

Evaluation of DNA metabarcoding for identifying fish eggs: a case study on the West Florida Shelf (#81223)

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Evaluation of DNA metabarcoding for identifying fish eggs: a case study on the West Florida Shelf

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A critical factor in fisheries management is the protection of spawning sites for ecologically and economically important fish species. DNA barcoding (i.e., amplification and sequencing of the mitochondrial cytochrome c oxidase I (COI) gene) of fish eggs has emerged as a powerful technique for identifying spawning sites. However, DNA barcoding of individual fish eggs is time-consuming and expensive. In an attempt to reduce costs and effort for long-term fisheries monitoring programs, here we used DNA metabarcoding, in which DNA is extracted and amplified from a composited sample containing all the fish eggs collected at a given site, to identify fish eggs from 49 stations on the West Florida Shelf. A total of 37 taxa were recovered from 4,719 fish eggs. Egg distributions on the West Florida Shelf corresponded with the known habitat types occupied by these taxa, which included burrowers, coastal pelagic, epipelagic, mesopelagic, demersal, deep demersal, commensal, and reef-associated taxa. Metabarcoding of fish eggs was faster and far less expensive than barcoding individual eggs; however, this method cannot provide absolute taxon proportions due to variable copy numbers of mitochondrial DNA in different taxa, different numbers of cells within eggs depending on developmental stage, and PCR amplification biases. In addition, some samples yielded sequences from more taxa than the number of eggs present, demonstrating the presence of contaminating DNA and requiring the application of a threshold proportion of sequences required for counting a taxon as present. Finally, we review the advantages and disadvantages of using metabarcoding versus individual fish egg barcoding for long-term monitoring programs.

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15

16 Abstract

17

18 A critical factor in fisheries management is the protection of spawning sites for ecologically and
19 economically important fish species. DNA barcoding (i.e., amplification and sequencing of the
20 mitochondrial cytochrome c oxidase I (COI) gene) of fish eggs has emerged as a powerful
21 technique for identifying spawning sites. However, DNA barcoding of individual fish eggs is
22 time-consuming and expensive. In an attempt to reduce costs and effort for long-term fisheries
23 monitoring programs, here we used DNA metabarcoding, in which DNA is extracted and
24 amplified from a composited sample containing all the fish eggs collected at a given site, to
25 identify fish eggs from 49 stations on the West Florida Shelf. A total of 37 taxa were recovered
26 from 4,719 fish eggs. Egg distributions on the West Florida Shelf corresponded with the known
27 habitat types occupied by these taxa, which included burrowers, coastal pelagic, epipelagic,
28 mesopelagic, demersal, deep demersal, commensal, and reef-associated taxa. Metabarcoding of
29 fish eggs was faster and far less expensive than barcoding individual eggs; however, this method
30 cannot provide absolute taxon proportions due to variable copy numbers of mitochondrial DNA
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32 PCR amplification biases. In addition, some samples yielded sequences from more taxa than the
33 number of eggs present, demonstrating the presence of contaminating DNA and requiring the
34 application of a threshold proportion of sequences required for counting a taxon as present.
35 Finally, we review the advantages and disadvantages of using metabarcoding versus individual
36 fish egg barcoding for long-term monitoring programs.

37 Introduction

38

39 A critical factor in fisheries management is the protection of spawning sites for
40 ecologically and economically important fish species. Studies commonly hindcast spawning sites
41 based on the locations where larvae from a given species have been identified, but this method is
42 imprecise because larvae can be days, weeks, or even months old at the time of capture (Cowen
43 and Sponaugle 2009). In contrast, predicting spawning sites based on the presence of eggs is
44 much more reliable since eggs behave as relatively passive particles (the exception being the
45 eggs of species that are not neutrally buoyant), most fish remain in this developmental stage for a
46 maximum of 1-2 days, and the development stage of the eggs can be determined, allowing
47 identification of eggs that are less than a few hours old, if so desired. However, since fish eggs
48 are difficult to identify visually, the spawning locations of many broadcast spawning species
49 remain unknown (Kawakami et al. 2010, Becker et al. 2015). DNA barcoding (i.e., amplification
50 and sequencing of the mitochondrial cytochrome c oxidase I (COI) gene) of individual fish eggs
51 has emerged as a powerful technique for identification of fish spawning sites (Shao et al. 2002,
52 Saitoh et al. 2009, Lelièvre et al. 2012, Burghart et al. 2014, Frantine-Silva et al. 2015, Harada et
53 al. 2015, Lewis et al. 2016, Leyva-Cruz et al. 2016, Lin et al. 2016, Hofmann et al. 2017, Ahern
54 et al. 2018, Duke et al. 2018, Burrows et al. 2019, Hou et al. 2020, Kerr et al. 2020,
55 Mateos-Rivera et al. 2020, Hou et al. 2022).

