

Insights on *Pinna nobilis* population genetic structure in Greece

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The fan mussel *Pinna nobilis* Linnaeus, 1758 is an endemic species of the Mediterranean Sea, protected by international agreements. It is one of the largest bivalves in the world, playing an important role in the benthic communities; yet it has been recently characterized as “Critically Endangered” by the IUCN, due to Mass Mortality Events. In this context, the assessment of the genetic variation of the remaining *P. nobilis* populations and the evaluation of connectivity among them are crucial elements for the conservation of the species. For this purpose, samples were collected from six regions of the Eastern Mediterranean Sea; the Islands of Karpathos, Lesvos and Crete; the Chalkidiki and Attica Peninsulas; and the Amvrakikos Gulf. Sampling was performed either by collecting tissue from the individuals or by using a non-invasive method, i.e., by scraping the inside of their shells aiming to collect their mucus and thus avoid stress induction to them. Conventional molecular techniques with the use of the COI and 16S rRNA genetic markers were selected for the depiction of the intra-population genetic variability. The analyses included 104 samples from the present study and publicly available sequences of individuals across the Mediterranean Sea. The results of this work a) suggest the use of eDNA as an efficient sampling method for protected bivalves and b) shed light to the genetic structure of *P. nobilis* population in the Eastern Mediterranean, knowledge that might prove to be fundamental for the species conservation and hence the ecosystem resilience. The haplotype analyses reinforced the evidence that there is a certain degree of connectivity among the distinct regions of the Mediterranean; yet there is evidence of population distinction within the basin, namely between the Western and the Eastern basins. The combination of both genetic markers in the same analysis along with the inclusion of a

large number of individuals produced more robust results, revealing a group of haplotypes being present only in the Eastern Mediterranean and providing insights for the species' most suitable conservation management.

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24 Short Title:

25 *P. nobilis* genetic structure in Greece

26 Keywords:

27 *Pinna nobilis*, population genetics, genetic structure, Mediterranean, COI, 16S rRNA, eDNA,
28 critically endangered, haplotype networks

29 Abstract

30 The fan mussel *Pinna nobilis* Linnaeus, 1758 is an endemic species of the Mediterranean Sea,
31 protected by international agreements. It is one of the largest bivalves in the world, playing an
32 important role in the benthic communities; yet it has been recently characterized as “Critically
33 Endangered” by the IUCN, due to Mass Mortality Events. In this context, the assessment of the
34 genetic variation of the remaining *P. nobilis* populations and the evaluation of connectivity
35 among them are crucial elements for the conservation of the species. For this purpose, samples
36 were collected from six regions of the Eastern Mediterranean Sea; the Islands of Karpathos,
37 Lesvos and Crete; the Chalkidiki and Attica Peninsulas; and the Amvrakikos Gulf. Sampling was
38 performed either by collecting tissue from the individuals or by using a non-invasive method,
39 i.e., by scraping the inside of their shells aiming to collect their mucus and thus avoid stress
40 induction to them. Conventional molecular techniques with the use of the COI and 16S rRNA
41 genetic markers were selected for the depiction of the intra-population genetic variability. The
42 analyses included 104 samples from the present study and publicly available sequences of
43 individuals across the Mediterranean Sea. The results of this work a) suggest the use of eDNA as
44 an efficient sampling method for protected bivalves and b) shed light to the genetic structure of
45 *P. nobilis* population in the Eastern Mediterranean, knowledge that might prove to be
46 fundamental for the species conservation and hence the ecosystem resilience. The haplotype
47 analyses reinforced the evidence that there is a certain degree of connectivity among the distinct
48 regions of the Mediterranean; yet there is evidence of population distinction within the basin,
49 namely between the Western and the Eastern basins. The combination of both genetic markers in
50 the same analysis along with the inclusion of a large number of individuals produced more
51 robust results, revealing a group of haplotypes being present only in the Eastern Mediterranean
52 and providing insights for the species' most suitable conservation management.

53 Introduction

54 During the autumn of 2016 a massive mortality phenomenon was observed on the Western
55 Mediterranean populations of *Pinna nobilis*, the largest endemic bivalve of the Mediterranean
56 Sea (Darriba, 2017). The mass mortality events (MME) reached quickly the Eastern
57 Mediterranean Sea (Katsanevakis et al., 2019). Although several pathogens have been proposed

58 as the MME agents (Catanese et al., 2018; Carella et al., 2019, 2023; Panarese et al., 2019), the
59 most likely one is the protozoan *Haplosporidium pinnae*, which is considered to affect the
60 digestive gland of the animal, resulting in stress, starvation and in a general dysfunction and
61 finally death of the organism (Box, Sureda & Deudero, 2009; Grau et al., 2022). Based on all the
62 above, the species status in the IUCN red list changed to Critically Endangered (CR) (Kersting et
63 al., 2019).

64 The decline of the pen shell's population was known several years before the MME (Centoducati
65 et al., 2007) due to threats such as the coastal construction activity, the degradation of its
66 habitats, the anchoring -especially at touristic hotspots, the wave action, the byssus exploitation
67 for production of sea silk and the illegal trawling activity (Hendriks et al., 2013; Basso et al.,
68 2015). Therefore, a series of regulations were established aiming to protect this species and
69 ensure its survival; national legislation and international conventions have been in force for the
70 past decades, such as the Barcelona Convention for the Protection of the Marine Environment
71 and the Coastal Region of the Mediterranean and the Council Directive 92/43/EEC on the
72 Conservation of natural habitats and of wild fauna and flora (Annex IV). Nevertheless, the
73 effectiveness of those measures was argued, since *P. nobilis* was still subject to illegal fishing for
74 personal or commercial consumption or for decorative purposes (Katsanevakis et al., 2011).
75 Undoubtedly, *P. nobilis* is a beneficial species for the benthic communities for a number of
76 reasons, since it offers various ecosystem services. As a filter feeder, it filters large amounts of
77 water contributing to seawater clarity (Basso et al., 2015), a process that benefits the meadows of
78 the cohabitant species *Posidonia oceanica* and/or *Cymodocea nodosa* (Trigos et al., 2014). Its
79 large valves provide a hard substrate within a sandy area for many sedentary organisms, so it is
80 fairly considered as an ecosystem engineer (Rabaoui et al., 2015). It sometimes also cohabits
81 with the crustaceans *Pontonia pinnophylax* or *Nepinnotheres pinnotheres* (Hassine, Zouari &
82 Rabaoui, 2008; Akyol & Ulaş, 2015), thus increasing even more the complexity and species
83 richness of the community it lives in. Recently, due to the attention it has attracted, *P. nobilis* has
84 been characterized as a flagship species (Scarpa et al., 2020). Without a doubt, this recognition is
85 significant not only for the conservation of the species itself and the ecosystem it is associated
86 with, but also for raising public awareness about marine environmental issues in general (Polgar
87 & Jaafar, 2018).

