

1 Patch and matrix characteristics determine the 2 outcome of ecosystem engineering: a case study on 3 the vegetation of mole rat mounds

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18

19 Abstract

20

21 **Background.** Burrowing mammals are important ecosystem engineers, especially in open
22 ecosystems in the arid and semi-arid regions, where they create patches that differ from the
23 surrounding matrix in their structure or ecosystem functions.

24 **Methods.** We evaluated the fine-scale effects of a subterranean ecosystem engineer, the western
25 blind mole rat mounds on the vegetation composition of sandy dry grasslands in Hungary. In this
26 model system we tested whether the characteristics of the patch (mound size) and the matrix
27 (total vegetation cover in the undisturbed grassland) influence the structural and functional
28 contrasts between the mounds and the undisturbed grasslands. We sampled the vegetation of 80
29 mounds and 80 undisturbed grassland plots, where we recorded the total vegetation cover, and
30  the list and cover of each vascular plant species. We used two proxies to characterise the patches
31 (mounds) and the matrix (undisturbed grassland): the perimeter of the mounds and the total
32 vegetation cover of the undisturbed grasslands. First, we compared the vegetation characteristics
33 of the mounds and the surrounding grasslands with general linear models. Second, we
34 characterised the contrasts between the patches (mounds) and the matrix (undisturbed grassland)
35 by relative response indices (RRIs) of the vegetation characteristics studied in the first step.


36 **Results.** Species composition of the vegetation of the mounds and undisturbed grasslands was
37 well separated in ~~case of~~ three out of the four study sites. Mounds were characterised by **smaller**
38 vegetation cover, **smaller** cover of perennial graminoids, and higher diversity, and evenness
39 compared to undisturbed grasslands. The contrast in vegetation cover between patches and the

40 matrix increased with decreasing patch size. Increasing vegetation cover in the matrix grasslands
41 increased the contrasts between the patches and the matrix in terms of total cover, diversity, and
42 evenness. Our results suggest that mole rat mounds provide improved establishment conditions
43 for subordinate species, because they are larger than other types of natural gaps and are
44 characterised by less intense belowground competition. The ecosystem engineering effect, i.e.,
45 the contrast between the patches and the matrix was the largest in the more closed grasslands.

46

47 Introduction

48

49 Ecosystem engineer organisms create patches that differ from the surrounding matrix in their
50 structure or ecosystem functions (Jones, Lawton & Shachak, 1994). s way they alter the
51 resource distribution in the landscape (Mallen-Cooper, Nakagawa & Eldridge, 2018; Neilly, Cale
52 & Eldridge, 2022; Valkó et al., 2022). Previous syntheses on ecosystem engineers found that
53 there are several factors determining the characteristics of the engineered patches, including the
54 traits of the engineer, the habitat, climate and soil type (Mallen-Cooper, Nakagawa & Eldridge,
55 2018; Root-Bernstein and Ebersperger, 2013). Burrowing mammals are important ecosystem
56 engineers, especially in the arid and semi-arid regions (Davidson, Detling & Brown, 2012, Valkó
57 et al., 2021; Whitford and Kay, 1998). Through their mound-building and burrowing activities,
58 they move large amounts of soil and create sparsely vegetated patches that often have different
59 vegetation compared to the surrounding habitat matrix (Coggan, Hayward & Gibb, 2018;
60 Mallen-Cooper, Nakagawa & Eldridge, 2018).

61

62 Mole rats inhabit arid and semi-arid regions in Africa (mole rats, Bathyergidae family, Visser,
63 Bennett & van Vuuren, 2019) and Eurasia (blind mole rats, Spalacidae family, Németh et al.,
64 2020a). They are highly specialised to subterranean lifestyle and most of the mole rat species
65 spend majority of the time below ground. Therefore, their ecosystem engineering effect is
66 different from ~~these~~ burrowing mammals that feed and graze aboveground, like marmots (Valkó
67 et al., 2021) or prairie dogs (Winter, Cully Jr. & Pontius, 2002). Mole rats can alter vegetation
68 composition by two mechanisms: 1) they generally feed on belowground plant organs, such as
69 roots and bulbs which can decrease the abundance of certain plant species, 2) they create mounds
70 with open soil surface that can play a role in vegetation dynamics as establishment gaps
71 (Reichman and Jarvis, 1998).

