

# Alien species revises systematic status: integrative species delimitation of two similar taxa of *Symbrenthia* Hübner, [1819] (Lepidoptera, Nymphalidae)

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Introduction of organisms to new range may impose detrimental effects on local organisms, especially when closely related species are involved. Species delimitation employing an integrative taxonomy approach may provide a quick assessment for the species status between taxa of interest, and to infer ecological competition and/or introgression that may be the associated with the introduction. A nymphalid butterfly, *Symbrenthia lilaea lunica*, was recently introduced to Taiwan, where a closely related local taxon, *S. l. formosanus*, can be found. We employed multiple species delimitation methods to study the species status between the two taxa, and the results revealed that they can be recognized as two distinct species. We further performed a niche modeling approach to investigate the ecological interaction between the two species. The taxonomic status of the two taxa, now elevated to species, has been revised and conservation facing rapid expansion of the introduced species discussed.

1 **Alien species revises systematic status: integrative species delimitation of two similar taxa**

2 **of *Symbrenthia* Hübner, [1819] (Lepidoptera, Nymphalidae)**

3

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19

20 **ABSTRACT**

21 Introduction of organisms to new range may impose detrimental effects on local organisms,  
22 especially when closely related species are involved. Species delimitation employing an  
23 integrative taxonomy approach may provide a quick assessment for the species status between  
24 taxa of interest, and to infer ecological competition and/or introgression that may be the  
25 associated with the introduction. A nymphalid butterfly, *Symbrenthia lilaea lunica*, was recently  
26 introduced to Taiwan, where a closely related local taxon, *S. l. formosanus*, can be found. We  
27 employed multiple species delimitation methods to study the species status between the two taxa,  
28 and the results revealed that they can be recognized as two distinct species. We further  
29 performed a niche modeling approach to investigate the ecological interaction between the two  
30 species. The taxonomic status of the two taxa, now elevated to species, has been revised and  
31 conservation facing rapid expansion of the introduced species discussed.

32

33 **Keywords** Systematics, Species delimitation, Biological introduction, Species interaction, Niche  
34 modeling

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## 38 **Introduction**

39 Biological invasions and range expansion of organisms usually impose unfavorable effects on  
40 local organisms that share similar ecological requirements (*Mooney & Cleland, 2001*),  
41 particularly when the expanding form is of continental origin and entering insular areas (*Sax &*  
42 *Gaines, 2008*). Competition may occur between native taxa and the invading one, notably when  
43 the counter taxa are closely related, where they are expected to share resource requirements  
44 (*Zwerschke et al., 2018*). This kind of scenarios have been well-documented in various  
45 organisms, such as ants (*von Aesch & Cherix, 2005*), birds (*Koenig, 2003*), molluscs (*Zwerschke*  
46 *et al., 2018*), plants (*Leger & Espeland, 2010; Čuda et al., 2015; Sheppard & Brendel, 2021*),  
47 etc., and even inspired E. O. Wilson to coin in his “taxon cycles” hypothesis (*Wilson, 1961*).  
48 Alternatively, if the involved taxa were previously allopatric populations or subspecies, gene  
49 flow may occur between taxa, resulting in changes in the genetic make-up of the local population  
50 (*Rhymer & Simberloff, 1996*). Consequently, whether the exotic taxon represents interspecific or  
51 intraspecific entity to the local taxon may lead to different types of ecological and evolutionary  
52 impacts to local community. Species delimitation employing an integrative taxonomy approach  
53 has becoming popular to help with taxonomic treatments/decisions among closely related taxa  
54 and to detect cryptic species (*Moraes et al., 2021*). This approach can also help with assessing  
55 whether an invading taxon is conspecific or not to the local taxon.

56 A case of recent invasion, either by anthropogenic introduction or by natural means such as  
57 wind, of a nymphalid butterfly currently recognized as *Symbrenthia lilaea lunica* Bascombe,  
58 Johnston & Bascombe, 1999 (Fig. 1A–1D) into Taiwan has been reported (Lu & Chen, 2014;  
59 Hsu et al., 2022). In Taiwan, there is a local native *Symbrenthia* Hübner, [1819] taxon, *S. l.*  
60 *formosanus* Fruhstorfer, 1908 (Fig. 1E–1H; Shirôzu & Ueda, 1992; Hsu et al., 2022). The taxon  
61 *lunica* Bascombe, Johnston & Bascombe, 1999 is a replacement name for *Papilio lucina* Stoll,  
62 1780; in Cramer, De uitlandsche kapellen 4(26b-28): 82 (Bascombe et al., 1999), distributed  
63 from northern Indochina to southern China according to Tsukada & Nishiyama (1985). Although  
64 some authors (e. g. Bozano & Floriani, 2012; Ek-Amnuay, 2012; Lang, 2012; Monastyrskii,  
65 2019) regarded *lucina* a synonym of the nominotypical *lilaea* described from India but have not  
66 formally synonymized these two names, others treat them as separate subspecies (e. g. Tsukada  
67 & Nishiyama, 1985; Hsu et al., 2022; Fric et al., 2022). *Symbrenthia lilaea lunica* from  
68 continental Asia was first documented in 2004 in Taiwan (Lu & Chen, 2014). Natural  
69 observations in Taiwan on immature biology and hostplant usage of both *S. l. lunica* and *S. l.*  
70 *formosanus* have documented that the two taxa use the same larval hostplants (Lu & Chen,  
71 2014). The invading or introduction event represents a ‘natural experiment’ in which to examine  
72 whether these two taxa are conspecific as current classification suggests, or if they represent  
73 distinct biological species. The answer to this natural experiment may help clarify what

74 ecological effect the introduced taxon may impose on the native taxa. If they turn out to  
75 represent distinct species, competition between them may occur as they do not differ in larval  
76 hostplant usage. If they turn out to be conspecific, gene introgression may occur since  
77 hybridization is liable to occur.

78 We investigated the species status between the two now sympatric taxa in Taiwan using  
79 multiple coalescent model based species delimitation methods and distance based species  
80 delineation, under a scheme for testing hypothesis of taxonomic status developed by *Braby et al.*  
81 (2012), following by a niche modeling survey to understand the ecological interaction between  
82 two taxa.

83

## 84 **Materials and Methods**

### 85 **Sampling**

86 The mitochondrial cytochrome oxidase subunit I (*COI*) gene has been successfully applied as a  
87 helpful marker with which to delimit closely related species (*Hebert et al., 2003*). We included a  
88 total of 13 specimens collected from various localities in our *COI*-based study. Four specimens  
89 of *S. l. formosanus* were collected around Taiwan, and six specimens of *S. l. lunica* were  
90 collected in Taiwan (mainland and Matsu archipelago), China, and Thailand. All samples were  
91 preserved in 70% ethanol and kept at -20°C for the subsequent molecular study. Moreover,

92 additional *COI* sequences of *S. l. formosanus* (AY788679) from Taiwan, *S. l. lunica* (EU368155,  
93 KJ649017, KX300094) from China, Vietnam, Myanmar respectively, and the nominate  
94 subspecies *S. l. lilaea* (Hewitson, 1864) (KP644228, KP644229) from India were obtained from  
95 GenBank. For the phylogenetic analyses, we used one sequence of *S. brabira* Moore, 1872  
96 (EU368154) as an outgroup, which was also obtained from GenBank.

97

#### 98 **DNA extraction, PCR amplification and DNA sequencing**

99 Genomic DNA was extracted from one leg of specimens using the Gentra Puregen tissue kit  
100 form QIAGEN (QIAGEN, Maryland, USA), following the manufacturer's protocol. A partial  
101 fragment from the *COI* gene was targeted for amplification by polymerase chain reaction (PCR).  
102 The *COI* gene was amplified using the universal primers COX-J-1460 (5'-  
103 TACAATTTATCGCCTAAACTTCAGCC-3') and COX-N-2191 (5'-  
104 CCCGGTAAAATTTAAAATATAAACTTC-3'). PCR reactions were performed in a 30µL  
105 volume eppendorf, containing 1µL of extracted DNA, 23.5µL of ddH<sub>2</sub>O, 3µL of 10X PCR  
106 reaction buffer, 0.6µL of each primer and 0.3µL of Power Taq (Genomics Biosci & Tech,  
107 Taiwan). The following PCR protocol was used: an initial denaturation at 95°C for 5 min,  
108 followed by 40 cycles of 30s denaturation at 95°C, 30s annealing at 50°C and 45s extension at  
109 72°C, and a final extension at 72°C for 10 min. Automatic sequencing was performed using an

110 ABI 3730XL DNA Analyzer (Applied Biosystems).

111

## 112 **Sequence Analyses and Phylogenetic Reconstruction**

113 Sequences were edited and assembled using Sequencher 4.10.1 (Gene Codes Corporation, Ann

114 Arbor, USA), and sequence alignments were performed using MUSCLE in MEGA 11 (*Tamura et*

115 *al.*, 2021), and pairwise genetic distances between different populations of *S. l. formosanus* and

116 *S. l. lunica* were also measured using MEGA 11 with the Kimura 2-parameter model.

