

Alien species revises systematic status: integrative species delimitation of two similar taxa of *Symbrenthia* Hübner, [1819] (Lepidoptera, Nymphalidae)

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Introduction of organisms to new range may impose detrimental effects on local organisms, especially when closely related species are involved. Species delimitation employing an integrative taxonomy approach may provide a quick assessment for the species status between taxa of interest, and to infer ecological competition and/or introgression that may be the associated with the introduction. A nymphalid butterfly, *Symbrenthia lilaea lunica*, was recently introduced to Taiwan, where a closely related local taxon, *S. l. formosanus*, can be found. We employed multiple species delimitation methods to study the species status between the two taxa, and the results revealed that they can be recognized as two distinct species. We further performed a niche modeling approach to investigate the ecological interaction between the two species. The taxonomic status of the two taxa, now elevated to species, has been revised and conservation facing rapid expansion of the introduced species discussed.

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ABSTRACT

Introduction of organisms to new range may impose detrimental effects on local organisms, especially when closely related species are involved. Species delimitation employing an integrative taxonomy approach may provide a quick assessment for the species status between taxa of interest, and to infer ecological competition and/or introgression that may be the associated with the introduction. A nymphalid butterfly, *Symbrenthia lilaea lunica*, was recently introduced to Taiwan, where a closely related local taxon, *S. l. formosanus*, can be found. We employed multiple species delimitation methods to study the species status between the two taxa, and the results revealed that they can be recognized as two distinct species. We further performed a niche modeling approach to investigate the ecological interaction between the two species. The taxonomic status of the two taxa, now elevated to species, has been revised and conservation facing rapid expansion of the introduced species discussed.

Keywords Systematics, Species delimitation, Biological introduction, Species interaction, Niche modeling

Introduction

Biological invasions and range expansion of organisms usually impose unfavorable effects on local organisms that share similar ecological requirements (*Mooney & Cleland, 2001*), particularly when the expanding form is of continental origin and entering insular areas (*Sax & Gaines, 2008*). Competition may occur between native taxa and the invading one, notably when the counter taxa are closely related, where they are expected to share resource requirements (*Zwerschke et al., 2018*). This kind of scenarios have been well-documented in various organisms, such as ants (*von Aesch & Cherix, 2005*), birds (*Koenig, 2003*), molluscs (*Zwerschke et al., 2018*), plants (*Leger & Espeland, 2010; Čuda et al., 2015; Sheppard & Brendel, 2021*), etc., and even inspired E. O. Wilson to coin in his “taxon cycles” hypothesis (*Wilson, 1961*). Alternatively, if the involved taxa were previously allopatric populations or subspecies, gene flow may occur between taxa, resulting in changes in the genetic make-up of the local population (*Rhymer & Simberloff, 1996*). Consequently, whether the exotic taxon represents interspecific or intraspecific entity to the local taxon may lead to different types of ecological and evolutionary impacts to local community. Species delimitation employing an integrative taxonomy approach has becoming popular to help with taxonomic treatments/decisions among closely related taxa and to detect cryptic species (*Moraes et al., 2021*). This approach can also help with assessing whether an invading taxon is conspecific or not to the local taxon.

A case of recent invasion, either by anthropogenic introduction or by natural means such as wind, of a nymphalid butterfly currently recognized as *Symbrenthia lila* Bascombe, Johnston & Bascombe, 1999 (Fig. 1A–1D) into Taiwan has been reported (Lu & Chen, 2014; Hsu et al., 2022). In Taiwan, there is a local native *Symbrenthia* Hübner, [1819] taxon, *S. l. formosanus* Fruhstorfer, 1908 (Fig. 1E–1H; Shirôzu & Ueda, 1992; Hsu et al., 2022). The taxon *lunica* Bascombe, Johnston & Bascombe, 1999 is a replacement name for *Papilio lucina* Stoll, 1780; in Cramer, De uitlandsche kapellen 4(26b-28): 82 (Bascombe et al., 1999), distributed from northern Indochina to southern China according to Tsukada & Nishiyama (1985). Although some authors (e. g. Bozano & Floriani, 2012; Ek-Amnuay, 2012; Lang, 2012; Monastyrskii, 2019) regarded *lunica* a synonym of the nominotypical *lila* described from India but have not formally synonymized these two names, others treat them as separate subspecies (e. g. Tsukada & Nishiyama, 1985; Hsu et al., 2022; Fric et al., 2022). *Symbrenthia lila lunica* from continental Asia was first documented in 2004 in Taiwan (Lu & Chen, 2014). Natural observations in Taiwan on immature biology and hostplant usage of both *S. l. lunica* and *S. l. formosanus* have documented that the two taxa use the same larval hostplants (Lu & Chen, 2014). The invading or introduction event represents a ‘natural experiment’ in which to examine whether these two taxa are conspecific as current classification suggests, or if they represent distinct biological species. The answer to this natural experiment may help clarify what

ecological effect the introduced taxon may impose on the native taxa. If they turn out to represent distinct species, competition between them may occur as they do not differ in larval hostplant usage. If they turn out to be conspecific, gene introgression may occur since hybridization is liable to occur.

We investigated the species status between the two now sympatric taxa in Taiwan using multiple coalescent model based species delimitation methods and distance based species delineation, under a scheme for testing hypothesis of taxonomic status developed by *Braby et al. (2012)*, following by a niche modeling survey to understand the ecological interaction between two taxa.

Materials and Methods

Sampling

The mitochondrial cytochrome oxidase subunit I (*COI*) gene has been successfully applied as a helpful marker with which to delimit closely related species (*Hebert et al., 2003*). We included a total of 13 specimens collected from various localities in our *COI*-based study. Four specimens of *S. l. formosanus* were collected around Taiwan, and six specimens of *S. l. lunica* were collected in Taiwan (mainland and Matsu archipelago), China, and Thailand. All samples were preserved in 70% ethanol and kept at -20°C for the subsequent molecular study. Moreover,

additional *COI* sequences of *S. l. formosanus* (AY788679) from Taiwan, *S. l. lunica* (EU368155, KJ649017, KX300094) from China, Vietnam, Myanmar respectively, and the nominate subspecies *S. l. lilaea* (Hewitson, 1864) (KP644228, KP644229) from India were obtained from GenBank. For the phylogenetic analyses, we used one sequence of *S. brabira* Moore, 1872 (EU368154) as an outgroup, which was also obtained from GenBank.

DNA extraction, PCR amplification and DNA sequencing

Genomic DNA was extracted from one leg of specimens using the Gentra Puregen tissue kit form QIAGEN (QIAGEN, Maryland, USA), following the manufacturer's protocol. A partial fragment from the *COI* gene was targeted for amplification by polymerase chain reaction (PCR). The *COI* gene was amplified using the universal primers COX-J-1460 (5'-TACAATTTATCGCCTAAACTTCAGCC-3') and COX-N-2191 (5'-CCCGGTAAAATTAATAAATAAACTTC-3'). PCR reactions were performed in a 30μL volume eppendorf, containing 1μL of extracted DNA, 23.5μL of ddH₂O, 3μL of 10X PCR reaction buffer, 0.6μL of each primer and 0.3μL of Power Taq (Genomics Biosci & Tech, Taiwan). The following PCR protocol was used: an initial denaturation at 95°C for 5 min, followed by 40 cycles of 30s denaturation at 95°C, 30s annealing at 50°C and 45s extension at 72°C, and a final extension at 72°C for 10 min. Automatic sequencing was preformed using an

110 ABI 3730XL DNA Analyzer (Applied Biosystems).

111

112 **Sequence Analyses and Phylogenetic Reconstruction**

113 Sequences were edited and assembled using Sequencher 4.10.1 (Gene Codes Corporation, Ann

114 Arbor, USA), and sequence alignments were performed using MUSCLE in MEGA 11(*Tamura et*

115 *al.*, 2021), and pairwise genetic distances between different populations of *S. l. formosanus* and

116 *S. l. lunica* were also measured using MEGA 11 with the Kimura 2-parameter model.