72

73 Mole rats are ideal organisms for the study of patches created by burrowing rodents as they
74 create clearly visible mounds with sharp boundaries that are distinct patches in the grassland
75 matrix. Mole rats are considered as ecosystem engineers in fynbos ecosystems, where their
76 mounds are characterised by lower vegetation cover and higher species richness compared to the
77 surrounding matrix (Davies et al., 1986; Hagenah and Bennett, 2012; Reichman and Jarvis,
78 1998). Their engineer effect was tested in temperate dry grasslands in Europe, in a study
79 comparing twelve burrows with undisturbed grasslands (Zimmermann et al., 2014). No

80 difference was found between the vegetation characteristics and species richness of the burrows
81 and the surrounding grasslands, but some differences in the species composition were found.
82 This weak engineering effect detected in European compared to African grasslands might be due
83 to generally more accentuated effects of ecosystem engineers at lower latitudes (Romero et al.,
84 2015). However, given the similar ecological function of African and Eurasian mole rats, we
85 believe that a more detailed analysis on their potential engineer effect in temperate Eurasia is
86 necessary.

87

88 The aim of our study was to evaluate the fine-scale effect of western blind mole rat mounds on
89 the vegetation composition of dry grasslands in Hungary. In this model system we tested whether
90 the characteristics of the patch (mound size) and the matrix (total vegetation cover in the
91 neighbouring undisturbed grassland) influence the structural and functional contrasts between the
92 mounds and the undisturbed grasslands. We tested the following hypotheses: i) Mounds have
93 different vegetation structure, species composition and diversity patterns compared to the
94 undisturbed grasslands. ii) Larger mounds are less affected by the edge effect so the structural
95 and functional contrasts between the mounds and the undisturbed grasslands increases with
96 increasing mound size. iii) The structural and functional contrasts between the mounds and the
97 undisturbed grasslands increases with increasing canopy cover in the grassland matrix. As mole
98 rats create a large number of mounds locally in an ecosystem type where gap dynamics are
99 crucial driver of the vegetation composition, the study system is ideal for testing these
100 hypotheses.

101

102 **Materials & Methods**

103


104 *Study system*

105

106 The western blind mole rat (*Nannospalax* [superspecies *leucodon*]) superspecies complex
107 includes several morphologically very similar but genetically isolated species (Németh et al.,
108 2020a). Besides genetic differences within the superspecies complex, involved taxa have the
109 same ecological function and lifestyle, hence here we did not distinguish between them. These
110 subterranean mammals are strictly protected and critically endangered in Central Europe. They
111 inhabit dry grasslands, old-fields, and sometimes also urban areas (Németh, Moldován & Szél,
112 2020b). We sampled four study sites that all hold large populations of the mole rats. The study
113 sites represent the largest known localities of the species in Hungary and are characterised by
114 pristine sandy grassland vegetation in a good conservation status (Németh Moldován & Szél,
115 2020b). The study sites are located in the Hortobágy National Park (Bagamér site – N 47.47028,
116 E 21.95873, and Hajdúbagos site – N 47.41340, E 21.67606) and in the Kiskunság National Park
117 (Ásotthalom site – N 46.22235, E 19.67164, and Baja site –N 46.19643, E 18.99083). **he** Trans-
118 Tisza Environmental, Nature Protection and Water Inspectorate approved this study
119 (6646/08/2014). The characteristic vegetation of the study sites is dry sandy grassland, the

120 dominant grass species are *Festuca pseudovina*, *F. rupicola*, *F. vaginata* and *Koeleria glauca*.
121 The sites provide habitat to several protected grassland plant species, such as *Astragalus va,rius*,
122 *Colchicum arenarium*, and the strictly protected *Pulsatilla flavescens* (Borhidi et al., 2012).

123

124 *Sampling design* 

125

126 In each study site in April 2020, we selected 20 mounds built by mole rats. We selected mounds
127 that were built one year before the survey. Instead of using a fix-sized sampling quadrat, we
128 considered one mound as one sampling unit as in this case we could capture the potential within-
129 mound variety in the vegetation. We adjusted a measuring tape along the mound edge and used
130 the same measuring tape for delineating a control plot with the same size and shape as the
131 mound. Control plots were designated in the undisturbed sandy grassland one meter from each
132 mound. We recorded the perimeter of each mound.

133

134 We sampled the vegetation of 80 mounds and 80 undisturbed grassland plots: we recorded the
135 total vegetation cover, and the list and cover of each vascular plant species. Plant nomenclature
136 follows the work of Király (2009).

