

Effects of yearling, juvenile and adult survival on reef manta ray (*Manta alfredi*) demography

Isabel M Smallegange, Isabelle BC van der Ouderaa, Yara Tibiriçá

Background. The trade in manta ray gill plates has considerably increased over the last two decades. The resulting increases in ray mortality, in addition to mortality caused by by-catch, has caused many ray populations to decrease in size. The aim of this study was to ascertain how yearling and juvenile growth and survival, and adult survival and reproduction affect reef manta ray (*Manta alfredi*) population change, in order to increase our understanding of manta ray demography and thereby improve conservation research and measures for these fish.

Methods. We developed a population projection model for reef manta rays, and used published life history data on yearling and juvenile growth and adult reproduction to parameterise the model. Because little is known about reef manta ray yearling and juvenile survival, we conducted our analyses using a range of plausible survival rate values for yearlings, juveniles and adults.

Results. The model accurately captured the observed patterns of variation in population growth rate, lifetime reproductive success and cohort generation time in different reef manta ray populations around the world. Our demographic analyses revealed a range of population consequences in response to variations in demographic rates. For example, an increase in yearling or adult survival rates always elicited greater responses in population growth rate, lifetime reproductive success and cohort generation time than the same increase in juvenile survival rate. The population growth rate increased linearly with increasing yearling or adult survival rates, whereas lifetime reproductive success and cohort generation time increased at an accelerating rate with increasing yearling or adult survival rates. Hence, even a small increase in survival rate could increase lifetime reproductive success by almost one pup, and cohort generation time by up to several years. Elasticity analyses revealed that at the lowest adult survival rate (0.54 per year), the population growth rate is most sensitive to changes in the rate with which juveniles survive but stay juveniles (i.e. do not mature into adults), whereas at higher adult survival rates, such as those observed in populations off the coasts of Mozambique (0.68 per year) and Japan (0.95 per year), it is most sensitive to changes in the adult survival rate.

Discussion. It is important to gain an in-depth understanding of reef manta ray life histories, particularly of yearling and adult survival rates, as these can influence reef manta ray population dynamics in a variety of ways. For declining populations in particular, it is crucial to know which life stage should be targeted for their conservation. For one such declining population off the coast of Mozambique, adult annual survival rate has the greatest effect on population growth, and by increasing adult survival by protecting adult aggregation sites, this population's decline could be halted or even reversed.

1 **Effects of yearling, juvenile and adult survival on reef manta ray (*Manta alfredi*)**
2 **demography**

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16 **ABSTRACT**

17 **Background.** The trade in manta ray gill plates has considerably increased over the last two
18 decades. The resulting increases in ray mortality, in addition to mortality caused by by-catch, has
19 caused many ray populations to decrease in size. The aim of this study was to ascertain how
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33 survival rate. The population growth rate increased linearly with increasing yearling or adult
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35 accelerating rate with increasing yearling or adult survival rates. Hence, even a small increase in
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37 generation time by up to several years. Elasticity analyses revealed that at the lowest adult
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44 dynamics in a variety of ways. For declining populations in particular, it is crucial to know which
45 life stage should be targeted for their conservation. For one such declining population off the
46 coast of Mozambique, adult annual survival rate has the greatest effect on population growth,
47 and by increasing adult survival by protecting adult aggregation sites, this population's decline
48 could be halted or even reversed.

49 **INTRODUCTION**

50 The global demand for plant and animal products is unsustainable (Berkes et al., 2006; Lenzen et
51 al., 2012), and trading in shark fins (Clarke et al., 2006), swim bladders (Sadovy & Cheung,
52 2003; Clark, 2004), sea cucumbers (Anderson et al., 2011) and seahorses (Foster & Vincent,
53 2004) occurs on a large scale. Since 1998, trading in products derived from manta and devil rays
54 has increased exponentially (Ward-Paige, Davis & Worm, 2013). Ray gill plates are a key
55 ingredient in traditional Chinese medicine, and cartilage serves as a filler in shark fin soup
56 (White et al., 2006; Ward-Paige, Davis & Worm, 2013). The exploitation of ray species has
57 resulted in population declines (Marshall et al., 2011a; Couturier et al., 2012), and increases their
58 risk of extinction. As a result, some rays, including the reef manta ray *Manta alfredi* and giant
59 manta ray *M. birostris*, are now listed as ‘Vulnerable’ on the International Union for
60 Conservation of Nature (IUCN) Red List of Threatened Species (Marshall et al., 2011a). Reef
61 manta rays have a slow life history, i.e. late maturity, a long gestation period and a low mean
62 lifetime reproductive success (Marshall et al., 2011a). Therefore, once a reef manta ray
63 population starts to decrease or contains critically few individuals (e.g. due to overfishing), it is
64 very difficult for the population to recover (Kashiwagi, 2014). Therefore, understanding how
65 manta ray populations’ growth rates are affected by variations in demographic rates such as
66 growth, survival and fertility is particularly important (Couturier et al., 2014; Kashiwagi, 2014).

67 Recently, *M. alfredi* and *M. birostris* were listed on Appendix II of the Convention on
68 International Trade in Endangered Species of Wild Fauna and Flora (CITES), meaning that any
69 international trade in manta rays from September 2014 onward must be regulated. However, in
70 many countries, particularly developing ones (e.g. Sri Lanka and countries in East Africa, such
71 as Mozambique), manta ray populations are decreasing at an alarming rate (Marshall et al.,

72 2011a; Ward-Paige, Davis & Worm, 2013). Although manta ray ecotourism occurs in many of
73 these regions, only in 32% of them are manta rays protected (Ward-Paige, Davis & Worm,
74 2013). For example, despite their importance in ecotourism (Tibiriçá et al., 2011), manta rays are
75 not protected under Mozambique law, despite the fact that there has been an 88% decrease in
76 reef manta ray sightings off Praia do Tofo, Mozambique (Rohner et al., 2013). In addition, the
77 main reef manta ray aggregation areas off the coast of southern Mozambique are not inside
78 marine protected areas (Pereira et al., 2014), and there has been a rapid increase in the use of gill
79 nets by artisanal fisheries within inshore regions that are frequented by the rays, which has
80 significantly increased reef manta ray by-catch (Marshall, Dudgeon & Bennett, 2011b; Pereira et
81 al., 2014). A comprehensive understanding of reef manta ray demographics, and their responses
82 to different mortality regimes, is therefore urgently needed to improve conservation efforts and
83 management policies (Ward-Paige, Davis & Worm, 2013).

84 Although manta rays are often easy to approach, we currently do not have sufficient
85 demographic data to fully understand their population dynamics (Ward-Paige, Davis & Worm,
86 2013). If conservation management policies are to be effective, knowing which age classes
87 (yearlings, juveniles or adults) within a population are the most sensitive to disturbance is
88 essential. For example, demographic analyses of the population dynamics of other long-lived
89 organisms, such as turtles and killer whales (*Orcinus orca*), have revealed that population
90 persistence is most sensitive to adult survival, whereas protecting young (e.g. through protective
91 rearing schemes) has a much smaller impact on population persistence (Brault & Caswell, 1993;
92 Heppell, Crowder & Crouse, 1996). Therefore, a very small decrease in the annual survival rate
93 of juveniles and/or adults likely has serious repercussions on the persistence of populations of
94 long-lived species such as manta rays (Ward-Paige, Davis & Worm, 2013; Kashiwagi, 2014).

