

Response of soil extracellular enzyme activity and stoichiometry to short-term warming and phosphorus addition in desert steppe

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Abstract

Background. P is regarded as one of the major limiting factors in grassland ecosystems. Soil available phosphorus deficiency could affect soil extracellular enzyme activity, which plays an essential role in microbial metabolism. Yet, it is still unclear how soil available phosphorus affects soil extracellular enzyme activity and microbial nutrient limitation of desert grassland in the context of climate warming.

Methods. The study carried out a short-term open-top chambers (OTCs) experiment in desert steppe to examine the effects of warming, P addition and their interaction on soil properties, the activities of soil extracellular enzyme s and stoichiometries.

Results. The findings demonstrated that soil acquisition enzyme stoichiometry of C: N: P was 1.2:1:1.5 in this experiment region, which deviated from the global mean scale (1:1:1). Warming increased soil AN (ammonium nitrogen and nitrate nitrogen) contents, decreased MBC (microbial biomass carbon) and MBN (microbial biomass nitrogen). Phosphorus addition raised the soil available phosphorus contents and MBP (microbial biomass phosphorus). Soil extracellular enzyme activities and stoichiometries in desert grassland are largely impacted by soil AN, MBC: MBP, and MBN: MBP. These results revealed that the changes of soil available nutrients and stoichiometries induced by short-term warming and P addition could influence soil microbial activities and alleviate soil microbial carbon and phosphorus limitation. Our findings highlight the critical role played by soil available phosphorus in regulating soil extracellular enzyme activity and microbial nutrient limitation of desert grassland. It will be better to comprehend the microbiological mechanisms underlying these events with further research on the soil microbial communities.

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Introduction

Phosphorus (P), one of the most significant mineral elements, is crucial for plant development, substance synthesis and energy metabolism (Sharma et al.,2020), Yet, due to the effects of climate change, human activities, land use patterns and other reasons, there has been a severe shortage of soil available P worldwide (Hinsinger, 2001; Hou et al., 2020). P has evolved into a major limiting factor for plant growth in grassland ecosystems (Du et al., 2020), which may have negative effects on primary productivity and other ecological processes (Tang et al.,2018; Hou et al.,2020). Since 1850-1900, the temperature of the atmosphere has risen by around 1.1 °C, and it is predicted to reach or exceed 1.5-2.0 °C in the future (IPCC, 2021). The availability of nutrients in the soil may change as a result of climate warming (Feike et al.,2012; Hu et al., 2022), which requires soil microbes to secrete extracellular enzymes in response to the variation of soil nutrients (Raiesi and Salek-Gilani,2018; Cui et al.,2019a), this is because the activity of soil extracellular enzymes regulates the decomposition of soil organic matter and the cycling of nutrients, which in turn affects the function of ecosystem services (Güsewell, 2004; Hu et al., 2016). Soil extracellular enzyme secretion is frequently the rate-limiting step of microbial metabolism (Nannipieri et al.,2018). In this process, the temperature affects the soil microbial activity (Zi et al.,2018). According to the studies, global warming has the potential to change soil available phosphorus by altering mean annual temperature and precipitation(Hou et al.,2018),

and also have an impact on phosphorus fixation, mineralization, and weathering by indirectly affecting plant growth and soil microbial community structure (Delgado-Baquerizo et al., 2013). Evaluating the degree of nutrient availability and limitation of the microbial environment can be achieved by analyzing soil C, N, and P acquisition enzyme activities and stoichiometries (Moorhead et al., 2013; Dong et al., 2019).

The previous findings demonstrated that climate change can increase the availability of nutrients by accelerating the microbial breakdown of soil organic matter (Bardgett et al., 2008). The demand for carbon and phosphorus from microbes may change as a result of soil short-term warming, leading to greater microbial phosphorus limitation (Zheng et al., 2020). The addition of P enhanced the soil available phosphorus, reduced phosphorus limitation, and increased carbon sequestration (Manzoni et al., 2012; Wang et al., 2022). P addition had both enhancing and inhibitory impacts on the soil extracellular enzyme activity (Jing et al., 2016; Wang et al., 2020a), even the input rates of various phosphorus fertilizers affected soil acquiring enzyme activities (Waring et al., 2014; Tian et al., 2016). These findings imply that potential mechanisms influencing soil extracellular enzyme activity and microbial metabolism induced by soil available P in the context of climate change are controversial.

A typical fragile grassland ecosystem, the desert steppe is especially vulnerable to climate change (Zuo et al., 2020), which is important to sustain ecological security and socioeconomic development (Kang et al., 2007). The degradation of desert grasslands and the decline of ecosystem stability and service functions may be driven by climate change, overgrazing, and unreasonable exploitation (Reynolds et al., 2007; Deng et al., 2014; Zhang et al., 2019). Given the importance of the soil available nutrients for primary production and other ecological processes in grassland ecosystems (Tang et al., 2018), understanding the relationship between soil available phosphorus, soil extracellular enzyme activity, and its stoichiometry under warming conditions can help provide insight into the role of soil available phosphorus in controlling ecological processes, which is crucial to actively respond to climate warming and develop effective conservation strategies for desert grassland (Liu et al., 2017; Wang et al., 2021). Here, we set up a field experiment to investigate the response of soil extracellular enzyme activity and its stoichiometry to warming and P addition in desert grassland, we sought to address the following two issues: (i) how soil extracellular enzyme activities, microbial nutrient limitation reacts to short-term warming and P addition; and (ii) what are the major factors affecting soil extracellular enzyme activities and stoichiometries. Based on the previous studies (Waring et al., 2014; Zheng et al., 2020), we hypothesized that (i) the degree of nutrient limitation of soil microorganisms was closely related to the content of soil available phosphorus; (ii) warming and P addition would significantly reduce soil extracellular enzyme activity in desert grassland.

Materials & Methods

Study site

The experiment was conducted on a desert steppe in Eastern Yanchi County (37°04'-38°10' N, 106°03'-107°04' E), Ningxia Hui Autonomous Region, northwest China. It is located on the southwest edge of the Mu Us Desert, which has a semiarid continental monsoon climate.

