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# Characterizing transcriptomic responses to sediment stress across location and morphology in reef-building corals

Jill Ashey<sup>1</sup>, Hailey McKelvie<sup>2</sup>, John Freeman<sup>2</sup>, Polina Shpilker<sup>2</sup>, Lauren H. Zane<sup>1</sup>, Danielle M. Becker<sup>1</sup>, Lenore Cowen<sup>2</sup>, Robert H. Richmond<sup>3</sup>, Valerie J. Paul<sup>4</sup>, Francois O. Seneca<sup>5</sup> and Hollie M. Putnam<sup>1</sup>

<sup>1</sup> Department of Biological Sciences, University of Rhode Island, Kingston, Rhode Island, United States

<sup>2</sup> Department of Computer Science, Tufts University, Medford, Massachusetts, United States

<sup>3</sup> Kewalo Marine Lab, University of Hawaii at Manoa, Honolulu, Hawaii, United States

<sup>4</sup> Smithsonian Marine Station, Smithsonian, Fort Pierce, Florida, United States

<sup>5</sup> Centre Scientifique de Monaco, Monaco, Monaco

# ABSTRACT

Anthropogenic activities increase sediment suspended in the water column and deposition on reefs can be largely dependent on colony morphology. Massive and plating corals have a high capacity to trap sediments, and active removal mechanisms can be energetically costly. Branching corals trap less sediment but are more susceptible to light limitation caused by suspended sediment. Despite deleterious effects of sediments on corals, few studies have examined the molecular response of corals with different morphological characteristics to sediment stress. To address this knowledge gap, this study assessed the transcriptomic responses of branching and massive corals in Florida and Hawai'i to varying levels of sediment exposure. Gene expression analysis revealed a molecular responsiveness to sediments across species and sites. Differential Gene Expression followed by Gene Ontology (GO) enrichment analysis identified that branching corals had the largest transcriptomic response to sediments, in developmental processes and metabolism, while significantly enriched GO terms were highly variable between massive corals, despite similar morphologies. Comparison of DEGs within orthogroups revealed that while all corals had DEGs in response to sediment, there was not a concerted gene set response by morphology or location. These findings illuminate the species specificity and genetic basis underlying coral susceptibility to sediments.

**Subjects** Bioinformatics, Genomics, Marine Biology, Molecular Biology **Keywords** Transcriptomics, Sediment, Corals, RNASeq, Morphology

# **INTRODUCTION**

Coral reefs are incredibly diverse marine ecosystems, providing numerous ecological and economic services such as biodiversity, cultural value, coastal protection, fisheries, and tourism (*Reaka-Kudla, 1997; Sumaila & Cisneros-Montemayor, 2010; Costanza et al., 2014*). Reef-building corals form a critical nutritional symbiotic relationship with unicellular photosynthetic algal endosymbionts in the family Symbiodiniaceae

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Corresponding author Jill Ashey, jillashey@uri.edu

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How to cite this article Ashey J, McKelvie H, Freeman J, Shpilker P, Zane LH, Becker DM, Cowen L, Richmond RH, Paul VJ, Seneca FO, Putnam HM. 2024. Characterizing transcriptomic responses to sediment stress across location and morphology in reef-building corals. PeerJ 12:e16654 DOI 10.7717/peerJ.16654 (*LaJeunesse et al., 2018*). The carbohydrates produced by the algal photosynthesis are translocated to the coral to be used as its primary energy source, supporting the daily respiratory carbon demand of tropical corals (*Muscatine & Porter, 1977; Muscatine et al., 1984*). This coral-algal symbiosis fuels reef productivity and accretion (*Roth, 2014*), but is sensitive to changing environmental conditions that can impact the symbiosis such as light, nutrients, temperature, pH, and sediment (*Hoegh-Guldberg et al., 2007; Davy, Allemand & Weis, 2012*). For example, under exposure to sedimentation, or the downward fall of sediment from the water column toward the benthos (*Schlaefer, Tebbett & Bellwood, 2021*), corals display reduced photosynthetic efficiency (*Weber, Lott & Fabricius, 2006; Rushmore, Ross & Fogarty, 2021*), increased respiration rates (*Riegl & Branch, 1995; Browne et al., 2014*), decreased calcification (*Bak, 1978*), and rapid consumption of energy reserves (*Sheridan et al., 2014*). While sediment transport naturally occurs on reefs, suspended sediment caused by anthropogenic activities such as dredging, runoff, and coastal development have increased (*Rogers, 1990; Fabricius, 2005; Erftemeijer et al., 2012; Miller et al., 2016; Cunning et al., 2019*).

Deposited sediment and suspended sediment are the two primary ways that sediment interacts with corals (Rogers, 1990; Fabricius, 2005; Erftemeijer et al., 2012). Deposited sediment occurs when sediment particles settle directly on the coral surface, making physical contact with the tissue. Passive removal of sediment includes gravity or flow removing it from the coral surface (Lasker, 1980; Jones, Fisher & Bessell-Browne, 2019). In response to deposited sediment, corals can also initiate an acute response to attempt to move the sediment using active mechanisms. Active sediment removal mechanisms include ciliary and tentacle movement, increased mucus production, and hydrostatic inflation (Rogers, 1990; Stafford-Smith & Ormond, 1992; Stafford-Smith, 1993; Bessell-Browne et al., 2017). However, these active mechanisms are often very energetically costly, and thus cannot be sustained for long periods of time (Riegl & Branch, 1995; Erftemeijer et al., 2012). If the sediment deposition rate exceeds the coral's sediment clearance rate, sediment will accumulate on the coral, reducing heterotrophic feeding and light transmission to algal endosymbionts and creating hypoxic conditions near the coral tissue, which often leads to tissue necrosis and coral mortality (Philipp & Fabricius, 2003; Weber, Lott & Fabricius, 2006; Weber et al., 2012).

Corals can also, or alternatively, interact with suspended sediment, which occurs when particles such as clay, silt, and sand are moved into the water column by some natural or anthropogenic disturbance and remain in the water column for a period of time (*Rogers, 1990*; *Fabricius, 2005*; *Erftemeijer et al., 2012*). Suspended sediment reduces the amount of light that reaches the coral, impeding the ability of the algal endosymbionts to photosynthesize and provide the coral host with sufficient energy for metabolism and growth (*Rogers, 1990*; *Fabricius, 2005*; *Erftemeijer et al., 2012*; *Bessell-Browne et al., 2017*). Reduced photosynthetic efficiency can induce corals to switch to heterotrophic feeding, a much less efficient way to obtain carbon than through its endosymbionts (*Muscatine & Porter, 1977*; *Anthony & Fabricius, 2000*; *Houlbrèque & Ferrier-Pagès, 2009*). Additionally, heterotrophic feeding in the presence of sediments may lead the coral to ingest sediment particles, disrupting its nutritional intake and potentially acting as a vector for harmful

bacteria and toxins (*Erftemeijer et al., 2012*; *Studivan et al., 2022*). Suspended sediment has also been observed to induce immune responses and increase disease prevalence in corals (*Pollock et al., 2014*; *Sheridan et al., 2014*).

In addition to sediment type, the morphology of the coral can modulate its interaction with sediments. For example, massive, plating, and encrusting corals have a higher planar surface area and thus higher capacity to trap sediments in comparison to branching corals with high three-dimensional and more vertical structure. Sediment removal from massive corals often requires active removal mechanisms (Dallmeyer, Porter & Smith, 1982; Rogers, 1990; Stafford-Smith, 1993). However, massive corals may also be more resilient to high suspended sediment concentrations because their greater surface area allows for increased opportunities to capture light, maximizing the photosynthetic efficiency of their algal endosymbionts (Fabricius, 2005; Erftemeijer et al., 2012). In contrast, the relatively small surface area and vertical branches of branching coral species means sediment is minimally trapped and can be more easily removed by gravity or currents (Lasker, 1980; Rogers, 1983; Stafford-Smith, 1993). Branching coral species typically have faster clearance rates than non-branching species, and active removal mechanisms are required less frequently, allowing branching corals to devote that energy towards other functions such as reproduction and growth (Stafford-Smith, 1993; Jones, Fisher & Bessell-Browne, 2019). Collectively, these studies support that morphology plays a role in response to sediment stress.

While most research has focused primarily on physiological responses to sediment stress, there is a small but growing body of work on gene expression of corals exposed to sediment stress. Early microarray studies found that the upregulation of heat shock proteins (HSPs) occurred in response to sediment stress (Wiens et al., 2000; Hashimoto et al., 2004). As a generalized stress response protein, HSPs have also been implicated in coral response to other stressors, such as deoxygenation, temperature and ocean acidification (DeSalvo et al., 2008; DeSalvo et al., 2010; Kaniewska et al., 2012; Alderdice et al., 2022). Another general response to sediment stress is upregulation of biomarkers of oxidative stress (Morgan, Edge & Snell, 2005), more specifically, thioredoxin, a protein that modulates redox and cell-to-cell signaling (Tomanek, 2015). A potentially more specific gene responding to sediment stress is indicated by the differential expression of urokinase plasminogen activator surface receptor (uPAR) transcripts in the coral Diploria strigosa along a sedimentation/pollution gradient in Castle Harbor, Bermuda (Morgan, Edge & *Snell*, 2005). uPAR is associated with proteolysis, wound healing and inflammation, and is hypothesized to contribute to coral tissue remodeling in response to elevated levels of sedimentation (Morgan, Edge & Snell, 2005). Genes related to immunity, as well as energy metabolism, were also implicated in the transcriptomic response to sediment stress in corals from two locations, Singapore (Goniastrea pectinata and Mycedium elephantotus) and Eilat, Israel (G. pectinata only) using RNASeq (Bollati et al., 2021). While there were some methodological differences between their experiments, shared mechanisms were identified across different species and populations, demonstrating a conserved response to sediment stress across species and sediment types. However, the corals evaluated in that study, G. pectinata and M. elephantotus, both have similar morphological characteristics

(massive and encrusting), indicating that this response may only be relevant to massive and encrusting corals (*Bollati et al., 2021*).

While these initial molecular analyses have provided important insights into the coral transcriptomic response to sediment stress, less is known about shared molecular responses by morphology and location/sediment type. To this end, our study aims to fill these knowledge gaps by examining gene expression across different coral morphologies and the use of multiple locations/types of sediment. Here we quantified the transcriptomic responses of corals with different colony morphologies in response to different types of sediment stress. Floridian corals (Acropora cervicornis, Montastraea cavernosa and Orbicella faveolata) were exposed to sterilized white carbonate sediment for 18 days, whereas Hawaiian corals (Montipora capitata, Pocillopora acuta (formerly Pocillopora *damicornis*) and *Porites lobata*) were exposed to unsterilized terrigenous red soil for up to 7 days. In this study, A. cervicornis and P. acuta were categorized as branching corals, while M. cavernosa, O. faveolata and P. lobata were categorized as massive corals. The morphology of *M. capitata* was considered as intermediate between branching and plating, as *M. capitata* tends to form plates growing horizontally with branches sprouting upward (Veron, 2002). The methodological differences prevent us from making direct statistical comparisons between the experiments. However, it is still relevant to highlight general biological processes and mechanisms related to sediment stress responses across morphology and location.

# **METHODS**

To characterize a broad range of responses to different kinds of sediment stress, two independent experiments were performed in Florida and Hawai'i; the separate experiments will be referred to as either Florida or Hawai'i throughout the text (Fig. 1). In the first experiment, Hawaiian corals were exposed to live terrigenous red soil for up to 7 days. In the other experiment, Floridian corals were exposed to sterilized coral rubble sediment for 18 days. We acknowledge that differences in experimental sediment regimes do not allow for direct statistical comparisons between these two experiments at all levels. However, qualitatively examining the patterns found in response to sediment across two experiments and in orthogroups across taxa in response to sediment provides the unique opportunity to describe these findings and test for shared transcriptomic patterns in response to sediment stress. Portions of this text were previously published as part of a preprint (*Ashey et al., 2023*; https://www.biorxiv.org/content/10.1101/2023.01.30. 526279v1.full).