56 Through several pilot studies and the long-term Spawning Habitat & Early-life Linkages
57 to Fisheries (SHELF) program funded by the Florida RESTORE Act Center of Excellence
58 Program (FLRACEP), we have used DNA barcoding to identify over 8,500 individual fish eggs
59 from over 320 locations in the Gulf of Mexico (GOM) and Florida Straits in the past decade
60 (Burghart et al. 2014, Burrows et al. 2019, Keel et al. 2022, Kerr et al. 2020, Kerr et al. 2022).
61 These data have provided tremendous insight into the spatial distribution of fish early life stages
62 in this region, provided the first known spawning grounds for several taxa (Kerr et al. 2020),
63 demonstrated a disparity between the composition of co-occurring egg and larval communities
64 (Burghart et al. 2014), identified distinct distributions of eggs from neritic versus oceanic taxa
65 with a community transition at the shelf break (Burrows et al. 2019), and documented the
66 potential of mesoscale cyclic eddies to entrain the eggs of reef-associated taxa and transport them
67 to deeper waters (Kerr et al. 2020). Building a long-term time series of fish egg community
68 composition is valuable for detecting changes in fish-egg community composition over time;
69 however, DNA barcoding of individual fish eggs is expensive and time-consuming. Two recent
70 studies have demonstrated the use of metabarcoding, in which DNA is extracted and amplified
71 from an aggregate sample containing all the fish eggs collected at a given site, to characterize the
72 spawning community (Duke and Burton 2020, Miranda-Chumacero et al. 2020). Here, we sought
73 to evaluate the performance of metabarcoding as a potential way to increase throughput and
74 reduce both financial and human resource costs for the purposes of supporting a long-term fish
75 egg monitoring program. We applied DNA metabarcoding to identify 4,719 fish eggs collected
76 from 49 samples on the West Florida Shelf, and recovered eggs from a total of 37 taxa.

77

78 **Materials & Methods**

79

80 *Sample collection*

81 We collected planktonic fish eggs from an a priori-defined grid on the West Florida Shelf
82 (Fig. 1A) aboard the R/V *Hogarth* on two cruises (August 6-16, 2019 and September 24-26,
83 2019; Supplemental Table 1). At each station, we performed a 15-minute tow at the ocean
84 surface with a bongo (double conical) plankton net (333 μm mesh, 61 cm mouth diameter)
85 equipped with plastic, 1-liter cod-ends, and a General Oceanics 2030R mechanical flowmeter.
86 After each tow, we washed down the nets with ambient seawater. The right-hand cod-ends were
87 drained of excess seawater using a sieve and rinsed back into the jar using 95% isopropanol,
88 leaving the final concentration >50% isopropanol. In the case of high biomass, we split the
89 samples into two cod-ends to allow proper alcohol preservation. We stored the samples at 4°C
90 until processing.

91 In the laboratory, we picked at least 100 percomorph eggs (except when less were
92 present) per sample using a stereomicroscope, gridded petri dishes, and fine-tipped forceps
93 during the month of December 2019. When fewer than 100 eggs were present, we processed the
94 entire sample. Each sample was separated into a labeled 1.5 mL screw cap, o-ring tube with 70%
95 isopropanol for genetic identification. The number of eggs processed for each sample ranged
96 from 2 – 272 (Supplemental Table 1).

97

98 *DNA extraction, PCR, sequencing*

99 To start the extraction process, we carefully removed the isopropanol with a sterile
100 pipette tip. Next, we rinsed the eggs twice with molecular grade PCR water to remove any
101 additional alcohol or other contaminants. To break open the chorion of the eggs, we added 0.4 g
102 of 1 mm beads to dry eggs along with 500 μl of HotSHOT alkaline lysis buffer (25 mM NaOH,
103 0.2 mM disodium EDTA, pH 12; Truett et al. 2000). We placed the tubes in a Fisher Scientific™
104 Bead Mill 4 Homogenizer for 5 minutes at 5 m/s and centrifuged briefly to reveal un-popped
105 eggs. We manually broke any un-popped eggs with sterile toothpicks. We then incubated the
106 tubes at 95°C for 30 minutes, vortexing occasionally, and put them on ice for 3 minutes. Finally,
107 we added 500 μl of HotSHOT neutralization buffer (40 mM Tris-HCl, pH 5; Truett et al. 2000)
108 to each tube before storage at -20°C until the Polymerase Chain Reaction (PCR) step.

