.5 µg/L and intra-assay CVs were 9.64% (MBP), 9.01% (MBzP), 5.43%  
165 (MEHP), 8.7% (MEOHP), and 4.78% (MEHHP).

166

#### 167 *Data analysis*

168 We characterized distributions for exposures and covariates, and imputed urinary cotinine,  
169 arsenic, and phthalates values below the method LODs as  $LOD/\sqrt{2}$  prior to analysis (Hornung &  
170 Reed, 1990). We also calculated %MEHP as  $[MEHP/(MEHP + MEOHP + MEHHP) \times 100]$  on a  
171 molar basis, to assess the impact of phthalates metabolism (Hauser, 2008). To evaluate the  
172 unadjusted effects of environmental exposures on response to cervical cancer therapy, defined as  
173 the difference in tumor size before and after treatment, we used a series of individual regression  
174 models, including only urine creatinine as a covariate in addition to either cotinine, arsenic, or  
175 each phthalate as the sole predictor. We also constructed a series of comprehensive multiple  
176 linear regression models to evaluate the impact of confounder adjusted environmental exposures  
177 on cervical cancer treatment response, including age, baseline tumor size, and urine cotinine  
178 (arsenic and phthalates models) as covariates.

179

180 Given our use of spot collections for biomarkers of exposure, we adjusted for diurnal variation in  
181 urine volume by including creatinine as a covariate in the regression models (Kim et al., 2011).  
182 However, in a second set of regression models, we also used a more ‘traditional’ creatinine  
183 correction, in which urinary cotinine, arsenic, and phthalates were divided by urine creatinine  
184 and the ‘normalized’ variables entered into the regression models. We examined the distribution  
185 of residuals from all regression models to verify the tenability of the normality assumption and to

186 identify outlying and influential observations for further examination. Stata v.12 (StataCorp LP,  
187 College Station, TX USA) was used for the statistical analysis, and statistical significance was  
188 defined as  $p < 0.05$  for a two-tailed test.

189

### 190 3. Results

191 As described in Table 1, 37 participants were 52 years of age on average at the time of the  
192 cervical cancer diagnosis (range 26-76 years). Approximately 59% ( $n = 22$ ) of patients  
193 responded well to the cancer treatment, presenting an 80-100% reduction of the initial tumor  
194 size, although the tumor size was reduced by less than 80%, in 41% of women. Most participants  
195 (59%) were non-smokers (including 13 women who reported ETS exposure), 6 were former  
196 smokers (16%), and 9 (24%) women self-identified as smokers.

197

198 Table 2 shows the distributions for measured urinary cotinine, arsenic, and phthalates levels on a  
199 creatinine basis. Values are reported on a wet-weight basis in the Supplementary Material (Table  
200 S1). Urinary cotinine values ranged from  $< \text{LOD}-395.4 \mu\text{g/g}$  creatinine with a geometric mean of  
201  $13.9 \mu\text{g/g}$  creatinine. Total arsenic values ranged from  $< \text{LOD}-115.4 \mu\text{g/g}$  creatinine with a  
202 geometric mean of  $13.1 \mu\text{g/g}$  creatinine. Using ANOVA, cotinine was higher in smokers  
203 compared to non-smokers ( $p=0.04$ ) and to former smokers ( $p=0.08$ ). On a creatinine basis,  
204 geometric mean values for 35 women with sufficient urine volume available for phthalates  
205 determination were  $8.8 \mu\text{g/g}$  MBP,  $5 \mu\text{g/g}$  MBzP,  $15.8 \mu\text{g/g}$  MEHP,  $3.5 \mu\text{g/g}$  MEOHP, and  $8.7$   
206  $\mu\text{g/g}$  MEHHP. The highest maximum values were measured for MBP ( $295.5 \mu\text{g/g}$ ) and MBzP  
207 ( $182.9 \mu\text{g/g}$ ), whereas maximum values for MEHP ( $91.9 \mu\text{g/g}$ ), MEOHP ( $34.1 \mu\text{g/g}$ ), and  
208 MEHHP ( $88.6 \mu\text{g/g}$ ). The geometric mean %MEHP was 68.2, ranging from a minimum value of

209 10.7% to a maximum value of 97.2%.

210

211 Table 3 describes the multiple linear regression analysis of cervical cancer treatment response,  
212 measured as the change in tumor size, upon environmental exposures, adjusted for confounding  
213 variables. All effect estimates were of small magnitude and 95% confidence intervals (95%CI)  
214 included the null hypothesis; we detected no statistically significant confounder-adjusted  
215 associations between urinary cotinine, total arsenic, or phthalate metabolites and the change in  
216 tumor size. In contrast, the results for the unadjusted regression analysis of total arsenic were  
217 statistically significant ( $\beta=0.01$ , 95% CI 0.0003, 0.02;  $P=0.045$ ), although other results were  
218 similarly null. The results were also similar when we used a traditional creatinine correction  
219 procedure in lieu of adjustment for urinary creatinine as a covariate in regression models  
220 (Supplementary Material Table S2). However, as described by Figure 1 we detected a  
221 statistically significant positive association between %MEHP and response to cervical cancer  
222 therapy, in the multivariable regression model adjusted for confounders ( $\beta=0.015$ ; 95% CI 0.003,  
223 0.03;  $P=0.019$ ). The unadjusted association was not statistically significant for %MEHP  
224 ( $\beta=0.004$ ; 95% CI -0.01, 0.02;  $P=0.634$ ).