95 The aim of this study was to ascertain how yearling and juvenile growth and survival, and
96 adult reproduction and survival, affect populations of reef manta rays. To this end, we developed
97 a stage-structured population projection model (PPM) (Caswell, 2001) that we parameterised
98 using published life history data obtained from populations off the coasts of southern
99 Mozambique (Marshall, Dudgeon & Bennett, 2011b) and the Yaeyama Islands, Japan
100 (Kashiwagi, 2014). Sufficient data were available to parameterise growth and reproduction in the
101 PPM, but detailed information on the survival of yearling and juvenile reef manta rays is scarce
102 (Marshall et al., 2011a; Dulvy et al., 2014); therefore, we used the model to investigate how
103 different annual survival rates of yearlings, juveniles and adults affect the population growth
104 rate, mean lifetime reproductive success and cohort generation time. We assessed the
105 performance of the model by comparing the predicted values of these three population biology
106 descriptors with empirical observations. Subsequently, we conducted elasticity analyses for all
107 combinations of yearling, juvenile and adult survival rates to ascertain which demographic rate
108 (rate at which individuals survive and stay in the same life stage; survive and grow into the next
109 life stage; reproduce offspring) of which life stage (yearling, juvenile or adult) has the greatest
110 influence on the population growth rate. Elasticity analysis is widely used by conservation
111 biologists, because the results obtained can be used to prioritise conservation research and
112 management for those life stages that have the greatest effect on population growth (Benton &
113 Grant, 1999; Carslake, Townley & Hodgson, 2009). Because much less is known about yearling
114 and juvenile survival rates than adult survival rates (Marshall et al., 2011a; Dulvy et al. 2014),
115 investigating a range of yearling and juvenile survival rates will elucidate if and how reef manta
116 ray population responses vary with variations in survival rates. For all of the combinations of
117 yearling, juvenile and adult survival rates, we used the calculated population growth rates to

118 project a population of 500 individuals forward over a period of 10 years, in order to investigate
119 the population consequences of different yearling, juvenile and adult mortality regimes.

120

121 **METHODS**

122 ***M. alfredi* life cycle**

123 The life cycle of reef manta rays is generally divided into three life stages: yearlings, non-
124 reproducing juveniles and reproducing adults (Fig. 1) (Marshall et al., 2011a; Kashiwagi, 2014).
125 Male manta rays reach maturity after six years, and females are thought to mature at 8–10 years
126 of age; longevity is estimated to be at least 40 years (Marshall et al., 2011a). On average, adult
127 females produce one pup every two years, but fertility rates can range from one pup every one to
128 five years (Marshall et al., 2011a). Reef manta ray life history data have been collected from
129 various populations, including those off the coasts of Mozambique and the Yaeyama Islands,
130 Japan (Table 1). These two populations differ remarkably in their estimated annual survival rates
131 and population growth rates; the population off the coast of Japan is stable, and juveniles and
132 adults exhibit high survival rates (0.95 per year) (Kashiwagi, 2014), whereas the population off
133 the coast of Mozambique is declining, and the adult survival rate is estimated to be as low as
134 0.68 ± 0.147 SE (standard error) per year (Marshall, Dudgeon & Bennett, 2011b) (Table 1). In
135 the present study, we used the life history data of these two populations to serve as reference
136 points for our demographic analyses.

137

138 **Population model**

139 The population model was based on a three-stage life cycle (Fig. 1). The addition of further life
140 stages may have increased model accuracy, but these are the only currently distinguishable

141 stages in *M. alfredi*. The rate at which individuals survive and remain in the same life stage (as
 142 opposed to growing into the next life stage) equals P_i , where i indicates Y (yearling), J (juvenile)
 143 or A (adult), and was calculated following Caswell (2001):

144

$$145 \quad P_i = \sigma_i(1 - \gamma_i) \quad (\text{Equation 1})$$

146

147 where σ_i ($i = Y, J, A$) is the estimated survival rate for each life stage (Table 1). The parameter γ_i
 148 is the transition rate from one life stage to the next (expressed per year); in this case, from
 149 yearling to juvenile (γ_Y) or from juvenile to adult (γ_J). Each transition rate γ_i was calculated as γ_i
 150 $= 1/D_i$, where D_i is the duration (in years) of either the yearling ($i = Y$) or juvenile life stage ($i =$
 151 J) (Table 1). The rate (per year) at which individuals survive and grow into the next life stage is
 152 defined as:

153

$$154 \quad G_i = \sigma_i \gamma_i \quad (\text{Equation 2})$$

155

156 where i indicates Y (yearling) or J (juvenile). The number of offspring produced each year equals
 157 F_A . These equations result in the following population projection matrix, which has a projection
 158 interval of one year:

159

$$160 \quad \mathbf{A} = \begin{bmatrix} P_Y & 0 & F_A \\ G_Y & P_J & 0 \\ 0 & G_J & P_A \end{bmatrix} \quad (\text{Equation 3})$$

161

162 **Parameterisation and model performance**

163 Following Kashiwagi (2014), and as is common practice (Caswell [2001]), the population model
164 was parameterised for females under the assumption that their growth and survival rates are not
165 too dissimilar to those of male reef manta rays. We set the stage transition rates γ_i in Equations 1
166 and 2 constant at $\gamma_Y = 1/D_Y = 1/1 = 1$ and $\gamma_J = 1/D_J = 1/8 = 0.125$ (Table 1), and assumed
167 that females produce one pup every two years, so that $F_A = 0.5$ per year. Because little is known
168 about yearling and juvenile survival rates (Marshall et al., 2011a; but see Kashiwagi, 2014), we
169 conducted each demographic analysis (explained in the next section) for all combinations of
170 values of yearling annual survival rate (σ_Y) and juvenile annual survival rate (σ_J) within the
171 interval $[0.5, 1]$ in increments of 0.005 (Table 1). We conducted each analysis using the observed
172 adult annual survival rate of reef manta rays off the coast of Mozambique, which is $\sigma_A = 0.68$
173 (Marshall, Dudgeon & Bennett, 2011b), a 20%-reduced adult annual survival rate of $\sigma_A = 0.54$,
174 and 20%- and 40%-increased adult annual survival rates of $\sigma_A = 0.82$ and $\sigma_A = 0.95$, respectively
175 (Table 1). The final value of $\sigma_A = 0.95$ is equal to the observed non-yearling annual survival rate
176 of reef manta rays in a stable population off the coast of Japan (Kashiwagi, 2014) (Table 1). To
177 assess the performance of our population model, we compared our predictions of population
178 growth rate, lifetime reproductive success and cohort generation time with the empirical
179 observations.

180

181 **Demographic analyses**

182 Firstly, we calculated the population growth rate λ from the dominant eigenvalue of matrix **A**
183 (Equation 3) for each of the abovementioned combinations of yearling, juvenile and adult annual
184 survival rates. Secondly, for each of the survival rate combinations, we performed an elasticity
185 analysis to investigate how sensitive the population growth rate λ is to perturbations of each of

186 the different growth, survival and fertility rates in the population projection matrix \mathbf{A} (Equation
 187 3). To this end, we calculated the elasticity matrix \mathbf{E} , where each element on row m and column
 188 n of matrix \mathbf{E} , e_{mn} , represents the proportional contribution of each associated demographic rate
 189 P_i , G_i and F_A in the population projection matrix \mathbf{A} (Equation 3) to the population growth rate λ .
 190 The elasticities were calculated as follows (Caswell, 2001):

191

$$192 \quad e_{mn} = \frac{a_{mn}}{\lambda} \frac{\partial \lambda}{\partial a_{mn}} \quad (\text{Equation 4})$$

193

194 where a_{mn} are the elements of \mathbf{A} . The second part of the equation describes the sensitivities of λ
 195 to changes in the elements a_{mn} of \mathbf{A} (Caswell, 2001). The elasticities sum to 1, and give the
 196 proportional contributions of the matrix elements to the population growth rate λ . Therefore, the
 197 higher an elasticity value is relative to other elasticity values, the greater is the effect of the
 198 associated demographic rate on the population growth rate.