#### **Collections and exposures**

#### Hawai'i

All coral collections were obtained under Hawai'i SAP permit (SAP 2015-48). *Montipora capitata, Porites lobata* and *Pocillopora acuta* adult corals were collected in early June 2015 from Kāne'ohe Bay, O'ahu, Hawai'i (21°25′59.1″N 157°47′11.1″W). One fragment was collected per colony, and fragments were acclimated to tank conditions for 1 week post collection. Following acclimation, 12 fragments per species were placed in six 60-L

	Species	Treatments & experiment duration	Experimental set-up
ŋ	a) Acropora carricornis (right) Branching	b) Dates: 5/25/2016 – 6/9/2016 Day 0 Day 18	c)
Florid	Mortestrea cavernosa (above)         Orbicelle fravelatar (right)	Control (0 mg/L)           Treatment 1 (30 mg/L)           Treatment 2 (100 mg/L)           Treatment 3 (300 mg/L)           Treatment 4 (1000 mg/L)	
Hawai'i	d) Montipora capitata (right) Intermediate	e)       Dates: 6/8/2015 - 6/15/2015         Dates: 6/8/2015 - 6/15/2015       Da	n

Figure 1 Experimental design for Florida (A–C) and Hawai'i (D–F) experiments. (A) Species used and their morphologies in parentheses for Florida study: *Acropora cervicornis* (branching morphology), *Montastrea cavernosa* (massive morphology) and *Orbicella faveolata* (massive morphology). (B) Illustration of timeline and sediment treatments. (C) Experimental tanks from the Florida experiment (image by Francois Seneca). (D) Species used and their morphologies in parentheses for Hawai'i study: *Montipora capitata* (intermediate morphology), *Pocillopora acuta* (branching morphology) and *Porites lobata* (massive morphology) (E) Illustration of timeline and sediment treatments. (F) Experimental tanks from the Hawai'i experiment (image by Francois Seneca). Images of coral species were obtained from Corals of the World (*Veron et al., 2016a, 2016b, 2016c, 2016d, 2016e, 2016f*) (C) Corals of the World.

flow-through outdoor tanks (two replicate tanks per treatment and two replicate fragments per tank for n = 4 samples per treatment, except *M. capitata* where n = 3) at the Point Lab of the Hawai'i Institute of Marine Biology (HIMB). Corals in Hawai'i have become increasingly exposed to sedimentation, agricultural runoff, and pollution that contains organic material (Pastorok & Bilyard, 1985; Hunter & Evans, 1995; Ogston & Field, 2010; Weber et al., 2012; Abaya et al., 2018), and the increased prevalence of organic matter in sediment and outflows can negatively affect corals (Hodgson, 1990; Weber, Lott & Fabricius, 2006; Weber et al., 2012; Loiola, Oliveira & Kikuchi, 2013). Given the prevalence of organic matter in sediment in Hawai'i, fragments were exposed to sediment originating from a live terrigenous Hawaiian red soil collected from the highest elevation point on Moku O Lo'e (Coconut Island). The total suspended sediment (TSS) concentration was determined by the technique described in Cortés & Risk (1985). The unfiltered unsterilized terrigenous red dirt was mixed with fresh water to mimic the sediment moved during a storm and extract the silt and clay (<60 micron) used in the experiment as in Bahr et al. (2020). Particles were filtered through mesh sieves to remove debris and produce the final silt-clay mixture. Sediment was added from a concentrated stock at day 0 and again at day 4 to reach a low and a high total suspended sediment (TSS) concentration, immediately

following addition. To test the efficacy of our additions, triplicate samples were taken from the experimental tanks following sediment dosing to measure the sediment load. The sediment samples were filtered and oven dried on Whatman paper, and the sediment weight measured with a precision scale at  $150 \pm 73$  and  $235 \pm 63$  mg/L, respectively (Philipp & Fabricius, 2003; Browne et al., 2014). Powerheads (Eheim Universal 300 L/h) were placed at the bottom of each tank to ensure water motion. Seawater was pumped directly from the lagoon in front of the Point Lab and was filtered through a sand filter to remove sediment before being distributed to the experiment tanks at a flow-through rate of 12 L/h. The temperature exposure was the natural profile of the lagoon water by HIMB during the time of the experiment (Station 51207; NOAA Buoy in Kane'ohe Bay, Hawai'i; Fig. S1A). Salinity and total pH fluctuated between 34–35 psu and 7.90–8.01 during the experimental period, following the water conditions of the lagoon (MPCO2 Mooring Kāne'ohe, NOAA Buoy). The irradiance within the 60-liter experimental tanks was reduced by 30% with shade cloth to obtain a maximal daylight value fluctuating approximately between 60-90 lux, measured with a HOBO pendant data logger (HOBO Pendant Temp/Light 64 K logger) placed at the bottom of the tanks. Tanks received natural day/night light cycles. The coral fragments were not fed in this experiment. Coral fragments (n = 4 replicate fragments per species), two from each of the two tanks, in each of the three treatments were exposed (except *M*. *capitata* where n = 3) and were collected and preserved in liquid nitrogen at day 4 (24 samples on June 12, 2015), and day 7 (24 samples on June 15, 2015) of the exposure. Samples were then stored at -80 °C until extraction.

#### Florida

Adult Acropora cervicornis, Montastraea cavernosa, and Orbicella faveolata colonies were collected from the Key West nursery of the Florida Keys National Marine Sanctuary on March 23, 2016 under NOAA National Marine Sanctuaries Permit #FKNMS-2015-016 and #FKNMS-2016-017. On March 28, 2016, coral colonies were transported from Key West to Fort Pierce in bins filled with aerated seawater and placed in acclimatation tanks at the Smithsonian Marine Station (Fort Pierce, FL, USA) prior to the start of the experiment. After acclimation, 15 fragments per species were placed in fifteen 12-L static outdoor tanks (three replicate tanks per treatment and 1 fragment per tank for n = 3 samples per treatment). The water source was oceanic water and had a consistent salinity of 35 psu. The experimental tanks were static, and reverse osmosis water was added daily to maintain the salinity. Tanks were submerged in recirculating water baths to keep them at a constant temperature of 28 °C. pH was not recorded during the experiment. Corals in Florida are subject to an increased number of sedimentation events, in particular through dredging and coastal development (Barnes et al., 2015; Miller et al., 2016; Cunning et al., 2019); these activities disturb bottom/seafloor sediments, which reduces light availability and can smother corals, leading to a multitude of adverse effects (Erftemeijer et al., 2012; Jones et al., 2016). Given the increased prevalence of dredging and coastal development in Florida, coral fragments were exposed to sterilized carbonate sand sediment intended to mimic dredging and development activities. Sediment was collected from Key Largo

(25°08'22.0"N 80°23'37.6"W) and was filtered through a 63 micron sieve to remove debris and obtain a fine grain mixture. Sediment was then run through active carbon filters for a week in seawater to reduce chemical pollution. Sterilization of the sediment stock was performed using an autoclave program for liquid (20 min at 121 °C and 2.1 bar). Despite being sterilized, the sediment still had the capacity to induce adverse effects in corals. Previous studies have found that sterilized sediment can deplete energy reserves, induce immune responses, and decrease photosynthetic capabilities (Browne et al., 2014; Junjie et al., 2014; Sheridan et al., 2014). Thus, it is likely that the sterilized sediment stressor used in the Florida experiment induced an ecologically relevant response. The total suspended sediment (TSS) concentration was determined by the technique described in Bahr et al. (2020). On the first day of exposure (May 23, 2016), sediment was added from the sterilized concentrated stock (458 g/L). Sediment was added again on May 25, 27, and 30 to mimic repeat dredging activities. There were five total suspended sediment (TSS) 0, 30, 100, 300, 1,000 mg/L targets (Rogers, 1979; Junjie et al., 2014). To approximately obtain these TSS concentrations, 1, 3, 10, and 30 mL of stock was added to each 12-L treatment tank. Powerheads (Eheim Universal 300 L/h) were placed at the bottom of each tank to ensure water motion. Tanks received natural day/night cycles and were under 50% reduction shade, measured with a HOBO pendant data logger (HOBO Pendant Temp/Light 64 K logger) placed at the bottom of the tanks. The coral fragments were not fed in this experiment. On June 9, coral fragments from all tanks were collected and frozen in liquid nitrogen. Samples were then stored at -80 °C until extraction.

# **RNA** extraction

Prior to RNA extraction, frozen coral samples were crushed with a manual hydraulic press (12 tons pressure) and a metal mortar and pestle chilled with liquid nitrogen. Approximately 100 mg of frozen coral powder was used in the RNA extractions starting with 1 mL of TRIzol<sup>TM</sup> Reagent (Invitrogen, Waltham, MA, USA) and 100 uL of 0.1 mm ceramic beads in 2.0 mL tubes. Coral tissues were lysed with two rounds of 20 s at 6,500 bpm on a tissue homogenizer (FastPrep), and a rest on ice of 30 s in between rounds. Tissue slurries were then incubated on ice for 5 min. After a brief spin down to gather liquid at the bottom of the tube, 300 uL of molecular grade chloroform was added to the slurry. Tubes were hand shaken for 15 s and rested for 3 min on ice. Phase separation was obtained by centrifugation at 12 xg for 15 min at 4 °C. Top aqueous phase containing RNA was transferred to new 1.5 mL tubes and mixed again with 200 uL of chloroform. The previous three steps were repeated. An equal volume of chilled 70% molecular grade ethanol was added to the RNA in solution. From this point, the Direct-zol<sup>TM</sup> RNA MiniPrep (Zymo Research; Cat# R2070, Orange, CA, USA) protocol with DNase treatment was followed. Total RNAs were eluted in DEPC-treated water and quantified using the Qubit fluorometer. RNA quality was assessed using a NanoDrop spectrophotometer in addition to a visual check for degradation via a 1% agarose TAE gel. Samples with at least 9 ug/uL of RNA were sent for sequencing.

#### **RNA** sequencing

RNA samples were diluted to accommodate for library production starting with 100 ng of total RNA. Samples were then loaded onto Neoprep cards and processed following the TruSeq stranded mRNA Library Prep for NeoPrep kit (Document # 15049725 v03; Illumina, San Diego, CA, USA) protocol. Quality controlled libraries were then sequenced through HiSeq 50 cycle single read sequencing v4 by the High Throughput Genomics Core Facility at the University of Utah.

#### **Bioinformatic analysis**

Workflows and data are located at https://github.com/JillAshey/SedimentStress. Quality checks of raw and trimmed reads were performed using FASTQC (v0.11.8, Java-1.8; Andrews, 2010) and MultiQC (v1.7, Python-2.7.15; Ewels et al., 2016). Reads that did not pass quality control were trimmed with Trimmomatic (v0.30, Java-1.8; Bolger, Lohse & Usadel, 2014). Genomes, protein sequences, and transcript sequences were obtained from the following locations: Acropora cervicornis (Baums Lab, v1.0\_171209; https://usegalaxy. org/u/skitch/h/acervicornis-genome); Montastraea cavernosa (Matz Lab, July 2018 version; https://matzlab.weebly.com/data-code.html); Montipora capitata (http:// cyanophora.rutgers.edu/montipora/; Version 2, Stephens et al., 2021); Orbicella faveolata (NCBI, assembly accession GCF\_002042975.1; Prada et al., 2016); Pocillopora acuta (http://cyanophora.rutgers.edu/Pocillopora\_acuta/; Version 1, (Stephens et al., 2021); Porites lutea (used to analyze P. lobata data in current study; http://plut.reefgenomics.org/ download/; Version 1.1, Robbins et al., 2019). We have archived all references used for this analysis at https://osf.io/8qn6c/ (DOI 10.17605/OSF.IO/8QN6C) to enable reproducible analyses for this project. After trimming, reads were mapped to their respective genomes using STAR (v2.5.3; Dobin et al., 2013). The aligned read files from STAR (BAM file format) were assembled to the references and count data were generated using StringTie (v2.1.1-GCCcore-7.3.0; Pertea et al., 2015). Assembly quality was assessed with gffcompare (v0.11.5; Pertea & Pertea, 2020). The StringTie prepDE python (v2.7.15; Pertea et al., 2015) script was used to generate a gene count matrix.

To generate current gene ontology information for all species, functional annotation was performed on all genomes using the following workflow (https://github.com/ JillAshey/FunctionalAnnotation). First, protein sequences from each species were identified using BLAST (*blastp*; v2.11.0; *Altschul et al.*, 1990) against NCBI's nr database (1e-5 e-value; database accessed and updated on Oct. 12, 2021) and the Swissport database (1e-5 e-value; database accessed and updated on Oct. 22, 2021; *Bairoch & Apweiler*, 1997). The XML files generated from the BLAST output were then used as input for BLAST2GO (v.5.2.5) to generate gene ontology (GO) terms (*Götz et al.*, 2008). Protein sequences were also used as input to InterProScan (v5.46–81.0; Java v11.0.2), which identified homologous sequences and assigned GO terms (*Jones et al.*, 2014). Using BLAST2GO, the XML file generated from InterProScan was merged with the nr and Swissprot BLAST2GO output, creating final functional annotation tables, which were saved in csv format (https://osf.io/8qn6c/).

