

Prioritising non-native fish species for management actions in three Polish rivers using the newly developed tool - Dispersal-Origin-Status-Impact scheme

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Background. Biological invasions are a major threat to global biodiversity, with freshwater ecosystems being among the most susceptible to the successful establishment of non-native species and their respective potential impacts. In Poland, the introduction and spreading of non-native fish has led to biodiversity loss and ecosystem homogenisation.

Methods. Our study applies the Dispersal-Origin-Status-Impact (DOSI) assessment scheme, which is a population-level specific assessment that integrates multiple factors, including dispersal mechanisms, origin, status, and impacts, providing a nuanced framework for assessing invasion risks at local and regional levels. We used this tool to evaluate the risks associated with non-native fish species across three major Polish rivers (Pilica, Bzura, and Skrwa Prawa) and to prioritise them for management actions.

Results. Using DOSI, we assessed eight non-native species identified in the three studied rivers: seven in both Pilica and Bzura and four in Skrwa Prawa. The DOSI assessment scheme identified high variability in the ecological impacts and management priorities among the identified non-native species. Notably, species such as the Ponto-Caspian gobies exhibited higher risk levels due to their rapid spread and considerable ecological effects, contrasting with other species that demonstrated lower impact levels and, hence, received a lower priority for intervention.

Conclusion. The adoption of the DOSI scheme in three major rivers in Poland has provided valuable insights into the complexities of managing biological invasions, suggesting that localised, detailed assessments are crucial for effective conservation strategies and highlighting the importance of managing non-native populations locally.

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30 **Abstract**

31 **Background.** Biological invasions are a major threat to global biodiversity, with freshwater
32 ecosystems being among the most susceptible to the successful establishment of non-native species
33 and their respective potential impacts. In Poland, the introduction and spreading of non-native fish
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44 native species. Notably, species such as the Ponto-Caspian gobies exhibited higher risk levels due
45 to their rapid spread and considerable ecological effects, contrasting with other species that
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47 **Conclusion.** The adoption of the DOSI scheme in three major rivers in Poland has provided
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49 detailed assessments are crucial for effective conservation strategies and highlighting the
50 importance of managing non-native populations locally.

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54 Introduction

55 Non-native species actively or passively translocated by human actions in regions they have no
56 evolutionary history with (Soto et al., 2024), are recognised among the major threats to global
57 biodiversity, affecting all aspects of ecosystems (Simberloff et al., 2013; Cepic, Bechtold &
58 Wilfing, 2022). These impacts are modulated and often magnified by synergistic interactions with
59 other drivers such as habitat loss, which is considered ‘immense, insidious and usually irreversible’
60 (Strayer, 2010; Caffrey et al., 2014). Freshwater ecosystems are, among all ecosystems, the most
61 vulnerable to being affected by external drivers such as climate change, pollution, and biological
62 invasions (Havel et al., 2015; Haubrock et al., 2021; Cuthbert et al., 2023). Moreover, in the last
63 three decades, biodiversity declined faster in freshwater ecosystems than in marine and terrestrial
64 ecosystems (Collier et al., 2016; Reid et al., 2019, but see van Klink et al., 2020), with non-native
65 species introductions being among the main extinction drivers (Blackburn et al., 2014). The
66 intrinsic connectivity of freshwater ecosystems due to e.g. the canalization of large rivers,
67 facilitates the spread of non-native species and ultimately increases the homogenisation of
68 ecosystems (Marr et al., 2013). Consequently, mitigation of the effects of non-native species has
69 become one of the most pressing problems ecologists, decision makers, and stakeholders face
70 (Simberloff, 2015).

71 Considering the growing distribution of countless non-native species and the increasing
72 evidence of their staggering negative effects on recipient ecosystems (PBES, 2023) that are
73 increasingly difficult to monitor and manage (Moon, Blackman & Brewer, 2015; Crowley,
74 Hinchliffe & McDonald, 2017), there is a rising need for reliable, accessible, and robust tools to
75 assess the potential threat different populations of these non-native species present. Within the last
76 two decades, several protocols (including both ‘risk assessment protocols (Hawkins et al., 2015)
77 and ‘risk screening (a.k.a. ‘risk identification) (Vilizzi et al., 2022); please also see Srèbalienè et
78 al. (2019) and Hill et al. (2020) for a comparison of impact and risk assessment methods) have
79 been developed and implemented worldwide, targeting various taxonomic groups and evaluating
80 current and potential impacts of non-native species. Most of the available assessment protocols
81 share a common feature: they enable the classification of non-native species based on the level of
82 risk they do or may present to a specific assessment area. They, however, differ in complexity (e.g.
83 number of assessed aspects of the species), the underlying scoring system, and the range of impacts
84 assessed. However, although they are designed and tested by scientific experts, a recently

85 conducted – yet criticized – comprehensive consistency analysis revealed considerable
86 inconsistency among taxonomic groups, scoring systems, expertise of assessors, and impact
87 evaluated (environmental only or with socio-economic; González-Moreno et al., 2019). One
88 pressing issue is that most of these protocols are employed at the national (Tarkan et al., 2017) or
89 continental scale (Haubrock et al., 2021; Vilizzi et al., 2021), which is valuable for national
90 information systems or larger political entities like the European Union but lacks granularity
91 considering the variability of non-native species populations (Haubrock et al., 2024). These
92 generalised approaches can lead to underestimating or overestimating impacts at particular sites
93 by assuming that local effects can be generalised at the species level and be superimposed across
94 regions and ecosystems with similar conditions. *Vice versa*, an assessment at the national scale,
95 even when informed by local risk screenings, may still underestimate the threat a non-native
96 species presents at specific sites, as generalizations across regions and ecosystems with similar
97 conditions may overlook critical local variations. Another important issue is that several of the
98 currently available protocols consider only environmental impacts (González-Moreno et al., 2019)
99 as socio-economic impacts are usually difficult to quantify due to the lack of information, despite
100 it being widely accepted that the economic consequences of biological invasions prerequisite an
101 efficient allocation of financial resources e.g. management actions (Lodge et al., 2016; Bang et al.,
102 2022; Soto et al., 2023; Tarkan et al., 2024). This further underlines the urgency to easily
103 differentiate and prioritise non-native species for management interventions, resulting in more
104 efficient actions (Lodge et al., 2016).

105 In Poland, over 60% (17 out of 28) of non-native freshwater fish species were introduced
106 more than three decades ago and now form self-sustained populations in the wild or do not breed
107 in natural condition but are kept in aquaculture and used for stocking several water bodies
108 (Grabowska, Kotusz & Witkowski, 2010). One of the most important pathways aiding the range
109 extension of non-native aquatic species in inland waters of Poland is the European central invasion
110 corridor (Jażdżewski, 1980; Bij de Vaate et al., 2002; Fig. 1). This route was used by several non-
111 native fish species to spread in Polish inland waters (Grabowska, Pietraszewski & Ondračková,
112 2008; Semenchenko et al., 2011). In response to changing temporal invasion dynamics of non-
113 native species in Polish freshwater ecosystems, alongside recent European Union regulations, the
114 national project run by the government institution *The General Directorate for Environmental*
115 *Protection* was completed in 2018. It aimed to determine the degree of invasiveness of non-native

116 species in Poland and identify species that pose the greatest threat to invaded ecosystems. To
117 achieve that goal the *Harmonia*⁺ protocol was implemented in Poland and named *Harmonia*^{+PL}
118 (Tokarska-Guzik et al., 2019). Among the non-native fish species considered in this recent national
119 project (<https://www.gov.pl/web/gdos/inwazyjne-gatunki-obce-ias>), species recently established
120 in Poland include four species of Ponto-Caspian gobies (round, monkey, western tubenose and
121 racer goby; *Neogobius melanostomus*, *N. fluviatilis*, *Proterorhinus semilunaris* and *Babka*
122 *gymnotrachelus* respectively), the Chinese sleeper *Perccottus glenii*, and the topmouth gudgeon
123 *Pseudorasbora parva*, but also one species present in European inland waters (including Poland)
124 since the end of the 18th Century, namely the brown bullhead *Ameiurus nebulosus*. The last species
125 included was pirapitinga *Piaractus brachypomus* that is very occasionally recorded as single
126 individuals released by aquarists (Grabowska, Kotusz & Witkowski, 2010).

127 All those species analysed via *Harmonia*^{+PL} protocol, despite their wide distribution across
128 the country (except pirapitinga), were categorised as a low priority in the case of gobies and as a
129 medium priority in the case of Chinese sleeper and topmouth gudgeon. This resulted in the removal
130 of all four goby species from the list of harmful non-native species considered a national Polish
131 concern following the implementation of EU regulations (1143/2014). Furthermore, these changes
132 translate directly into the management of gobies: Although it is still forbidden to introduce them
133 or move them within the environment, it is now allowed to keep them (e.g., in the aquarium or
134 private pond), stock, sell, or exchange them. This can, in practice, result in e.g., intentional
135 introductions via anglers using gobies as live baits (Drake & Mandrak, 2014). Although there is
136 limited evidence of monkey, western tubenose, and racer goby negatively affecting ecosystems
137 they are introduced to (Grabowska et al., 2023), this is not the case for the round goby (Cerwenka
138 et al., 2023). Thus, the only fish species among the non-native species currently occurring in Polish
139 waters that remained on the lists of Union or Polish concern are the Chinese sleeper, the topmouth
140 gudgeon, the pumpkinseed (*Lepomis gibbosus*; which was not assessed by *Harmonia*^{+PL}), and the
141 brown bullhead (EU regulations 1143/2014 and its implementation at the national level in Dz. U.
142 2021 poz. 1718).

143 However, there is growing recognition that biological invasions are context-specific, with
144 considerable variations in the potential of individuals to spread and exert impacts among
145 populations influenced by diverse environmental and biological factors (Soto et al., 2024;
146 Haubrock et al., 2024). Consequently, there is a need for accurate and standardised assessment

147 protocols that consider the varied effects (both presence and impact) within populations of the
148 same species. The first steps have already been made by (Soto et al., 2024), who sorted out the
149 confusion in biological invasion nomenclature and proposed a new assessment scheme - The
150 Dispersal-Origin-Status-Impact (DOSI). The advantage of this approach stems from its thorough
151 yet adaptable framework, which can be applied to specific populations or at broader regional or
152 ecosystem scales in precise and scientific communication. Therefore, some populations might be
153 identified at different scales of prioritisation and can change over time due to the inherent temporal
154 dynamics of an invasion (e.g. population expanding or higher impacts). DOSI improves upon
155 previous management practices by assisting stakeholders and managers, who often face resource
156 constraints (Adelino et al., 2021) in selecting non-native species populations for management
157 actions, thereby enabling them to assess and prioritise non-native species.

