Sex ratio and relatedness in the Griffon vulture (*Gyps fulvus*) population of Serbia

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Background

Once a widespread species across the region of Southeast Europe, the Griffon vulture is now confined to small and isolated populations across the Balkan Peninsula. The population from Serbia represents its biggest and most viable population that can serve as an important reservoir of genetic diversity from which the birds can be used for the region's reintroduction programmes. The available genetic data for this valuable population are scarce and as a protected species that belongs to the highly endangered vulture group, it needs to be well described so that it can be properly managed and used as a restocking population. Considering the serious recent bottleneck event that the Griffon vulture population from Serbia experienced we estimated the overall relatedness among the birds from this population. Sex ratio, another important parameter that shows the vitality and strength of the population was evaluated as well.

Methods

During the annual monitoring that was performed in the period from 2013-2021, we collected blood samples from individual birds that were marked in the nests. In total, 169 samples were collected and each was used for molecular sexing while 58 presumably unrelated birds from different nests were used for inbreeding and relatedness analyses. The relatedness was estimated using both biparentally (10 microsatellite loci) and uniparentally (*Cytb* and *D-loop I* of mitochondrial DNA) inherited markers.

Results

The level of inbreeding was relatively high and on average it was 8,3% while the mean number of relatives for each bird was close to three. The sex ratio was close to 1:1 and for the analysed period of nine years, it didn't demonstrate a statistically significant deviation from the expected ratio of 1:1, suggesting that this is a stable and healthy population. Our data suggest that, even though a relatively high level of inbreeding can be detected among the individual birds, the Griffon vulture population from Serbia can be used as a source population for restocking and reintroduction programmes in the region. These data combined with previously observed genetic differentiation between the populations from the

Iberian and Balkan Peninsulas suggest that the introduction of foreign birds should be avoided and that local birds should be used instead.

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27 region's reintroduction programmes. The available genetic data for this valuable
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29 vulture group, it needs to be well described so that it can be properly managed and
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49 be used as a source population for restocking and reintroduction programmes in the
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52 foreign birds should be avoided and that local birds should be used instead.

53

54 ## Introduction
55 The Eurasian Griffon vulture (*Gyps fulvus*) is a colonial, cliff-nesting obligate scavenger
56 specialized in feeding on softer parts of mammal carcasses (Johnson et al. 2006). As a
57 member of a vulture group of birds, it provides an essential ecosystem service of
58 carcass removal in nature which prevents the spread of infectious diseases (Moleon
59 2014). They are monogamous, long-living birds with delayed maturity and a maximum
60 of one fledgling per pair per year (Del Hoyo et al. 1994; Fergusson-Lees & Christie
61 2001; García-Ripollés & López-López 2011; Pirastru et al. 2021). Although once
62 common species in Europe, these vultures faced a serious population decline in the
63 mid-20th century leading to the extinction of many populations across the European
64 range and the vast depletion of breeding birds across the breeding range (Martens et al.
65 2005; Pantovic & Andevski 2018; Pirastru et al. 2021; Schenk 1972; Tucker & Heath
66 1994). Thanks to the successful cross-continental conservation efforts that included
67 reintroduction and conservation measures implemented across the former natural range
68 of the Griffon vulture in Europe, the future of this ecologically important species seems
69 brighter (Dobrev et al. 2021; Le Gouar et al. 2008; Pirastru et al. 2021; Potena et al.
70 2009). The European populations are now showing a significant increment in size with
71 this species now being categorized as 'Least Concern' on the IUCN Red List (BirdLife
72 International. 2021). However, even with these successful measures, this species is still
73 facing serious population declines in some habitats and has become extinct in others,
74 which is why it is a subject of conservation strategies by numerous laws, directives, and
75 conventions (reviewed in Pirastru 2021). Due to the bottlenecks and inbreeding
76 depressions that most of these populations experienced in recent history, constant
77 monitoring efforts are needed in order to assess the population's ability to maintain
78 healthy status and possibly be future reintroduction resources.

