

# Using DNA metabarcoding to assess insect diversity in citrus orchards

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**Background.** DNA metabarcoding is rapidly emerging as a cost-effective approach for large-scale biodiversity assessment and pest monitoring. The current study employed metabarcoding to assess insect diversity in citrus orchards in Ganzhou city, Jiangxi, China in 2018-2019. Insects were sampled using Malaise traps deployed in three citrus orchards producing a total of 43 pooled monthly samples.

**Methods.** The Malaise trap samples were sequenced following DNA metabarcoding workflow. Generated sequences were curated and analyzed using two cloud-based data storage and analytical platforms, the Barcode of Life Data System (BOLD) and Multiplex Barcode Research And Visualization Environment (mBRAVE).

**Results.** These platforms assigned the sequences to 2,141 Barcode Index Numbers (BINs), a species proxy. Only 63% of the BINs were shared among the three sampling sites ( $J=0.63$ ) while BIN sharing between any two sites did not exceed 75% ( $J=0.73$  to  $0.75$ ). Shannon diversity index ( $H'$ ) showed a similar pattern of BIN assortment at the 3 sampling sites. Comparison of BIN records against all those on BOLD made it possible to identify 40% of the BINs to a species, 57% to a genus, 97% to a family and 99% to an order. BINs which received a species match on BOLD were placed in one of four categories based on this assignment: pest, parasitoid, predator, or pollinator. As this study provides the first baseline data on insect biodiversity in Chinese citrus plantations, it is a valuable resource for research in a broad range of areas such as pest management and monitoring beneficial insects in citrus gardens.

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25 **Abstract**

26 **Background.** DNA metabarcoding is rapidly emerging as a cost-effective approach for large-  
27 scale biodiversity assessment and pest monitoring. The current study employed metabarcoding to  
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44 pest management and monitoring beneficial insects in citrus gardens.

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52 **Introduction**

53 China ranks among the top three countries globally for citrus cultivation and production (Caserta  
54 et al. 2020). In fact, it ranked first in the world in 2017-2018, producing 32 million tons of fresh  
55 citrus (CGA 2020). Citrus fruit is considered a superior agricultural product and its production  
56 represents an important industry in the rural areas of southern China. However, crop yields are  
57 severely impacted by an array of pests. A large number of insect species, both pest and beneficial  
58 taxa (Niu et al. 2014) occur on citrus but their identification is difficult. Cryptic morphology and  
59 lack of taxonomic expertise are the major challenges in large-scale insect diversity assessments.

60 DNA barcoding (Hebert et al. 2003) the characterization of sequence variation in a standard 658  
61 bp fragment of cytochrome *c* oxidase I gene (COI), has gained global acceptance for specimen  
62 identification and species discovery (Ashfaq et al. 2017; Gwiazdowski et al. 2015; Hebert et al.  
63 2004; Taberlet et al. 2012). This method circumvents the limitations of morphology by  
64 identifying unknown organisms by matching their barcode sequences to reference sequences.  
65 The simplicity and reliability of this method has motivated expansion of the DNA barcode  
66 reference library in the Barcode of Life Data System (BOLD) ([www.boldsystems.org](http://www.boldsystems.org)). The  
67 BOLD system assigns eligible barcode sequences (>507 bp, <1% ambiguous bases, no stop  
68 codons, no contamination) to Barcode Index Numbers (BINs) which serve as proxy for species  
69 (Ratnasingham & Hebert 2013). BOLD currently holds seven million insect barcodes which  
70 have been assigned to more than 700,000 BINs. Implementation of the BIN system has enhanced  
71 the ability of DNA barcoding to discern and count species (Hebert et al. 2016), assess biodiversity  
72 composition (Ashfaq et al. 2018; Telfer et al. 2015), to map species distributions (Ashfaq et al.  
73 2017), and to track species movements across borders (Ren et al. 2017). This success has led to  
74 the use of BINs in the analysis of bulk samples and biodiversity studies by high-throughput  
75 sequencing (HTS) (Cristescu 2014).

76 DNA metabarcoding is a developing approach that identifies the species present in a mixed  
77 sample (bulk DNA or environmental DNA) based on HTS of a specific DNA marker (Comtet et  
78 al. 2015; Hajibabaei et al. 2011; Moriniere et al. 2016; Yu et al. 2012). It differs from  
79 conventional DNA barcoding (usually based on Sanger sequencing of individual specimens)

80 because HTS allows taxonomy to be assigned to hundreds or even thousands of species in a bulk  
81 sample. It achieves this goal by generating amplicons of the barcode region from bulk DNA  
82 extracts which are then sequenced and assigned to operational taxonomic units (OTUs) that are  
83 queried against reference sequences to determine their source species (Cristescu 2014). Studies  
84 have now employed this approach to assess species composition in biological communities such  
85 as aquatic and terrestrial arthropods (Beng et al. 2016; Braukmann et al. 2019; Elbrecht &  
86 Steinke 2018; Ji et al. 2013) making metabarcoding an increasingly cost-effective approach for  
87 large-scale biodiversity studies.

88 Widespread interest in metabarcoding has resulted in data proliferation and the development of  
89 computational tools to aid data analysis. As the BIN system has offered a novel approach to  
90 circumvent morphological bottlenecks to discriminate species, pairing of BINs with HTS has  
91 accelerated biodiversity assessments. Prior studies have used metabarcoding to survey insect  
92 diversity in different ecological settings in China (Huang et al. 2022), but reports on the use of  
93 this technology to explore insect diversity in fruit gardens in this region are lacking. The current  
94 study aimed to fill this gap by coupling metabarcoding with the BIN system to explore insect  
95 diversity in citrus plantations. The composition of insect communities in citrus orchards was  
96 analyzed using DNA metabarcoding followed by data analysis on BOLD and mBRAVE. The  
97 species revealed by BIN matches on BOLD were then searched in citrus pest database – Citrus  
98 pest information system (CPIS, cpis.hzau.edu.cn) and the literature to allow their classification  
99 into pest, parasitoid, predator, or pollinator (Niu et al. 2014; Smaili et al. 2020; Urbaneja et al.  
100 2020). The results provide a valuable resource for research on citrus pest management and  
101 beneficial insects. Exploration of pest and beneficial insect species in Chinese citrus orchards  
102 would allow scientists to screen the appropriate agents for pest management programs and  
103 promote pest biological control.

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## 107 **Materials and Methods**

108 A single Malaise trap (deWaard et al. 2019; Geiger et al. 2016) was deployed in three citrus  
109 orchards (GAN, QIU, SHI) about one kilometer apart in Ganzhou city, Jiangxi province, China  
110 (Figure 1). Samples were collected from April 2018 to July 2019 by replacing the collection  
111 bottles every month – so 43 samples were obtained from the three traps (Table S1). The  
112 collection bottles were stored at -20 °C until used. Specimens from each bottle were sorted into  
113 two size categories: small (e.g., parasitic wasps, *Drosophila*) and large (e.g., butterflies, locust).  
114 Large specimens were subsampled to obtain a tissue block (legs or partial abdomen) of a size  
115 similar to the small ones while small specimens were used in entirety. This was done to achieve  
116 comparable tissue-mass representation for all specimens in the DNA extracts. The  
117 specimens/tissue samples from each bottle were mixed, frozen in liquid nitrogen and then ground  
118 to a fine powder using a disposable mortar and pestle. DNA was extracted with the TIANamp  
119 Genomic DNA kit (DP304, TIANGEN Biotech, Beijing, China) following manufacturer's  
120 protocols. Briefly, 500 mg of the powder was lysed in 2 ml Buffer GA overnight at 56 °C in the  
121 presence of 20 µl proteinase K. The lysate was centrifuged at 5000 g for 10 min and the  
122 supernatant was then aliquoted 40 µl into ten equal volumes for DNA extraction. The quality and  
123 purity of the DNA was assessed using a Nanodrop 2000 and 1% agarose gel electrophoresis.  
124 Three of the ten DNA extracts from each sample were randomly selected for PCR, creating a  
125 total of 129 DNA extracts for analysis.