158 To test the relevance and applicability of DOSI, we applied it to non-native species, in
159 three Polish rivers: Bzura, Pilica, and Skrwa Prawa, tributaries of the Vistula River, i.e., the Polish
160 section of the European central corridor of invasions aiming to assess different populations of non-
161 native fish species in rivers of different size. For this, monitoring studies were conducted at least
162 twice on each river, allowing to document several non-native species by examining the entire
163 length of the rivers (i.e. from their sources to their mouths), enabling to obtain an understanding
164 of ongoing changes in the distribution and abundance of these species. The DOSI scheme
165 implementation should provide insight into the threat of non-native species at the population level,
166 enable comparisons with results from the previously conducted Harmonia^{+PL} to identify potential
167 discrepancies and thereby direct future management efforts to particular localities. The DOSI
168 application may also reveal variability in the level of risk that different populations of the same
169 non-native species may pose in different water bodies, as the population level is usually overlooked
170 by more general metrics (e.g. Harmonia^{+PL}).

171

172 **Materials and Methods**

173 *Study sites and data collection*

174 Data for the current study consisted of results published in the national journal issued by the Polish
175 Angling Association (*Scientific Annual of the Polish Angling Association*; (Głowacki et al., 2024;
176 Jażdżewski et al., 2012; Penczak, 2006) and unpublished data from monitoring the Pilica, Bzura

177 and Skrwa Prawa Rivers (Fig. 1) performed by the Department of Ecology and Vertebrate
178 Zoology, University of Lodz, in 2013 and 2018. They are all tributaries of the Vistula River,
179 however, they differ in length and size (Pilica 332.5 km, 9258 km²; Bzura 166 km, 7788 km²;
180 Skrwa Prawa 117.6, 1704 km², length and catchment area, respectively). Each of the analysed
181 rivers, the Pilica, Bzura, and Skrwa Prawa, were sampled using the same methodology. One-run
182 electrocatch per constant unit effort (CPUE) was conducted using certified equipment. The effort
183 unit was established following Becklemishev's rule (Backiel and Penczak, 1989; Penczak, 1967),
184 which asserts that the sampling site length is adequate if no new species are collected with further
185 sampling. Electrofishing was performed by two persons, each using an anode with a dip net from
186 the boat or by wading, depending on the river depth.

187 The Pilica River was sampled in 2003-2005 (Penczak et al., 2006) and again in 2014-2017
188 (Głowacki et al., 2024) at 64 sites along the river; results from previous decades of sampling are
189 also presented in Penczak et al. (2006). Data for the Bzura River were collected in 2013
190 (unpublished) and 2009-2011 (Penczak et al., 2012) from 15 and 17 sites, respectively. The Skrwa
191 Prawa was sampled in 2002-2003 and 2010-2011 (Jądzewski et al., 2012) at 18 sites.

192 *The Dispersal-Origin-Status-Impact (DOSI) assessment scheme*

193 The DOSI assessment scheme (Fig. 2) exclusively focuses on negative impacts, emphasizing that
194 these potential threats are significantly more important and distinct than any potential benefits
195 (Carneiro et al. 2024). However, DOSI's objective is to prioritise non-native populations for
196 management interventions by considering local risks only, without considering the feasibility or
197 availability of appropriate methods, or the species' potential to spread beyond their current
198 locations. The focus on the population level distinguishes DOSI from other assessment tools, like
199 the Harmonia^{+PL} protocol, that are commonly applied at varying regional scales (i.e. assessment
200 regions) without a strict focus on the population level. The Harmonia^{+PL} protocol looks at non-
201 native species at the national level and consists of 30 questions divided into the two main modules
202 “invasion process” and “impact” and a final score calculated based on combined results obtained
203 for both modules.

204 DOSI prioritisation is structured around a hierarchy of primary dispersal mechanisms,
205 distinguishing between non-native populations that can (a) spread independently and invade areas
206 beyond the introduction site, (b) rely mainly on human assistance and the presence of pathways

207 and vectors, (c) have the capability for both assisted and independent spread (i.e., evaluated for
208 both a and b), and (d) the populations' status, which defines the state of a population within the
209 target site and the local impact it exerts. This means, that populations that can spread independently
210 and with assistance, and those showing changes in abundance and range, are ranked higher than
211 those with only one type of dependency. This is because the former scenarios indicate a greater
212 and more harmful invasion potential. Similarly, populations with one static and one expanding
213 dependency are also ranked higher. Conversely, if a population is determined to have no known
214 local impact, it is lowered in the priority ranking and thus requires a different response (Fig. 3).

215 To test the DOSI assessment scheme for each river, we considered all non-native fish species
216 identified. We assessed each identified non-native fish species in the Pilica River (seven non-
217 native species), the Bzura River (seven non-native species), and the Skrwa Prawa River (four non-
218 native species; Table 1) using DOSI to provide an objective overview for the prioritisation of each
219 rivers' non-native species populations (Fig. 3). Information on changes in abundance growth or
220 range extension were not always precise based on the field samplings, thus we filled information
221 gaps based on our expert knowledge of the study sites and the respective non-native species
222 invasion histories. Consequently, we discussed the DOSI assessment outcomes for the assessed
223 species with the previous screening based on Harmonia^{PL} to identify discrepancies and ultimately
224 test if the population level considered by DOSI provides relevant variability.

225

226 **Results**

227 Within the three tested rivers (i.e. Pilica, Bzura and Skrwa Prawa rivers), eight non-native species
228 were identified, three of which (i.e. the monkey goby, racer goby, and western tubenose goby)
229 were of Ponto-Caspian origin, another three (i.e. the topmouth gudgeon, Chinese sleeper, and gibel
230 carp) originated from Eastern Asia, while one (brown bullhead) originated from North America,
231 and another one (common carp) from the Danube catchment (Tables S1-S3). All goby species as
232 well as Chinese sleeper and topmouth gudgeon in each evaluated river (Pilica, Bzura, Skrwa
233 Prawa) were classified as independently dispersing, whereas brown bullhead, gibel carp, and
234 common carp as spreading depending on human assistance.

235 The DOSI ranking was not consistent among species and rivers highlighting the context-
236 dependency of invasions (Fig. 4). The monkey goby was designated as Highest Priority in the

237 Pilica and Bzura Rivers (the species was absent in the Skrwa Prawa and could not be evaluated
238 there) due to its increasing range and abundance leading to competitive pressure on native species
239 (Błońska et al., 2016; Błońska et al. under review). The monkey goby ranking was constant across
240 the analysed sites (Highest). The second one was the western tubenose goby, which also received
241 the status High Priority in all rivers. Although the species is also continually extending its range
242 and abundance, no negative impact has been observed (yet). The third was the topmouth gudgeon,
243 which was ranked as Medium Priority based on static range and abundance in both Pilica and
244 Bzura.

245 Both the racer goby and the gibel carp were ranked as Highest Priority in the Skrwa Prawa
246 and Pilica River, respectively. In other sites, both species were ranked as High or Medium Priority.
247 These discrepancies result from the inconsistent dynamics of both species. Besides range extension
248 and abundance increase, they displayed a negative effect on native biota at one site while having
249 no influence at another (e.g. gibel carp in the Pilica vs. Bzura River). The Chinese sleeper was
250 scored with Medium Priority in both the Pilica and Bzura Rivers and High Priority in the Skrwa
251 Prawa, where its abundance was increasing rather than static. The only species designated with
252 Low Priority was the brown bullhead, whose decreasing range and abundance are probably due to
253 the less suitable riverine habitat for this species compared to more stagnant waters such as oxbow
254 lakes.

255 The DOSI ranking differentiated among populations (ranging from High to Low Priority)
256 and was not complementary with the Harmonia^{+PL} results, which differentiated the six previously
257 assessed species into moderately invasive and potentially invasive (Table 2).

258

259 Discussion

260 In the current study, we evaluated the risk posed by non-native species in three temperate lowland
261 rivers in Poland (the Pilica, Bzura and Skrwa Prawa River) by applying the DOSI scheme (as
262 assessment protocol) and comparing our results to the Harmonia^{+PL} screening protocol outcomes.
263 Although DOSI and Harmonia^{+PL} are not directly comparable for this reason, as DOSI focuses on
264 population-level prioritization while Harmonia^{+PL} identify species risk based on broader ecological
265 implications of species introduction, this distinction allows DOSI to provide more granular, site-

266 specific risk rankings that may vary across locations. This variability was reflected in our findings,
267 where species were not consistently assigned the same rank across the three rivers, demonstrating
268 how DOSI can capture localized differences in species impact, even within similar ecological
269 contexts. Across the three rivers, species were not always designated with the same rank. High and
270 Medium Priority ranks dominated (six times each) with four Highest Priority and only one Low
271 Priority, underlining the DOSI's ability to prioritise non-native species at the population level.

