

Response of dung beetles diversity to remediated soils ecosystems in the Ecuadorian Amazon

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Dung beetles are considered to be habitat quality bioindicators. However, studies regarding the Scarabaeinae diversity in remediated ecosystems previously affected by hydrocarbon activity are null. We evaluated the diversity of dung beetles present in remediated soils previously contaminated with hydrocarbons and heavy metals (Agricultural soils and Sensitive ecosystems) and in non-contaminated soils (Natural forest and Palm plantations) in the Ecuadorian Amazon. Four sampling sites within each type of soil were established. At each site, six pitfall traps were installed and eleven samples were carried out monthly over one year. Each month, the traps remained active for 24 hours during a period of five consecutive days, resulting in a total sampling effort of 880 monitoring days with 330 trap-days per site. A total of 7,506 individuals belonging to 13 genera and 37 species of Scarabaeinae were captured. Mean values of abundance, richness, and diversity differed between ecosystems within each month. The non-contaminated soils ecosystems presented a higher abundance, richness, and diversity of beetles than the remediated soils ecosystems. Natural forest and Palm plantations presented higher abundance, richness, and diversity than Sensitive ecosystems and Agricultural soils, respectively. It can be concluded that dung beetle diversity in soils non-degraded (Natural forest and Palm plantations) do not only depend on the characteristics of the ecosystems, but also on the month of sampling.

Response of Dung Beetles Diversity to Remediated Soils Ecosystems in the Ecuadorian Amazon

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Abstract

Biodiversity in remediated soil ecosystems previously affected by hydrocarbon activities is poorly understood. Therefore, bioindicators such as dung beetles could be a valuable tool to elucidate the benefits of remediation processes. We evaluated the diversity of dung beetles on remediated soil ecosystems (Agricultural soils and Sensitive ecosystems) and on non-contaminated soils (Natural forests and Palm plantations) in Sucumbíos and Orellana provinces, Ecuadorian Amazon. The study was conducted at four sampling sites per ecosystem type (a total of 16 sites). At each sampling site, six pitfall traps were installed and monitored monthly for a 120-hour period during one year. We collected 37 species and 7,506 individuals of dung beetles. We observed significant differences in mean species abundance, richness and diversity between non-contaminated soil ecosystems and remediated soil ecosystems, with Natural forests presenting the highest values, and Agricultural soils the lowest values. Regarding sampling month, we also found significant differences among ecosystems. It is remarkable that between the Agricultural soils (remediated soil ecosystems) and Palm plantations (non-contaminated soils) we found the highest species similarity (34.2%). Variation in dung beetle diversity among ecosystems may aid in decision making to improve remediation processes at sites affected by hydrocarbon activities.

Subjects Biodiversity, Ecology, Entomology

41 **Keywords** Scarabaeinae, tropical rain forest, degradation, ecological restoration

42

43 **Introduction**

44 The extraction of hydrocarbon resources has led to the fragmentation of tropical forests around
45 the world (Abrahams, Griffin & Matthews, 2015; Bogaert et al., 2011; Thomas, Brittingham &
46 Stoleson, 2014). In addition, oil extraction processes often cause accidental spills, which result in
47 contamination of soil and water sources, affecting the environment and hence all life forms
48 (Sajna et al., 2015; Souza, Vessoni-Penna & De Souza Oliveira, 2014). In Ecuador, the Amazon
49 Forest is the most affected site by hydrocarbon activities (Villacís, 2016; Villacís et al., 2016a;
50 Villacís et al., 2016b), where crude oil spills of around 650,000 barrels have negatively impacted
51 the diversity of flora and fauna (Yáñez & Bárcenas, 2012; Ministerio del Ambiente, 2016). Due
52 to this situation, the Ecuadorian government has remediated 1,200,098 m³ of contaminated soil
53 (PETROAMAZONAS EP, 2018). The remediation process includes the collection and washing
54 of solid waste; the suction and transport of fluids; the cleaning of contaminated soil; and the
55 revegetation of intervened areas (García-Villacís, 2021). After remediation, each site is
56 catalogued as either Agricultural soil or Sensitive ecosystem in accordance with specific
57 permissible levels of hydrocarbons, cadmium, nickel, and lead established in environmental
58 regulations (Ministerio de Energía y Minas, 2010).

59 García-Villacís et al. (2021) analyzing the benefits of the remediation process in Ecuador
60 included ecosystem variables such as acidification, terrestrial-, and freshwater-eutrophication,
61 and freshwater ecotoxicity. However, organisms that serve as bioindicators of ecosystem
62 alterations were not included. Changes in the abundance, diversity, and composition of these
63 organisms can elucidate the effects of environmental disturbance (Kremen, 1992; Grand et al.,
64 2017). Among the most commonly used bioindicators are dung beetles (Coleoptera,
65 Scarabaeinae), which are distributed across a wide range of geographical locations (Lumaret &
66 Lobo, 1996; Herzog et al., 2013), have high levels of diversity (Espinoza & Noriega, 2018), and
67 are sensitive to microclimatic changes caused by deforestation and forest fragmentation (Campos
68 & Hernández, 2015; Davis & Sutton, 1998; Davis et al., 2001). Dung beetle communities are
69 commonly decreasing under environmental changes that also affect overall ecosystem health
70 (Otavo, Parrado-Rosselli & Noriega, 2013), since these organisms fulfill important ecological
71 functions such as secondary seed dispersal, decomposition of organic matter, and nutrient
72 cycling (Rangel-Acosta & Martínez-Hernández, 2017; Fernandes et al., 2019).

73 Several studies have been conducted to assess the diversity of dung beetles in degraded
74 ecosystems (Andresen, 2005; Feer & Hingrat, 2005; Nichols et al., 2007; Sánchez-de-Jesús et al.,
75 2016). Therefore, dung beetle diversity could be a useful tool to better understand the impact of
76 remediation activities on sites previously affected by hydrocarbon activities. It was hypothesized
77 that sites with soil degradation (Sensitive ecosystems and Agricultural soils) decrease dung
78 beetle diversity when compared to non-contaminated soils (Natural forests and Palm
79 plantations). To address this hypothesis, changes in abundance, richness, and diversity of dung
80 beetle communities were evaluated in two types of remediated ecosystems.