88 *P. nobilis* has been the focus of numerous molecular studies conducted in various regions of the
89 Mediterranean Sea over the past decades. A study by Katsares et al. (2008) revealed low genetic
90 differentiation among the examined populations in Thermaikos Gulf (Greece), possibly
91 attributed to the species' pelagic larval stage and the resulting high gene flow. Similar findings
92 were observed in studies conducted along the Tunisian coasts (Rabaoui et al., 2011), which also
93 indicated the absence of a genetic barrier between the Aegean Sea and the Tunisian coasts. A
94 study across a wider area of the Western Mediterranean by Sanna et al. (2013) provided
95 additional insights on *P. nobilis* populations; it was the first one to include a considerable
96 number of samples and, actually indicated a distinct genetic structure between the Western
97 Mediterranean (Sardinia, Corsica, Sicily) and the Eastern Mediterranean (Aegean Sea and
98 Tunisian coasts). Furthermore, it identified the northern Adriatic Sea (Venice) as a distinct
99 population. Interestingly, two other areas in the Adriatic Sea, the natural marine parks of Mljet
100 and Telascica, showed greater similarity to the Western Mediterranean samples than to those
101 from Venice (Ankon, 2017).

102 In 2015, microsatellite markers were used for the first time at *P. nobilis* samples from the
103 Balearic coasts (González-Wangüemert et al., 2015) suggesting their usefulness for the genetic
104 diversity and connectivity assessments. Wesselmann et al. (2018) combined both mitochondrial
105 and microsatellite markers along with lagrangian simulations to suggest a series of insightful
106 conclusions for the populational genetics of *P. nobilis* with the upper aim of enhancing its
107 conservation. In the Gulf of Lion (North-Western Mediterranean Sea) *P. nobilis* populations
108 exhibited high genetic diversity across various locations, although there was no significant
109 genetic differentiation among these populations, thus indicating a genetically homogeneous
110 population spanning the entire coastline (Peyran et al., 2021). Clearly, the small-scale surveys
111 seem to indicate populational homogeneity at a genetic level; however, on a larger
112 Mediterranean scale, where a greater number of samples are included, the distinction becomes
113 more evident.

114 The majority of these publicly available sequences are partial sequences of the mitochondrial
115 DNA, and for the most part COI and 16S rRNA genes. Even though the mtDNA is more widely
116 used for phylogeographic purposes, it can reveal a significant level of differentiation among and
117 within populations, as has been shown in several studies for marine bivalves (Parker et al., 1998;
118 Matsumoto, 2003; Wood et al., 2007; Feng et al., 2011; Fernández-Pérez et al., 2018;

119 Ramadhaniaty, Setyobudiandi & Madduppa, 2018). It should also be noted, that mtDNA in
120 certain bivalves, such as *Donax trunculus* (Theologidis et al., 2008) and *Mytilus* spp. (Zouros,
121 2013), has a biparental inheritance which, undoubtedly, affects population diversity estimates
122 based on it.

123 The aim of the present study was to a) investigate the genetic diversity of the *P. nobilis*
124 populations at the Eastern Mediterranean Sea, an area that has not been well studied in this
125 regard, and b) compare it with similar studies from the Western Mediterranean in an attempt to
126 provide further insights into population structuring of this critically endangered species, which
127 will offer a good estimation on the fitness and diversity of the Greek populations.

128 Material and Methods

129 Sampling area

130 For the purpose of the study 105 samples were analyzed, collected within the period of August
131 2018 – April 2021. The samples were collected from six locations of the Eastern Mediterranean
132 Sea and particularly from the Islands of Karpathos, Lesvos and Crete, the Chalkidiki and Attica
133 Peninsulas, and the Amvrakikos Gulf (Fig. 1, Supplementary Table 1). Depth at each collection
134 point was recorded by the divers using a diving computer. Samples were collected under a
135 relevant research permit (175828/2195 of 14/11/2018) issued by the Greek Ministry of
136 Environment and Energy, General Directorate for the Forests & Forest's Environment,
137 Department of Wildlife and Hunting Management.

138 eDNA sampling

139 The sampling method for Karpathos' samples was non-lethal, non-invasive and low impact
140 aiming at the minimization of the disturbance towards the bivalves, since the tissue removal may
141 provoke stress and make the animal more susceptible to diseases. Initially, a rod of 0.5 cm
142 diameter was placed at the opening of the valves of each animal by the SCUBA divers in order to
143 keep them slightly open, carefully taking into account the fragility of the shell's outermost part.
144 Consequently, a sampling brush, resembling a buccal swab was used (Supplementary Fig. 1) to
145 scrape tissue remnants and mucus from the interior of the valves. The sampling brushes (one for
146 each individual) were placed in small zip bags and stored at -20 °C until further processing.

147 Additionally, the shells' width and height (above and below sediment) were recorded by the
148 divers using a caliper.

149 Tissue sampling

150 All the other samples were collected from sacrificed individuals under research permits, since the
151 initial aim of the sampling was the investigation of the infection of *P. nobilis* from the parasite
152 *H. pinnae*. Specifically, 50-100 mg of different tissues (mantle, gills, digestive gland) from each
153 individual were removed, preserved in absolute ethanol and stored at 4 °C until further
154 processing. As previously, the shells' width and height were recorded by the divers, for the
155 majority of the individuals.

156 DNA extraction

157 DNA was extracted according to the protocol of Sambrook, Fritsch & Maniatis (1989), and as
158 previously described in Grau et al. (2022), both from the brushes as well as from the tissues.
159 Specifically, in the case of the latter, small pieces of the collected tissues were chopped with
160 sterile scissors; triplicate extractions were performed for each tissue. Each replicate sample was
161 washed with 800 ul of sterile distilled water for 15 min, following centrifugation at 13000 g for 2
162 min, as in Darriba (2017). The supernatant was removed and the wash was repeated. Afterwards,
163 each sample was washed with 600 ul of lysis buffer (0.5 M Tris, 0.1 M EDTA, 2% SDS, pH 8.8)
164 for 15 min, following centrifugation at 13000 g for 2 min and removal of the supernatant. The
165 washes with the lysis buffer were repeated twice. The pellet was mixed with 600 ul of lysis
166 buffer and 6 ul of proteinase K (20 mg/ml) and incubated at 55°C overnight. DNA was extracted
167 by precipitation with isopropanol and ammonium acetate (5 M) (Sambrook, Fritsch & Maniatis,
168 1989). In the final step of the DNA extraction protocol; i.e. the elution of the DNA pellet,
169 replicate samples were pooled and their concentration was measured in a NanoDrop 1000
170 spectrophotometer (Grau et al., 2022).