272 *Population-level assessment*

273 Among the non-native species surveyed using DOSI were fish strongly associated with riverine
274 habitats, specifically Ponto-Caspian gobies. The evaluated rivers are in proximity to the Central
275 invasion corridor in Europe, which serves as the main expansion route for these species in Poland
276 (Semenchenko et al., 2011). Both monkey and western tubenose gobies were designated as Highest
277 and High Priority, respectively, constantly occurring across considered rivers that resulted from
278 increasing range and abundance. They are among the fastest spreading non-native species in
279 Poland, with the monkey goby having extended its range by 340 km in the last five years (Bylak
280 & Kukuła, 2024) and the tubenose goby by 255 km in seven years (Grabowska et al., 2021). Once
281 established, they often become abundant and may pose a threat to native species due to competition
282 (Borcherding, Heubel & Storm, 2019; Błażejowski et al., 2022) and predation (Grabowska et al.,
283 2023), even though they do not display aggressive behaviour (Kessel et al., 2011; Błońska et al.,
284 2016; Błońska, Kobak & Grabowska, 2017). Although they do not affect native species directly,
285 their high abundance and similar resource requirements can threaten native species (Błońska et al.,
286 2016a; Błońska et al. under review). A distinct example is the racer goby, which was ranked
287 differently in each evaluated river, from Medium in Bzura, High in Pilica to Highest Priority in
288 Skrwa Prawa. Although this variability in DOSI rankings among these goby species can likely be
289 explained by differences in habitat requirements (Płachocki et al., 2020; Bylak & Kukuła, 2024),
290 it should be noted that racer goby is not as efficient in expanding its range as monkey and tubenose
291 gobies, but it can significantly affect recipient communities (Grabowska et al., 2023). Observations
292 under laboratory conditions for instance revealed that racer gobies aggressively outcompete native
293 species when resources are limited (Kakareko et al., 2013; Grabowska et al., 2016). This adverse
294 effect on native species was also observed in the field (Kakareko et al., 2016). Impact of racer
295 goby was not confirmed directly in the analysed rivers, however, its extending range and
296 abundance ranked it with higher priority in Pilica and Skrwa Prawa, which in the case of the latter

297 one was reflected by decrease in population of white-finned gudgeon (*Romanogobio albipinnatus*),
298 golden loach (*Sabanejewia baltica*) and European bullhead (*Cottus gobio*).

299 Another group of assessed non-native species consisted of species that naturally express a
300 preference for stagnant waters and often occur in various natural and artificial water bodies in the
301 vicinity of river valleys from where individuals or in relatively small groups may accidentally enter
302 a main course of a river. Some of them, like the Chinese sleeper, are locally very common and
303 even dominate in some oxbow lakes or other parts of flood plains (Koščo et al., 2003; Grabowska
304 et al., 2011; Reshetnikov, 2013; Rechulicz, Płaska & Nawrot, 2015) where water current is slower
305 or even blocked like in old side arms, bays or marinas etc., and occasionally are flushed to the
306 main river channel during high water episodes. It is claimed that the Chinese sleeper uses rivers
307 for fast long-distance dispersal during floods (Reshetnikov, 2013). It also occurs as an accidentally
308 introduced species in fish ponds and spreads with stocking material of commercial species
309 (Reshetnikov, 2013; Grabowska et al., 2020). We acknowledge that the frequency of this species’
310 reporting in rivers, but also that of numerous other non-native fish species, will increase in the
311 foreseeable future (Witkowski & Grabowska, 2012; Seebens et al., 2021). However, the opposite
312 may be the case for the brown bullhead that used to occupy similar types of waters as the Chinese
313 sleeper but its range and abundance have decreased in Poland since the 1980s, when its intentional
314 introductions by local angling associations were very common (Witkowski A, 1996; Grabowska,
315 Kotusz & Witkowski, 2010) but nowadays is treated as a “pest” to be removed (Harmonia^{+PL};
316 Grabowska et al., 2018). This ultimately underlines the importance of local assessment for non-
317 native species.

318 The assessed non-native species also include species that, in most cases, directly originated
319 from fish ponds and accidentally escaped to adjacent streams and rivers. One of them is the gibel
320 carp, a cosmopolitan, eurytopic species; currently being the most widespread non-native fish in
321 Poland’s inland waters (Witkowski A, 1996; Grabowska, Kotusz & Witkowski, 2010). In fish
322 ponds, it is often stocked with accompanying carp, and it is introduced into special types of
323 commercial fishery, i.e. “put-and-take” recreational angling ponds (Grabowska, Kotusz &
324 Witkowski, 2010). Another non-native species found in fish ponds, the topmouth gudgeon, spreads
325 unintentionally in its non-native range as a contamination of stocking material of other Asian
326 cyprinids, such as carp or silver carp (Witkowski A, 1996; Grabowska, Kotusz & Witkowski,
327 2010; Gozlan et al., 2010). Both gibel carp and topmouth gudgeon are often found in rivers in a

328 large abundance, particularly after cleaning and other maintenance practices in fish ponds
329 (Witkowski A, 2009; Takács et al., 2017). However, such a situation was not observed in the
330 studied rivers as only single or few individuals of these species were caught during the sampling.

331 Although there is some evidence that species like Chinese sleeper, gibel carp, and topmouth
332 gudgeon have impacts on native species, economy, and even culture in stagnant waters (Gozlan et
333 al., 2010; Tarkan et al., 2012; Kutsokon et al., 2021), their ephemeral presence in rivers do not
334 create a serious threat for riverine ecosystems. Thus, they were scored as low or medium priority
335 due to a lack of abundance growth and impacts. However, these species are currently expanding
336 their invasive ranges and must be treated with consciousness and their occurrence in rivers should
337 be monitored.

338 *DOSI and Harmonia^{+PL}*

339 The impact of non-native species can differ substantially across sites, generalising at larger
340 geographically or political scales complicated or even flawed (Haubrock et al., 2024). Here, we
341 found substantial differences in the scores non-native species obtained across the three studied
342 rivers, and, considering the number of Highest and High Priority species, DOSI even suggested
343 that the Pilica and Skrwą Prawa Rivers are under higher pressure than the Bzura River, where most
344 species was identified as of Medium Priority (5 out of 7). DOSI also identified noteworthy
345 differences to *Harmonia^{+PL}*, which is applied at the country level and previously assessed all non-
346 native species that were also assessed by DOSI in this study (except for the gibel and common
347 carp). Indeed, the highest discrepancies were among Ponto-Caspian gobies, which were assigned
348 a High or Highest Priority in most analysed rivers following DOSI, while in *Harmonia^{+PL}* they
349 were ranked as potentially invasive non-native species (Grabowska et al., 2018b; Kakareko et al.,
350 2018a, 2018b). It can be partly explained by the differences in scoring scheme applied in DOSI
351 and *Harmonia^{+PL}* assessment.

352 Thus, even that in the case of Ponto-Caspian gobies they got the highest score assessing
353 their invasion process (what indicated that at time of the assessment they were still in the expansion
354 phase with a high risk of further spread), their “impact” was scored as low or moderate and it
355 influenced the final risk assessment score. It resulted from the fact that the knowledge of the impact
356 of that species on biota and inanimate elements of the ecosystem was low or there were not
357 convincing studies proving such potential impact (reviewed in Grabowska et al., 2023). A

358 contrasting case was the brown bullhead, recorded only in one of three analysed rivers and
359 accordingly only ranked as Low Priority by DOSI, was assessed as a moderately invasive non-
360 native species in Harmonia^{+PL} (Grabowska et al., 2018c). Topmouth gudgeon and Chinese sleeper
361 received similar scores (moderate) in both protocols (Grabowska et al., 2018a; Kakareko et al.,
362 2018c). Those three species got much higher scores in the “impact” module of Harmonia protocol
363 which increased the results of their risk screening.

364 *Management following DOSI*

365 The findings from the DOSI scheme highlight the importance of distinguishing between non-
366 native species that spread independently and those that spread through human assistance. This
367 differentiation is crucial for developing effective management strategies tailored to the specific
368 mechanisms of spread for each species. In the evaluated rivers, five species have been identified
369 to spread independently, whereas three species have been spreading primarily through human
370 assistance.

371 For species that spread independently, such as Ponto-Caspian gobies and Chinese sleeper
372 in River Pilica, Bzura and Skrwa Prawa, population management is essential. Effective strategies
373 should focus on the decimation of the population by implementing targeted removal programs to
374 reduce the population size, limiting propagule and colonization pressure through measures such as
375 ecosystem restorations to make the environment less conducive for these species to reproduce and
376 spread (Dorenbosch et al., 2017), and lowering exerted impacts by ongoing monitoring and
377 intervention to mitigate the negative impacts on native species and ecosystems. Current efforts in
378 some regions, such as existing management actions, have already shown success in lowering the
379 abundances of these species (e.g. Dorenbosch et al., 2017). Continued and enhanced efforts are
380 necessary to ensure long-term control and protection of native biodiversity (Leuven et al., 2017).

381 For species spreading through human assistance, such as gibel carp in River Pilica,
382 managing the pathways of introduction is critical. Relevant pathways include monitoring and
383 regulating the transport and release of fish stock to prevent contamination with non-native species,
384 educating and regulating activities such as fishing and boating to reduce unintentional
385 introductions, and ensuring that water management practices, such as the maintenance of fish
386 ponds and river channels, do not inadvertently facilitate the spread of non-native species. Effective
387 management of these pathways is possible through stringent regulation, public education, and

388 collaboration between stakeholders, including local communities, conservation organizations, and
389 government agencies.

390 Based on the DOSI assessment, it is recommended to enhance monitoring and research, as
391 continuous monitoring and research are essential to track the spread and impact of non-native
392 species. Implementing targeted management plans for high-priority species in each river is also
393 crucial. Increasing public awareness and involvement through education and engagement in
394 monitoring and control activities is necessary, as well as strengthening regulations and
395 enforcement to prevent the introduction and spread of non-native species through human activities.
396 By addressing both independent and assisted spread, we can develop a comprehensive approach
397 to managing non-native species and protecting the integrity of river ecosystems in Poland.

398

399 **Conclusion**

400 The application of the DOSI scheme in evaluating the risk posed by non-native species in three
401 temperate lowland rivers in Poland demonstrates that ranking non-native species is both feasible
402 and effective. The study highlights substantial differences between DOSI's population-level
403 assessments, and the species-level risk screening provided by Harmonia^{+PL}. These differences
404 underscore the importance of localized and population-specific evaluations in understanding and
405 managing non-native species. DOSI's ability to assess the risk at the population level provides
406 nuanced insights that are critical for effective management. By identifying the specific threats and
407 prioritising non-native species based on their local impact and spread, DOSI enables more targeted
408 and relevant management decisions. This approach helps in determining the most appropriate
409 management strategies, whether it involves population management for independently spreading
410 species or pathway management for those spreading through human assistance. Applying DOSI in
411 combination with monitoring surveys could enhance the accuracy and timeliness of risk
412 assessments, allowing for more proactive intervention strategies. Expanding the application of
413 DOSI to other geographical regions and aquatic environments may reveal further insights into its
414 effectiveness in addressing varying ecological contexts.