126

## 127 **PCR amplification**

128 PCR employed a first round to amplify the target region of COI while the second added  
129 adapters to allow discrimination of the sequences derived from each DNA extract (Prosser et al.  
130 2016). PCR1 employed the following primer pair - AncientLepF3  
131 (TTATAATTGGDGGWTTTGGWAATTG) (Prosser et al. 2016) and LepR1  
132 (TAAACTTCTGGATGTCCAAAAAATCA) (Hebert et al. 2004) and the following  
133 thermocycling regime (initial denaturation for 5 min at 95 °C, then 25 cycles of denaturation for  
134 30 s at 95 °C, followed by annealing for 30 s at 55 °C and extension for 60 s at 72 °C, and a final  
135 extension for 10 min at 72 °C). For PCR2, adapters were added to primers employed in PCR1.  
136 All PCR reactions had a total volume of 25 µl and included 12.5 µl 2×PCR Master Mix, 2 µl  
137 each of sequencing primers described above (10 µM), and 1 µl of template purified from the

138 PCR1 product or water (negative template control). PCR2 employed the following  
139 thermocycling regime (Initial denaturation for 45 s at 98 °C, then 6 cycles of denaturation for 15  
140 s at 98 °C, followed by annealing for 30 s at 60 °C and extension for 30 s at 72 °C followed by  
141 final extension for 1 min at 72 °C). Amplicons were purified using QIAquick Gel Extraction Kit  
142 and KAPA Library Quant (Illumina) DNA Standards & Primer Premix was used for amplicon  
143 quantification. Satisfactory amplification was achieved from 126 of the 129 DNA extracts (Table  
144 S1).

145

### 146 **Next generation sequencing**

147 Purified PCR2 products (126 reactions) were sequenced separately on an Illumina Miseq PE300  
148 platform following standard protocols. Each reaction had two sequencing replicates to generate  
149 252 products for sequencing. The amplicon libraries were prepared using MiSeq Reagent Kit v3  
150 (Illumina, Inc.). Briefly, the pooled library was thawed on ice along with HT1 (hybridization  
151 buffer), and then diluted to 2nM in EB buffer. 5 µl of the library were mixed with 5 µl of NaOH  
152 at 0.2 N in a microcentrifuge tube with brief vortexing and 1 min of centrifugation at 280 g.  
153 After 5 min incubation at room temperature, 990 µl pre-chilled HT1 was added to the tube  
154 containing denatured library providing 1 ml of a 10 pM denatured library. The denatured library  
155 was diluted to 8 pM (480 µl of the 10 pM denatured library, 120 µl of the pre-chilled HT1) by  
156 inverted mixing and then pulse centrifugation. Subsequently, the libraries were loaded onto the  
157 reagent cartridge to set up the sequencing run. After cleaning the flow cell, the reagents were  
158 loaded into the flow cell to initiate the sequencing.

159

### 160 **Sequencing data analysis**

161 There are a range of cloud-based platforms and tools that are freely available to analyze NGS  
162 data (Bani Baker et al. 2020). One such tool, mBRAVE, the Multiplex Barcode Research and  
163 Visualization Environment, is a data storage and analytics platform with standardized pipelines  
164 and a sophisticated web interface designed to transform raw HTS data into biological insights  
165 ([www.mbrave.net](http://www.mbrave.net)). mBRAVE integrates common analytical methods and links to the BOLD  
166 System for access to reference datasets (Young et al. 2021; Zieritz et al. 2022) and assignment of

167 sequences to BINs, the features that are unique to this platform only. Results from the 252  
168 sequence libraries were uploaded to mBRAVE ([www.mbrave.net](http://www.mbrave.net)) under the project “MBR-  
169 MTCHN1” where they were analyzed using a standard pipeline (Young et al. 2021; Zieritz et al.  
170 2022) involving sequence trimming (25 bp on each side), quality filtering (minimum QV 25), de-  
171 replication, identification, and OTU generation. mBRAVE has direct access to the DNA barcode  
172 reference libraries on BOLD which allows comparison of the sequence data with the selected  
173 libraries and interpretation of the outcome. This also allows the assignment of the generated  
174 OTUs to the Barcode Index Numbers (BINs) and taxonomy on BOLD. The sequences were run  
175 against the DNA barcode reference library for Insecta on BOLD that represents 0.5 million BINs  
176 and 213,000 named species (the system reference library for mBRAVE ID engine – Insecta). The  
177 library also includes about 60,000 insect barcodes from China (DS-CHINAINS).

178 Sequences on mBRAVE were organized by ‘sequence runs’ and assigned to BINs and Linnaean  
179 taxonomy. The resultant data was downloaded from mBRAVE to summarize the results by  
180 Malaise trap location. Pairwise comparisons of BINs among trap locations were visualized by  
181 VennDiagram, an R package in the R statistical environment. To reduce the likelihood of false  
182 positives, a cleaning step was employed which excluded read counts in the BIN table that  
183 represented less than 0.01% of the total read count for their respective sample. BINs represented  
184 by a single sequence (singletons) were also excluded from the final BIN count. Concordance or  
185 discordance between a BIN and the associated species was determined using “BIN Discordance”  
186 tool on BOLD.

187 Diversity analysis within and among three Malaise trap sites was performed using BINs (as  
188 species proxies) recovered from the NGS sequence data. Alpha diversity at individual sites was  
189 analyzed by Shannon-Wiener index (Shannon 1948) while the diversity comparisons among the  
190 Malaise sites were conducted by Jaccard similarity coefficient (Jaccard 1912). Spread of BINs  
191 over time was determined by calculating BIN incidences in the Malaise sampling events for each  
192 site.

193

## 194 **Results**

### 195 **General patterns of citrus insects**

196 The 252 NGS libraries included 14 Malaise trap collections from GAN, 15 from QIU and 14  
197 from SHI (Table S1). In total, these runs yielded 9.5 million (M) DNA sequences which  
198 dereplicated to 2.7M barcodes averaging 250 bp in length. The 1,515,627 sequences remaining  
199 after filtration were assigned to BINs. Most sequences belonged to Diptera (64%), Hymenoptera  
200 (16.1%), Lepidoptera (15.3%), Hemiptera (2.9%) and Coleoptera (1.3%) (Figure 2). The BIN  
201 system (Ratnasingham and Hebert 2013) linked the cumulative NGS sequences from the three  
202 Malaise sites to a total of 2,141 BINs, and their counts were similar for the three orchards (GAN  
203 = 1,795; QIU = 1,792; SHI = 1,712) (Table 1). These BINs were used as a proxy for species to  
204 analyze insect biodiversity assemblages at the collection sites. Alpha diversity analysis showed a  
205 similar Shannon index ( $H'$ ) value (GAN = 6.9; QIU = 6.9; SHI = 6.8) for the three Malaise sites  
206 (Figure 3). Biodiversity overlap determined by Jaccard similarity coefficient ( $J$ ) suggested a high  
207 level of BIN sharing among the three sites in pairwise comparisons ( $J = 0.73-0.75$ ) or in total ( $J$   
208 = 0.63) (Figure 3). At all three sampling sites, most BINs (GAN=769; QIU=739; SHI=751) were  
209 encountered only once in the 14/15 collection events and just a few (GAN=12; QIU=15;  
210 SHI=17) were detected in all samples. However, the variance ( $R^2$ ) of BIN occurrence in the  
211 collection events at the three sampling sites remained low (Figure 4).

