Taxon-dependent diversity response along a temperate elevation gradient covered by grassland (#89456)

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Taxon-dependent diversity response along a temperate elevation gradient covered by grassland

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Elevational gradients constitute excellent systems for understanding the mechanisms that generate and maintain global biodiversity patterns. The climatic gradient associated with elevation produces a sequence of different habitat types that influences the distribution of animal species. The study of the same habitat type along an elevation gradient is an ideal scenario to compare alternatives to the energy hypotheses. Our aim was to investigate how changes in climatic conditions along the elevational gradient drive α - and β -diversity of four taxa in a mountain system located within a grassland biome. We sampled ants, spiders, birds and plants, and measured climatic variables at six elevational bands (with 10 sampling sites each) established between 470 and 1000 masl on a mountain from the Ventania Mountain System, Argentina. Species richness per site and β-diversity (turnover and nestedness) between the lowest band and upper sites were estimated. To assess the response of β-diversity to elevation and each climatic variable we used generalized linear models (GLMs) and ranked climatic models following the Akaike information criterion. For most taxa, species richness declined at high elevations and energy, through temperature, was the major driver of species richness for ants and plants, and through productivity for birds, prevailing over water availability. The major β-diversity component was turnover for plants, spiders and birds, and nestedness for ants. The unique environmental conditions of the upper bands favoured the occurrence of specialist and endemic species. In highelevation areas global climate warming may lead to biota homogenization through the loss of specialists and to an increase in species richness.

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a Abstract

Elevational gradients constitute excellent systems for understanding the mechanisms that
generate and maintain global biodiversity patterns. The climatic gradient associated with
elevation produces a sequence of different habitat types that influences the distribution of animal
species. The study of the same habitat type along an elevation gradient is an ideal scenario to
compare alternatives to the energy hypotheses. Our aim was to investigate how changes in
climatic conditions along the elevational gradient drive $\alpha\text{-}$ and $\beta\text{-}$ diversity of four taxa in a
mountain system located within a grassland biome. We sampled ants, spiders, birds and plants,
and measured climatic variables at six elevational bands (with 10 sampling sites each)
established between 470 and 1000 masl on a mountain from the Ventania Mountain System,
Argentina. Species richness per site and β -diversity (turnover and nestedness) between the lowest
band and upper sites were estimated. To assess the response of β -diversity to elevation and each
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following the Akaike information criterion. For most taxa, species richness declined at high
elevations and energy, through temperature, was the major driver of species richness for ants and
plants, and through productivity for birds, prevailing over water availability. The major $\beta\text{-}$
diversity component was turnover for plants, spiders and birds, and nestedness for ants. The
unique environmental conditions of the upper bands favoured the occurrence of specialist and
endemic species. In high-elevation areas global climate warming may lead to biota
homogenization through the loss of specialists and to an increase in species richness.



Keywords

38 climatic drivers, energy hypothesis, mountain, multi-taxa, nestedness, turnover

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Introduction

41	Life is heterogeneously distributed on Earth and the attributes of biological communities most
42	often vary in space, giving rise to patterns. The study of the spatial or geographical processes
43	underlying these patterns is a topical issue in ecology (Hawkins et al., 2003). In this regard,
44	elevational gradients are considered valuable systems to better understand the mechanisms
45	involved in their generation and maintenance (Lomolino, 2001; McCain, 2009; Peters et al.,
46	2016; Ramos et al., 2021). Mountain gradients offer a variety of conditions useful to answer
47	particular questions about diversity drivers (McCain, 2009; Sundqvist, Sanders & Wardle, 2013).
48	In mountain systems temperature decreases with elevation while water availability depends on
49	the context (McCain, 2009). In general, the climatic gradient associated with an increase in
50	elevation creates a sequence of habitat types (i.e. steppe, forest and grassland).
51	The structuring of animal communities is not only determined by climate but also by changes in
52	habitat type, either directly through resource provision, or indirectly by modifying environmental
53	conditions (i.e. soil temperature and humidity; (Werenkraut, Fergnani & Ruggiero, 2015).
54	However, it is still unclear how climatic drivers affect diversity along elevational gradients with
55	a single habitat type (e.g. grassland).





Species richness has been the most studied component of diversity along elevational gradients
(McCain & Grytnes, 2010) and energy and water are known to be the main drivers of the
geographical distribution and diversity patterns of species (Hawkins et al., 2003). An increase in
incident energy favours an increase in taxonomic richness by providing support to species that
otherwise would be limited by physiological constraints. On this basis, species richness in
mountains is expected to decrease with decreasing temperature (Sanders et al., 2007). In animal
communities, energy and water conditions may limit species richness by affecting the
availability of trophic resources given by the productivity of the systems (Mittelbach et al.,
2001). Temperature decreases monotonically with elevation, while water availability may
decrease or increase and their combined effect results in different patterns of primary
productivity (McCain, 2009). Changes in habitat type creates new niche opportunities for animal
species and contributes to diversify the forms of environmental resource exploitation, thus
increasing species diversity (Tews et al., 2004). Moreover, the fact that in most habitats plant
communities determine the structure of the environment has a strong influence on the
distribution of and interaction between animal species (Atauri, De Lucio & Lucio, 2001; Tews et
al., 2004; Fergnani, Sackmann & Ruggiero, 2010). There is little information on the combined
response of various taxa to environmental factors along an elevation gradient, and the few
available studies have included different habitat types in the analyses (Peters et al., 2016; Di
Nuzzo et al., 2 ¹ .). An elevation gradient with a single habitat type is optimal for comparing
variants of the energy hypotheses, and whether diversity patterns respond to a single main driver
or depend on the taxon.