171 PCR amplifications

172 For the PCR amplification of the tissue samples, no specific tissue was chosen but rather a
173 mixture of all the extracted DNAs, for each individual, in similar concentrations. Initially PCR
174 amplifications were performed for the COI and 16S rRNA genes with previously used primers

175 and conditions (Folmer et al., 1994; Sanna et al., 2013, 2014; Leray et al., 2013); however the
176 amplifications were not successful. Therefore, new primers were designed (Table 1) based on the
177 available *P. nobilis* sequences in GenBank (Sayers et al., 2023).

178 Each PCR contained 2 ul of DNA template (about 20 ng/ul), 4 ul of 5X KAPA HiFi Fidelity
179 Buffer (Roche Molecular Systems, Inc.), 1 ul of each primer (10 μ m), 0.8 ul of dNTPs (10 mM
180 each), 1 ul of KAPA HiFi HotStart DNA Polymerase (1 U/uL) (Roche Molecular Systems, Inc.)
181 in a total volume of 20 ul. Amplifications were performed at a BioRad T100 thermal cycler. The
182 PCR protocol was the same for the two genes; namely a denaturation step at 95 °C for 5 min
183 followed by 35 cycles of 98 °C for 20 sec, 53 °C for 30 sec, 72 °C for 30 sec and a final
184 extension step at 72°C for 5 min.

185 Amplification of the 16S rRNA yielded in some cases a double PCR product; in this case,
186 purification of both the PCR products was carried out from a 2% agarose gel using the
187 NucleoSpin Gel and PCR Clean-up (MACHEREY-NAGEL). For the COI amplicons, a sodium
188 acetate-absolute ethanol cleanup protocol was conducted. All purified PCR products were
189 sequenced in an automated sequencer ABI 3730.

190 Analyses

191 The ABI chromatograms were checked and corrected by eye using the BioEdit Sequencing
192 Alignment Editor software (Hall, 2011) and MEGA X sequence analysis software (Kumar et al.,
193 2018). 16S rRNA sequences, COI sequences and concatenated 16S rRNA-COI sequences
194 (following the approach of Sanna et al., 2013) from the present study were aligned with the
195 Clustal W package (Thompson, Higgins & Gibson, 1994) embedded in BioEdit and MEGA X.
196 In addition, publicly available sequences of the corresponding genes of *P. nobilis*, for which
197 sample location information was available, were also downloaded from GenBank and added to
198 the aforementioned alignments (Supplementary Table 2; Supplementary Fig. 2). Phylogenetic
199 trees were constructed using the IQ-TREE web server (Trifinopoulos et al., 2016) with automatic
200 identification of substitution model and FreeRate heterogeneity, 100 bootstraps, 1000 replicates
201 of the SH-aLRT branch test and approximate Bayes test.

202 DnaSP software (Rozas et al., 2017) was used to estimate the following variables: number of
203 haplotypes (h), haplotype diversity (Hd), number of polymorphic loci (Ps), nucleotidic diversity

204 (Pi) and Fst values. With the use of DnaSP .nex archives (nexus format) median joining
205 haplotype networks were generated in PopART (Leigh & Bryant, 2015).
206 A map showing the distribution of the most abundant 16S rRNA-COI haplotypes in the different
207 locations was generated using ggplot2 (v. 3.4.2) (Wickham, 2016) and scatterpie (v.0.2.1) (Yu,
208 2023). Populations were defined based on the Spalding et al. (2007) ecoregions, except in the
209 case of the Aegean Sea where the ecoregion was divided into North and South Aegean Sea. F_{ST}
210 values between populations were calculated using the adegenet (v.2.1.7) (Jombart, 2008), pegas
211 (v.1.2) (Paradis, 2010) and hierfstat packages (v.0.5.11) (Goudet & Jombart, 2022). AMOVA
212 (Analysis of molecular variance) was performed to determine genetic variation between
213 populations using poppr (v.2.9.4) (Kamvar, Tabima & Grünwald, 2014) and pegas packages.
214 Isolation by distance was tested using a Mantel test between a matrix of genetic distances
215 (calculated using Edwards' distance) and a matrix of geographic distance between populations
216 (calculated using Euclidean geographic distances) using adegenet and MASS (v.7.3.57)
217 (Venables & Ripley, 2002) packages. Discriminant Analysis of Principal Components (DAPC)
218 and Principal Component Analysis (PCA) were performed with the adegenet package. hierBAPS
219 was also performed as a method for hierarchical clustering of the sequence data to reveal nested
220 population structure, with the use of rhierbaps (v.1.1.4) (Cheng et al., 2013), phytools (v.1.5.1)
221 (Revell, 2012) and ggtree (v.3.2.1) (Yu et al., 2017) packages. Plots were created using ggplot2.
222 All the aforementioned analyses were performed in R version 4.1.1 (R Core Team, 2021).
223 The map of the sampling sites was generated with the QGIS software. Raw DNA sequences from
224 the present study are available from the European Nucleotide Archive (ENA) (Burgin et al.,
225 2023) using the urls <http://www.ebi.ac.uk/ena/data/view/OX406989-OX407068> (16S rRNA) and
226 <http://www.ebi.ac.uk/ena/data/view/OX407172-OX407248> (COI).

227 Results

228 From the 60 eDNA samples, amplification was successful in 36 for the 16S rRNA and 33
229 samples for the COI gene. All the tissue samples had successful amplifications for both genes.
230 When examining the shell's dimensions, the relationship between the height above sediment
231 with the height buried inside the sediment was quite linear for all the study areas (Supplementary
232 Fig. 3; Supplementary Table 3). However, in the regression between the total height and the
233 width, the population from Lesvos followed a different pattern (Supplementary Fig. 4;

234 Supplementary Table 3). Currently, to the best of our knowledge, the only population with living
235 individuals of *P. nobilis* is the one in Amvrakikos Gulf; all the other sampled locations have no
236 individuals surviving the MME.

237 All the available sequences of the COI and 16S rRNA genes for *P. nobilis* were obtained from
238 the NCBI database in order to be included in the statistical analyses. However, it should be
239 highlighted that each survey -where the sequences derived from- aimed at a distinct gene region,
240 and this resulted in a small overlap when the sequences were aligned all together. Furthermore,
241 some surveys did not include both of the genes amplified in the present study, i.e., they focused
242 either only on the 16S rRNA gene or on the COI gene (Supplementary Fig. 2). For these reasons,
243 it was not possible to create only one dataset that could include all this information and therefore
244 three datasets were created instead (Table 2).