#### Gene expression and ontology analysis

All gene expression and ontology analyses were done in RStudio (v1.3.959) using v4.0.2 of R. First, genes were filtered using genefilter's (v1.70.0; Gentleman et al., 2021) pOverA function; genes were retained for expression analysis only if counts were greater than or equal to 5 in at least 85% of the samples, which minimizes differential expression results from low count genes with lower confidence. Because different samples may have been sequenced to different depths, size factors were calculated as the standard median ratio of a sample over a 'pseudosample' (for each gene, the geometric mean of all samples; Anders & Huber, 2010; Love, Huber & Anders, 2014). After confirming size factors were estimated to be less than 4, filtered gene counts were normalized using DESeq's (v1.28.1) vst function (Love, Huber & Anders, 2014). Treatment was set as a factor; in the HI experiment, Time (Day 4 and Day 7) was not significant as a factor and so corals sampled on different days were combined by treatment for further analysis. Differential gene expression was assessed using the *DESeq* function with the Wald likelihood test ratio (p-adjusted < 0.05; *Love*, Huber & Anders, 2014). PCA plots with differentially expressed genes were generated using the plotPCA function (Love, Huber & Anders, 2014). To estimate the power to detect a range of effect sizes (1.25, 1.5, 1.75, and 2), the R package RNASeqPower (v1.38.0; Hart et al., 2013). Depth of sequencing was calculated using the formula derived from Sims et al. (2014) as LN/G, where L is read length, N is the average number of reads and G is the transcriptome length. The power analysis using the filtered gene datasets (Table S1) indicates the power to detect an effect was similar across the six species (effects size standard error of the mean ranged from 0.012 to 0.06 across the range of effect sizes tested).

Gene ontology analysis was completed with GOSeq (v1.40.0; *Young et al., 2010*), which corrects for the higher-confidence in differential expression as a function of gene length as follows: Genes that passed the *pOverA* filter and were marked as differentially expressed genes by DESeq above were used to calculate the probability weighting function using function *nullp* with the bias data being gene length (*Young et al., 2010*). To identify category enrichment amongst differentially expressed genes, the *goseq* function was performed with the Wald method (*Young et al., 2010*). Significantly enriched GO terms from each of the biological processes (BP), molecular functions (MF), or cellular components (CC) ontologies were denoted as those with an over-represented *p*-value < 0.05. BP GO term analysis is presented here; CC and MF GO term analyses can be found in the Supplemental Material (See Figs. S2–S7). GO term information was organized under their parent GO slim terms (obtained from http://www.informatics.jax.org/gotools/data/ input/map2MGIslim.txt; accessed on April 4, 2021) for qualitative comparison across species. Overlap of GO terms between species was compared and visualized using ComplexUpset's (v1.3.3; *Lex et al., 2014*) upset function.

#### **Orthology analysis**

To test for functional similarities in response to sediment stress across taxa, all species were characterized into orthogroups using OrthoFinder (v2.3.3) with dependencies Diamond (v0.9.22), MCL (v14.137), FastME (v2.1.6.1) and BLAST+ (v2.8.1) and default parameters

Table 1 Summary of read sequencing, quality and alignment by species. See Table S1 for more details.						
Species	Location	# of samples	Avg. raw reads	Avg. clean reads	Avg. # of reads removed	Avg. % of reads mapped
A. cervicornis	Florida	13	16,630,621	16,499,909	130,712	70.62
M. cavernosa	Florida	15	15,921,959	15,740,195	181,765	50.18
O. faveolata	Florida	14	14,111,857	13,905,076	206,781	65.54
M. capitata	Hawaiʻi	11	16,360,670	16,280,212	80,459	63.75
P. acuta	Hawaiʻi	12	14,441,259	14,359,654	81,605	68.94
P. lobata	Hawaiʻi	12	13,995,393	13,858,883	136,510	52.83

(*Emms & Kelly, 2015, 2019*). Orthogroups were filtered to those common in all species. The differentially expressed genes (DEGs) in the orthogroups for each species were identified using the DESeq2 results. Commonality in functional orthogroups containing DEGs were compared between species and visualized using ComplexUpset's (v1.3.3; *Lex et al., 2014*) *upset* function.

# RESULTS

# Read sequencing and quality

Sequencing of 77 samples (n = 11-15 per species; Table 1) yielded a total of 1,173,800,658 raw reads with an average of 15,244,164 ± 3,285,387 raw reads per sample (Table S2). Quality filtering and trimming removed an average of 140,560 ± 119,359 reads per sample, leaving an average of 15,103,605 ± 3,290,800 cleaned reads per sample for analysis (Table S2). Reads were aligned to the species-specific genome, and alignment rates ranged from an average of 50.18% in *M. cavernosa* to an average of 70.62% in *A. cervicornis* (Table 1; Table S2).

# Gene expression

For Florida species, *A. cervicornis* had 215 unique differentially expressed genes (DEGs), *M. cavernosa* had 62 DEGs, and *O. faveolata* had 8 DEGs between ambient and sediment exposure treatments (Table 2; Table S3). Among these, 67 genes in *A. cervicornis*, 19 in *M. cavernosa*, and 0 in *O. faveolata* were upregulated (Log Fold Change, LFC > 0, padj < 0.05), while 154 genes in *A. cervicornis*, 44 in *M. cavernosa*, and 8 in *O. faveolata* were downregulated (LFC < 0, padj < 0.05; Table 2; Table S3). In PCA plots of all genes for Hawaiian species, 'Days' was not visually separated from 'Treatment'; therefore, 'Days' was dropped as an independent variable in further analyses. For Hawaiian coral species, when comparing ambient and sediment exposure treatments, *M. capitata* had 157, *P. acuta* had 263, and *P. lobata* has 153 unique DEGs (Table 2; Table S3). 79 genes in *M. capitata*, 129 in *P. acuta* and 128 in *P. lobata* were downregulated (LFC < 0, padj < 0.05; Table 2; Table S3). The presence in *M. capitata*, 129 in *P. acuta* and 128 in *P. lobata* were upregulated (LFC > 0, padj < 0.05; Table 2; Table S3). The sense in *M. capitata*, 135 in *P. acuta* and 128 in *P. lobata* were downregulated (LFC < 0, padj < 0.05; Table 2; Table S3). Therestingly, there were more downregulated DEGs than upregulated in all species examined, with the exception of *M. capitata*. It is possible that the lower numbers of DEGs in the Florida coral group was due, in part, to a level of acclimation to

Species	Loca- tion	Morpholo- gy	Unique DEGs	Up- regulated DEGs	Down- regulated DEGs	Unique GO terms assigned to DEGs	Orthogroups containing DEGs	DEGs in orthog-roups
A.cervicornis	Florida	Branching	215	67	154	278	102	119
M. cavernosa	Florida	Massive	62	19	44	158	28	31
O. faveolata	Florida	Massive	8	0	8	27	2	2
M. capitata	Hawai'i	Intermediate	157	79	78	237	80	87
P. acuta	Hawaiʻi	Branching	263	129	135	380	100	123
P. lobata	Hawai'i	Massive	153	32	128	198	56	66

 Table 2
 Summary of differentially expressed genes (DEGs), gene ontology (GO) terms and orthogroups by species. See Tables S2–S5 for more details.

the sediment stress, as the Florida experiment was longer (18 days) than the Hawai'i experiment (up to 7 days).

PCA of the DEGs for each species showed that all sediment treatments clustered away from the control on the PC1 axis (PC1 variance explained was 51% for *A. cervicornis*, 46% for *M. cavernosa*, 51% for *O. faveolata*, 79% for *M. capitata*, 64% for *P. acuta*, and 50% for *P. lobata*), supporting a concerted gene expression response to sediment stress across taxa (Figs. 2A–2F). There was slight separation between the mid and high treatments on the PC2 axis for both *P. acuta* and *P. lobata*, indicating differential responses to sediment concentration (PC2 variance explained was 13% for *P. acuta* and 22% for *P. lobata*; Figs. 2E and 2F).

#### Gene ontology

Significantly enriched gene ontology (GO) terms were identified in the DEGs of the species in the present study. In the Florida experiment, corals exposed to sterilized carbonate sediment for 18 days resulted in 278 unique GO terms assigned to *A. cervicornis* DEGs, 158 to *M. cavernosa* DEGs, and 27 to *O. faveolata* DEGs (Table 2; Table S2). *A. cervicornis* shared GO terms related to developmental processes and signal transduction with both *M. cavernosa* and *O. faveolata* (Table 3; Fig. 3; Table S4). Specifically, ovarian cumulus expansion (GO:0001550), positive regulation of skeletal muscle tissue development (GO: 0048643) and regulation of Rho protein signal transduction (GO:0035023) were shared between *A. cervicornis* and *M. cavernosa*, while chondrocyte development (GO:0002063) and positive regulation of signal transduction (GO:0009967) were shared between *A. cervicornis* and *O. faveolata* (Table 3; Fig. 3; Table S4). Despite similar morphologies, *M. cavernosa* and *O. faveolata* did not share any GO terms.

In the Hawaiian coral species that were exposed to unsterilized red terrigenous sediment for up to 7 days, 237, 380, and 198 GO terms were assigned to *M. capitata*, *P. acuta*, and *P. lobata* DEGs, respectively (Table 2; Table S2). Grouping by GO slim term, we found that GO enrichment occurred in functions related to developmental processes, protein metabolism, cell organization and biogenesis, signal transduction, and stress response, among others (Table 3; Fig. 3; Table S4). Only one shared GO term was found in all three





Figure 2 Principal component analysis (PCA) of differentially expressed genes in (A) A. cervicornis, (B) M. cavernosa, (C) O. faveolata,(D) M. capitata, (E) P. acuta, and (F) P. lobata.Full-size 🖾 DOI: 10.7717/peerj.16654/fig-2

of *M. capitata*, *P. acuta*, and *P. lobata*, namely microtubule-based processes (GO:0007017; Table 3; Fig. 3; Table S4). *M. capitata* and *P. acuta*, both of which were branching corals, shared 20 GO terms, primarily relating to developmental processes and signal transduction (Table 3; Fig. 3; Table S4). This was the largest number of GO terms shared between species. *P. acuta* and *P. lobata* shared functional responses related to cell organization and biogenesis, cell-cell signaling, and developmental processes. Developmental processes included five shared GO terms, specifically embryonic axis specifications (GO:000578), embryonic organ development (GO:0014036), and neural plate anterior/posterior regionalization (GO:0021999). *P. acuta* had the most GO terms identified across both experiments (Table 3; Fig. 3; Table S4). Aside from the microtubule-based process, *M. capitata* and *P. lobata* did not share other GO terms.

There was overlap of GO terms between location and morphology (Table 3; Figs. 4 and 5). *M. capitata* was present in all in common three-way interactions; it shared positive regulation of skeletal muscle tissue development (GO:0048643) with *M. cavernosa* and

Species	# of shared GO terms	GO slim category	GO term
A. cervicornis + M. cavernosa + M. capitata	1	Developmental processes	Positive regulation of skeletal muscle tissue development (GO: 0048643)
M. capitata + M. cavernosa + P. acuta	1	Transport	Riboflavin transport (GO:0032218)
M. capitata + P. acuta + P. lobata	1	Other biological processes	Microtubule-based process (GO:0007017)
A. cervicornis +		Cell-cell signaling & transport	Regulation of neurotransmitter uptake (GO:0051580)
M. cavernosa		Developmental processes	Ovarian cumulus expansion (GO:0001550)
		Other metabolic processes	Diadenosine tetraphosphate biosynthetic process (GO:0015966)
		Signal transduction	Regulation of Rho protein signal transduction (GO:0035023)
		No GO slim category identified	Negative regulation of chondrocyte proliferation (GO:1902731)
	11		Positive regulation of cell junction assembly (GO:1901890)
			Negative regulation of histone H4-K16 acetylation (GO:2000619)
			Plasma membrane raft assembly (GO:0044854)
			Cellular response to exogenous dsRNA (GO:0071360)
			Cellular response to platelet-derived growth factor stimulus (GO: 0036120)
			Positive regulation of extrinsic apoptotic signaling pathway <i>via</i> death domain receptors (GO:1902043)
A. cervicornis + O. faveolata	2	Developmental processes	Chondrocyte development (GO:0002063)
		Signal transduction	Positive regulation of signal transduction (GO:0009967)
A. cervicornis + M. capitata	1	No GO slim category identified	positive regulation of adipose tissue development (GO:1904179)
A. cervicornis + P. acuta	5	Other biological processes	Maintenance of gastrointestinal epithelium (GO:0030277)
		Signal transduction	Wnt receptor signaling pathway (GO:0016055)
		Protein metabolism	Mo-molybdopterin cofactor biosynthetic process (GO:0006777)
		Stress response	Cellular response to starvation (GO:0009267)
		No GO slim category identified	Mesenchymal stem cell maintenance involved in nephron morphogenesis (GO:0072038)
A. cervicornis + P. lobata	3	RNA metabolism	Positive regulation of transcription by RNA polymerase II (GO: 0045944)
			DNA methylation-dependent heterochromatin assembly (GO: 0006346)
		Cell organization and biogenesis	DNA methylation-dependent heterochromatin assembly (GO: 0006346)
M. capitata + M. cavernosa	4	Cell organization and biogenesis	Positive regulation of neuron projection development (GO: 0010976)
		Developmental processes	Positive regulation of neuron projection development (GO: 0010976)
			Retina development in camera-type eye (GO:0060041)
		Other biological processes	Negative regulation of mitochondrial membrane potential (GO: 0010917)
		No GO slim category identified	Ventricular compact myocardium morphogenesis (GO:0003223)
			(Continued)

Table 3 Shared Biological Processes gene ontology (GO) terms between species. See Tables S2 & S3 for more details.