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Changes in the species composition of communities (β -diversity) are crucial for understanding the causes of variation in spatial patterns of species distribution (Liira et al., 2008). It has been found that changes in the taxonomic composition of species may be associated with changes in environmental conditions or geographical distance (MacArthur, 1972; Hubbell, 2001). Association with changes in environmental factors provide evidence that differences in species composition are due to differences in the degree of species specialization for the different dimensions of their ecological niche (MacArthur, 1972), while association with geographical distance suggests species dispersal limitations (Condit et al., 2002). Therefore, the degree of taxonomic similarity between communities separated by a short geographical distance and in the absence of geographical barriers to dispersal will depend on the similarity of their environments and the species capacity to adapt to the environment (Qian, Ricklefs & White, 2005; Steinitz et al., 2006). Mountains are characterized by a high degree of environmental variation over a short geographical distance, making them an ideal setting to investigate changes in dissimilarity of species composition resulting from environmental changes (Jankowski et al., 2009; Tang et al., 2012). However, the relative contribution of the components of β -diversity (i.e. turnover and nestedness) to these changes remains unclear. In this regard, mountains with homogeneous habitat types provide an opportunity to explore changes in species composition due to other factors. A nested pattern is expected when a subset of generalist species found at lower elevations can support the environmental conditions at upper elevations, while a turnover pattern is likely to occur if species are specialists that only persist at higher elevations. Therefore, the analysis of the patterns of β-diversity components among assemblages allows us to delve deeper into the causes underlying changes in α -diversity along elevational gradients.

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A multi-taxa approach is a promising method to elucidate complex community diversity patterns along elevational gradients (Peters et al., 2016; Di Nuzzo et al., 2021). Since assemblages composed of the same taxon tend to share many ecological niche characteristics, they are expected to respond similarly to environmental changes, giving rise to idiosyncratic responses. The multi-taxa approach involving phylogenetically distant taxa allows to reduce this effect. Indeed, broadening the taxonomic scope may lead to changes in some diversity patterns and reveals the importance of potential climate and environmental drivers of diversity (Di Nuzzo et al., 2021). Our aim was to investigate the effect of changes in climatic conditions on the diversity of four taxa (i.e. birds, ants, spiders, and plants) along an elevational gradient within a grassland biome.

Methods

Study area

The Ventania Mountain range is located at the southwest of Buenos Aires province, Argentina, in the Pampas plain. It is an isolated system characterized by steep slopes and a rugged terrain with numerous ravines, and the landscape is dominated by grasslands without trees or shrubs. This mountain system is ancient and dates back to 280-500 million years ago; it supports a rich biodiversity and a high level of endemism for several taxa (Kristensen & Frangi, 1995). Some examples of endemic vegetation are *Senecio ventanensis* (Asteraceae), *Poa iridifolia* (Poaceae), *Adiantum thalictroides* (Adiantaceae) and *Olsynium junceum* (Iridaceae) (Cuevas & Zalba,





122	2009). Moreover, a sun-spider species collected in the present study was described as a new
123	species, Gaucha casuhati (Botero-Trujillo, Ott & Carvalho, 2017).
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125	This study was conducted in the Ernesto Tornquist Provincial Park (38° 03′ 41″ S, 61° 59′ 18″
126	W) located in the centre of the Ventania System, which is one of the few conservation areas in
127	the Pampean grasslands. Permission of Organismo Provincial para el Desarrollo Sostenible de la
128	Provincia de Buenos Aires (number: 057/11). The peaks range from about 450 to 1,200 masl.
129	The climate is temperate, with a mean annual temperature of 15°C and a mean annual rainfall of
130	700 mm (Kristensen & Frangi, 1995).
131	Study design
132	A series of six elevational bands were established covering the entire elevational gradient of the
133	Ventana Mountain, from the valley at the foot of the mountain to near the top, separated by
134	approximately 100 m of elevation (i.e. at average heights of 471, 517, 625, 723, 844 and 1001
135	masl). Fieldwork was carried out in spring and summer (mid-November to mid-January).
136	
137	We used phylogenetically distant taxa: ants, spiders, birds and plants. These groups were
138	selected because they have proven to be sensitive to environmental changes along elevational
139	gradients, are very abundant and diverse, and provide fundamental ecosystem functions
140	(Sekercioglu, 2012; Peters et al., 2016). Ants tend to exhibit high local diversity, performing
141	numerous ecological functions in a wide variety of niches (Hölldobler & Wilson, 1990). They
142	are thermophiles and their diversity tends to decrease at low temperatures (Hölldobler & Wilson,
143	1990; Sanders et al., 2007). Spiders are a ubiquitous component of invertebrate assemblages and





important generalist predators in ecosystems (Wise, 1995). Their richness is associated with prey availability, which is often highly correlated with habitat structural complexity and primary productivity (Aisen et al., 2017). Birds are sensitive to alterations in vegetation structure (Phifer et al., 2017) and respond to changes in primary productivity along gradients (Harrower et al., 2017). Plants are sensitive to changes in abiotic conditions and provide habitat for animals, influencing their diversity and distribution (Nic Lughadha et al., 2005). Water and energy are crucial for plant physiological processes, thus having direct effects on plant diversity (O'Brien, 2006).