81

82 **Materials & Methods**

83

84 **Ethics statement**

85 This study was authorized under research permit 016-2018-IC-DPAO/AVS and authorization for
86 the mobilization of specimens 005-2019-MOV-FAU-DPAO/AVS, both issued by the Ministerio
87 del Ambiente del Ecuador.

88

89 **Study area**

90 The study was performed in the Sucumbíos (0° 5'S, 76° 53'W) and Orellana (0° 56'S, 75° 40'W)
91 provinces in the Ecuadorian Amazon. The land use of the localities is distributed in 70% for
92 natural forest, 22% for crops and pastures, and 8% for urban and industrialized areas (Ministerio
93 de Agricultura y Ganadería, 2015). The predominant climate is Humid Tropical Rainforest
94 according to the Holdridge climate classification (Holdridge, 1967), with a mean annual
95 temperature of mean 24.5 ° C and heavy rainfall throughout the year (4132 mm). Mean of
96 climate variables during the study period are showed in table 1.

97

98 **Selection of collection sites**

99 Sampling sites were established after the completion of the remediation process (~1 year) in
100 ecosystems previously affected by hydrocarbon activities (Sensitive ecosystems and Agricultural
101 soils). The designation as sensitive ecosystems or agricultural soils is based mainly on
102 contaminant levels in accordance with environmental regulations (lower in Sensitive ecosystems)
103 but also on the landscape characteristics of each site (Table 2). In addition, two types of non-
104 contaminated soil ecosystems used as controls (Natural forests and Palm plantations) were also
105 included. Before the beginning of the study, a composite soil sample (5 per sampling site) was
106 collected in order to determine the total petroleum hydrocarbon and polycyclic aromatic
107 hydrocarbon concentrations (Table 2) using GC2 014 gas chromatographs (Shimadzu Scientific
108 Instruments, Inc, Columbia, MD, USA). In addition, the concentration of cadmium (Cd), nickel
109 (Ni), and lead (Pb) in soils (Table 2) was determined by using atomic absorption spectrometry
110 (AA-6800; Shimadzu Scientific Instruments, Inc, Columbia, MD, USA) as indicated by the EPA
111 SW-846 method (Le Blanc & Majors, 2001). The analyses were performed at the Soil
112 Laboratory of the Universidad de las Fuerzas Armadas, Ecuador.

113

114 **Sampling design**

115 Sixteen sample sites (four per each ecosystem type) were established. In each sample site, six
116 pitfall traps baited with pig dung were placed 10 m apart. Pitfall traps consisted of plastic
117 containers of 0.8 L (15 cm depth × 10 cm diameter) buried up to their rims in the soil and
118 containing a solution 50:50 of water with alcohol. The traps remained active at the sites for 120 h
119 per month during one year (February 2018 to January 2019). The amount of dung per trap was
120 ~50 g and was replaced every 24-36 h. In April, 2019 the traps were not evaluated due to conflict
121 with the landowners, who prevented the entrance to the sampling sites in La Joya de los Sachas
122 locality.

123 Dung beetles were preserved in 70% ethanol, and some specimens were pinned and
124 identified to species using dichotomous keys (Chamorro et al., 2018; Vaz-De-Mello et al., 2011)
125 and comparing with voucher specimens from Museo de Historia Natural “Gustavo Orcés”,

126 (Escuela Politécnica Nacional, Quito, Ecuador). Vouchers were deposited in the Museum of
127 Zoological Researches (Universidad de las Fuerzas Armadas, Sangolquí, Ecuador).

128

129 **Data analysis**

130 All analyses were performed using the software INFOSTAT (Di Rienzo et al., 2008) in interface
131 with R (R Core Team, 2013). To detect the effectiveness of the inventories, species accumulation
132 curves were created using the Clench method (Moreno & Halffter, 2000; Soberon & Llorente,
133 1991). In addition, the richness observed in each type of ecosystem was evaluated using the non-
134 parametric estimator Chao 1 (Colwell & Elsensohn, 2014).

135 The monthly data were grouped to estimate the total species abundance, richness and
136 Shannon diversity index of each plot (Magurran, 1998). Pooled data were used to compare the
137 four types of ecosystems due to the homogeneity among the four sampling sites per ecosystem
138 type.

139 Abundance, richness, and diversity were analyzed by using repeated measures (by
140 month). Differences between ecosystems were analyzed using analysis of variance with mixed
141 models for a complete randomized design with six replications. In addition, we performed
142 orthogonal contrast for treatments. The first contrast evaluated differences in abundance,
143 richness, and diversity between remediated soil ecosystems and non-contaminated soil
144 ecosystems. The second contrast evaluated the differences between Agricultural soils and Palm
145 plantations, and the third contrast evaluated the differences between Sensitive ecosystems and
146 Natural forests. Furthermore, we tested for differences between ecosystems, months and
147 interactions by using a DGC post hoc test ($P < 0.05$). The normality of the data was verified
148 using the Shapiro-Wilks test, and the homoscedasticity was modeled using independent
149 variances.

150 The Sørensen index was used to compare the similarity of dung beetle species
151 composition between each type of ecosystem evaluated. Finally, a dendrogram was prepared
152 using this information (Beals, 1984).

153

154 **Results**

155 We collected 7,506 individual beetles of the Scarabaeinae subfamily, belonging to 13 genera and
156 37 species (Table 3). Specific abundance varied greatly, ranging from one to 1,502 individuals
157 (an average of 202.86 individuals \pm 52.81 SE per species). *Canthon aequinotialis* (20% of total
158 abundance), *Ontherus sulcator* (13%), *Dichotomius ohausi* (10%), and *Deltochilum howdeni*
159 (9%), accounted for 52% of all individuals collected. Fifty-one percent of the species were
160 classified as rare, with a relative frequency of less than ten percent. Twenty-two percent of the
161 total species collected were found in all four evaluated ecosystems, while 12 others were
162 exclusive to one of them: 11 in the natural forest and one in the Sensitive Ecosystem.