According to meteorological data from the Yanchi Meteorological Station, the average annual temperature is 8.8°C, and the average annual precipitation is 298.15mm, with the majority of the precipitation falling between July and September (1980-2021). The habitat type is the desert steppe. The soil type is classified as Arenosol (IUSS, 2015), with 4.12 g·kg⁻¹ of soil organic matter, 0.40 g·kg⁻¹ of total nitrogen, 0.32 mg·kg⁻¹ of total phosphorus, 2.30 mg·kg⁻¹ of available phosphorus, and pH 8.57. The dominant species in the region are *Agropyron mogolicum*, *Lespedeza potaninii*, *Caragana korshinskii*, *Stipa bungeana*, and *Polygala tenuifolia* Willd.

Experimental Design

We conducted a randomized split-plot design with two temperature treatments (CK, control; W, warming) as the main plot and three P addition levels (0 g·m⁻²·yr⁻¹, 5 g·m⁻²·yr⁻¹, and 10 g·m⁻²·yr⁻¹) as the subplot in April 2022. There were six treatments: P₀, P₅, P₁₀, WP₀, WP₅, and WP₁₀, each replicated four times, for a total of 24 treatment plots. A PVC sheet was put into the soil at a depth of 0.8 m to divide each subplot from the main plot, which was spaced 3 m apart to provide a buffer zone. The phosphorus fertilizer was supplied by triple superphosphate [Ca (H₂PO₄)₂·H₂O], which was evenly distributed to the treatment plots before rainfall from early June to August. Although the rate of phosphorus fertilizer input in this study is higher than the rate at which phosphorus is deposited in the atmosphere of northern China (Zhu et al., 2016), it meets with the current agricultural fertilization in China (Cui et al., 2020; Guo et al., 2022).

Based on the meteorological monitoring of the study site from 1980 to 2021 (Figure S1) and the existing research foundation of the team (Ma et al., 2019; Ma et al., 2021), we modified the experimental warming device. OTCs (Open Top Chambers) were used as a passive temperature rise device in this investigation. The stainless steel and high-transmittance glass material (5mm thick) were used to construct a regular octagonal prism structure, which has a substrate area of 5.6 m² and a vertical height of 1.8 m. They were installed in the sample plot permanently to avoid disruption. The air and soil temperatures at 15cm above and below ground were automatically recorded every half-hour using HOBO MX2302A and HOBO MX2201 data loggers, respectively (Onset Computer Corporation, Bourne, Massachusetts, USA).

Soil sampling and measurements

Three soil cores (5 cm in diameter, 15 cm in depth) were collected randomly from each plot after removing any surface debris on August 25, 2022. These soil samples were cleaned of plant roots, stone, or other pollutants, and manually homogenized to one composite sample, then returned to the laboratory immediately. The soil was divided into two parts, one of which was air-dried and sieved through a 2 mm mesh to measure soil organic carbon (SOC), total nitrogen (TN), total phosphorus (TP), and available phosphorus (AP). The other of which was stored at 4 °C to measure soil soluble organic carbon (DOC), ammonium nitrogen (NH₄⁺-N), nitrate nitrogen (NO₃⁻-N), microbial biomass, and soil C-, N-, P- acquiring enzyme activities.

Soil moisture content (SMC) was determined by oven-drying the soil at 105°C for 48 hours. A PHS-3E glass pH electrode (Leici, Shanghai, China) was used to measure the pH of the soil in a suspension of air-dried soil and distilled water (1:5, w/v) (Anderson and Ingram, 1993). A TOC analyzer was used to measure the soil DOC concentration (Vario TOC, Elementar, Hanau,

Germany). Soil available nitrogen ($\text{NH}_4^+\text{-N}$ and $\text{NO}_3^-\text{-N}$) was determined by the KCl extraction method with continuous flow analyzer; AP was determined by NaHCO_3 extraction with molybdenum antimony anti-colorimetric method; Soil SOC and TN were determined by potassium dichromate external heating method and $\text{H}_2\text{SO}_4\text{-H}_2\text{O}_2$ digestion with Kjeldahl method; Soil TP was determined by $\text{HClO}_4\text{-H}_2\text{SO}_4$ digestion with molybdenum antimony anti-colorimetric method (Nelson, 1996). Microbial biomass (MBC, MBN, MBP) was extracted and analyzed by chloroform fumigation (Brookes et al., 1985; Vance et al., 1987), respectively.

Soil enzyme extraction and vector analysis

Soil C ($\beta\text{-1,4-glucosidase}$, BG), N (leucine aminopeptidase, LAP, and $\beta\text{-1,4-N-acetylglucosaminidase}$, NAG), and P acquisition enzyme (alkaline phosphatase, ALP) activities were determined using a modified standard fluorescence technique (Saiya-Cork et al., 2002; German et al., 2011). Soil extracellular enzyme activity (EEA) was expressed in units of $\text{nmol g}^{-1} \text{h}^{-1}$ (Sinsabaugh et al., 2008; Sinsabaugh et al., 2009; Waring et al., 2014). The ratios of C, N, and P acquisition enzymes were calculated by the following formulae to determine soil extracellular enzyme stoichiometry (EES), respectively. Scatter plots of soil ecoenzymatic stoichiometry were used to identify microbial resource-limited (Hill et al., 2012; Jiang et al., 2022).

$$\text{Soil e C: N} = \ln(\text{BG}) / \ln(\text{LAP} + \text{NAG})$$

$$\text{Soil e C: P} = \ln(\text{BG}) / \ln(\text{ALP})$$

$$\text{Soil e N: P} = \ln(\text{LAP} + \text{NAG}) / \ln(\text{ALP})$$

Soil C, N, and P-acquiring enzyme stoichiometries were assessed using vector analysis. Vector length and vector angle were calculated by the following formulae, respectively (Moorhead et al., 2016). Limitations on N and P are shown by vector angles of less than 45° and greater than 45° , respectively, while a relative carbon limitation is shown by vector length.