137

138 *Data analysis*

139

140 We used two proxies to characterise the patches (mounds) and the matrix (undisturbed
141 grassland): the perimeter of the mounds and the total vegetation cover of the undisturbed
142 grasslands. We calculated the Shannon diversity and the evenness of the vegetation in each
143 sampling unit.

144

145 First, we compared the vegetation characteristics of the mounds and the surrounding grasslands
146 with general linear models, where the **fix** factor was microsite type (mound vs. undisturbed
147 grassland), and site **was random** factor. In the analysis, the following dependent variables were
148 used: total vegetation cover, cover of perennial graminoids, species richness, Shannon diversity,
149 and evenness.

150

151 Second, we characterised the contrasts between the patches (mounds) and the matrix
152 (undisturbed grassland) by relative response indices (RRIs, Armas, Ordiales & Pugnaire, 2004;
153 Perkins and Hatfield, 2014) of the vegetation characteristics studied in the first step. RRIs were
154 calculated based on the following equation: $RRI = (X_M - X_G) / (X_M + X_G)$; where X_M is the
155 value of a dependent variable (e.g. Shannon diversity) in a mound and X_G is the value of the
156 same dependent variable in the undisturbed grassland plot paired with the mound plot. Value of
157 RRI ranges between -1 and $+1$. The closer is $|RRI|$ to 1, the higher the contrast between the
158 patches and the matrix. With generalized linear mixed models (GLMMs), we tested the effects of
159 patch size (fixed factor) and total vegetation cover in matrix grassland (fixed factor) on the

160 contrasts between the patches and the matrix (i.e., |RRI|s of the studied vegetation
161 characteristics). Study site was used as random factor. GLMs and GLMMs were calculated in the
162 SPSS 17.0 program.

163 We applied non-metric multidimensional scaling (NMDS) using Bray-Curtis index of
164 dissimilarity to test differences in the species composition of the mounds and undisturbed
165 grasslands in CANOCO 5.0 program (ter Braak and Šmilauer, 2012). We made the calculations
166 for each site separately.

167

168 Results

169

170 We recorded in total 112 vascular plant species, out of which 102 species occurred on mounds
171 and 93 in undisturbed grasslands. 19 species occurred exclusively on mounds and 10 exclusively
172 in the matrix. 64 species were more frequent on mounds than in the undisturbed grasslands, 18
173 occurred with the same frequency, and 30 species were more frequent in the undisturbed
174 grasslands. Out of the six protected plant species recorded at the study sites, one occurred only
175 on mounds (*Pulsatilla flavescens*), four occurred both on mounds and grasslands (*Colchicum*
176 *arenarium*, *Dianthus serotinus*, *Onosma arenaria*, *Stipa borystenica*) and one only in
177 undisturbed grasslands (*Astragalus varius*).

178

179 Species composition of the vegetation of the mounds and undisturbed grasslands was well
180 separated in ~~case of~~ three study sites (Hajdúbagos, Bagamér and Ásotthalom), whilst showed
181 considerable similarity in the case of the Baja site (Figure 1). The matrix grass species (*Festuca*
182 spp.) and some typical perennial (such as *Thymus glabrescens*, *Potentilla arenaria*, *Euphorbia*
183 *cyparissias*) species of the studied dry grasslands characterised the undisturbed grasslands.
184 Vegetation of mounds were characterised by several disturbance-tolerant species (such as
185 *Erophila verna*, *Eryngium campestre*, *Poa bulbosa*, *Rumex acetosella*, *Vicia lathyroides*).

186

187 Mounds were characterised by smaller vegetation cover ($F = 87.168, p = 0.003$), smaller cover of
188 perennial graminoids ($F = 93.503, p = 0.002$), higher Shannon diversity ($F = 16.422, p = 0.027$)
189 and evenness ($F = 15.780, p = 0.029$) compared to undisturbed grasslands (Figure 2). Species
190 richness on the mounds and in the undisturbed grasslands was not different ($F = 6.820,$
191 $p = 0.080$).

192

193 The average perimeter of the mounds was $2.18 \text{ m} \pm 0.88 \text{ SD}$, and the total vegetation cover in the
194 undisturbed grasslands was $66.94\% \pm 14.17 \text{ SD}$. There was a high contrast in total vegetation
195 cover between small patches and the matrix, which decreased with increasing patch size
196 (Table 1). Increasing vegetation cover in the matrix grasslands increased the contrasts between
197 the patches and the matrix in terms of total cover, Shannon diversity and evenness.