117 The best-fit nucleotide substitution model for phylogenetic analysis was inferred using

118 jModelTest 2.1.10 (*Posada, 2008*) based on Akaike information criterion (AIC). Phylogenetic

119 trees were reconstructed under maximum likelihood (ML) and Bayesian inference (BI). ML

120 analysis was performed using RAxML v8.2.10 (*Stamatakis, 2014*) with 1000 bootstrap replicates

121 to assess the reliability of the tree. BI analysis was performed using MrBayes 3.2.6 (*Ronquist et*

122 *al.*, 2012). For MrBayes, the substitution model inferred from jModelTest was applied. The

123 Bayesian Markov Chain Monte Carlo (MCMC) analysis for 10⁹ generations with sampling every

124 1000 generations was run to ensure the average standard deviation of split frequencies were

125 below 0.01. The first 30% of trees were discarded as burn-in. FigTree v1.4.4 was used to

126 visualize the consensus tree.

127

Molecular species delimitation analyses

Many molecular species delimitation programs have been proposed and broadly applied in speciation studies, which provides important evidence for integrative taxonomy. Among molecular species delimitation programs, the Poisson Tree Processes model (PTP) (Zhang *et al.*, 2013), the Automatic Barcode Gap Discovery (ABGD) (Puillandre *et al.*, 2012), and the Generalized Mixed Yule Coalescent model (GMYC) (Fujisawa & Barraclough, 2013) were developed as single locus-based approaches for species delimitation. Therefore, we delineated species limits among *S. l. formosanus*, *S. l. lunica*, and *S. l. lilaea* by employing the Molecular Operational Taxonomic Unit concept set by these three programs.

For PTP, we used the tree inferred by MrBayes as input tree on the web server (<https://species.h-its.org/ptp/>), with 100000 MCMC generations and 100 thinning. Subsequently, PhyloMap was used to visualize the results of PTP. For ABGD, we performed the analyses on the web version of ABGD (<https://bioinfo.mnhn.fr/abi/public/abgd/>), with default settings of relative gap width (X=1.5) and the Kimura two-parameter (K2P) model for nucleotide substitution. For GMYC, we used the phylogenetic tree inferred by MrBayes 3.2.6. The results from MrBayes were forced bifurcated by the “multi2di” and “chronos” function in the package “ape” in R 4.1.2. A single-threshold GMYC analysis was performed in the R package splits v1.0-20. We chose the single-threshold model because of the limited improvements of multiple-

threshold model.

Species distribution model of *S. l. formosanus* and *S. l. lunica* in Taiwan

Symbrenthia lilaea lunica was not known to occur in Taiwan until recently, although it inhabits Matzu and Kinmen, which are small outlying islands of Taiwan and close to mainland Asia. However, *S.l. lunica* arrived to the main island of Taiwan due to anthropogenic activities or via natural dispersal, with the first credible record found in Xinzhu in northwestern Taiwan in 2004 (Lu & Chen, 2014). Since then, the range of *S. l. lunica* has expended quickly, and is currently found in lowland areas throughout Taiwan (Lu & Chen, 2014; Hsu et al., 2022). It is an interesting issue whether competitive exclusion has happened between *S. l. lunica* and native *S. l. formosanus*, especially if the species delimitation analyses decide they represent different species.

The occurrence data of *S. l. formosanus* and *S. l. lunica* were obtained from the Global Biodiversity Information Facility (GBIF) (<https://gbif.org/>, accessed 26 July 2021), Taiwan Moth Information Center (<https://twmoth.terrsi.gov.tw/>, accessed 26 July 2021), and the specimen collection at National Taiwan Normal University. To test the interaction between these two species, we separated the occurrence data into two stages based on year. Because the first

documentation of *S. l. lunica* was in 2004, we divided the time period based on the median year (2012). The early invasion was defined as the data recorded from 1911–2012, and the late invasion was defined as the data recorded from 2013–2021. Repeated data was excluded using R 4.1.2, and we ensured the presences of only one presenting point in each raster to avoid overfitting. In total, 48 and 43 localities were obtained during the early invasion stage for *S. l. formosanus* and *S. l. lunica*, respectively (Fig. 2A), and 132 and 192 localities were obtained during the later invasion stage for *S. l. formosanus* and *S. l. lunica*, respectively (Fig. 2B). These data were organized using Microsoft Excel for the subsequent analyses.

A total 19 bioclimate variables (period: 1979–2013) were collected from CHELSA (<https://chelsa-climate.org/>, accessed on 14 July 2021) at a spatial resolution of 30 arc-seconds (1km²). These bioclimate variables were derived from temperature and precipitation, which are considered to be related to the distribution and survival of small arthropods and have been widely used in the prediction of species distribution (De Meyer *et al.*, 2010; Xu *et al.*, 2020). In order to avoid the effect of multicollinearity, these 19 variables were selected by the “vifstep” and “vifcor” function with the threshold of 10 and 0.6 separately in “usdm” package in R 4.1.2 (selected variables shown in Fig. 3).

MaxEnt (3.4.4) (Phillips *et al.*, 2006) was applied to predict the habitat suitability of *S. l. formosanus* and *S. l. lunica* based on the occurrence data. 10% of the data were selected to run a

random test and the remaining data were run following the default settings. Presence-only data were generated pseudo-absences and 10000 random background points were randomly selected by the MaxEnt model. The results were output after 10 cross-validation replicates.

The predictions generated from MaxEnt modeling were evaluated according to the threshold independent area under the receiver operating characteristic (ROC) curve (AUC) values. ROC curves were used to plot the true-positive rate against the false-positive rate and the AUC was used as a measure of the goodness of fit of the model. The AUC value ranges from 0 to 1, with higher values indicating higher predictive performance. The logistic output was chosen as an estimate of the probability of presence conditioned by bio-environmental variables per grid cell. Jackknifing was used to screen for the contribution of each bio-environmental variable used in the model.

We performed principal component analyses (PCA) to test the niche overlap of these two species in both the early invasion stage and late invasion stage. The 19 bioclimatic variables were obtained from the CHELSA database based on the GPS of each observation point. The analyses were conducted in R 4.1.2 using the function “prcomp”, with scatterplots built using the function “ggbiplot”. Additionally, in order to evaluate the niche shift pattern between the two *Symbrenthia* species in Taiwan, we apply methods modified from *Bates et al. (2020)* to quantify the niche shift between *S. l. formosanus* and *S. l. lunica* by calculating niche overlap, presented

by Schoener's D, and niche expansion of *S. l. lunica*.

Results

Taxonomic decisions

Phylogenetic reconstruction of *Symbrenthia* COI samples (Fig. 4) revealed that all samples of *lunica+lilaea* form a monophyletic group sister to *formosanus* samples, which also formed a monophyletic group. The p-distance was 0.0017 between *lunica* and *lilaea* and 0.0505–0.0525 between *lunica+lilaea* and *formosanus*. PTP, ABGD and GMYC all recognize a two species scenario, with *lunica+lilaea* and *formosanus* each representing a distinct species. Therefore, *formosanus* is recognized as a species distinct from *lunica+lilaea*, with the combination as *Symbrenthia formosanus* Fruhstorfer, 1908 (**stat. rev.**). The taxon *S. l. lunica* Bascombe, Johnston & Bascombe, 1999 is proposed to be lumped with *S. l. lilaea*, Hewitson, 1864 (**syn. nov.**) herein as the two may not be distinguished by COI barcode nor adult and immature morphology. We thus will call them *S. formosanus* and *S. lilaea* respectively in the remaining text of this article.

Environmental factors which contribute to the distribution of *Symbrenthia* species in question

After applying "usdm" package to remove those highly correlated factors, "bio 2", "bio 8", "bio 12" (Fig. 3A & 3B) are used to construct the species distribution model of "early invasive stage", while "bio 7", "bio 8", "bio 9", "bio 12", "bio 18" (Fig. 3C & 3D) are used to construct the species distribution model of "late invasive stage". According to the results of the jackknife test, the factors show different contribution patterns in the early invasive stage. In the early invasive stage, "bio 2" (annual precipitation) and "bio 12" (air temperature) contribute reversely between these two species; annual precipitation contributes more than mean diurnal air temperature range in the distribution model of *S. formosanus*, whereas mean diurnal air temperature range contributes more than annual precipitation amount in the model of *S. lilaia*.

Comparing the jackknife results of both species between the two invasive stages, "bio 8" contributes the most among all models. According to this, the mean daily air temperatures of the wettest quarter may play a key role in the distribution of these two *Symbrenthia* species in Taiwan.