109 We PCR amplified a 226 base pair (bp) region of the mitochondrial cytochrome c oxidase
110 I (COI) gene using primer set Mini_SH-E, with forward reads 5'-
111 CACGACGTTGTAAAACGACACYAAICAYAAAGAYATIGGCAC-3', and reverse read 5'-
112 GGATAACAATTTACACAGGCTTATRTTTRTTTATICGIGGAAIGC-3' (Shokralla et al.
113 2015). Each 50- μl PCR contained final concentrations of 1x Apex NH₄ buffer, 1.5 mM Apex
114 MgCl₂, 0.2 μM Apex dNTPs, 1 U Apex RedTaq (Genesee Scientific), 0.2 μM forward and
115 reverse primers, 10 $\mu\text{g}/\mu\text{l}$ bovine serum albumin (New England BioLabs Inc.), and 2 μl of target
116 DNA (Burrows et al. 2019). The thermocycler conditions were as follows, 95°C for 5 min,

117 followed by 35 cycles of (94°C for 40 s, 46°C for 1 min, 72°C for 30 s) and 72°C for 5 min. We
118 verified successful PCR amplification by running products on a 1.5% agarose gel stained with
119 ethidium bromide. All samples were deemed successful and cleaned with the Zymo DNA Clean
120 and Concentrator -25 Kit if the bands were bright, and the Zymo DNA Clean and Concentrator -
121 5 Kit if the bands were faint. We quantified the amplified DNA using a Qubit™ dsDNA HS
122 Assay Kit, normalized the samples to equal concentrations, and sent to Genewiz for next-
123 generation Illumina sequencing using the Genewiz Amplicon-EZ pipeline and partial Illumina
124 adapters, forward 5'-ACACTCTTCCCTACACGACGCTCTTCCGATCT-3', reverse 5'-
125 GACTGGAGTTCAGACGTGTGCTCTTCCGATCT-3' to genetically identify the fish eggs
126 present in each sample.

127

128 *Sequence analysis*

129 To process the raw sequence data and obtain filtered and trimmed amplicon sequence
130 variants (ASV), we used the Divisive Amplicon Denoising Algorithm (DADA2) v1.12 package
131 (Callahan et al. 2016) in the *R* statistical environment (R Core Team 2022). ASVs were first
132 matched with species-level records in the Barcode of Life Database (BOLD;
133 <http://www.boldsystems.org/>)(Sujeevan and Hebert 2007), then BLASTn comparison (Altschul
134 et al. 1990) against the National Center for Biotechnology Information (NCBI) nucleotide
135 database (<https://blast.ncbi.nlm.nih.gov/Blast.cgi>) if no match was made in BOLD. For some
136 samples, more taxa were identified than the number of eggs present, signaling the presence of
137 false positives and requiring the establishment of a threshold percentage of sequences required to
138 consider a taxon present in a given sample (see Discussion). We applied a 2% threshold based on
139 total number of eggs within each sample and number of taxa assigned. This empirically derived
140 threshold is based on the principle that it would not be possible for a sample containing *n* number
141 of eggs to have greater than *n* taxa present unless contamination was present. The final ASV
142 table containing sequences that comprised >2% of the total sequence reads from any given
143 sample is available in GRIIDC (Kerr et al 2023).

144

145 *Quality control and data visualization*

146 We identified sequences to the finest taxonomic resolution possible based on comparison
147 to BOLD. In some cases, we could not distinguish between multiple potential identifications
148 based on the sequenced portion of the COI gene. Therefore, to refine our identifications we
149 referenced the geographic distribution of each taxon using published guides (McEachran and
150 Fechhelm 1998, McEachran and Fechhelm 2005) and FishBase (Froese and Pauly 2019). We
151 excluded any taxa not found in the Gulf of Mexico, such as those with Indo-Pacific or eastern
152 Atlantic distributions. Twenty-one of the identifications were made at the species-level, while the
153 remaining 16 identifications were to two or more closely related fishes; each distinct
154 identification is referred to as a “taxon”.

155 We also generated spatial heatmaps to visualize how metabarcoding-based identifications
156 compared with the known habitat types occupied by each taxon (Fig. 1). Due to the qualitative

157 nature of metabarcoding identifications, we used a presence-based approach. We first
158 categorized each of the 37 distinct taxa into one of eight habitat types, which included burrower,
159 commensal, reef-associated, demersal, deep demersal, coastal pelagic, epipelagic, and
160 mesopelagic, based on information from FishBase (Froese and Pauly 2019). We then quantified
161 the number of distinct taxa in each habitat type at each sampling station. These presence-absence
162 count data were then used to generate heatmaps in QGIS (QGIS Development Team 2022)
163 where spatial weighting was set to 55,000 km (roughly equivalent to the distance between any
164 two stations).