198

199

200 Discussion

201

202 We found that western blind mole rats created patches with different structural attributes (smaller
203 vegetation cover, smaller cover of perennial grasses) compared to the surrounding sand grassland
204 matrix. The vegetation of the patches and the matrix were different in terms of total vegetation
205 cover, cover of perennial grasses as well as Shannon diversity and evenness, which confirmed
206 our first hypothesis. This suggests that the mole rats play an important role in the gap dynamics
207 in the study system. We found a more open vegetation with smaller cover of perennial grasses on
208 the mounds, which is a general pattern observed on mounds of various burrowing mammals
209 (e.g., prairie dogs: Winter, Cully Jr. & Pontius, 2002; pikas: Wesche, Nadrowski & Retzer, 2007;
210 marmots: Valkó et al., 2021). These are notable effects considering the small size of the mounds.

211

212 In the studied sandy grassland ecosystems, the vegetation is open, but there is an intense
213 belowground competition between the perennial graminoids and other subordinate plant species
214 (Borhidi, Kevey & Lendvai, 2012). It is possible that the gaps created by mole rats can provide
215 improved establishment conditions for subordinate species, because the mounds are larger than
216 the other types of natural gaps. Also, as mole rats feed on roots and other belowground organs
217 (Corbet 1984), they can locally decrease the belowground competition which gives an
218 establishment advantage to subordinate species over perennial graminoids. The differences
219 between the vegetation of the mole rat mounds and the matrix is interesting also because of the
220 subterranean lifestyle of the mole rats. In most burrowing mammals, mounds are not only
221 biogeomorphological features but the mammals' activities on the mound surface also shape
222 vegetation composition, e.g., by trampling and manuring (e.g., foxes, Godó et al., 2018;
223 marmots, Valkó et al., 2021; pikas, Yoshihara et al., 2010), but this is not the case in mole rats
224 that spend most of their life underground.

225

226 Compared to the only other published study on the ecosystem engineer effect of mole rat species
227 in temperate Eurasia (Zimmermann et al., 2014), we found more and stronger evidence for the
228 engineer effects. The detected weaker evidence of the engineer effect in the previous study can
229 be either a result of the particular study design (one study site, small sample size, fixed plot size)
230 in the other study or by the slightly different habitat types considered (sandy grasslands in our
231 study and loess steppes in Zimmermann et al., 2014).

232

233 The species composition of the mounds and undisturbed grasslands was not different when
234 considering all the sites together; however, looking at the site level we found marked differences
235 in three of the four study sites. This suggests that the effect of mole rats on the vegetation should
236 be considered at the local and not the regional scale. This is in line with another study on the
237 effect of fine-scale environmental heterogeneity on the species composition of grasslands, where
238 the effects of environmental heterogeneity were more pronounced on the local than on the
239 regional scale (Deák et al., 2021).

240

241 Most of the species recorded in the study (74%) occurred both on the mounds and in the intact
242 grasslands. This implies that mounds are not unique establishment microsites for the majority of
243 plant species; however, mounds can be considered as temporal biodiversity hotspots, where due
244 to the low level of competition by perennial graminoids, subordinate species can establish. There
245 was no difference between the species richness of the mounds and the undisturbed grasslands,
246 but vegetation patches on the mounds were more diverse and the species were more evenly
247 distributed compared to the undisturbed grassland matrix.

248

249 Our second hypothesis was not supported as we found that smaller mounds were structurally
250 more different from the surrounding matrix than larger ones in terms of total vegetation cover.
251 This suggests that the height of the mounds (approximately 10 cm) and their steep slopes provide
252 a sharp vegetation boundary which prevents the clonal growth of the surrounding vegetation on
253 the mound. The colonization of the mounds by plants is probably driven by random dispersal
254 processes (i.e., seed rain), and a higher number of incoming diaspores can be expected on the
255 larger surface of larger patches. Also, besides total vegetation cover, we found that mound size
256 did not affect the other studied variables. Even there was variation in mound size they were
257 rather small so the effect of patch size might be relevant in other scales.