The species distribution model and niche shifting of the two *Symbrenthia* species in different time stages

According to the species distribution model results, *S. formosanus* does not show an obvious change between the early and late invasive stages (Fig. 5A & 5C). For both invasive stages, the

presence probability of *S. formosanus* seems to be higher in the suburban areas and places with less human activity. For *S. lilaia*, the distribution model presents different results between the two time stages (Fig. 5B & 5D). Particularly, presence probability in the southwest part of Taiwan is higher in the later invasive stage (Fig. 5D). The SDM results of both species show that the presence probability decreases in the Pingtung area, the southernmost county of Taiwan. Although there may be biological importance to this observation, it is most likely a result of uneven presence observation point density in the later stage. Most of the presence points for the late invasive stage SDM are from northern Taiwan.

From the results of the early and late invasive stages (Fig 6), the niche overlap value increased during the recent years (past-2012 D: 0.48; 2013-2021 D: 0.64), and the niche expansion value of *S. lilaia* remained zero between the two different time stages. Together, these mean that, during these two periods of time, the niche of this alien species did not extend beyond the niche of the native species. According to the ENM model and the niche shift results, competitive exclusion seems to not be occurring between these two species over these 18 years.

Discussion

Taxonomic status of the introduced and native *Symbrenthia* butterflies

The introduced and native *Symbrenthia* butterflies in question of the study were regarded as

conspecific subspecies prior to the present study (e. g., *Hsu et al.*, 2022; *Fric et al.*, 2022). It has been argued that species delimitation is difficult for allopatric populations or subspecies of similar forms (*King*, 1993; *Braby et al.*, 2012), but in the present case, the introduction of continental *S. lilaea* to Taiwan has proven that insular *S. formosanus* ought to represent a species endemic to the island, instead of being a geographical race of the former. Distinctions between them include: 1) distal band on hindwing uppersides of both sexes form a continuous orange stripe in *S. lilaea* (Fig. 1A & 1C), whereas it is interrupted by darkened veins in *S. formosanus* (Fig. 1E & 1G); 2) distal tip of uncus is acute in *S. lilaea* (Fig. 7A), whereas it is blunt in *S. formosanus* (Fig. 7D); 3) distal margin of valva is rounded in *S. lilaea* (Fig. 7A), whereas it is angled, somewhat squared in *S. formosanus* (Fig. 7D); 4) ampulla is stout, slightly down-curved in *S. lilaea* (Fig. 7A), whereas it is slender, strongly bent downwards in *S. formosanus* (Fig. 7D); 5) posterior margin of sterigma is concave in *S. lilaea* (Fig. 7C), whereas it is truncate in *S. formosanus* (Fig. 7F); 6) yellow eggs are laid in cluster in *S. lilaea* (Fig. 8A), in contrast to green eggs laid singly in *S. formosanus* (Fig. 8D); 7) larvae are gregarious in *S. lilaea* (Fig. 8B–8C), but solitary in *S. formosanus* (Fig. 8E–8F).

Vouchers.

***Symbrenthia formosanus* Fruhstorfer, 1908**

272 *Ssp. formosanus*: **XINBEI CITY [= TAIPEI Co.]**: 1♀, Xindian, Sikanshui, 5. X. 2001 (H. S.
 273 Que leg.); 1♂, Wulai, 150m, reared from *Boehmeria nivea*, emerged 8. XI. 2004, HSU 04L21 (Y.
 274 F. Hsu leg.); 1♂, Wulai, 13. II. 2005, reared from *Debregeasia orientalis*, emerged 12. III. 2005,
 275 HSU 05B6 (L. H. Wang leg.); 1♀, same locality, 5. VI. 2005, reared from *B. densiflora*, emerged
 276 23. VI. 2005, HSU 05F9 (Y. F. Hsu leg.); 1♂, same locality, 5. VI. 2005, reared from *D.*
 277 *orientalis*, emerged 5. VI. 2005, HSU 05F10 (Y. F. Hsu leg.); 1♂, same locality, 13. VI. 2005,
 278 reared from *B. nivea*, emerged 2. VII. 2005, HSU 05F36 (Y. F. Hsu leg.); 1♀, same locality, 16.
 279 I. 2007 (Y. F. Hsu leg.); 1♂, Wulai, ca 200m, 5. VI. 2005 (Y. F. Hsu leg.). **TAIPEI CITY**: 1♀,
 280 Daan, Baozangyan, 28. III. 2019 (L. Huang leg.). **TAOYUAN CITY [= TAOYUAN Co.]**: 1♂,
 281 Fuxing, Xuanyuan, ca 1000m, 16. VI. 2005 (L. H. Wang leg.); 1♀, Fuxing, Daman, 10. VI.
 282 2005, reared from *B. nivea*, emerged 22-23. VI. 2005, HSU 05F28 (Y. F. Hsu leg.); 1♀, same
 283 locality, 10. VI. 2005, reared from *D. orientalis*, emerged 23. VI. 2005, HSU 05F29 (Y. F. Hsu
 284 leg.); 1♀, Fuxing, Gaoyi, 600m, 7. III. 2008 (Y. F. Hsu leg.). **YILAN Co.**: 1♂, 2♀, Nanao,
 285 14Km S. Nanao, 6. VI. 2005, reared from *B. nivea*, emerged 8-17. VI. 2005, HSU 05F17 (Y. F.
 286 Hsu leg.); 3♂, 1♀, Datong, Qilan, 1. XII. 2005, reared from *B. nivea*, emerged III. 2006, HSU
 287 05M7 (C. C. Lu & H. Y. Lee leg.). **XINZHU Co.**: 2♂, Jianshi, Xiuluan, 800m, 21. VI. 2005 (L.
 288 H. Wang leg.); 1♀, Jianshi, Guanwu, Dalu Forest Trail, 2. X. 2006 (L. H. Wang leg.); 1♂,
 289 Jianshi, Yulao, ca 1500m, 25. II. 2009 (C. K. Wang leg.). **MIAOLI Co.**: 1♂, Zhuolan, Liyutan,

290 300m, 2. VII. 2006 (Y. F. Hsu leg.). **TAIZHONG CITY [= TAIZHONG Co.]**: 1♀, Heping,
 291 Shangguguan, 900m, 30. X. 2006 (Y. F. Hsu leg.); 1♂, same locality, 29. VII. 2007 (Y. F. Hsu
 292 leg.); 1♂, Heping, Guguan, 700m, 5. VII. 1998 (Y. T. Lo leg.); 1♂, Heping, Guguan, IV. 2005
 293 (L. H. Wang leg.). **NANTOU Co.**: 1♂, Renai, Tunyuan/ Tianchi, 2000/2800m, 31. VII- 1. VIII.
 294 1998 (Y. T. Lo leg.); 1♂, 1♀, Renai, Meiyuan, ca 400m, 22. XI. 2004, reared from *B. nivea*,
 295 emerged 16-22. XII. 2004, HSU 04L49 (Y. F. Hsu leg.); 1♀, Yuchi, Shiguanyin, 480m, 22. XI.
 296 2004, reared from *B. densiflora*, emerged 21. XII. 2004, HSU 04L52 (Y. F. Hsu leg.); 1♂, Lugu,
 297 Fenghuanggu, 600m, 4. XII. 2004, reared from *B. densiflora*, emerged 23. XII. 2004, HSU 04M5
 298 (Y. F. Hsu leg.); 1♂, 3♀, Lugu, Fenghuanggu, 4. XII. 2004, reared from *B. nivea*, emerged 17-
 299 20. XII. 2004, HSU 04M4 (Y. F. Hsu leg.), 1♀, same locality, 11. XII. 2004, reared from *B.*
 300 *densiflora*, emerged 14. II. 2005, HSU 04M14 (Y. F. Hsu leg.), 1♂, same locality, 12. XII. 2004
 301 (Y. F. Hsu leg.); 1♂, Renai, Beidongyanshan, ca 1800m, 17. XII. 2005 (L. W. Wu & Y. F. Hsu
 302 leg.) (genitalia preparation YFH 1583), 1♀, same locality, 17. XII. 2005, reared from *D.*
 303 *orientalis*, emerged 4. I. 2006, HSU 05M32 (Y. F. Hsu leg.); 3♂, Renai, Tunyuan, 1550m, 28.
 304 VII. 2006 (Y. F. Hsu leg.); 1♀, Renai, Nanshanxi, ca 900m, 1. IX. 2007 (Y. F. Hsu leg.); 1♂,
 305 Lugu, ca 600m, 19. VII. 2006, reared from *B. nivea*, emerged 4. VIII. 2006, HSU 06G12 (Y. F.
 306 Hsu leg.); 1♂, 2♀, Lugu, ca 500m, 1. I. 2007, reared from *B. nivea*, emerged 18. I. 2007, HSU
 307 07A7 (Y. F. Hsu leg.); 1♀, Renai, Huisun, 600-700m, 20. II. 2010 (Y. F. Hsu & H. C. Huang