212 When the sequences were compared against all insect records on BOLD, 40% (875) of the BINs  
213 showed a match to a known species while 59% (1229) were placed to a genus, 97% (2076) to a  
214 family, and 100% (2141) to an order. However, a check of the correspondence between the BINs  
215 and their associated taxonomy revealed 235 discordances at the species level, 98 at the genus  
216 level and 20 at the family level. In GAN, 220 discordant BINs were found at the species, 85 at  
217 the genus and 15 at the family. In QIU, 218 discordant BINs were found at the species, 87 at the  
218 genus and 18 in the family. In SHI, 216 discordant BINs were found at the species, 89 at the  
219 genus and 19 at the family (Table 1).

220

## 221 **Screening pest and beneficial insect species**

222 Most (96%) of the 2,141 BINs were linked to 14 insect orders, predominantly Diptera (49%),  
223 Hymenoptera (22%), Lepidoptera (14%), Coleoptera (5%), and Hemiptera (4.5%), and of them  
224 875 were linked to a known species. Among the 875 BINs that were identified to a known insect

225 species, 443 were pests, 223 pest/pollinators, 140 parasitoids, 52 predators, 2 predator/pollinators  
226 and 15 were pollinators (Figure 5, Table S2).

227

## 228 **Discussion**

229 Challenges of morphology and lack of taxonomic expertise have limited the understanding of  
230 insect pest and natural enemy complexes in citrus orchards, compromising the efficacy of pest  
231 management tactics (Niu et al. 2014). The present study circumvented these limitations by  
232 employing DNA metabarcoding to assess insect diversity in three citrus plantations. The  
233 coupling of Malaise traps with metabarcoding has been successfully used to develop inventory  
234 for insect faunas in Sweden (Karlsson et al. 2020) and Germany (Moriniere et al. 2016). Malaise  
235 traps are useful for capturing flying insects and have been frequently used for barcode-based  
236 insect diversity assessments (Hardulak et al. 2020; Karlsson et al. 2020). DNA metabarcoding  
237 offers simultaneous identification of multiple species in bulk samples (Yu et al. 2012). Prior  
238 studies have used metabarcoding to examine insect biodiversity, pest prevalence (Piper et al.  
239 2022), and to evaluate predator-prey relationships (Yang et al. 2022). For example, (Huang et al.  
240 2022) used metabarcoding to reveal the composition of Diptera communities in a subtropical  
241 system, while (Kirse et al. 2021) used this method to analyze seasonal shifts in arthropod  
242 diversity in Malaise trap catches. The effective implementation of metabarcoding for  
243 simultaneous, multi-species identification of complex mixed communities have helped scaling  
244 up pest surveillance efforts (Piper et al. 2019).

245 BINs were used as a proxy to count species and to link the generated sequences to species  
246 represented on BOLD. This approach not only revealed the presence of potentially more than  
247 2,000 insect species, but it also linked 866 (40%) of the BINs to known insects. Interestingly, the  
248 number of BINs revealed at each of the three sites were similar, but 25% of the BINs were  
249 unique to each site. This result supports the utility of BINs as an effective approach for counting  
250 species (Hebert et al. 2016) and assessing insect diversity (Telfer et al. 2015). While most BINs  
251 were assigned to single taxon, 12% were linked to more than one species or genus. The lower  
252 number of BINs with species matches on BOLD indicates the incompleteness of the current  
253 barcode reference library, emphasizing the need to develop regional DNA barcode reference

254 libraries. BIN discordance has been reported in many barcode studies (De Leon et al. 2020;  
255 Gibbs 2018) and the issue has generally been linked to misidentifications (Hebert et al. 2004),  
256 heteroplasmy, or incomplete lineage sorting (Kang et al. 2016; Weber et al. 2019). These issues  
257 can only be resolved by detailed taxonomic analysis. Nonetheless, use of BINs to identify or  
258 count species in pooled samples where sequenced specimens may not be validated by  
259 morphology has limitations (Elbrecht et al. 2017). To investigate the spread of BINs over time  
260 we calculated BIN incidences in the collection events. This analysis revealed that almost half of  
261 the BINs at each site were encountered only once in the total collection events at each site and  
262 just a fraction of them were detected in all samples. The points to the seasonality of the activity  
263 of various insect species (Wolda 1988).

264 Insects were classified as either pest or beneficial taxa based on the barcode sequence/BIN  
265 association with known species. The number of pest species detected was far higher than  
266 documented in the literature on citrus pests in China (Niu et al. 2014; Urbaneja et al. 2020). The  
267 present data provides a valuable resource for research in citrus pest management and monitoring.  
268 For example, the Asian citrus psyllid, *Diaphorina citri*, a devastating exotic pest of citrus in the  
269 US, was detected in samples collected from January to April, 2019. Its detection suggests the  
270 need for further monitoring targeted at this species to assist the timely implementation of control  
271 measures to reduce the risk of damage to the citrus production.

272 Numerous studies have been conducted to reveal the trophic relationships between predator and  
273 prey by detecting host DNA from the gut contents or feces of the predators using DNA  
274 barcoding (Furlong 2015; Galan et al. 2018; McClenaghan et al. 2019; Rytkonen et al. 2019;  
275 Verdasca et al. 2022). In our current study, we deployed the Malaise traps in the center of each  
276 citrus orchard, that reduced the possibility of adventitious predators. However, possibility of  
277 migration of insect predators from the neighboring crops to the sampling sites who ended up in  
278 the traps in search of their prey cannot be ruled out. Likewise, it is possible that some insects  
279 revealed in our study are not truly citrus pests instead they had been consumed by a predator or  
280 had moved to the gardens in search of nectar or to benefit from honeydew secreted by other  
281 arthropods. To minimize this deviation at most, we manually verified the identified insect  
282 species in the CPIS database and some published references described in materials and methods  
283 section, but contamination with non-target DNA originating from predator-prey interactions

284 (Eitzinger et al. 2019) or airborne eDNA from off-site insects cannot be ruled out (Roger et al.  
285 2022).

286 Biological control is the most environmentally safe and cost-effective pest management strategy.  
287 The present study revealed the presence of known species of Diptera and Hymenoptera which  
288 may have potential as biological control agents. For example, BINs linked to 32 species of  
289 Braconidae were encountered which belong to three genera (9 *Apanteles*, 14 *Cotesia*, 9  
290 *Microplitis*) that are potential biological control agents for Lepidoptera (Wharton 1993). In  
291 addition, syrphid flies (48 species) revealed by BIN-linkages also can be useful natural enemies  
292 of aphids (Nelson et al. 2012).

293 In summary, metabarcoding of bulk insect collections provides a cost- and time-effective way to  
294 assess insect communities in citrus orchards. However, the present study only examined one  
295 region. Constructing a comprehensive insect inventory for citrus orchards across China will  
296 require surveys at the representative sites in other citrus areas.

## 297 **Conclusions**

298 Our study analyzed the composition of insect communities in citrus orchards using DNA  
299 metabarcoding followed by data analysis on BOLD and mBRAVE. The species revealed by BIN  
300 matches on BOLD were then searched in citrus pest database and the literature to allow their  
301 classification into pest, parasitoid, predator, or pollinator. Our results provide a valuable resource  
302 for research on citrus pest management and beneficial insect exploration.