245 The first dataset included 100 sequences (N) (concatenation of COI and 16SrRNA genes) of 982
246 bp from Greece (Eastern Mediterranean Sea). It revealed 34 haplotypes and 45 polymorphic
247 sites. The haplotypic diversity was high (Hd: 0.91 ± 0.017) while the nucleotide diversity was
248 low (Pi: 0.00304 ± 0.00029). The haplotype network had a star-like shape, with two central
249 haplotypes from which all the other haplotypes derive (Fig. 2). The regions of Epanomi and
250 Aggelochori from the North Aegean Sea along with Chios Island and Korinthiakos Gulf formed
251 a distinct group compared to all the other regions. A similar indication of differentiation
252 appeared also in the South Aegean with samples mainly from Karpathos being distinct from the
253 other ones. AMOVA showed that there is a very small variation (0.52%) between the
254 populations but it is much higher, and statistically significant, between the samples (82.02%)
255 (Table 3). F_{ST} values were quite low (Supplementary Table 4), which also implied that there is
256 no differentiation among the populations. When we tested for isolation by distance with the
257 Mantel test, there was no clear isolation by distance pattern (Observation: -0.9748659; Simulated
258 p-value: 1); however, in the scatterplot between the genetic distances and the geographic
259 distances, two clouds of points appeared indicating that there might be distant patches
260 (Supplementary Fig. 5). DAPC, after a-score optimization (10 PCs retained), showed that there
261 are a few admixed individuals (Supplementary Fig. 6) but overall DAPC classification is
262 consistent with the original populations; reassignment to actual population was higher for the
263 South Aegean Sea (>80%), followed by the North Aegean Sea (>60%) and the Ionian Sea
264 (>40%). PCA also showed that the population clusters were not very clear (Supplementary Fig.

265 7). hierBAPS clustered the sequences into four groups (cluster log marginal likelihood: -
266 444.275995882745) (Supplementary Fig. 8).

267 The second dataset included 294 sequences (N) of 714 bp from the Western and Eastern
268 Mediterranean Sea (concatenation of both genes). It revealed the highest number of haplotypes
269 (104) and polymorphic sites (72) among all datasets. The haplotypic diversity (Hd: 0.961 ± 0.005)
270 was high and the nucleotide diversity was moderately high (Pi: 0.00511 ± 0.00019). The
271 haplotype network showed a clear differentiation among the three subregions; Adriatic Sea,
272 Western and Eastern Mediterranean Sea (Fig. 3). A few central, highly frequent haplotypes from
273 the Western Mediterranean Sea split into many closely related unique haplotypes in a star-like
274 scheme. The same structure was observed in the haplotypes that occurred in the Eastern
275 Mediterranean (Ionian, North Aegean, South Aegean, Levantine) although there were a few that
276 were closer to the Western Mediterranean ones. The Venice samples (Adriatic Sea), although
277 distinct, showed a higher relatedness to the Western Mediterranean samples than the Eastern
278 ones. When we plotted the 19 most abundant haplotypes (67% cumulative abundance), it was
279 again evident that there is a population differentiation across the Mediterranean Sea (Figure 4).
280 AMOVA also showed a high variation between the populations (25.07%), but again a higher one
281 between the samples (67.33%), with both values being statistically significant (Table 3);
282 however, F_{ST} values were quite low (Supplementary Table 5).

283 When we tested for isolation by distance with the Mantel test, there was no clear pattern
284 (Observation: 0.05073212; Simulated p-value: 0.401) and there was one single consistent cloud
285 of point in the scatterplot, without discontinuities indicating patches (Supplementary Fig. 9).

286 DAPC, after a-score optimization (12 PCs retained), showed that there are again a few
287 admixed individuals (Supplementary Fig. 10) but overall DAPC classification is consistent
288 with the original populations; reassignment to actual population reached 100% for the South
289 Aegean Sea population and was higher than 40% for all populations, except for the Ionian Sea
290 one (<10%). PCA also did not reveal a clear clustering (Supplementary Fig. 11). hierBAPS
291 clustered the sequences into three groups (cluster log marginal likelihood: -1485.08119596456)
292 (Supplementary Fig. 12).

293 The third dataset included 450 sequences (N) of the COI gene (243 bp) again from the whole
294 Mediterranean Sea. It revealed 48 haplotypes and 36 polymorphic sites. The haplotypic diversity
295 (Hd: 0.652 ± 0.024) was lower than the one observed in the other two datasets and the nucleotide

296 diversity was moderately high ($P_i: 0.00475 \pm 0.00028$). AMOVA showed a low, but statistically
297 significant, variation between the populations (9.56%) and a much higher variation between the
298 samples (89.26%), again statistically significant (Table 3). F_{ST} values were again quite low
299 (Supplementary Table 6). Again, there was no clear pattern of isolation by distance
300 (Observation: 0.2545314; Simulated p-value: 0.213) and one single consistent cloud of points in
301 the scatterplot (Supplementary Fig. 13). DAPC, after a-score optimization (9 PCs retained),
302 showed a very similar pattern for most of the individuals (Supplementary Fig. 14). Reassignment
303 to actual population was very low for the Ionian Sea, North Aegean Sea and Tunisian
304 Plateau/Gulf of Sidra (<10% in all cases); however, it was >40% for the Western Mediterranean
305 and Levantine Sea populations. PCA again did not reveal a clear clustering (Supplementary Fig.
306 15) and hierBAPS clustered the sequences into three groups (cluster log marginal likelihood: -
307 1013.22748489699) (Supplementary Fig. 16).

308 Discussion

309 Population genetic structure

310 This study contributes significantly to the knowledge of *P. nobilis* genetic structure as it provides
311 data from regions of Greece that had not been sampled before. The results of this study indicate
312 that within the Eastern Mediterranean Sea there is no differentiation among the different
313 geographic regions that were sampled implying a high connectivity among them, i.e. the
314 isolation by distance of the populations of North and South Aegean Sea, as well as of Ionian and
315 Aegean Sea is not supported. Similar results have been found for the horse mussel (*Modiolus*
316 *barbatus*), a fact which was attributed to the very long (up to 6 months) pelagic larval stage of
317 the species (Giantsis et al., 2019), which by far exceeds that of *P. nobilis*. Previously, it has been
318 suggested that transplantations may have been responsible for the absence of geographic
319 structure of *Mytilus galloprovincialis* populations in the Aegean Sea (Giantsis, Kravva &
320 Apostolidis, 2012). This might have been the case also for *P. nobilis*, as transplantations had
321 been proposed as a conservation action for the protection of the species (Katsanevakis, 2016;
322 Acarli, 2021), although they were most likely performed only on a local scale and, thus, they
323 should not have influenced the genetic structure of *P. nobilis* at the scale of the Aegean Sea.
324 However, since they were not documented in detail, it is impossible to fully assess their potential

325 effect on the populations. A slight population differentiation is observed between the regions of
326 the North and South Aegean Sea (Fig. 2). For the South Aegean, this could be attributed to the
327 higher number of collected samples compared to the other regions, leading to a higher haplotypic
328 diversity in this case. On the other hand, it could be attributed to the fact that the island of
329 Karpathos is part of a Marine Protected Area (MPA). Although the design of MPAs is generally
330 not based on genetic and genomic data (Sandström et al., 2016; Xuereb et al., 2020), in certain
331 cases it has been shown that they succeed in capturing most of the genetic diversity of their
332 keystone species (Miller & Ayre, 2008), and combined with the protection measures for those
333 species, they might end up preserving a higher number of haplotypes.