Table 3 (continued)				
Species	# of shared GO terms	GO slim category	GO term	
M. capitata + P. acuta	20	Cell adhesion	Heterophilic cell adhesion (GO:0007157)	
		Developmental processes	Embryonic hindlimb morphogenesis (GO:0035116)	
			Positive regulation of osteoclast differentiation (GO:0045672)	
		Other biological processes	Positive regulation of bone resorption (GO:0045780)	
			Positive regulation of cellular pH reduction (GO:0032849)	
			Oxidation reduction (GO:0055114)	
			Response to activity (GO:0014823)	
			Response to glucagon stimulus (GO:0033762)	
			Response to pH (GO:0009268)	
		RNA metabolism	Positive regulation of transcription factor activity (GO:0051091)	
		Signal transduction	Blue light signaling pathway (GO:0009785)	
		Stress response	Response to pain (GO:0048265)	
		Transport	Carbon dioxide transport (GO:0015670)	
			Secretion (GO:0046903)	
		No GO slim category identified	Angiotensin-activated signaling pathway (GO:0038166)	
			Cellular response to retinoic acid (GO:0071300)	
			Positive regulation of dipeptide transmembrane transport (GO: 2001150)	
			Positive regulation of mitochondrial membrane permeability (GO: 0035794)	
			Negative regulation of glucocorticoid secretion (GO:2000850)	
			Regulation of chloride transport (GO:2001225)	
M. cavernosa + P. acuta	2	Developmental processes	Ectoderm development (GO:0007398)	
		No GO slim category identified	Anterior head development (GO:0097065)	
M. cavernosa + P. lobata	3	Developmental processes	Somatic muscle development (GO:0007525)	
		No GO slim category identified	Regulation of cell junction assembly (GO:1901888)	
			Regulation of protein serine/threonine kinase activity (GO: 0071900)	
O. faveolata + P. acuta	1	Other biological processes	Response to stimulus (GO:0050896)	
P. acuta + P. lobata	11	Cell-cell signaling	SPEMANN organizer formation (GO:0060061)	
		Cell organization and biogenesis	Neural crest cell fate specification (GO:0014036)	
		Developmental processes	Embryonic axis specification (GO:0000578)	
			Embryonic organ development (GO:0048568)	
			Eye development (GO:0001654)	
			Neural crest cell fate specification (GO:0014036)	
			Neural plate anterior/posterior regionalization (GO:0021999)	
		Other biological processes	Regulation of circadian rhythm (GO:0042752)	
		No GO slim category identified	Canonical Wnt signaling pathway involved in neural crest cell differentiation (GO:0044335)	
			Cellular hypotonic response (GO:0071476)	
			Endocardial cushion development (GO:0003197)	
			Negative regulation of cardiac cell fate specification (GO:2000044)	

Table 3 (continued)			
Species	# of shared GO terms	GO slim category	GO term
A. cervicornis	256	See Tables S2 and S3	
M. capitata	209	See Tables S2 and S3	
M. cavernosa	136	See Tables S2 and S3	
O. faveolata	24	See Tables S2 and S3	
P. acuta	339	See Tables S2 and S3	
P. lobata	181	See Tables S2 and S3	





*A. cervicornis*, riboflavin transport (GO:0032218) with *M. cavernosa* and *P. acuta*, and microtubule-based process (GO:0007017) with *P. acuta* and *P. lobata*. *M. capitata* and *M. cavernosa* also shared three GO terms related to developmental processes (positive regulation of neuron projection development (GO:0010976) and retina development in camera-type eye (GO:0060041; Table 3; Table S4). Among others, *M. cavernosa* from the FL experiment, shared GO terms relating to ectoderm development (GO:0007398) and regulation of protein serine/threonine kinase activity (GO:0071900) with *P. acuta* and *P. lobata*, respectively. *A. cervicornis* and *P. acuta*, both branching corals, shared 11 GO terms, including Mo-molybdopterin cofactor biosynthetic process (GO:0006777) and Wnt



**Figure 4** Counts of biological process GO terms grouped under GO slim categories by location. GO slim categories are on the x-axis, while the number of Biological Processes GO terms in each GO slim category is on the y-axis. The bars are colored by location: green bar = Florida, orange bar = Hawai'i. Full-size DOI: 10.7717/peerj.16654/fig-4

receptor signaling pathway (GO:0016055). With the exception of *O. faveolata*, the massive corals, *M. cavernosa* and *P. lobata*, shared GO terms relating to regulation of cell junction assembly (GO:1901888), regulation of protein serine/threonine kinase activity (GO: 0071900), and somatic muscle development (GO:0007525). In total, more significantly enriched GO terms were identified at HI and in the branching corals (Figs. 4 and 5).

#### Orthogroups

In total, 21,688 total orthogroups were identified and 9,216 were common to all species. Amongst the Florida species, common orthogroups contained 119 DEGs from *A. cervicornis*, 31 from *M. cavernosa*, and two from *O. faveolata* (Table 2; Table S5). In the Hawai'i species, common orthogroups included 87 DEGs from *M. capitata*, 123 from *P. acuta*, and 66 from *P. lobata* (Table 2; Table S5).

Similarly to GO terms, there was overlap in orthogroups between species. There was one 4-way interaction, in which *A. cervicornis*, *M. capitata*, *P. acuta*, and *P. lobata* shared two orthogroups (Fig. 6; Table S5). *A. cervicornis* and *P. acuta* shared 3 orthogroups, the most orthogroups shared between a set of species (Fig. 6; Table S6). Although there may not



**Figure 5** Counts of biological process GO terms grouped under GO slim categories by morphology. GO slim categories are on the x-axis, while the number of Biological Processes GO terms in each GO slim category is on the y-axis. The bars are colored by morphology: blue bar = branching, gray bar = intermediate, yellow bar = massive. Full-size DOI: 10.7717/peerj.16654/fig-5

have been a high number of overlapping orthogroups, the function of those overlapping orthogroups were similar between species. For example, one orthogroup (OG0000487) contained DEGs related to microtubule-based processes and structural constituent of cytoskeleton in *M. capitata* and *P. acuta* (Table S6).

# DISCUSSION

In this study, we conducted two separate experiments to characterize the molecular underpinnings of corals with differing morphological characteristics responding to sediment stressors in two locations, Florida and Hawai'i. The methodological differences prevent us from making direct statistical comparisons between the experiments. However, it is still relevant to highlight general biological processes and mechanisms related to sediment stress responses across morphology and location.

# Responses to unsterilized red sediment in Hawai'i

It is well established that morphology can play a role in modulating a coral's response to sediment stress, but it is unknown if gene expression varies in corals with differing morphological traits. Our study found that, across morphologies in Hawaiian corals,



Figure 6UpSet plot of the intersection of orthogroups across location (green bar or circle = Florida; orange bar or circle = Hawai'i) andmorphology (blue stripe = branching; gray stripe = intermediate; yellow stripe = massive).Full-size  $\square$  DOI: 10.7717/peerj.16654/fig-6

developmental processes were primarily affected by unsterilized sediment. Previous work has demonstrated that sediment deposition, high turbidity, and low light have been shown to adversely affect development and reproductive output across morphologies (Kojis & Quinn, 1984; Gilmour, 1999; Humphrey et al., 2008), though M. capitata development and fecundity are not always negatively affected by high sedimentation rates (Padilla-Gamiño et al., 2014). In the present study, the branching coral, P. acuta, shared a relatively high number of overlapping developmental processes-related responses with corals with intermediate (M. capitata; 20 shared terms) and massive (P. lobata; 11 shared terms) morphology. Shared responses corresponded to reproduction and developmental processes like embryonic axis specifications, embryonic organ development, eye development, neural crest cell fate specification, neural plate anterior/posterior regionalization, embryonic hindlimb morphogenesis and others. Unexpectedly, these terms primarily correspond to vertebrate developmental processes, complicating our ability to interpret these results. Given our use of specific annotation databases (i.e., NCBI, SwissProt, InterPro), it is possible that these databases are all dominated by vertebrate-related annotations and thus, invertebrate protein sequences are assigned vertebrate-centric annotations. While it is clear that developmental processes are affected by short term sediment stress across morphologies, it is unclear how vertebrate specific gene functions may translate to developmental processes in basal metazoans. More work is necessary to evaluate equivalency across annotation softwares.

During the year, *M. capitata, P. acuta*, and *P. lobata* develop their gametes and then release when conditions are optimal, typically June and July in Hawai'i (*Stimson, 1978*; *Richmond & Jokiel, 1984*; *Padilla-Gamiño & Gates, 2012*; *Brown et al., 2020*). Because the corals in our study were exposed during this time period, it is possible that we captured gene expression signatures unique to corals at or near the peak of their reproductive phenotype. Although the corals were not sampled at different times of year under the same experimental conditions, these results suggest that because the corals were sampled during their reproductive period, signals of developmental processes may be higher than they might have been at other points in the year. Given the high number of shared and unique developmental process GO terms in the Hawai'i corals, however, it is likely that sediment has the potential to have reproductive and developmental consequences, which can have subsequent impacts on population growth and dynamics.

The only specific shared response across the three Hawaiian species was the downregulation of microtubule-based processes. Microtubules, tubulin polymers that maintain structure and shape to eukaryotic cells, are major components of cilia, which coral utilize for activities such as feeding and clearing sediment (Westbroek, Yanagida & Isa, 1980; Rogers, 1990; Erftemeijer et al., 2012). Downregulation of processes relating to cilia biogenesis/degradation and motility was also observed in the P. acuta host and its endosymbionts in response to combined acute heat and sediment stress (*Poquita-Du et al.*, 2019, 2020). The authors of these studies hypothesized that the corals were likely diverting energy resources away from feeding and active sediment clearing in order to prioritize energy for basic homeostasis. While our experiment did not include heat stress, it is probable that downregulation of microtubule or cilia-related processes is a generally conserved response to sediment stress. Exhaustion of sediment-clearing activity of corals and eventual loss/breakdown of cilia cells can also be a consequence of continued exposure to high levels of sediment stress (Stafford-Smith & Ormond, 1992; Stafford-Smith, 1993; *Erftemeijer et al.*, 2012). Given the acute nature of the stress in the Hawai'i experiment (*i.e.*, sediment was added at day 0 and day 4), the corals here may have exhausted their ciliary action abilities, thus leading to downregulation of microtubule-based processes.