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305

## 306 **References**

- 307 Ashfaq M, Akhtar S, Rafi MA, Mansoor S, and Hebert PDN. 2017. Mapping global biodiversity  
308 connections with DNA barcodes: Lepidoptera of Pakistan. *PLoS One* 12:e0174749.  
309 10.1371/journal.pone.0174749  
310 Ashfaq M, Sabir JSM, El-Ansary HO, Perez K, Levesque-Beaudin V, Khan AM, Rasool A, Gallant C, Addesi  
311 J, and Hebert PDN. 2018. Insect diversity in the Saharo-Arabian region: Revealing a little-studied  
312 fauna by DNA barcoding. *PLoS One* 13:e0199965. 10.1371/journal.pone.0199965

- 313 Bani Baker Q, Hammad M, Al-Rashdan W, Jararweh Y, Al-Smadi M, and Al-Zinati M. 2020.  
314 Comprehensive comparison of cloud-based NGS data analysis and alignment tools. *Informatics*  
315 *in Medicine Unlocked* 18. 10.1016/j.imu.2020.100296
- 316 Beng KC, Tomlinson KW, Shen XH, Surget-Groba Y, Hughes AC, Corlett RT, and Slik JW. 2016. The utility  
317 of DNA metabarcoding for studying the response of arthropod diversity and composition to  
318 land-use change in the tropics. *Sci Rep* 6:24965. 10.1038/srep24965
- 319 Braukmann TWA, Ivanova NV, Prosser SWJ, Elbrecht V, Steinke D, Ratnasingham S, de Waard JR, Sones  
320 JE, Zakharov EV, and Hebert PDN. 2019. Metabarcoding a diverse arthropod mock community.  
321 *Mol Ecol Resour* 19:711-727. 10.1111/1755-0998.13008
- 322 Caserta R, Teixeira-Silva NS, Granato LM, Dorta SO, Rodrigues CM, Mitre LK, Yochikawa JTH, Fischer ER,  
323 Nascimento CA, Souza-Neto RR, Takita MA, Boscariol-Camargo RL, Machado MA, and De Souza  
324 AA. 2020. Citrus biotechnology: What has been done to improve disease resistance in such an  
325 important crop? *Biotechnology Research and Innovation*. 10.1016/j.biori.2019.12.004
- 326 CGA. 2020. Citrus Growers' Association of Southern Africa. Key industry statistics for citrus growers  
327 (Available from: [www.cga.co.za](http://www.cga.co.za)).
- 328 Comtet T, Sandionigi A, Viard F, and Casiraghi M. 2015. DNA (meta)barcoding of biological invasions: a  
329 powerful tool to elucidate invasion processes and help managing aliens. *Biological Invasions*  
330 17:905-922. 10.1007/s10530-015-0854-y
- 331 Cristescu ME. 2014. From barcoding single individuals to metabarcoding biological communities:  
332 towards an integrative approach to the study of global biodiversity. *Trends Ecol Evol* 29:566-571.  
333 10.1016/j.tree.2014.08.001
- 334 De Leon LF, Cornejo A, Gavilan RG, and Aguilar C. 2020. Hidden biodiversity in Neotropical streams: DNA  
335 barcoding uncovers high endemism of freshwater macroinvertebrates at small spatial scales.  
336 *PLoS One* 15:e0231683. 10.1371/journal.pone.0231683
- 337 deWaard JR, Levesque-Beaudin V, deWaard SL, Ivanova NV, McKeown JTA, Miskie R, Naik S, Perez KHJ,  
338 Ratnasingham S, Sobel CN, Sones JE, Steinke C, Telfer AC, Young AD, Young MR, Zakharov EV,  
339 and Hebert PDN. 2019. Expedited assessment of terrestrial arthropod diversity by coupling  
340 Malaise traps with DNA barcoding (1). *Genome* 62:85-95. 10.1139/gen-2018-0093
- 341 Eitzinger B, Abrego N, Gravel D, Huotari T, Vesterinen EJ, and Roslin T. 2019. Assessing changes in  
342 arthropod predator-prey interactions through DNA-based gut content analysis-variable  
343 environment, stable diet. *Mol Ecol* 28:266-280. 10.1111/mec.14872
- 344 Elbrecht V, and Steinke D. 2018. Scaling up DNA metabarcoding for freshwater macrozoobenthos  
345 monitoring. *Freshwater Biology*. 10.1111/fwb.13220
- 346 Elbrecht V, Vamos EE, Meissner K, Aroviita J, Leese F, and Yu D. 2017. Assessing strengths and  
347 weaknesses of DNA metabarcoding-based macroinvertebrate identification for routine stream  
348 monitoring. *Methods in Ecology and Evolution* 8:1265-1275. 10.1111/2041-210x.12789
- 349 Furlong MJ. 2015. Knowing your enemies: Integrating molecular and ecological methods to assess the  
350 impact of arthropod predators on crop pests. *Insect Sci* 22:6-19. 10.1111/1744-7917.12157
- 351 Galan M, Pons JB, Tournayre O, Pierre E, Leuchtman M, Pontier D, and Charbonnel N. 2018.  
352 Metabarcoding for the parallel identification of several hundred predators and their prey:  
353 Application to bat species diet analysis. *Mol Ecol Resour* 18:474-489. 10.1111/1755-0998.12749
- 354 Geiger MF, Moriniere J, Hausmann A, Haszprunar G, Wagele W, Hebert PD, and Rulik B. 2016. Testing  
355 the Global Malaise Trap Program - How well does the current barcode reference library identify  
356 flying insects in Germany? *Biodivers Data J*:e10671. 10.3897/BDJ.4.e10671
- 357 Gibbs J. 2018. DNA barcoding a nightmare taxon: assessing barcode index numbers and barcode gaps for  
358 sweat bees. *Genome* 61:21-31. 10.1139/gen-2017-0096