258

259 We confirmed our third hypothesis as the contrasts between the matrix and the patch were higher
260 in the case of more closed matrix vegetation. This suggests that the importance of the ecosystem
261 engineering effect is the highest in the more closed grasslands, where the engineer organisms
262 increase more the structural and functional heterogeneity. For the conservation of the plant
263 species associated to dry grasslands, creating proper establishment microsites is crucial (Klaus et
264 al., 2018). Our results suggest that mounds of mole rats can provide suitable establishment
265 microsites for approximately 91% of the species pool of the studied sandy dry grasslands. Thus,
266 they might be potentially feasible as establishment gaps in restoration projects or when
267 introducing particular rare species (Kiss et al., 2021; Limb et al., 2010). The spatio-temporal
268 dynamics of the creation and the re-vegetation of the mounds can be an important driver of
269 establishment of subordinate species in the studied grassland ecosystems. Further studies are
270 needed for testing the effectiveness of such natural gaps in increasing species richness during
271 restoration.

272

273

274 **Conclusions**

275

276 We found that the subterranean mole rats create patches in temperate sandy grasslands that differ
277 from the matrix in species composition and vegetation characteristics. The contrast between the
278 patches and the matrix were the sharpest in grasslands with more closed canopy cover. We
279 suggest that the contrasts between the patches and the matrix which was proposed in this study as

280 a proxy for the strength of the engineering effect, can be a useful variable also in other studies.
281 The effects of patch and matrix characteristics on the contrasts should be studied in other
282 ecosystems and other organisms for a deeper understanding of the mechanisms beyond
283 ecosystem engineering.

284

285

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287

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289 topic.

290

291 **References**

292

293 Armas C, Ordiales R, Pugnaire FI. 2004. Measuring plant interactions: a new comparative index.
294 *Ecology* 85:2682–2686. <https://doi.org/10.1890/03-0650>

295 Borhidi A, Kevey B, Lendvai G. 2012. *Plant Communities of Hungary*. Budapest: Akadémiai
296 Kiadó.

297 Coggan NV, Hayward MW, Gibb H. 2018. A global database and “state of the field” review of
298 research into ecosystem engineering by land animals. *Journal of Animal Ecology* 87:974–
299 994. <https://doi.org/10.1111/1365-2656.12819>

300 Corbet G. 1984. *The Encyclopedia of Mammals*. New York: Facts on File Publications.

301 Davidson AD, Detling JK, Brown JH. 2012. Ecological roles and conservation challenges of
302 social, burrowing, herbivorous mammals in the world's grasslands. *Frontiers in Ecology
303 and the Environment* 10(9):477–486. <https://doi.org/10.1890/110054>

304 Davies KC, Jarvis JU. 1986. The burrow systems and burrowing dynamics of the mole-rats
305 *Bathorygus suillus* and *Cryptomys hottentotus* in the fynbos of the south-western Cape,
306 South Africa. *Journal of Zoology* 209(1):125–147. [https://doi.org/10.1111/j.1469-
307 7998.1986.tb03570.x](https://doi.org/10.1111/j.1469-7998.1986.tb03570.x)

308 Deák B, Kovács B, Rádai Z, Apostolova I, Kelemen A, Kiss R, Lukács K, Palpurina S,
309 Sopotlieva D, Báthori F, Valkó O. 2021. Linking environmental heterogeneity and plant
310 diversity: the ecological role of small natural features in homogeneous landscapes. *Science
311 of the Total Environment* 763:144199. <https://doi.org/10.1016/j.scitotenv.2020.144199>.

312 Godó L, Tóthmérész B, Valkó O, Tóth K, Radócz S, Kiss R, Kelemen A, Török P, Švamberková
313 E, Deák B. 2018. Ecosystem engineering by foxes is mediated by isolation in grassland
314 fragments. *Ecology and Evolution* 8(14):7044–7054. <https://doi.org/10.1002/ece3.4224>

315 Hagenah N, Bennett NC. 2013. Mole rats act as ecosystem engineers within a biodiversity
316 hotspot, the Cape Fynbos. *Journal of Zoology* 289(1):19–26.
317 <https://doi.org/10.1111/j.1469-7998.2012.00958.x>

318 Jones CG, Lawton JH, Shachak M. 1994. Organisms as ecosystem engineers. In: *Ecosystem
319 Management*. New York, NY: Springer.