225

#### 226 **4. Discussion**

227 We evaluated the impact of exposure to ETS, arsenic, and phthalates on locally advanced  
228 cervical cancer treatment using urine biomarkers from 37 women aged 26-76 years. More than a  
229 half of the study participants responded well to cancer treatment, presenting a reduction of the  
230 initial tumor size between 80% and 100%. We found no meaningful effects on the change in  
231 tumor size for urine cotinine or total arsenic, or for five monoester metabolites of three widely

232 distributed phthalate diesters. Yet, our analysis of %MEHP suggested an enhanced response to  
233 cervical cancer treatment, indicating that less efficient conversion of the primary hydrolytic  
234 DEHP phthalate monoester MEHP, to its secondary-oxidative metabolites MEOHP and  
235 MEHHP, may confer benefit (Hauser, 2008). It is tempting to speculate the existence of a  
236 common metabolic pathway impacting hydrolysis and oxidation of DEHP and response to the  
237 cervical cancer treatment protocol. However, given the limited sample size available for our  
238 study, and the absence of more comprehensive indicators to characterize cervical cancer therapy  
239 response, these observations should be considered preliminary and will require confirmation in a  
240 more definitive investigation.

241

242 Several human biomonitoring studies from nearby European areas reported urinary levels of  
243 cotinine, arsenic, and phthalates measured in general population samples. Geometric mean  
244 urinary cotinine was modestly lower in 120 Romanian mothers 25 to 45 years of age ( $9.10 \mu\text{g/g}$   
245 creatinine) (Lupsa et al., 2015) than for women in our study. However, the geometric mean total  
246 urine arsenic value for 4,730 Germans aged 18 to 69 ( $3.1 \mu\text{g/g}$ ) was substantially lower than for  
247 women in our study (Becker et al., 2003). In our study, we determined urinary metabolites of  
248 DBP, BzBP, and DEHP as those phthalates most prevalent in Central and Eastern European  
249 populations (Černá et al., 2015). The highest urinary phthalates values in our study population  
250 were measured for MBP, MBzP, MEHP and MEHHP, while MEHP had the highest geometric  
251 mean. The geometric mean concentrations of MBzP and MEHP were higher in our study than  
252 were reported for 117 women from the Czech Republic ( $4.35$  and  $3.13 \mu\text{g/g}$  creatinine,  
253 respectively), 125 women from Slovakia ( $3.81$  and  $3.22 \mu\text{g/g}$  creatinine, respectively), and 115  
254 women from Hungary ( $3.82$  and  $3.70 \mu\text{g/g}$  creatinine, respectively) (Černá et al., 2015). In

255 contrast, levels of MEOHP and MEHHP were lower in our study, than reported for Czech (11.68  
256 and 18.45  $\mu\text{g/g}$  creatinine, respectively), Slovakian (11.54 and 18.20  $\mu\text{g/g}$  creatinine,  
257 respectively), and Hungarian (10.37 and 15.54  $\mu\text{g/g}$  creatinine, respectively) women (Černá et  
258 al., 2015). Overall, the distribution of exposures among our study population was unique, yet we  
259 identified similarities with values reported for several nearby, non-clinical European populations.

260

261 Our study had several important limitations, and so, our results should be interpreted only as  
262 hypothesis generating. The small number of study participants may have limited our ability to  
263 detect modest associations, and it precluded an analysis of multiple exposures included in a  
264 single regression model. We hope to expand the sample size in the future to generate more  
265 precise effect estimates and to more comprehensively analyze exposure to the mixture of tobacco  
266 smoke, arsenic, and phthalates. While our study was prospective in nature, we measured  
267 exposure at only a single time point prior to initiation of cancer treatment; the short in vivo  $\frac{1}{2}$ -  
268 lives and the episodic nature of phthalates exposure may have misclassified some patients  
269 (Fromme et al., 2007). Our use of a total arsenic variable, including comparatively innocuous  
270 organic species with toxic inorganic species, may have further misclassified exposure for some  
271 women (Marchiset-Ferlay, Savanovitch & Sauvante-Rochat, 2012). Yet, misclassification is  
272 unlikely to have differed by study outcome and so any bias will have led to underestimated  
273 effects. Finally, we did not incorporate recent data suggesting an important role for gene  
274 expression on cancer treatment response in our study population (Balacescu et al., 2014). Still,  
275 we do anticipate a link to exposure and so bias was unlikely. A larger future investigation should  
276 incorporate longitudinal collection of urine specimens during cancer treatment to reduce  
277 exposure misclassification as well as gene expression information to assess the impact on

278 treatment response.