- 359 Gwiazdowski RA, Foottit RG, Maw HE, and Hebert PDN. 2015. The hemiptera (insecta) of Canada:  
360 constructing a reference library of DNA barcodes. *PLoS One* 10:e0125635.  
361 10.1371/journal.pone.0125635
- 362 Hajibabaei M, Shokralla S, Zhou X, Singer GA, and Baird DJ. 2011. Environmental barcoding: a next-  
363 generation sequencing approach for biomonitoring applications using river benthos. *PLoS One*  
364 6:e17497. 10.1371/journal.pone.0017497
- 365 Hardulak LA, Moriniere J, Hausmann A, Hendrich L, Schmidt S, Doczkal D, Muller J, Hebert PDN, and  
366 Haszprunar G. 2020. DNA metabarcoding for biodiversity monitoring in a national park:  
367 Screening for invasive and pest species. *Mol Ecol Resour* 20:1542-1557. 10.1111/1755-  
368 0998.13212
- 369 Hebert PDN, Cywinska A, Ball SL, and deWaard JR. 2003. Biological identifications through DNA  
370 barcodes. *Proc Biol Sci* 270:313-321. 10.1098/rspb.2002.2218
- 371 Hebert PDN, Penton EH, Burns JM, Janzen DH, and Hallwachs W. 2004. Ten species in one: DNA  
372 barcoding reveals cryptic species in the neotropical skipper butterfly *Astraptes fulgerator*.  
373 *Proceedings of the National Academy of Sciences of the United States of America* 101:14812-  
374 14817. 10.1073/pnas.0406166101
- 375 Hebert PDN, Ratnasingham S, Zakharov EV, Telfer AC, Levesque-Beaudin V, Milton MA, Pedersen S,  
376 Jannetta P, and deWaard JR. 2016. Counting animal species with DNA barcodes: Canadian  
377 insects. *Philos Trans R Soc Lond B Biol Sci* 371. 10.1098/rstb.2015.0333
- 378 Huang J, Miao X, Wang Q, Menzel F, Tang P, Yang D, Wu H, and Vogler AP. 2022. Metabarcoding reveals  
379 massive species diversity of Diptera in a subtropical ecosystem. *Ecol Evol* 12:e8535.  
380 10.1002/ece3.8535
- 381 Jaccard P. 1912. The Distribution of the Flora in the Alpine Zone.1. *New Phytologist* 11:37-50.  
382 10.1111/j.1469-8137.1912.tb05611.x
- 383 Ji Y, Ashton L, Pedley SM, Edwards DP, Tang Y, Nakamura A, Kitching R, Dolman PM, Woodcock P,  
384 Edwards FA, Larsen TH, Hsu WW, Benedick S, Hamer KC, Wilcove DS, Bruce C, Wang X, Levi T,  
385 Lott M, Emerson BC, and Yu DW. 2013. Reliable, verifiable and efficient monitoring of  
386 biodiversity via metabarcoding. *Ecol Lett* 16:1245-1257. 10.1111/ele.12162
- 387 Kang AR, Kim MJ, Park IA, Kim KY, and Kim I. 2016. Extent and divergence of heteroplasmy of the DNA  
388 barcoding region in *Anapodisma miramae* (Orthoptera: Acrididae). *Mitochondrial DNA A DNA*  
389 *Mapp Seq Anal* 27:3405-3414. 10.3109/19401736.2015.1022730
- 390 Karlsson D, Forshage M, Holston K, and Ronquist F. 2020. The data of the Swedish Malaise Trap Project,  
391 a countrywide inventory of Sweden's insect fauna. *Biodivers Data J* 8:e56286.  
392 10.3897/BDJ.8.e56286
- 393 Kirse A, Boursat SJ, Langen K, and Fonseca VG. 2021. Metabarcoding Malaise traps and soil eDNA reveals  
394 seasonal and local arthropod diversity shifts. *Sci Rep* 11:10498. 10.1038/s41598-021-89950-6
- 395 McClenaghan B, Nol E, and Kerr KCR. 2019. DNA metabarcoding reveals the broad and flexible diet of a  
396 declining aerial insectivore. *The Auk* 136. 10.1093/auk/uky003
- 397 Moriniere J, Cancian de Araujo B, Lam AW, Hausmann A, Balke M, Schmidt S, Hendrich L, Doczkal D,  
398 Fartmann B, Arvidsson S, and Haszprunar G. 2016. Species Identification in Malaise Trap  
399 Samples by DNA Barcoding Based on NGS Technologies and a Scoring Matrix. *PLoS One*  
400 11:e0155497. 10.1371/journal.pone.0155497
- 401 Nelson EH, Hogg BN, Mills NJ, and Daane KM. 2012. Syrphid flies suppress lettuce aphids. *BioControl*  
402 57:819-826. 10.1007/s10526-012-9457-z
- 403 Niu J-Z, Hull-Sanders H, Zhang Y-X, Lin J-Z, Dou W, and Wang J-J. 2014. Biological control of arthropod  
404 pests in citrus orchards in China. *Biological Control* 68:15-22. 10.1016/j.biocontrol.2013.06.005

- 405 Piper AM, Batovska J, Cogan NOI, Weiss J, Cunningham JP, Rodoni BC, and Blacket MJ. 2019. Prospects  
406 and challenges of implementing DNA metabarcoding for high-throughput insect surveillance.  
407 *Gigascience* 8. 10.1093/gigascience/giz092
- 408 Piper AM, Cunningham JP, Cogan NOI, and Blacket MJ. 2022. DNA Metabarcoding Enables High-  
409 Throughput Detection of Spotted Wing Drosophila (*Drosophila suzukii*) Within Unsorted Trap  
410 Catches. *Frontiers in Ecology and Evolution* 10. 10.3389/fevo.2022.822648
- 411 Prosser SW, deWaard JR, Miller SE, and Hebert PD. 2016. DNA barcodes from century-old type  
412 specimens using next-generation sequencing. *Mol Ecol Resour* 16:487-497. 10.1111/1755-  
413 0998.12474
- 414 Ratnasingham S, and Hebert PDN. 2013. A DNA-based registry for all animal species: the barcode index  
415 number (BIN) system. *PLoS One* 8:e66213. 10.1371/journal.pone.0066213
- 416 Ren J-M, Ashfaq M, Hu X-N, Ma J, Liang F, Hebert PDN, Lin L, Germain JF, and Ahmed MZ. 2017. Barcode  
417 index numbers expedite quarantine inspections and aid the interception of nonindigenous  
418 mealybugs (*Pseudococcidae*). *Biological Invasions* 20:449-460. 10.1007/s10530-017-1546-6
- 419 Roger F, Ghanavi HR, Danielsson N, Wahlberg N, Löndahl J, Pettersson LB, Andersson GKS, Boke Olén N,  
420 and Clough Y. 2022. Airborne environmental DNA metabarcoding for the monitoring of  
421 terrestrial insects—A proof of concept from the field. *Environmental DNA* 4:790-807.  
422 10.1002/edn3.290
- 423 Rytönen S, Vesterinen EJ, Westerduin C, Leviakangas T, Vatka E, Mutanen M, Valimäki P, Hukkanen M,  
424 Suokas M, and Orell M. 2019. From feces to data: A metabarcoding method for analyzing  
425 consumed and available prey in a bird-insect food web. *Ecol Evol* 9:631-639. 10.1002/ece3.4787
- 426 Shannon C. 1948. A mathematical theory of communication. *The Bell System Technical Journal*:379–423  
427 and 623–656.
- 428 Smaili MC, Boutaleb-Joutei A, and Blenzar A. 2020. Beneficial insect community of Moroccan citrus  
429 groves: assessment of their potential to enhance biocontrol services. *Egyptian Journal of*  
430 *Biological Pest Control* 30. 10.1186/s41938-020-00241-0
- 431 Taberlet P, Coissac E, Hajibabaei M, and Rieseberg LH. 2012. Environmental DNA. *Mol Ecol* 21:1789-  
432 1793. 10.1111/j.1365-294X.2012.05542.x
- 433 Telfer AC, Young MR, Quinn J, Perez K, Sobel CN, Sones JE, Levesque-Beaudin V, Derbyshire R,  
434 Fernandez-Triana J, Rougerie R, Thevanayagam A, Boskovic A, Borisenko AV, Cadel A, Brown A,  
435 Pages A, Castillo AH, Nicolai A, Glenn Mockford BM, Bukowski B, Wilson B, Trojahn B, Lacroix CA,  
436 Brimblecombe C, Hay C, Ho C, Steinke C, Warne CP, Garrido Cortes C, Engelking D, Wright D,  
437 Lijtmaer DA, Gascoigne D, Hernandez Martich D, Morningstar D, Neumann D, Steinke D, Marco  
438 DeBruin DD, Dobias D, Sears E, Richard E, Damstra E, Zakharov EV, Laberge F, Collins GE, Blagoev  
439 GA, Grainge G, Ansell G, Meredith G, Hogg I, McKeown J, Topan J, Bracey J, Guenther J, Sills-  
440 Gilligan J, Addesi J, Persi J, Layton KK, D'Souza K, Dorji K, Grundy K, Nghidinwa K, Ronnenberg K,  
441 Lee KM, Xie L, Lu L, Penev L, Gonzalez M, Rosati ME, Kekkonen M, Kuzmina M, Iskandar M,  
442 Mutanen M, Fatahi M, Pentinsaari M, Bauman M, Nikolova N, Ivanova NV, Jones N, Weerasuriya  
443 N, Monkhouse N, Lavinia PD, Jannetta P, Hanisch PE, McMullin RT, Ojeda Flores R, Mouttet R,  
444 Vender R, Labbee RN, Forsyth R, Lauder R, Dickson R, Kroft R, Miller SE, MacDonald S, Panthi S,  
445 Pedersen S, Sobek-Swant S, Naik S, Lipinskaya T, Eagalle T, Decaens T, Kosuth T, Braukmann T,  
446 Woodcock T, Roslin T, Zammit T, Campbell V, Dinca V, Peneva V, Hebert PD, and deWaard JR.  
447 2015. Biodiversity inventories in high gear: DNA barcoding facilitates a rapid biotic survey of a  
448 temperate nature reserve. *Biodivers Data J*:e6313. 10.3897/BDJ.3.e6313
- 449 Urbaneja A, Grout TG, Gravena S, Wu F, Cen Y, and Stansly PA. 2020. Chapter 16 - Citrus pests in a global  
450 world. In: Talon M, Caruso M, and Gmitter FG, eds. *The Genus Citrus*: Woodhead Publishing,  
451 333-348.