#### Responses to sterilized white sediment in Florida

No specific responses were shared among *A. cervicornis*, *M. cavernosa*, and *O. faveolata*, despite being exposed to the same sediment for the same amount of time. Given the length of time of this experiment, it is possible that we only captured the longer-term stress responses of these corals and that their responses may have been more similar at the beginning of the exposures. Morphology also may have contributed to the divergent responses between these species. In previous sediment stress studies, morphology contributed to physiological response variability (*Stafford-Smith*, *1993*; *Fabricius*, *2005*; *Erftemeijer et al.*, *2012*). For example, *Rogers* (*1979*) evaluated the effect of shading (as a proxy for turbidity) for 5 weeks on several coral species from San Cristobal Reef, Puerto Rico. The branching coral, *A. cervicornis*, had entirely bleached, while the massive coral, *M. cavernosa*, was visibly unaffected and appeared to have little response. On the other hand, *M. annularis* had substantial bleaching after 5 weeks, despite being a close relative of

*M. cavernosa* (*Rogers, 1990*). Different branching coral species exhibit a wide range of sediment tolerances; after 12 weeks of exposure, the lowest sediment treatments that caused full colony mortality were 30 mg/L<sup>-1</sup> for *M. aequituberculata* and 100 mg/L<sup>-1</sup> for *A. millepora* (*Flores et al., 2012*). *Stafford-Smith* (1993) examined sediment rejection efficiency in 22 species of Australian corals from a range of morphologies, finding that there was a wide range of active-rejection efficiencies between species. For instance, *Gardineroseries planulata* is a competent rejector of a variety of sediment sizes, but only for a short period of time. *Favia stelligera* and *Leptoria phrygia* had moderate clearance rates, but tissue mortality occurred within one to two days. *M. aequituberculata* and *Porites* spp. had low rejection efficiency and had bleached tissues, but no tissue mortality for up to 8 days. Morphology was a driving factor in those results, as branching corals had faster clearing rates than massive corals (*Stafford-Smith, 1993*). Thus, it is likely that morphology played a role in driving differences in long-term response in the Floridian corals, highlighting the challenge of predicting responses to sedimentation across species.

Despite differing morphologies, the branching coral, A. cervicornis, shared responses related to developmental processes and signal transduction with both massive corals, M. cavernosa and O. faveolata, independently. Interestingly, no specific responses were shared between M. cavernosa and O. faveolata, despite similar morphologies. Downregulation of Rho protein signal transduction was observed in A. cervicornis and M. cavernosa (with the exception of upregulation in the M. cavernosa T2vT3 treatment comparison). Rho proteins are part of a superfamily of signaling GTPase proteins, which typically control the assembly and organization of the cytoskeleton, as well as participate in functions such as cell adhesion, contraction, migration, morphogenesis, and phagocytosis (Mackay & Hall, 1998; Moon & Zheng, 2003; Phuyal & Farhan, 2019). In corals, Rho GTPases participate in cytoskeleton remodeling during phagocytosis, as well as cell division of endosymbionts within symbiotic gastrodermal cells (*Li et al., 2014*). Rho GTPase pathways have been identified in coral polyp bailout responses to heat stress and hyperosmosis, as well as in bleached corals in response to low flow environments (Chuang & Mitarai, 2020; Fifer et al., 2021; Gösser et al., 2021). In this study, downregulation of Rho protein pathways suggests that minimal cytoskeleton maintenance, assembly and organization is occurring and that the corals may not be able to properly maintain their cellular structures under sedimentation.

A. cervicornis and O. faveolata both downregulated chondrocyte development, a developmental response. Chondrocytes are cells in cartilage that make up the cytoskeletal matrix in humans and other animals, including some marine invertebrates (*Philpott & Person, 1970; Cowden & Fitzharris, 1975; Libbin et al., 1976; Archer & Francis-West, 2003; Kamisan et al., 2013*). In corals, the cytoskeleton matrix is made up of calcium carbonate, as opposed to cartilage, which forms through rapid accretion of protein rich skeletal organic matrix and extracellular calcium carbonate crystals to form a stony skeleton (*Vandermeulen & Watabe, 1973; Akiva et al., 2018*). Skeletal matrix formation begins when planktonic coral larvae settle and begin to secrete calcium carbonate, which helps to anchor the coral to the substrate (*Akiva et al., 2018*). The skeleton grows as the coral animal continues to secrete calcium carbonate, building up a large and intricate 3D skeletal

structure (*Tambutté et al., 2011*). Sedimentation can hinder coral skeletal growth by depositing sediment on the tissue and diverting energy away from growth, as well as decreasing the amount of light that reaches the coral, thus affecting the possibility for light-enhanced calcification, which is responsible for most of the skeletal growth in corals (*Dodge, Aller & Thomson, 1974; Erftemeijer et al., 2012*). Downregulation of chondrocyte development may be related to the disruption of skeletal matrix formation, which may influence skeletal density and growth. Decreased skeletal density was observed in corals from inshore reefs which experience higher levels of sedimentation as compared to corals from offshore reefs (*Lough & Barnes, 1992*). Therefore, the downregulation of genes relating to chondrocyte development across morphologies, suggests that corals with a range of morphological characteristics may have decreased skeletal growth in response to sediment stress.

#### Differences in response based on morphological characteristics

This study combined data from two independent experiments in order to characterize the molecular mechanisms that corals use to respond to different sedimentation stressors. The experiments used different sediment types (unsterilized red clay sediment in Hawai'i, sterilized carbonate sand sediment in Florida) and lengths of exposure (up to 7 days in Hawai'i, 18 days in Florida). Differences in experimental methodology prevent us from making direct statistical comparisons between the two experiments; however, the shared morphologies between experiments enables us to identify broad generalizations about conserved gene regulation in response to sediment stress. These comparisons are relevant, as knowledge of what genes and biological processes are broadly affected by sediment stress can help coral reef management.

We did not identify a generalized response across morphology nor gene expression patterns across taxa. There were two groups of three species (M. capitata, M. cavernosa, P. acuta and A. cervicornis, M. capitata, M. cavernosa) that shared specific responses. However, no group contained a single morphology and gene expression patterns varied among species. For example, M. capitata (intermediate), M. cavernosa (massive), and *P. acuta* (branching) all expressed genes relating to riboflavin transport, the transport of certain vitamins in cells, though they had very different expression patterns. Riboflavin transport genes were upregulated in P. acuta, but downregulated in M. cavernosa; *M. capitata*, on the other hand, differentially expressed two genes relating to riboflavin transport, one of which was upregulated and the other downregulated. Thus, even though a shared response was identified, the direction of the response varied greatly. This result demonstrates that responses may be shared across morphologies, locations and sediment types, but it may be difficult to predict the directionality of the response. This response may also be due a level of acclimation attained by M. cavernosa during the longer exposure in the Florida experiment. The other group, which contained A. cervicornis (branching), M. capitata (intermediate), and M. cavernosa (massive), all downregulated genes relating to positive regulation of skeletal muscle tissue development. This term refers to the activation, maintenance, or increase of the rate of skeletal muscle tissue development (Buckingham et al., 2003; Grefte et al., 2007). Given that this term is downregulated, it

means that there is little to no activation, maintenance, or increase of the rate of muscle tissue development in these corals. These gene expression patterns highlight the complexity of characterizing responses to different kinds of sedimentation stress in species with different morphotypes.

Some terms were shared across both morphology and location, indicating a generalized sediment stress response. For example, DNA methylation-dependent heterochromatin assembly was shared between A. cervicornis (branching) and P. lobata (massive). Opposite expression patterns were again observed, in which upregulation occurred in A. cervicornis and downregulation in P. lobata. DNA methylation-dependent heterochromatin assembly refers to repression of transcription by DNA methylation leading to the formation of heterochromatin (Jones & Wolffe, 1999; Grewal & Moazed, 2003). In the case of A. cervicornis, upregulation suggests that repression of transcription by DNA methylation and subsequent heterochromatin formation is occurring. Therefore, certain portions of DNA cannot be accessed, giving A. cervicornis more stringent control on gene expression. On the other hand, the downregulation of these genes in *P. lobata* means less repression of transcription by DNA methylation occurring and heterochromatin is not being fully formed, leaving much of the DNA accessible to transcription machinery and ultimately, reducing the amount of control that *P. lobata* has on gene expression. To date, no work has examined how sediment stress affects epigenetic mechanisms, such as DNA methylation, in coral. However, previous studies have found changes to DNA methylation in response to stress and environmental change (Putnam, Davidson & Gates, 2016; Liew et al., 2018; Dimond & Roberts, 2020; Rodríguez-Casariego, Mercado-Molina & Garcia-Souto, 2020). Additionally, it has been observed in corals that genes with weak methylation signals are more likely to demonstrate differential gene expression (Dixon, Bay & Matz, 2014; Entrambasaguas et al., 2021). Epigenetic modifications and their regulation of gene transcription are highly species and context dependent. Furthermore, the directionality of epigenetic regulation on gene expression or repression can vary depending on the underlying genetic machinery and the environment. This is exemplified in the two species with overlapping response terms. Namely, A. cervicornis exhibits a more regulated control on gene expression in contrast to *P. lobata*, which exhibits a less regulated profile of gene regulation. Thus, it is likely that sedimentation stress from each location impacted DNA methylation and heterochromatin in different ways, causing opposing expression patterns.

More general responses were shared over morphology and location, as identified by the orthogroup analysis. The orthogroup analysis grouped homologous gene sequences in different species related to one another by linear descent. The resulting 'orthogroup' represents a group of similar gene sequences across multiple species (*Emms & Kelly, 2015, 2019*). Orthogroups were shared across morphology and location, though in relatively low numbers (sharing between one and three orthogroups). This sharing may represent a group of orthogroups that form a core group of genes in response to sediment stress. Although we cannot directly compare the genes or orthogroups between experiments, the shared orthogroups represent potential overlap in sedimentation response over location and morphology.

The branching corals, A. cervicornis from the Florida experiment and P. acuta from the Hawai'i experiment, shared a metabolic response, 'Mo-molybdopterin cofactor biosynthetic process', which describes the creation of the Mo-molybdopterin cofactor, an essential component for catalytic activity of certain enzymes (Kisker, Schindelin & Rees, 1997; Mendel, 2013). Molybdenum (Mo) is a trace metal synthesized de novo through GTP-based processes; cyclic pyranopterun monophosphate (cPMP) is initially formed, which is then converted to the molybdopterin cofactor (Mendel, 2013). This essential cofactor catalyzes the oxidation and reduction of molecules in enzymatic processes regulating nitrogen, sulfur, and carbon (Daniels et al., 2008; Iobbi-Nivol & Leimkühler, 2013). Similar to other results in the present study, both species demonstrated opposing differential gene expression for this term. Mo-molybdopterin cofactor biosynthetic process was downregulated in A. cervicornis, but upregulated in P. acuta, suggesting that while different sediment types and exposure durations can induce similar differentially expressed genes, it can produce different expression patterns for those genes. Molybdopterin are co-factors for oxidoreductases, a family of enzymes that catalyze the transfer of electrons between molecules (Kisker, Schindelin & Rees, 1997). Upregulation of molybdopterin synthesis may suggest that these types of enzymes are more metabolically active. Stressful conditions have made these kinds of enzymes more active in plants and corals (Bouchard & Yamasaki, 2008; Zdunek-Zastocka & Sobczak, 2013). Upregulation of metabolism related genes was also observed in a study that examined transcriptomic responses of corals in response to two different sediment experiments (Bollati et al., 2021). Downregulation of Mo-molybdopterin cofactor synthesis may indicate that the coral does not have enough energy to synthesize molybdopterin which in turn makes it so the activity of these specific enzymes is decreased or stopped altogether, suggesting a decrease in metabolism for A. cervicornis. It is also possible that the unsterilized sediment was providing molybdopterin to P. acuta, making it necessary for P. acuta to upregulate genes relating to molybdopterin processing to manage the influx (Fujimoto & Sherman, 1951; Siebert et al., 2015). Because the sediment was sterilized in the Florida experiment, no molybdenum would be present in the sediment, indicating that molybdenum-related enzymes may not have been functioning at a high level, leading to downregulation.

# CONCLUSION

This study incorporated data from two separate experiments to more fully characterize the molecular mechanisms induced by sedimentation in corals. We found that developmental processes are primarily impacted in branching corals across location, highlighting potential future research avenues with regards to sediment stress and reproductive potential and output. While few specific genes were shared across morphology and location, orthogroup analysis uncovered potential overlap in generalized sediment stress responses. Direct comparisons across species are necessary to further elucidate the genetic basis of coral susceptibility to sediment stress.

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# ADDITIONAL INFORMATION AND DECLARATIONS

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# **Competing Interests**

The authors declare that they have no competing interests.