199 Thirdly, for each combination of yearling, juvenile and adult annual survival rates, we
 200 calculated mean lifetime reproductive success (R_0) by taking the dominant eigenvalue of the
 201 matrix $\mathbf{R} = \mathbf{FN}$. The matrix \mathbf{F} is a fertility matrix that describes the production of new
 202 individuals:

203

$$204 \quad \mathbf{F} = \begin{bmatrix} 0 & 0 & F_A \\ 0 & 0 & 0 \\ 0 & 0 & 0 \end{bmatrix} \quad (\text{Equation 5})$$

205

206 The matrix \mathbf{N} is calculated as $\mathbf{N} = (\mathbf{I} - \mathbf{U})^{-1}$, where \mathbf{I} is the identity matrix and \mathbf{U} the transient
 207 matrix that describes the growth and survival rates of the different stages:

208

$$209 \quad \mathbf{U} = \begin{bmatrix} P_Y & 0 & 0 \\ G_Y & P_J & 0 \\ 0 & G_J & P_A \end{bmatrix} \quad (\text{Equation 6})$$

210

211 Fourthly, for each combination of yearling, juvenile and adult annual survival rates, we
 212 calculated cohort generation time as the mean age of offspring production in a cohort of
 213 yearlings (Caswell, 2009):

214

$$215 \quad T_c = \text{diag}(\mathbf{F}\mathbf{N}\mathbf{e}_Y)^{-1} \mathbf{F}\mathbf{N}\mathbf{U}\mathbf{N}\mathbf{e}_Y, \quad (\text{Equation 7})$$

216

217 where the vector \mathbf{e}_Y is a vector with 1 in the first entry (for yearlings) and zeros in the second and
 218 third entries for juveniles and adults, respectively. Finally, we used the population growth rates
 219 that were calculated at step one to project a population of 500 individuals forward over a period
 220 of 10 years. All of the demographic analyses were conducted in MATLAB[®] R2014b
 221 (MathWorks[®], MA, USA).

222

223 RESULTS

224 Model performance

225 Overall, the predictions from our PPM matched the empirical observations well. Firstly, the
 226 predicted values for the population growth rate λ ranged from 0.64 to 1.13, depending on the
 227 values of yearling, juvenile and adult survival rates (Fig. 2; Table 2). This range includes the
 228 range of observed population growth rate values, but also slightly exceeds the range of observed
 229 values (Table 2). Similarly, the range of predicted values of lifetime reproductive success R_0

230 (0.06–6.20) (Fig. 3; Table 2) included the range of observed values of R_0 , but the highest
231 predicted value of R_0 exceeded the highest observed value of R_0 (Table 2). The predicted values
232 for cohort generation time were very low (Fig. 4; Table 2), and much lower than the observed
233 cohort generation times in most cases (Table 2). Only when adult annual survival rate was at its
234 highest ($\sigma_A = 0.95$) (Fig. 4D) did the predicted cohort generation time match the observed value
235 (Table 2).

236

237 **Summary of the demographic analyses**

238 Because the results of our demographic analyses are complex, we first provide a summary to aid
239 in the interpretation of the specific results (below). Because little is known about survival rates of
240 yearling and juvenile reef manta rays, we explored the effects of a range of values of yearling
241 and juvenile annual survival rates on lifetime reproductive success, population growth rate and
242 cohort generation time. We also varied the adult annual survival rate from as low as 0.54, which
243 is 20% lower than the observed annual survival rate of adults (0.68 per year) off the coast of
244 Mozambique (Marshall, Dudgeon & Bennett, 2011b), to as high as 0.95 per year, which equals
245 the observed adult annual survival rate in the stable population off the coast of the Yaeyama
246 Islands (Kashiwagi, 2014). The effect of an increase in adult annual survival rate across this
247 range of values was straightforward: with increasing adult annual survival rate, the values of all
248 three population descriptors also increased. However, variations in yearling and juvenile annual
249 survival rates had different and varying effects on the population descriptors that we
250 investigated. In the case of population growth rate, changes in these two survival rates had
251 additive effects on the population growth rate, but interactive (multiplicative) effects on mean
252 lifetime reproductive success, whereas cohort generation time was unaffected by variations in the

253 juvenile annual survival rate. In addition, the effect of an increase in juvenile annual survival rate
254 was always of a far greater magnitude on population growth rate, mean lifetime reproductive
255 success and cohort generation time than the effect that the same increase in yearling or adult
256 annual survival rate had on these population descriptors.

257

258 **Specific results of the demographic analyses**

259 Calculating the population growth rate λ for all of the different values of yearling, juvenile and
260 adult annual survival rates revealed that for the observed adult annual survival rate of $\sigma_A = 0.68$
261 (Marshall, Dudgeon & Bennett, 2011b), populations can only persist if both yearling and
262 juvenile annual survival rates are high ($\sigma_Y > 0.7$ and $\sigma_J > 0.95$) (Fig. 2B; populations persist to
263 the right of the blue line, indicating population stability at $\lambda = 1$). When applying the lower value
264 of adult annual survival rate ($\sigma_A = 0.54$), populations can only persist if both yearling and
265 juvenile annual survival rates are almost at unity (Fig. 2A; populations persist to the right of the
266 blue line, indicating population stability at $\lambda = 1$). At higher values of σ_A ($\sigma_A = 0.82$ and $\sigma_A =$
267 0.95), populations can persist at much lower values of yearling and juvenile annual survival rates
268 (Fig. 2C, D; populations persist to the right of the blue line, indicating population stability at $\lambda =$
269 1), e.g. if $\sigma_A = 0.95$, the yearling survival rate (σ_Y) can be as low as 0.5 , as long as the juvenile
270 survival rate is $\sigma_J = 0.8$ (Fig. 2D). Because the isoclines in each panel are neither horizontal nor
271 vertical, we can infer that for a constant value of σ_Y (or σ_J), the population growth rate depends
272 on what the value of σ_J (or σ_Y) is. However, because the isoclines in each plot are parallel, we
273 can infer that these effects are additive, and that therefore there is no interactive effect between
274 σ_Y and σ_J on λ (i.e. the magnitude of an effect of σ_Y on λ does not depend on the value of σ_J , and
275 vice versa).

276 We then investigated how variations in yearling, juvenile and adult survival rates affected
277 the elasticity of the population growth rate λ to each of the demographic rates of each life stage
278 in the PPM (Equation 3). This revealed that, depending on the survival rate values, λ was either
279 most sensitive to P_A , the rate at which adults survive and remain in the adult stage, or P_J , the rate
280 at which juveniles survive and remain in the juvenile life stage (Fig 2; white areas in each panel
281 denote survival rate values where λ is most sensitive to P_A , and grey areas denote survival rate
282 values for which λ is most sensitive to P_J). Interestingly, with increasing values of the adult
283 annual survival rate σ_A (going from Fig. 2A to Fig. 2D), the region of yearling survival rate (σ_Y)
284 values for which λ is most sensitive to P_A (white areas) increases, whereas the region of yearling
285 survival rate values for which λ is most sensitive to P_J (grey areas) decreases. These shifts
286 indicate that the elasticity results were independent of juvenile annual survival rate (σ_J); rather,
287 whether or not λ is most sensitive to perturbations in P_J or P_A critically depended on the values
288 of σ_Y and σ_A (Fig. 2).

289 We then investigated the effect of variations in yearling, juvenile and adult survival rates
290 on mean lifetime reproductive success. The results were qualitatively similar to those obtained
291 for the population growth rate: with increasing values of the adult annual survival rate σ_A ,
292 populations can persist at ever lower values of yearling and juvenile annual survival rates (Fig. 3;
293 populations persist to the right of the blue line, indicating population stability at $R_0 = 1$). In
294 contrast to the results for population growth rate, however, the isoclines in each panel are not
295 parallel and are unevenly spaced (Fig. 3), indicating that the yearling and juvenile annual
296 survival rates σ_Y and σ_J have an interactive effect on lifetime reproductive success, i.e. the
297 magnitude of an effect of σ_Y on lifetime reproductive success depends on the value of σ_J , and
298 vice versa. The uneven spacing of the isoclines for each value of adult annual survival rate (Fig.

299 3) indicates that, with increasing values of yearling and juvenile annual survival rates, lifetime
300 reproductive success increases at an ever higher rate. This increase in lifetime reproductive
301 success is greater with increasing values of yearling or adult annual survival rates than with
302 increasing values of juvenile annual survival rates.