163 *Canthidium aurifex*, *Eurysternus atrosericus*, *Ontherus sulcator*, *Onthophagus osculatii*, and *O.*
164 *nyctopus* are new provincial records, whereas *O. hircus* is a new record for Ecuador (Table 3).

165 As the sampling time increased, the accumulated richness of the dung beetle decreased
166 from the seventh month in the Natural forests, whereas in the Sensitive ecosystems, Agricultural

167 soils and Palm plantations, the curves did not stabilize (Fig. 2). The non-parametric estimator of
168 richness (Chao 1) showed that the efficiency of the inventories reached 99.20 % in natural forest,
169 97.63 % in sensitive ecosystems, 94.34 % in agricultural soils, and 87.35 % in palm plantations.

170 The average values of abundance, richness, and the Shannon index differed between
171 ecosystems within each month (Table 4).

172 The orthogonal contrast showed that non-contaminated soil ecosystems contained higher
173 abundance, richness and diversity of beetles in comparison to remediated soil ecosystems ($F_{1,132}$
174 = 313.51, $P < 0.0001$). Natural forest presented higher abundance, richness and diversity than
175 Sensitive ecosystems ($F_{1,132} = 313.51$, $P < 0.0001$) and Palm plantations presented higher
176 abundance, richness and diversity than Agricultural soils ($F_{1,132} = 51.60$, $P < 0.0001$; Fig. 3).
177 The highest monthly average values for abundance (January and November), richness (January,
178 February, September, October and November) and Shannon's index (September and November)
179 were recorded in the Natural forests (Fig. 4).

180 Cluster analysis showed that Agricultural ecosystems and Palm plantations presented a
181 species similarity of 32.4%, and both were similar to the Sensitive ecosystems at 26.8%. The
182 Natural forests were similar to other ecosystems only at 8.53% (Fig. 5).

183

184 Discussion

185 Our study provides the first quantitative data on dung beetle communities in ecosystems affected
186 by hydrocarbon activities in the Ecuadorian Amazon. For all evaluated ecosystems, the richness
187 of dung beetle communities was greater than 87%, which suggests that minimum changes in
188 species inventories could exist if sampling effort is increased (Feinsinger, 2001).

189 The species presented in this study represent 17% of the 220 species of dung beetles
190 registered in Ecuador (Chamorro et al., 2018) and more than 50% of previously registered
191 species in the Orellana and Sucumbios provinces. Five species are new for these provinces,
192 whereas *Ontophagus hircus* was recorded for the first time in Ecuador. This demonstrates that
193 beetle diversity must be studied to understand soil disturbance in tropical forests (Beiroz et al.,
194 2017).

195

196 Dung beetle diversity

197 The community structure in the non-contaminated soil ecosystems trends toward high
198 abundance, richness, and diversity when compared to remediated soil ecosystems. The results are
199 similar to those reported by Da Silva, Vaz-de-Mello & Di Mare (2013) and Batilani-Filho &
200 Hernandez (2017), who also found higher values of abundance, richness, and diversity of dung
201 beetles in remnant Atlantic forests (Southern Brazil) and lower values in soils affected by
202 agriculture and deforestation. This could be because natural forests are the habitat of birds and
203 mammals, which provide food resources for dung beetles (Campos & Hernández, 2015; Niero &
204 Hernández, 2017).

205 Dung beetles are very sensitive to habitat disturbance (Audino, Louzada & Comita, 2014;
206 Campos & Hernández, 2015; Da Silva, Vaz-de-Mello & Di Mare, 2013). Changes in dung beetle

207 community structure have been reported under low vegetation cover (Nichols et al., 2007) as a
208 result of intense solar radiation on the soil surface, which accelerates the decomposition rate of
209 food sources (Méndez et al., 2019). Moreover, chemical perturbations affect dung beetle
210 communities in several ways such as changes in composition, diversity and population. (Hutton
211 & Giller, 2003; Correa et al., 2022). Therefore, the presence of hydrocarbons and heavy metals
212 in Agricultural soils and Sensitive ecosystems as well as landscape modifications at oil
213 exploitation sites could influence abundance, richness, and diversity of dung beetle assemblages.

214 Similarly, cluster analyses showed an acute division between the Natural forests and
215 other ecosystems evaluated. This could be because natural forest fragments within human
216 modified landscapes constitute wildlife refuges (Blaum et al., 2009). This trend of decreased
217 dung beetle diversity between the Natural forest and the other ecosystem types follows a general
218 decreasing gradient of diversity and an increase in anthropomorphic disturbances due to
219 contamination and land use (Aninta et al., 2019; McCain, 2005; Stevens & Gavilanez, 2015). For
220 example, Scarabaeinae diversity in Palm plantations was similar to that found in Agricultural
221 soils. This is consistent with previous studies of Fitzherbert et al. (2008) and Harada et al.
222 (2020), which reported that agrochemicals could favor the degradation of soil and nutrients and
223 hence diminish dung beetle diversity.

224

225 **Temporal variation of dung beetle diversity**

226 The diversity of dung beetles is determined by regional rainfall patterns (Novais et al., 2016).
227 Our results indicated that the diversity of dung beetles in the Natural forests was higher during
228 the month with lower levels of precipitation (October 238 mm month⁻¹) which is consistent with
229 the study of Ibarra-Polesel, Damborsky & Porcel (2015), who studied dung beetles in subtropical
230 ecosystems. However, previous studies demonstrate that higher beetle diversity is linked to the
231 months with elevated values of precipitation (Escobar et al., 2008; Nunes et al., 2016; Rangel-
232 Acosta & Martínez-Hernández, 2017).