$$\text{Relative C limitation} = \text{Vector Length} = \sqrt{[\ln(\text{BG}) / \ln(\text{LAP} + \text{NAG})]^2 + [\ln(\text{BG}) / \ln(\text{ALP})]^2}$$

$$\text{N/P limitation} = \text{Vector angle} (^\circ) = \text{Degrees} \{ \text{ATAN2} [(\ln(\text{BG}) / \ln(\text{ALP})), (\ln(\text{BG}) / \ln(\text{LAP} + \text{NAG}))] \}$$

Statistical analyses

A two-way analysis of variance using a split-plot design was carried out to investigate the effects of warming, P addition, and their interaction effects on soil extracellular enzyme activity and stoichiometries of desert grassland. One-way ANOVA was used to compare the differences in soil properties, soil extracellular enzyme activity, and stoichiometries of the different treatments. Significant differences were tested using Duncan's multiple comparison tests ($P < 0.05$), and the data for soil enzyme activity had been transformed using a natural logarithm before statistical analysis. Before performing an ANOVA, all data were tested for normality of the residuals and homogeneity of variances in SPSS 25.0 for Windows (SPSS Inc., Chicago, Illinois, USA). Pearson correlation analysis was used to determine the association between soil characteristics, microbial biomass and their stoichiometric. Mantel test was used to explore the correlation between carbon, nitrogen, and phosphorus acquisition enzyme activities and soil environmental factors. RDA (redundancy analysis) was used to explore the corresponding relationships of soil properties, microbial biomass, and extracellular enzyme activity in Canoco 5.0. For the graphics,

R 4.2.3 and Origin 2022 (Origin Laboratory Corporation, Northampton, MA, USA) were applied.

Results

Soil temperature and soil moisture content

OTCs had the anticipated warming effect in all temperature-increased treatment plots (Fig. 1), with an average increase of 1.17°C in air temperature 15cm above the ground and 0.65°C in soil temperature 15cm below ground. Warming considerably decreased soil moisture content by 13.67%. While P addition significantly increased SMC by 11.20% and 19.52% at P₅ and P₁₀, respectively. Soil moisture content was not significantly affected by warming × P addition (Table 1).

Soil properties

No significant effects of SOC, TN, or TP by warming (Table. 1). P addition had distinct effects on SOC under different P addition rates, experiencing a decline of 8.02% at P₅ and an increase of 2.44% at P₁₀. P addition significantly increased TP and decreased TN. Warming × P addition resulted in a rise of SOC and TP. As for TN, WP₁₀ decreased by 4.48%, while WP₅ grew by 13.43%. Warming and P addition did not significantly affect SOC: TN, but decreased SOC: TP. Warming × P addition increased SOC: TP at WP₅ while lowering it by 22.59% at WP₁₀. Soil TN: TP decreased by 2.58%, and 29.38% at WP₅ and WP₁₀, respectively.

The soil pH was not significantly affected by warming, P addition, and their interactions (Table 1). P addition and warming both lowered soil DOC contents, but their interaction effects significantly increased DOC 12.22% and 15.06% at WP₅ and WP₁₀, respectively. Warming and warming × P addition significantly increased the AN content. P addition decreased AN by 6.67% at P₅ and did not change at P₁₀. The addition of P had a considerable impact on available P content, and the higher the rate of P addition, the greater the increase in soil available P content. Nevertheless, there was no discernible influence of available P by warming. Warming and P addition decreased DOC: AN. Warming × P addition significantly affected DOC: AN, with 33.05% decrease in WP₅ and no significant difference in WP₁₀. P addition and warming × P addition had a significant negative impact on DOC: AP and AN: AP.

Soil microbial biomass and their stoichiometry

Compared to the control plots (P₀), all of the treatments led to a reduction in MBC and MBN (Figs. 2A and 2B). Warming and P addition did not significantly affect MBC and MBN, but their interaction significantly decreased them. MBP responded differently from MBC and MBN (Fig. 2C), warming lowered MBP by 9.01%, P addition and warming × P addition considerably increased MBP. Ratios of MBC: MBN, MBC: MBP, and MBN: MBP were significantly reduced by warming and P addition. Warming × P addition increased MBC: MBN (Fig. 2D) and a substantial decrease in MBC: MBP (Fig. 2E) and MBN: MBP (Fig. 2F).

Soil enzyme activity and their stoichiometry

Soil BG was significantly lowered by warming, P addition, and the effects of their interaction (Fig.3A). As for warming × P addition, the activity of BG at WP₅ was higher than that at WP₁₀. LAP+NAG was significantly decreased by P addition (Fig.3B), but not significantly affected by

warming and their interaction effects. Soil ALP was significantly reduced by warming, P addition, and warming \times P addition (Fig.3C). Warming and P addition reduced ln (BG): ln (LAP+NAG) (Fig.3D), while demonstrating a different trend by Warming \times P addition, rising by 4.10% at WP₅, and falling by 13.93% at WP₁₀. Regarding ln (BG): ln (ALP) increased at WP₅ (Fig. 3E), with no significant impacts from other treatments. Warming, P addition and the interaction effects raised ln (LAP+NAG): ln (ALP) (Fig.3F). In addition, a significant correlation between BG, NAG+LAP, and ALP activity was also discovered (Fig. 4).

Vector analysis

All the treatments were over the 1:1 line (Fig. 5A), which showed that microbial nutrients were severely phosphorus limited. Warming, P addition, and their interaction effects reduced vector angle (Fig. 5B). Vector length and vector angle were decreased as a result of warming and P addition (Figs. 5C and 5D). Warming \times P addition increased vector length by 11.64% at WP₅ and decreased by 7.53% at WP₁₀.