303 Regarding cohort generation time, for each value of adult annual survival rate (σ_A),
304 cohort generation time increases at an accelerating rate with increasing values of yearling annual
305 survival rate (σ_Y); hence, a slight increase in a high value of σ_Y results in a far larger increase in
306 cohort generation time than a slight increase in a low value of σ_Y . However, there is no effect of
307 juvenile annual survival rate (σ_J), because the increase in cohort generation time with increasing
308 values of σ_Y is the same for each value of σ_J (Fig. 4). Cohort generation time also increases at an
309 accelerating rate with increasing values of adult annual survival rate (σ_A); consequently, a slight
310 increase in a high value of σ_A results in a far larger increase in cohort generation time than a
311 slight increase in a low value of σ_A (Fig. 4).

312 Finally, we used the predicted population growth rates (Fig. 2) to project a starting
313 population of 500 individuals forward over 10 years to investigate the population consequences
314 of variations in yearling, juvenile and adult survival rates. The combinations of yearling, juvenile
315 and adult survival rate values at which populations are stable and the projected population size
316 remains at 500 individuals after 10 years (indicated by the green lines in Fig. 5) were the same as
317 those obtained from the population growth rate (Fig. 2) and lifetime reproductive success (Fig. 3)
318 analyses. In each panel in Fig. 5, combinations of survival rate values to the right of the green
319 line correspond to population increases. Comparing the different panels shows that the increase
320 in population size is greater at higher values of the adult annual survival rate (σ_A) (Fig. 5). The
321 lowest observed population size of reef manta rays off the coast of Mozambique is 149

322 individuals (Marshall, Dudgeon & Bennett, 2011b), and is indicated by red lines in Fig. 5.
323 Matching this lowest observed population size to our population projections reveals that it
324 corresponds to ever lower values of yearling annual survival rate (σ_Y) and juvenile annual
325 survival rate (σ_J) as the adult annual survival rate (σ_A) increases in value. This suggests that the
326 decrease in population size over 10 years is less at higher values of the adult annual survival rate
327 than at lower values.

328

329 **DISCUSSION**

330 **Model performance**

331 In this study, we developed a population model for reef manta rays that we used to conduct a
332 detailed analysis of reef manta ray demography. With this analysis, we aim to increase our
333 understanding of the drivers of population change in reef manta rays, and how perturbations to
334 demographic rates, such as a decrease in survival due to targeted fishing and by-catch, affect
335 their population fluctuations. Different techniques exist that relate the dynamics of populations to
336 the demographic rates of individuals, and include physiologically structured population models
337 (Metz & Diekmann, 1986), delay-differential equation models (Nisbet & Gurney, 2003),
338 individual-based models (Grimm & Railsback, 2005), integral projection models (IPMs)
339 (Easterling, Ellner & Dixon, 2000), and PPMs (Caswell, 2001). These methodologies all link
340 individual state to population structure, but differ in their mathematical approaches. Structured
341 population models such as PPMs and IPMs are particularly useful for investigating how
342 demographic changes affect population dynamics. They are closely and easily linked to field and
343 experimental data, and require relatively straightforward mathematical techniques from matrix
344 calculus (Coulson, 2012). IPMs have the added benefit that they can be used to investigate

345 simultaneous ecological and rapid evolutionary change in quantitative characters, life history
346 evolution and population dynamics (Smallegange & Coulson, 2013). However, IPMs are data
347 hungry, because their parameterisation requires extensive, long-term datasets on the life history
348 trajectories of individuals (Coulson, 2012). Because these data are currently not available for reef
349 manta rays, we developed a PPM that included the three life stages that can currently be
350 distinguished in reef manta rays: yearlings, juveniles and adults (Marshall et al., 2011a). Future
351 studies should, however, aim to develop a reef manta ray IPM that can take any evolutionary
352 responses in life history parameters to environmental change into consideration. We
353 parameterised our PPM using published life history data; therefore, the model outcomes
354 critically depended on the soundness and accuracy of these data. The data that we used to model
355 growth and reproduction are sound, as several studies on the growth and reproduction of reef
356 manta rays report very similar results (Marshall & Bennet, 2010; Marshall, Dudgeon & Bennett,
357 2011b; Kashiwagi, 2014). Less is known about individual survival in this species. Marshall,
358 Dudgeon & Bennett (2011b) studied adult survival in a reef manta ray population off the
359 southern coast of Mozambique, and our own preliminary, capture-mark-recapture analysis of
360 sight-re-sight data of adult reef manta rays off the southern coast of Mozambique (200 km south
361 of the study site of Marshall, Dudgeon & Bennett [2011b]) obtained an adult survival rate of 0.67
362 (± 0.16 SE) per year (I.B.C. van der Ouderaa & Y. Tibiriçá, unpublished data), which is not
363 significantly different to the value of 0.68 (± 0.15 SE) per year obtained by Marshall, Dudgeon
364 & Bennett (2011b). The survival of yearling and juvenile reef manta rays has been less studied,
365 as individuals at these life stages do not frequently visit the aggregation sites where demographic
366 data on adults are typically collected (Marshall, Dudgeon & Bennett, 2011b). For this reason, we
367 used a range of annual survival rate values for yearlings and juveniles.

368 Overall, we found that the performance of our model was satisfactory; mean lifetime
369 reproductive success and population growth rates observed in different reef manta ray
370 populations were all within the ranges that we predicted from our population model. However,
371 the predicted population growth rate and lifetime reproductive success values sometimes
372 exceeded the observed values; this was probably due to the fact that we also investigated the
373 population consequences of annual survival rates of yearlings, juveniles and adults that were
374 lower and higher than the observed survival rates. For the reef manta ray population off the coast
375 of the Yaeyama Islands, the annual survival rates of all three life stages, as well as the population
376 growth rate, have been estimated: the yearling annual survival rate is estimated to be 0.63 and
377 juvenile and adult annual survival rates are both estimated as 0.95 (Kashiwagi, 2014). The
378 population growth rate associated with these values as predicted by our population model was
379 ~ 1.01 (Fig. 2D; $\sigma_Y = 0.63$, $\sigma_J = \sigma_A = 0.95$), which is very close to the actual population growth
380 rate of the Yaeyama Islands reef manta ray population, which is estimated as 1.02 per year
381 (Kashiwagi, 2014). The only discrepancy between the predicted and observed values was cohort
382 generation time at the lower adult annual survival rates of $0.54 \leq \sigma_A \leq 0.82$ (Fig. 4A–C). At these
383 low survival rates, adults do not live very long, which lowers the average age at which adults
384 reproduce and results in a low cohort generation time. Cohort generation time values have
385 probably been obtained from stable populations (Marshall et al., 2011a; Ward-Paige, Davis &
386 Worm, 2013), in which annual adult survival rates are much higher. Indeed, at $\sigma_Y = \sigma_A = 0.95$
387 (as found for the stable reef manta ray population off the coast of the Yaeyama Islands
388 [Kashiwagi, 2014]), the predicted cohort generation time was 18.5 years, which is very close to
389 the observed generation time of 19.4 years (Ward-Paige, Davis & Worm, 2013). Overall, it is

390 rewarding that predictions from our population model match observations on the key population
391 descriptors of lifetime reproductive success, population growth rate and cohort generation time
392

393 **Demographic analyses**

394 The demographic analysis revealed that the effects of variations in yearling and juvenile survival
395 rates on population growth rate, mean lifetime reproductive success and cohort generation time
396 are not straightforward, but some general patterns did emerge. Firstly, an increase in yearling or
397 adult annual survival rate always elicited a greater response in population growth rate, mean
398 lifetime reproductive success and cohort generation time than the same increase in juvenile
399 annual survival rate. This suggests that a perturbation in yearling or adult annual survival rate
400 will have far greater consequences for reef manta ray population dynamics than the same
401 magnitude of perturbation in juvenile annual survival rate. Secondly, increases in any of the three
402 population descriptors with increasing yearling or adult survival rate values was either linear, in
403 the case of population growth rate, or was at an accelerating rate, in the case of mean lifetime
404 reproductive success and cohort generation time. The accelerating rates of increase are of
405 particular interest, because if yearling or adult annual survival rates are already high, a slight
406 increase can increase mean lifetime reproductive success by almost one pup (Fig. 3D), and
407 cohort generation time by a year to several years (Fig. 4D). Both of these effects can
408 significantly affect population structure and fluctuations. Therefore, in order to obtain an
409 accurate insight into reef manta ray population dynamics, accurate estimates of yearling and
410 juvenile survival rates should be obtained from natural populations.