117 The best-fit nucleotide substitution model for phylogenetic analysis was inferred using

118 jModelTest 2.1.10 (*Posada, 2008*) based on Akaike information criterion (AIC). Phylogenetic

119 trees were reconstructed under maximum likelihood (ML) and Bayesian inference (BI). ML

120 analysis was performed using RAxML v8.2.10 (*Stamatakis, 2014*) with 1000 bootstrap replicates

121 to assess the reliability of the tree. BI analysis was performed using MrBayes 3.2.6 (*Ronquist et*

122 *al.*, 2012). For MrBayes, the substitution model inferred from jModelTest was applied. The

123 Bayesian Markov Chain Monte Carlo (MCMC) analysis for  $10^9$  generations with sampling every

124 1000 generations was run to ensure the average standard deviation of split frequencies were

125 below 0.01. The first 30% of trees were discarded as burn-in. FigTree v1.4.4 was used to

126 visualize the consensus tree.

127

## 128 **Molecular species delimitation analyses**

129 Many molecular species delimitation programs have been proposed and broadly applied in  
130 speciation studies, which provides important evidence for integrative taxonomy. Among  
131 molecular species delimitation programs, the Poisson Tree Processes model (PTP) (*Zhang et al.*,  
132 2013), the Automatic Barcode Gap Discovery (ABGD) (*Puillandre et al.*, 2012), and the  
133 Generalized Mixed Yule Coalescent model (GMYC) (*Fujisawa & Barraclough*, 2013) were  
134 developed as single locus-based approaches for species delimitation. Therefore, we delineated  
135 species limits among *S. l. formosanus*, *S. l. lunica*, and *S. l. lilaea* by employing the Molecular  
136 Operational Taxonomic Unit concept set by these three programs.

137 For PTP, we used the tree inferred by MrBayes as input tree on the web server  
138 (<https://species.h-its.org/ptp/>), with 100000 MCMC generations and 100 thinning. Subsequently,  
139 PhyloMap was used to visualize the results of PTP. For ABGD, we performed the analyses on  
140 the web version of ABGD (<https://bioinfo.mnhn.fr/abi/public/abgd/>), with default settings of  
141 relative gap width ( $X=1.5$ ) and the Kimura two-parameter (K2P) model for nucleotide  
142 substitution. For GMYC, we used the phylogenetic tree inferred by MrBayes 3.2.6. The results  
143 from MrBayes were forced bifurcated by the “multi2di” and “chronos” function in the package  
144 “ape” in R 4.1.2. A single-threshold GMYC analysis was performed in the R package splits  
145 v1.0-20. We chose the single-threshold model because of the limited improvements of multiple-

146 threshold model.

147

#### 148 **Species distribution model of *S. l. formosanus* and *S. l. lunica* in Taiwan**

149 *Symbrenthia lilaea lunica* was not known to occur in Taiwan until recently, although it inhabits

150 Matzu and Kinmen, which are small outlying islands of Taiwan and close to mainland Asia.

151 However, *S.l. lunica* arrived to the main island of Taiwan due to anthropogenic activities or via

152 natural dispersal, with the first credible record found in Xinzhu in northwestern Taiwan in 2004

153 (*Lu & Chen, 2014*). Since then, the range of *S. l. lunica* has expended quickly, and is currently

154 found in lowland areas throughout Taiwan (*Lu & Chen, 2014; Hsu et al., 2022*). It is an

155 interesting issue whether competitive exclusion has happened between *S. l. lunica* and native *S. l.*

156 *formosanus*, especially if the species delimitation analyses decide they represent different

157 species.

158

159 The occurrence data of *S. l. formosanus* and *S. l. lunica* were obtained from the Global

160 Biodiversity Information Facility (GBIF) (<https://gbif.org/>, accessed 26 July 2021), Taiwan Moth

161 Information Center (<https://twmoth.terrsi.gov.tw/>, accessed 26 July 2021), and the specimen

162 collection at National Taiwan Normal University. To test the interaction between these two

163 species, we separated the occurrence data into two stages based on year. Because the first

164 documentation of *S. l. lunica* was in 2004, we divided the time period based on the median year  
165 (2012). The early invasion was defined as the data recorded from 1911–2012, and the late  
166 invasion was defined as the data recorded from 2013–2021. Repeated data was excluded using  
167 R 4.1.2, and we ensured the presences of only one presenting point in each raster to avoid  
168 overfitting. In total, 48 and 43 localities were obtained during the early invasion stage for *S. l.*  
169 *formosanus* and *S. l. lunica*, respectively (Fig. 2A), and 132 and 192 localities were obtained  
170 during the later invasion stage for *S. l. formosanus* and *S. l. lunica*, respectively (Fig. 2B). These  
171 data were organized using Microsoft Excel for the subsequent analyses.

172 A total 19 bioclimate variables (period: 1979–2013) were collected from CHELSA ([https://](https://chelsa-climate.org/)  
173 [chelsa-climate.org/](https://chelsa-climate.org/), accessed on 14 July 2021) at a spatial resolution of 30 arc-seconds (1km<sup>2</sup>).  
174 These bioclimate variables were derived from temperature and precipitation, which are  
175 considered to be related to the distribution and survival of small arthropods and have been  
176 widely used in the prediction of species distribution (*De Meyer et al., 2010; Xu et al., 2020*). In  
177 order to avoid the effect of multicollinearity, these 19 variables were selected by the “vifstep”  
178 and “vifcor” function with the threshold of 10 and 0.6 separately in “usdm” package in R 4.1.2  
179 (selected variables shown in Fig. 3).

180 MaxEnt (3.4.4) (*Phillips et al., 2006*) was applied to predict the habitat suitability of *S. l.*  
181 *formosanus* and *S. l. lunica* based on the occurrence data. 10% of the data were selected to run a

182 random test and the remaining data were run following the default settings. Presence-only data  
183 were generated pseudo-absences and 10000 random background points were randomly selected  
184 by the MaxEnt model. The results were output after 10 cross-validation replicates.

185       The predictions generated from MaxEnt modeling were evaluated according to the  
186 threshold independent area under the receiver operating characteristic (ROC) curve (AUC)  
187 values. ROC curves were used to plot the true-positive rate against the false-positive rate and the  
188 AUC was used as a measure of the goodness of fit of the model. The AUC value ranges from 0  
189 to 1, with higher values indicating higher predictive performance. The logistic output was chosen  
190 as an estimate of the probability of presence conditioned by bio-environmental variables per grid  
191 cell. Jackknifing was used to screen for the contribution of each bio-environmental variable used  
192 in the model.

193       We performed principal component analyses (PCA) to test the niche overlap of these two  
194 species in both the early invasion stage and late invasion stage. The 19 bioclimatic variables  
195 were obtained from the CHELSA database based on the GPS of each observation point. The  
196 analyses were conducted in R 4.1.2 using the function “prcomp”, with scatterplots built using the  
197 function “ggbiplot”. Additionally, in order to evaluate the niche shift pattern between the two  
198 *Symbrenthia* species in Taiwan, we apply methods modified from *Bates et al. (2020)* to quantify  
199 the niche shift between *S. l. formosanus* and *S. l. lunica* by calculating niche overlap, presented

200 by Schoener's D, and niche expansion of *S. l. lunica*.

201

## 202 **Results**

### 203 **Taxonomic decisions**

204 Phylogenetic reconstruction of *Symbrenthia* COI samples (Fig. 4) revealed that all samples of  
205 *lunica+lilaea* form a monophyletic group sister to *formosanus* samples, which also formed a  
206 monophyletic group. The p-distance was 0.0017 between *lunica* and *lilaea* and 0.0505–0.0525  
207 between *lunica+lilaea* and *formosanus*. PTP, ABGD and GMYC all recognize a two species  
208 scenario, with *lunica+lilaea* and *formosanus* each representing a distinct species. Therefore,  
209 *formosanus* is recognized as a species distinct from *lunica+lilaea*, with the combination as  
210 *Symbrenthia formosanus* Fruhstorfer, 1908 (**stat. rev.**). The taxon *S. l. lunica* Bascombe,  
211 Johnston & Bascombe, 1999 is proposed to be lumped with *S. l. lilaea*, Hewitson, 1864 (**syn.**  
212 **nov.**) herein as the two may not be distinguished by COI barcode nor adult and immature  
213 morphology. We thus will call them *S. formosanus* and *S. lilaea* respectively in the remaining  
214 text of this article.