Relationships between soil properties and soil extracellular enzyme activities

Pearson correlation showed a significant relationship between soil AP and DOC: AP, AN: AP, MBN: MBP, MBC: MBP. Soil SMC content was significantly related to DOC: AP, AN: AP. Mantel correlation analysis showed that soil total nutrients and stoichiometry had no significant effects on soil EEAs and EES. However, a positive link between soil C and P acquisition enzyme activities and soil available nutrients, microbial biomass, and their stoichiometry was observed, soil N acquisition enzyme activity was closely related to DOC: AN, DCO: AP, and MBP. (Fig. 6). By using a redundancy analysis, we further examined the connection between the soil available nutrients, soil EEAs, EES, and vectors (Fig.7). The results revealed that the important factors affecting soil EEAs and EES were AN, MBC: MBP, and MBN: MBP. There had a significant positive correlation between AN and e C: P. e N: P was positively correlated with AP. MBC was positively correlated with e C: N. Vector length was positively correlated with AN, while vector angle was significantly negatively correlated with available P (Fig. 7).

Discussion

Short-term warming

In this study, we conducted a field experiment to investigate the effects of short-term warming and P addition on the soil extracellular enzyme activity and stoichiometries. Short-term warming increased soil AN content (Table 1), and reduced microbial biomass C and N (Fig.2). Soil C-, N-, and P-acquiring enzymes were decreased by warming (Fig. 3), which was consistent with our hypotheses. No significant effect on the soil total nutrients and stoichiometries by warming (Table 1). Correlation analysis and redundancy analysis revealed a significant positive correlation between soil EEAs and soil available nutrients, microbial biomass and their stoichiometry (Figs. 6 and 7), these results were in keeping with the previous studies (Cui et al., 2021; Li et al., 2022b). Further research on soil microbial communities of desert grassland is required to better understand the exact microbiological mechanisms by which microbial biomass controls the soil extracellular enzyme activity.

Soil moisture content may also affect soil EEA (Allison and Treseder, 2008; Li et al., 2022a). Warming considerably decreased SMC in our study (Table 1), however, SMC had no significant impact on soil EEAs and EES (Figs. 6 and 7), and it was not a limiting factor for soil EEA. The experiment demonstrated that soil microorganisms may adapt to dry, infertile soil conditions by adjusting soil extracellular enzyme activity in response to their demands and the surrounding environment (Gong et al., 2015; Wang et al., 2017). Meanwhile, we discovered that warming slightly raised $\ln(\text{BG})$: $\ln(\text{ALP})$ and $\ln(\text{LAP+NAG})$: $\ln(\text{ALP})$, decreased $\ln(\text{BG})$: $\ln(\text{LAP+NAG})$ (Fig. 3), vector length and vector angle (Fig. 5), it is likely because warming reduced the soil microbial biomass C and N, and consequently influenced the secretion of C- and N-acquiring enzymes to adapt to environmental changes (Turner and Wright, 2014). Taken together, short-term warming changed the activities of soil microbial in desert steppe.

Phosphorus addition

Soil phosphatases typically showed higher activity in P-limited ecosystems, and soil C and N-acquiring enzyme activity reflects microbial demand for energy and nutrients (Turner and Wright, 2014). In our study, P addition lowered soil extracellular enzyme activities (Fig. 3), which had a great impact on the activity of the P acquisition enzyme (Marklein and Houlton, 2012; Shi et al., 2021). P addition significantly promoted the soil total P, available P and microbial biomass P content while concurrently reducing DOC, AN, MBC, and MBN contents (Table 1, Fig. 2). Varying alterations were found in MBC, MBN, and MBP, suggesting that soil microorganisms may preferentially fix P (Bünemann et al., 2012), which affected the stoichiometries of soil microbial biomass. Correlation analysis demonstrated a negative correlation between soil C-, P-acquiring enzyme activity and available P, MBP (Fig. 6). The relationship is most likely explained by the fact that P addition increased plant growth and nutrient uptake while reducing the number of nutrients available for the synthesis of soil enzymes (Wang et al., 2008). Yet, contrary to the results of some research, this might be influenced by the factors like the rate of P addition, type of ecosystem and study site (Colvan et al., 2001; Tian et al., 2016; Wang et al., 2020b).

In terms of soil extracellular enzyme stoichiometry, $\ln(\text{BG})$: $\ln(\text{LAP+NAG})$ decreased by P addition, whereas $\ln(\text{BG})$: $\ln(\text{ALP})$ and $\ln(\text{LAP+NAG})$: $\ln(\text{ALP})$ increased (Fig. 3), suggesting that P addition decreased microbial demand for carbon and soil microbial demand for alkaline phosphatase production costs in phosphorus-limited soils, and also affects microbial activity associated with C and N cycling enzymes (Allison et al., 2010). P addition decreased vector length and vector angle (Fig. 5), demonstrating that P addition can alleviate both the soil microbial carbon and phosphorus limitation in desert grassland, the result confirmed our hypothesis that soil microbial nutrient limitation of desert grassland is closely related to soil AP content.

The interactive effects of short-term warming and P addition

In this study, warming \times P addition significantly reduced soil EEAs (Fig. 3). As for C-acquiring enzyme activity, WP_5 was higher than WP_{10} , which may be related to the significant increase in soil DOC and AN content. This, in turn, may have caused the low level of P_5 addition rate to

secrete more C and N acquisition enzymes at higher temperatures, increasing microbial carbon and nitrogen demand (Fig. 2). At WP₁₀, however, there were less available carbon and nitrogen in the soil, and fewer C- acquiring enzymes are secreted by the microorganisms. WP₁₀ was consistently slightly higher than WP₅ throughout the changes in N-, and P-acquiring enzyme activities, suggesting that the amount of available phosphorus may be a vital factor (Table 1). P-acquiring enzymes were most significantly impacted by warming × P addition (Fig. 3). It showed a negative feedback relationship between high available phosphorus content and soil EEA (Allison et al., 2007). The positive connection between the activities of the BG, LAP+NAG, and ALP indicates that these soil enzymes changed significantly under experiment warming and P addition (Fig. 4), and it further illustrates that the microbial acquisition to C, N, and P was altered. About soil EES, warming × P addition resulted in a reduction of ln (BG): ln (ALP) and ln (LAP+NAG): ln (ALP), increased ln (BG): ln (LAP+NAG) at WP₅ and declined at WP₁₀ (Fig. 3). Combining the results of the redundancy analysis and correlation analysis, the primary factors that influenced the soil EEAs and EES were the soil AN, MBC: MBP, and MBN: MBP. It can be shown that the interaction between warming and P addition had an impact on soil physicochemical properties, which altered the stoichiometry of soil extracellular enzymes (Zheng et al., 2015; Zhang et al., 2018).