# **Author Contributions**

- Jill Ashey performed the experiments, analyzed the data, prepared figures and/or tables, authored or reviewed drafts of the article, funding acquisition, and approved the final draft.
- Hailey McKelvie analyzed the data, authored or reviewed drafts of the article, data validation, and approved the final draft.
- John Freeman analyzed the data, authored or reviewed drafts of the article, data validation, and approved the final draft.
- Polina Shpilker analyzed the data, authored or reviewed drafts of the article, data validation, and approved the final draft.
- Lauren H. Zane analyzed the data, authored or reviewed drafts of the article, and approved the final draft.
- Danielle M. Becker analyzed the data, authored or reviewed drafts of the article, and approved the final draft.
- Lenore Cowen analyzed the data, authored or reviewed drafts of the article, resources, funding acquisition, and approved the final draft.
- Robert H. Richmond conceived and designed the experiments, performed the experiments, authored or reviewed drafts of the article, resources, funding acquisition, and approved the final draft.

- Valerie J. Paul conceived and designed the experiments, performed the experiments, authored or reviewed drafts of the article, resources, funding acquisition, and approved the final draft.
- Francois O. Seneca conceived and designed the experiments, performed the experiments, authored or reviewed drafts of the article, resources, funding acquisition, and approved the final draft.
- Hollie M. Putnam analyzed the data, authored or reviewed drafts of the article, resources, funding acquisition, and approved the final draft.

#### **Field Study Permissions**

The following information was supplied relating to field study approvals (*i.e.*, approving body and any reference numbers):

Hawaiian corals were collected in Kāne'ohe Bay, O'ahu, Hawai'i under Hawai'i SAP permit (SAP 2015-48). Floridian corals were collected from the Key West nursery of the Florida Keys National Marine Sanctuary under NOAA National Marine Sanctuaries Permit #FKNMS-2015-016 and #FKNMS-2016-017.

#### **DNA** Deposition

The following information was supplied regarding the deposition of DNA sequences: The raw sequences are available at NCBI BioProject: PRJNA911752.

### **Data Availability**

The following information was supplied regarding data availability:

The code for analyses is available at GitHub and Zenodo:

- https://github.com/JillAshey/SedimentStress.

- Ashey, J. (2023). SedimentStress [Data set]. In PeerJ. Zenodo. https://doi.org/10.5281/ zenodo.10183389.

- https://github.com/JillAshey/FunctionalAnnotation.

- Ashey, J. (2023). Functional annotation [Data set]. In PeerJ. Zenodo. https://doi.org/ 10.5281/zenodo.10211582.

The archived references used in this study are available at OSF: Ashey, Jill, and Hollie Putnam. 2022. "Sediment Stress." OSF. December 27. https://doi.org/10.17605/OSF.IO/ 8QN6C.

The raw sequences are available at NCBI BioProject: PRJNA911752.

#### **Supplemental Information**

Supplemental information for this article can be found online at http://dx.doi.org/10.7717/ peerj.16654#supplemental-information.

#### REFERENCES

Abaya LM, Wiegner TN, Beets JP, Colbert SL, Carlson KM, Kramer KL. 2018. Spatial distribution of sewage pollution on a Hawaiian coral reef. *Marine Pollution Bulletin* 130(May):335–347 DOI 10.1016/j.marpolbul.2018.03.028.

- Akiva A, Neder M, Kahil K, Gavriel R, Pinkas I, Goobes G, Mass T. 2018. Minerals in the pre-settled coral *Stylophora pistillata* crystallize via protein and ion changes. *Nature Communications* 9(1):1880 DOI 10.1038/s41467-018-04285-7.
- Alderdice R, Perna G, Cárdenas A, Hume BCC, Wolf M, Kühl M, Pernice M, Suggett DJ, Voolstra CR. 2022. Deoxygenation lowers the thermal threshold of coral bleaching. *Scientific Reports* 12(1):18273 DOI 10.1038/s41598-022-22604-3.
- Altschul SF, Gish W, Miller W, Myers EW, Lipman DJ. 1990. Basic local alignment search tool. *Journal of Molecular Biology* 215(3):403–410 DOI 10.1016/S0022-2836(05)80360-2.
- Anders S, Huber W. 2010. Differential expression analysis for sequence count data. *Genome Biology* 11(10):R106 DOI 10.1186/gb-2010-11-10-r106.
- Andrews S. 2010. A quality control tool for high throughput sequence data. Available at http:// www.bioinformatics.babraham.ac.uk/projects/fastqc.
- Anthony KRN, Fabricius KE. 2000. Shifting roles of heterotrophy and autotrophy in coral energetics under varying turbidity. *Journal of Experimental Marine Biology and Ecology* 252(2):221–253 DOI 10.1016/S0022-0981(00)00237-9.
- Archer CW, Francis-West P. 2003. The chondrocyte. *The International Journal of Biochemistry* & *Cell Biology* 35(4):401-404 DOI 10.1016/S1357-2725(02)00301-1.
- Ashey J, McKelvie H, Freeman JD, Shpilker P, Zane LH, Becker-Polinski DM, Cowen L, Richmond RH, Paul V, Seneca F, Putnam HM. 2023. Characterizing transcriptomic responses to sediment stress across location and morphology in reef-building corals. BioRxiv preprint DOI 10.1101/2023.01.30.526279.
- Bahr KD, Rodgers KS, Jokiel PL, Prouty NG, Storlazzi CD. 2020. Pulse sediment event does not impact the metabolism of a mixed coral reef community. *Ocean and Coastal Management* 184(3):105007 DOI 10.1016/j.ocecoaman.2019.105007.
- **Bairoch A, Apweiler R. 1997.** The SWISS-PROT protein sequence data bank and its supplement TrEMBL. *Nucleic Acids Research* **25(1)**:31–36 DOI 10.1093/nar/25.1.31.
- Bak RPM. 1978. Lethal and sublethal effects of dredging on reef corals. *Marine Pollution Bulletin* 9(1):14–16 DOI 10.1016/0025-326X(78)90275-8.
- Barnes BB, Hu C, Kovach C, Silverstein RN. 2015. Sediment plumes induced by the port of miami dredging: analysis and interpretation using landsat and MODIS data. *Remote Sensing of Environment* 170(18):328–339 DOI 10.1016/j.rse.2015.09.023.
- Bessell-Browne P, Negri AP, Fisher R, Clode PL, Duckworth A, Jones R. 2017. Impacts of turbidity on corals: the relative importance of light limitation and suspended sediments. *Marine Pollution Bulletin* 117(1):161–170 DOI 10.1016/j.marpolbul.2017.01.050.
- Bolger AM, Lohse M, Usadel B. 2014. Trimmomatic: a flexible trimmer for illumina sequence data. *Bioinformatics* 30(15):2114–2120 DOI 10.1093/bioinformatics/btu170.
- Bollati E, Rosenberg Y, Simon-Blecher N, Tamir R, Levy O, Huang D. 2021. Untangling the molecular basis of coral response to sedimentation. *Molecular Ecology* 31(3):884–901 DOI 10.1111/mec.16263.
- **Bouchard JN, Yamasaki H. 2008.** Heat stress stimulates nitric oxide production in symbiodinium *Microadriaticum*: a possible linkage between nitric oxide and the coral bleaching phenomenon. *Plant & Cell Physiology* **49(4)**:641–652 DOI 10.1093/pcp/pcn037.
- Brown NP, Forsman ZH, Tisthammer KT, Richmond RH. 2020. A resilient brooding coral in the broadcast spawning *Porites lobata* species complex: a new endemic, introduced species, mutant, or new adaptive potential? *Coral Reefs* **39**(3):809–818 DOI 10.1007/s00338-020-01922-w.

- Browne NK, Precht E, Last KS, Todd PA. 2014. Photo-physiological costs associated with acute sediment stress events in three near-shore turbid water corals. *Marine Ecology Progress Series* 502:129–143 DOI 10.3354/meps10714.
- Buckingham M, Bajard L, Chang T, Daubas P, Hadchouel J, Meilhac S, Montarras D, Rocancourt D, Relaix F. 2003. The formation of skeletal muscle: from somite to limb. *Journal of Anatomy* 202(1):59–68 DOI 10.1046/j.1469-7580.2003.00139.x.
- Chuang P, Mitarai S. 2020. Signaling pathways in the coral polyp bail-out response. *Coral Reefs* 39(6):1535–1548 DOI 10.1007/s00338-020-01983-x.
- Cortés JN, Risk MJ. 1985. A reef under siltation stress: Cahuita, Costa Rica. *Bulletin of Marine Science* 36(2):339–356.
- Costanza R, Groot R, Sutton P, Ploeg S, Anderson SJ, Kubiszewski I, Farber S, Turner RK. 2014. Changes in the global value of ecosystem services. *Global Environmental Change: Human and Policy Dimensions* 26(May):152–158 DOI 10.1016/j.gloenvcha.2014.04.002.
- **Cowden RR, Fitzharris T. 1975.** The *Histochemistry* and structure of tentacle cartilage tissues in the marine polychaete, *Sabella melanostigma*. *Histochemie* **43(1)**:1–10 DOI 10.1007/BF00490150.
- **Cunning R, Silverstein RN, Barnes BB, Baker AC. 2019.** Extensive coral mortality and critical habitat loss following dredging and their association with remotely-sensed sediment plumes. *Marine Pollution Bulletin* **145(August)**:185–199 DOI 10.1016/j.marpolbul.2019.05.027.
- **Dallmeyer DG, Porter JW, Smith GJ. 1982.** Effects of particulate peat on the behavior and physiology of the jamaican reef-building coral *Montastrea annularis*. *Marine Biology* **68(3)**:229–233 DOI 10.1007/BF00409589.
- **Daniels JN, Wuebbens MM, Rajagopalan KV, Schindelin H. 2008.** Crystal structure of a molybdopterin synthase–precursor Z complex: insight into its sulfur transfer mechanism and its role in molybdenum cofactor deficiency. *Biochemistry* **47**(2):615–626 DOI 10.1021/bi701734g.
- Davy SK, Allemand D, Weis VM. 2012. Cell biology of cnidarian-dinoflagellate symbiosis. Microbiology and Molecular Biology Reviews: MMBR 76(2):229–261 DOI 10.1128/MMBR.05014-11.
- **DeSalvo MK, Sunagawa S, Voolstra CR, Medina M. 2010.** Transcriptomic responses to heat stress and bleaching in the elkhorn coral *Acropora palmata*. *Marine Ecology Progress Series* **402**:97–113 DOI 10.3354/meps08372.
- DeSalvo MK, Voolstra CR, Sunagawa S, Schwarz JA, Stillman JH, Coffroth MA, Szmant AM, Medina M. 2008. Differential gene expression during thermal stress and bleaching in the caribbean coral *Montastraea faveolata*. *Molecular Ecology* 17(17):3952–3971 DOI 10.1111/j.1365-294X.2008.03879.x.
- **Dimond JL, Roberts SB. 2020.** Convergence of DNA methylation profiles of the reef coral *Porites astreoides* in a novel environment. *Frontiers in Marine Science* **6**:792 DOI 10.3389/fmars.2019.00792.
- **Dixon GB, Bay LK, Matz MV. 2014.** Bimodal signatures of germline methylation are linked with gene expression plasticity in the coral *Acropora millepora*. *BMC Genomics* **15**:1109 DOI 10.1186/1471-2164-15-1109.
- Dobin A, Davis CA, Schlesinger F, Drenkow J, Zaleski C, Jha S, Batut P, Chaisson M, Gingeras TR. 2013. STAR: ultrafast universal RNA-Seq aligner. *Bioinformatics* 29(1):15–21 DOI 10.1093/bioinformatics/bts635.
- Dodge RE, Aller RC, Thomson J. 1974. Coral growth related to resuspension of bottom sediments. *Nature* 247(5442):574–577 DOI 10.1038/247574a0.