- 452 Verdasca MJ, Godinho R, Rocha RG, Portocarrero M, Carneiro LG, Rebelo R, and Rebelo H. 2022. A  
453 metabarcoding tool to detect predation of the honeybee *Apis mellifera* and other wild insects by  
454 the invasive *Vespa velutina*. *Journal of Pest Science* 95:997-1007. 10.1007/s10340-021-01401-3
- 455 Weber AA, Stohr S, and Chenuil A. 2019. Species delimitation in the presence of strong incomplete  
456 lineage sorting and hybridization: Lessons from *Ophioderma* (Ophiuroidea: Echinodermata). *Mol*  
457 *Phylogenet Evol* 131:138-148. 10.1016/j.ympev.2018.11.014
- 458 Wharton RA. 1993. Bionomics of the Braconidae. *Annual Review of Entomology* 38:121-143.  
459 10.1146/annurev.en.38.010193.001005
- 460 Wolda H. 1988. Insect Seasonality: Why? *Annual Review of Ecology and Systematics* 19:1-18.  
461 10.1146/annurev.es.19.110188.000245
- 462 Yang T, Song X, Zhong Y, Wang B, and Zhou C. 2022. Field investigation- and dietary metabarcoding-  
463 based screening of arthropods that prey on primary tea pests. *Ecol Evol* 12:e9060.  
464 10.1002/ece3.9060
- 465 Young RG, Milian-Garcia Y, Yu J, Bullas-Appleton E, and Hanner RH. 2021. Biosurveillance for invasive  
466 insect pest species using an environmental DNA metabarcoding approach and a high salt trap  
467 collection fluid. *Ecol Evol* 11:1558-1569. 10.1002/ece3.7113
- 468 Yu DW, Ji Y, Emerson BC, Wang X, Ye C, Yang C, and Ding Z. 2012. Biodiversity soup: metabarcoding of  
469 arthropods for rapid biodiversity assessment and biomonitoring. *Methods in Ecology and*  
470 *Evolution* 3:613-623. 10.1111/j.2041-210X.2012.00198.x
- 471 Zieritz A, Lee PS, Eng WWH, Lim SY, Sing KW, Chan WN, Loo JS, Mahadzir FN, Ng TH, Yeo DCJ, Gan LX,  
472 Gan JY, Gibbins C, Zoqratt MZHM, and Wilson JJ. 2022. DNA metabarcoding unravels unknown  
473 diversity and distribution patterns of tropical freshwater invertebrates. *Freshwater Biology*  
474 67:1411-1427. 10.1111/fwb.13926

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**Table 1** (on next page)

Insect BINs recovered from three Malaise traps in citrus orchards and their assignment and discordance at different taxonomic ranks.

1

2 **Table 1: Insect BINs recovered from three Malaise traps in citrus orchards and their**  
3 **assignment and discordance at different taxonomic ranks.**

	Location			
	GAN	QIU	SHI	All locations
Total BINs	1,795	1,782	1,712	2,141
BINs assigned to order	1,784	1,776	1,707	2,128
BINs assigned to family	1,744	1,739	1,664	2,076
BINs assigned to genus	1,053	1,034	1,019	1,229
BINs assigned to species	766	755	733	875
BINs with family-level discordance	15	18	19	20
BINs with genus-level discordance	85	87	89	98
BINs with species-level discordance	220	218	216	235

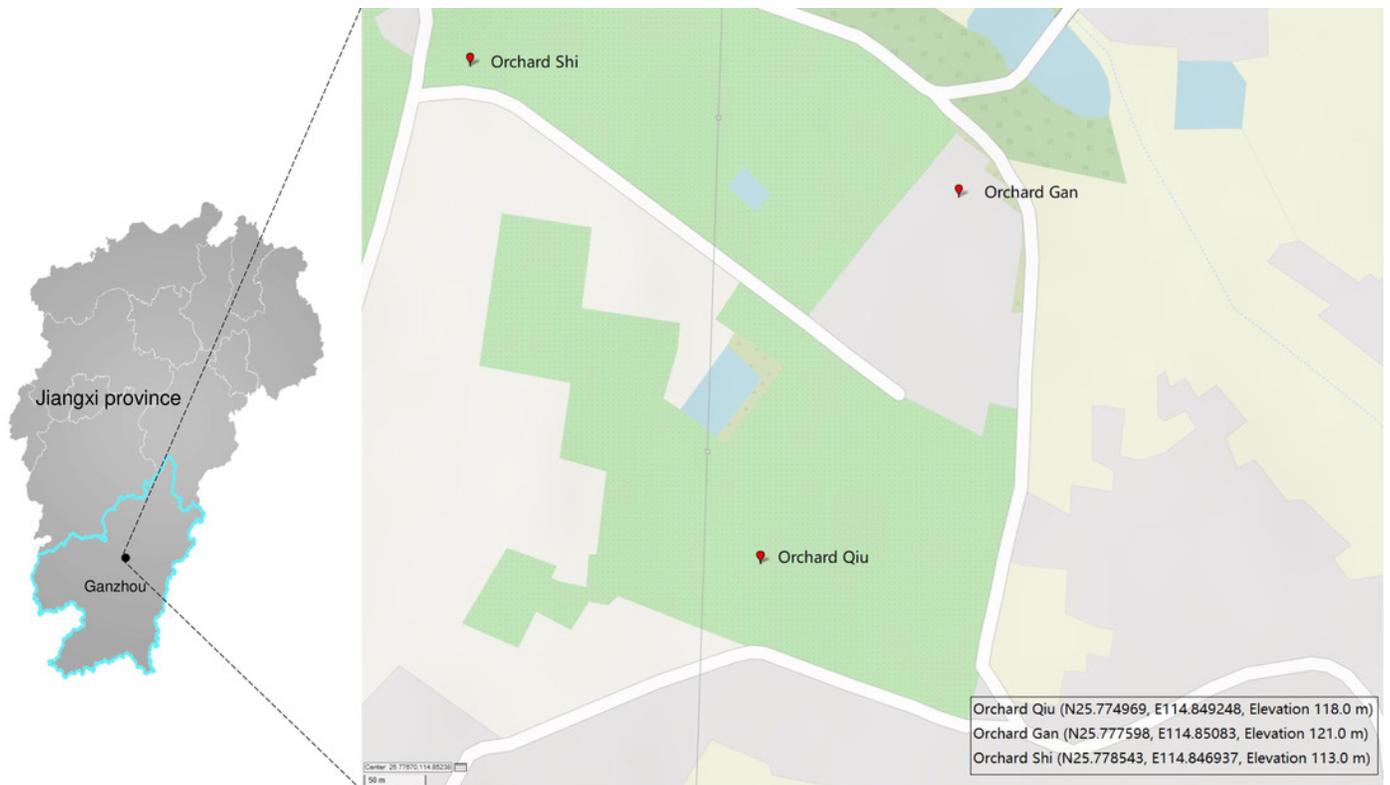
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5