Additionally, the natural log ratio of soil C-N-P- acquiring enzymes in this study was 1.2:1:1.3 (Table S1), which differs from the global mean scale of 1:1:1 (Sinsabaugh et al., 2009), demonstrating that desert grassland in the study region was largely C and P limitation. This is partially consistent with the findings of the previous study, demonstrating that desert grassland was severely P-limited in the temperate grassland of northern China (Peng and Wang, 2016). The ratio of soil C: N: P acquisition enzymes adjusted to 1.3:1:1.2 at WP₅, it is shown that a low phosphorus addition rate causes both soil microbial carbon and phosphorus limitation in the region, while the addition of P₁₀ alleviated it. Also, previous research had revealed that adding P may boost soil microbial nitrogen demand, and lead to an increase in N-acquiring enzyme activity (Wang et al., 2020b). Nevertheless, it did not occur in our study, and this issue may need to be confirmed by further research. Vector analysis, however, revealed that relative carbon limitation increased at WP₅ and reduced at WP₁₀. As for phosphorus limitation, a higher level of phosphorus addition rate (P₁₀) was beneficial to alleviate phosphorus limitation in the region. Warming × P addition changed the soil available nutrient contents and their stoichiometries, while soil nutrient stoichiometry influences microbial C, N, and P metabolism by regulating soil elemental balance (Cui et al., 2019b). In this process, the secretion of soil microbial enzymes in P-deficiency areas of desert grassland was stimulated to adapt to nutrient limitation (Schimel et al., 2007; Xiao et al., 2018). From the above results, in this study, soil available nutrients, microbial biomass, and their stoichiometry are the key factors contributing to microbial nutrient limitation in desert grassland.

Conclusions

In this study, we investigated how soil C, N, and P acquisition enzyme activities and stoichiometries responded to short-term warming and P addition in desert grassland. The results

showed that warming did not affect soil available P but altered soil AN, MBC, and MBN. P addition significantly raised soil available phosphorus contents, which had positive effects on the activities of BG, LAP+NAG, and ALP. For the different rates of P addition, P₁₀ was preferable to P₅ for eliminating the relative carbon and phosphorus limitation. Warming and P addition slightly raised ln (BG): ln (ALP) and ln (LAP+NAG): ln (ALP), decreased ln (BG): ln (LAP+NAG). About the interaction of warming and P addition, resulting in a reduction of ln (BG): ln (ALP) and ln (LAP+NAG): ln (ALP), increased ln (BG): ln (LAP+NAG) at WP₅ and declined at WP₁₀. The primary factors that influenced the soil EEAs and EES were the soil AN, MBC: MBP, and MBN: MBP. In conclusion, Short-term warming, P addition, and their interaction significantly affected soil extracellular enzymes and stoichiometries, which in turn changed microbial resource acquisition techniques in desert steppe. More extensive research on soil microbial communities is required to confirm the potential role that soil available phosphorus may play in actively adapting to global warming and microbial nutrient constraint.

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Table 1 (on next page)

Warming, P addition and their interaction effects on soil properties, nutrients contents and stoichiometries.

SMC, soil moisture content. SOC, soil organic carbon. TN, soil total nitrogen. TP, soil total phosphorus. DOC, dissolved organic carbon. AN, soil ammonium nitrogen ($\text{NH}_4^+\text{-N}$) and nitrate nitrogen ($\text{NO}_3^-\text{-N}$). AP, soil available phosphorus. The results of two-way split-plot ANOVAs for the treatments are shown in the table (* $P < 0.05$, ** $p < 0.01$, *** $p < 0.001$, and “—” indicate no significance). Lowercase letters show significant differences among the treatments by Duncan’s test at $p < 0.05$.

Treatment		pH	SWC	SOC	TN	TP	SOC:TN	SOC:T P	TN:T P	DOC	AN	AP	DOC:A N	DOC:A P	AN:A P
Control	P ₀	8.55± 0.11a	9.73± 0.41b	6.98± 1.39ab	0.67± 0.12ab	0.34± 0.02c	10.46± 0.62a	20.27± 3.45a	1.94± 0.33a	20.79± 1.50abc	6.75± 0.52cd	3.58± 0.39d	3.09± 0.21a	5.83± 0.26a	1.9± 0.17a
		8.53± 0.09ab	10.82 ± 0.93a	6.42± 0.79b	0.55± 0.12b	0.4± 0.04ab c	11.92± 0.98a	16.18± 0.51bc	1.37± 0.16b c	17.77± 1.27c	6.30± 0.56d	30.43± 1.89b	2.83± 0.26a	0.59± 0.07c	0.21± 0.02c
	P ₁ o	8.48± 0.11ab	11.63 ± 1.20a	7.15± 1.34ab	0.62± 0.1ab	0.48± 0.09a	11.47± 0.58a	14.87± 2.19c	1.29± 0.13c	18.38± 2.19c	6.73± 0.34cd	41.60± 2.80a	2.74± 0.36a	0.44± 0.07d	0.16± 0.01c
		8.39± 0.05b	8.40± 0.26c	7.05± 0.99ab	0.66± 0.13ab	0.37± 0.08bc	10.77± 0.68a	19.3± 3.27ab	1.8± 0.36a b	19.75± 3.47bc	7.61± 1.40bc	3.80± 0.48d	2.62± 0.34a	5.21± 0.7a	1.99± 0.15a
	P ₅	8.44± 0.05ab	9.43± 0.46b	8.34± 0.9a	0.76± 0.13a	0.41± 0.07ab c	11.11± 1.35a	20.56± 2.33a	1.89± 0.43a	23.33± 2.64ab	11.26± 0.78a	19.73± 3.15c	2.07± 0.11b	1.2± 0.21b	0.58± 0.07b
		8.53± 0.11ab	9.59± 0.22b	7.33± 0.41ab	0.64± 0.06ab	0.47± 0.05ab	11.48± 0.88a	15.69± 2.06bc	1.37± 0.17b c	23.92± 2.66a	8.19± 0.82b	46.90± 6.78a	2.95± 0.52a	0.52± 0.12cd	0.18± 0.03c
Warming		—	*	—	—	—	—	—	**	—	**	—	—	*	*
P addition		—	**	—	—	**	—	*	*	—	*	***	*	***	***
Warming × P		—	—	—	—	—	—	—	—	*	**	***	**	***	*