215

### 216 **Environmental factors which contribute to the distribution of *Symbrenthia* species in**

### 217 **question**

218 After applying "usdm" package to remove those highly correlated factors, "bio 2", "bio 8", "bio  
219 12" (Fig. 3A & 3B) are used to construct the species distribution model of "early invasive stage",  
220 while "bio 7", "bio 8", "bio 9", "bio 12", "bio 18" (Fig. 3C & 3D) are used to construct the  
221 species distribution model of "late invasive stage". According to the results of the jackknife test,  
222 the factors show different contribution patterns in the early invasive stage. In the early invasive  
223 stage, "bio 2" (annual precipitation) and "bio 12" (air temperature) contribute reversely between  
224 these two species; annual precipitation contributes more than mean diurnal air temperature range  
225 in the distribution model of *S. formosanus*, whereas mean diurnal air temperature range  
226 contributes more than annual precipitation amount in the model of *S. lilaea*.

227 Comparing the jackknife results of both species between the two invasive stages, "bio 8"  
228 contributes the most among all models. According to this, the mean daily air temperatures of the  
229 wettest quarter may play a key role in the distribution of these two *Symbrenthia* species in  
230 Taiwan.

231

### 232 **The species distribution model and niche shifting of the two *Symbrenthia* species in** 233 **different time stages**

234 According to the species distribution model results, *S. formosanus* does not show an obvious  
235 change between the early and late invasive stages (Fig. 5A & 5C). For both invasive stages, the

236 presence probability of *S. formosanus* seems to be higher in the suburban areas and places with  
237 less human activity. For *S. lilaea*, the distribution model presents different results between the  
238 two time stages (Fig. 5B & 5D). Particularly, presence probability in the southwest part of  
239 Taiwan is higher in the later invasive stage (Fig. 5D). The SDM results of both species show that  
240 the presence probability decreases in the Pingtung area, the southernmost county of Taiwan.  
241 Although there may be biological importance to this observation, it is most likely a result of  
242 uneven presence observation point density in the later stage. Most of the presence points for the  
243 late invasive stage SDM are from northern Taiwan.

244 From the results of the early and late invasive stages (Fig 6), the niche overlap value  
245 increased during the recent years (past-2012 D: 0.48; 2013-2021 D: 0.64), and the niche  
246 expansion value of *S. lilaea* remained zero between the two different time stages. Together, these  
247 mean that, during these two periods of time, the niche of this alien species did not extend beyond  
248 the niche of the native species. According to the ENM model and the niche shift results,  
249 competitive exclusion seems to not be occurring between these two species over these 18 years.

250

## 251 **Discussion**

### 252 **Taxonomic status of the introduced and native *Symbrenthia* butterflies**

253 The introduced and native *Symbrenthia* butterflies in question of the study were regarded as

254 conspecific subspecies prior to the present study (e. g., *Hsu et al., 2022; Fric et al., 2022*). It has  
255 been argued that species delimitation is difficult for allopatric populations or subspecies of  
256 similar forms (*King, 1993; Braby et al., 2012*), but in the present case, the introduction of  
257 continental *S. lilaea* to Taiwan has proven that insular *S. formosanus* ought to represent a species  
258 endemic to the island, instead of being a geographical race of the former. Distinctions between  
259 them include: 1) distal band on hindwing uppersides of both sexes form a continuous orange  
260 stripe in *S. lilaea* (Fig. 1A & 1C), whereas it is interrupted by darkened veins in *S. formosanus*  
261 (Fig. 1E & 1G); 2) distal tip of uncus is acute in *S. lilaea* (Fig. 7A), whereas it is blunt in *S.*  
262 *formosanus* (Fig. 7D); 3) distal margin of valva is rounded in *S. lilaea* (Fig. 7A), whereas it is  
263 angled, somewhat squared in *S. formosanus* (Fig. 7D); 4) ampulla is stout, slightly down-curved  
264 in *S. lilaea* (Fig. 7A), whereas it is slender, strongly bent downwards in *S. formosanus* (Fig.  
265 7D); 5) posterior margin of sterigma is concave in *S. lilaea* (Fig. 7C), whereas it is truncate in *S.*  
266 *formosanus* (Fig. 7F); 6) yellow eggs are laid in cluster in *S. lilaea* (Fig. 8A), in contrast to  
267 green eggs laid singly in *S. formosanus* (Fig. 8D); 7) larvae are gregarious in *S. lilaea* (Fig. 8B-  
268 8C), but solitary in *S. formosanus* (Fig. 8E–8F).

269

270 **Vouchers.**

271 ***Symbrenthia formosanus* Fruhstorfer, 1908**

- 272 *Ssp. formosanus*: **XINBEI CITY [= TAIPEI Co.]**: 1♀, Xindian, Sikanshui, 5. X. 2001 (H. S.  
273 Que leg.); 1♂, Wulai, 150m, reared from *Boehmeria nivea*, emerged 8. XI. 2004, HSU 04L21 (Y.  
274 F. Hsu leg.); 1♂, Wulai, 13. II. 2005, reared from *Debregeasia orientalis*, emerged 12. III. 2005,  
275 HSU 05B6 (L. H. Wang leg.); 1♀, same locality, 5. VI. 2005, reared from *B. densiflora*, emerged  
276 23. VI. 2005, HSU 05F9 (Y. F. Hsu leg.); 1♂, same locality, 5. VI. 2005, reared from *D.*  
277 *orientalis*, emerged 5. VI. 2005, HSU 05F10 (Y. F. Hsu leg.); 1♂, same locality, 13. VI. 2005,  
278 reared from *B. nivea*, emerged 2. VII. 2005, HSU 05F36 (Y. F. Hsu leg.); 1♀, same locality, 16.  
279 I. 2007 (Y. F. Hsu leg.); 1♂, Wulai, ca 200m, 5. VI. 2005 (Y. F. Hsu leg.). **TAIPEI CITY**: 1♀,  
280 Daan, Baozangyan, 28. III. 2019 (L. Huang leg.). **TAOYUAN CITY [= TAOYUAN Co.]**: 1♂,  
281 Fuxing, Xuanyuan, ca 1000m, 16. VI. 2005 (L. H. Wang leg.); 1♀, Fuxing, Daman, 10. VI.  
282 2005, reared from *B. nivea*, emerged 22-23. VI. 2005, HSU 05F28 (Y. F. Hsu leg.); 1♀, same  
283 locality, 10. VI. 2005, reared from *D. orientalis*, emerged 23. VI. 2005, HSU 05F29 (Y. F. Hsu  
284 leg.); 1♀, Fuxing, Gaoyi, 600m, 7. III. 2008 (Y. F. Hsu leg.). **YILAN Co.**: 1♂, 2♀, Nanao,  
285 14Km S. Nanao, 6. VI. 2005, reared from *B. nivea*, emerged 8-17. VI. 2005, HSU 05F17 (Y. F.  
286 Hsu leg.); 3♂, 1♀, Datong, Qilan, 1. XII. 2005, reared from *B. nivea*, emerged III. 2006, HSU  
287 05M7 (C. C. Lu & H. Y. Lee leg.). **XINZHU Co.**: 2♂, Jianshi, Xiuluan, 800m, 21. VI. 2005 (L.  
288 H. Wang leg.); 1♀, Jianshi, Guanwu, Dalu Forest Trail, 2. X. 2006 (L. H. Wang leg.); 1♂,  
289 Jianshi, Yulao, ca 1500m, 25. II. 2009 (C. K. Wang leg.). **MIAOLI Co.**: 1♂, Zhuolan, Liyutan,

290 300m, 2. VII. 2006 (Y. F. Hsu leg.). **TAIZHONG CITY [= TAIZHONG Co.]**: 1♀, Heping,  
291 Shangguguan, 900m, 30. X. 2006 (Y. F. Hsu leg.); 1♂, same locality, 29. VII. 2007 (Y. F. Hsu  
292 leg.); 1♂, Heping, Guguan, 700m, 5. VII. 1998 (Y. T. Lo leg.); 1♂, Heping, Guguan, IV. 2005  
293 (L. H. Wang leg.). **NANTOU Co.**: 1♂, Renai, Tunyuan/ Tianchi, 2000/2800m, 31. VII- 1. VIII.  
294 1998 (Y. T. Lo leg.); 1♂, 1♀, Renai, Meiyuan, ca 400m, 22. XI. 2004, reared from *B. nivea*,  
295 emerged 16-22. XII. 2004, HSU 04L49 (Y. F. Hsu leg.); 1♀, Yuchi, Shiguanyin, 480m, 22. XI.  
296 2004, reared from *B. densiflora*, emerged 21. XII. 2004, HSU 04L52 (Y. F. Hsu leg.); 1♂, Lugu,  
297 Fenghuanggu, 600m, 4. XII. 2004, reared from *B. densiflora*, emerged 23. XII. 2004, HSU 04M5  
298 (Y. F. Hsu leg.); 1♂, 3♀, Lugu, Fenghuanggu, 4. XII. 2004, reared from *B. nivea*, emerged 17-  
299 20. XII. 2004, HSU 04M4 (Y. F. Hsu leg.), 1♀, same locality, 11. XII. 2004, reared from *B.*  
300 *densiflora*, emerged 14. II. 2005, HSU 04M14 (Y. F. Hsu leg.), 1♂, same locality, 12. XII. 2004  
301 (Y. F. Hsu leg.); 1♂, Renai, Beidongyanshan, ca 1800m, 17. XII. 2005 (L. W. Wu & Y. F. Hsu  
302 leg.) (genitalia preparation YFH 1583), 1♀, same locality, 17. XII. 2005, reared from *D.*  
303 *orientalis*, emerged 4. I. 2006, HSU 05M32 (Y. F. Hsu leg.); 3♂, Renai, Tunyuan, 1550m, 28.  
304 VII. 2006 (Y. F. Hsu leg.); 1♀, Renai, Nanshanxi, ca 900m, 1. IX. 2007 (Y. F. Hsu leg.); 1♂,  
305 Lugu, ca 600m, 19. VII. 2006, reared from *B. nivea*, emerged 4. VIII. 2006, HSU 06G12 (Y. F.  
306 Hsu leg.); 1♂, 2♀, Lugu, ca 500m, 1. I. 2007, reared from *B. nivea*, emerged 18. I. 2007, HSU  
307 07A7 (Y. F. Hsu leg.); 1♀, Renai, Huisun, 600-700m, 20. II. 2010 (Y. F. Hsu & H. C. Huang