79   Due to conservation efforts that included the opening of the feeding station and
80   prohibition of poison usage, the Serbian Griffon vulture population is one of the
81   successful conservation stories. From only ten breeding pairs recorded in the 1990s, it
82   has reached 164 breeding pairs in 2019 and now is the largest population in the Balkan
83   Peninsula (Dobrev et al. 2021; Marinkovic et al. 2021). The historical bottleneck, seen in
84   the last decade of the 20th century, was a peak of a rapid demographic decline in
85   Serbian Griffon vulture populations. This decline resulted from mass poisonings and
86   loss of food sources due to the change in livestock management (Davidovic et al. 2020).
87   A study by Davidovic et al. (2020) showed that this population's genetic variability
88   (based on the variability of ten microsatellites loci) was similar to numerous, unscathed
89   populations such as the Iberian Griffon vulture population and therefore is proposed to
90   serve as a stock population for further reintroduction efforts in the Balkan Peninsula. In
91   addition, analysis of the mitochondrial DNA (mtDNA) gene, *Cytb*, further demonstrated
92   that the Griffon vulture population of Serbia exhibits substantial levels of genetic
93   diversity as well as the presence of private haplotypes which stresses the need of using
94   the local birds in the future reintroduction programmes planned for the region (Davidovic
95   et al. 2022). However, determining the status of the endangered population under
96   conservation measures should be based on more than just one or two parameters of
97   genetic variability. Parameters such as breeding parameters, genetic analysis of
98   mitochondrial genome, adult and offspring sex ratio as well as relatedness among
99   presumed unrelated birds are of great importance for establishing the status of both
100   biological and genetic health of the population in question (Arshad et al. 2009a; Arshad
101   et al. 2009b; Earnhardt 1999; Friar et al. 2001; Godoy et al. 2004).
102   Breeding parameters in the Serbian Griffon vulture population are relatively high
103   despite the substantial increase in population size, implying that the population is far
104   from its saturation level (Marinkovic et al. 2021) further strengthening this population
105   prospect as a reintroduction resource in the Balkan Peninsula (Davidovic et al. 2020;
106   Davidovic et al. 2022). The latest research on the spatial ecology of Griffon vultures in
107   the Balkans identified seven key vulture zones (one in Serbia, one shared between
108   North Macedonia and Bulgaria, one shared between Bulgaria and Greece, two in
109   Bulgaria, one in western Greece and one shared between Kvarner Archipelago islands
110   in Croatia and the Julian Alps - Italy, Austria and Slovenia) (Peshev et al. 2021). This
111   research also identified that the young birds tend to migrate but always return to their
112   nesting ground after migration due to the strong natal philopatry specific for this species
113   and so far no or little exchange of migrant birds between these vulture zones in the
114   Balkans was detected (Peshev et al. 2021).
115   Sex ratio, the proportion of females and males, is an important factor contributing
116   to the success of the newly established population (Sarrazin & Legendre 2000). Fisher's
117   (1930) argument that equal investment in the sex of offspring represents an
118   evolutionary stable strategy and Trivers' (1972) additional comments in theory on sexual