- 320 Király G. 2009. New Hungarian Herbal. The Vascular Plants of Hungary. Identification Keys.
321 Aggtelek National Park Directorate, Jósvalő (in Hungarian).
- 322 Kiss R, Deák B, Tóthmérész B, Miglécz T, Tóth K, Török P, Lukács K, Godó L, Körmöczi Z,
323 Radócz S, Kelemen A, Sonkoly J, Kirmer A, Tischew S, Švamberková E, Valkó O. 2021.
324 Establishment gaps: biodiversity hotspots to support the colonization of target species in
325 species-poor grasslands. *Restoration Ecology* 29(S1):e13135.
326 <https://doi.org/10.1111/rec.13135>
- 327 Klaus VH, Hoever CJ, Fischer M, Hamer U, Kleinebecker T, Mertens D, Schäfer D, Prati D,
328 Hölzel N. 2018. Contribution of the soil seed bank to the restoration of temperate
329 grasslands by mechanical sward disturbance. *Restoration Ecology* 26:S114–S122.
330 <https://doi.org/10.1111/rec.12626>
- 331 Limb RF, Engle DM, Bidwell TG, Althoff DP, Anderson AB, Gipson PS, Howard HR. 2010.
332 Restoring biopedturbation in grasslands with anthropogenic focal disturbance. *Plant*
333 *Ecology* 210(2):331–342. <https://doi.org/10.1007/s11258-010-9760-7>
- 334 Mallen-Cooper M, Nakagawa S, Eldridge DJ. 2019. Global meta-analysis of soil-disturbing
335 vertebrates reveals strong effects on ecosystem patterns and processes. *Global Ecology and*
336 *Biogeography* 28(5):661–679.
- 337 Neilly H, Cale P, Eldridge DJ. 2022. Incubation mound building by the Australian megapode
338 (malleefowl, *Leipoa ocellata*) creates novel, resource-rich patches in a semi-arid woodland.
339 *Journal of Ecology* 110(6):1432–1441. <https://doi.org/10.1111/1365-2745.13883>
- 340 Németh A, Csorba G, Laczkó L, Mizsei E, Bereczki J, Pásztor JA, Petró P, Sramkó G. 2020a.
341 Multi-locus genetic identification of a newly discovered population reveals a deep genetic
342 divergence in European blind mole rats (Rodentia: Spalacidae: Nannospalax). *Annales*
343 *Zoologici Fennici* 57(1):89–98. <https://doi.org/10.5735/086.057.0110>
- 344 Németh A, Moldován O, Szél L. 2020b. Do they always stand in the way? – Challenges of Blind
345 Mole Rat Protection in Urban Environments in Hungary. *Természetvédelmi Közlemények*
346 26:52–69.
- 347 Perkins LB, Hatfield G. 2014. Competition, legacy, and priority and the success of three invasive
348 species. *Biological Invasions* 16:2543–2550. <https://doi.org/10.1007/s10530-014-0684-3>
- 349 Reichman OJ, Jarvis JUM. 1989. The influence of three sympatric species of fossorial mole-rats
350 (Bathyergidae) on vegetation. *Journal of Mammalogy* 70(4):763–771.
351 <https://doi.org/10.2307/1381710>
- 352 Root-Bernstein M, Ebensperger LA. 2013. Meta-analysis of the effects of small mammal
353 disturbances on species diversity, richness and plant biomass. *Austral Ecology* 38, 289–
354 299. <https://doi.org/10.1111/j.1442-9993.2012.02403.x>
- 355 ter Braak CJF, Šmilauer P. 2012. *CANOCO reference manual and user's guide: Software for*
356 *ordination, Wageningen version 5.0*. Ithaka: Microcomputer Power.
- 357 Valkó O, Borza S, Godó L, Végvári Z, Deák B. 2022. The Eurasian crane (*Grus grus*) as
358 ecosystem engineer in grasslands – conservation values, ecosystem services and