165

166 Results

167

168 We performed DNA metabarcoding on 4719 fish eggs from 49 samples, and obtained an
169 average of 57,185 sequence reads per sample. The analyses presented here consider any taxa
170 comprising $\geq 2\%$ the sequence reads from a given sample to be “present” in the sample and any
171 sequences comprising $< 2\%$ the sequence reads from a given sample as “absent” since
172 metabarcoding data are not quantitative due to methodological biases (see Discussion). We
173 identified a total of 37 distinct taxa, with 1-11 distinct taxa per sample (Table 1; Supplemental
174 Table 1). Twelve of the taxa (32%) were only detected in a single sample, eight taxa (22%) were
175 detected in two samples, and the remainder were present in three or more samples. Tuna eggs
176 (sequences could not be distinguished between *Auxis thazard/rochei*, *Euthynnus alletteratus*, and
177 *Katsuwonus pelamis*) were identified in 26 samples, constituting the most widespread taxon in
178 the dataset. Other taxa found at more than 10 sites included *Decapterus punctatus/tabl*
179 (round/roughear scad), *Lutjanus griseus* (grey snapper), *Prionotus martis* (Gulf of Mexico barred
180 searobin), *Pristipomoides aquilonaris* (wenchman), and *Xyrichtys sp.* (razorfish).

181 To validate the metabarcoding-based identification of fish eggs, we constructed spatial
182 heatmaps based on presence-absence data (Fig. 1A-I), which were consistent with the expected
183 distribution of the identified taxa. Specifically, reef-associated species were found broadly
184 throughout the sampling area (Fig. 1D), corresponding with the patchy distribution of structured
185 hard-bottom throughout the West Florida Shelf (Hine and Locker 2011). Coastal pelagic taxa
186 were found inshore (Fig. 1G), epipelagic taxa were broadly distributed throughout the region
187 (Fig. 1H), and mesopelagic taxa were found in deeper water farther offshore (Fig. 1I). Notably,
188 we observed an apparent “hotspot” of demersal taxa (Fig. 1E) toward the southern end of the
189 sampling region, which may indicate an area of interest for the management of recreationally or
190 commercially important demersal fishes.

191

192 Discussion

193 DNA barcoding has gained popularity for identifying fish eggs; however, the vast
194 majority of studies analyze individual fish eggs. Processing individual fish eggs yields
195 quantitative data; that is, we directly determine the exact proportion of the collected fish egg
196 community comprised by each taxon. Information on the abundance of planktonic fish eggs from

197 each taxon is valuable for estimating the biomass of parent fish stocks using the daily egg
198 production method (Stratoudakis et al. 2006, Burrows et al. 2019). However, the number of eggs
199 and sampling stations that can be processed with this method is limited by financial and labor
200 resources. Metabarcoding of DNA extracted from all the collected fish eggs present at a given
201 sample could be an advantageous alternative since it is faster and less expensive (Cristescu
202 2014). We are only aware of two studies to date that have applied metabarcoding to examine the
203 community composition of fish eggs, one in marine waters and one in freshwater (Duke and
204 Burton 2020, Miranda-Chumacero et al. 2020). Here we add to this emerging field of study by
205 assessing the potential of metabarcoding as an alternate method for long-term monitoring of fish
206 egg community composition. The advantages and disadvantages of DNA barcoding individual
207 eggs versus the metabarcoding method applied here are summarized in Table 2 and discussed
208 below.

209 Compared to DNA barcoding of individual fish eggs, in which a single sequence is
210 obtained from each egg, metabarcoding is not quantitative due to variable copy number of
211 mitochondrial DNA in different taxa, different numbers of cells within eggs depending on
212 developmental stage, chimeric sequences, and PCR amplification biases (Bik et al. 2012,
213 Hatzenbuehler et al. 2017, Duke and Burton 2020). A ground-truthing study by Duke and Burton
214 (2020) demonstrated that fish egg metabarcoding reliably detected taxa that comprised over six
215 percent of a mock community and three percent of a natural community, with variable recovery
216 of rarer community members. Despite some variability, Duke and Burton (2020) found a positive
217 relationship between the proportion of reads from a given taxon and the proportion of eggs from
218 that taxon in the mock communities. Therefore, although metabarcoding data cannot provide
219 absolute taxon proportions, this technique can yield valuable information about abundant taxa, as
220 well as rarer taxa above a given threshold.