415

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422

423

424 **References**

- 425 Adelino JRP, Heringer G, Diagne C, Courchamp F, Faria LDB, Zenni RD. 2021. The economic
426 costs of biological invasions in Brazil: a first assessment. *NeoBiota* 67:349–374. DOI:
427 10.3897/neobiota.67.59185.
- 428 Backiel T., Penczak T. 1989. The fish and fisheries in the Vistula River and its tributary, the
429 Pilica River. In: Dodge D. P. ed. *Proceedings of the International Large River Symposium* .
430 Canadian Special Publication of Fisheries and Aquatic Sciences 106, 488–503.
- 431 Bang A, Cuthbert RN, Haubrock PJ, Fernandez RD, Moodley D, Diagne C, Turbelin AJ, Renault
432 D, Dalu T, Courchamp F. 2022. Massive economic costs of biological invasions despite
433 widespread knowledge gaps: a dual setback for India. *Biological Invasions* 24:2017–2039.
434 DOI: 10.1007/s10530-022-02780-z.
- 435 Bij de Vaate A, Jażdżewski K, Ketelaars HAM, Gollasch S, Van der Velde G. 2002.
436 Geographical patterns in range extension of Ponto-Caspian macroinvertebrate species in
437 Europe. *Canadian Journal of Fisheries and Aquatic Sciences* 59:1159–1174. DOI:
438 10.1139/f02-098.
- 439 Blackburn TM, Essl F, Evans T, Hulme PE, Jeschke JM, Kühn I, Kumschick S, Marková Z,
440 Mrugała A, Nentwig W, Pergl J, Pyšek P, Rabitsch W, Ricciardi A, Richardson DM,
441 Sendek A, Vilà M, Wilson JRU, Winter M, Genovesi P, Bacher S. 2014. A Unified
442 Classification of Alien Species Based on the Magnitude of their Environmental Impacts.
443 *PLoS Biology* 12:e1001850. DOI: 10.1371/journal.pbio.1001850.
- 444 Błażejowski M, Król J, Kakareko T, Mierzejewska K, Hliwa P. 2022. Daily and seasonal
445 dynamics of littoral zone fish communities in the lowland Włocławek Reservoir (central
446 Poland), with a special emphasis on alien invasive gobies. *Journal of Limnology* 81. DOI:
447 10.4081/JLIMNOL.2022.2059.
- 448 Błońska D, Kobak J, Grabowska J. 2017. Shelter competition between the invasive western
449 tubenose goby and the native stone loach is mediated by sex. *Journal of Limnology* 76.
450 DOI: 10.4081/jlimnol.2016.1557.
- 451 Błońska D, Kobak J, Kakareko T, Grabowska J. 2016. Can the presence of alien Ponto–Caspian
452 gobies affect shelter use by the native European bullhead? *Aquatic Ecology* 50:653–665.
- 453 Borcharding J, Heubel K, Storm S. 2019. Competition fluctuates across years and seasons in a 6-
454 species-fish community: empirical evidence from the field. *Reviews in Fish Biology and*
455 *Fisheries* 29:589–604. DOI: 10.1007/s11160-019-09567-x.

- 456 Bylak A, Kukuła K. 2024. Remarkable monkey goby expansion through a new invasion route
457 into submountain rivers. *Biological Invasions* 26:1303–1312. DOI: 10.1007/s10530-024-
458 03261-1.
- 459 Caffrey J, Baars J-R, Barbour J, Boets P, Boon P, Davenport K, Dick J, Early J, Edsman L,
460 Gallagher C, Gross J, Heinimaa P, Horrill C, Hudin S, Hulme P, Hynes S, MacIsaac H,
461 McLoone P, Millane M, Moen T, Moore N, Newman J, O’Conchuir R, O’Farrell M,
462 O’Flynn C, Oidtmann B, Renals T, Ricciardi A, Roy H, Shaw R, Weyl O, Williams F, Lucy
463 F. 2014. Tackling Invasive Alien Species in Europe: the top 20 issues. *Management of*
464 *Biological Invasions* 5:1–20. DOI: 10.3391/mbi.2014.5.1.01.
- 465 Cepic M, Bechtold U, Wilfing H. 2022. Modelling human influences on biodiversity at a global
466 scale—A human ecology perspective. *Ecological Modelling* 465:109854. DOI:
467 10.1016/j.ecolmodel.2021.109854.
- 468 Cerwenka AF, Brandner J, Dashinov D, Geist J. 2023. Small but Mighty: The Round Goby
469 (*Neogobius melanostomus*) as a Model Species of Biological Invasions. *Diversity* 15:528.
470 DOI: 10.3390/d15040528.
- 471 Collier KJ, Probert PK, Jeffries M. 2016. Conservation of aquatic invertebrates: concerns,
472 challenges and conundrums. *Aquatic Conservation: Marine and Freshwater Ecosystems*
473 26:817–837. DOI: 10.1002/aqc.2710.
- 474 Crowley SL, Hinchliffe S, McDonald RA. 2017. Conflict in invasive species management.
475 *Frontiers in Ecology and the Environment* 15:133–141. DOI: 10.1002/fee.1471.
- 476 Cuthbert RN, Darriet F, Chabrerie O, Lenoir J, Courchamp F, Claeys C, Robert V, Jourdain F,
477 Ulmer R, Diagne C, Ayala D, Simard F, Morand S, Renault D. 2023. Invasive
478 hematophagous arthropods and associated diseases in a changing world. *Parasites &*
479 *Vectors* 16:291. DOI: 10.1186/s13071-023-05887-x.
- 480 Dorenbosch M, van Kessel N, Liefveld W, Schoor M, van der Velde G, Leuven R. 2017.
481 Application of large wood in regulated riverine habitats facilitates native fishes but not
482 invasive alien round goby (*Neogobius melanostomus*). *Aquatic Invasions* 12:405–413. DOI:
483 10.3391/ai.2017.12.3.13.
- 484 Drake DAR, Mandrak NE. 2014. Ecological Risk of Live Bait Fisheries: A New Angle on
485 Selective Fishing. *Fisheries* 39:201–211. DOI: 10.1080/03632415.2014.903835.
- 486 Głowacki Ł., Zięba G., Pietraszewski D., Tsydel M., Tybulczuk S., Błońska D., Kruk A.,
487 Pyrzanowski K., Leśniak M., Janic B., Penczak T. 2024. Fish fauna in the Pilica river
488 catchment in the sixth decade of study part I. Pilica River . *Scientific Annual of the Polish*
489 *Angling Association*.
- 490 González-Moreno P, Lazzaro L, Vilà M, Preda C, Adriaens T, Bacher S, Brundu G, Copp GH,
491 Essl F, García-Berthou E, Katsanevakis S, Moen TL, Lucy FE, Nentwig W, Roy HE,
492 Srébalienė G, Talgø V, Vanderhoeven S, Andjelković A, Arbačiauskas K, Auger-
493 Rozenberg M-A, Bae M-J, Bariche M, Boets P, Boieiro M, Borges PA, Canning-Clode J,
494 Cardigos F, Chartosia N, Cottier-Cook EJ, Crocetta F, D’hondt B, Foggi B, Follak S,
495 Gallardo B, Gammelmo Ø, Giakoumi S, Giuliani C, Guillaume F, Jelaska LŠ, Jeschke JM,
496 Jover M, Juárez-Escario A, Kalogirou S, Kočić A, Kytinou E, Laverty C, Lozano V,
497 Maceda-Veiga A, Marchante E, Marchante H, Martinou AF, Meyer S, Minchin D,
498 Montero-Castaño A, Morais MC, Morales-Rodriguez C, Muhthassim N, Nagy ZÁ, Ogris N,
499 Onen H, Pergl J, Puntila R, Rabitsch W, Ramburn TT, Rego C, Reichenbach F, Romeralo
500 C, Saul W-C, Schrader G, Sheehan R, Simonović P, Skolka M, Soares AO, Sundheim L,
501 Tarkan AS, Tomov R, Tricarico E, Tsiamis K, Uludağ A, van Valkenburg J, Verreycken H,