leg.). **HUALIAN Co.:** 2♂, Xiulin, Huitouwan, 23. VII. 2005 (Y. F. Hsu leg.), 1♂, same locality,
 30. I. 2007 (Y. F. Hsu leg.); 4♂, 3♀, Xiulin, Zuocang Trail, 7. III. 2006 (Y.F. Hsu leg.); 1♀,
 Xiulin, Guangbeibabiao, 2200m, 28. V. 2007 (Y. F. Hsu leg.); 1♀, Xiulin, Huoranting, 1000m,
 10. XII. 2007 (M. H. Sun leg.); 1♂, Xiulin, Wujiabengshan, 2000m, 7. VI. 2008 (L.H. Wang
 leg.); 1♀, Zhouxi, Zhongping Forest Trail, 1. VI. 2007 (L. H. Wang leg.). **TAINAN CITY [=**
TAINAN Co.]: 1♀, Dongshan, Kantoushan, 600-800m, 4. II. 2010 (Y. F. Hsu leg.).
PINGDONG Co.: 1♂, Shizi, Nunaishan, 200m, 26. III. 2006 (Y. F. Hsu leg.); 1♂, Fangliao,
 Yuquan, 26. II. 2006 (Y. F. Hsu leg.); 2♂, 1♀, Shizi, Lilongshan, 500m, 17. II. 2008 (Y. F. Hsu
 leg.); 1♂, same locality, 1. III. 2008 (Y. F. Hsu leg.); 1♀, Wutai, Jiudawu, 500m, 11. V. 2008 (C.
 L. Huang leg.); 1♀, Sandimen, ca 300m, 2. I. 2009 (Y. F. Hsu leg.). **TAIDONG Co.:** 1♀,
 Yanping, Hongye, 11. I. 2009 (C. H. Lin, jr leg.); 1♀, Taidong, Pipa Lake, 4. VI. 2006 (Y. F.
 Hsu leg.); 1♀, Haiduan, Liyuan, 7. VI. 2020 (Y. F. Hsu leg.) (genitalia preparation JYL1036).
***Symbrenthia lilaea* (Hewitson, 1864)**
JILONG CITY: 1♀, Longgang Trai, 26. IX. 2006 (Y. F. Hsu leg.); 2♀, same locality, 2. X.
 2006 (Y. F. Hsu leg.); 1♀, same locality, 7. II. 2007 (C. K. Wang leg.); 1♂, Shenao, Sea Level,
 20. VI. 2006, reared from *B. nivea*, emerged 2. VII. 2006, HSU 06F17 (Y. F. Hsu leg.); 2♂,
 Haimentianxian, 26. IX. 2006 (Y. F. Hsu leg.). **XINBEI CITY [= TAIPEI Co.]:** 7♂, 7♀, Wulai,
 150m, 8. XI. 2004, reared from *B. nivea*, emerged 29. XI-5. XII. 2004, HSU 04L20 (Y. F. Hsu

leg.); 1♂, Wulai, 11. III. 2005 (C. R. Chen leg.), 1♀, same locality, 5. VI. 2005 (Y. F. Hsu leg.),
 1♂, 1♀, same locality, 11. VIII. 2005, reared from *B. densiflora*, emerged 25. VIII. -1. IX. 2005,
 HSU 05H8 (Y. F. Hsu leg.), 1♂, same locality, 23. IX. 2005, reared from *B. nivea*, emerged 4. X.
 2005, HSU 05J72 (J. R. Chen leg.), 3♀, same locality, 6. X. 2006, reared from *B. nivea*, emerged
 3-15. XI. 2006, HSU 06L3 (Y. F. Hsu leg.); 1♂, Wulai, Fushan, 700m, 4. VII. 2005, reared from
B. nivea, emerged 15. VII. 2005 (C. L. Huang leg.); 1♀, Wulai, Fushan, 16. VI. 2005 (C. L.
 Huang leg.); 1♂, Wulai, Baoqing Temple, ca 620m, 4. VIII. 2005 (J. R. Chen leg.); 1♀, Xindian,
 Yinhedong, 250m, 14. VI. 2005 (Y. F. Hsu leg.); 8♂, 1♀, Pinglin, Yuguang, 2. XI. -11. XII 2005,
 reared from *B. formosana*, emerged 29. XI. /18. XII. 2005, HSU 05L5 (C. C. Lu leg.); 1♀,
 Pinglin, Zhongxinlun, 2. XI. 2005 (C. C. Lu leg.). **TAIPEI CITY**: 1♀, Beitou, Guizikeng, 24.
 X. 2004 (H. C. Huang & P. Lo leg.); 2♀, Wenshan, Xianjiyan, 16. V. 2005 (Y. F. Hsu leg.); 2♀,
 Daan, NTNU campus, 9. VI. 2005, reared from *B. nivea*, emerged 23-24. VI. 2005, HSU 05F18
 (Y. F. Hsu leg.); 5♀, Wenshan, Gongguan campus, NTNU, 9. VI. 2005, reared from *B. nivea*,
 emerged 21-24. VI. 2005, HSU 05F18 (Y. F. Hsu leg.); 3♀, same locality, 15. I. 2007, reared
 from *B. nivea*, emerged 31. I. 2007, HSU 07A20 (Y. F. Hsu leg.); 1♂, Shilin, Tianxiyuan, 17.
 VII. 2011 (H. C. Huang leg.); 1♂, Neihu, Dagouxi, 25. I. 2014 (L. Huang leg.). **TAOYUAN**
CITY [= TAOYUAN Co.]: 2♀, Fuxing, Kapu, 800m, 14. X. 2007, reared from *B. densiflora*,
 emerged 29. XI. 2007, HSU 07K7 (Y. F. Hsu & H. C. Huang leg.). **YILAN Co.**: 1♂, Yuanshan,

344 Fushan Botanical Garden, ca 700m, 4-5. VIII. 2006 (Y. F. Hsu & H. C. Huang leg.); 1♂, 1♀,
 345 Toucheng, Guishan Is., 4-5. III. 2006 (H. C. Huang & C. L. Huang leg.); 3♀, same locality, 24.
 346 VI. 2006 (H. C. Huang leg.); 1♀, same locality, 24. VI. 2006 (C. L. Huang leg.); 2♀, same
 347 locality, 4. VIII. 2007 (H. C. Huang leg.). **XINZHU Co.:** 1♂, Qionglin, Feifengshan, ca 80m,
 348 19. I. 2005, reared from *B. densiflora*, emerged 14. II. 2005, HSU 05A4 (Y. F. Hsu leg.); 1♂,
 349 Qionglin, Feifengshan, 10-80m, 19. I. 2005, reared from *B. nivea*, emerged 8. II. 2005, HSU
 350 05A3 (Y. F. Hsu leg.); 1♀, Zhudong, 11. IV. 2006 (Y. F. Hsu leg.); 2♀, Zhudong, 100m, 28. III.
 351 2008 (Y. F. Hsu leg.). **MIAOLI Co.:** 8♂, 2♀, Nanzhuang, 200m, 28. XI. 2004, reared from *B.*
 352 *densiflora*, emerged 11-20. XII. 2004, HSU 04L68 (Y. T. Lo leg.); 1♂, Touwu, Xiangshan
 353 Village, Jinshui, 60m, 27. II. 2005, reared from *B. nivea*, emerged 7. III. 2005, HSU 05B36 (Y.
 354 F. Hsu leg.); 5♂, 4♀, Sanwan, Emei Bridge, 14. I. 2006, reared from *B. nivea*, emerged 3-7. II.
 355 2006, HSU 06A12 (Y. F. Hsu leg.); 1♂, Zhuolan, Liyutan, 2. VII. 2006 (Y. F. Hsu leg.).
 356 **NANTOU Co.:** 2♂, 3♀, Lugu, 500m, 1. I. 2007, reared from *B. nivea*, emerged 17-31. I. 2007,
 357 HSU 07A6 (Y. F. Hsu leg.). **HUALIAN Co.:** 1♀, Wanrong, Wanrong Forest Trail, 29. VIII.
 358 2010 (Z. H. Yen leg.). **TAINAN CITY [= TAINAN Co.]:** 2♀, Xinhua, Dakeng, 80m, 10/11. II.
 359 2006 (Y. F. Hsu leg.); 1♀, Xinhua, 4. II. 2006, reared from *B. nivea*, emerged 13. III. 2006, HSU
 360 06J41 (Y. F. Hsu leg.), 2♂, same locality, 20. I. 2007 (Y. F. Hsu leg.). **KAOHSIUNG CITY:**
 361 3♂, 1♀, Jiashian Dist., Butingshan, ca 700m, 26. VII, 2022, reared from *Pouzolzia elegans*,

emerged 14. VIII. 2022, HSU 22G36 (Y. F. Hsu leg.). **LIANJIANG Co.:** 1♂, 1♀, Nangan, 10. VII. 2003, reared from *B. nivea*, emerged 25. VII. 2003, HSU 03G20 (L. W. Wu leg.); 1♂, Nangan, Shengli Dam, 9-10. XII. 2006, reared from *B. nivea*, emerged 6. I. 2007, HSU 06M2 (H. C. Huang leg.); 1♂, 1♀, Beigan, Qinbi, 22. IV. 2007 (Y. F. Hsu & H. C. Huang leg.).