334 In the haplotype network of the Eastern Mediterranean (Fig. 2) the haplotypes of North Aegean
335 (Epanomi, Aggelochori, Chios) formed a subgroup shown in blue coloring; yet the Korinthiakos
336 Gulf (Ionian Sea) also shares them. These haplotypes were described by Katsares et al. (2008)
337 and were grouped with the ones from the Tunisian coasts in the research of Sanna et al. (2013),
338 reinforcing the hypothesis of the high connectivity within the Eastern Mediterranean basin. On
339 the other hand, the populations that were sampled within the present study (sampled in the period
340 2018-2021) did not share the above mentioned haplotypes. The intervening period between the
341 studies coincided with the outbreak of the MME, thus raising questions on the association of the
342 populations genetic structuring and the mass mortality events the populations of the species
343 underwent. Unfortunately, it is not possible to estimate the potential genetic structuring of the
344 populations if the MME had not occurred, since the majority of the populations have not
345 survived it. The only exception is the population in Amvrakikos Gulf, for which there is no pre-
346 MME genetic information.

347 The findings of this study support the distinction of the *P. nobilis* individuals into three regions
348 of the Mediterranean Sea. The case of the Adriatic Sea is explained in detail in Sanna et al.
349 (2013); it is a semi-enclosed sea where the genetic flow from the rest of the Mediterranean Sea is
350 not that high. The other basins of the Mediterranean Sea are distinct for a number of other
351 species (*Penaeus (Melicertus) kerathurus*: Zitari-Chatti et al., 2009; *Pomatoschistus tortonesei*:
352 Mejri et al., 2009; *Holothuria polii*: Gharbi & Said, 2011; *Carcinus aestuarii*: Deli, Said &
353 Chatti, 2015), including *P. nobilis* (Sanna et al., 2013). The present study analyzed a high
354 number of samples from the Eastern Mediterranean Sea in order to confirm this pattern. The
355 concatenation of the COI and 16S rRNA genes that was used in the present study has also proved

356 useful and more informative in other genetic studies of bivalves (Yuan, He & Huang, 2009; Feng
357 et al., 2011; Slynko et al., 2018), and shows that there is a certain level of differentiation between
358 the *P. nobilis* populations in the Western vs Eastern Basin. This finding suggests that the already
359 known oceanographic barriers at the Sicily Strait and at the Otranto Strait might be limiting the
360 dispersal of the species and minimizing the gene flow (Čekovská et al., 2020). Due to its short
361 pelagic larval duration stage, *P. nobilis* is a species which is considered to be rather affected by
362 currents and fronts; at the same time, it could have a weak recovery to gene flow from other
363 locations (Pascual et al., 2017) and it could exhibit strong population structuring, as has been
364 shown for other bivalves also characterized by a short planktonic larval stage (Ye, Wu & Li,
365 2015).

366 eDNA and mtDNA marker sequencing

367 eDNA has been used widely for biodiversity assessments (Pereira et al., 2021) and for the
368 detection of cryptic, threatened (Hunter et al., 2018) and invasive species (Ardura et al., 2015).
369 This study was the first, to our knowledge, to use eDNA collected separately from each
370 individual for population genetics assessment on a critically endangered species, although its
371 potential has been advocated for in the literature (Barnes & Turner, 2016; Adams et al., 2019).
372 Our results suggest that the approach can be replicated to other organisms where minimal
373 disturbance and non-invasive methods are in order. In addition, it can be employed in the few
374 remaining populations of *P. nobilis* around the Mediterranean, such as the ones in Ebro Delta
375 (Prado et al., 2020), Occitan coast (Peyran et al., 2022) and the one in Amvrakikos Gulf.
376 Successful amplification for our chosen markers was possible for about half of the samples,
377 which is lower compared to the amplification success from the tissue samples, as was originally
378 expected. However, this number is still considered adequate for the estimation of population
379 genetics indices. Another advantage of this approach is the certainty that each sample of genetic
380 material corresponds to a specific individual which would not have been possible if the eDNA
381 matrix was e.g. water or sediment collected from the study sites; however, there have been
382 studies on population-level inferences from eDNA water samples mostly regarding large
383 populations of fish (Sigsgaard et al., 2020).
384 The results of the present study are based on the sequencing of two mtDNA genes and there is
385 the possibility that they would be different if another approach was used instead or in

386 complement to ours, such as sequencing of microsatellites markers (Meenakshi, Remya & Sanil,
387 2010; Vanhaecke et al., 2012) or ddRAD sequencing (Darschnik et al., 2019; Ortiz et al., 2021)
388 or even the addition of more mtDNA markers (e.g. D-loop) (Pourkazemi, Skibinski &
389 A.Beardmore, 1999; Parmaksiz, 2019). However, as mentioned previously, *P. nobilis* is a
390 critically endangered species and the amount of available samples for deciphering population
391 genetic structure is quite limited; thus, it is challenging to detect the remaining populations of the
392 species and obtain the appropriate number of samples, with a subsequent high DNA quality,
393 while, at the same time, ensuring the well-being of the organisms.

394 Conclusions and Prospects ahead

395 The present study is the first one including such a high number of *P. nobilis* specimens from
396 different areas of the Eastern Mediterranean basin. Therefore it significantly contributes to the
397 knowledge of the genetic variability of the pen shell's populations. In light of the MME,
398 coordinated studies on the genetic diversity of *P. nobilis* throughout the Mediterranean Sea
399 should be performed towards the aim of the conservation and management of the remaining
400 populations of the species. An orchestrated attempt of a pan-Mediterranean investigation appears
401 to be indispensable. Scientific cooperation and use of common standards should be implemented
402 in order to obtain more FAIR data and therefore lead more efficiently to knowledge (Wilkinson
403 et al., 2016). In future conservational plans on a national level, the Eastern Mediterranean basin
404 should be considered as homogenous, based on the findings herein. It is obvious that more
405 samples from the Southern-Eastern Mediterranean (e.g., from Turkey, Syria, Lebanon, Israel,
406 Egypt, Libya) Sea would shed more light on the population genetics status of the species.
407 Furthermore, more detailed approaches, as mentioned in the previous section, should be
408 employed to unravel the genetic structure of *P. nobilis* throughout the Mediterranean.