233 Higher dung beetle diversity during the months with lowest levels of rainfall in the
234 evaluated ecosystems may be due to the interference of factors other than rainfall. For example,
235 alteration of microclimates and microhabitats (Medina, Escobar & Kattan, 2002; Noriega &
236 Realpe, 2018; Sánchez-Hernández et al., 2022), changes in trophic structure (Novais et al.,
237 2016), as well as altitudinal gradient effects (Noriega & Realpe, 2018). These factors could
238 affect mainly the mobility, displacement, and genetic flow of organisms between ecosystems
239 (Harvey, Gonzalez & Somarriba, 2006). For example, the Natural forests in the provinces where
240 the study was conducted are under a higher degree of environmental disturbance (deforestation,
241 forest fragmentation, oil spills, population growth, etc.) (Rivera-Parra et al. 2020) and, in
242 general, the entire Amazon basin is under massive degradation due to deforestation (Marin et al.
243 2022).

244

245 **Implications for the conservation**

246 The diversity of dung beetles provides a useful tool for assessing the temporary status of
247 remediated sites previously affected by hydrocarbon activities. Although differences in beetle
248 diversity were found between remediated ecosystems, similar recommendations for conservation
249 measures can be made for both Agricultural soils and Sensitive ecosystems. Therefore,
250 Agricultural soils and Sensitive ecosystems should not only be based on the levels of
251 hydrocarbons and heavy metals but also on the diversity of bioindicators.

252 The presence of dung beetles in the remediation ecosystems provides a guideline for
253 implementing strategies to conserve the existing diversity. However, the conservation of
254 biodiversity in remediated ecosystems depends not only on remediation activities, but also on
255 other anthropogenic activities in the Amazonian tropical forests.

256 Overall, dung beetle diversity could be used for conservation planning and
257 management of hydrocarbon- and heavy metal-contaminated ecosystems. In addition, our study
258 provides a baseline for future research that may include other environmental variables and
259 activities that modify dung beetle diversity.

260

261 **Conclusions**

262 Our study shows that in sites where hydrocarbons and heavy metals were present, the abundance,
263 richness and diversity of dung beetles were lower compared to non-contaminated sites. In
264 addition, dung beetle diversity changed throughout the year and was significantly higher in
265 months with low precipitation.

266

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273

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Figure 1

Location of collection sites in the Sucumbíos and Orellana provinces

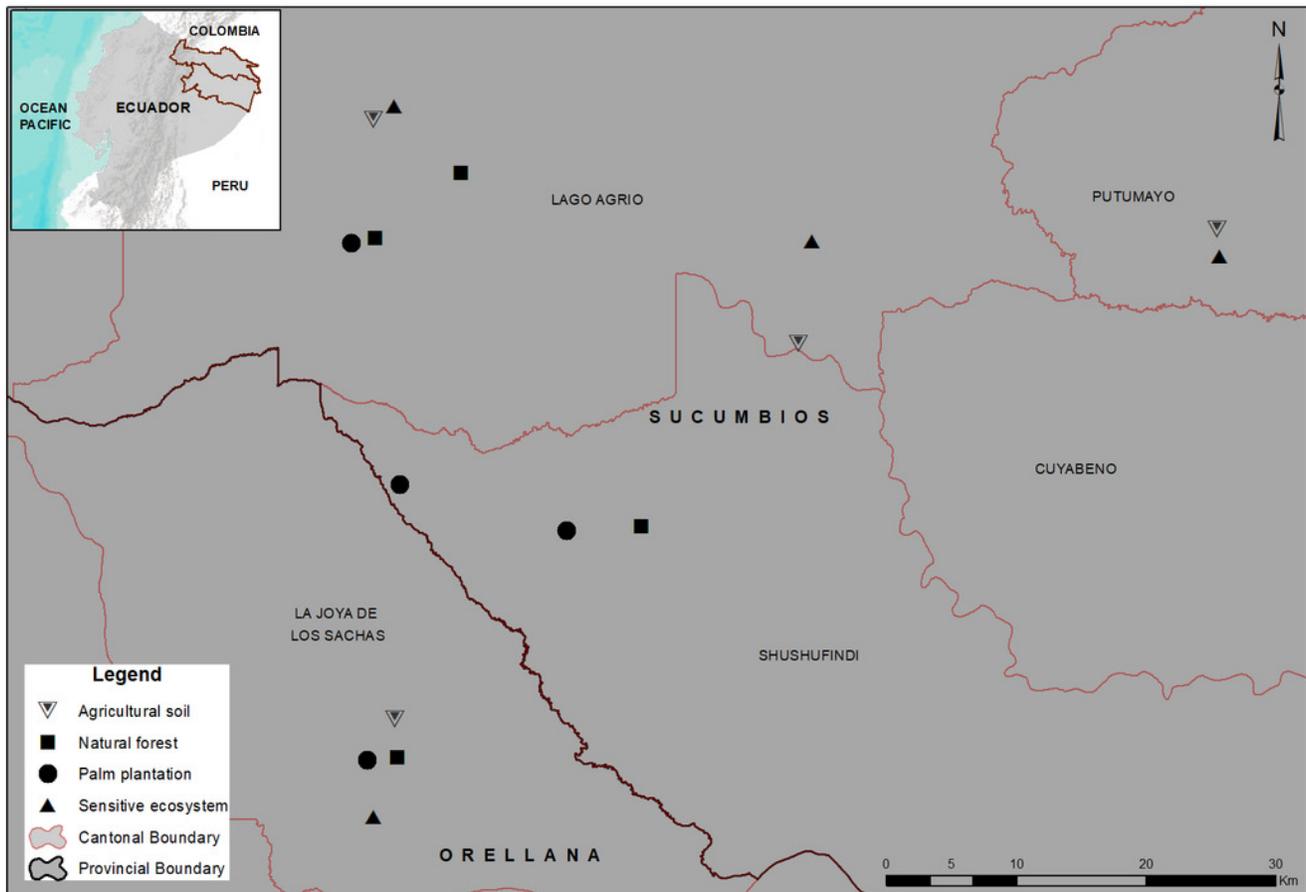


Figure 2

Species accumulation curves of Scarabaeinae communities recorded in remediated soil ecosystems and non-contaminated soil ecosystems in Ecuadorian Amazon