Niche overlap between two *Symbrenthia* species in Taiwan

Many studies have reported that when a newly introduced species is present, competitive exclusion could be observed between the alien and similar native species (Mooney & Cleland, 2001; Paine et al., 2008; Muthukrishnan et al., 2018). However, our study presents a different aspect of this interaction. Both the SDM and the niche overlap results showed that the degree of overlap of the presenting area between these two species increases over time (Fig. 6). This means that competitive exclusion may not be present between the alien and the native species. This result may be explained by the following two alternative scenarios:

Firstly, obvious competition for the two species may not be observed due to insufficient time of introduction of the alien species. *S. lilaia* was first found on the main island of Taiwan as recently as 2004, and thus may still be in the process of population establishment and early growing stages (McGeoch & Jetz, 2019). Consequently, the competitive exclusion effect between these two species may have not yet occurred or not yet occurred to a level observable by

our available data.

Secondly, perhaps no competitive exclusion will occur between the two *Symbrenthia* species due to abundant host plant resources. Some studies have shown that host plants are much more important for the distribution of herbivorous insects when compared with the abiotic environmental factors (Wiens *et al.*, 2010; Simões & Peterson, 2018). These two butterfly species feed on several species in the family Urticaceae, and most of these host plants are common and abundant in Taiwan (Yang *et al.*, 1996). The food supply to the caterpillars of *Symbrenthia* may therefore be beyond the demand of both species combined, resulting in the absence of interspecific competition.

It awaits to be seen which scenario is more likely to occur, but it may be worthwhile to notice that *S. formosanus* was abundant in the Yangmingshan National Park of northern Taiwan (Chang, 1994), but in a butterfly survey conducted with sampling on monthly basis there from the beginning of 2021 to date, only *S. lilaia* has been recorded (Hsu *et al.*, unpublished data). This observation suggests competition between the two species may actually have initiated.

Distribution difference of *S. lilaia* between two different invasive stage

The SDM of *S. lilaia* shows different result patterns between the early and late invasive time stages (Fig. 5B & 5D). Especially in southwest portion of Taiwan, the presence probability

increases significantly in the late invasive stages. This phenomenon can be the result of the expansion of the distribution area of this alien species. The first record of *S. lilaea* is in Xinzhu county located in the northwest part of Taiwan. The distribution area of this species gradually expanded during the 18 years since it was first observed in Taiwan, and this butterfly species can today be observed in nearly all lowland areas around Taiwan.

The species distribution model (SDM) has been widely used as a tool to detect the potential invasive area of invasive species (*Wiens et al., 2010; Ahmed et al., 2020*). Based on the niche conservatism of the invasive species, we are usually able to predict the invasive area based on areas in the invaded region with similar environment to the source area from which the species originated. However, the SDM difference of *S. lilaea* between two different invasive stages suggests that the SDM may have inaccurately predicted potential areas of invasive species presence. The niche may be hard to quantify, even though some studies have suggested methods to measure it (*Framiout & Monnet, 2018; Lei et al., 2019*), but introduced species usually still undergo niche expansion in the newly invaded area (*Datta et al., 2019; Bates et al., 2020*).

Although SDM is still a widely used tool to evaluate potential impact of invasive species, the inaccuracy of the model is inevitable due to the reasons addressed above. We suggest to include more biotic factors of the invasive species when predicting potential invasive regions rather than relying on the SDM results alone. By combining the biotic variables with the SDM constructed

416 by the abiotic variables, the results should be closer to the realistic distribution pattern.

417

418 CONCLUSIONS

419 Species delimitation employing an integrative taxonomy approach has helped to clarify

420 taxonomic entities of an introduced and a native *Symbrenthia* butterfly taxa regarded conspecific

421 to date, leading to a decision to recognize each as a distinct species. This result suggests that

422 interspecific competition may occur by the introduction of the alien species, rather than gene

423 introgression. Subsequently, a niche modeling investigation was following, and the result

424 showed that competition between the two species interest has not yet occurred or just initiated.

425

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434

435 Data Availability

436 The following information was supplied regarding data availability:

437 Voucher specimens are housed at Department of Life Science, National Taiwan University,

438 Taipei, R. O. C. and Biodiversity Research Center, Academia Sinica, Taiwan, R. O. C. The *COI*

439 sequences described here are accessible via GenBank accession number LC727399 to

440 LC727411.

441

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Figure 1 Adult specimens of *Symbrenthia lilaea lunica* and *S. l. formosanus*. (A—B) Male specimen of *S. l. lunica*, TAIWAN: Miaoli County, Sanmwan, Emei Bridge. (C—D) Female specimen of *S. l. lunica*, TAIWAN: Xinzhu County, Zhudong, 100m. (E—F) Male specimen of *S. l. formosanus*, TAIWAN: New Taipei City, Wulai, 150m. (G—H) Female specimen of *S. l. formosanus*, TAIWAN: Taoyuan City, Fuxing, Daman. Scale bar = 1 cm (A—H).

Figure 2 Distribution points of *Symbrenthia lilaea formosanus* and *S. lilaea lunica* during two different invasive stages in Taiwan. (A) The distribution points of *S. l. formosanus* and *S. l. lilaea* in the early invasive stage (1911-2012). (B) The distribution points of *S. l. formosanus* and *S. l. lilaea* in the late invasive stage (2013-2021).

Figure 3 The Jackknife of regularized training gain of different bioclimatic factors of *Symbrenthia formosanus* and *S. lilaea* between different invasive time stages. (A) represents the jackknife results of *S. formosanus* in the early invasive stage. (B) represents the jackknife results of *S. lilaea* (=lunica) in the early invasive stage. (C) represents the jackknife results of *S. formosanus* in the late invasive stage. (D) represents the jackknife results of *S. lilaea* (=lunica) in the late invasive stage.

Figure 4 Systematic and species delimitation results of *Symbrenthia lilaea* and *S. formosanus*. (A) Phylogenetic tree reconstructed by BI analysis of 20 *COI* sequences of *Symbrenthia*. Nodal support from the BI posterior probabilities is along with ML bootstrap pseudo-replicates (BI posterior probabilities/MLbootstrap). Columns after taxon names correspond to the results of single-locus species delimitation analyses: Poisson Tree Processes model (PTP), Automatic Barcode Gap Discovery (ABGD) and Generalized Mixed Yule Coalecent model (GMYC). (B) PhyloMap visualization the result of PTP species delimitation, including 20 taxa of *Symbrenthia*. The first axis (horizontal) explained 62.64% and the second axis (vertical) explained 36.08% of sequence variance among samples.

Figure 5 The Species Distribution Model (SDM) of *Symbrenthia formosanus* and *S. lilaea* between different invasive time stages predicted from MaxEnt. (A) represents the SDM of *S. formosanus* in the early invasive stage. (B) represents the SDM of *S. lilaea* in the early invasive stage. (C) represents the SDM of *S. formosanus* in the late invasive stage. (D) represents the SDM of *S. lilaea* in the late invasive stage.

Figure 6 Shifts in climatic niche between alien (*Symbrenthia lilaea*) and native (*S. formosanus*) species. (A) The climatic niche shift between two *Symbrenthia* species in the early

invasive stage. (B) The climatic niche shift between two *Symbrenthia* species in the late invasive stage. Red describes the climatic niche of *S. lilaea* and blue describes the climatic niche of *S. formosanus*.