411 One way of gaining a general insight into the population consequences of differences in
412 demographic rates is by using population models to project a population forward in time and

413 investigate its future size relative to its original size, which we did for a period of 10 years for all
414 combinations of yearling, juvenile and adult annual survival rates. The reef manta ray population
415 off the coast of Mozambique decreased by 88% between 2005 and 2011 due to variations in the
416 local environment, anthropogenic pressures and large-scale oceanographic influences (Rohner et
417 al., 2013). Our population projections confirm that the low observed adult annual survival rate of
418 adult reef manta rays off the coast of Mozambique of 0.68 per year (Marshall, Dudgeon &
419 Bennett, 2011b) nearly always results in population decline, unless yearling and juvenile annual
420 survival rates are close to unity. However, because reef manta ray by-catch has recently
421 significantly increased in this region (Marshall, Dudgeon & Bennett, 2011b; Pereira et al., 2014),
422 it is unlikely that the juvenile survival rate is close to unity. In the stable reef manta ray
423 population off the coast of the Yaeyama Islands, the yearling annual survival rate has been
424 estimated to be as low as 0.63, probably because of predation (Kashiwagi, 2014). If we assume
425 that this yearling annual survival rate also applies to the Mozambique population, the
426 Mozambique population is predicted to continue to decrease in size, even if the juvenile annual
427 survival rate is at unity (Fig. 5B). Therefore, unless the survival rates of reef manta rays in
428 populations off the coast of Mozambique increase (by reducing direct fishing and by-catch), it is
429 unlikely that this population will stop declining.

430

431 **Conservation implications**

432 Many manta ray populations across the globe are declining, according to the IUCN Red List for
433 Threatened Species (Marshall et al., 2011a; but see Kashiwagi [2014] for an exception). One
434 way of increasing our understanding of how such declines can be reduced or even halted is by
435 conducting elasticity analyses of a demographic model. The results of such analyses identify

436 which demographic rates of which life stages have the greatest effect on population growth. By
437 targeting conservation research and management on those rates and life stages, vulnerable
438 populations can be protected from further decline (Benton & Grant, 1999; Carslake, Townley &
439 Hodgson, 2009). Our elasticity analysis revealed that the population growth rate is most sensitive
440 to change in either the adult survival rate or the rate at which juveniles survive but stay in the
441 juvenile stage (i.e. do not mature into adults). To exemplify how the results of this analysis can
442 be used, we compared the values predicted by our analysis to the values obtained in two reef
443 manta ray populations off the coasts of Mozambique and the Yaeyama Islands. In the case of
444 Mozambique, the observed adult annual survival rate is 0.68 (Marshall, Dudgeon & Bennett,
445 2011b), and the observed population growth rate is estimated as 0.77 per year (Rohner et al.,
446 2013); according to our elasticity analysis (Fig. 2B), at these values, the population growth rate
447 will be most sensitive to changes in the adult annual survival rate. To prevent this population
448 declining further, the adult survival rate should be increased by reducing targeted and by-catch
449 fishing through the protection of aggregation sites that are often frequented by adults. Our
450 analyses indicate that if the adult survival rate increases to 0.95 per year, then the population
451 growth rate is close to unity and the Mozambique reef manta ray population would be stable.
452 Regarding the stable population off the coast of the Yaeyama Islands, the adult annual survival
453 rate is 0.95, and according to our elasticity analysis, this population will also be most sensitive to
454 changes in the adult annual survival rate. Although this population is not currently subject to
455 direct fishing pressure (Kashiwagi, 2014), our results predict that any changes in adult survival
456 will greatly affect it.

457 A previous demographic analysis that was based on a generic reef manta ray life cycle
458 and not on a specific manta ray population found that the intrinsic population growth rate r was

459 most sensitive to changes in the offspring production rate and not the mortality rate (Dulvy et al.,
460 2014). However, unlike our elasticity analysis, Dulvy et al.'s (2014) sensitivity analysis
461 investigated how *additive* perturbations in life history parameters affect the intrinsic population
462 growth rate, whereas we investigated how *proportional* perturbations in demographic rates affect
463 the long-term population growth rate. We used the second part of Equation 4 to run a sensitivity
464 analysis in order to investigate how *additive* perturbations affect the population growth rate, and
465 found that the population growth rate is most sensitive to perturbations in G_J or P_A , depending on
466 the values of yearling, juvenile and adult annual survival rates (Appendix: Fig. A1). However,
467 these results are not very informative, because the demographic rates in our population model are
468 measured in different units; survival rates are probabilities, and only have values of between zero
469 and unity, whereas reproduction rate has no such restrictions. Therefore, it is difficult to compare
470 the sensitivity of the population growth rate to changes in survival and growth rates with the
471 sensitivity of the reproductive rate. Therefore, we focus on the results of the elasticity analysis,
472 which suggest that reef manta ray populations off the coast of Mozambique and Japan are most
473 sensitive to perturbations in the adult annual survival rate. The demographic rates that comprise
474 our population matrix are determined by the underlying parameters σ_i (survival rate) and γ_i
475 (stage-specific transition rate); however, because the adult annual survival rate P_A equals σ_A and
476 is independent of γ_i , the population growth rate is indeed most sensitive to perturbations in adult
477 survival at high adult survival rates, which is typical for long-lived animals (Brault & Caswell,
478 1993; Caswell, 2001). This indicates that effective management and legislation is urgently
479 needed to avoid the local extinction of the reef manta ray population off the coast of
480 Mozambique. The following two approaches should be taken: (1) the species should be protected
481 against fishing, including accidental catch, at the national level; and (2) aggregation areas should

482 be protected. The behaviour of reef manta rays at cleaning stations makes targeted fishing a
483 potential threat, but also creates an opportunity for site-specific protection. By protecting
484 aggregation sites, adults, which are regular visitors to such sites (Marshall, Dudgeon & Bennett,
485 2011b; Kashiwagi, 2014), should exhibit increased survival rates, which will result in an increase
486 in the population growth rate.

487

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491

492 **REFERENCES**

493 **Anderson SC, Flemming JM, Watson R, Lotze HK. 2011.** Serial exploitation of global sea
494 cucumber fisheries. *Fish and Fisheries* **12**:317–339 DOI 10.1111/j.1467-
495 2979.2010.00397.x.

496 **Benton TG, Grant A. 1999.** Elasticity analysis as an important tool in evolutionary and
497 population ecology. *Trends in Ecology and Evolution* **14**:467–471.

498 **Berkes F, Hughes TP, Steneck RS, Wilson JA, Bellwood DR, Crona B, Folke C, Gunderson**
499 **LH, Leslie HM, Norberg J, Nyström M, Olsson P, Österblom H, Scheffer M, Worm B.**
500 **2006.** Globalization, roving bandits, and marine resources. *Science* **311**:1557–1558 DOI
501 10.1126/science.1122804.

502 **Brault S, Caswell H. 1993.** Pod-specific demography of killer whales (*Orcinus orca*). *Ecology*
503 **74**:1444–1454.