409 Acknowledgements

410 We would like to thank the Management Agency of the Dodecanese Protected Areas (MADPA)
411 and especially Mr Dinos Protopapas and Mr Giorgos Prearis (captain of the R/V Saria) for
412 providing assistance during our sampling campaign in Karpathos island. The authors would like
413 to also thank the scuba divers for their help in sampling, namely M. Salomidi, V. Gerakaris and
414 Hippocampus Bali dive center. D. Karagiannis of the National Reference Laboratory for Mollusc

415 Diseases (Greek Government) provided the samples MYT1-MYT9 from Lesvos, which were
416 collected under a permit from local authorities (MEE/GDDDP89926/1117). Samples TS1-TS6
417 from Lesvos were collected under a permit by the Department of Agriculture and Fisheries,
418 Decentralized Administration of the Aegean (No. 52321/6-9-2018). Samples from Crete,
419 Chalkidiki and Attica Peninsulas were collected under a relevant research permit (175828/2195
420 of 14/11/2018) issued by the Greek Ministry of Environment and Energy, General Directorate
421 for the Forests & Forest's Environment, Department of Wildlife and Hunting Management. We
422 would also like to thank Dr Katerina Vasileiadou (ORCID: 0000-0002-5057-6417) and Ms
423 Xenia Sarropoulou (ORCID: 0000-0003-3671-9693) for their invaluable contribution to the
424 manuscript and the statistical analyses.

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Figure 1

Map of the sampling locations.

Map of the sampling locations of the current study (numbers in square brackets indicate the number of samples). Credits: Giorgos Chatzigeorgiou. CC0. Map created using the Free and Open Source QGIS.

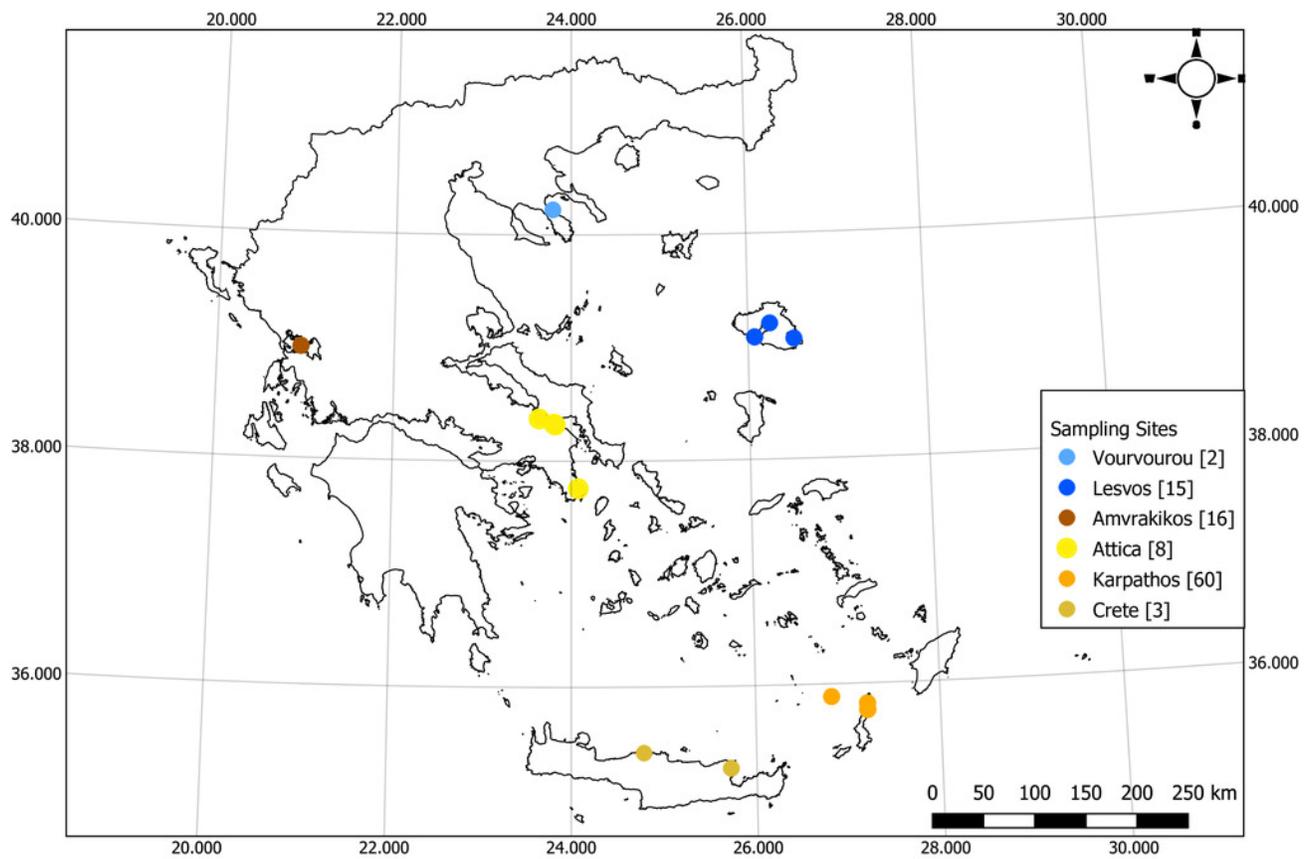


Figure 2

Haplotype network for the Eastern Mediterranean of 16S rRNA-COI dataset.

Haplotype network for the 16S rRNA-COI dataset of Greece (Eastern Mediterranean). Circle size depicts the haplotype frequency; color coding according to sample location; details on the number of samples, sequence size and number of haplotypes are provided in Table 2.