- 502 Vettraino AM, Vilar L, Wiig Ø, Witzell J, Zanetta A, Kenis M. 2019. Consistency of impact
503 assessment protocols for non-native species. *NeoBiota* 44:1–25. DOI:
504 10.3897/neobiota.44.31650.
- 505 Gozlan RE, Andreou D, Asaeda T, Beyer K, Bouhadad R, Burnard D, Caiola N, Cakic P,
506 Djikanovic V, Esmaeili HR, Falka I, Golicher D, Harka A, Jeney G, Kováč V, Musil J,
507 Nocita A, Povz M, Poulet N, Virbickas T, Wolter C, Serhan Tarkan A, Tricarico E,
508 Trichkova T, Verreycken H, Witkowski A, Guang Zhang C, Zweimueller I, Robert Britton
509 J. 2010. Pan-continental invasion of *Pseudorasbora parva* : towards a better understanding
510 of freshwater fish invasions. *Fish and Fisheries* 11:315–340. DOI: 10.1111/j.1467-
511 2979.2010.00361.x.
- 512 Grabowska J, Błońska D, Ondračková M, Kakareko T. 2023. The functional ecology of four
513 invasive Ponto–Caspian gobies. *Reviews in Fish Biology and Fisheries* 33:1329–1352. DOI:
514 10.1007/s11160-023-09801-7.
- 515 Grabowska J., Kakareko T., Mazurska K. 2018a. *Perccottus glenii* Dybowski, 1877. In:
516 *Harmonia+PL—procedure for negative impact risk assessment for invasive alien species*
517 *and potentially invasive alien species in Poland*. .
518 [http://projekty.gdos.gov.pl/files/artykuly/126888/Perccottus-](http://projekty.gdos.gov.pl/files/artykuly/126888/Perccottus-glenii_trawianka_EN_icon.pdf)
519 [glenii_trawianka_EN_icon.pdf](http://projekty.gdos.gov.pl/files/artykuly/126888/Perccottus-glenii_trawianka_EN_icon.pdf).
- 520 Grabowska J., Kakareko T., Mazurska K. 2018b. *Proterorhinus semilunaris* (Pallas, 1811). In:
521 *Harmonia+PL—procedure for negative impact risk assessment for invasive alien species*
522 *and potentially invasive alien species in Poland*.
523 [http://projekty.gdos.gov.pl/files/artykuly/126893/Proterorhinus-marmoratus_babka-](http://projekty.gdos.gov.pl/files/artykuly/126893/Proterorhinus-marmoratus_babka-marmurkowata_EN_icon.pdf)
524 [marmurkowata_EN_icon.pdf](http://projekty.gdos.gov.pl/files/artykuly/126893/Proterorhinus-marmoratus_babka-marmurkowata_EN_icon.pdf).
- 525 Grabowska J., Kakareko T., Mazurska K. 2018c. *Ameiurus nebulosus* (Le Sueur, 1819). In:
526 *Harmonia+PL—procedure for negative impact risk assessment for invasive alien species*
527 *and potentially invasive alien species in Poland*.
528 [http://projekty.gdos.gov.pl/files/artykuly/126845/Ameiurus-nebulosus_sumik-](http://projekty.gdos.gov.pl/files/artykuly/126845/Ameiurus-nebulosus_sumik-karlowaty_EN_icon.pdf)
529 [karlowaty_EN_icon.pdf](http://projekty.gdos.gov.pl/files/artykuly/126845/Ameiurus-nebulosus_sumik-karlowaty_EN_icon.pdf).
- 530 Grabowska J, Kakareko T, Błońska D, Przybylski M, Kobak J, Jermacz, Copp GH. 2016.
531 Interspecific competition for a shelter between non-native racer goby and native European
532 bullhead under experimental conditions - Effects of season, fish size and light conditions.
533 *Limnologia* 56:30–38. DOI: 10.1016/j.limno.2015.11.004.
- 534 Grabowska J, Kotusz J, Witkowski A. 2010. Alien invasive fish species in Polish waters: an
535 overview. *Folia Zoologica* 59:73–85. DOI: 10.25225/fozo.v59.i1.a1.2010.
- 536 Grabowska J, Kvach Y, Rewicz T, Pupins M, Kutsokon I, Dykyy I, Antal L, Zięba G, Rakauskas
537 V, Trichkova T, Čeirāns A, Grabowski M. 2020. First insights into the molecular population
538 structure and origins of the invasive Chinese sleeper, *Perccottus glenii*, in Europe. *NeoBiota*
539 57:87–107. DOI: 10.3897/neobiota.57.48958.
- 540 Grabowska J, Pietraszewski D, Ondračková M. 2008. Tubenose goby *Proterorhinus marmoratus*
541 (Pallas, 1814) has joined three other Ponto-Caspian gobies in the Vistula River (Poland).
542 *Aquatic Invasions* 3:261–265. DOI: 10.3391/ai.2008.3.2.20.
- 543 Grabowska J, Pietraszewski D, Przybylski M, Tarkan AS, Marszał L, Lampart-Kałużniacka M.
544 2011. Life-history traits of Amur sleeper, *Perccottus glenii*, in the invaded Vistula River:
545 early investment in reproduction but reduced growth rate. *Hydrobiologia* 661:197–210.
546 DOI: 10.1007/s10750-010-0524-0.

- 547 Grabowska J, Tarkan AS, Błońska D, Karakuş NT, Janic B, Przybylski M. 2021. Prolific
548 pioneers and reserved settlers. Changes in the life-history of the western tubenose goby
549 (*Proterorhinus semilunaris*) at different invasion stages. *Science of the Total Environment*
550 750:142316.
- 551 Haubrock PJ, Copp GH, Johović I, Balzani P, Inghilesi AF, Nocita A, Tricarico E. 2021. North
552 American channel catfish, *Ictalurus punctatus*: a neglected but potentially invasive
553 freshwater fish species? *Biological Invasions* 23:1563–1576. DOI: 10.1007/s10530-021-
554 02459-x.
- 555 Haubrock PJ, Soto I, Ahmed DA, Ansari AR, Tarkan AS, Kurtul I, Macêdo RL, Lázaro-Lobo A,
556 Toutain M, Parker B, Błońska D, Guareschi S, Cano-Barbacil C, Dominguez Almela V,
557 Andreou D, Moyano J, Akalin S, Kaya C, Bayçelebi E, Yoğurtçuoğlu B, Briski E, Aksu S,
558 Emiroğlu Ö, Mammola S, De Santis V, Kourantidou M, Pincheira-Donoso D, Britton JR,
559 Kouba A, Dolan EJ, Kirichenko NI, García-Berthou E, Renault D, Fernandez RD, Yapıcı S,
560 Giannetto D, Nuñez MA, Hudgins EJ, Pergl J, Milardi M, Musolin DL, Cuthbert RN. 2024.
561 Biological invasions are a population-level rather than a species-level phenomenon. *Global*
562 *Change Biology* 30. DOI: 10.1111/gcb.17312.
- 563 Havel JE, Kovalenko KE, Thomaz SM, Amalfitano S, Kats LB. 2015. Aquatic invasive species:
564 challenges for the future. *Hydrobiologia* 750:147–170. DOI: 10.1007/s10750-014-2166-0.
- 565 Hawkins CL, Bacher S, Essl F, Hulme PE, Jeschke JM, Kühn I, Kumschick S, Nentwig W, Pergl
566 J, Pyšek P, Rabitsch W, Richardson DM, Vilà M, Wilson JRU, Genovesi P, Blackburn TM.
567 2015. Framework and guidelines for implementing the proposed IUCN Environmental
568 Impact Classification for Alien Taxa (EICAT). *Diversity and Distributions* 21:1360–1363.
569 DOI: 10.1111/ddi.12379.
- 570 Hill J, Copp G, Hardin S, Lawson K, Lawson L, Tuckett Q, Vilizzi L, Watson C. 2020.
571 Comparing apples to oranges and other misrepresentations of the risk screening tools FISK
572 and AS-ISK – a rebuttal of Marcot et al. (2019). *Management of Biological Invasions*
573 11:325–341. DOI: 10.3391/mbi.2020.11.2.10.
- 574 Jażdżewski K. 1980. Range extensions of some gammaridean species in European inland waters
575 caused by human activity. *Crustaceana. Supplement*:84–107.
- 576 Jażdżewski M., Błońska D., Marszał L., Przybylski M., Janic B., Pietraszewski D., Tybulczuk S.,
577 Zieliński P., Grabowska J., Zięba G. 2012. Monitoring of fish fauna in the Skrwa Prawa
578 River system: continuation in 2010 and 2011 . *Scientific Annual of the Polish Angling*
579 *Association* 25:5–29.
- 580 Kakareko T, Kobak J, Grabowska J, Jermacz Ł, Przybylski M, Poznańska M, Pietraszewski D,
581 Copp GH. 2013. Competitive interactions for food resources between invasive racer goby
582 *Babka gymnotrachelus* and native European bullhead *Cottus gobio*. *Biological Invasions*
583 15:2519–2530. DOI: 10.1007/s10530-013-0470-7.
- 584 Kakareko T, Kobak J, Poznańska M, Jermacz L, Copp GH. 2016. Underwater evaluation of
585 habitat partitioning in a European river between a non-native invader, the racer goby and a
586 threatened native fish, the European bullhead. *Ecology of Freshwater Fish* 25:60–71. DOI:
587 10.1111/eff.12191.
- 588 Kakareko T., Grabowska J., Mazurska K. 2018a. *Neogobius gymnotrachelus* (Pallas, 1811). In:
589 *Harmonia+PL—procedure for negative impact risk assessment for invasive alien species*
590 *and potentially invasive alien species in Poland*.
591 [http://projekty.gdos.gov.pl/files/artykuly/126875/Neogobius-gymnotrachelus_babka-](http://projekty.gdos.gov.pl/files/artykuly/126875/Neogobius-gymnotrachelus_babka-lysa_EN_icon.pdf)
592 [lysa_EN_icon.pdf](http://projekty.gdos.gov.pl/files/artykuly/126875/Neogobius-gymnotrachelus_babka-lysa_EN_icon.pdf).

- 593 Kakareko T., Grabowska J., Mazurska K. 2018b. *Neogobius fluviatilis* (Pallas, 1811). In:
594 *Harmonia+PL—procedure for negative impact risk assessment for invasive alien species*
595 *and potentially invasive alien species in Poland*.
596 [http://projekty.gdos.gov.pl/files/artykuly/126874/Neogobius-fluviatilis_babka-](http://projekty.gdos.gov.pl/files/artykuly/126874/Neogobius-fluviatilis_babka-szczupla_EN_icon.pdf)
597 [szczupla_EN_icon.pdf](http://projekty.gdos.gov.pl/files/artykuly/126874/Neogobius-fluviatilis_babka-szczupla_EN_icon.pdf).
- 598 Kakareko T., Grabowska J., Mazurska K. 2018c. *Pseudorasbora parva* (Schlegel, 1842). In:
599 *Harmonia+PL—procedure for negative impact risk assessment for invasive alien species*
600 *and potentially invasive alien species in Poland*.
601 [http://projekty.gdos.gov.pl/files/artykuly/126894/Pseudorasbora-parva_Czebaczek-](http://projekty.gdos.gov.pl/files/artykuly/126894/Pseudorasbora-parva_Czebaczek-amurski_EN_icon.pdf)
602 [amurski_EN_icon.pdf](http://projekty.gdos.gov.pl/files/artykuly/126894/Pseudorasbora-parva_Czebaczek-amurski_EN_icon.pdf).
- 603 Kessel N Van, Dorenbosch M, De Boer MRM, Leuven RSEW, Van Der Velde G. 2011.
604 *Competition for shelter between four invasive gobiids and two native benthic fish species*.
605 van Klink R, Bowler DE, Gongalsky KB, Swengel AB, Gentile A, Chase JM. 2020. Meta-
606 analysis reveals declines in terrestrial but increases in freshwater insect abundances. *Science*
607 368:417–420. DOI: 10.1126/science.aax9931.
- 608 Koščo J., Lusk S., Halačka K., Lusková V. 2003. The expansion and occurrence of the Amur
609 sleeper (*Percottus glenii*) in eastern Slovakia. *Folia Zoologica* 52:329–336.
- 610 Kutsokon I, Tkachenko M, Bondarenko O, Pupins M, Snigirova A, Berezovska V, Čeirāns A,
611 Kvach Y. 2021. The role of invasive Chinese sleeper *Percottus glenii* Dybowski, 1877 in
612 the Ilgas Nature Reserve ecosystem: an example of a monospecific fish community.
613 *BioInvasions Records* 10:396–410. DOI: 10.3391/bir.2021.10.2.18.
- 614 Leuven R, Boggero A, Bakker E, Elgin A, Verreycken H. 2017. Invasive species in inland
615 waters: from early detection to innovative management approaches. *Aquatic Invasions*
616 12:269–273. DOI: 10.3391/ai.2017.12.3.01.
- 617 Lodge DM, Simonin PW, Burgiel SW, Keller RP, Bossenbroek JM, Jerde CL, Kramer AM,
618 Rutherford ES, Barnes MA, Wittmann ME, Chadderton WL, Apriesnig JL, Beletsky D,
619 Cooke RM, Drake JM, Egan SP, Finnoff DC, Gantz CA, Grey EK, Hoff MH, Howeth JG,
620 Jensen RA, Larson ER, Mandrak NE, Mason DM, Martinez FA, Newcomb TJ,
621 Rothlisberger JD, Tucker AJ, Warziniack TW, Zhang H. 2016. Risk Analysis and
622 Bioeconomics of Invasive Species to Inform Policy and Management. *Annual Review of*
623 *Environment and Resources* 41:453–488. DOI: 10.1146/annurev-environ-110615-085532.
- 624 Marr SM, Olden JD, Leprieur F, Arismendi I, Čaleta M, Morgan DL, Nocita A, Šanda R, Serhan
625 Tarkan A, García-Berthou E. 2013. A global assessment of freshwater fish introductions in
626 mediterranean-climate regions. *Hydrobiologia* 719:317–329. DOI: 10.1007/s10750-013-
627 1486-9.
- 628 Moon K, Blackman DA, Brewer TD. 2015. Understanding and integrating knowledge to
629 improve invasive species management. *Biological Invasions* 17:2675–2689. DOI:
630 10.1007/s10530-015-0904-5.
- 631 Penczak T. 1967. The biological and technical principles of the fishing by use of direct-current
632 field. *Przegląd Zoologiczny* 11:114–131.
- 633 Penczak T., Kruk A., Marszał L., Galicka W., Tybulczuk S., Tszydel M. 2012. Fish fauna
634 regeneration of the Bzura and Ner Rivers after industrial sewage reduction. *Scientific*
635 *Annual of the Polish Angling Association* 25:85–93.
- 636 Penczak T., Kruk A., Zięba G., Marszał L., Koszaliński H., Tybulczuk S., Galicka W. 2006. Fish
637 fauna in the Pilica River system in the fifth decade of study part I. Pilica River. *Scientific*
638 *Annual of the Polish Angling Association* 19:103–122.