308 leg.). **HUALIAN Co.:** 2♂, Xiulin, Huitouwan, 23. VII. 2005 (Y. F. Hsu leg.), 1♂, same locality,  
 309 30. I. 2007 (Y. F. Hsu leg.); 4♂, 3♀, Xiulin, Zuocang Trail, 7. III. 2006 (Y.F. Hsu leg.); 1♀,  
 310 Xiulin, Guangbeibabiao, 2200m, 28. V. 2007 (Y. F. Hsu leg.); 1♀, Xiulin, Huoranting, 1000m,  
 311 10. XII. 2007 (M. H. Sun leg.); 1♂, Xiulin, Wujiabengshan, 2000m, 7. VI. 2008 (L.H. Wang  
 312 leg.); 1♀, Zhouxi, Zhongping Forest Trail, 1. VI. 2007 (L. H. Wang leg.). **TAINAN CITY [=**  
 313 **TAINAN Co.]:** 1♀, Dongshan, Kantoushan, 600-800m, 4. II. 2010 (Y. F. Hsu leg.).  
 314 **PINGDONG Co.:** 1♂, Shizi, Nunaishan, 200m, 26. III. 2006 (Y. F. Hsu leg.); 1♂, Fangliao,  
 315 Yuquan, 26. II. 2006 (Y. F. Hsu leg.); 2♂, 1♀, Shizi, Lilongshan, 500m, 17. II. 2008 (Y. F. Hsu  
 316 leg.); 1♂, same locality, 1. III. 2008 (Y. F. Hsu leg.); 1♀, Wutai, Jiudawu, 500m, 11. V. 2008 (C.  
 317 L. Huang leg.); 1♀, Sandimen, ca 300m, 2. I. 2009 (Y. F. Hsu leg.). **TAIDONG Co.:** 1♀,  
 318 Yanping, Hongye, 11. I. 2009 (C. H. Lin, jr leg.); 1♀, Taidong, Pipa Lake, 4. VI. 2006 (Y. F.  
 319 Hsu leg.); 1♀, Haiduan, Liyuan, 7. VI. 2020 (Y. F. Hsu leg.) (genitalia preparation JYL1036).  
 320 ***Symbrenthia lilaea* (Hewitson, 1864)**  
 321 **JILONG CITY:** 1♀, Longgang Trai, 26. IX. 2006 (Y. F. Hsu leg.); 2♀, same locality, 2. X.  
 322 2006 (Y. F. Hsu leg.); 1♀, same locality, 7. II. 2007 (C. K. Wang leg.); 1♂, Shenao, Sea Level,  
 323 20. VI. 2006, reared from *B. nivea*, emerged 2. VII. 2006, HSU 06F17 (Y. F. Hsu leg.); 2♂,  
 324 Haimentianxian, 26. IX. 2006 (Y. F. Hsu leg.). **XINBEI CITY [= TAIPEI Co.]:** 7♂, 7♀, Wulai,  
 325 150m, 8. XI. 2004, reared from *B. nivea*, emerged 29. XI-5. XII. 2004, HSU 04L20 (Y. F. Hsu

326 leg.); 1♂, Wulai, 11. III. 2005 (C. R. Chen leg.), 1♀, same locality, 5. VI. 2005 (Y. F. Hsu leg.),  
327 1♂, 1♀, same locality, 11. VIII. 2005, reared from *B. densiflora*, emerged 25. VIII. -1. IX. 2005,  
328 HSU 05H8 (Y. F. Hsu leg.), 1♂, same locality, 23. IX. 2005, reared from *B. nivea*, emerged 4. X.  
329 2005, HSU 05J72 (J. R. Chen leg.), 3♀, same locality, 6. X. 2006, reared from *B. nivea*, emerged  
330 3-15. XI. 2006, HSU 06L3 (Y. F. Hsu leg.); 1♂, Wulai, Fushan, 700m, 4. VII. 2005, reared from  
331 *B. nivea*, emerged 15. VII. 2005 (C. L. Huang leg.); 1♀, Wulai, Fushan, 16. VI. 2005 (C. L.  
332 Huang leg.); 1♂, Wulai, Baoqing Temple, ca 620m, 4. VIII. 2005 (J. R. Chen leg.); 1♀, Xindian,  
333 Yinhedong, 250m, 14. VI. 2005 (Y. F. Hsu leg.); 8♂, 1♀, Pinglin, Yuguang, 2. XI. -11. XII. 2005,  
334 reared from *B. formosana*, emerged 29. XI. /18. XII. 2005, HSU 05L5 (C. C. Lu leg.); 1♀,  
335 Pinglin, Zhongxinlun, 2. XI. 2005 (C. C. Lu leg.). **TAIPEI CITY**: 1♀, Beitou, Guizikeng, 24.  
336 X. 2004 (H. C. Huang & P. Lo leg.); 2♀, Wenshan, Xianjiyan, 16. V. 2005 (Y. F. Hsu leg.); 2♀,  
337 Daan, NTNU campus, 9. VI. 2005, reared from *B. nivea*, emerged 23-24. VI. 2005, HSU 05F18  
338 (Y. F. Hsu leg.); 5♀, Wenshan, Gongguan campus, NTNU, 9. VI. 2005, reared from *B. nivea*,  
339 emerged 21-24. VI. 2005, HSU 05F18 (Y. F. Hsu leg.); 3♀, same locality, 15. I. 2007, reared  
340 from *B. nivea*, emerged 31. I. 2007, HSU 07A20 (Y. F. Hsu leg.); 1♂, Shilin, Tianxiyuan, 17.  
341 VII. 2011 (H. C. Huang leg.); 1♂, Neihu, Dagouxi, 25. I. 2014 (L. Huang leg.). **TAOYUAN**  
342 **CITY [= TAOYUAN Co.]**: 2♀, Fuxing, Kapu, 800m, 14. X. 2007, reared from *B. densiflora*,  
343 emerged 29. XI. 2007, HSU 07K7 (Y. F. Hsu & H. C. Huang leg.). **YILAN Co.**: 1♂, Yuanshan,