Sampling of ants and spiders and taxonomic identification

At each elevational band, we established ten sampling sites spaced at least 50 m apart and installed a pitfall trap at each sampling site, making up a total of 60 traps. Pitfall traps are widely used to obtain representative samples of ant and spider assemblages (Agosti et al., 2000; Pinto et al., 2018). These consisted of plastic containers (500 ml in volume and 85 mm in diameter) partially filled with a solution of propylene glycol: water (1:2) Traps were inserted flush with the ground surface and protected by a plastic cover to avoid flooding by rain. They remained open for three one-week periods. Captured specimens were identified to species level whenever possible or to morphospecies (hereafter referred to as species) based on diagnostic characters of the genera or families, using taxonomic keys and specialist consultations.

Bird survey

- Birds were surveyed using the point-count method with a fixed 50-m radius (Ralph et al., 1996).
- At each elevational band, ten point-count sites were established systematically 150 m apart from





each other. Surveys were conducted on clear and calm days, from dawn to the following 4 h. At each point-count site, we recorded all birds seen or heard over 5 min, except for birds flying overhead. The same two observers conducted all surveys.

Plant sampling and identification

Plant species were estimated using 1-m² and 12-m² quadrats for herbaceous plants and shrubs, respectively (Kent, 2012). This procedure was repeated twice near the point-count sites for birds (thus resulting in ten points per elevational band), and the species recorded in each subsample were pooled to obtain a species list by site. Plant species that could not be identified in the field were hereolized for further identification in the laboratory. Plant taxonomy follows the Catálogo de Plantas Vasculares del Cono Sur and the online update (http://www.darwin.edu.ar).

Environmental gradient characterization

At each elevational band, an automatic sensor of temperature and relative humidity (HOBO U23002) was placed 10 cm above the ground level and gathered data at 6-h intervals for one month. Primary productivity was estimated at each level using a soil-adjusted vegetation index (SAVI) (Huete, 1988). SAVI maps covered the entire length of the elevational gradient and the average value for each elevational level was extracted. The maps were derived from Landsat-5 satellite images, taken in the springs of 2010 and 2011.



Data analyses

The trend of each environmental variable along the elevational gradient was described by a
scatter plot as a function of elevation. Locally weighted regressions were fitted and trends were
overlapped on the plots to visually analyze the elevational pattern of each environmental variable
(Cleveland, Grosse & Shyu, 1993). These were fitted, by least squares with a span of 1.5, using
the loess function of R sof re (Ramos et al., 2021). Then, species renness was calculated as
the number of species of each taxon per sampling site. The elevational trend of species richness
was analyzed using generalized linear models (GLMs) with this variable as a function of
elevation. The distributions used were Poisson for total richness, ants, epiders and birds, and
negative binomial for plants to avoid overdispersion (R, packages stats, glmmTMB and visreg).
We used a model selection approach (Burnham & Anderson, 2002) to identify the main
environmental factors driving changes in species richness along the elevational gradient. For
each taxon, we ran GLMs with species richness as the response variable and each environmental
variable as the independent variable, and ranked them following the Akaike information criterion
(AIC) using the model.sel function of the R package MuMin (Barton et al., 2016). Models with a
difference in AIC values less than two were considered equivalent to the minimum AIC model
(Burnham & Anderson, 2002), and hence they were selected as the best set of models (Ramos et
al., 2018).
Changes in species composition along the elevational gradient were estimated as taxonomic
dissimilarity considering the turnover and nestedness components (Baselga, 2010). For each
taxon, all species recorded at the sampling sites of the lowest elevational band were pooled. The
change in species composition of the assemblages along the elevational gradient was estimated
by comparing the dissimilarity at the lowest band with that recorded at each sampling site of the





upper bands, based on species presence data (Santoandré et al., 2019). Finally, we compared patterns of taxonomic dissimilarity between taxa along the elevation gradient using a GLM with Beta distribution, since the dissimilarity values fell within the (0-1) interval (Ramos et al., 2018). We only compared the trends of the β -diversity components because the magnitude of the effect was influenced by data pooling.

Results

216	We identified a total of 176 plant species (1,417 presence records), 32 ant species (485), 48
217	spider species (206) and 31 bird species (236) collected and/or observed during samplings
218	(Supplementary material). The environmental variables showed different elevational patterns:
219	temperature (Fig. 1A) and SAVI (Fig. 1B) decreased with elevation, while relative humidity
220	increased with elevation (Fig. 1C).
221	
222	The regression analysis of changes in the species richness of ants, birds and plants along the
223	elevation gradient showed a decreasing trend at hight-elevations (ants: DF= 57, Chisq p-value=
224	2.822e-7; birds: DF= 58, Chisq p-value= 1.450e-7; plants: DF= 57, Chisq p-value= 1.312e-6)
225	(Fig. 2). However, ants and plants richness showed a sharp decrease from the upper half of the
226	elevational gradient onwards. On the other hand, spiders richness did not change significantly
227	along the altitudinal gradient (DF= 56, Chisq, p-value = 0.378) (Fig. 2). (Model selection analysis
228	indicated that temperature was the main driver for the species richness pattern of ants and plants,
229	and productivity for birds along the elevational gradient (Table 1).