- 359 disservices related to a large iconic bird species. *Land Degradation and Development*
360 33(12):2155–2165. <https://doi.org/10.1002/ldr.4314>
- 361 Valkó O, Tölgyesi C, Kelemen A, Bátori Z, Gallé R, Rádai Z, Bragina TM, Bragin YA, Deák B.
362 2021. Steppe Marmot (*Marmota bobak*) as ecosystem engineer in arid steppes. *Journal of*
363 *Arid Environments* 184:104244. <https://doi.org/10.1016/j.jaridenv.2020.104244>
- 364 Visser JH, Bennett NC, van Vuuren BJ. 2019. Phylogeny and biogeography of the African
365 Bathyergidae: a review of patterns and processes. *PeerJ* 7:e7730.
366 <https://doi.org/10.7717/peerj.7730>
- 367 Wesche K, Nadrowski K, Retzer V. 2007. Habitat engineering under dry conditions: the impact
368 of pikas (*Ochotona pallasii*) on vegetation and site conditions in southern Mongolian
369 steppes. *Journal of Vegetation Science* 18 (5):665–674.
- 370 Whitford WG, Kay FR. 1999. Biopedturbation by mammals in deserts: a review. *Journal of Arid*
371 *Environments* 41(2):203–230. <https://doi.org/10.1006/jare.1998.0482>
- 372 Winter SL, Cully Jr. JF, Pontius JS. 2002. Vegetation of prairie dog colonies and non-colonized
373 shortgrass prairie. *Journal of Rangeland Management* 55:502–508.
- 374 Yoshihara Y, Okuro T, Buuveibaatar B, Undarmaa J, Takeuchi K. 2010. Clustered animal
375 burrows yield higher spatial heterogeneity. *Plant Ecology* 206 (2):211–224.
376 <https://doi.org/10.1007/s11258-009-9636-x>
- 377 Zimmermann Z, Szabó G, Csathó AI, Sallainé Kapocsi J, Szentes S, Juhász M, Házi J, Komoly
378 C, Virágh K, Harkányiné Szekeres Z, Lampert R, Sutyinszki Z, Bartha S. 2014. The impact
379 of the lesser blind mole rat [*Nannospalax* (superspecies *leucodon*)] on the species
380 composition and diversity of a loess steppe in Hungary. *Applied Ecology and Environmental*
381 *Research* 12(2):577–588.
382

Table 1 (on next page)

Table 1. The effect of patch size and matrix cover (fixed factors) on the contrasts between the vegetation of the mounds and the grassland matrix.

Contrasts were expressed by the absolute values of the relative response indices ($|RRI|$) calculated between the vegetation characteristics of the mounds and the undisturbed grasslands.

1 **Table 1.** The effect of patch size and matrix cover (fixed factors) on the contrasts between the
 2 vegetation of the mounds and the grassland matrix. Contrasts were expressed by the absolute
 3 values of the relative response indices (|RRI|) calculated between the vegetation characteristics of
 4 the mounds and the undisturbed grasslands.

	Patch size			Matrix cover		
	direction	F	p	direction	F	p
Total cover, RRI	↓	6.355	0.014	↑	31.240	0.000
Species richness, RRI		0.715	0.401		0.294	0.589
Perennial grass cover, RRI		0.360	0.550	↑	9.527	0.003
Shannon diversity, RRI		0.219	0.641	↑	4.135	0.045
evenness, RRI		0.012	0.912	↑	9.303	0.003

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Figure 1

Figure 1. Differences in the species composition of mounds and undisturbed grasslands in the four studied sites (NMDS ordination).

We plotted the 20 most abundant species at each site on the panels. Species names are abbreviated using the first three letters of the genus and species names.