- Emms DM, Kelly S. 2015. OrthoFinder: solving fundamental biases in whole genome comparisons dramatically improves orthogroup inference accuracy. *Genome Biology* 16:157 DOI 10.1186/s13059-015-0721-2.
- Emms DM, Kelly S. 2019. OrthoFinder: phylogenetic orthology inference for comparative genomics. *Genome Biology* 20(1):238 DOI 10.1186/s13059-019-1832-y.
- Entrambasaguas L, Ruocco M, Verhoeven KJF, Procaccini G, Marín-Guirao L. 2021. Gene body DNA methylation in seagrasses: inter- and intraspecific differences and interaction with transcriptome plasticity under heat stress. *Scientific Reports* 11:14343 DOI 10.1038/s41598-021-93606-w.
- Erftemeijer PLA, Riegl A, Hoeksema BW, Todd PA. 2012. Environmental impacts of dredging and other sediment disturbances on corals: a review. *Marine Pollution Bulletin* **64(9)**:1737–1765 DOI 10.1016/j.marpolbul.2012.05.008.
- Ewels P, Magnusson M, Lundin S, Käller M. 2016. MultiQC: summarize analysis results for multiple tools and samples in a single report. *Bioinformatics* 32(19):3047–3048 DOI 10.1093/bioinformatics/btw354.
- **Fabricius KE. 2005.** Effects of terrestrial runoff on the ecology of corals and coral reefs: review and synthesis. *Marine Pollution Bulletin* **50(2)**:125–146 DOI 10.1016/j.marpolbul.2004.11.028.
- Fifer J, Bentlage B, Lemer S, Fujimura AG, Sweet M, Raymundo LJ. 2021. Going with the flow: how corals in high-flow environments can beat the heat. *Molecular Ecology* **30**(9):2009–2024 DOI 10.1111/mec.15869.
- Flores F, Hoogenboom MO, Smith LD, Cooper TF, Abrego D, Negri AP. 2012. Chronic exposure of corals to fine sediments: lethal and sub-lethal impacts. *PLOS ONE* 7(5):e37795 DOI 10.1371/journal.pone.0037795.
- **Fujimoto G, Sherman GD. 1951.** Molybdenum content of typical soils and plants of the Hawaiian Islands. *Agronomy Journal* **43(9)**:424–429 DOI 10.2134/agronj1951.00021962004300090003x.
- Gentleman R, Carey VJ, Huber W, Hahne F. 2021. Genefilter: genefilter: methods for filtering genes from high-throughput experiments (version 1.70.0). R package. *Available at https://bioconductor.org/packages/release/bioc/html/genefilter.html*.
- Gilmour J. 1999. Experimental investigation into the effects of suspended sediment on fertilisation, larval survival and settlement in a scleractinian coral. *Marine Biology* 135:451–462 DOI 10.1007/s002270050645.
- Gösser F, Raulf A, Mosig A, Tollrian R, Schweinsberg M. 2021. Signaling pathways of heat- and hypersalinity-induced polyp bailout in *Pocillopora acuta*. *Coral Reefs* **40(6)**:1713–1728 DOI 10.1007/s00338-021-02191-x.
- Götz S, García-Gómez JM, Terol J, Williams TD, Nagaraj SH, Nueda MJ, Robles M, Talón M, Dopazo J, Conesa A. 2008. High-throughput functional annotation and data mining with the Blast2GO suite. *Nucleic Acids Research* 36(10):3420–3435 DOI 10.1093/nar/gkn176.
- Grefte S, Kuijpers-Jagtman AM, Torensma R, Von den Hoff JW. 2007. Skeletal muscle development and regeneration. *Stem Cells and Development* 16(5):857–868 DOI 10.1089/scd.2007.0058.
- Grewal SIS, Moazed D. 2003. Heterochromatin and epigenetic control of gene expression. *Science* 301(5634):798–802 DOI 10.1126/science.1086887.
- Hart SN, Therneau TM, Zhang Y, Poland GA, Kocher JP. 2013. Calculating sample size estimates for RNA sequencing data. *Journal of Computational Biology* 20(12):970–978 DOI 10.1089/cmb.2012.0283.

- Hashimoto K, Shibuno T, Murayama-Kayano E, Tanaka H, Kayano T. 2004. Isolation and characterization of stress-responsive genes from the scleractinian coral *Pocillopora damicornis*. *Coral Reefs* 23(4):485–491 DOI 10.1007/s00338-004-0410-1.
- Hodgson G. 1990. Tetracycline reduces sedimentation damage to corals. *Marine Biology* 104(3):493–496 DOI 10.1007/BF01314355.
- Hoegh-Guldberg O, Mumby PJ, Hooten AJ, Steneck RS, Greenfield P, Gomez E, Harvell CD, Sale PF, Edwards AJ, Caldeira K, Knowlton N, Eakin CM, Iglesias-Prieto R, Muthiga N, Bradbury RH, Dubi A, Hatziolos ME. 2007. Coral reefs under rapid climate change and ocean acidification. *Science* 318(5857):1737–1742 DOI 10.1126/science.1152509.
- Houlbrèque F, Ferrier-Pagès C. 2009. Heterotrophy in tropical scleractinian corals. *Biological Reviews of the Cambridge Philosophical Society* 84(1):1–17 DOI 10.1111/j.1469-185X.2008.00058.x.
- Humphrey C, Weber M, Lott C, Cooper T, Fabricius K. 2008. Effects of suspended sediments, dissolved inorganic nutrients and salinity on fertilisation and embryo development in the coral *Acropora millepora* (Ehrenberg, 1834). *Coral Reefs* 27(4):837–850 DOI 10.1007/s00338-008-0408-1.
- Hunter CL, Evans CW. 1995. Coral reefs in Kāne'ohe Bay, Hawai'i: two centuries of western influence and two decades of data. *Bulletin of Marine Science* 57(2):501–515.
- **Iobbi-Nivol C, Leimkühler S. 2013.** Molybdenum enzymes, their maturation and molybdenum cofactor biosynthesis in *Escherichia coli*. *Biochimica et Biophysica Acta* **1827(8–9)**:1086–1101 DOI 10.1016/j.bbabio.2012.11.007.
- Jones R, Bessell-Browne P, Fisher R, Klonowski W, Slivkoff M. 2016. Assessing the impacts of sediments from dredging on corals. *Marine Pollution Bulletin* 102(1):9–29 DOI 10.1016/j.marpolbul.2015.10.049.
- Jones P, Binns D, Chang H, Fraser M, Li W, McAnulla C, McWilliam H, Maslen J, Mitchell A, Nuka G, Pesseat S, Quinn AF, Sangrador-Vegas A, Scheremetjew M, Yong S, Lopez R, Hunter S. 2014. InterProScan 5: genome-scale protein function classification. *Bioinformatics* 30(9):1236–1240 DOI 10.1093/bioinformatics/btu031.
- Jones R, Fisher R, Bessell-Browne P. 2019. Sediment deposition and coral smothering. *PLOS ONE* 14(6):e0216248 DOI 10.1371/journal.pone.0216248.
- Jones PL, Wolffe AP. 1999. Relationships between chromatin organization and DNA methylation in determining gene expression. *Seminars in Cancer Biology* **9**(5):339–347 DOI 10.1006/scbi.1999.0134.
- Junjie RK, Browne NK, Erftemeijer PLA, Todd PA. 2014. Impacts of sediments on coral energetics: partitioning the effects of turbidity and settling particles. *PLOS ONE* **9(9)**:e107195 DOI 10.1371/journal.pone.0107195.
- Kamisan N, Naveen SV, Ahmad RE, Kamarul T. 2013. Chondrocyte density, proteoglycan content and gene expressions from native cartilage are species specific and not dependent on cartilage thickness: a comparative analysis between rat, rabbit and goat. *BMC Veterinary Research* 9(1):62 DOI 10.1186/1746-6148-9-62.
- Kaniewska P, Campbell PR, Kline DI, Rodriguez-Lanetty M, Miller DJ, Dove S, Hoegh-Guldberg O. 2012. Major cellular and physiological impacts of ocean acidification on a reef building coral. PLOS ONE 7(4):e34659 DOI 10.1371/journal.pone.0034659.
- Kisker C, Schindelin H, Rees DC. 1997. Molybdenum-cofactor-containing enzymes: structure and mechanism. *Annual Review of Biochemistry* 66(1):233–267 DOI 10.1146/annurev.biochem.66.1.233.

- Kojis BL, Quinn NJ. 1984. Seasonal and depth variation in fecundity of *Acropora palifera* at two reefs in papua new guinea. *Coral Reefs* 3(3):165–172 DOI 10.1007/BF00301961.
- LaJeunesse TC, Parkinson JE, Gabrielson PW, Jeong HJ, Reimer JD, Voolstra CR, Santos SR. 2018. Systematic revision of symbiodiniaceae highlights the antiquity and diversity of coral endosymbionts. *Current Biology: CB* 28(16):2570–2580.e6 DOI 10.1016/j.cub.2018.07.008.
- Lasker HR. 1980. Sediment rejection by reef corals: the roles of behavior and morphology in Montastrea cavernosa (Linnaeus). Journal of Experimental Marine Biology and Ecology 47(1):77–87 DOI 10.1016/0022-0981(80)90139-2.
- Lex A, Gehlenborg N, Strobelt H, Vuillemot R, Pfister H. 2014. UpSet: visualization of intersecting sets. *IEEE Transactions on Visualization and Computer Graphics* 20(12):1983–1992 DOI 10.1109/TVCG.2014.2346248.
- Li H, Huang Z, Ye S, Lu C, Cheng P, Chen S, Chen C. 2014. Membrane labeling of coral gastrodermal cells by biotinylation: the proteomic identification of surface proteins involving cnidaria-dinoflagellate endosymbiosis. *PLOS ONE* **9**(1):e85119 DOI 10.1371/journal.pone.0085119.
- Libbin RM, Ozer R, Person P, Hirschman A. 1976. In vitro accumulation of mineral components by invertebrate cartilage. *Calcified Tissue Research* 22(1):67–75 DOI 10.1007/BF02010347.
- Liew YJ, Zoccola D, Li Y, Tambutté E, Venn AA, Michell CT, Cui G, Deutekom ES, Kaandorp JA, Voolstra CR, Forêt S, Allemand D, Tambutté S, Aranda M. 2018. Epigenomeassociated phenotypic acclimatization to ocean acidification in a reef-building coral. *Science Advances* 4(6):eaar8028 DOI 10.1126/sciadv.aar8028.
- Loiola M, Oliveira MDM, Kikuchi RKP. 2013. Tolerance of brazilian brain coral *Mussismilia braziliensis* to sediment and organic matter inputs. *Marine Pollution Bulletin* 77(1–2):55–62 DOI 10.1016/j.marpolbul.2013.10.033.
- Lough JM, Barnes DJ. 1992. Comparisons of skeletal density variations in *Porites* from the central great barrier reef. *Journal of Experimental Marine Biology and Ecology* 155(1):1–25 DOI 10.1016/0022-0981(92)90024-5.
- Love MI, Huber W, Anders S. 2014. Moderated estimation of fold change and dispersion for RNA-Seq data with DESeq2. *Genome Biology* 15(12):1–21 DOI 10.1186/s13059-014-0550-8.
- Mackay DJ, Hall A. 1998. Rho GTPases. *The Journal of Biological Chemistry* 273(33):20685–20688 DOI 10.1074/jbc.273.33.20685.
- Mendel RR. 2013. The molybdenum cofactor. *The Journal of Biological Chemistry* 288(19):13165–13172 DOI 10.1074/jbc.R113.455311.
- Miller MW, Karazsia J, Groves CE, Griffin S, Moore T, Wilber P, Gregg K. 2016. Detecting sedimentation impacts to coral reefs resulting from dredging the Port of Miami, Florida USA. *PeerJ* 4(November):e2711 DOI 10.7717/peerj.2711.
- Moon SY, Zheng Y. 2003. Rho GTPase-activating proteins in cell regulation. *Trends in Cell Biology* 13(1):13–22 DOI 10.1016/s0962-8924(02)00004-1.
- Morgan MB, Edge SE, Snell TW. 2005. Profiling differential gene expression of corals along a transect of waters adjacent to the bermuda municipal dump. *Marine Pollution Bulletin* 51(5-7):524–533 DOI 10.1016/j.marpolbul.2005.09.023.
- Muscatine L, Falkowski PG, Porter JW, Dubinsky Z. 1984. Fate of photosynthetic fixed carbon in light- and shade-adapted colonies of the symbiotic coral *Stylophora pistillata*. *Proceedings of the Royal Society B: Biological Sciences* 222:181–202 DOI 10.1098/rspb.1984.0058.
- Muscatine L, Porter JW. 1977. Reef corals: mutualistic symbioses adapted to nutrient-poor environments. *Bioscience* 27(7):454–460 DOI 10.2307/1297526.