231	Species composition showed an increase in total dissimilarity with elevation for ants (DF=47,
232	Chisq p-value= 1.368e-8), spiders (DF= 47, Chisq p-value =9.001e-4), birds (DF=47, Chisq p-
233	value= 5.910e-9) and plants (DF= 47, Chisq p-value= 1.890e-9) (Fig 3A). Species turnover
234	increased with elevation in spiders (DF= 40, Chisq p-value= 9.657e-4), birds (DF= 47, Chisq p-
235	value = 0.027) and plants (DF= 47, Chisq p-value 1.382e-12) (Fig 3B), but no significant trends
236	were observed in ants (DF= 47, Chisq p-value = 0.554) (Fig 3). On the other hand, the
237	nestedness component increased with elevation in ants (DF=47, Chisq p-value = 1.977e-9), but
238	decreased with elevation in spiders (DF= 46, Chisq p-value=3.479e-4) and plants (DF= 47,
239	Chisq p-value = 6.424e-9), while it showed no significant trends for birds (DF= 47, Chisq p-
240	value= 0.801) (Fig 3C).

Discussion

Energy, primarily through temperature and productivity changes, were the major driver of species richness for most taxa, prevailing over water availability. The species richness of these taxa decreased at hight elevations and changes in species composition (β-diversity) were taxon-dependent. Thus, turnover was observed in plants, spiders and birds, and nestedness in ants.

Despite the fact that upper elevations harboured lower species richness for most taxa, they provided suitable environmental conditions for species, including indemic ones, that were not found at the lowest sites. These results emphasize the key role played by this isolated and ancient mountain system in preserving regional biodiversity.

Temperature was the main environmental variable laining the decreasing elevational richness pattern for ants and plants. This is in agreement with several elevational richness studies





involving a single taxon in different habitat types and mountain contexts (McCain & Grytnes, 2010; Marathe et al., 2020). However, energy acting indirectly, via productivity, explained the decline in bird richness, as suggested by Harrower et al. (2017) for grassland songbird species. In our study, the absence of maximum productivity at intermediate elevations suggests that the Ventania System is composed of mountains with wet bases (McCain, 2007, 2009; Tellería, 2020). This may indicate that water availability was not a limiting factor, while energy had a different effect on species richness along the elevational gradient, possibly due to the similar responses across taxa to the thermal selection pressures and resources limitations that occur with increasing elevation. In the absence of environmental changes induced by habitat type change, energy availability overshadows the other factors (Harrower et al., 2017; Roeder, Roeder & Bujan, 2021).

The comparison of the species composition between the lowest band and each sampling site of the upper bands indicated that even though the taxonomic dissimilarity increased for all taxa, the main component of β-diversity differed among them. Thus, species turnover was the most important component for spiders, birds and plants, and species loss (nestedness) for ants. Species turnover has been reported as the major driver of changes in species composition along elevational gradients (Foord & Dippenaar-Schoeman, 2016; Gebrehiwot et al., 2019). Moreover, this component has been generally associated with changes in habitat type (Foord & Dippenaar-Schoeman, 2016; Aisen et al., 2017; Iijima & Morimoto, 2021). Our results show that species turnover may also occur within the same habitat type along an elevational gradient, probably due to the combined effect of two processes. First, climatic changes during the last millennia could have led to a latitudinal shift between the Patagonian (steppe), Espinal (xerophytic forest) and





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Pampas (grassland) ecoregions. As a result, species from different ecoregions would have colonized the Ventania System at different historical times, and some of them could have persisted due to the high environmental heterogeneity of these mountains (Frangi & Bottino, 1995). Second, the speciation events that occurred in these ancient mountains most likely account for the presence of endenic species at different elevations (Cuevas & Zalba, 2009). The increase in nestedness of ant assemblages suggests that climatic factors at high elevations would have prevented many ant species present at the base from inhabiting upper elevations, where the extreme environmental conditions may not be compatible with their ecological niche. In conclusion, species with different abiotic requirements are expected to find suitable environmental conditions at different elevations. This process may result in turnover in species composition due to the presence of specialists (Marathe et al., 2021); or in nestedness when only a few base-dwelling generalist species can withstand the environmental conditions at upper elevations. The taxa present in this ancient and isolated mountain system are expected to be differentially affected under the current global warming scenario (Barros et al., 2015). In this line of reasoning, ant richness would increase as a response to the relaxation of environmental filters that limit base species to disperse to upper elevations, while the diversity of plants, spiders, and birds would decrease due to the loss of species only found at upper elevations. The latter alternative could be due not only to the new unfavorable environmental conditions, but also to the competitive exclusion of lower elevation species over upper ones (Freeman, Strimas-Mackey & Miller,

2022). Mountain ecosystems are biodiversity hotspots especially threatened by global change

(Löffler et al., 2011; Rahbek et al., 2019), because of the presence of species with small



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distribution areas and high levels of specialization (Viterbi et al., 2020). Finally, global warming may lead to biota homogenization in the Ventania Mo ain System because it may promote a 300 decrease in β -diversity and endemism but, at the same time, an increase in species richness.