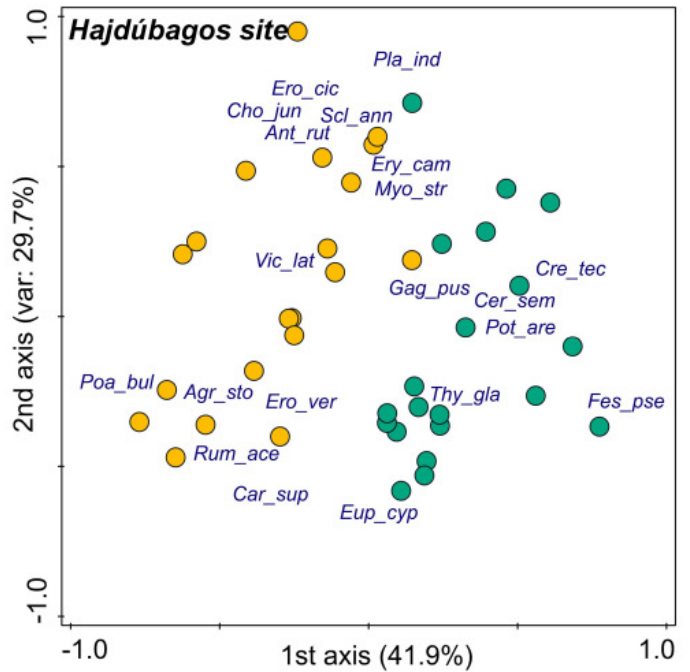
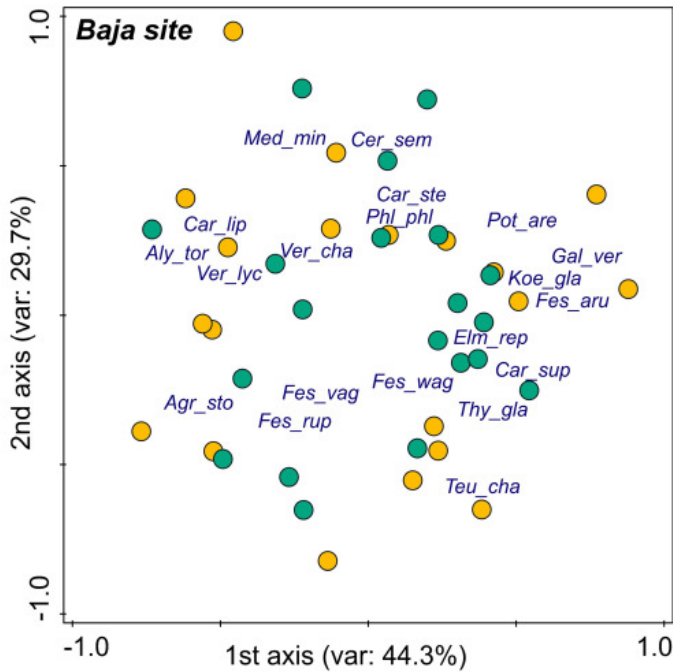
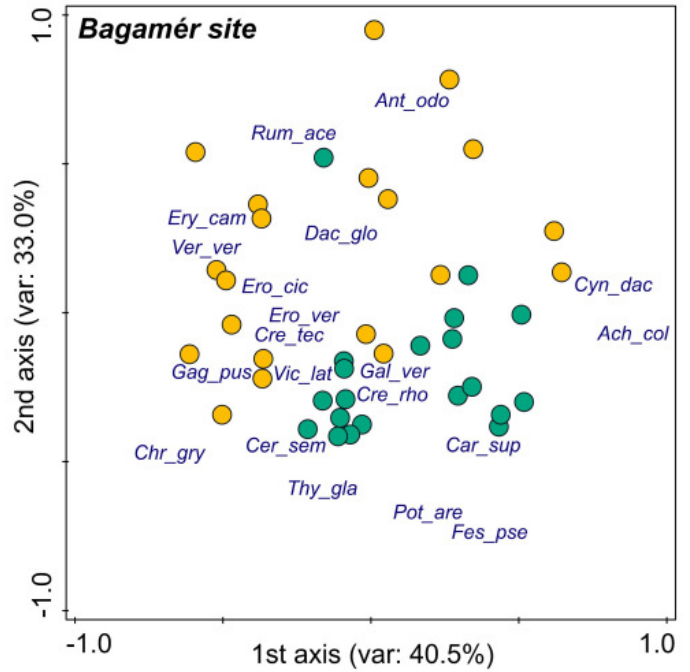
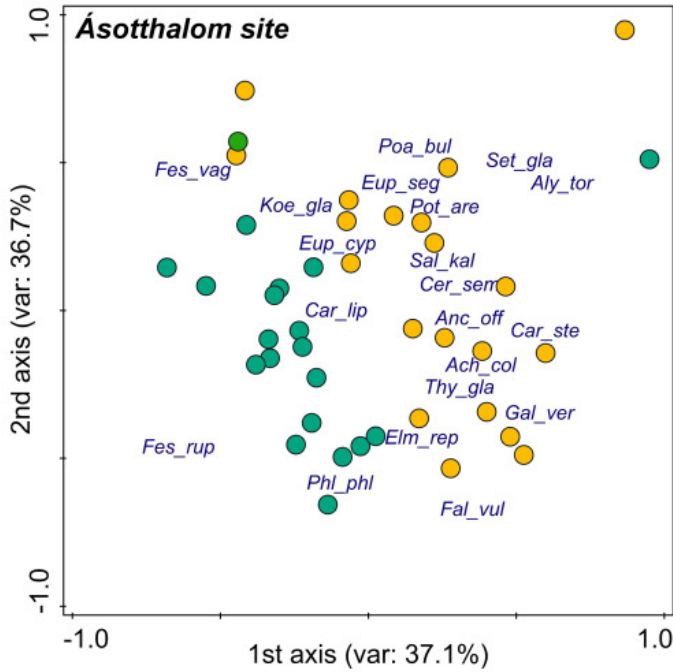


Figure 2

Figure 2. Vegetation characteristics on the patches (mole rat mounds) and matrix (undisturbed grasslands).