221 Another potential flaw with DNA metabarcoding of fish eggs is the detection of false
222 positives either due to environmental DNA (eDNA) stuck on the fish egg surfaces or
223 contamination introduced during processing (Fritts et al. 2019, Duke and Burton 2020). Duke
224 and Burton (2020) found that most false positives comprised a very small percentage of the
225 sequences recovered from a given sample. These data and other studies of fish early life stages
226 suggest setting a threshold proportion of sequences required for counting a taxon as present;
227 however, there is no consensus on what that threshold value should be and it may need to be
228 specific to each study area (Mariac et al. 2018, Duke and Burton 2020, Miranda-Chumacero et
229 al. 2020). In the present study, we empirically derived a threshold based on the total number of
230 eggs within each sample and the number of taxa assigned, based on the principle that it would
231 not be possible for a sample containing n number of eggs to have greater than n taxa unless
232 contamination was present. We found that setting a 2% threshold (i.e., sequences comprising less
233 than 2% of the total sample reads were considered false positives and removed) ensured that the
234 maximum number of taxa never exceeded the maximum number of eggs in a sample.

235 Unlike individual egg DNA barcoding where a single sequence is recovered from each
236 egg and the results are quantitative, we found that the metabarcoding results varied depending on

237 the threshold applied, **thus** false positives (i.e., sequences above the threshold that were not
238 derived from eggs) and false negatives (i.e., sequences below the threshold that were derived
239 from eggs but comprised a percentage of the sequences below the cutoff) remain problematic and
240 can have a major effect on reported spawning sites. For example, the 2% threshold used to
241 analyze the data in this study resulted in an average of 3.67 taxa per sample (range 1-11).
242 Applying a more conservative 5% threshold would have resulted in an average of 2.18 taxa per
243 sample (range 1-6). Although 32 of the 37 taxa identified in this study would still be detected in
244 at least one sample with the 5% threshold, five taxa (*Acanthocybium solandri*/*Scomberomorus*
245 *cavalla*, *Engraulis eurystole*, *Haemulon plumieri*, *Scomberomorus maculatus/regalis*, *Synagrops*
246 *bellus/spinosus*) would have been removed completely. Eggs from all five of these taxa have
247 been previously recovered on the West Florida Shelf through individual egg barcoding (Keel et
248 al. 2022, Kerr et al. 2022) **so** in this case, we believe that increasing the threshold would likely
249 result in false negatives. This example demonstrates the large effect that small differences in the
250 threshold can have on DNA metabarcoding results.

251 Since the inception of the Fish Barcode of Life (FISH-BOL) Initiative, the COI gene is
252 widely used for genetic identification of fishes and this gene is capable of distinguishing between
253 the majority of described fish species (Teletchea 2009, Ward et al. 2009). Numerous studies have
254 demonstrated the advantage of using multiple genetic markers for metabarcoding (Evans et al.
255 2016, Sawaya et al. 2019, Duke and Burton 2020); however, databases tend to be more limited
256 for other markers and may need to be supplemented for geographic regions of interest. One
257 advantage of individual egg barcoding is the ability to return to DNA samples extracted from
258 specific eggs to analyze population genetics or apply additional primer sets in the case where the
259 conserved region of the COI gene used for barcoding cannot distinguish between certain species
260 complexes. For example, Burrows et al. (2019) applied additional PCR primers to distinguish
261 between the economically important species *Thunnus thynnus* and *Katsuwonus pelamis*, as well
262 as between *Scomberomorus cavalla* and *Acanthocybium solandri*, to achieve a definitive
263 identification. The ability to return to specific eggs to refine taxonomic uncertainties is lost in
264 DNA metabarcoding, where the DNA from all eggs within a sample is combined. Finally, it is
265 possible that the shorter sequence length used for Illumina sequencing in metabarcoding
266 compared to Sanger sequencing used for individual egg barcoding might hinder the assignment
267 of sequences to species level. We did not experience lower taxonomic resolution due to the
268 shorter sequence length obtained in this study, which is consistent with other analyses that have
269 shown the ability to reliably assign 140 bp reads with relatively high success rates (Shokralla et
270 al. 2015, Kimmerling et al. 2018, Mariac et al. 2018).