Figure 7 Genitalia of *Symbrenthia lilaea* and *S. formosanus*. (A—B) Male genitalia of *S. lilaea* (Gen. Prep. JYL-877, NTNU). (C) Female genitalia of *S. lilaea* (Gen. Prep. JYL-880, NTNU). (D—E) Male genitalia of *S. formosanus* (Gen. Prep. YFH-1583, NTNU). (F) Female genitalia of *S. formosanus* (Gen. Prep. JYL-1036, NTNU). Scale bar = 1mm (A—B; D—E), 1mm (C; F)

Figure 8 Eggs and immature stages of *Symbrenthia lilaea* and *S. formosanus*. (A) Egg mass of *S. lilaea*. (B) Young larvae of *S. lilaea*. (C) mature larvae of *S. lilaea*. (D) Egg of *S. formosanus*. (E) Young larvae of *S. formosanus*. (F) mature larvae of *S. formosanus*.

Figure S1 The niche overlap results infer from the 19 bioclimatic variables by principal component analyses (PCA). (A) The niche overlap results of the early invasive stage. (B) The niche overlap results of the late invasive stage.

632 **Figure S2 The receiver operating characteristic curve (ROC) and the threshold-independent**
 633 **area under the ROC curve (AUC) values of the SDM.** (A) the ROC and AUC of *S.*
 634 *formosanus* in the early invasive stage. (B) the ROC and AUC of *S. lilaea* (=lunica) in the early
 635 invasive stage. (C) the ROC and AUC of *S. formosanus* in the late invasive stage. (D) the ROC
 636 and AUC of *S. lilaea* (=lunica) in the late invasive stage.

637

638 **Table S1 pairwise distance of *Symbrenthia* samples**

639

640 **Table S2 *Symbrenthia* samples information**

Figure 1

Adult specimens of *Symbrenthia lilaea lunica* and *S. l. formosanus*

(A - B) Male specimen of *S. l. lunica*, TAIWAN: Miaoli County, Sanmwan, Emei Bridge. (C - D) Female specimen of *S. l. lunica*, TAIWAN: Xinzhu County, Zhudong, 100m. (E - F) Male specimen of *S. l. formosanus*, TAIWAN: New Taipei City, Wulai, 150m. (G - H) Female specimen of *S. l. formosanus*, TAIWAN: Taoyuan City, Fuxing, Daman. Scale bar = 1cm (A - H).



I

A



B



C



D



E



F



G



H

Figure 2

Distribution points of *Symbrenthia lilaea formosanus* and *S. lilaea lunica* during two different invasive stages in Taiwan

(A) The distribution points of *S. l. formosanus* and *S. l. lilaea* in the early invasive stage (1911-2012). (B) The distribution points of *S. l. formosanus* and *S. l. lilaea* in the late invasive stage (2013-2021).

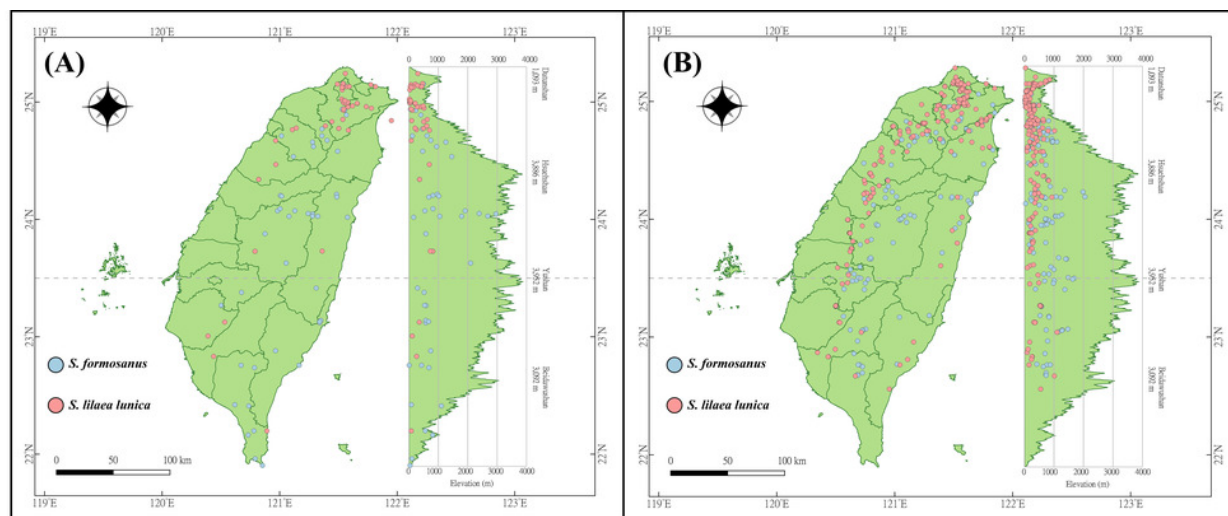


Figure 3

The Jackknife of regularized training gain of different bioclimatic factors of *Symbrenthia formosanus* and *S. lilaea* between different invasive time stages

(A) represents the jackknife results of *S. formosanus* in the early invasive stage. (B) represents the jackknife results of *S. lilaea* (=lunica) in the early invasive stage. (C) represents the jackknife results of *S. formosanus* in the late invasive stage. (D) represents the jackknife results of *S. lilaea* (=lunica) in the late invasive stage.

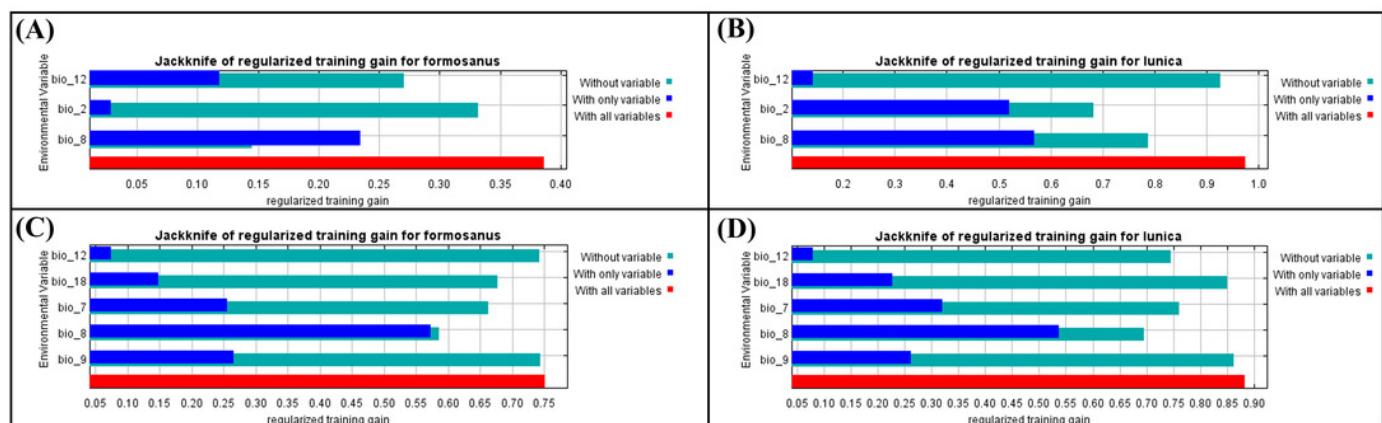
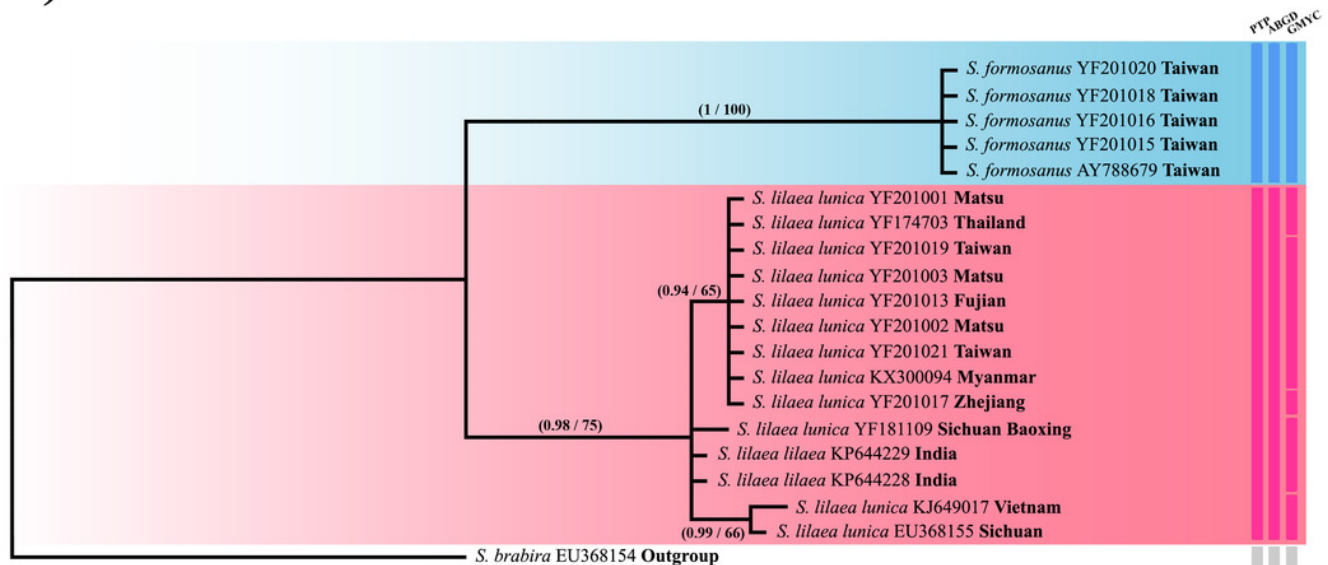