- 639 Płachocki D, Kobak J, Poznańska-Kakareko M, Kakareko T. 2020. Environmental factors
640 associated with the occurrence of the Ponto–Caspian gobies in a lowland river belonging to
641 the central European invasion corridor. *River Research and Applications* 36:25–35. DOI:
642 10.1002/rra.3543.
- 643 Rechulicz J, Płaska W, Nawrot D. 2015. Occurrence, dispersion and habitat preferences of Amur
644 sleeper (*Perccottus glenii*) in oxbow lakes of a large river and its tributary. *Aquatic Ecology*
645 49:389–399. DOI: 10.1007/s10452-015-9532-5.
- 646 Reid AJ, Carlson AK, Creed IF, Eliason EJ, Gell PA, Johnson PTJ, Kidd KA, MacCormack TJ,
647 Olden JD, Ormerod SJ. 2019. Emerging threats and persistent conservation challenges for
648 freshwater biodiversity. *Biological reviews* 94:849–873.
- 649 Reshetnikov A. 2013. Spatio-temporal dynamics of the expansion of rotan *Perccottus glenii* from
650 West-Ukrainian centre of distribution and consequences for European freshwater
651 ecosystems. *Aquatic Invasions* 8:193–206. DOI: 10.3391/ai.2013.8.2.07.
- 652 Roy HE, Pauchard A, Stoett P, Renard Truong T, Bacher S, Galil B, Hulme P, Ikeda T, Sankaran
653 KV, McGeoch M, Meyerson L, Nunez M, Ordonez A, Rahlao S, Schwindt E, Seebens H,
654 Sheppard A, and Vandvik V. 2023. *Summary for Policymakers of the Thematic Assessment
655 Report on Invasive Alien Species and their Control of the Intergovernmental Science-Policy
656 Platform on Biodiversity and Ecosystem Services*. Bonn, Germany.
- 657 Seebens H, Bacher S, Blackburn TM, Capinha C, Dawson W, Dullinger S, Genovesi P, Hulme
658 PE, van Kleunen M, Kühn I, Jeschke JM, Lenzner B, Liebhold AM, Pattison Z, Pergl J,
659 Pyšek P, Winter M, Essl F. 2021. Projecting the continental accumulation of alien species
660 through to 2050. *Global Change Biology* 27:970–982. DOI: 10.1111/gcb.15333.
- 661 Semenchenko V, Grabowska J, Grabowski M, Rizevsky V, Pluta M. 2011. Non-native fish in
662 Belarusian and Polish areas of the European central invasion corridor. *Oceanological and
663 Hydrobiological Studies* 40:57–67. DOI: 10.2478/s13545-011-0007-6.
- 664 Simberloff D. 2015. Non-native invasive species and novel ecosystems. *F1000Prime Reports* 7.
665 DOI: 10.12703/P7-47.
- 666 Simberloff D, Martin J-L, Genovesi P, Maris V, Wardle DA, Aronson J, Courchamp F, Galil B,
667 García-Berthou E, Pascal M, Pyšek P, Sousa R, Tabacchi E, Vilà M. 2013. Impacts of
668 biological invasions: what’s what and the way forward. *Trends in Ecology & Evolution*
669 28:58–66. DOI: 10.1016/j.tree.2012.07.013.
- 670 Soto I, Balzani P, Carneiro L, Cuthbert RN, Macêdo R, Serhan Tarkan A, Ahmed DA, Bang A,
671 Bacela-Spychalska K, Bailey SA, Baudry T, Ballesteros-Mejia L, Bortolus A, Briski E,
672 Britton JR, Buřič M, Camacho-Cervantes M, Cano-Barbacil C, Copilaş-Ciocianu D,
673 Coughlan NE, Courtois P, Csabai Z, Dalu T, De Santis V, Dickey JWW, Dimarco RD,
674 Falk-Andersson J, Fernandez RD, Florencio M, Franco ACS, García-Berthou E, Giannetto
675 D, Glavendekic MM, Grabowski M, Heringer G, Herrera I, Huang W, Kamelamela KL,
676 Kirichenko NI, Kouba A, Kourantidou M, Kurtul I, Laufer G, Lipták B, Liu C, López-
677 López E, Lozano V, Mammola S, Marchini A, Meshkova V, Milardi M, Musolin DL,
678 Nuñez MA, Oficialdegui FC, Patoka J, Pattison Z, Pincheira-Donoso D, Piria M., Probert
679 AF, Rasmussen JJ, Renault D, Ribeiro F, Rilov G, Robinson TB, Sanchez AE, Schwindt E,
680 South J, Stoett P, Verreycken H, Vilizzi L, Wang YJ, Watari Y, Wehi PM, Weiperth A,
681 Wiberg-Larsen P, Yapıcı S, Yoğurtçuoğlu B, Zenni RD, Galil BS, Dick JTA, Russell JC,
682 Ricciardi A, Simberloff D, Bradshaw CJA, Haubrock PJ. 2024a. Taming the terminological
683 tempest in invasion science. *Biological Reviews*.

- 684 Soto I, Haubrock PJ, Cuthbert RN, Renault D, Probert AF, Tarkan AS. 2023. Monetary impacts
685 should be considered in biological invasion risk assessments. *Journal of Applied Ecology*
686 60:2309–2313. DOI: 10.1111/1365-2664.14510.
- 687 Srébalienė G, Olenin S, Minchin D, Narščius A. 2019. A comparison of impact and risk
688 assessment methods based on the IMO Guidelines and EU invasive alien species risk
689 assessment frameworks. *PeerJ* 7:e6965. DOI: 10.7717/peerj.6965.
- 690 Strayer DL. 2010. Alien species in fresh waters: ecological effects, interactions with other
691 stressors, and prospects for the future. *Freshwater Biology* 55:152–174. DOI:
692 10.1111/j.1365-2427.2009.02380.x.
- 693 Takács P, Czeglédi I, Ferincz Á, Sály P, Specziár A, Vitál Z, Weiperth A, Erős T. 2017. Non-
694 native fish species in Hungarian waters: historical overview, potential sources and recent
695 trends in their distribution. *Hydrobiologia* 795:1–22. DOI: 10.1007/s10750-017-3147-x.
- 696 Tarkan AS, Bayçelebi E, Giannetto D, Özden ED, Yazlık A, Emiroğlu Ö, Aksu S, Uludağ A,
697 Aksoy N, Baytaşoğlu H, Kaya C, Mutlu T, Kırankaya ŞG, Ergüden D, Per E, Üremiş İ,
698 Candan O, Kekillioğlu A, Yoğurtçuoğlu B, Ekmekçi FG, Başak E, Özkan H, Kurtul I, Innal
699 D, Killi N, Yapıcı S, Ayaz D, Çiçek K, Mol O, Çınar E, Yeğen V, Angulo E, Cuthbert RN,
700 Soto I, Courchamp F, Haubrock PJ. 2024. Economic costs of non-native species in Türkiye:
701 A first national synthesis. *Journal of Environmental Management* 358:120779. DOI:
702 10.1016/j.jenvman.2024.120779.
- 703 Tarkan AS, Gaygusuz Ö, Gürsoy Gaygusuz Ç, Saç G, Copp GH. 2012. Circumstantial evidence
704 of gibel carp, *Carassius gibelio*, reproductive competition exerted on native fish species in a
705 mesotrophic reservoir. *Fisheries Management and Ecology* 19:167–177. DOI:
706 10.1111/j.1365-2400.2011.00839.x.
- 707 Tarkan AS, Vilizzi L, Top N, Ekmekçi FG, Stebbing PD, Copp GH. 2017. Identification of
708 potentially invasive freshwater fishes, including translocated species, in Turkey using the
709 Aquatic Species Invasiveness Screening Kit (AS-ISK). *International Review of*
710 *Hydrobiology* 102:47–56. DOI: 10.1002/iroh.201601877.
- 711 Tokarska-Guzik B., Urbisz A., Mazurska K., Solarz W., Bzdega K., Danielewicz W., Jackowiak
712 B., Sudnik-Wójcikowska B., Nowak T., Chmura D., Gąbka M, Wołkowycki D., Hołdyński
713 C, Celka Z, Szymura M, Gazda A, Adamowski W, Brzosko E, Chmiel J, Sachajdakiewicz I,
714 Kompała-Bąba A, Popiela A, Michalska-Hejduk D, Sotek Z, Zajac A, Zajac M, Kolada A,
715 Korniak T, Krzysztofiak A, Melon E, Myśliwy M, Rosadziński S, Sierka E, Szcześniak E,
716 Śliwiński M, Halerewicz A, Halladin-Dąbrowska A, Kopeć D, Otręba A, Purcel A,
717 Wiatrowska B, Woziwoda B, Wylazłowska J, Bąbelewski P, Bomanowska A, Kamiński R,
718 Krajewski Ł, Mirski P, Podlaska M, Sobisz Z, Szewczyk M, Woźniak G, Żołnierz L, Król
719 W, Krzysztofiak L, Najberek K, Pasierbiński A, Smieja A, Zarychta A. 2019. The
720 Harmonia+PL procedure for the negative impact risk assessment of invasive alien species
721 and potentially invasive alien species in Poland. In: Frey L. ed. *Streszczenia Referatów i*
722 *Plakatów 58. Zjazdu PTB*. Warszawa: Polskie Towarzystwo Botaniczne, 126–127.
- 723 Vilizzi L, Copp GH, Hill JE, Adamovich B, Aislabie L, Akin D, Al-Faisal AJ, Almeida D,
724 Azmai MNA, Bakiu R, Bellati A, Bernier R, Bies JM, Bilge G, Branco P, Bui TD, Canning-
725 Clode J, Cardoso Ramos HA, Castellanos-Galindo GA, Castro N, Chaichana R, Chainho P,
726 Chan J, Cunico AM, Curd A, Dangchana P, Dashinov D, Davison PI, de Camargo MP,
727 Dodd JA, Durland Donahou AL, Edsman L, Ekmekçi FG, Elphinstone-Davis J, Erős T,
728 Evangelista C, Fenwick G, Ferincz Á, Ferreira T, Feunteun E, Filiz H, Forneck SC,
729 Gajduchenko HS, Gama Monteiro J, Gestoso I, Giannetto D, Gilles AS, Gizzi F, Glamuzina