Figure 1

Air temperature at 15cm above the ground and soil temperature at 15cm below the ground in the warming plots (inside of the OTCs) and the control plots (outside of the OTCs).

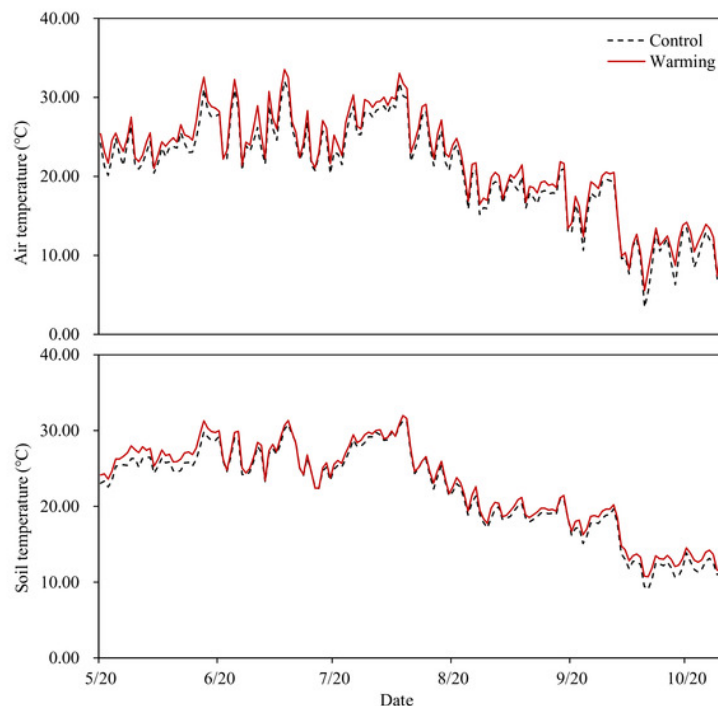


Figure 2

Warming, P addition and the interaction effects on soil microbial biomass and stoichiometries.

MBC, microbial biomass carbon. MBN, microbial biomass nitrogen. MBP, microbial biomass phosphorus. The results of two-way split-plot ANOVAs for the treatments are shown in the figures (* $P < 0.05$, ** $p < 0.01$, *** $p < 0.001$, and ns indicate no significance), lowercase letters above the multi-factor group histograms show significant differences among the treatments by Duncan's test at $p < 0.05$. Error bars show means \pm SE (n=4).

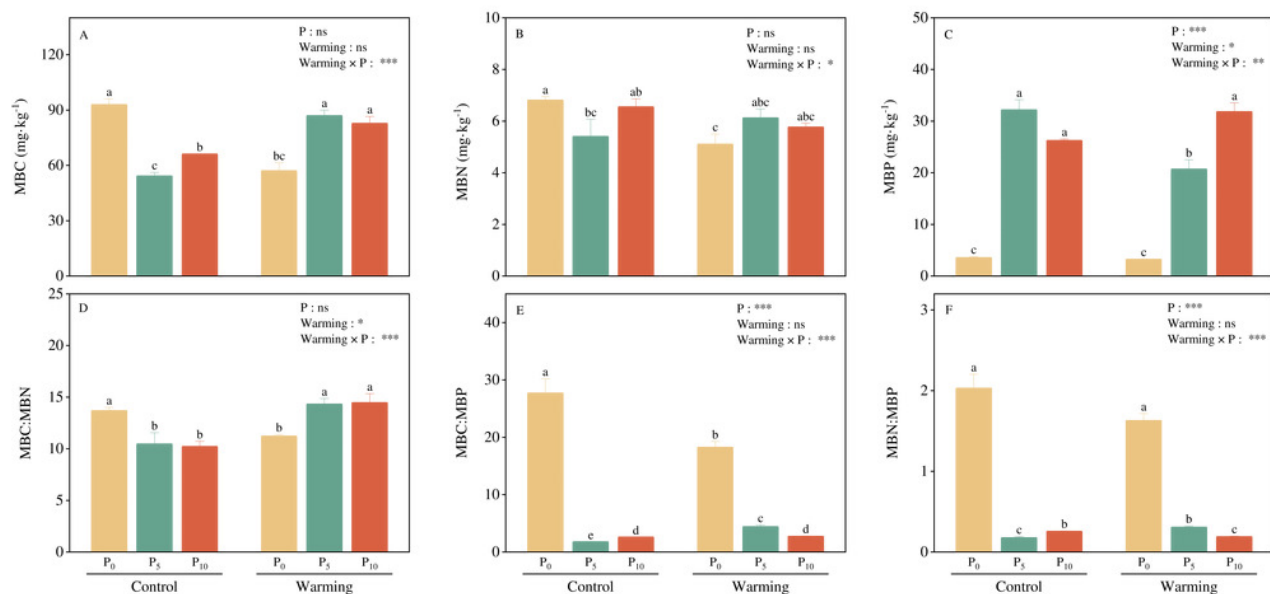


Figure 3

Warming, P addition and the interaction effects on soil extracellular enzyme activities and stoichiometries.