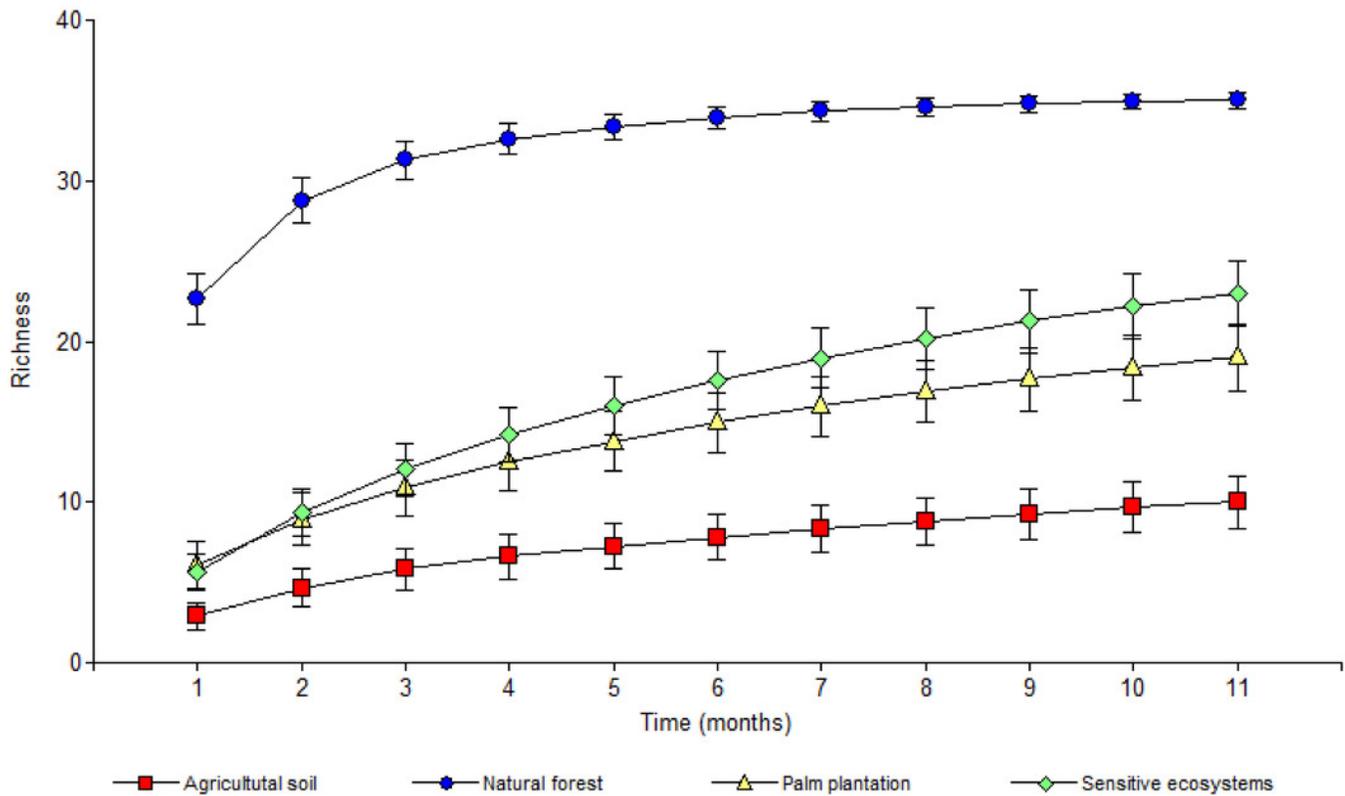


Figure 3

Abundance, richness, and diversity of beetles that were collected in remediated soil ecosystems and non-contaminated soil ecosystems in the Ecuadorian Amazonia

Bars represent means \pm standard error (n = 44)