119 selection states that at birth in monogamous species expected sex ratio should be 1:1
120 and that the equal ratio may be a useful indicator of population trajectory and overall
121 health. In sexually monomorphic species, where there are no visible and easily
122 detectable morphological differences between sexes, the importance of monitoring and
123 maintaining an equal sex ratio grow stronger especially if the population is small and
124 endangered or is under consideration to be used as a reintroduction resource
125 (Legendre 2004). The success of reintroduction projects heavily depends on the
126 accurate determination of the sex of founding individuals (Griffiths & Tiwari 1995; IUCN
127 1998). In such species, the most common strategy is a random sampling of individuals
128 hoping for a balanced sex ratio in the founding group and this strategy heavily relies on
129 the implication that the sex ratio is equal at birth and that mortality and dispersal of
130 adults are equal between sexes (Bosé et al. 2007). Stochastic fluctuations of sex ratio in
131 small populations of monogamous species could substantially reduce effective
132 population size by increasing reproductive variance among individuals as well as
133 increasing extinction probability (Garofalo et al. 2016; Legendre et al. 1999). Therefore,
134 an equal sex ratio is one of the factors contributing to the maximum population growth
135 rate (Legendre 2004). On the other hand, the adult sex ratio, especially in monogamous
136 species, can sometimes be skewed toward males since different environmental
137 pressures affects opposite sexes differently. Traditional sex determination of individual
138 adult Griffon vulture is difficult, as is the case with other monomorphic species, and
139 requires highly invasive methods like cloacal examination (Tacha & Lewis 1979) and
140 laparoscopy (Fry 1983). Thus, molecular sexing was developed to replace these
141 invasive and difficult methods for fieldwork (Fridolfsson & Ellegren 1999). There are two
142 possible ways of acquiring samples for molecular sexing: capturing adult birds or
143 sampling chicks in the nests. Considering the difficulties of capturing adult birds as well
144 as the physical stress they experience in the process it is more efficient and humane to
145 sample birds while they are in the nests (Beja-Pereira et al. 2009). This sampling
146 strategy initially provides only the information about the offspring sex ratio, but thanks to
147 the marking of the chicks, these birds can later be monitored as adults thus enabling
148 information about the adult sex ratio. On the other hand, it is important to know the sex
149 ratio of the hatched birds because the sex ratio among adults can be influenced by
150 external factors like different life traits or deaths caused by accidents that can randomly
151 change the sex ratio among adult birds.
152   The Serbian Griffon vulture population experienced a similar fate as other
153 populations in Europe in the form of serious population decline. Although it made it
154 prone to drift and bottleneck effects this population retained significant genetic variability
155 similar to those detected in stable and large populations (Davidovic et al. 2020;
156 Davidovic et al. 2022). Nevertheless, there is always a concern about the relatedness
157 and kinship of remaining birds due to their natal philopatric behaviour and inbreeding
158 imprint on the population's overall fitness. For genetic-based estimates of relatedness,

159   in populations with no available pedigree data, traditional markers of choice are
160   microsatellites, usually a panel of eight - 30 microsatellite loci (Attard et al. 2018; Galla
161   et al. 2020; Jones et al. 2010; Pemberton 2008; Taylor 2015; Wang 2011). Traditionally,
162   the relatedness of two individuals is the probability that two randomly chosen alleles are
163   identical by descent (IBD) (Speed & Balding 2015). Currently, available relatedness
164   estimation methods can produce varying results since each takes into account different
165   assumptions about the input data (Hogg et al. 2018). The program COANCESTRY
166   (Wang 2011) reflects population reality because it takes inbreeding into account. It is
167   especially useful in inferring relationships in populations with historically recorded
168   bottleneck events. COANCESTRY estimates two types of identification of related
169   individuals. The first estimate is relatedness (r) which is a continuous measure of the
170   proportions of the alleles of the dyad (pair of individuals) that are IBD relative to a
171   reference population. The second type infers kinship which are discrete categories of
172   dyad relationships (parent-offspring, full-sibling, half-sibling). Additionally, an inbreeding
173   coefficient can be calculated for an individual as well as for the population in question.
174   Inbreeding coefficients may serve as predictors of the population`s overall fitness and
175   molecular-based pedigrees may be useful in the management and further reintroduction
176   efforts enabling the choice of unrelated founder individuals (Galla et al. 2020).
177   In this paper, we evaluated the sex ratio and estimated the relatedness among
178   Griffon vulture offspring hatched in Serbia. The aims of this study were to first assess
179   the impact of the bottleneck event that occurred in the early 1990s and second to further
180   investigate the potential of using this population as a stock population for reintroduction
181   in the Balkan Peninsula.
182

Materials & Methods
Population location and sampling

185   The sampling of the Griffon vulture population that inhabits the gorge of the river
186   Uvac in the Southwestern part of Serbia was performed in the period from 2013-2021.
187   The sampling was conducted directly in the nests as described in Davidovic et al.
188   (2020). Blood samples were collected via wing venal puncture using sterile equipment
189   with respect to animal welfare following relevant guidelines and regulations. During the
190   sampling, a thorough morphological analysis of chicks was performed (weight, bird
191   length, wing length, tail length, hindquarters and finger length, tarsus length, beak
192   length, width and height) and they were tagged by the licensed personnel. None of the
193   birds were sacrificed or injured during the sampling process and all the birds were
194   immediately returned to their original nests, or in the case of the captured/recaptured
195   subadult birds, released to nature. In total, we have collected the blood samples from
196   169 young birds that were stored either in sodium citrate or QUEENS buffer at -20°C. In
197   among 169 blood samples taken from the birds, four belonged to the birds that were