BG, β -1,4-glucosidase. LAP, leucine aminopeptidase. NAG, β -1,4-N-acetylglucosaminidase. ALP, alkaline phosphatase. The results of two-way split-plot ANOVAs for the treatments are shown in the figures (* $P < 0.05$, ** $p < 0.01$, *** $p < 0.001$, and ns indicate no significance), lowercase letters above the multi-factor group histograms show significant differences among the treatments by Duncan's test at $p < 0.05$. Error bars show means \pm SE (n=4).

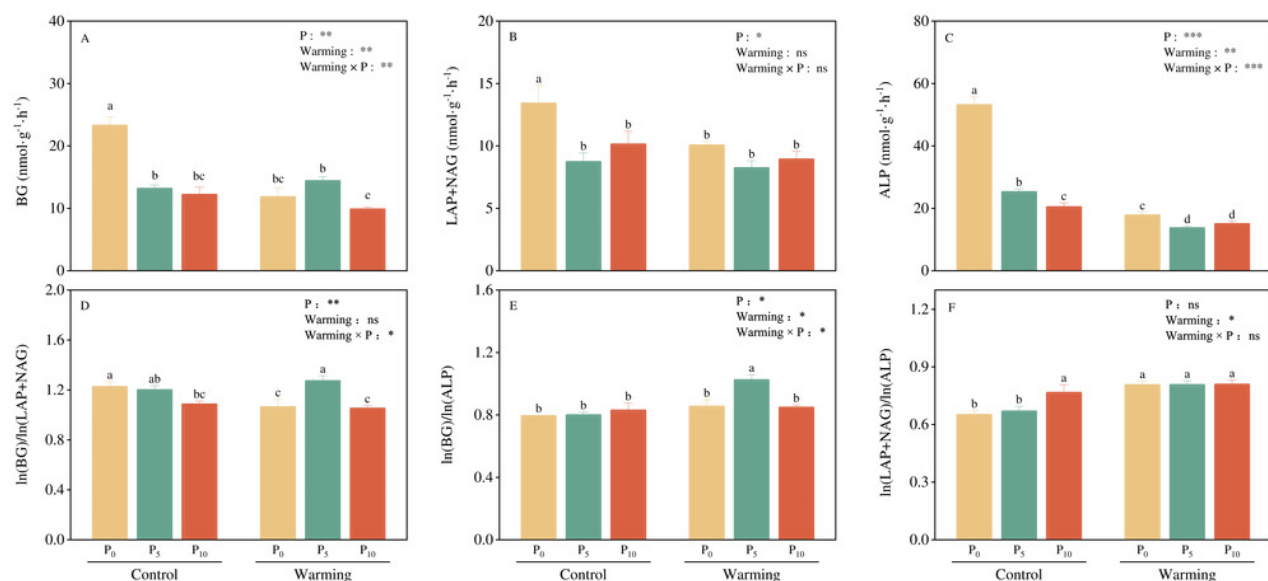


Figure 4

Regression analysis of the soil C, N, and P acquisition enzyme activities.

BG, β -1,4-glucosidase. LAP, leucine aminopeptidase. NAG, β -1,4-N-acetylglucosaminidase.

ALP: alkaline phosphatase. All data were natural log(ln)-transformed.

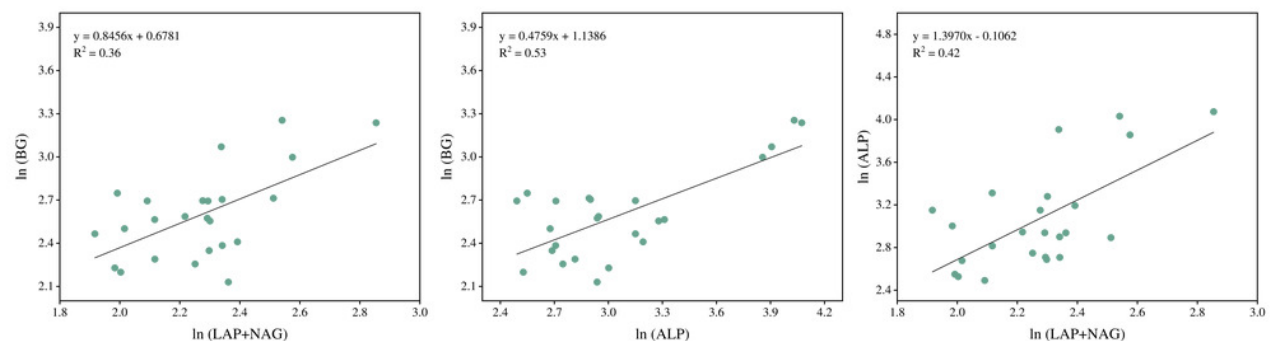


Figure 5

Scatter plots of the eco-enzyme stoichiometry of relative proportions of C/C+P to C/C+N(A). Regression analysis of vector length and vector angle(B). Warming, P addition and the interaction effects on vector length(C) and vector angle(D).

The results of two-way split-plot ANOVAs for the treatments are shown in the figures (* $P < 0.05$, ** $p < 0.01$, *** $p < 0.001$, and ns indicate no significance), lowercase letters above the multi-factor group boxplots show significant differences among the treatments by Duncan's test at $p < 0.05$. Error bars show means \pm SE ($n=4$).