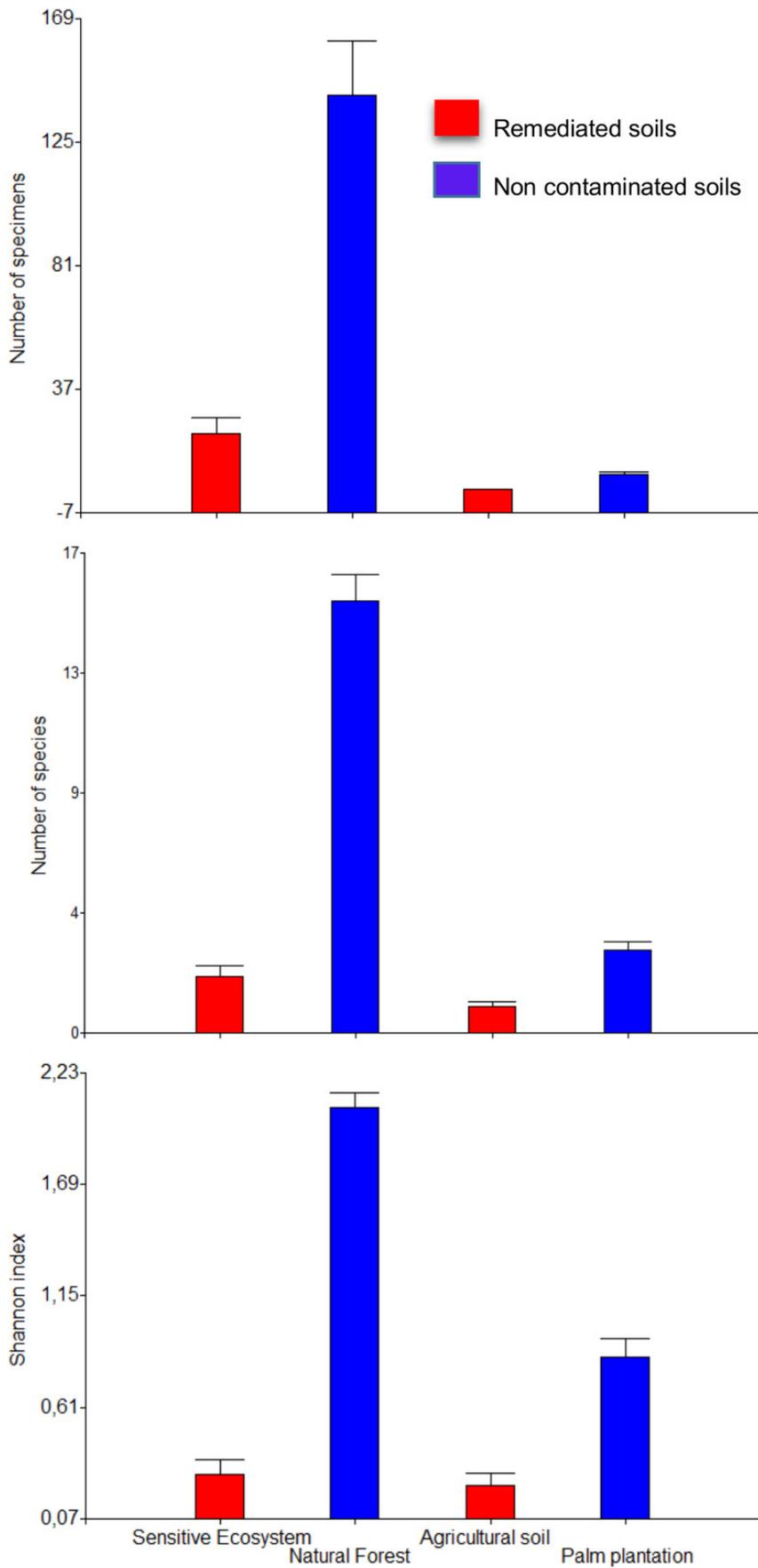


Figure 4

Abundance (top), richness (middle), and diversity (bottom) of beetles that were collected monthly on remediated soil ecosystems and non-contaminated soil ecosystems in the in the Ecuadorian Amazon

Values are means \pm standard error (n = 4). Different letters in each point indicate significant differences (DGC, P < 0.05)

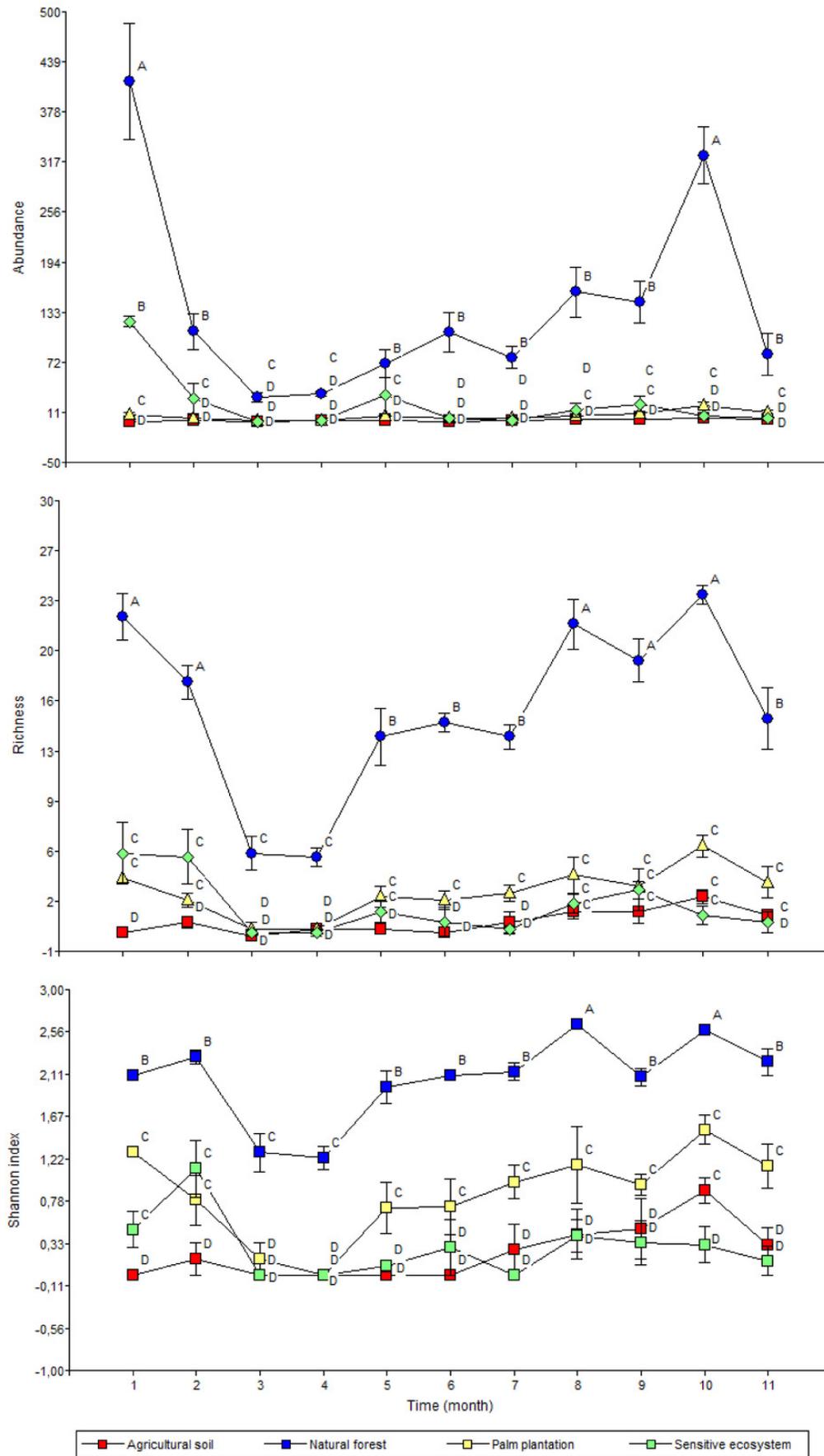


Figure 5

Species similarity clusters based on the Bray-Curtis distance of the Sorensen similarity percentage in remediated soil ecosystems and non-contaminated soil ecosystems in the Ecuadorian Amazon