# Figure 1

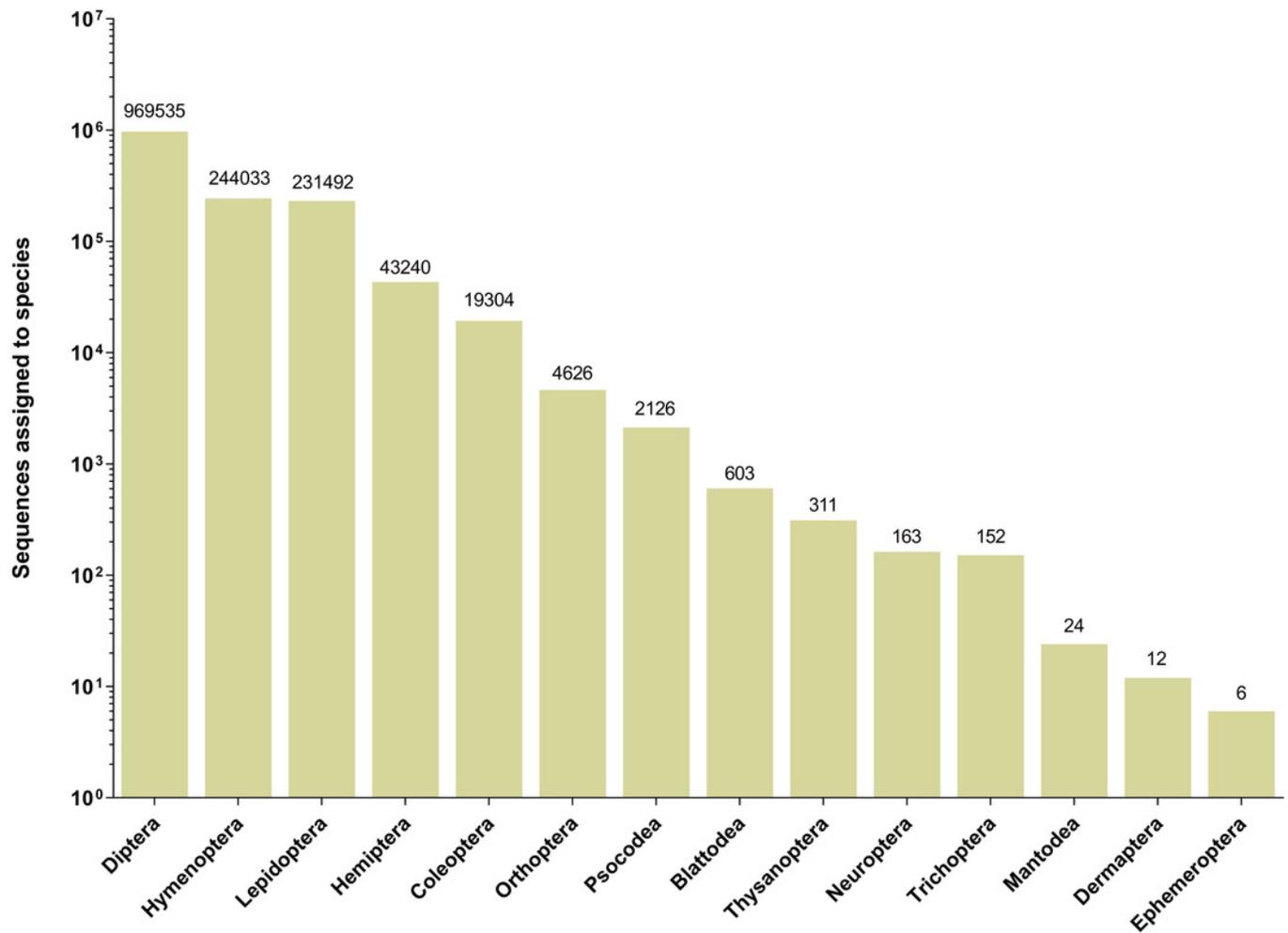
Three sites in Ganzhou (Jiangxi province, China) where Malaise traps were deployed.

The map was created with GPSVISUALIZER (<https://www.gpsvisualizer.com>).



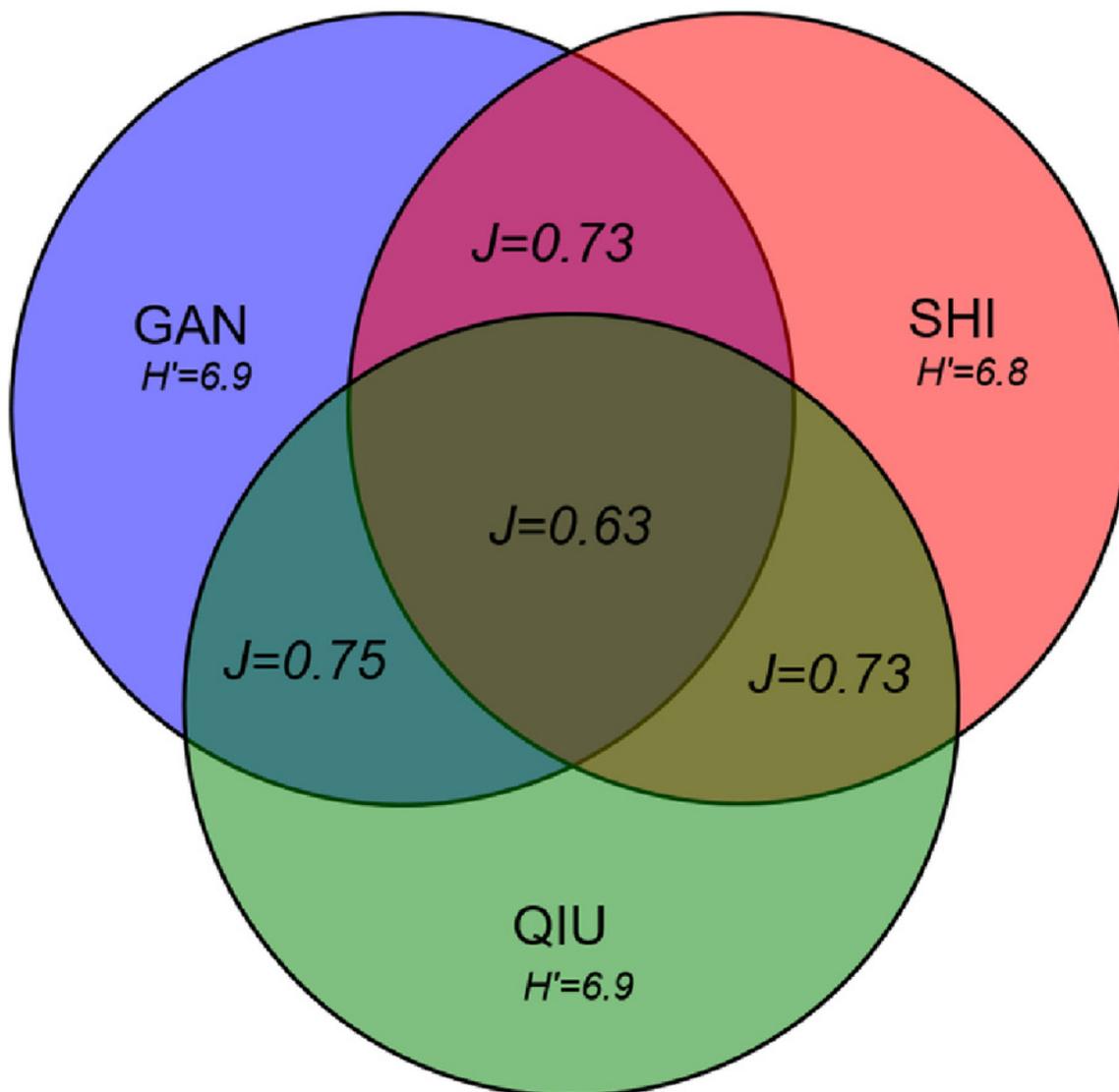
## Figure 2

Number of sequences assigned to 14 insect orders based on Malaise trap collections from three Chinese citrus orchards