- 504 **Carslake D, Townley S, Hodgson DJ. 2009.** Patterns and rules for sensitivity and elasticity in
505 population projection matrices. *Ecology* **90**:3258–3267.
- 506 **Caswell H. 2001.** *Matrix population models*. Sunderland: Sinauer Associates.
- 507 **Caswell H. 2009.** Stage, age and individual stochasticity in demography. *Oikos* **118**:1763–1782.
- 508 **Clarke SC. 2004.** Understanding pressures on fishery resources through trade statistics: a pilot
509 study of four products in the Chinese dried seafood market. *Fish and Fisheries* **5**:53–74 DOI
510 10.1111/j.1467-2960.2004.00137.x.
- 511 **Clarke SC, McAllister MK, Milner-Gulland EJ, Kirkwood GP, Michielsens CGJ, Agnew
512 DJ, Pikitch EK, Nakano H, Shivji MS. 2006.** Global estimates of shark catches using trade
513 records from commercial markets. *Ecology Letters* **9**:1115–1126 DOI 10.1111/j.1461-
514 0248.2006.00968.x.
- 515 **Coulson T. 2012.** Integral projection models, their construction and use in posing hypotheses in
516 ecology. *Oikos* **121**:1337–1350.
- 517 **Couturier LIE, Dudgeon CL, Pollock KH, Jaine FRA, Bennett MB, Townsend KA, Weeks
518 SJ, Richardson AJ. 2014.** Population dynamics of the reef manta ray *Manta alfredi* in
519 eastern Australia. *Coral Reefs* **33**:329–342 DOI: 10.1007/s00338-014-1126-5.
- 520 **Couturier LIE, Marshall AD, Jaine FRA, Kashiwagi T, Pierce SJ, Townsend KA, Weeks
521 SJ, Bennett MD, Richardson AJ. 2012.** Biology, ecology and conservation of the
522 Mobulidae. *Journal of Fish Ecology* **80**:1075–1119.
- 523 **Dulvy NK, Baum JK, Clarke S, Compagno LJ, Cortés E, Domingo A, Fordham S, Fowler
524 S, Francis MP, Gibson C, Martínez J, Musick JA, Soldo A, Stevens JD, Valenti S. 2008.**
525 You can swim but you can't hide: the global status and conservation of oceanic pelagic
526 sharks and rays. *Aquatic Conservation: Marine and Freshwater Ecosystems* **18**:459–482.

- 527 **Dulvy NK, Pardo SA, Simpfendorfer CA, Carlson JK. 2014.** Diagnosing the dangerous
528 demography of manta rays using life history theory. *PeerJ* **2**:e400; DOI 10.7717/peerj.400
- 529 **Easterling MR, Ellner SP, Dixon PM. 2000.** Size-specific sensitivity: applying a new
530 structured population model. *Ecology* **81**:694–708.
- 531 **Foster SJ, Vincent ACJ. 2004.** Life history and ecology of seahorses: implications for
532 conservation and management. *Journal of Fish Biology* **65**:1–61 DOI 10.1111/j.0022-
533 1112.2004.00429.x.
- 534 **Grimm V, Railsback, SF. 2005.** *Individual-based modeling and ecology*. Princeton: Princeton
535 University Press.
- 536 **Heppell SS, Crowder LB, Crouse DT. 1996.** Models to evaluate headstarting as a management
537 tool for long-lived turtles. *Ecological Applications* **6**:556–565.
- 538 **Kashiwagi T. 2014.** Conservation biology and genetics of the largest living rays: manta rays.
539 PhD Thesis, University of Queensland.
- 540 **Lenzen M, Moran D, Kanemoto K, Foran B, Lobefaro L, Geschke A. 2012.** International
541 trade drives biodiversity threats in developing nations. *Nature* **486**:109–112 DOI
542 10.1038/nature11145.
- 543 **Marshall AD, Bennett MB. 2010.** Reproductive ecology of the reef manta ray *Manta alfredi* in
544 southern Mozambique. *Journal of Fish Biology* **77**:169–190.
- 545 **Marshall AD, Kashiwagi T, Bennett MB, Deakos M, Stevens G, McGregor F, Clark T,**
546 **Ishihara H, Sato K. 2011a.** *Manta alfredi*. The IUCN Red List of Threatened Species
547 Version 2015.2.

- 548 **Marshall AD, Dudgeon CL, Bennett MB. 2011b.** Size and structure of a photographically
549 identified population of manta rays *Manta alfredi* in southern Mozambique. *Marine Biology*
550 **158**:1111–1124.
- 551 **Metz JAJ, Diekmann O. 1986.** *The dynamics of physiologically structured populations.* Berlin:
552 Springer.
- 553 **Nisbet RM, Gurney WSC. 2003.** *Modelling fluctuating populations.* New Jersey: Blackburn
554 Press.
- 555 **Pereira MAM, Litulo C, Santos R, Leal M, Fernandes RS, Tibiriçá Y, Williams J,**
556 **Atanassov B, Carreira F, Massingue A, Marques da Silva I. 2014.** *Mozambique marine*
557 *ecosystems review.* Maputo: Biodinâmica/CTV, 139.
- 558 **Rohner CA, Pierce SJ, Marshall AD, Weeks SJ, Bennett MB, Richardson AJ. 2013.** Trends
559 in sightings and environmental influences on a coastal aggregation of manta rays and whale
560 sharks. *Marine Ecology Progress Series* **482**:153–168.
- 561 **Sadovy Y, Cheung WL. 2003.** Near extinction of a highly fecund fish: the one that nearly got
562 away. *Fish and Fisheries* **4**:86–99 DOI 10.1046/j.1467-2979.2003.00104.x.
- 563 **Smallegange IM, Coulson T. 2013.** Towards a general, population-level understanding of eco-
564 evolutionary change. *Trends in Ecology and Evolution* **28**:143–148.
- 565 **Tibiriçá Y, Birtles A, Valentine P, Miller DK. 2011.** Diving tourism in Mozambique: an
566 opportunity at risk? *Tourism in Marine Environments* **7**:141–151.
- 567 **Ward-Paige CA, Davis B, Worm B. 2013.** Global population trends and human use patterns of
568 *Manta* and *Mobula* rays. *PLoS ONE* **8**:e74835 DOI: 10.1371/journal.pone.0074835.

569 **White WT, Giles J, Dharmadi D, Potter IC. 2006.** Data on the bycatch fishery and
570 reproductive biology of mobulid rays (Myliobatiformes) in Indonesia. *Fisheries Research*
571 **82:65–73.**

Table 1. Life history data of different reef manta ray populations. Shown are annual survival rates, σ_i , duration of different life stages, D_i , where $i = Y$ (yearlings), $i = J$ (juveniles) and $i = A$ (adults), and fertility rate of adults, F_A . Indicated are values estimated from data collected from populations off the coast of southern Mozambique and off the coast of Yaeyama Islands, Japan. Also shown are the values that were used in our demographic analyses.

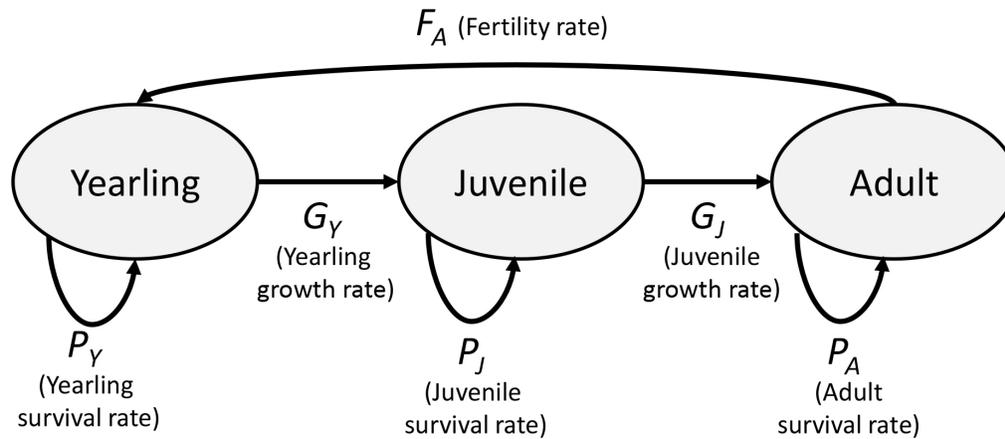
	Explanation	Value in analyses	Observed value	Location of observation	Reference for observed value
σ_Y	Annual survival rate of yearlings	0.5 - 1.0	0.63	Japan	Kashiwagi, 2014
σ_J	Annual survival rate of juveniles	0.5 - 1.0	0.95	Japan	Kashiwagi, 2014
σ_A	Annual survival rate of adults	{0.54, 0.68, 0.82, 0.95}	0.68	Mozambique	Marshall, Dudgeon, & Bennett, 2011b
			0.95	Japan	Kashiwagi, 2014
D_Y	Duration of yearling stage (years)	1	1	not specified/ Japan	Marshall et al., 2011a; Kashiwagi, 2014
D_J	Average duration of (female) juvenile stage (years)	9	8-10	not specified/ Japan	Marshall et al., 2011a; Kashiwagi, 2014
D_A	Duration of adult stage (years)	31	31	not specified/ Japan	Marshall et al., 2011a; Kashiwagi, 2014
F_A	Average number of pups per year	0.5	0.5	Mozambique	Marshall & Bennett 2010

Table 2. Predicted and observed population descriptors for different reef manta ray populations. The population descriptors are: population growth rate (λ , expressed per year), mean lifetime reproductive success (R_0), and cohort generation time (T_c , years). Predicted values given are the minimum and maximum values from our demographic analyses (Fig. 2-4); observed values are taken from different locations around the world (locations are indicated).