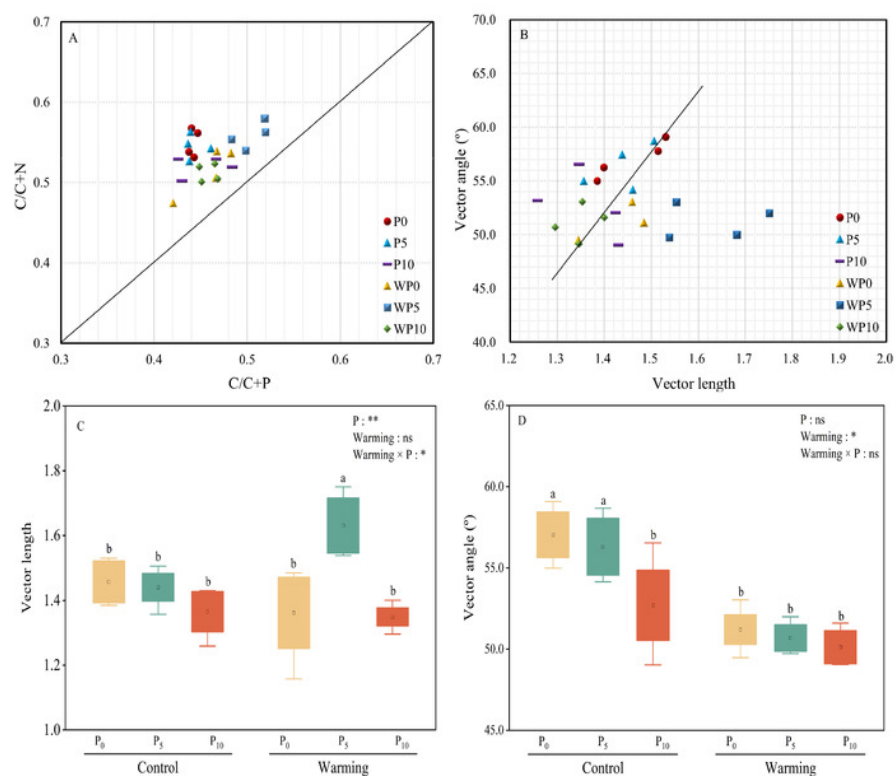


Figure 6

Correlation analysis on soil properties, soil extracellular enzyme activities and stoichiometries.

SOC, soil organic carbon. TN, total nitrogen. TP, total phosphorus. SMC, soil moisture content.

DOC, dissolved organic carbon. AN, soil ammonium nitrogen ($\text{NH}_4^+\text{-N}$) and nitrate nitrogen

($\text{NO}_3^-\text{-N}$). AP, soil available phosphorus. MBC, microbial biomass carbon. MBN, microbial

biomass nitrogen. MBP, microbial biomass phosphorus. C acquisition enzyme, β -1,4-

glucosidase. N acquisition enzyme, leucine aminopeptidase and β -1,4-N-

acetylglucosaminidase. P acquisition enzyme, alkaline phosphatase.

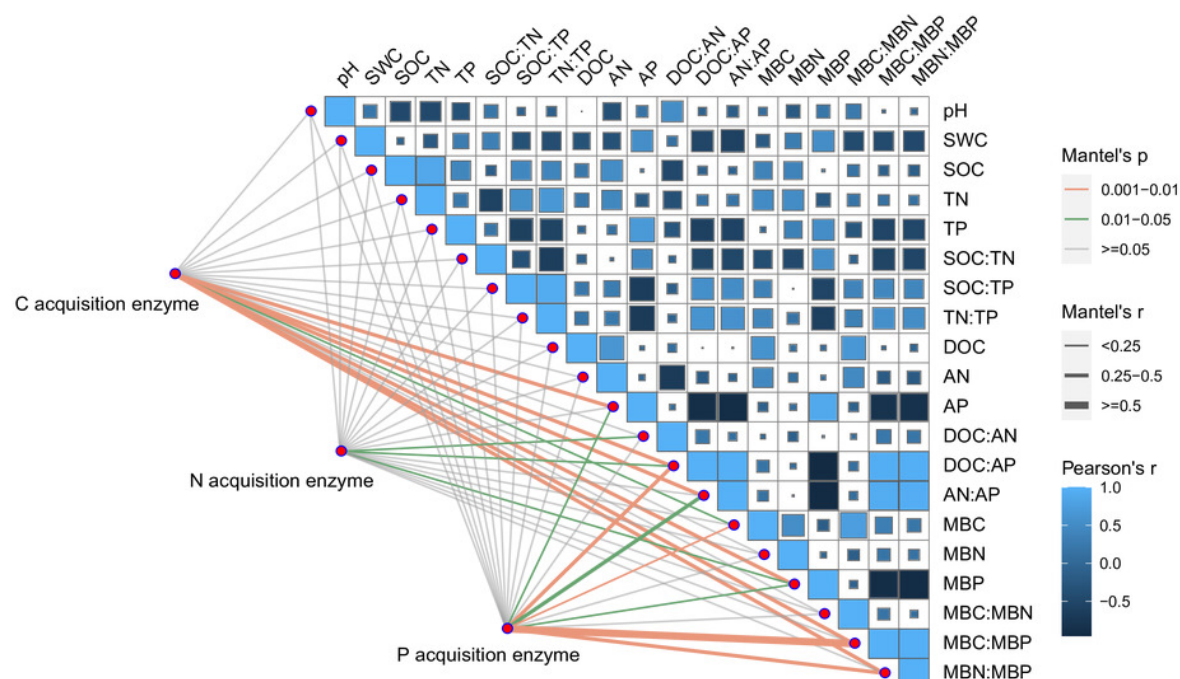


Figure 7

Redundancy analysis (RDA) on soil properties, soil extracellular enzyme activities and stoichiometries, vector length, and vector angle.

SMC, soil moisture content. DOC, dissolved organic carbon. AN, soil ammonium nitrogen ($\text{NH}_4^+\text{-N}$) and nitrate nitrogen ($\text{NO}_3^-\text{-N}$). AP, soil available phosphorus. MBC, microbial biomass carbon. MBN, microbial biomass nitrogen. MBP, microbial biomass phosphorus. BG, β -1,4-glucosidase. LAP, leucine aminopeptidase. NAG, β -1,4-N-acetylglucosaminidase. ALP, alkaline phosphatase. e C: N, e C: P, and e N: P represent $\ln(\text{BG}) : \ln(\text{LAP} + \text{NAG})$, $\ln(\text{BG}) : \ln(\text{ALP})$ and $\ln(\text{LAP} + \text{NAG}) : \ln(\text{ALP})$, respectively. VL, vector length. VA, vector angle.