198   recaptured in 2013 but were previously marked in the year of their hatching 2011 (one
199   bird) and 2009 (three birds). All the procedures applied in this study were reviewed and
200   approved by the Ministry of Nature Protection of the Republic of Serbia, the Ministry of
201   Agriculture, Forestry and Water Management of the Republic of Serbia (323-07-
202   09135/2020-05/1) and the ethical committee of the Institute for Biological Research
203   "Siniša Stanković", Belgrade (323-09135-2020-05).

204   **DNA extraction**
205   The DNA extraction was performed by using the GeneJET Genomic DNA
206   Purification Kit (Thermo Fisher Scientific Cat.No. K0721) and a modified salting-out
207   protocol as described in Davidovic et al. (2022). The quality of the DNA extracts was
208   checked both by the spectrophotometer (NanoPhotometer, IMPLEN, Germany) and
209   agarose gel electrophoresis.

210   **Molecular sexing**
211   Molecular sexing of all sampled birds was performed using the protocol
212   described in Chang et al. (2008). Fragments for the Chromo-Helicase-DNA binding
213   protein (*CHD*) located on chromosomes *Z* (*CHD-Z*) and *W* (*CHD-W*) were amplified in
214   two separate reactions using the following set of primers: P2 reverse primer 5'
215   TCTGCATCGCTAAATCCTTT 3' complementary for both *W* and *Z* chromosomes, and
216   two forward primers *CHD-ZW*-common 5' GATCAGCTTTAATGGAAGTGAAG 3' and
217   *CHD-W*-specific 5' GGTTTTCACACATGGCACA 3'. Amplification was performed in a
218   volume of 25 µL with the following final concentrations of reaction components: 1×Taq
219   Buffer with $(NH_4)_2SO_4$, 2.5 mM MgCl2, 0.8 mM dNTP mix, 1 U of reverse Taq
220   polymerase (all components were produced by Thermo Fisher Scientific, EU) and 10
221   pmol of each forward and reverse primer. All PCRs were performed using the following
222   program: one cycle of initial denaturation at 94 °C for 5 min, after which there were 38
223   cycles of 35 s at 94 °C, 35 s at the annealing temperature at 56 °C and 35 s at 72 °C.
224   The step of final elongation was performed at 72 °C for 10 min. All amplified fragments
225   were evaluated by agarose gel electrophoresis. The lengths of the PCR fragments were
226   as follows: 148 bp for *P2/CHD-ZW*-common set of primers and 258 bp for *P2/CHD-W*-
227   specific set of primers. Both PCR fragments (148bp and 258 bp) were detected for the
228   heterogametic female birds and for the homogametic male birds only one fragment (148
229   bp) was detected. Fragments of interest were amplified in a 2720 Thermal Cycler
230   (Applied Biosystems, UK). A portion of the total analysed birds (54 birds) whose sex
231   was previously determined for the period of 2013-2018 were taken from the Hribsek
232   (2022).

233   **Sequencing**
234   Sequencing of mtDNA regions *Cytb* and *D-loop I* were performed for the chosen birds in
235   order to determine their maternal relationship. The chosen fragments were amplified
236   with the following set of primers: for *Cytb* forward GF-L13740 5'
237   TAATCAACAACTCCCTAATCGACCTAC 3', reverse GF-H15014 5'