279

280 To the best of our knowledge, this preliminary report describes the first investigation on the  
281 impact of widely distributed environmental pollutants on the effectiveness of locally advanced  
282 cervical cancer therapy. These results should help to reassure clinicians that even levels of  
283 cotinine, total arsenic, MBzP, and MEHP higher than reported from other European study  
284 populations are unlikely to interfere with the effectiveness of radio and chemotherapy for  
285 invasive cervical cancer. Still, %MEHP may prove important. To better understand the impact of  
286 these results, particularly in women diagnosed with advanced stage cervical cancer, a larger  
287 sample size is needed for detecting potentially modest effects, and to comprehensively assess the  
288 exposure. Current cancer treatments are both expensive and induce serious side effects, and so  
289 characterizing the potential impact of widely distributed environmental pollutants is critical to  
290 establish new indicators for predicting the effectiveness of cervical cancer treatments.

291

## 292 **Conclusions**

293 Exposure to tobacco smoke, arsenic, and phthalates did not appear to impact cervical cancer  
294 treatment at the levels of exposure experienced by our study population. Phthalates metabolism  
295 may be associated with locally advanced cervical cancer treatment response, although the  
296 clinical relevance is unclear. A more comprehensive investigation with a larger sample size will  
297 be necessary for a more definitive result.

298

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302 **References**

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**Table 1** (on next page)

Demographic and clinical characteristics of women receiving cervical cancer treatment and participating in the study (n=37)

1

Characteristics	Mean (n)	SD (%)	Min.	50 <sup>th</sup> %tile	Max.
Age (years)	52.2	11.1	26	54	76
Body mass index (kg/m <sup>2</sup> )	27.1	5.4	17	26.8	41.2
Smoker (yes)	(9)	(24.3)	-	-	-
Tumor Stage					
IIB	(13)	(35.2)	-	-	-
IIIA	(14)	(37.8)	-	-	-
IIIB	(10)	(27)	-	-	-
Change in tumor size (cm)	3.3	1.3	1	3	6

2

**Table 2** (on next page)

Urinary cotinine, arsenic, and phthalates metabolites measured in n=37 women undergoing cervical cancer treatment and participating in the study ( $\mu\text{g/g}$  creatinine)

LOD = method limit of detection; MBP = mono butyl phthalate; MBzP = mono benzyl phthalate; MEHP = mono (2-ethylhexyl) phthalate; MEOHP = mono (2-ethyl-5-oxohexyl) phthalate; MEHHP = mono (2-ethyl-5-hydroxyhexyl) phthalate; %MEHP =  $100 \times (\text{MEHP}/(\text{MEHP} + \text{MEOHP} + \text{MEHHP}))$  on a molar basis

1

Analyte	Min.	25 <sup>th</sup> %tile	50 <sup>th</sup> %tile	75 <sup>th</sup> %tile	Max.	Geometric mean
Cotinine	< LOD	< LOD	9.3	35.3	395.4	13.9
Arsenic	< LOD	7.4	13.2	20.8	115.4	13.1
MBP	< LOD	< LOD	9.8	22	295.5	8.8
MBzP	< LOD	1.8	5.4	10.7	182.9	5
MEHP	1.5	10.6	15.5	28.3	91.9	15.8
MEOHP	< LOD	< LOD	2.8	8.2	34.1	3.5
MEHHP	< LOD	4.6	8.8	13.8	88.6	8.7
%MEHP	10.7	63.6	85.4	91.6	97.2	68.2

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**Table 3**(on next page)

Associations for urinary cotinine, arsenic, and phthalates ( $\mu\text{g/l}$ ) with cervical cancer therapy response, adjusted for covariates using multiple linear regression models <sup>a</sup>

<sup>a</sup> Adjusted for baseline tumor size (cm), age (years), urinary creatinine (mg/l), and urinary cotinine for arsenic and phthalates ( $\mu\text{g/l}$ ); <sup>b</sup> %MEHP =  $100 \times (\text{MEHP}/(\text{MEHP} + \text{MEOHP} + \text{MEHHP}))$  on a molar basis.

MBP = mono butyl phthalate; MBzP = mono benzyl phthalate; MEHP = mono (2-ethylhexyl) phthalate; MEOHP = mono (2-ethyl-5-oxohexyl) phthalate; MEHHP = mono (2-ethyl-5-hydroxyhexyl) phthalate.

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Predictors	n	$\beta$	95% CI		P-value
Cotinine	37	-0.001	-0.005	0.003	0.590
Arsenic	37	0.005	-0.002	0.013	0.173
MBP	35	0.0003	-0.001	0.002	0.752
MBzP	35	-0.001	-0.013	0.010	0.831
MEHP	35	0.0007	-0.006	0.007	0.817
MEOHP	35	-0.016	-0.044	0.012	0.256
MEHHP	35	0.0006	-0.006	0.007	0.855
<b>%MEHP<sup>b</sup></b>	<b>35</b>	<b>0.015</b>	<b>0.003</b>	<b>0.03</b>	<b>0.019</b>

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**Figure 1** (on next page)

Association between %MEHP and response to cervical cancer treatment<sup>a</sup>

<sup>a</sup> Adjusted for initial size of the tumor, urine cotinine, creatinine, and age %MEHP = 100 x (MEHP/(MEHP + MEOHP + MEHHP)) on a molar basis