	Predicted range	Observed value	Explanation of observed value	Location of observation	Reference for observed value
λ	0.64 - 1.13	0.77	Calculated from the observation of 88% decline between 2005 – 2011	Mozambique	Rohner et al., 2013
		0.98	Calculated from the observation of 80% decline over 75 years	not specified	Marshall et al., 2011a
		1.02	Estimated using POPAN models covering 1987 – 2009	Japan	Kashiwagi, 2014
$R_0^{(1)}$	0.06 – 6.20	0.72	Calculated using IUCN data (Marshall et al. 2011a) ⁽¹⁾ : $T_c = 15$ and $\lambda = 0.98$	not specified	Marshall et al., 2011a
		0.02	Worst-case scenario calculated using slowest life history values ⁽¹⁾ : $T_c = 19.4$ and $\lambda = 0.77$	not specified	Marshall et al., 2011a; Rohner et al., 2013
T_c	3.89 – 20.40	19.4	Mean of minimum (6.75 years) and maximum (32 years) age of adults	Tropical Easter Pacific & Atlantic; Hawaii	Ward-Paige, Ward-Paige, Davis & Worm, 2013
		15	Mean of minimum (10 years) and maximum (40 years) age of adults	not specified	Marshall et al., 2011a

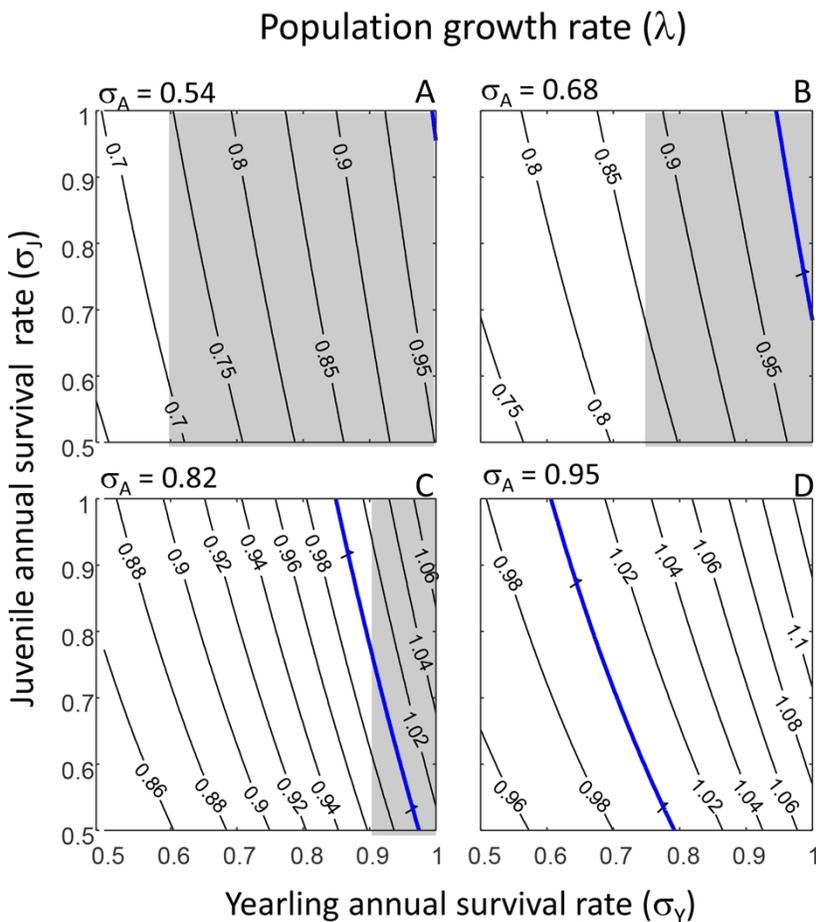
⁽¹⁾ R_0 was calculated by taking the exponent of $T_c \times \log(\lambda)$ (Caswell 2001)

574



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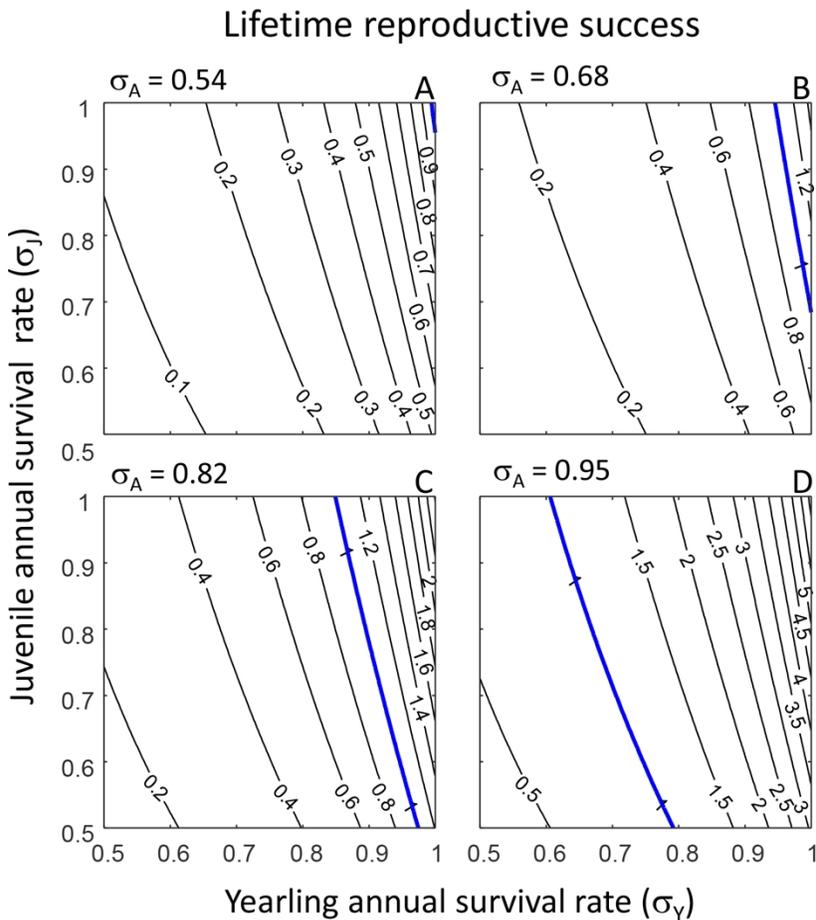
576 **Figure 1 Life cycle of *Manta alfredi*.** We distinguished three life stages: yearlings (Y), juveniles
 577 (J) and adults (A). The rate at which individuals survive and remain in the same life stage equals
 578 P_i , where i indicates Y (yearling), J (juvenile) or A (adult); the rate at which individuals survive
 579 and grow to the next life stage equals G_i , where i indicates Y (yearling) or J (juvenile); the rate at
 580 which adults produce yearlings equals F_A . See also Equations 1–3.