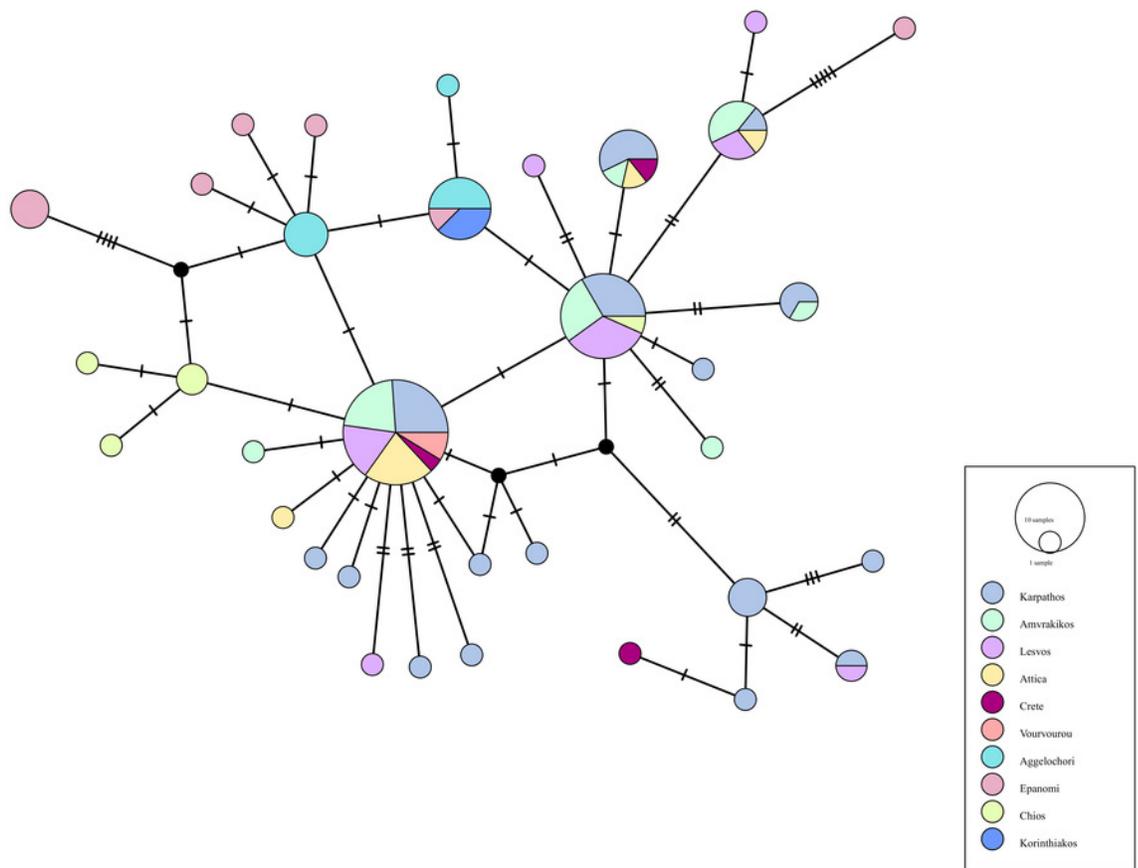


Figure 3

Haplotype network for the 16S rRNA-COI dataset of the Central, Western and Eastern Mediterranean Sea.

Haplotype network for the 16S rRNA-COI dataset of the Central, Western and Eastern Mediterranean Sea. Circle size depicts the haplotype frequency; color coding according to sample location; details on the number of samples, sequence size and number of haplotypes are provided in Table 2.

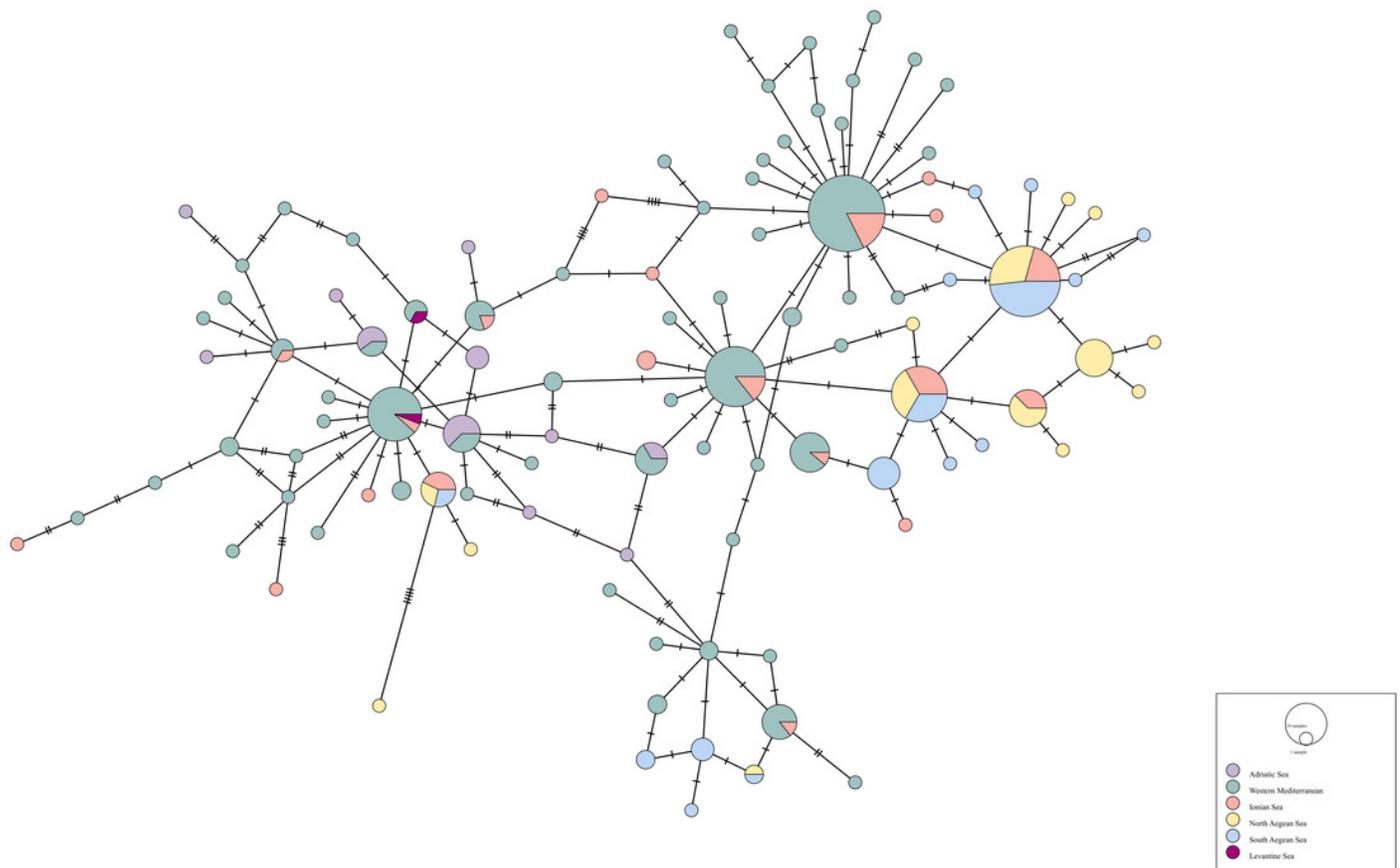


Figure 4

Map showing the distribution of the most abundant haplotypes

Map showing the distribution of the most abundant haplotypes of dataset 2 (16S rRNA-COI Central, Western and Eastern Mediterranean Sea). Credits: Christina Pavludi. CC0. Map created using R.