238 CCTTTTGGGCCGAGAACTCT 3' and for *D-loop I* forward GF-L14961 5'
239 GCTCAACCACTAAATACTCTAATAG 3', reverse GF-H16307 5'
240 CAGTTTAGGGGGGAAGGAAGG 3'.
241   The same PCR program was used for the amplification of both *Cytb* and *D-loop I*
242 fragments: one cycle of initial denaturation at 94 °C for 5 min, after which there were 35
243 cycles of 35 s at 94 °C, 35 s at the annealing temperature at 55 °C and 90 s at 72 °C.
244 The step of final elongation was performed at 72 °C for 10 min. The lengths of the PCR
245 fragments were as follows: 1320 bp for the GF-L13740/GF-H15014 for *Cytb* set of
246 primers and 1391 bp for the GF-L14961/GF-H16307 set of primers for *D-loop I*.
247 Amplification of the mtDNA regions was performed using the same reaction
248 components as the ones used for the PCR amplification of *CHD* genes. The sequencing
249 was performed by Macrogen Europe using the same set of primers that were used for
250 the PCR amplification of *Cytb* while the *D-loop I* region was sequenced with a different
251 pair of primers in order to obtain a better quality sequence and avoid the T and C tract
252 (GF-L15147 5' TGTACATTACACTATTTGCCCCATA 3' and GF-H15547 5'
253 GCAGGGGGAAAGTAAGATCC 3'). Mitochondrial haplotypes were determined by
254 comparison with the mtDNA reference sequence for Griffon vulture, NC_036050.1
255 (Mereu et al. 2017), following the nomenclature that includes the nucleotide position and
256 adequate substitution different to the reference mitogenome. *Cytb* sequences used in
257 this analysis were published in Davidovic et al. (2022) and can be accessed by the
258 following accession numbers: OL962644, OL962685, OL962664, OL962652,
259 OL962667, OL962636, OL962648, OL962634, OL962646 and OL962649. Newly
260 sequenced *D-loop I* sequences were deposited in GenBank under the following
261 accession numbers: ON988042-ON988051.
262 **Data analysis**
263   For the estimation of relatedness, *r*, among the pair of birds (dyads) we used 57
264 presumably unrelated birds originating from 56 different nests from the total collected
265 sample of 169 birds. In this analysis we used the published genetic data of ten
266 microsatellite loci (Davidovic et al. 2020). In order to test the results of relatedness
267 estimates, we used two birds from the same nest (as known full-siblings, marked S95
268 and S53) thus having a total of 57 individuals in our analysis. The estimation was
269 performed using the COANCESTRY program (Wang 2011). All seven relatedness
270 estimators [TrioMl, (Wang 2007); Wang, (Wang 2002); LynchLi, (Li et al. 1993; Lynch
271 1988); LynchRd (Lynch & Ritland 1999); Ritland, (Ritland 1996); QuellerGt, (Queller&
272 Goodnight 1989); DyadMl, (Milligan 2003)] including three inbreeding estimators
273 [TrioMl, (Wang 2007); Ritland, (Ritland 1996); DyadMl, (Milligan 2003)] implemented in
274 the COANCESTRY program were tested in this analysis. Bootstrapping of 10000
275 samples over loci were performed in order to obtain the 95% confidence interval (95CI)
276 of each estimator for each dyad. Using the COANCESTRY program with the same
277 parameter settings, as previously defined, we evaluated the inbreeding coefficients, *F*,