- 344 Fushan Botanical Garden, ca 700m, 4-5. VIII. 2006 (Y. F. Hsu & H. C. Huang leg.); 1♂, 1♀,
- 345 Toucheng, Guishan Is., 4-5. III. 2006 (H. C. Huang & C. L. Huang leg.); 3♀, same locality, 24.
- 346 VI. 2006 (H. C. Huang leg.); 1♀, same locality, 24. VI. 2006 (C. L. Huang leg.); 2♀, same
- 347 locality, 4. VIII. 2007 (H. C. Huang leg.). **XINZHU Co.:** 1♂, Qionglin, Feifengshan, ca 80m,
- 348 19. I. 2005, reared from *B. densiflora*, emerged 14. II. 2005, HSU 05A4 (Y. F. Hsu leg.); 1♂,
- 349 Qionglin, Feifengshan, 10-80m, 19. I. 2005, reared from *B. nivea*, emerged 8. II. 2005, HSU
- 350 05A3 (Y. F. Hsu leg.); 1♀, Zhudong, 11. IV. 2006 (Y. F. Hsu leg.); 2♀, Zhudong, 100m, 28. III.
- 351 2008 (Y. F. Hsu leg.). **MIAOLI Co.:** 8♂, 2♀, Nanzhuang, 200m, 28. XI. 2004, reared from *B.*
- 352 *densiflora*, emerged 11-20. XII. 2004, HSU 04L68 (Y. T. Lo leg.); 1♂, Touwu, Xiangshan
- 353 Village, Jinshui, 60m, 27. II. 2005, reared from *B. nivea*, emerged 7. III. 2005, HSU 05B36 (Y.
- 354 F. Hsu leg.); 5♂, 4♀, Sanwan, Emei Bridge, 14. I. 2006, reared from *B. nivea*, emerged 3-7. II.
- 355 2006, HSU 06A12 (Y. F. Hsu leg.); 1♂, Zhuolan, Liyutan, 2. VII. 2006 (Y. F. Hsu leg.).
- 356 **NANTOU Co.:** 2♂, 3♀, Lugu, 500m, 1. I. 2007, reared from *B. nivea*, emerged 17-31. I. 2007,
- 357 HSU 07A6 (Y. F. Hsu leg.). **HUALIAN Co.:** 1♀, Wanrong, Wanrong Forest Trail, 29. VIII.
- 358 2010 (Z. H. Yen leg.). **TAINAN CITY [= TAINAN Co.]:** 2♀, Xinhua, Dakeng, 80m, 10/11. II.
- 359 2006 (Y. F. Hsu leg.); 1♀, Xinhua, 4. II. 2006, reared from *B. nivea*, emerged 13. III. 2006, HSU
- 360 06J41 (Y. F. Hsu leg.), 2♂, same locality, 20. I. 2007 (Y. F. Hsu leg.). **KAOHSIUNG CITY:**
- 361 3♂, 1♀, Jiashian Dist., Butingshan, ca 700m, 26. VII, 2022, reared from *Pouzolzia elegans*,

362 emerged 14. VIII. 2022, HSU 22G36 (Y. F. Hsu leg.). **LIANJIANG Co.:** 1♂, 1♀, Nangan, 10.  
363 VII. 2003, reared from *B. nivea*, emerged 25. VII. 2003, HSU 03G20 (L. W. Wu leg.); 1♂,  
364 Nangan, Shengli Dam, 9-10. XII. 2006, reared from *B. nivea*, emerged 6. I. 2007, HSU 06M2  
365 (H. C. Huang leg.); 1♂, 1♀, Beigan, Qinbi, 22. IV. 2007 (Y. F. Hsu & H. C. Huang leg.).

366

### 367 **Niche overlap between two *Symbrenthia* species in Taiwan**

368 Many studies have reported that when a newly introduced species is present, competitive  
369 exclusion could be observed between the alien and similar native species (*Mooney &*  
370 *Cleland, 2001; Paini et al., 2008; Muthukrishnan et al., 2018*). However, our study presents a  
371 different aspect of this interaction. Both the SDM and the niche overlap results showed that the  
372 degree of overlap of the presenting area between these two species increases over time (Fig. 6).  
373 This means that competitive exclusion may not be present between the alien and the native  
374 species. This result may be explained by the following two alternative scenarios:

375 Firstly, obvious competition for the two species may not be observed due to insufficient  
376 time of introduction of the alien species. *S. lilaea* was first found on the main island of Taiwan as  
377 recently as 2004, and thus may still be in the process of population establishment and early  
378 growing stages (*McGeoch & Jetz, 2019*). Consequently, the competitive exclusion effect  
379 between these two species may have not yet occurred or not yet occurred to a level observable by

380 our available data.

381 Secondly, perhaps no competitive exclusion will occur between the two *Symbrenthia*  
382 species due to abundant host plant resources. Some studies have shown that host plants are much  
383 more important for the distribution of herbivorous insects when compared with the abiotic  
384 environmental factors (*Wiens et al., 2010; Simões & Peterson, 2018*). These two butterfly  
385 species feed on several species in the family Urticaceae, and most of these host plants are  
386 common and abundant in Taiwan (*Yang et al., 1996*). The food supply to the caterpillars of  
387 *Symbrenthia* may therefore be beyond the demand of both species combined, resulting in the  
388 absence of interspecific competition.

389 It awaits to be seen which scenario is more likely to occur, but it may be worthwhile to  
390 notice that *S. formosanus* was abundant in the Yangmingshan National Park of northern Taiwan  
391 (*Chang, 1994*), but in a butterfly survey conducted with sampling on monthly basis there from  
392 the beginning of 2021 to date, only *S. lilaea* has been recorded (*Hsu et al., unpublished data*).  
393 This observation suggests competition between the two species may actually have initiated.

394

#### 395 **Distribution difference of *S. lilaea* between two different invasive stage**

396 The SDM of *S. lilaea* shows different result patterns between the early and late invasive time  
397 stages (Fig. 5B & 5D). Especially in southwest portion of Taiwan, the presence probability

398 increases significantly in the late invasive stages. This phenomenon can be the result of the  
399 expansion of the distribution area of this alien species. The first record of *S. lilaea* is in Xinzhu  
400 county located in the northwest part of Taiwan. The distribution area of this species gradually  
401 expanded during the 18 years since it was first observed in Taiwan, and this butterfly species can  
402 today be observed in nearly all lowland areas around Taiwan.

403       The species distribution model (SDM) has been widely used as a tool to detect the potential  
404 invasive area of invasive species (*Wiens et al., 2010; Ahmed et al., 2020*). Based on the niche  
405 conservatism of the invasive species, we are usually able to predict the invasive area based on  
406 areas in the invaded region with similar environment to the source area from which the species  
407 originated. However, the SDM difference of *S. lilaea* between two different invasive stages  
408 suggests that the SDM may have inaccurately predicted potential areas of invasive species  
409 presence. The niche may be hard to quantify, even though some studies have suggested methods  
410 to measure it (*Framiout & Monnet, 2018; Lei et al., 2019*), but introduced species usually still  
411 undergo niche expansion in the newly invaded area (*Datta et al., 2019; Bates et al., 2020*).

412 Although SDM is still a widely used tool to evaluate potential impact of invasive species, the  
413 inaccuracy of the model is inevitable due to the reasons addressed above. We suggest to include  
414 more biotic factors of the invasive species when predicting potential invasive regions rather than  
415 relying on the SDM results alone. By combining the biotic variables with the SDM constructed

416 by the abiotic variables, the results should be closer to the realistic distribution pattern.

417

## 418 **CONCLUSIONS**

419 Species delimitation employing an integrative taxonomy approach has helped to clarify

420 taxonomic entities of an introduced and a native *Symbrenthia* butterfly taxa regarded conspecific

421 to date, leading to a decision to recognize each as a distinct species. This result suggests that

422 interspecific competition may occur by the introduction of the alien species, rather than gene

423 introgression. Subsequently, a niche modeling investigation was following, and the result

424 showed that competition between the two species interest has not yet occurred or just initiated.

425

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434

435 **Data Availability**

436 The following information was supplied regarding data availability:

437 Voucher specimens are housed at Department of Life Science, National Taiwan University,

438 Taipei, R. O. C. and Biodiversity Research Center, Academia Sinica, Taiwan, R. O. C. The *COI*

439 sequences described here are accessible via GenBank accession number LC727399 to

440 LC727411.

441

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- 577

578 **Figure 1 Adult specimens of *Symbrenthia lilaea lunica* and *S. l. formosanus*.** (A–B) Male  
579 specimen of *S. l. lunica*, TAIWAN: Miaoli County, Sanmwan, Emei Bridge. (C–D) Female  
580 specimen of *S. l. lunica*, TAIWAN: Xinzhu County, Zhudong, 100m. (E–F) Male specimen of  
581 *S. l. formosanus*, TAIWAN: New Taipei City, Wulai, 150m. (G–H) Female specimen of *S. l.*  
582 *formosanus*, TAIWAN: Taoyuan City, Fuxing, Daman. Scale bar = 1 cm (A–H).

583

584 **Figure 2 Distribution points of *Symbrenthia lilaea formosanus* and *S. lilaea lunica* during**  
585 **two different invasive stages in Taiwan.** (A) The distribution points of *S. l. formosanus* and *S.*  
586 *l. lilaea* in the early invasive stage (1911-2012). (B) The distribution points of *S. l. formosanus*  
587 and *S. l. lilaea* in the late invasive stage (2013-2021).

588

589 **Figure 3 The Jackknife of regularized training gain of different bioclimatic factors of**  
590 ***Symbrenthia formosanus* and *S. lilaea* between different invasive time stages.** (A) represents  
591 the jackknife results of *S. formosanus* in the early invasive stage. (B) represents the jackknife  
592 results of *S. lilaea* (=lunica) in the early invasive stage. (C) represents the jackknife results of *S.*  
593 *formosanus* in the late invasive stage. (D) represents the jackknife results of *S. lilaea* (=lunica) in  
594 the late invasive stage.