Figure 4

Systematic and species delimitation results of *Symbrenthia lilaea* and *S. formosanus*

(A) Phylogenetic tree reconstructed by BI analysis of 20 *COI* sequences of *Symbrenthia*. Nodal support from the BI posterior probabilities is along with ML bootstrap pseudo-replicates (BI posterior probabilities/MLbootstrap). Columns after taxon names correspond to the results of single-locus species delimitation analyses: Poisson Tree Processes model (PTP), Automatic Barcode Gap Discovery (ABGD) and Generalized Mixed Yule Coalecent model (GMYC). (B) PhyloMap visualization the result of PTP species delimitation, including 20 taxa of *Symbrenthia*. The first axis (horizontal) explained 62.64% and the second axis (vertical) explained 36.08% of sequence variance among samples.

(A)



(B)

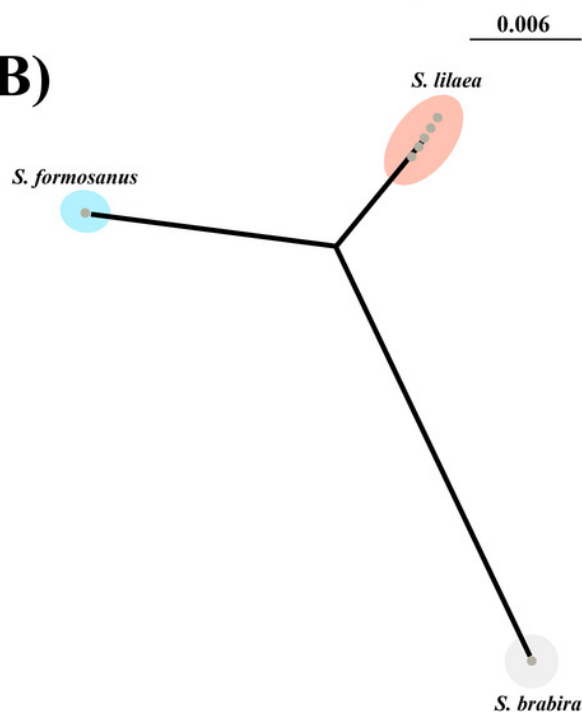


Figure 5

The Species Distribution Model (SDM) of *Symbrenthia formosanus* and *S. lilaea* between different invasive time stages predicted from MaxEnt

(A) represents the SDM of *S. formosanus* in the early invasive stage. (B) represents the SDM of *S. lilaea* in the early invasive stage. (C) represents the SDM of *S. formosanus* in the late invasive stage. (D) represents the SDM of *S. lilaea* in the late invasive stage.

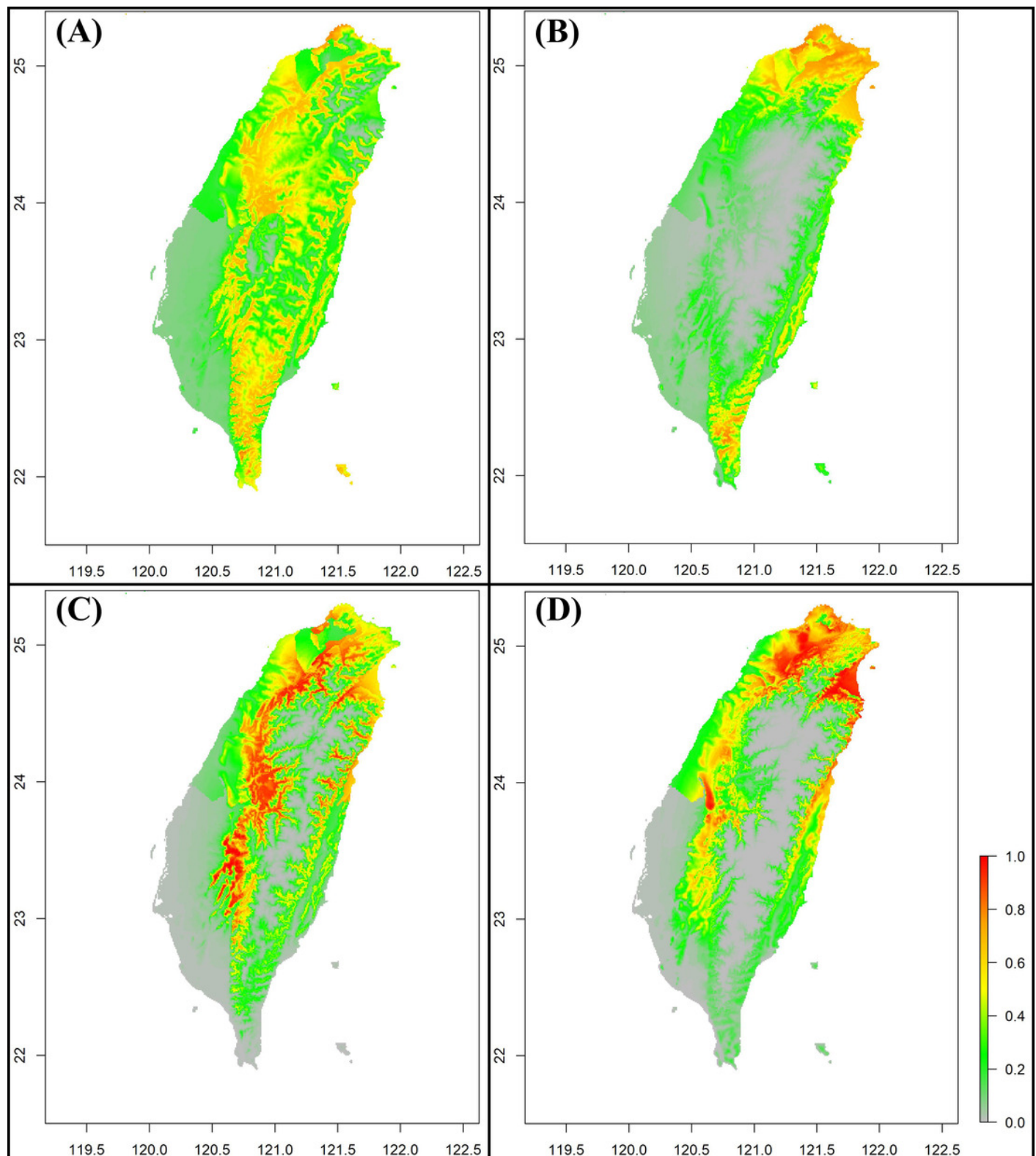


Figure 6

Shifts in climatic niche between alien (*Symbrenthia lila*) and native (*S. formosanus*) species

(A) The climatic niche shift between two *Symbrenthia* species in the early invasive stage. (B) The climatic niche shift between two *Symbrenthia* species in the late invasive stage. Red describes the climatic niche of *S. lila* and blue describes the climatic niche of *S. formosanus*.