278   for each analysed bird. The threshold values of r_{xy} coefficient were as follows: for r_{xy} <
279   0.09375 the birds in the dyad were unrelated, r_{xy} < 0.1875 and ≥ 0.09375 were
280   considered as third-order relatives (relationship), r_{xy} values < 0.375 and ≥ 0.1875 were
281   considered as second-order relatives (relationship) and r_{xy} values ≥ 0.375 were
282   considered as having full-sibling or parent-offspring relationship (del Pazo et al. 2021;
283   Lynch & Ritland 1999). In order to confirm acquired parent-offspring, full-siblings and
284   half-siblings relationships additional analysis in the software package ML-Relate
285   (Kalinowski et al. 2006) was performed.
286   Pearson's χ^2 goodness-of-fit statistics (Wilson & Hardy 2002) was used to test
287   the statistical significance for departure from an expected sex ratio of 0.5.
288
289   **Results**
290
291   **Sex ratio**
292   The results of the molecular sexing for the period of 2013-2021 are presented in
293   Table 1. The number of hatched chicks per year was in constant increase with the
294   exception of the years 2014 and 2021. The number of sampled chicks per year varied
295   from 13 to 29 which heavily depended on the weather conditions during the sampling
296   process as well as the availability of the nests that are mostly located on cliffs with
297   difficult access. Some of the nests were visited multiple times during this period. The
298   chicks originating from the supposed same breeding pair of birds that were hatched in
299   different years were sampled. The results of molecular sexing for the chicks originating
300   from the same nests are presented in Table S1. These data show the gender alteration
301   in the nest by year. In some of the nests, one of the sexes was more prominent than the
302   other while there are also nests with equal distribution of sexes throughout the sampling
303   period. During the period of nine years, we were able to sample between 12.04%-
304   17.68% of all hatched chicks per year and the only year with low sampling success was
305   2019 when only 8.97% of the hatched chicks were sampled mainly because of the bad
306   weather conditions (Table 1). Due to the incident in which a number of young birds fell
307   from the nests into the lake, the sampling success of the hatched chicks in 2020 was
308   somewhat higher compared to previous years. Four birds that fell from the nests were
309   rescued, marked and their blood samples were taken for genetic analyses.
310   The sex ratio among hatched chicks for each year is presented in Table 1. There
311   was no statistically significant deviation from the expected sex ratio of 0.5 (Table 2.)
312   although, somewhat skewed results were observed in 2013 (62.50% males towards
313   37.50% females) and 2020 (37.93% males towards 62.07% females) (Table 1.). In
314   addition, the sex was determined for the four birds recaptured in 2013, although they
315   were first captured and marked in 2009 (three females S033, S052, BE423) and in 2011
316   (male S078). These birds were not included in the statistics for the year 2013 since they
317   were born in different years before the blood sampling from the fledglings became a

318   part of the regular monitoring. No statistically significant departure from the expected
319   sex ratio was observed when all sexed individuals in the period 2013-2021 were
320   compared (Table 2.) even though a slightly higher proportion of the female fledglings
321   can be observed (Table 1.).
322
323   **Relatedness**
324   Relatedness was estimated for the 57 birds originating from 56 different nests.
325   According to the reproductive behaviour of Griffon vultures, birds originating from the
326   different nests should not be closely related which is why these 57 birds were chosen
327   for the relatedness analysis. During a lifetime, one pair of Griffon vultures uses the
328   same nest throughout their reproductive period so there should be no siblings among
329   chicks originating from different nests. The genetic data used for the estimation of
330   relatedness were obtained from the analysis of ten microsatellite loci analysed in the
331   previous work of Davidovic et al. (2020) and presented in Table S2.
332   The relatedness was calculated using all seven available estimators in the
333   COANCESTRY software. Of all estimators, the TrioML estimator (Wang 2007) exhibited
334   the lowest variance and it was chosen to represent the relatedness among the pair of
335   birds (dyads) in the analysed sample (Tale S3, Table 3.) (Wang 2011). TrioML uses a
336   likelihood method that estimates the IBD coefficients using the multilocus genotypes of
337   the focal dyad as well as a reference individual (Wang 2007; Wang 2011). In the same
338   way, this estimator can calculate individual inbreeding coefficients using the estimates
339   of the nine IBD coefficients (Wang 2007; Wang 2011). The mean value of the
340   relatedness r_{xy} among the sampled birds together with the acquired variance is
341   presented in Table 3. According to the TrioML estimator, each bird from the analysed
342   sample of the Griffon vulture population from Serbia on average has almost three
343   relatives of the full-sibling or parent-offspring relationship, almost eight half-siblings or
344   second-order relatives and around nine third-order relatives (Table 3, Table S3). When
345   95CI for TrioML estimator is taken into account, the number of relatives per bird is
346   lower, around one bird of second-order relationship and two birds of third-order
347   relationship while the number of birds with full-sibling or parent-offspring relationship is
348   rather low and close to zero (Table 3, Table S4). Overall, each bird in the sample has
349   around 20 relatives according to the TrioML estimator and around three relatives
350   according to 95CI for the same estimator (Table 3, Table S3-S4). The complete list of
351   mean relatedness for each estimator with appropriate variance and correlation between
352   estimators is presented in Table S5. The number of relatives that each individual bird
353   has according to the TrioML estimator is presented in Table S6. Inbreeding coefficients,
354   *F*, for each bird used in this analysis are presented in Table S7. According to the TrioML
355   estimator which exhibits the lowest variance the mean *F* = 0.08268 (Table S7.).
356   The highest positive correlation between the different relatedness estimators can
357   be observed between the TrioML and DyadML estimators, which further confirms that