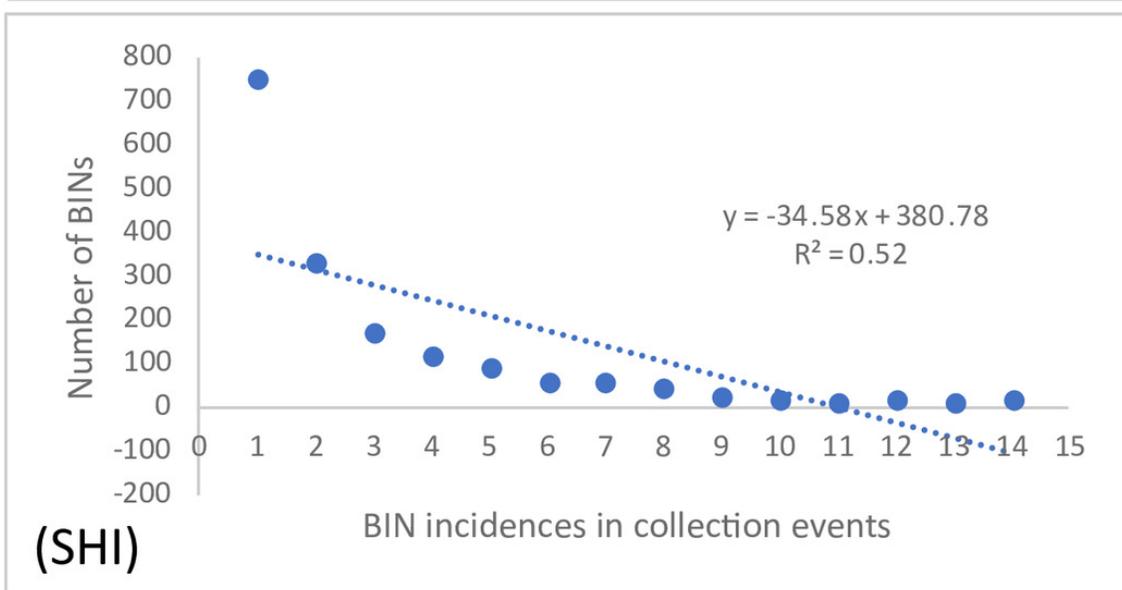
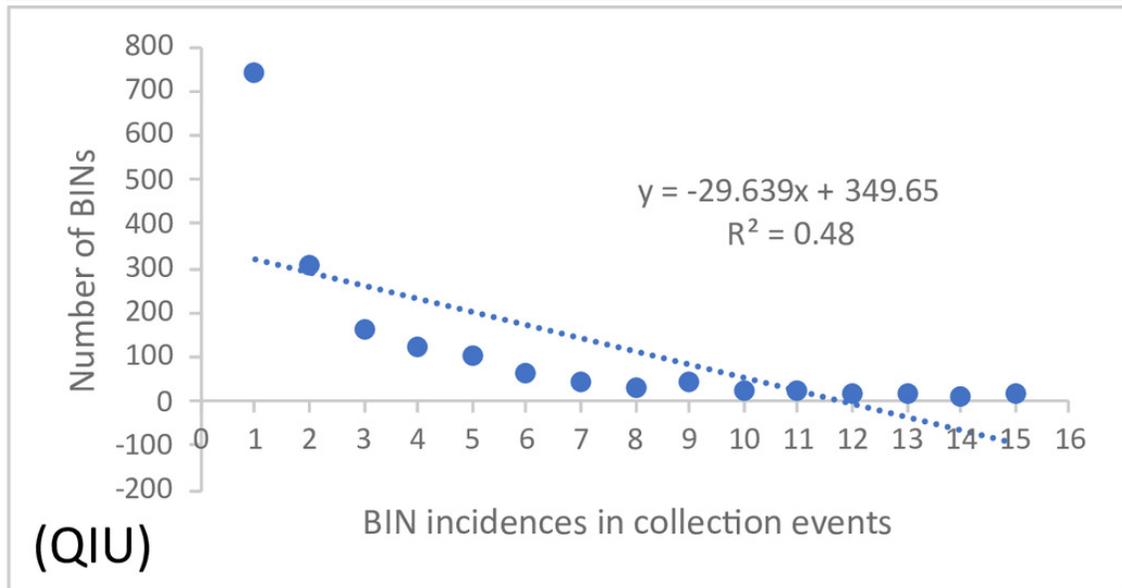
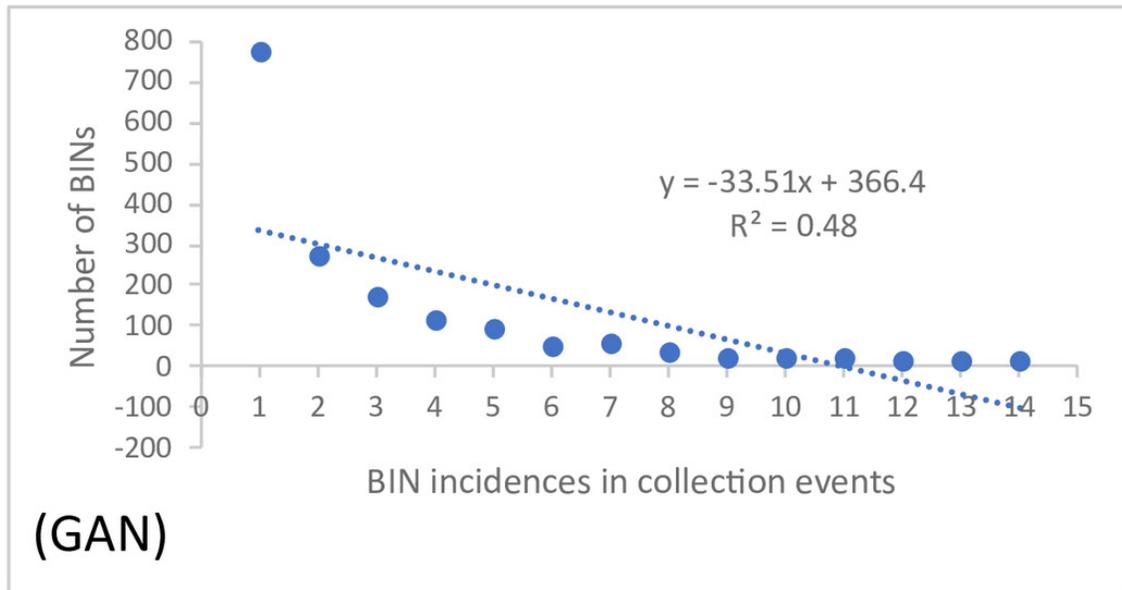
## Figure 3

Venn diagram depicting the alpha diversity (Shannon index,  $H'$ ) and the overlap (Jaccard index,  $J$ ) in BIN composition among insect collections from three citrus orchards.



## Figure 4

Incidence of BINs in the Malaise collections at three sampling sites, GAN, QIU and SHI.



## Figure 5

Number of BINs associated with pest and beneficial insect species.