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318	Barros VR, Boninsegna JA, Camilloni IA, Chidiak M, Magrin GO, Rusticucci M. 2015. Climate
319	change in Argentina: trends, projections, impacts and adaptation. Wiley Interdisciplinary
320	Reviews: Climate Change 6:151–169. DOI: doi: 10.1002/wcc.316.
321	Barton PS, Sato CF, Kay GM, Florance D, Lindenmayer DB. 2016. Effects of environmental
322	variation and livestock grazing on ant community structure in temperate eucalypt
323	woodlands. Insect Conservation and Diversity 9:124–134. DOI: 10.1111/icad.12151.
324	Baselga A. 2010. Partitioning the turnover and nestedness components of beta diversity. <i>Global</i>
325	<i>Ecology and Biogeography</i> 19:134–143. DOI: 10.1111/j.1466-8238.2009.00490.x.
326	Botero-Trujillo R, Ott R, Carvalho LS. 2017. Systematic revision and phylogeny of the South
327	American sun-spider genus Gaucha Mello-Leitão (Solifugae: Mummuciidae), with
328	description of four new species and two new generic synonymies. Staatliche
329	Naturhistorische Sammlungen Dresden 75:3–44.
330	Burnham KP, Anderson DR. 2002. Model selection and multimodel inference: a practical
331	information-theoretic approach. New York: Springer. DOI:
332	10.1016/j.ecolmodel.2003.11.004.
333	Cleveland W, Grosse E, Shyu W. 1993. Chapter Local Regression Models of Statistical Models
334	in S, edited by JM Chambers and TJ Hastie, 309–376.
335	Condit R, Pitman N, Leigh EG, Chave J, Terborgh J, Foster RB, Núñez PV, Aguilar S, Valencia
336	R, Villa G, Muller-Landau HC, Losos E, Hubbell SP. 2002. Beta-diversity in tropical
337	forest trees. Science 295:666–669. DOI:
338	10.1126/SCIENCE.1066854/SUPPL_FILE/CONDITWEBTABLE.XLS.
339	Cuevas YA, Zalba SM. 2009. Control de pinos invasores en el parque provincial ernesto
340	tornquist (buenos aires): areas prioritarias y analisis de costos. <i>BioScriba</i> 2:76–89.



341	Fergnani PN, Sackmann P, Ruggiero A. 2010. Richness-environment relationships in epigaeic
342	ants across the Subantarctic-Patagonian transition zone. Insect Conservation and
343	Diversity 3:278–290. DOI: 10.1111/j.1752-4598.2010.00105.x.
344	Foord SH, Dippenaar-Schoeman AS. 2016. The effect of elevation and time on mountain spider
345	diversity: a view of two aspects in the Cederberg mountains of South Africa. Journal of
346	Biogeography 43:2354–2365. DOI: 10.1111/jbi.12817.
347	Frangi JL, Bottino OJ. 1995. Comunidades vegetales de la Sierra de la Ventana, Provincia de
348	Buenos Aires, Argentina. Revista de la Facultad de Agronomía 71:93–133.
349	Freeman BG, Strimas-Mackey M, Miller ET. 2022. Interspecific competition limits bird species
350	ranges in tropical mountains. Science 377:416–420. DOI: 10.1126/SCIENCE.ABL7242
351	Gebrehiwot K, Demissew S, Woldu Z, Fekadu M, Desalegn T, Teferi E. 2019. Elevational
352	changes in vascular plants richness, diversity, and distribution pattern in Abune Yosef
353	mountain range, Northern Ethiopia. Plant Diversity 41:220–228. DOI:
354	10.1016/J.PLD.2019.06.005.
355	Harrower WL, Srivastava DS, McCallum C, Fraser LH, Turkington R. 2017. Temperate
356	grassland songbird species accumulate incrementally along a gradient of primary
357	productivity. PLoS ONE 12:1–18. DOI: 10.1371/journal.pone.0186809.
358	Hawkins BA, Field R, Cornell H V., Currie DJ, Guégan JF, Kaufman DM, Kerr JT, Mittelbach
359	GG, Oberdorff T, O'Brien EM, Porter EE, Turner JRGG. 2003. Energy, water, and
360	broad-scale geographic patterns of species richness. <i>Ecology</i> 84:3105–3117. DOI:
361	10.1890/03-8006.
362	Hölldobler B, Wilson E. 1990. <i>The ants</i> . Berlin, Heidelberg: Springer-Verlag. DOI:
363	10.1017/CBO9781107415324.004.



364	Hubbell SP. 2001. The unified neutral theory of biodiversity and biogeography. Princeton
365	University Press.
366	Huete AR. 1988. A soil-adjusted vegetation index (SAVI). Remote Sensing of Environment
367	25:295–309. DOI: 10.1016/0034-4257(88)90106-X.
368	Iijima D, Morimoto G. 2021. Bird community heterogeneity along four gradients of different
369	orientations on a temperate mountain. Ornithological Science 20:65-82.
370	Jankowski JE, Ciecka AL, Meyer NY, Rabenold KN. 2009. Beta diversity along environmental
371	gradients: implications of habitat specialization in tropical montane landscapes. Journal
372	of Animal Ecology 78:315–327. DOI: 10.1111/J.1365-2656.2008.01487.X.
373	Kent M. 2012. Vegetation description and data analysis: a practical approach. Chichester, West
374	Sussex, UK; Hoboken, NJ: Wiley-Blackwell.
375	Kristensen MJ, Frangi JL. 1995. La sierra de la vantana: una isla de biodiversidad. Ciencia Hoy
376	5:25–34.
377	Liira J, Schmidt T, Aavik T, Arens P, Augenstein I, Bailey D, Billeter R, Bukáček R, Burel F,
378	De Blust G, De Cock R, Dirksen J, Edwards PJ, Hamersky R, Herzog F, Klotz S, Kühn I,
379	Le Coeur D, Miklová P, Roubalova M, Schweiger O, Smulders MJM, Van Wingerden
380	WKRE, Bugter R, Zobel M. 2008. Plant functional group composition and large-scale
381	species richness in European agricultural landscapes. Journal of Vegetation Science
382	19:3–14. DOI: 10.3170/2007-8-18308.
383	Löffler J, Anschlag K, Baker B, Finch O-D, Diekkrüger B, Wundram D, Schröder B, Pape R,
384	Lundberg A. 2011. Mountain ecosystem reponse to global change. Erdkunde 65:189–
385	213.