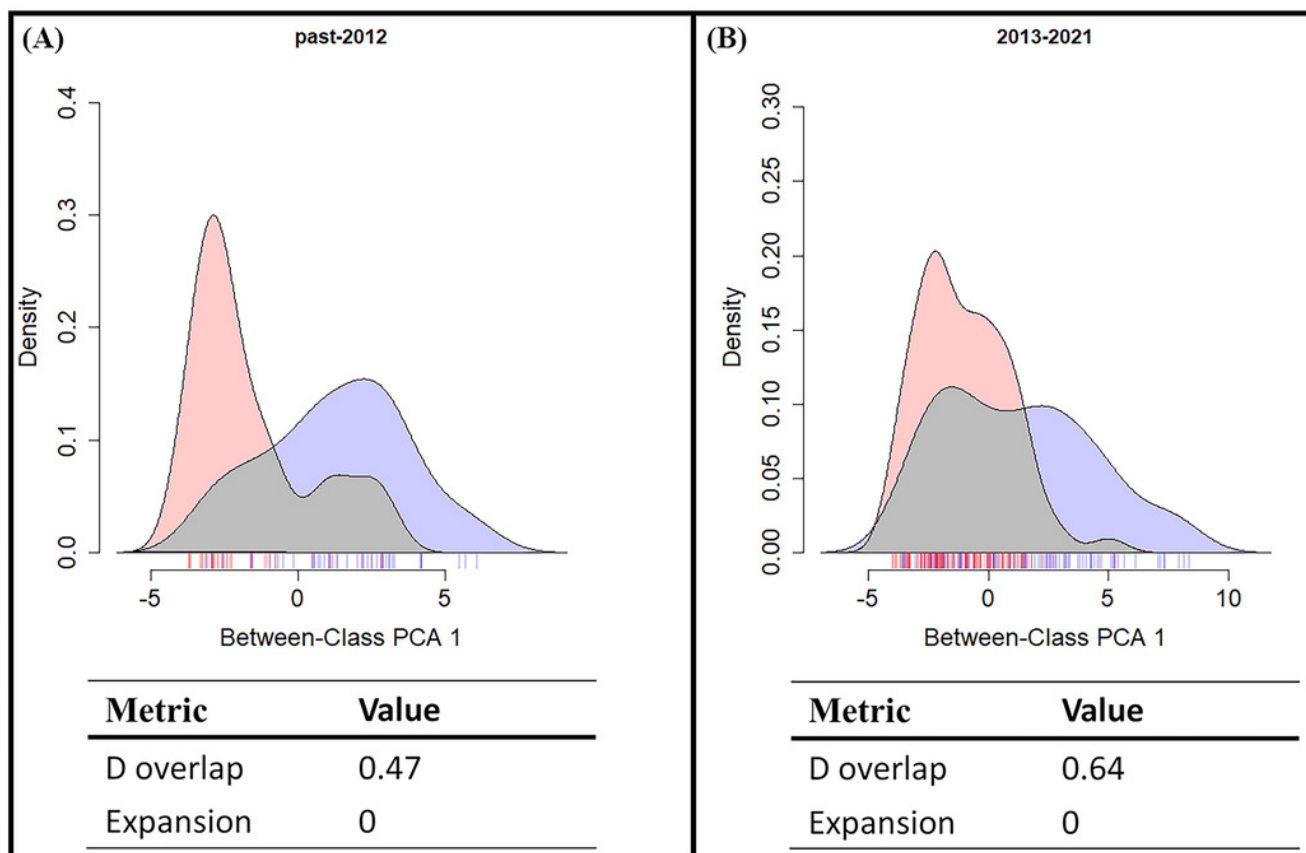


Figure 7

Genitalia of *Symbrenthia lilaea* and *S. formosanus*

(A - B) Male genitalia of *S. lilaea* (Gen. Prep. JYL-877, NTNU). (C) Female genitalia of *S. lilaea* (Gen. Prep. JYL-880, NTNU). (D - E) Male genitalia of *S. formosanus* (Gen. Prep. YFH-1583, NTNU). (F) Female genitalia of *S. formosanus* (Gen. Prep. JYL-1036, NTNU). Scale bar = 1mm (A - B; D - E), 1mm (C; F).

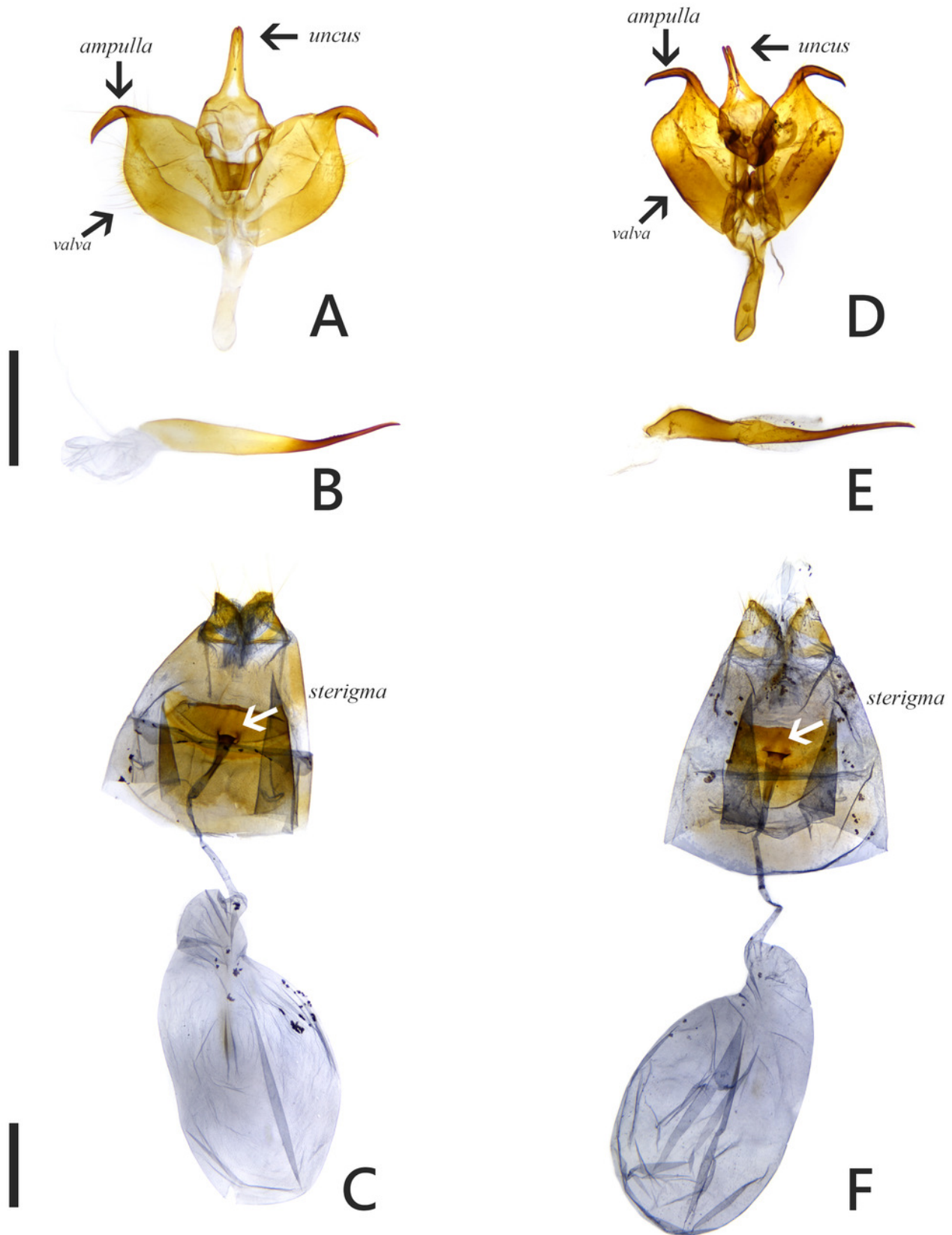


Figure 8

Eggs and immature stages of *Symbrenthia lilaea* and *S. formosanus*

(A) Egg mass of *S. lilaea*. (B) Young larvae of *S. lilaea*. (C) mature larvae of *S. lilaea*. (D) Egg of *S. formosanus*. (E) Young larvae of *S. formosanus*. (F) mature larvae of *S. formosanus*.