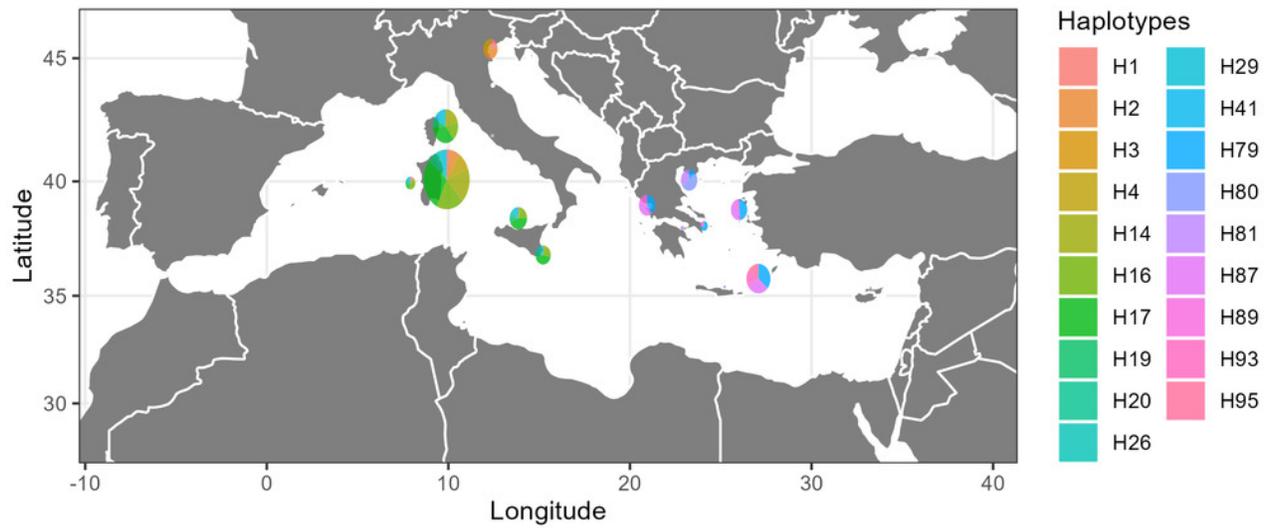


Table 1 (on next page)

Primers used in the present study.

Primers used in the present study.

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Table 1: Primers used in the present study.

Target gene	Forward primer (5' - 3')	Reverse primer (5' - 3')	Amplification length (bp)	Reference
COI	5' - CAGCTTTTGTAGAGGGCG - 3'	5' - CCAAATTACACCAGTCAGCC - 3'	722	this study
	5' - GATCCGGGATAGTAGGTAC - 3'	5' - CMGGATGACCAAARAACC - 3'	645	this study
	5' - ATGGCYGTCGATTTAGC - 3'	5' - CMGGATGACCAAARAACC - 3'	298	this study
LCO 1490		HCO 2198	710	Folmer et al., 1994
COI	mlCOIintF	jgHCO2198	313	Leray et al., 2013
	5' - GGTTGAACTATHATCCNCC - 3'	5' - GAAATCATYCCAAAAGC - 3'	338	Sanna et al., 2013
16S rRNA	5' - GGTAGCGAAATTCCTAGCC - 3'	5' - AAKGGTSGAACAGACCC - 3'	408	this study
16S rRNA	5' - TGCTCAATGCCCAAGGGGTAAAT - 3'	5' - AACTCAGATCACGTAGGG - 3'	450	Sanna et al., 2013
nad3	5' - CCTTATGARTGYGGBTTT - 3'	5' - TCHATAAGYTCATARTAYARCCC - 3'	203	Sanna et al., 2014

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Table 2 (on next page)

Genetic diversity estimates: N: number of sequences, Bp: Base pairs, h: number of haplotypes, Hd: Haplotype diversity, Ps: Polymorphic sites, Pi: Nucleotide diversity.

Genetic diversity estimates: N: number of sequences, Bp: Base pairs, h: number of haplotypes, Hd: Haplotype diversity, Ps: Polymorphic sites, Pi: Nucleotide diversity.

1 Table 2: Genetic diversity estimates: N: number of sequences, Bp: Base pairs, h: number of haplotypes, Hd: Haplotype diversity, Ps:
2 Polymorphic sites, Pi: Nucleotide diversity.

	Dataset	N	Bp	h	Hd	Ps	Pi
1	16S rRNA-COI Greece (Eastern Mediterranean)	100	982	34	0.91 ± 0.017	45	0.00304 ± 0.00029
2	16S rRNA-COI Central, Western and Eastern Mediterranean Sea	294	714	104	0.961 ± 0.005	72	0.00511 ± 0.00019
3	COI Mediterranean Sea	450	243	48	0.652 ± 0.024	36	0.00475 ± 0.00028

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Table 3 (on next page)

AMOVA table using genetic distances based on haplotype frequencies of the *P. nobilis* populations.

AMOVA table using genetic distances based on haplotype frequencies of the *P. nobilis* populations.

1 Table 3: AMOVA table using genetic distances based on haplotype frequencies of the *P. nobilis* populations.

Scenarios		Source of variation	Degrees of freedom	Sum of squares	Components of covariance (Sigma)	Variation (%)	P-value
Greece (Eastern Mediterranean) (16S rRNA-COI)	Ionian Sea (Amvrakikos, Korinthiakos) -	Between Populations	2	18.51356	0.01582342	0.52	0.28
	North Aegean Sea (Lesvos, Aggelochori, Epanomi, Chios, Vourvourou) -	Between samples Within Population	9	57.22781	0.53133279	17.46	0.01
	South Aegean Sea (Attica, Diafani, Tristomo, Astakida, Crete)	Within samples	88	219.5786	2.49521173	82.02	0.01
Central, Western and Eastern Mediterranean Sea (16S rRNA-COI)	Adriatic Sea (Venice) -	Between Populations	5	213.2511	0.8919529	25.07	0.01
	Ionian Sea (Amvrakikos, Korinthiakos, Sicily) -	Between samples Within Population	13	77.38444	0.2702837	7.60	0.01
	North Aegean Sea (Lesvos, Aggelochori, Epanomi, Chios, Vourvourou) -						
	South Aegean Sea (Attica, Diafani, Tristomo, Astakida, Crete) -						
	Levantine Sea (Cyprus) - Western Mediterranean (Italy, Sicily, Corsica, Sardinia)	Within samples	275	658.6637	2.3951407	67.33	0.01
Mediterranean Sea (COI)	Adriatic Sea (Venice) -	Between Populations	6	32.72616	0.09382413	9.56	0.01
	Ionian Sea (Amvrakikos, Korinthiakos, Sicily) -	Between samples Within Population	19	20.06477	0.01168851	1.19	0.06
	North Aegean Sea (Lesvos, Aggelochori, Epanomi, Chios, Vourvourou) -						
	South Aegean Sea (Attica, Diafani, Tristomo, Astakida, Crete) -						
	Levantine Sea (Cyprus) -						
	Western Mediterranean (Bizerta Lagoon, Italy, Sicily, Corsica, Sardinia, France, Spain) -						
	Tunisian Plateau/Gulf of Sidra (El Ketef, Stah Jaber, Kerkennah Island, El Bibane Lagoon)	Within samples	423	370.5275	0.87595168	89.26	0.01

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