271 PCR biases represent an important barrier to the feasibility of metabarcoding studies for
272 quantitative analyses of fish early life stages (Lamb et al. 2019, Zinger et al. 2019). Efforts to
273 make metabarcoding of fish larvae more quantitative have shown success, although these
274 methods have not yet been applied to eggs and have only been examined in a limited number of
275 studies and regions. Applying a different approach to quantify fish early life stages, Kimmerling
276 et al. (2018) used high coverage metagenomic sequencing (sequencing of total DNA from a

277 given sample, without first applying PCR to enrich for the COI gene). Kimmerling et al. (2018)
278 adjusted the sequence coverage for each sample to obtain ~20 COI-derived reads per larva,
279 allowing samples with more larvae to be sequenced more deeply. These methods yielded
280 quantitative results when normalized by the relative size of each larva in the sample, showing the
281 promise of this technique (Kimmerling et al. 2018). However, this method requires a large
282 amount of sequencing since the percentage of metagenomic sequence reads that belonged to the
283 COI gene was extremely low (approximately 1 in every 18,000 sequences), which will present a
284 substantial barrier for long-term monitoring efforts. With a newly developed Metabarcoding by
285 Capture using a Single Probe (MCSP) method, Mariac et al. (2018) achieved ~6,000 times
286 enrichment of COI sequences compared to an unenriched sample. By analyzing a mock
287 community, the relative frequencies of sequences recovered from larval swarms in the Amazon
288 basin with the MCSP method correlated extremely well with true frequencies derived from
289 Sanger sequencing of individual fish larvae (Mariac et al. 2018). Since this method relies on
290 hybridization instead of PCR amplification, MCSP is subject to fewer biases. However, it should
291 be noted that MCSP still required the application of a threshold value for the minimum number
292 of reads per taxon in order to count that taxon as present. Even with the application of an
293 empirically defined threshold established through the analysis of mock communities, a small
294 number of false positives were still encountered (Mariac et al. 2018).

295 Finally, although we cannot directly compare the metabarcoding results presented here to
296 our prior surveys of individual fish eggs in the GOM, it is notable that the vast majority of the
297 taxa identified here have also been recovered from this region in our prior work (Keel et al.
298 2022, Kerr et al. 2022). The only exceptions found with metabarcoding that we have not
299 observed with our more spatially and temporally expansive individual egg barcoding were
300 *Makaira nigricans* (Atlantic blue marlin), *Lepidopus altifrons* (crested scabbardfish), and
301 *Chilomycterus sp.* (burrfish), all of which are known to occur in the GOM. We examined the
302 data for these taxa to determine if they were present in very low abundances and thus were likely
303 false positives; however, that was not the case. *Chilomycterus sp.* comprised ~87% of the
304 sequences in one sample, *Lepidopus altifrons* comprised ~20% of the sequences in another
305 sample, and *Makaira nigricans* was found in two samples, where it made up ~16% of the
306 sequences of each sample. This suggests that eggs from these taxa were truly present in these
307 samples, and demonstrates an advantage of metabarcoding; namely, by enabling the processing
308 of samples from more stations, we can capture rarer or more episodic spawning events.

309

310 **Conclusions**

311

312 In this study, we assessed the performance of DNA metabarcoding to increase throughput
313 and reduce financial and labor costs associated with a long-term fish egg monitoring program. A
314 total of 37 taxa were identified from 49 stations on the West Florida Shelf. Egg identifications
315 were consistent with prior species distributions observed from individual egg DNA barcoding
316 and spatial heatmaps of eggs corresponded to known habitat types occupied by adults. The

317 increased throughput allowed by metabarcoding resulted in the identification of taxa not
318 previously detected in this region, possibly representing episodic spawning events. One
319 disadvantage of metabarcoding is that this method is not quantitative and requires the application
320 of a threshold proportion of sequences required to count a taxon as present. The choice of DNA
321 barcoding methods therefore depends on the goals of the study, and fisheries monitoring efforts
322 may benefit from a combination of the two approaches, with individual egg barcoding providing
323 quantitative information and metabarcoding expanding the number or geographic range of
324 samples that can be processed.

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328

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332 the plankton samples, and to Natalie Sawaya for guidance with bioinformatics.

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335 **Tables**

336

337 **Table 1.** Taxa comprising more than 2% of the sequences from any sample, habitat
338 classification, and the number of stations where the taxa were identified.

339

340 **Table 2.** Comparison of individual egg DNA barcoding versus metabarcoding.

341

342 **Figures**

343

344 **Figure 1.** West Florida Shelf study region. (A) Station locations were defined a priori, and
345 heatmaps were based on the number of taxa identified at each station for each habitat type: (B)
346 Burrower, (C) Commensal, (D) Reef-Associated, (E) Demersal, (F) Deep Demersal, (G) Coastal
347 Pelagic, (H) Epipelagic, and (I) Mesopelagic.