- 730 B, Glamuzina L, Goldsmit J, Gollasch S, Gouilletquer P, Grabowska J, Harmer R, Haubrock
731 PJ, He D, Hean JW, Herczeg G, Howland KL, İlhan A, Interesova E, Jakubčinová K,
732 Jelmert A, Johnsen SI, Kakareko T, Kanongdate K, Killi N, Kim J-E, Kırankaya ŞG,
733 Kňazovická D, Kopecký O, Kostov V, Koutsikos N, Kozic S, Kuljanishvili T, Kumar B,
734 Kumar L, Kurita Y, Kurtul I, Lazzaro L, Lee L, Lehtiniemi M, Leonardi G, Leuven RSEW,
735 Li S, Lipinskaya T, Liu F, Lloyd L, Lorenzoni M, Luna SA, Lyons TJ, Magellan K,
736 Malmstrøm M, Marchini A, Marr SM, Masson G, Masson L, McKenzie CH, Memedemin
737 D, Mendoza R, Minchin D, Miossec L, Moghaddas SD, Moshobane MC, Mumladze L,
738 Naddafi R, Najafi-Majd E, Năstase A, Năvodaru I, Neal JW, Nienhuis S, Nimtim M, Nolan
739 ET, Occhipinti-Ambrogi A, Ojaveer H, Olenin S, Olsson K, Onikura N, O'Shaughnessy K,
740 Paganelli D, Parretti P, Patoka J, Pavia RTB, Pellitteri-Rosa D, Pelletier-Rousseau M,
741 Peralta EM, Perdikaris C, Pietraszewski D, Piria M, Pitois S, Pompei L, Poulet N, Preda C,
742 Puntila-Dodd R, Qashqaei AT, Radočaj T, Rahmani H, Raj S, Reeves D, Ristovska M,
743 Rizevsky V, Robertson DR, Robertson P, Ruykys L, Saba AO, Santos JM, Sarı HM,
744 Segurado P, Semenchenko V, Senanan W, Simard N, Simonović P, Skóra ME, Slovák
745 Švolíková K, Smeti E, Šmídová T, Špelić I, Srébalienė G, Stasolla G, Stebbing P, Števo
746 B, Suresh VR, Szajbert B, Ta KAT, Tarkan AS, Tempesti J, Therriault TW, Tidbury HJ,
747 Top-Karakuş N, Tricarico E, Troca DFA, Tsiamis K, Tuckett QM, Tutman P, Uyan U,
748 Uzunova E, Vardakas L, Velle G, Verreycken H, Vintsek L, Wei H, Weiperth A, Weyl
749 OLF, Winter ER, Włodarczyk R, Wood LE, Yang R, Yapıcı S, Yeo SSB, Yoğurtçuoğlu B,
750 Yunnie ALE, Zhu Y, Zięba G, Žitňanová K, Clarke S. 2021. A global-scale screening of
751 non-native aquatic organisms to identify potentially invasive species under current and
752 future climate conditions. *Science of The Total Environment* 788:147868. DOI:
753 10.1016/j.scitotenv.2021.147868.
- 754 Vilizzi L, Hill JE, Piria M, Copp GH. 2022. A protocol for screening potentially invasive non-
755 native species using Weed Risk Assessment-type decision-support tools. *Science of The*
756 *Total Environment* 832:154966. DOI: 10.1016/j.scitotenv.2022.154966.
- 757 Witkowski A. 1996. Introduced fish species in Poland: pros and cons. *Archives of Polish*
758 *Fisheries* 4:101–112.
- 759 Witkowski A. 2009. On the expansion and occurrence of an invasive species - *Pseudorasbora*
760 *parva* (Temminck et Schlegel, 1846) (Teleostei: Cyprinidae: Gobioninae) in Poland.
761 *Fragmenta Faunistica* 52:25–32.
- 762 Witkowski A, Grabowska J. 2012. The Non-Indigenous Freshwater Fishes of Poland: Threats to
763 the Native Ichthyofauna and Consequences for the Fishery: A Review. *Acta Ichthyologica*
764 *Et Piscatoria* 42:77–87. DOI: 10.3750/AIP2011.42.2.01.
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Figure 1

Map of the rivers (Pilica, Bzura and Skrwa Prawa) assessed using the Dispersal-Origin-Status-Impact (DOSI) scheme.



Figure 2

Flow diagram illustrating the proposed classification scheme for populations entering a novel environment.

A species' **D**ISPERSAL mechanism can be assisted from its place of origin either *deliberately* (a_i) or *accidentally* (a_{ii}), or it can migrate *independently* of direct human intervention (b_i) by being *facilitated* or by exploiting human-driven environmental changes (b_{ii}), such as canals. The **O**RIGIN of a species that has its distribution shifted according to the mechanisms described can be *allochthonous* (2a) (not from 'here', with 'here' defined by the spatial scale of interest) or *autochthonous* (2b) (from 'here', as with local species moving within the region of focus). The definition of *allochthonous* or *autochthonous* can also depend on the time elapsed since the species' arrival (e.g., geological time, ancient introductions). **S**TATUS refers to the state of the species' population(s), defined by *abundance* or *range size* (*expanding*, *static*, or *shrinking*). These assessments depend on the duration of the species' presence, the measurement effort applied to assess population change, and the effectiveness of interventions (if any). The **I**MPACT category assesses whether the species causes harm to one or more sectors (ecology, economy, culture, human health). This assessment ranges from little to extensive harm or determines if the species is benign (no effect).

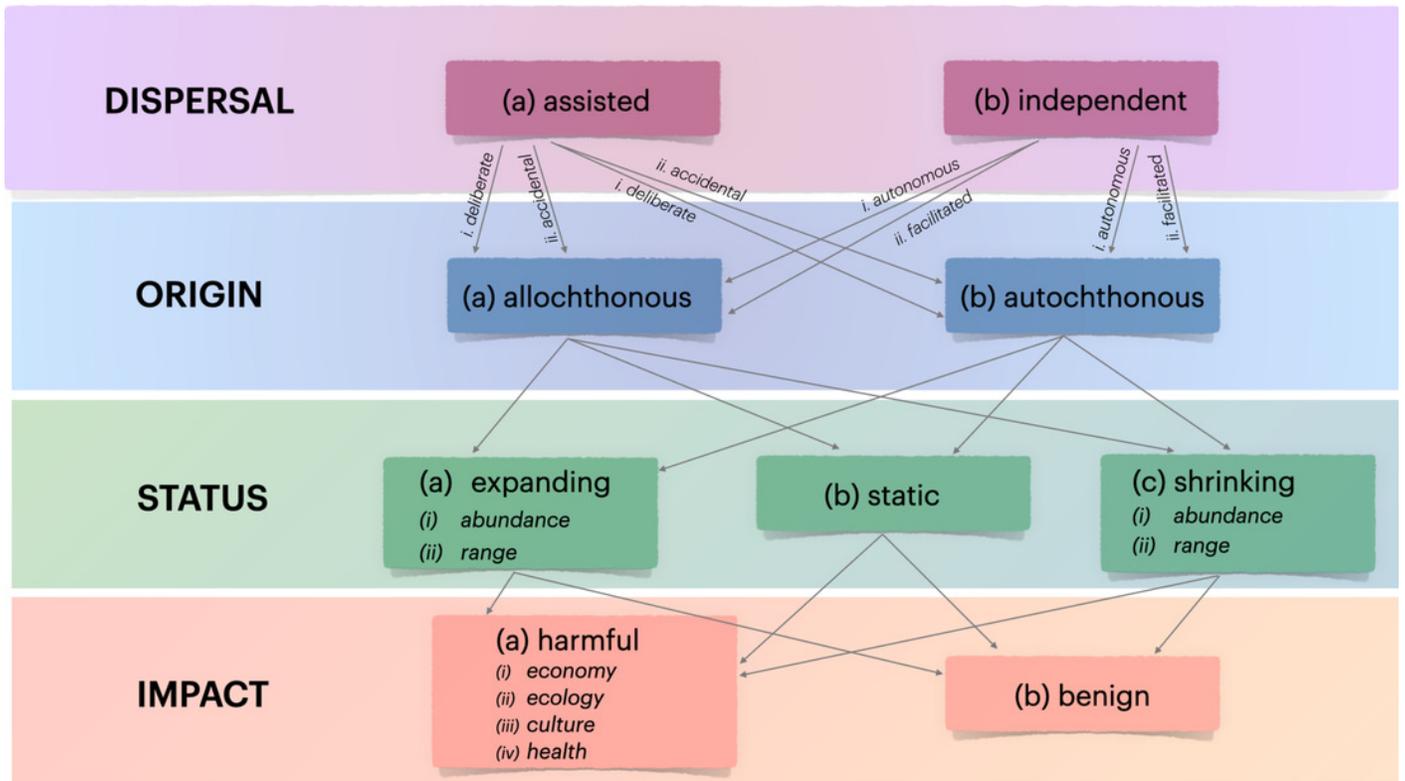


Figure 3

Priority ranking for management interventions of non-native populations based on the Dispersal-Origin-Status-Impact (DOSI) assessment scheme (Supplement 2)

(a) populations dispersing primarily without human assistance, and (b) populations dependent on human assistance for dispersal. See supplement figure for a definition of the various priority classes.