358   the relatedness estimates provided by the TrioML estimator best describes the
359   relatedness of the birds in the given sample (Table S4, Figure S1). Furthermore, TrioML
360   exhibited positive and statistically significant correlations with the other estimators
361   (Table S5.), further confirming the proposed relatedness among the samples.
362   Relatedness estimates of 95CI TrioML were used to reconstruct the family tree which is
363   presented in Figure S2, and the dyads that demonstrated the full-sibling or parent-
364   offspring relationship are presented in Figure 1. In order to assess if the obtained
365   relationship can be categorized as a full-sibling or parent-offspring relationship, we have
366   used the ML-Relate software. This software suggested that the dyads of interest exhibit
367   the following relationship: S31/SF1 - full-siblings, S95/S53 - full-siblings (same nest),
368   SA3/S01 - full-siblings, BE423/S34 - parent-offspring and S033/S43 – full-siblings.
369   To further evaluate the maternal relationship between the proposed full-siblings
370   we have analysed the mtDNA of the birds from these dyads. The resulting haplotypes
371   are presented in Table 4. Birds in the dyads S31/SF1 and S95/S53 possess the same
372   mtDNA haplotypes which suggest that the birds in dyads have direct maternal ancestor
373   which supports the full-sibling relationship. For the dyads SA3/S01 and S033/S43 the
374   direct maternal ancestor can be excluded since their mtDNA haplotypes are different.
375   The sequencing of the mtDNA regions for the BE423 sample showed that there is no
376   direct maternal link thus negating the parent-offspring relationship in BE423/S34 dyad.
377   The observed discrepancy between the mtDNA and microsatellite data could be the
378   result of the marker choice or specific demographic history of the Griffon vulture
379   population from Serbia that underwent a strong bottleneck event at the end of the 20th
380   century.
381

382   Discussion

383

384   Assessing the sex ratio in the population is an important factor in determining the
385   strength and health of a population and is of great importance considering the managing
386   strategies for small and endangered populations (Ghorpade et al. 2012; Griffiths &
387   Tiwari 1995; Lambertucci et al. 2013; Robertson et al. 2006). In monogamous species,
388   the optimal sex ratio among adult birds that maximizes population growth is 1:1
389   (Legendre 2004) and it is expected to be in equilibrium at birth (Cockburn et al. 2002).
390   In our sample of the Griffon vulture population from Serbia, we didn't detect a
391   statistically significant deviation from the expected sex ratio of 1:1. The balanced sex
392   ratio in Griffon vulture populations were also detected in the native population in Spain
393   (Wink et al. 1998) and populations in France, both native and reintroduced, (Bosé et al.
394   2007), while in Italy the sex balance in the reintroduced population was shifted towards
395   males (Garofalo et al. 2016). The skewed sex ratio in Italy could be the result of external
396   factors such as the sex-biased stock of founders used for reintroduction. It was
397   recognized as an alarm pointing to the importance of proper restocking strategies and