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350 **Supplemental Files**

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352 **Supplemental Data S1.** Detailed table of taxa identified in each sample, including scientific
353 name, common name, and habitat type. Metadata for each sample include the station number,
354 latitude and longitude, maximum water depth, date collected, and number of eggs processed.

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Table 1 (on next page)

Taxa comprising more than 2% of the sequences from any sample, habitat classification, and the number of stations where the taxa were identified.

1 **Table 1:** Taxa comprising more than 2% of the sequences from any sample, habitat
 2 classification, and the number of stations where the taxa were identified.
 3

| Scientific Name | Common Name | Habitat | # Stations |
|---|--|-----------------|------------|
| <i>Acanthocybium solandri</i> , <i>Scomberomorus cavalla</i> | Wahoo/King Mackerel | Epipelagic | 3 |
| <i>Acanthostracion quadricornis</i> | Scrawled Cowfish | Reef | 6 |
| <i>Auxis thazard/rochei</i> , <i>Euthynnus alletteratus</i> , <i>Katsuwonus pelamis</i> | Bullet/Frigate Tuna, Little Tunny, Skipjack | Epipelagic | 26 |
| <i>Brama dussumieri/caribbea</i> | Lesser Bream/Carribbean Pomfret | Deep demersal | 4 |
| <i>Callechelys muraena</i> | Blotched Snake Eel | Burrower | 1 |
| <i>Caranx crysos</i> | Blue Runner | Coastal pelagic | 5 |
| <i>Chaetodipterus faber</i> | Atlantic Spadefish | Reef | 1 |
| <i>Chilomycterus schoepfii/antillarum</i> | Striped/Web Burrfish | Reef | 1 |
| <i>Coryphaena hippurus</i> | Common Dolphinfish | Epipelagic | 2 |
| <i>Cyclopsetta fimbriata</i> | Spotfin Flounder | Demersal | 3 |
| <i>Decapterus punctatus/tabl</i> | Round/Roughear Scad | Coastal pelagic | 16 |
| <i>Diplogrammus pauciradiatus</i> | Spotted Dragonet | Reef | 1 |
| <i>Diplospinus multistriatus/Eustomias polyaster</i> | Striped escolar/Dragonfish | Mesopelagic | 5 |
| <i>Echeneis naucrates/neucratoides</i> , <i>Remora remora/osteochir</i> | Live/Whitefin Sharksucker, Marlin/Sharksucker | Commensal | 8 |
| <i>Engraulis eurystole</i> | European/Silver anchovy | Coastal pelagic | 1 |
| <i>Gordiichthys irretitus</i> | Horsehair Eel | Burrower | 1 |
| <i>Haemulon plumieri</i> | White Grunt | Reef | 1 |
| <i>Lepidopus altifrons</i> | Crested Scabbardfish | Mesopelagic | 1 |
| <i>Lutjanus campechanus</i> | Red Snapper | Reef | 2 |
| <i>Lutjanus griseus</i> | Grey Snapper | Reef | 11 |
| <i>Makaira nigricans</i> | Blue Marlin | Epipelagic | 2 |
| <i>Oxyporhamphus similis/micropterus</i> | Halfbeaks | Epipelagic | 4 |
| <i>Prionotus martis</i> | Gulf of Mexico Barred Searobin | Demersal | 11 |
| <i>Prionotus ophryas/scitulus</i> | Bandtail/Leopard Searobin | Demersal | 2 |
| <i>Prionotus roseus</i> | Bluespotted Searobin | Demersal | 2 |
| <i>Prionotus rubio/tribulus</i> | Blackwing/Bighead Searobin | Demersal | 8 |
| <i>Pristipomoides aquilonaris</i> | Wenchman | Reef | 18 |

| | | | |
|--|---|-----------------|----|
| <i>Prognichthys occidentalis</i> | Bluntnose Flyingfish | Epipelagic | 2 |
| <i>Rachycentron canadum</i> | Cobia | Reef | 1 |
| <i>Rhomboplites aurorubens</i> | Vermilion Snapper | Reef | 2 |
| <i>Saurida normani/brasiliensis</i> | Shortjaw/Brazilian Lizardfish | Demersal | 2 |
| <i>Scomberomorus maculatus/regalis</i> | Atlantic Spanish Mackerel/Cero | Coastal pelagic | 1 |
| <i>Synagrops bellus/spinosus</i> | Blackmouth/Keelcheek Bass | Mesopelagic | 1 |
| <i>Synodus intermedius/foetens/macrostigmus/sp</i> | Sand Diver/Inshore/ Largespot/Lizardfish sp. | Demersal | 5 |
| <i>Thunnus atlanticus/albacares/sp</i> | Blackfin/Yellowfin/Tuna sp. | Epipelagic | 8 |
| <i>Trachinocephalus myops</i> | Snakefish | Demersal | 1 |
| <i>Xyrichtys novacula/sp</i> | Pearly Razorfish/Razorfish sp. | Burrower | 11 |

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Table 2 (on next page)

Comparison of individual egg DNA barcoding versus metabarcoding.