386	Lomolino MV. 2001. Elevation gradients of species-density: Historical and prospective views.
387	<i>Global Ecology and Biogeography</i> 10:3–13. DOI: 10.1046/j.1466-822x.2001.00229.x.
388	MacArthur R. 1972. Geographical Ecology: Patterns in the distribution of species. New York:
389	Harper & Row.
390	Marathe A, Priyadarsanan DR, Krishnaswamy J, Shanker K. 2020. Spatial and climatic variables
391	independently drive elevational gradients in ant species richness in the Eastern Himalaya.
392	PLOS ONE 15:e0227628. DOI: 10.1371/JOURNAL.PONE.0227628.
393	Marathe A, Shanker K, Krishnaswamy J, Priyadarsanan DR. 2021. Species and functional group
394	composition of ant communities across an elevational gradient in the Eastern Himalaya.
395	Journal of Asia-Pacific Entomology 24:1244–1250. DOI:
396	10.1016/J.ASPEN.2021.08.009.
397	McCain CM. 2007. Could temperature and water availability drive elevational species richness
398	patterns? A global case study for bats. <i>Global Ecology and Biogeography</i> 16:1–13. DOI:
399	10.1111/J.1466-8238.2006.00263.X.
400	McCain CM. 2009. Global analysis of bird elevational diversity. Global Ecology and
401	Biogeography 18:346–360. DOI: 10.1111/j.1466-8238.2008.00443.x.
402	McCain CM, Grytnes J. 2010. Elevational Gradients in Species Richness. <i>eLS</i> :1–10. DOI:
403	10.1002/9780470015902.a0022548.
404	Mittelbach GG, Steiner CF, Scheiner SM, Gross KL, Reynolds HL, Waide RB, Willig MR,
405	Dodson SI, Gough L. 2001. What is the observed relationship between species richness
406	and productivity? Ecology 82:2381–2396. DOI: 10.1890/0012-
407	9658(2001)082[2381:WITORB]2.0.CO;2.



408	Nic Lughadha E, Baillie J, Barthlott W, Brummitt NA, Cheek MR, Farjon A, Govaerts R,
409	Hardwick KA, Hilton-Taylor C, Meagher TR, Moat J, Mutke J, Paton AJ, Pleasants LJ,
410	Savolainen V, Schatz GE, Smith P, Turner I, Wyse-Jackson P, Crane PR. 2005.
411	Measuring the fate of plant diversity: towards a foundation for future monitoring and
412	opportunities for urgent action. Philosophical Transactions of the Royal Society B:
413	Biological Sciences 360:359–372. DOI: 10.1098/RSTB.2004.1596.
414	Di Nuzzo L, Vallese C, Benesperi R, Giordani P, Chiarucci A, Di Cecco V, Di Martino L, Di
415	Musciano M, Gheza G, Lelli C, Spitale D, Nascimbene J. 2021. Contrasting multitaxon
416	responses to climate change in Mediterranean mountains. Scientific Reports 2021 11:1
417	11:1–12. DOI: 10.1038/s41598-021-83866-x.
418	O'Brien EM. 2006. Biological relativity to water-energy dynamics. <i>Journal of Biogeography</i>
419	33:1868–1888. DOI: 10.1111/j.1365-2699.2006.01534.x.
420	Peters MK, Hemp A, Appelhans T, Behler C, Classen A, Detsch F, Ensslin A, Ferger SW,
421	Frederiksen SB, Gebert F, Haas M, Helbig-Bonitz M, Hemp C, Kindeketa WJ,
422	Mwangomo E, Ngereza C, Otte I, Röder J, Rutten G, Schellenberger Costa D, Tardanico
423	J, Zancolli G, Deckert J, Eardley CD, Peters RS, Rödel MO, Schleuning M, Ssymank A,
424	Kakengi V, Zhang J, Böhning-Gaese K, Brandl R, Kalko EKV, Kleyer M, Nauss T,
425	Tschapka M, Fischer M, Steffan-Dewenter I. 2016. Predictors of elevational biodiversity
426	gradients change from single taxa to the multi-taxa community level. Nature
427	Communications 7. DOI: 10.1038/ncomms13736.
428	Phifer CC, Knowlton JL, Webster CR, Flaspohler DJ, Licata JA. 2017. Bird community
429	responses to afforested eucalyptus plantations in the Argentine pampas. Biodiversity and
430	Conservation 26:3073–3101. DOI: 10.1007/s10531-016-1126-6.