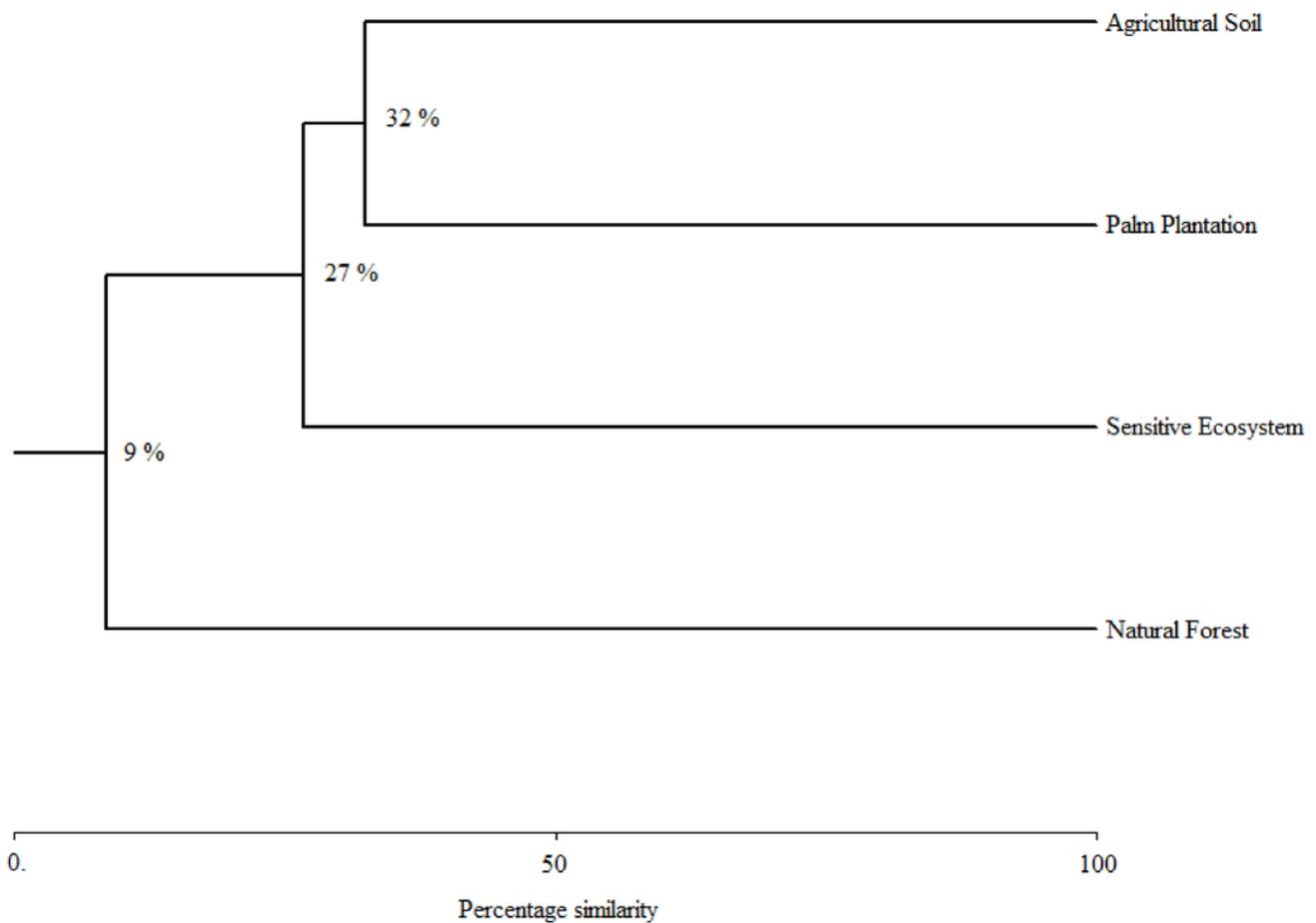


Table 1 (on next page)

Mean values of climatic variables in the ecosystems evaluated

1

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Order	Month/Year	Rainfall mm	Temperature °C	Humidity %
1	February/2018 ^r	368.81	24.36	90.86
2	March/2018 ^r	429.65	24.2	92.14
3	May/2018 ^r	372.02	24.2	93
4	June/2018 ^r	381.61	23.5	91.43
5	July/2018 ^d	322.25	23.96	90
6	August/2018 ^d	293.75	24.08	89
7	September/2018 ^d	248.15	24.48	88.17
8	October/2018 ^d	238.49	24.93	89.14
9	November/2018 ^d	298.9	24.92	88.5
10	December/2018 ^r	350.52	24.46	89
11	January/2019 ^r	376.31	24.76	89.5

3

^r= High rainfall, ^d= Low rainfall

4

Table 2 (on next page)

Soil characteristics of the remediated soil ecosystems and non-contaminated soil ecosystems in the Ecuadorian Amazon

1

Variable	Remediated soil ecosystems		Non-contaminated soil ecosystems	
	Sensitive ecosystems	Agricultural soils	Natural forest	Palm plantations
	Application of air and water to the soils with compressors and high-pressure pumps to release crude oil (Petroamazonas EP, 2018)			
Remediation Activities	Yes	Yes	No	No
Plot size (ha)	1	1	1	1
Total hydrocarbons (mg kg ⁻¹)	< 1 0000	< 2 500	No	No
Polycyclic Aromatic Hydrocarbons (mg kg ⁻¹)	< 1	< 2	No	No
Cadmium (mg kg ⁻¹)	< 1	< 2	No	No
Nickel (mg kg ⁻¹)	< 40	< 50	No	No
Lead (mg kg ⁻¹)	< 80	< 100	No	No
Agrochemical Use	No	No	No	Yes (herbicides and fungicides)
Tree cover presence	Yes	No	Yes	No
Number of present strata	2	0	4	2
Most abundant tree species	<i>Otoba parvifolia</i> , <i>Guarea</i> sp., <i>Pouroma</i> sp.	No	<i>Ceiba pentandra</i> , <i>Otoba parvifolia</i> , <i>Pouteria aubrevillei</i> , <i>Inga</i> sp., <i>Nectandra guararipo</i> , <i>Cordia alliodora</i> .	<i>Elaeis guianensis</i>
Mean number of trees DAP > 10 cm per ha	3.6	-	21.47	143
DAP mean ± SE (cm)	44.13±1.69	-	45.72±1.61	35.45±4.54
Mean total height ± SE (m)	14.54±0.65	-	16.23±0.25	18.45±2.21