1 **Table 2:** Comparison of individual egg DNA barcoding versus metabarcoding.

| Parameter | Individual eggs | Metabarcoding |
|--|--|---|
| Cost | \$5.15 per egg \$494.40 per site ^a | \$0.78 per egg ^a \$64.82 per site |
| Sequencing platform | Sanger | Illumina |
| Average sequence length | 500 base pairs | 200 base pairs |
| Quantitative | Yes | No |
| Ability to return to individual eggs with additional primers | Yes | No |
| Prevalence of false positives/negatives | Low/None | Frequent; dependent on the application of a threshold |

2 ^a Cost calculated based on 96 eggs per site.

3

Figure 1

West Florida Shelf study region and heatmap of identified taxa

(A) Station locations were defined a priori, and heatmaps were based on the number of taxa identified at each station for each habitat type: (B) Burrower, (C) Commensal, (D) Reef-Associated, (E) Demersal, (F) Deep Demersal, (G) Coastal Pelagic, (H) Epipelagic, and (I) Mesopelagic.

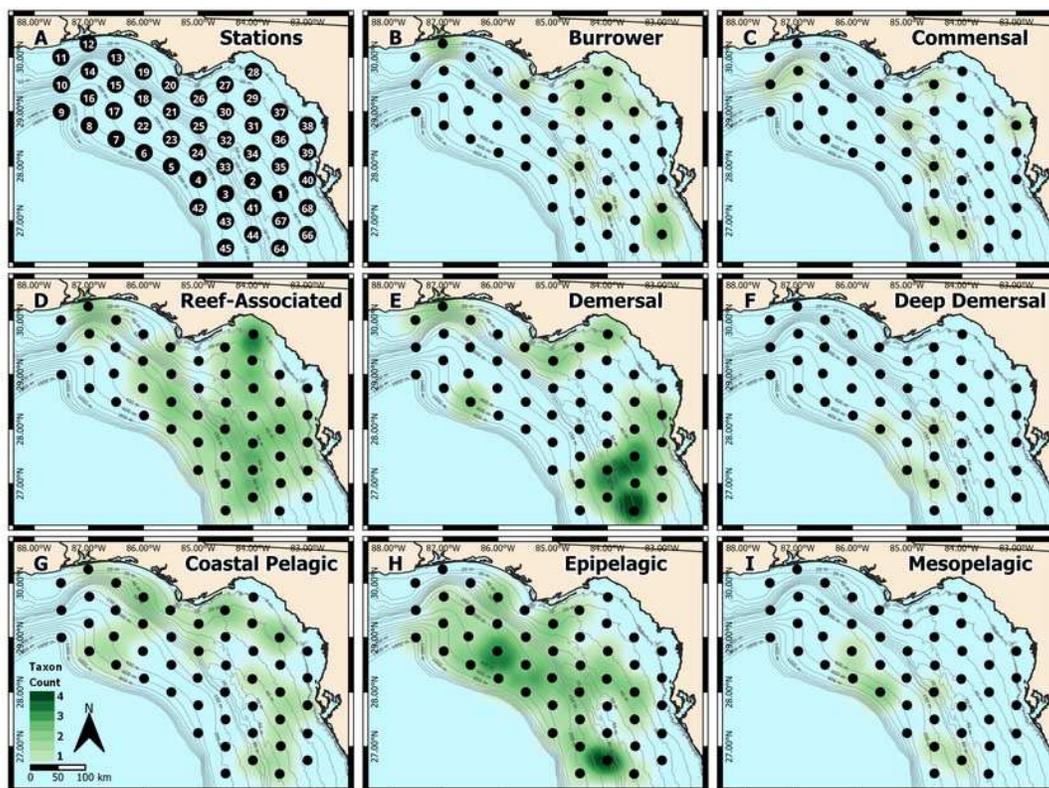


Figure 1: West Florida Shelf study region. Station locations were defined a priori (A), and heatmaps were based on the number of taxa identified at each station for each habitat type: Burrower (B), Commensal (C), Reef-Associated (D), Demersal (E), Deep Demersal (F), Coastal Pelagic (G), Epipelagic (H), and Mesopelagic (I).