595

596 **Figure 4 Systematic and species delimitation results of *Symbrenthia lilaea* and *S.***  
597 ***formosanus*.** (A) Phylogenetic tree reconstructed by BI analysis of 20 *COI* sequences of  
598 *Symbrenthia*. Nodal support from the BI posterior probabilities is along with ML bootstrap  
599 pseudo-replicates (BI posterior probabilities/MLbootstrap). Columns after taxon names  
600 correspond to the results of single-locus species delimitation analyses: Poisson Tree Processes  
601 model (PTP), Automatic Barcode Gap Discovery (ABGD) and Generalized Mixed Yule  
602 Coalescent model (GMYC). (B) PhyloMap visualization the result of PTP species delimitation,  
603 including 20 taxa of *Symbrenthia*. The first axis (horizontal) explained 62.64% and the second  
604 axis (vertical) explained 36.08% of sequence variance among samples.

605

606 **Figure 5 The Species Distribution Model (SDM) of *Symbrenthia formosanus* and *S. lilaea***  
607 **between different invasive time stages predicted from MaxEnt.** (A) represents the SDM of *S.*  
608 *formosanus* in the early invasive stage. (B) represents the SDM of *S. lilaea* in the early invasive  
609 stage. (C) represents the SDM of *S. formosanus* in the late invasive stage. (D) represents the  
610 SDM of *S. lilaea* in the late invasive stage.

611

612 **Figure 6 Shifts in climatic niche between alien (*Symbrenthia lilaea*) and native (*S.***  
613 ***formosanus*) species.** (A) The climatic niche shift between two *Symbrenthia* species in the early

614 invasive stage. (B) The climatic niche shift between two *Symbrenthia* species in the late invasive  
615 stage. Red describes the climatic niche of *S. lilaea* and blue describes the climatic niche of *S.*  
616 *formosanus*.

617

618 **Figure 7 Genitalia of *Symbrenthia lilaea* and *S. formosanus*.** (A–B) Male genitalia of *S.*  
619 *lilaea* (Gen. Prep. JYL-877, NTNU). (C) Female genitalia of *S. lilaea* (Gen. Prep. JYL-880,  
620 NTNU). (D–E) Male genitalia of *S. formosanus* (Gen. Prep. YFH-1583, NTNU). (F) Female  
621 genitalia of *S. formosanus* (Gen. Prep. JYL-1036, NTNU). Scale bar = 1mm (A–B; D–E),  
622 1mm (C; F)

623

624 **Figure 8 Eggs and immature stages of *Symbrenthia lilaea* and *S. formosanus*.** (A) Egg mass  
625 of *S. lilaea*. (B) Young larvae of *S. lilaea*. (C) mature larvae of *S. lilaea*. (D) Egg of *S.*  
626 *formosanus*. (E) Young larvae of *S. formosanus*. (F) mature larvae of *S. formosanus*.

627

628 **Figure S1 The niche overlap results infer from the 19 bioclimatic variables by principal**  
629 **component analyses (PCA).** (A) The niche overlap results of the early invasive stage. (B) The  
630 niche overlap results of the late invasive stage.

631

632 **Figure S2 The receiver operating characteristic curve (ROC) and the threshold-independent**

633 **area under the ROC curve (AUC) values of the SDM. (A) the ROC and AUC of *S.***

634 *formosanus* in the early invasive stage. (B) the ROC and AUC of *S. lilaea* (=lunica) in the early

635 invasive stage. (C) the ROC and AUC of *S. formosanus* in the late invasive stage. (D) the ROC

636 and AUC of *S. lilaea* (=lunica) in the late invasive stage.

637

638 **Table S1 pairwise distance of *Symbrenthia* samples**

639

640 **Table S2 *Symbrenthia* samples information**

# Figure 1

Adult specimens of *Symbrenthia lilaea lunica* and *S. l. formosanus*

(A - B) Male specimen of *S. l. lunica*, TAIWAN: Miaoli County, Sanmwan, Emei Bridge. (C - D) Female specimen of *S. l. lunica*, TAIWAN: Xinzhu County, Zhudong, 100m. (E - F) Male specimen of *S. l. formosanus*, TAIWAN: New Taipei City, Wulai, 150m. (G - H) Female specimen of *S. l. formosanus*, TAIWAN: Taoyuan City, Fuxing, Daman. Scale bar = 1cm (A - H).



I

A



B



C



D



E



F



G

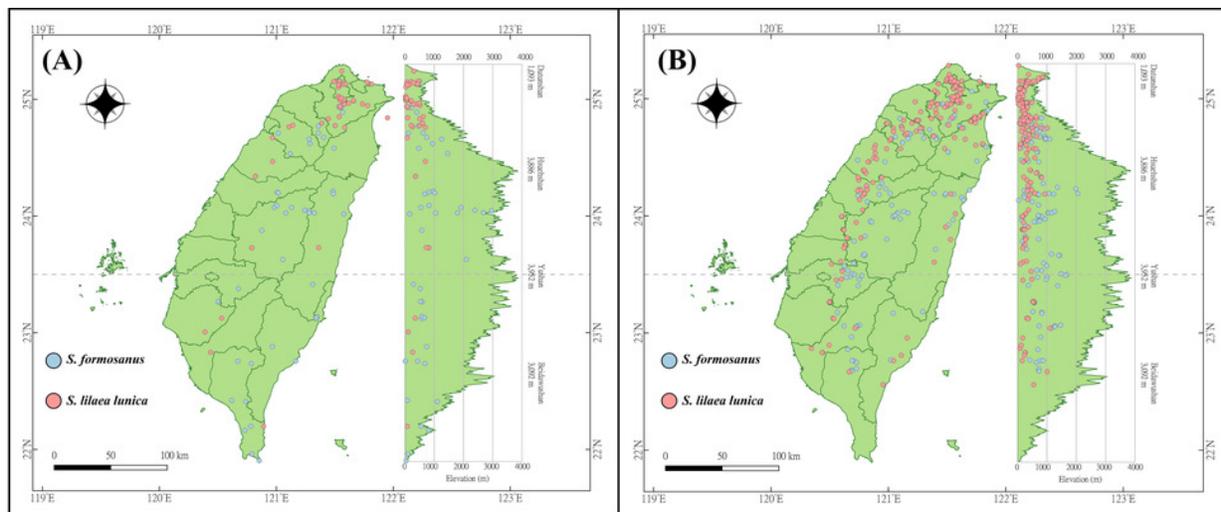


H

## Figure 2

Distribution points of *Symbrenthia lilaea formosanus* and *S. lilaea lunica* during two different invasive stages in Taiwan

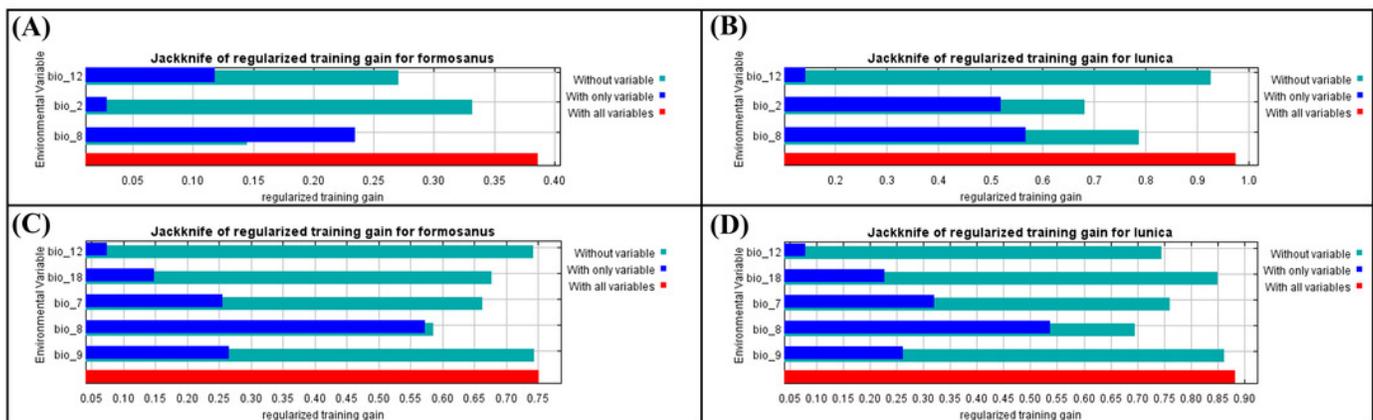
(A) The distribution points of *S. l. formosanus* and *S. l. lilaea* in the early invasive stage (1911-2012). (B) The distribution points of *S. l. formosanus* and *S. l. lilaea* in the late invasive stage (2013-2021).



## Figure 3

The Jackknife of regularized training gain of different bioclimatic factors of *Symbrenthia formosanus* and *S. lilaea* between different invasive time stages

(A) represents the jackknife results of *S. formosanus* in the early invasive stage. (B) represents the jackknife results of *S. lilaea* (=lunica) in the early invasive stage. (C) represents the jackknife results of *S. formosanus* in the late invasive stage. (D) represents the jackknife results of *S. lilaea* (=lunica) in the late invasive stage.