- **Ogston AS, Field ME. 2010.** Predictions of turbidity due to enhanced sediment resuspension resulting from sea-level rise on a fringing coral reef: evidence from Molokai, Hawaii. *Journal of Coastal Research* **26(6)**:1027–1037 DOI 10.2112/JCOASTRES-D-09-00064.1.
- Padilla-Gamiño JL, Gates RD. 2012. Spawning dynamics in the hawaiian reef-building coral Montipora capitata. Marine Ecology Progress Series 449(March 2014):145–160 DOI 10.3354/meps09530.
- Padilla-Gamiño JL, Hédouin L, Waller RG, Smith D, Truong W, Gates RD. 2014. Sedimentation and the reproductive biology of the Hawaiian reef-building coral *Montipora capitata*. *Biological Bulletin* 226(1):8–18 DOI 10.1086/BBLv226n1p8.
- Pastorok RA, Bilyard GR. 1985. Effects of sewage pollution on coral-reef communities. *Marine Ecology Progress Series* 21(1\2):175–189 DOI 10.3354/meps021175.
- Pertea G, Pertea M. 2020. GFF Utilities: GffRead and GffCompare. *F1000Research* 9(304):304 DOI 10.12688/f1000research.23297.2.
- Pertea M, Pertea GM, Antonescu CM, Chang T, Mendell JT, Salzberg SL. 2015. StringTie enables improved reconstruction of a transcriptome from RNA-Seq reads. *Nature Biotechnology* 33(3):290–295 DOI 10.1038/nbt.3122.
- Philipp E, Fabricius K. 2003. Photophysiological stress in scleractinian corals in response to short-term sedimentation. *Journal of Experimental Marine Biology and Ecology* 287(1):57–78 DOI 10.1016/S0022-0981(02)00495-1.
- Philpott DE, Person P. 1970. The biology of cartilage. II. Invertebrate cartilages: squid head cartilage. *Journal of Morphology* 131(4):417–430 DOI 10.1002/jmor.1051310405.
- **Phuyal S, Farhan H. 2019.** Multifaceted Rho GTPase signaling at the endomembranes. *Frontiers in Cell and Developmental Biology* 7:127 DOI 10.3389/fcell.2019.00127.
- Pollock FJ, Lamb JB, Field SN, Heron SF, Schaffelke B, Shedrawi G, Bourne DG, Willis BL. 2014. Sediment and turbidity associated with offshore dredging increase coral disease prevalence on nearby reefs. *PLOS ONE* 9(7):e102498 DOI 10.1371/journal.pone.0102498.
- Poquita-Du RC, Goh YL, Huang D, Chou LM, Todd PA. 2020. Gene expression and photo-physiological changes in pocillopora acuta coral holobiont following heat stress and recovery. *Microorganisms* 8(8):1227 DOI 10.3390/microorganisms8081227.
- **Poquita-Du RC, Huang D, Chou LM, Mrinalini, Todd PA. 2019.** Short term exposure to heat and sediment triggers changes in coral gene expression and photo-physiological performance. *Frontiers in Marine Science* 6:121 DOI 10.3389/fmars.2019.00121.
- Prada C, Hanna B, Budd AF, Woodley CM, Schmutz J, Grimwood J, Iglesias-Prieto R,
  Pandolfi JM, Levitan D, Johnson KG, Knowlton N, Kitano H, DeGiorgio M, Medina M.
  2016. Empty niches after extinctions increase population sizes of modern corals. *Current Biology: CB* 26(23):3190–3194 DOI 10.1016/j.cub.2016.09.039.
- **Putnam HM, Davidson JM, Gates RD. 2016.** Ocean acidification influences host DNA methylation and phenotypic plasticity in environmentally susceptible corals. *Evolutionary Applications* **9(9)**:1165–1178 DOI 10.1111/eva.12408.
- **Reaka-Kudla ML. 1997.** The global biodiversity of coral reefs: a comparison with rain forest. *Biodiversity II: Understanding and Protecting our Biological Resources* **120(4)**:486–493 DOI 10.2307/1791071.
- Richmond RH, Jokiel PL. 1984. Lunar periodicity in larva release in the reef coral Pocillopora damicornis at Enewetak and Hawai'i. *Bulletin of Marine Science* 34(2):280–287.

- Riegl B, Branch GM. 1995. Effects of sediment on the energy budgets of four scleractinian (Bourne 1900) and five alcyonacean (Lamouroux 1816) corals. *Journal of Experimental Marine Biology* and Ecology 186(2):259–275 DOI 10.1016/0022-0981(94)00164-9.
- Robbins SJ, Singleton CM, Chan CX, Messer LF, Geers AU, Ying H, Baker A, Bell SC, Morrow KM, Ragan MA, Miller DJ, Forêt S, ReFuGe2020 Consortium, Voolstra CR, Tyson GW, Bourne DG. 2019. A genomic view of the reef-building coral Porites lutea and its microbial symbionts. *Nature Microbiology* 4(12):2090–2100 DOI 10.1038/s41564-019-0532-4.
- **Rodríguez-Casariego JA, Mercado-Molina AE, Garcia-Souto D. 2020.** Genome-wide DNA methylation analysis reveals a conserved epigenetic response to seasonal environmental variation in the staghorn coral *Acropora cervicornis. Frontiers in Marine Science* **7**:560424 DOI 10.3389/fmars.2020.560424.
- Rogers CS. 1979. The effect of shading on coral reef structure and function. *Journal of Experimental Marine Biology and Ecology* 41(3):269–288 DOI 10.1016/0022-0981(79)90136-9.
- **Rogers CS. 1983.** Sublethal and lethal effects of sediments applied to common caribbean reef corals in the field. *Marine Pollution Bulletin* **14(10)**:378–382 DOI 10.1016/0025-326X(83)90602-1.
- **Rogers CS. 1990.** Responses of coral reefs and reef organisms to sedimentation. *Marine Ecology Progress Series* 62:185–202 DOI 10.3354/meps062185.
- Roth MS. 2014. The engine of the reef: photobiology of the coral-algal symbiosis. *Frontiers in Microbiology* 5(August):422 DOI 10.3389/fmicb.2014.00422.
- Rushmore ME, Ross C, Fogarty ND. 2021. Physiological responses to short-term sediment exposure in adults of the caribbean coral *Montastraea cavernosa* and adults and recruits of *Porites astreoides*. *Coral Reefs* **40(5)**:1579–1591 DOI 10.1007/s00338-021-02156-0.
- Schlaefer JA, Tebbett SB, Bellwood DR. 2021. The study of sediments on coral reefs: a hydrodynamic perspective. *Marine Pollution Bulletin* 169(August):112580 DOI 10.1016/j.marpolbul.2021.112580.
- Sheridan C, Grosjean P, Leblud J, Palmer CV, Kushmaro A, Eeckhaut I. 2014. Sedimentation rapidly induces an immune response and depletes energy stores in a hard coral. *Coral Reefs* 33(4):1067–1076 DOI 10.1007/s00338-014-1202-x.
- Siebert C, Pett-Ridge JC, Opfergelt S, Guicharnaud RA, Halliday AN, Burton KW. 2015. Molybdenum isotope fractionation in soils: influence of redox conditions, organic matter, and atmospheric inputs. *Geochimica et Cosmochimica Acta* 162(1):1–24 DOI 10.1016/j.gca.2015.04.007.
- Stafford-Smith MG. 1993. Sediment-rejection efficiency of 22 species of australian scleractinian corals. *Marine Biology* 115(2):229–243 DOI 10.1007/BF00346340.
- Stafford-Smith MG, Ormond RFG. 1992. Sediment-rejection mechanisms of 42 species of Australian scleractinian corals. *Marine and Freshwater Research* 43(4):683–705 DOI 10.1071/MF9920683.
- Stephens TG, Strand EL, Mohamed AR, Williams A, Chiles EN, Su X, Bhattacharya D, Putnam HM. 2021. Ploidy variation and its implications for reproduction and population dynamics in two sympatric hawaiian coral species. BioRxiv preprint DOI 10.1101/2021.11.21.469467.
- Stimson JS. 1978. Mode and timing of reproduction in some common hermatypic corals of Hawai'i and Enewetak. *Marine Biology* 48(2):173–184 DOI 10.1007/BF00395017.
- Studivan MS, Rossin AM, Rubin E, Soderberg N, Holstein DM, Enochs IC. 2022. Reef sediments can act as a stony coral tissue loss disease vector. *Frontiers in Marine Science* 8:36 DOI 10.3389/fmars.2021.815698.

- Sumaila UR, Cisneros-Montemayor AM. 2010. A global estimate of benefits from ecosystem-based marine recreation: potential impacts and implications for management. *Journal of Bioeconomics* 12(3):245–268 DOI 10.1007/s10818-010-9092-7.
- Tambutté S, Holcomb M, Ferrier-Pagès C, Reynaud S, Tambutté E, Zoccola D, Allemand D. 2011. Coral biomineralization: from the gene to the environment. *Journal of Experimental Marine Biology and Ecology* 408(1):58–78 DOI 10.1016/j.jembe.2011.07.026.
- Tomanek L. 2015. Proteomic responses to environmentally induced oxidative stress. *The Journal of Experimental Biology* 218(12):1867–1879 DOI 10.1242/jeb.116475.
- Vandermeulen JH, Watabe N. 1973. Studies on reef corals. Skeleton formation by newly settled planula larva of *Pocillopora damicornis*. *Marine Biology* 23(1):47–57 DOI 10.1007/bf00394111.
- **Veron JEN. 2002.** *New species described in Corals of the World.* Townsville, Queensland, Australia: Australian Institute of Marine Science, 206.
- Veron JEN, Stafford-Smith MG, Turak E, DeVantier LM. 2016a. Corals of the World. version 0.01. Available at http://www.coralsoftheworld.org/species\_factsheets/species\_factsheet\_summary/ acropora-cervicornis/?version=0.01 (accessed 21 September 2023).
- Veron JEN, Stafford-Smith MG, Turak E, DeVantier LM. 2016b. Corals of the World. version 0.01. Available at http://www.coralsoftheworld.org/species\_factsheets/species\_factsheet\_summary/ montipora-capitata/?version=0.01 (accessed 21 September 2023).
- Veron JEN, Stafford-Smith MG, Turak E, DeVantier LM. 2016c. Corals of the World. version 0.01. Available at http://www.coralsoftheworld.org/species\_factsheets/species\_factsheet\_summary/ montastraea-cavernosa/?version=0.01 (accessed 21 September 2023).
- Veron JEN, Stafford-Smith MG, Turak E, DeVantier LM. 2016d. Corals of the World. version 0.01. Available at http://www.coralsoftheworld.org/species\_factsheets/species\_factsheet\_summary/ orbicella-faveolata/?version=0.01 (accessed 21 September 2023).
- Veron JEN, Stafford-Smith MG, Turak E, DeVantier LM. 2016e. Corals of the World. version 0.01. Available at http://www.coralsoftheworld.org/species\_factsheets/species\_factsheet\_summary/ pocillopora-acuta/?version=0.01 (accessed 21 September 2023).
- Veron JEN, Stafford-Smith MG, Turak E, DeVantier LM. 2016f. Corals of the World. version 0.01. Available at http://www.coralsoftheworld.org/species\_factsheets/species\_factsheet\_summary/ porites-lobata/?version=0.01 (accessed 21 September 2023).
- Weber M, de Beer D, Lott C, Polerecky L, Kohls K, Abed RMM, Ferdelman TG, Fabricius KE. 2012. Mechanisms of damage to corals exposed to sedimentation. *Proceedings of the National Academy of Sciences of the United States of America* 109(24):E1558–E1567 DOI 10.1073/pnas.1100715109.
- Weber M, Lott C, Fabricius KE. 2006. Sedimentation stress in a scleractinian coral exposed to terrestrial and marine sediments with contrasting physical, organic and geochemical properties. *Journal of Experimental Marine Biology and Ecology* 336(1):18–32 DOI 10.1016/j.jembe.2006.04.007.
- Westbroek P, Yanagida J, Isa Y. 1980. Functional morphology of brachiopod and coral skeletal structures supporting ciliated epithelia. *Paleobiology* **6(3)**:313–330.
- Wiens M, Ammar MS, Nawar AH, Koziol C, Hassanein HM, Eisinger M, Müller IM, Müller WE. 2000. Induction of heat-shock (stress) protein gene expression by selected natural and anthropogenic disturbances in the octocoral Dendronephthya klunzingeri. *Journal of Experimental Marine Biology and Ecology* 245(2):265–276 DOI 10.1016/s0022-0981(99)00167-7.

- Young MD, Wakefield MJ, Smyth GK, Oshlack A. 2010. Gene ontology analysis for RNA-Seq: accounting for selection bias. *Genome Biology* 11(2):R14 DOI 10.1186/gb-2010-11-2-r14.
- Zdunek-Zastocka E, Sobczak M. 2013. Expression of pisum sativum PsAO3 gene, which encodes an aldehyde oxidase utilizing abscisic aldehyde, is induced under progressively but not rapidly imposed drought stress. *Plant Physiology and Biochemistry* 71(October):57–66 DOI 10.1016/j.plaphy.2013.06.027.