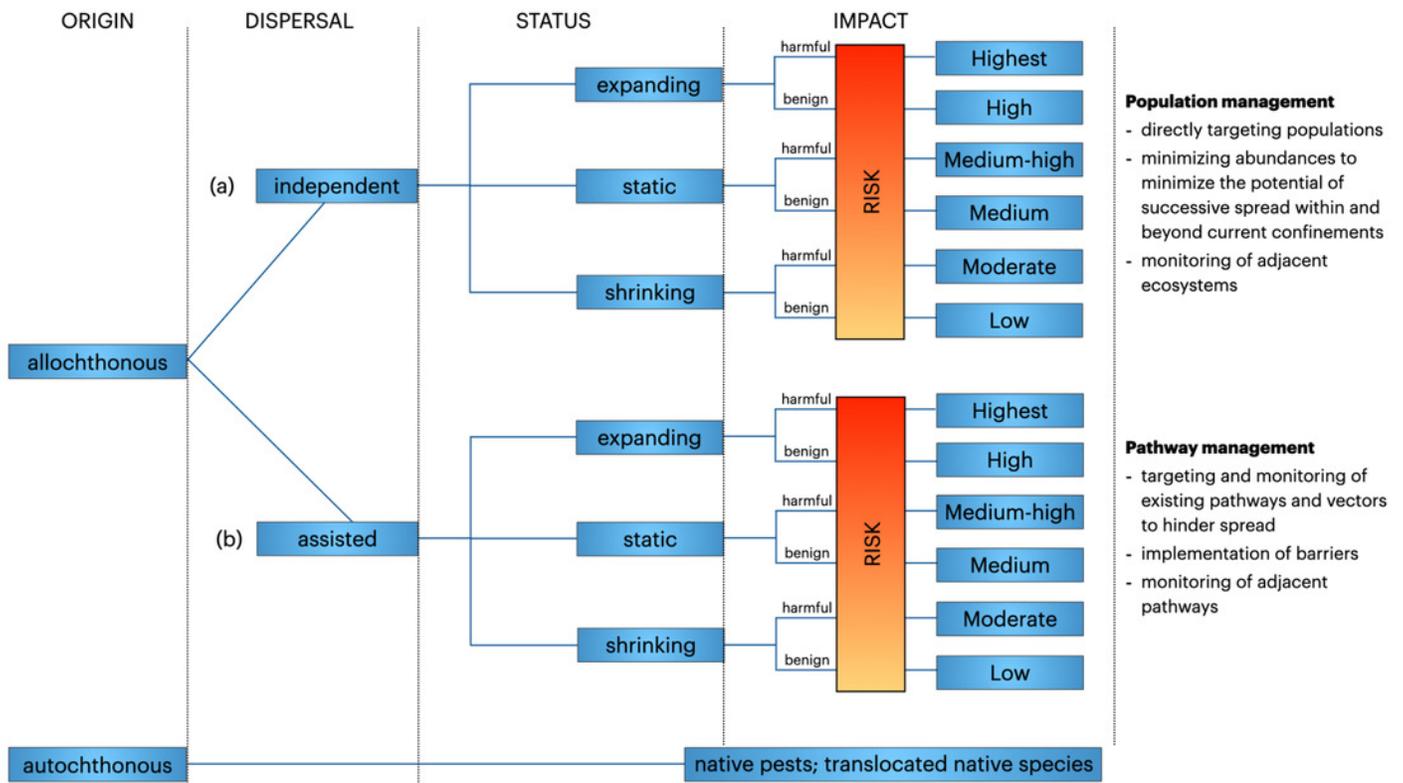


Figure 4

Ranking of established non-native fish species for management targeting populations in a) Pilica, b) Bzura and c) Skrwa Prawa Rivers following the assessment with the Dispersal-Origin-Status-Impact (DOSI) scheme.

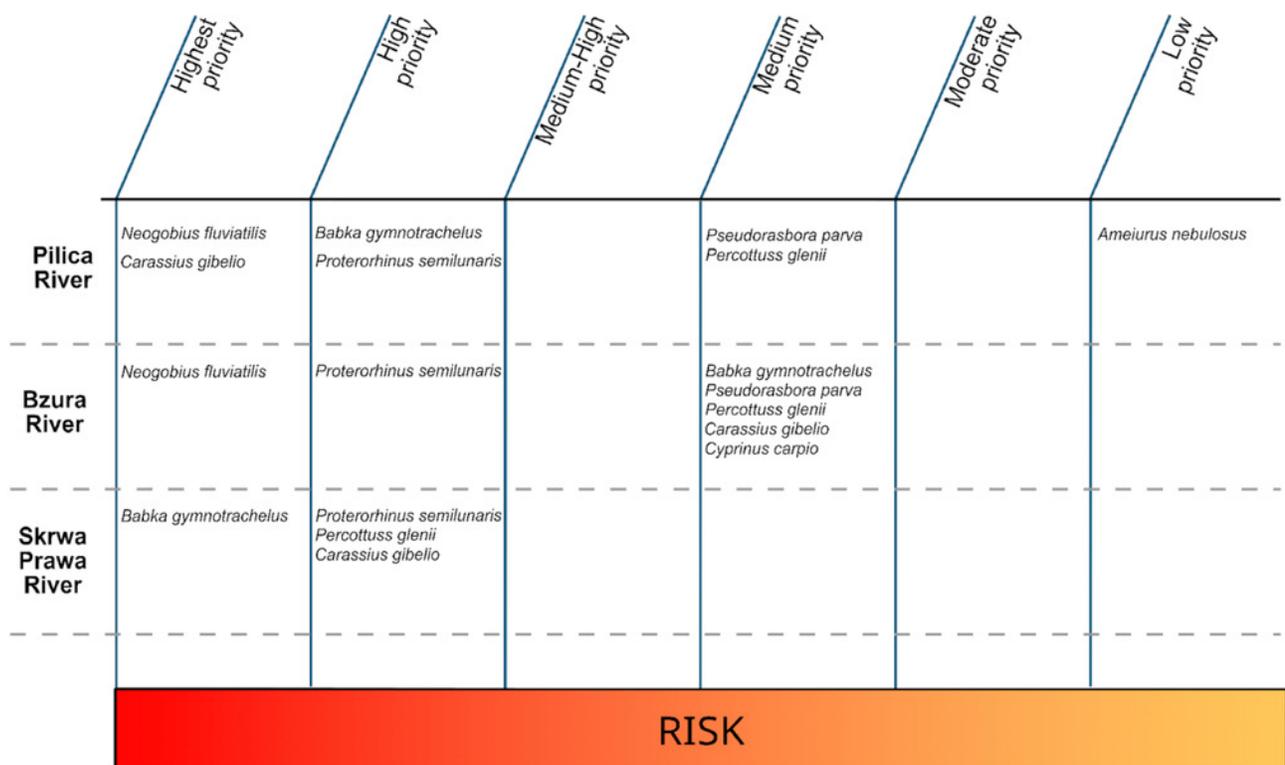


Table 1 (on next page)

Summary of non-native fish species occurrence found in each river (Pilica, Bzura and Skrwa Prawa).

- 1 **Table 1.** Summary of non-native fish species occurrence found in each river (Pilica, Bzura and
2 Skrwa Prawa).

Species	Common name	Pilica	Bzura	Skrwa Prawa
<i>Babka gymotrachelus</i>	racer goby	+	+	+
<i>Neogobius fluviatilis</i>	monkey goby	+	+	-
<i>Proterorhinus semilunaris</i>	western tubenose goby	+	+	+
<i>Percottuss glenii</i>	Chinese sleeper	+	+	+
<i>Ameiurus nebulosus</i>	brown bullhead	+	-	-
<i>Carassius gibelio</i>	gibel carp	+	+	+
<i>Pseudorasbora parva</i>	topmouth gudgeon	+	+	-
<i>Cyprinus carpio</i>	common carp	-	+	-

3

Table 2 (on next page)

Comparison of Dispersal-Origin-Status-Impact (DOSI) assessment scheme ranking and applied in Poland in 2018 Harmonia^{+PL} assessment of non-native freshwater fish in three evaluated rivers Pilica, Bzura and Skrwa Prawa.

1 **Table 2.** Comparison of Dispersal-Origin-Status-Impact (DOSI) assessment scheme ranking and applied
 2 in Poland in 2018 Harmonia^{+PL} assessment of non-native freshwater fish in three evaluated rivers Pilica,
 3 Bzura and Skrwa Prawa.

species	DOSI (Pilica)	DOSI (Bzura)	DOSI (Skrwa Prawa)	Harmonia ^{+PL}
<i>Babka gymotrachelus</i>	High	Medium	Highest	Potentially invasive
<i>Neogobius fluviatilis</i>	Highest	Highest	-	Potentially invasive
<i>Proterorhinus semilunaris</i>	High	High	High	Potentially invasive
<i>Percottuss glenii</i>	Medium	Medium	High	Moderately invasive
<i>Ameiurus nebulosus</i>	Low	-	-	Moderately invasive
<i>Carassius gibelio</i>	Highest	Medium	High	-
<i>Pseudorasbora parva</i>	Medium	Medium	-	Moderately invasive
<i>Cyprinus carpio</i>	-	Medium	-	-

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