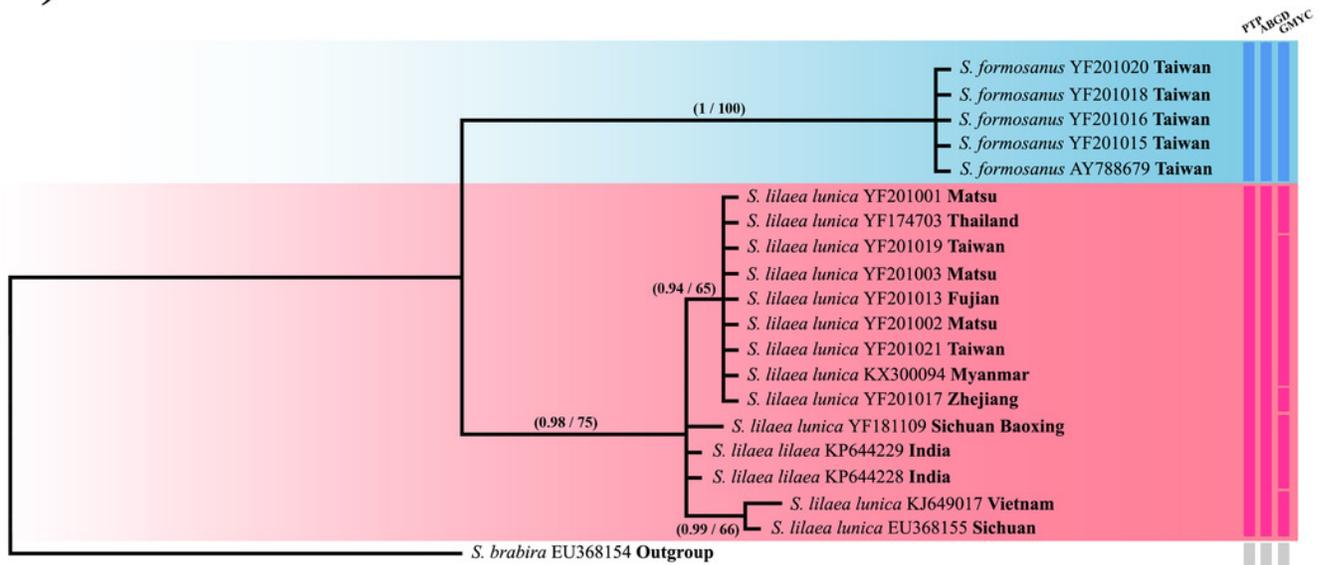


## Figure 4

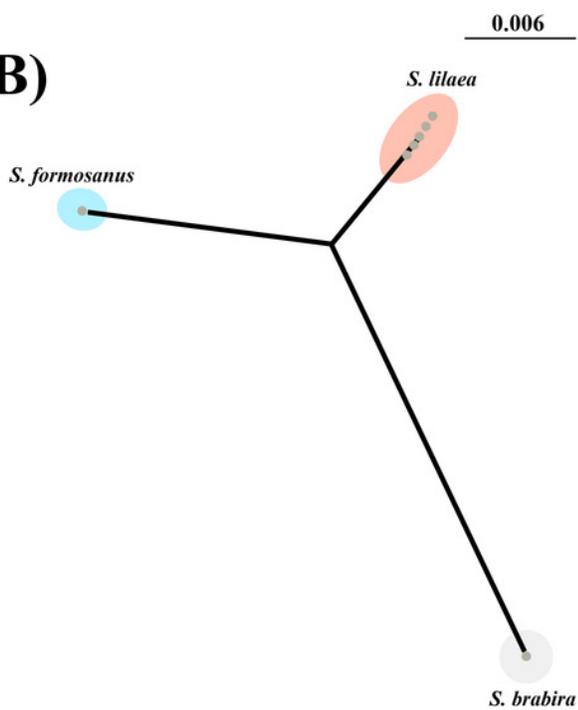
Systematic and species delimitation results of *Symbrenthia lilaea* and *S. formosanus*

(A) Phylogenetic tree reconstructed by BI analysis of 20 *COI* sequences of *Symbrenthia*. Nodal support from the BI posterior probabilities is along with ML bootstrap pseudo-replicates (BI posterior probabilities/MLbootstrap). Columns after taxon names correspond to the results of single-locus species delimitation analyses: Poisson Tree Processes model (PTP), Automatic Barcode Gap Discovery (ABGD) and Generalized Mixed Yule Coalecent model (GMYC). (B) PhyloMap visualization the result of PTP species delimitation, including 20 taxa of *Symbrenthia*. The first axis (horizontal) explained 62.64% and the second axis (vertical) explained 36.08% of sequence variance among samples.

(A)



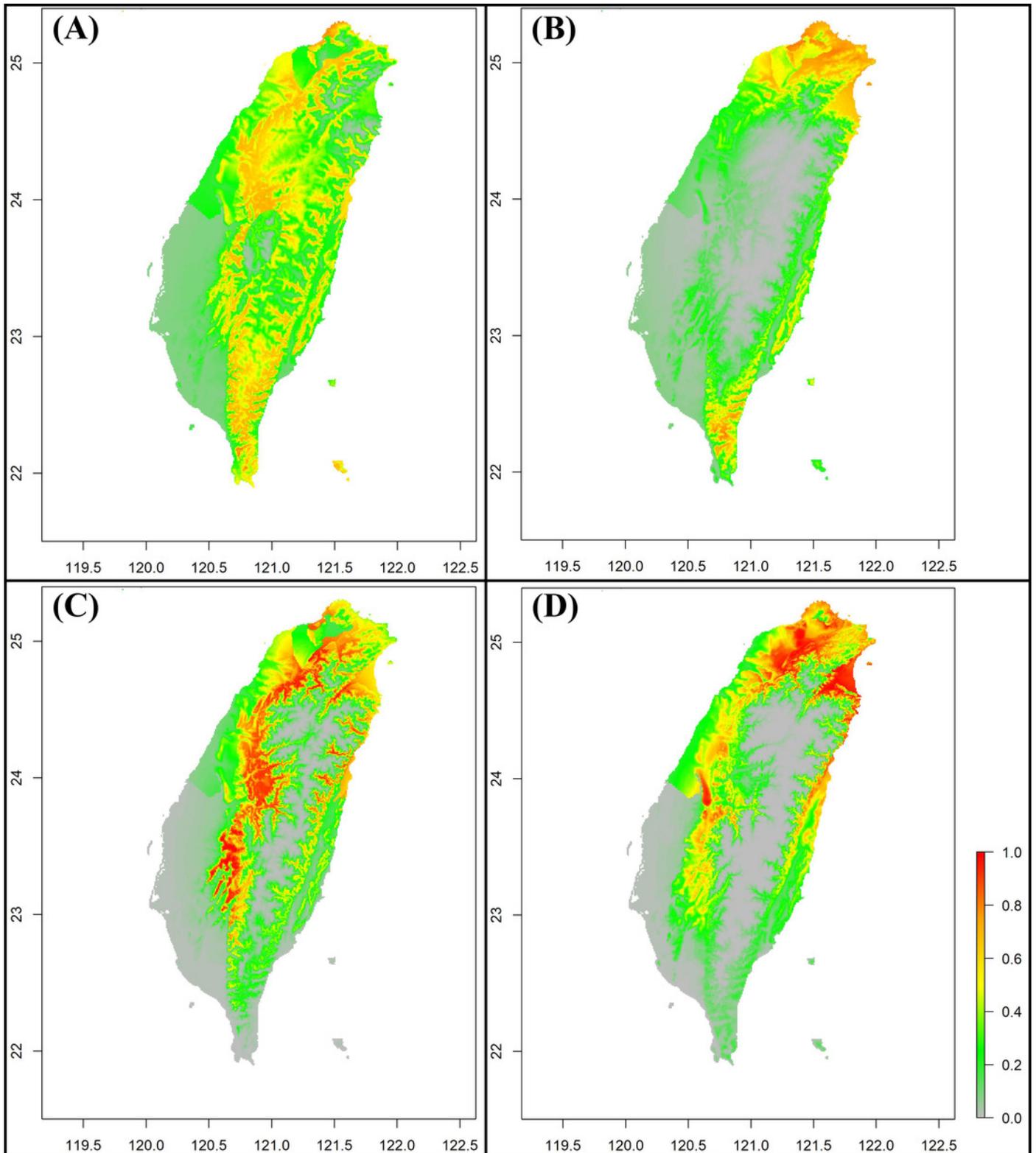
(B)



## Figure 5

The Species Distribution Model (SDM) of *Symbrenthia formosanus* and *S. lilaea* between different invasive time stages predicted from MaxEnt