431	Pinto CM, Santoandré S, Zurita G, Bellocq MI, Filloy J. 2018. Conifer plantations in grassland
432	and subtropical forest: Does spider diversity respond different to edge effect? Journal of
433	Forest Research 00:1–7. DOI: 10.1080/13416979.2018.1506248.
434	Qian H, Ricklefs RE, White PS. 2005. Beta diversity of angiosperms in temperate floras of
435	eastern Asia and eastern North America. Ecology Letters 8:15-22. DOI: 10.1111/J.1461-
436	0248.2004.00682.X.
437	Rahbek C, Borregaard MK, Colwell RK, Dalsgaard B, Holt BG, Morueta-Holme N, Nogues-
438	Bravo D, Whittaker RJ, Fjeldså J. 2019. Humboldt's enigma: What causes global patterns
439	of mountain biodiversity? Science 365:1108–1113. DOI:
440	10.1126/SCIENCE.AAX0149/ASSET/1BC19D96-73CF-4DE6-9877-
441	1B5AF15DD26B/ASSETS/GRAPHIC/365_1108_F5.JPEG.
442	Ralph CJ, Geupel GR, Pyle P, Martin TE, DeSante DF, Milá B. 1996. Manual de métodos de
443	campo para el monitoreo de aves terrestres. Albany, CA. DOI: 10.3145/epi.2006.jan.15.
444	Ramos CS, Isabel Bellocq M, Paris CI, Filloy J. 2018. Environmental drivers of ant species
445	richness and composition across the Argentine Pampas grassland. Austral Ecology
446	43:424–434. DOI: 10.1111/aec.12579.
447	Ramos CS, Picca P, Pocco ME, Filloy J. 2021. Disentangling the role of environment in cross-
448	taxon congruence of species richness along elevational gradients. Scientific Reports 11:1-
449	11. DOI: 10.1038/s41598-021-83763-3.
450	Roeder KA, Roeder DV, Bujan J. 2021. Ant thermal tolerance: a review of methods, hypotheses,
451	and sources of variation. Annals of the Entomological Society of America 114:459-469.
452	DOI: 10.1093/aesa/saab018.



Sanders NJ, Lessard J, Fitzpatrick MC, Dunn RR. 2007. Temperature, but not productivity or
geometry, predicts elevational diversity gradients in ants across spatial grains. Global
Ecology and Biogeography 16:640-649. DOI: doi: 10.1111/j.1466-8238.2007.00316.x.
Santoandré S, Filloy J, Zurita GA, Bellocq MI. 2019. Taxonomic and functional beta diversity of
ants along tree plantation chronosequences differ between contrasting biomes. Basic and
Applied Ecology 41:1–12. DOI: 10.1016/j.baae.2019.08.004.
Sekercioglu CH. 2012. Bird functional diversity and ecosystem services in tropical forests,
agroforests and agricultural areas. Journal of Ornithology 153:153-161. DOI:
10.1007/s10336-012-0869-4.
Steinitz O, Heller J, Tsoar A, Rotem D, Kadmon R. 2006. Environment, dispersal and patterns of
species similarity. Journal of Biogeography 33:1044–1054. DOI: 10.1111/j.1365-
2699.2006.01473.x.
Sundqvist MK, Sanders NJ, Wardle DA. 2013. Community and Ecosystem Responses to
Elevational Gradients: Processes, Mechanisms, and Insights for Global Change. Annual
Review of Ecology, Evolution, and Systematics 44:261–280. DOI: 10.1146/annurev-
ecolsys-110512-135750.
Tang Z, Fang J, Chi X, Feng J, Liu Y, Shen Z, Wang X, Wang Z, Wu X, Zheng C, Gaston KJ.
2012. Patterns of plant beta-diversity along elevational and latitudinal gradients in
mountain forests of China. <i>Ecography</i> 35:1083–1091. DOI: 10.1111/j.1600-
0587.2012.06882.x.
Tellería JL. 2020. Long-term altitudinal change in bird richness in a Mediterranean mountain
range: habitat shifts explain the trends. Regional Environmental Change 20. DOI:
10.1007/s10113-020-01657-y.





476	Tews J, Brose U, Grimm V, Tielbörger K, Wichmann MC, Schwager M, Jeltsch F. 2004. Animal				
477	species diversity driven by habitat heterogeneity/diversity: The importance of keystone				
478	structures. <i>Journal of Biogeography</i> 31:79–92. DOI: 10.1046/J.0305-0270.2003.00994.X.				
479	Viterbi R, Cerrato C, Bionda R, Provenzale A. 2020. Effects of Temperature Rise on Multi-Taxa				
480	Distributions in Mountain Ecosystems. Diversity 2020, Vol. 12, Page 210 12:210. DOI:				
481	10.3390/D12060210.				
482	Werenkraut V, Fergnani PN, Ruggiero A. 2015. Ants at the edge: a sharp forest-steppe boundary				
483	influences the taxonomic and functional organization of ant species assemblages along				
484	elevational gradients in northwestern Patagonia (Argentina). Biodiversity and				
485	Conservation 24:287–308. DOI: 10.1007/s10531-014-0808-1.				
486	Wise DH. 1995. Spiders in ecological webs. Cambridge: Cambridge University Press.				
487					
488	Table 1: Models that included temperature (temp) or productivity (SAVI) as the variable that				
489	best explained species richness for ants (ra), birds (rb) and plants (rp) along the elevation				
490	gradient of the Ventana Mountain, Buenos Aires province, Argentina. Species richness was				
491	explored using Generalized Linear Models (GLMs) with ln-link function. Only models with the				
492	lowest Akaike information criterion value are shown.				
493	Figure 1. Elevational patterns of each environmental variable. The line was fitted to a local				
494	regression model with 95% confidence interval. Figure generated in R software (2023).				
495	Figure 2: Species richness of the studied taxa (i.e. ants, spiders, birds and plants) along the				
496	elevational gradient of the Ventana Mountain, Buenos Aires province, Argentina. Lines indicate				