2

Table 3 (on next page)

Dung beetle species collected in remediated soil ecosystems and non-contaminated soil ecosystems in the Ecuadorian Amazon

1

No.	Species	Record type	Agricultural soils	Natural forest Total	Sensitive ecosystems Total	Palm plantations	Assemblage
1	<i>Canthon aequinoctialis</i> Harold, 1868	RE		1 490	12		1 502
2	<i>Ontherus sulcator</i> Fabricius, 1775	NR-P	18	135	747	90	990
3	<i>Dichotomius ohausi</i> Luederwaldt, 1923	RE	7	639	17	72	735
4	<i>Deltochilum howdeni</i> Martínez, 1955	NR-P		671	13	1	685
5	<i>Onthophagus haematopus</i> Harold, 1887	RE		493	7		500
6	<i>Eurysternus plebejus</i> Harold, 1880	RE	6	386	4	30	426
7	<i>Onthophagus osculatii</i> Guérin-Méneville, 1855	NR-P		395	7		402
8	<i>Coprophaneus telamon</i> Erichson, 1847	RE	11	307	22	33	373
9	<i>Onthophagus xanthomerus</i> Bates, 1887	RE		246	22	5	273
10	<i>Dichotomius</i> sp. 1	NA		232		1	233
11	<i>Dichotomius mamillatus</i> Felsche, 1901	RE		186	2	1	189
12	<i>Deltochilum amazonicum</i> Kolbe, 1905	RE		175		3	178
13	<i>Eurysternus atrosericus</i> Génier, 2009	NR-P	1	134	5	5	145
14	<i>Eurysternus squamosus</i> Génier, 2009	RE		107			107
15	<i>Canthon luteicollis</i> Erichson, 1847	RE		92	1	1	94
16	<i>Eurysternus caribaeus</i> Herbst, 1789	RE	4	76		13	93
17	<i>Eurysternus wittmerorum</i> Martínez, 1988	RE	1	61	7	10	79
18	<i>Canthidium</i> cf. <i>rufinum</i> Harold, 1867	RE		76			76
19	<i>Onthophagus hircus</i> Billberg, 1815	NR-E	10		40	6	56
20	<i>Onthophagus nyctopus</i> Bates, 1887	NR-P		52	2		54
21	<i>Dichotomius podalirius</i> Felsche, 1901	RE		49			49
22	<i>Uroxys</i> sp. 1	NA		48			48
23	<i>Eurysternus foedus</i> Guérin-Méneville, 1844	RE	3	30	1	6	40
24	<i>Phaneus chalcomelas</i> Perty, 1830	RE		28	3	1	32
25	<i>Scyballocanthon macullatus</i> Schmidt, 1920	RE		31	1		32
26	<i>Oxysternon silenus</i> d'Olsouefieff, 1924	RE	1	4	1	16	22
27	<i>Eurysternus hamaticollis</i> Balhasar, 1939	RE		19	1	1	21
28	<i>Onthophagus onore</i> Zunino & Halffter, 1997	RE		15			15
29	<i>Canthidium</i> sp. 1	NA		10	3	1	14
30	<i>Canthidium aurifex</i> Bates, 1887	NR-P		9	3		12
31	<i>Oxysternon conspicillatum</i> Weber, 1801	RE		8			8
32	<i>Deltochilum carinatum</i> Westwood, 1837	RE		7			7
33	<i>Onthophagus marginicollis</i> Harold, 1880	RE			6		6
34	<i>Scyballocanthon furvus</i> Schmidt, 1920	NA		4			4
35	<i>Canthidium onitoides</i> Perty, 1830	RE		3			3
36	<i>Malagoniella astyanax</i> Halffter, Pereira & Martínez, 1960	RE		2			2
37	<i>Canthon angustatus</i> Harold, 1867	RE		1			1
Abundance			62	6 221	927	296	7 506

Richness

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Data in bold indicate species with greatest specific abundance by ecosystem type in the study. RE = registered in the provinces studied, NR-P = newly registered in the provinces studied, NA = not evaluated, NRE = newly registered in Ecuador

Table 4(on next page)

Analysis of variance for the abundance, richness, and diversity of beetles that were collected monthly in remediated soil ecosystems and non-contaminated soil ecosystems in the Ecuadorian Amazon

Ecosystems were considered fixed factor and months random factor

Source	df	Abundance		Richness		Shannon	
		F	P	F	P	F	P
Ecosystems types	(3)	302.09	< 0.0001	462.21	< 0.0001	314.65	< 0.0001
Remediated soils ecosystems vs. Control ecosystems	1	374.32	< 0.0001	483.13	< 0.0001	313.51	< 0.0001
Agricultural soils vs. Palm plantations	1	991.54	< 0.0001	59.97	< 0.0001	51.60	< 0.0001
Sensitive ecosystems vs. Natural forest	1	289.20	< 0.0001	502.33	< 0.0001	427.16	< 0.0001
Months	10	26.09	< 0.0001	18.63	< 0.0001	9.56	< 0.0001
Ecosystems × Months	30	8.07	< 0.0001	2.41	0.0003	1.90	0.0073

1 df = degrees of freedom; F = result of F-Fisher value; P = result of probability.