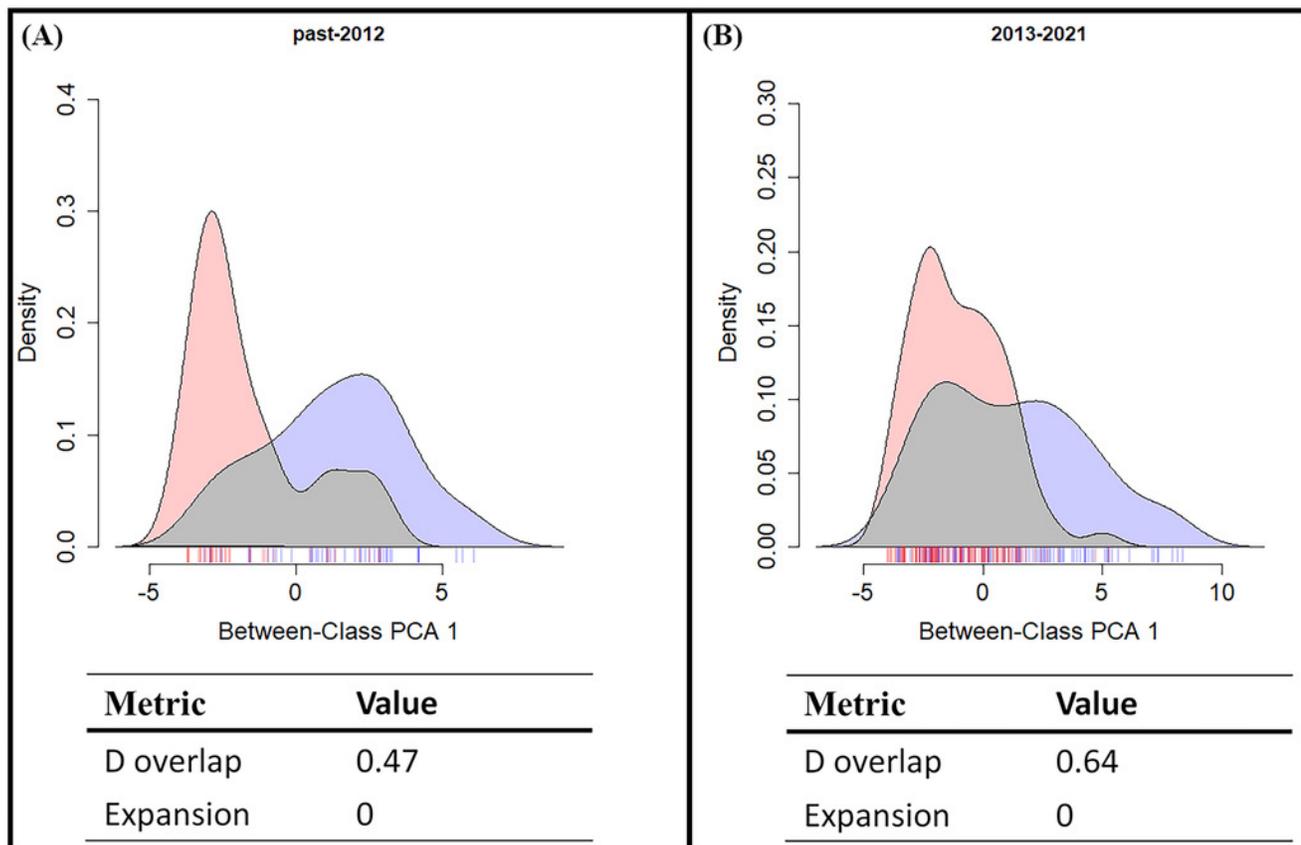
(A) represents the SDM of *S. formosanus* in the early invasive stage. (B) represents the SDM of *S. lilaea* in the early invasive stage. (C) represents the SDM of *S. formosanus* in the late invasive stage. (D) represents the SDM of *S. lilaea* in the late invasive stage.



## Figure 6

Shifts in climatic niche between alien (*Symbrenthia lilaea*) and native (*S. formosanus*) species

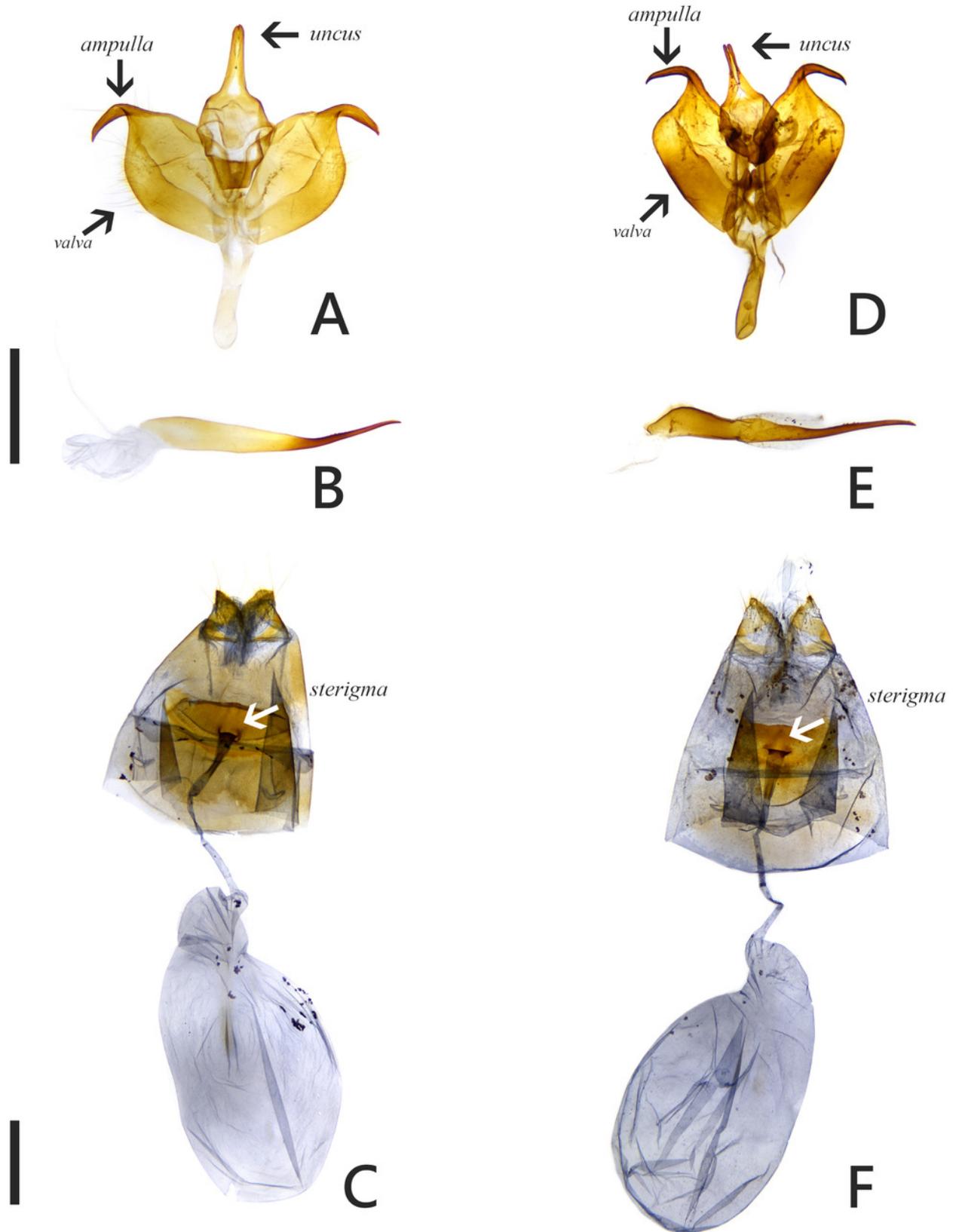
(A) The climatic niche shift between two *Symbrenthia* species in the early invasive stage. (B) The climatic niche shift between two *Symbrenthia* species in the late invasive stage. Red describes the climatic niche of *S. lilaea* and blue describes the climatic niche of *S. formosanus*.



## Figure 7

Genitalia of *Symbrenthia lilaea* and *S. formosanus*

(A - B) Male genitalia of *S. lilaea* (Gen. Prep. JYL-877, NTNU). (C) Female genitalia of *S. lilaea* (Gen. Prep. JYL-880, NTNU). (D - E) Male genitalia of *S. formosanus* (Gen. Prep. YFH-1583, NTNU). (F) Female genitalia of *S. formosanus* (Gen. Prep. JYL-1036, NTNU). Scale bar = 1mm (A - B; D - E), 1mm (C; F).



## Figure 8

Eggs and immature stages of *Symbrenthia lilaea* and *S. formosanus*

(A) Egg mass of *S. lilaea*. (B) Young larvae of *S. lilaea*. (C) mature larvae of *S. lilaea*. (D) Egg of *S. formosanus*. (E) Young larvae of *S. formosanus*. (F) mature larvae of *S. formosanus*.