497	the generalized linear model (GLM) for each taxon, with 95% confidence interval. Dots indicate				
498	the species richness at each sampling site. Points were jittered to avoid overlap.				
499	Figure 3: Taxonomic dissimilarity and its components between the lowest altitudinal band and				
500	each sampling site of the upper bands. Lines indicate the generalized linear model (GLM) for				
501	each taxon, using Beta distribution with 95% confidence interval. The x-axis indicates the				
502	elevation (masl) of each upper band.				



Table 1(on next page)

Table 1

Models that included temperature (temp) or productivity (SAVI) as the variable that best explained species richness for ants (ra), birds (rb) and plants (rp) along the elevation gradient of the Ventana Mountain, Buenos Aires province, Argentina. Species richness was explored using Generalized Linear Models (GLMs) with In-link function. Only models with the lowest Akaike information criterion value are shown.





Model	AIC	w(AIC)	
$ra = e^{(0.208 * temp - 2.342)}$	284	0.76	
$rb = e^{(5.409 * SAVI - 0.113)}$	215	0.99	
$rp = e^{(0.207 * temp - 1.254)}$	434	1.00	

AIC: Akaike information criterion; rounded Akaike weight: w(AIC).

1



Figure 1

Figure 1

Elevational patterns of each environmental variable. The line was fitted to a local regression model with 95% confidence interval. Figure generated in R software (2023).



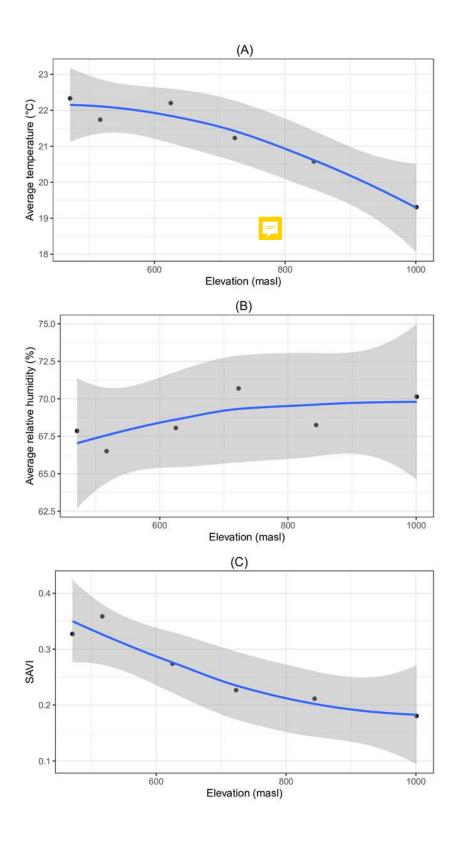




Figure 2

Figure 2

Species richness of the studied taxa (i.e. ants, spiders, birds and plants) along the elevational gradient of the Ventana Mountain, Buenos Aires province, Argentina. Lines indicate the generalized linear model (GLM) for each taxon, with 95% confidence interval. Dots indicate the species richness at each sampling site. Points were jittered to avoid overlap.



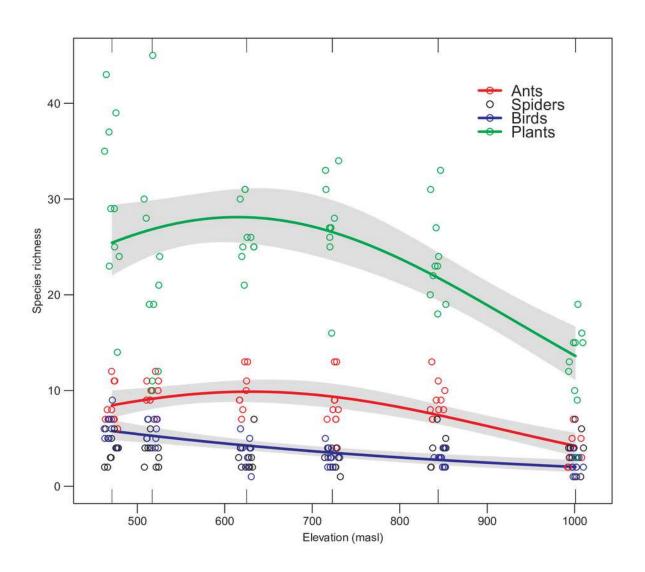




Figure 3

Figure 3

Taxonomic dissimilarity and its components between the lowest altitudinal band and each sampling site of the upper bands. Lines indicate the generalized linear model (GLM) for each taxon, using Beta distribution with 95% confidence interval. The x-axis indicates the elevation (masl) of each upper band.