581

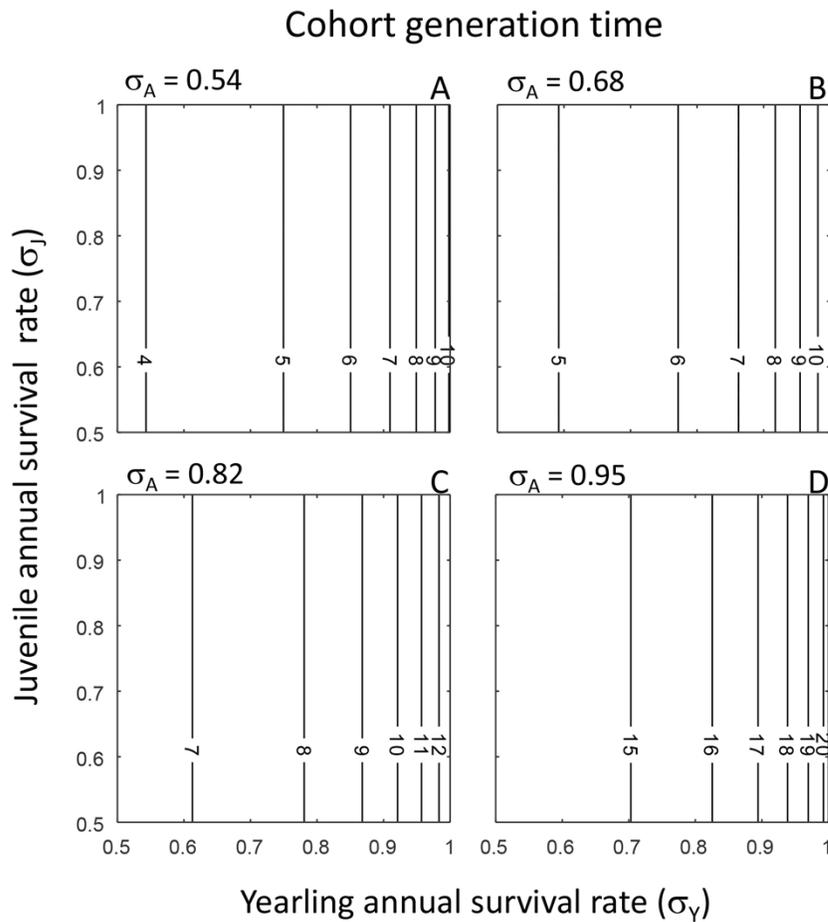
582 **Figure 2 Population growth rate and elasticity results.** Predicted population growth rate λ in
 583 relation to yearling annual survival rate (σ_Y) and juvenile annual survival rate (σ_J) shown for
 584 each of four values of adult annual survival rate (σ_A): $\sigma_A = 0.54$ (80% of observed rate) (A); σ_A
 585 = 0.68 (observed rate) (B); $\sigma_A = 0.82$ (120% of observed rate) (C); and $\sigma_A = 0.95$ (140% of
 586 observed rate) (D). In each panel, isoclines denote equal values of the population growth rate λ .
 587 The blue line in each panel denotes population stability at $\lambda = 1$; values higher than $\lambda = 1$ denote
 588 increasing populations and value lower than $\lambda = 1$ denote declining populations. The grey and
 589 white areas in panels denote the elasticity results: white areas (panel D is all white) denote
 590 parameter combinations where the population growth rate is most sensitive to P_A , the rate at

591 which adults survive and remain in the adult stage (Equation 3); grey areas denote parameter
592 combinations where the population growth rate is most sensitive to P_J , the rate at which juveniles
593 survive and remain in the juvenile life stage (Equation 3).



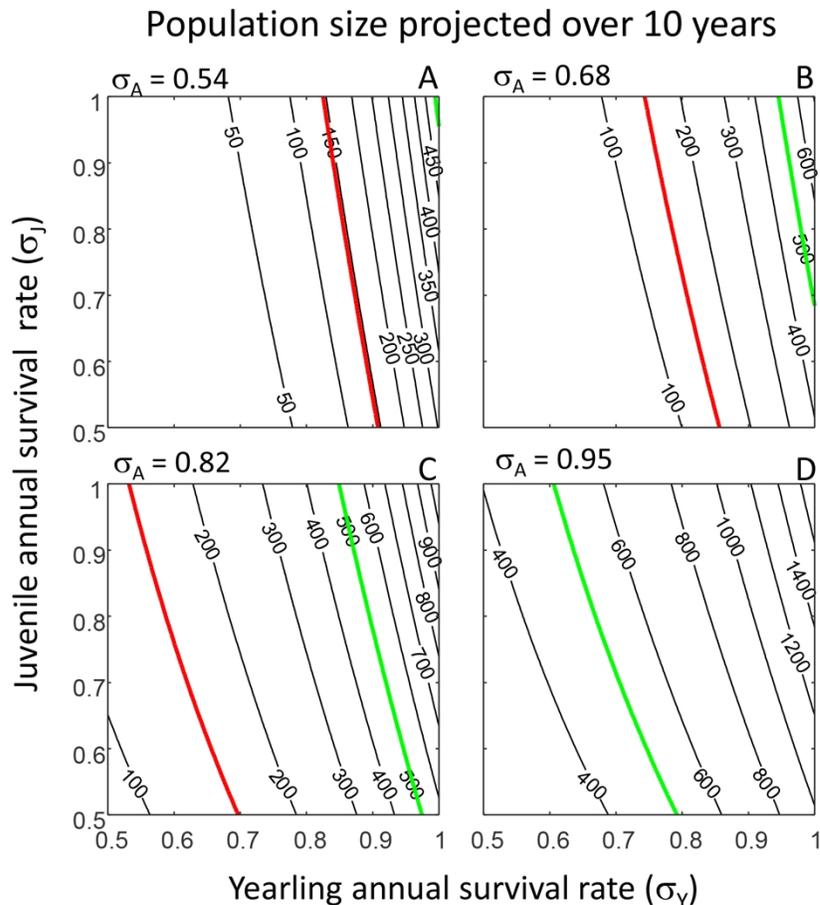
594

595 **Figure 3 Mean lifetime reproductive success.** Predicted lifetime reproductive success (R_0) in
 596 relation to yearling annual survival rate (σ_Y) and juvenile annual survival rate (σ_J) shown for
 597 each of four values of adult annual survival rate (σ_A): $\sigma_A = 0.54$ (80% of observed rate) (A); σ_A
 598 = 0.68 (observed rate) (B); $\sigma_A = 0.82$ (120% of observed rate) (C); and $\sigma_A = 0.95$ (140% of
 599 observed rate) (D). In each panel, isoclines denote equal values of lifetime reproductive success,
 600 R_0 . The blue line in each panel denotes population stability at $R_0 = 1$; values higher than $R_0 = 1$
 601 denote increasing populations and value lower than $R_0 = 1$ denote declining populations.



602

603 **Figure 4 Cohort generation time.** Predicted cohort generation time (T_c) in relation to yearling
 604 annual survival rate (σ_Y) and juvenile annual survival rate (σ_J) shown for each of four values of
 605 adult annual survival rate (σ_A): $\sigma_A = 0.54$ (80% of observed rate) (A); $\sigma_A = 0.68$ (observed rate)
 606 (B); $\sigma_A = 0.82$ (120% of observed rate) (C); and $\sigma_A = 0.95$ (140% of observed rate) (D). In each
 607 panel, isoclines denote equal values of cohort generation time.



608

609 **Figure 5 Population size projected over ten years.** A population of 500 individuals is
 610 projected over ten years using the predicted population growth rate λ (Fig. 2). Projected
 611 population sizes are shown in relation to yearling annual survival rate (σ_Y) and juvenile annual
 612 survival rate (σ_J) for each of four values of adult annual survival rate (σ_A): $\sigma_A = 0.54$ (80% of
 613 observed rate) (A); $\sigma_A = 0.68$ (observed rate) (B); $\sigma_A = 0.82$ (120% of observed rate) (C); and σ_A
 614 = 0.95 (140% of observed rate) (D). In each panel, isoclines denote equal values of projected
 615 population size. The green line in each panel denotes population stability where the projected
 616 population size is equal to the initial size of 500 individuals; above and below this line,
 617 populations are projected to increase or decrease respectively. The red line in each panel depicts

618 a population size of 149 individuals, which is equal to the lowest observed population size of
619 reef manta rays off the coast of Mozambique (Marshall, Dudgeon, & Bennett, 2011b).