398   the necessity to maintain sex balance in the groups of released birds as well as
399   performing constant monitoring of small and endangered populations (Garofalo et al.
400   2016). Maintaining a balanced sex ratio was also shown to be important for population
401   overall fitness in other monogamous and sexually monomorphic species of *Gyps* genus
402   such as the Oriental White-backed Vulture (*Gyps bengalensis*) or Long-billed Vulture
403   (*Gyps indicus*) (Arshad et al. 2009a) and distant groups of monogamous and sexually
404   monomorphic species like King penguins (*Aptenodytes patagonicus*) (Bordier et al.
405   2014). In our sample, over the course of nine years, we have detected two years with
406   skewed sex balance, although not statistically significant. In the year 2013, the balance
407   shifted towards males while the year 2020 demonstrated the shift towards females.
408   Strong and viable populations that inhabit a healthy environment with enough resources
409   should exhibit a balanced sex ratio (Fisher 1930) while it was shown that in the case of
410   the poor condition of parents or deprived resources parents tend to invest in less costly
411   sex (Brommer et al. 2003; Chakarov et al. 2015; Ferrer et al. 2009; Rosenfield et al.
412   2015; Trivers& Willard 1973). The study of pollutant influence in the environment, like
413   mercury contamination, on the formation of sex in birds demonstrated the skewed
414   balance toward females among fledglings in three different bird species where females
415   are considered as less costly sex: belted kingfisher (*Megaceryle alcyon*), eastern
416   bluebird (*Sialia sialis*) and tree swallow (*Tachycineta bicolor*) (Bouland et al. 2012). This
417   negative influence on overall health further confirmed the hypothesis that the parents
418   with poorer health tend to favour less costly sex among their offspring, the females
419   (Bouland et al. 2012). On the other hand, in raptors, where sexual dimorphism is
420   present, females are larger than males making males the less costly sex in terms of
421   parental care investment (Pleasants 1988., Trivers & Willard 1973). Although the Griffon
422   vulture belongs to the raptor group, they are sexually monomorphic, meaning there is
423   no observable difference in the cost of breeding males or females. The skewed balance
424   in 2013 could be explained either by the stochasticity during the sampling process or by
425   the influence of the environment that made the parents invest more in males. It has
426   been shown that in some circumstances when birds have greater access to food or are
427   in a better condition they tend to produce more males (Bradbury and Blakey 1998,
428   Sutherland 2002., Whittingham & Dunn 2000.). The skewed balance in 2020 may be
429   contributed to the fact that four birds that fell from the nests during an incident in the
430   Special nature reserve "Uvac'' were included in the sampling. The incident included the
431   illegal flyover of the helicopter which forced a number of birds to prematurely jump from
432   their nests into the lake. Among the four birds that were saved from drowning three
433   were females. Overall, the presented nine-year data show no deviation from the
434   balanced sex ratio among offspring which further supports the findings that the Griffon
435   vulture population of Serbia is strong and viable and strengthen its prospects to be used
436   as a stock population for further reintroduction efforts in the Balkan Peninsula
437   (Davidovic et al. 2020; Davidovic et al. 2022; Marinkovic et al. 2021).

438 Calculated absolute relatedness variation among relatedness estimators for our dataset
439 is expected since not all of them take into account inbreeding and/or genotyping errors
440 and have different mathematical methods for sample size accommodation and allele
441 weighing (Taylor 2015). Due to its lowest variance and the fact that it takes inbreeding
442 into account, TrioML was found to be the most reliable estimator for our data but other
443 estimators may be more suitable in other circumstances (Csillery et al. 2006; Hogg et al.
444 2018; Rollins et al. 2012; Taylor 2015). According to the TrioML estimator, each bird in
445 our sample has around 20 relatives and around three relatives according to 95CI for the
446 same estimator suggesting that percentage of related birds is around 5.3%. This value
447 is comparable to the one found in other birds of prey like Aplomado falcon (*Falco*
448 *femoralis*) whose values ranged from 5.2-8.8% (Johnson et al. 2021). In addition, the
449 detected level of relatedness exhibits nearly four times lower values of relatedness
450 compared to the values found in endangered Egyptian vulture (*Neophron percnopterus*)
451 in Spain (Blanco & Morinha 2021). Further evaluation of the mtDNA lineages of the
452 most closely related dyads showed the importance of the proper choice of relatedness
453 estimator and the need to include other markers in the analysis. While *Cytb* and *D-loop I*
454 sequences corroborated full-sibling relationship for our test birds from the same nest
455 (dyad S95/S53) and another pair originating from different nests (dyad S31/SF1), the
456 other two proposed full-sibling dyads (SA3/S01 and S033/S43) have different mtDNA
457 haplotypes negating the full-sibling status. The proposed parent-offspring relationship in
458 BE423/S34 dyad also can be rejected since not only do these birds have different
459 mtDNA haplotypes but records showed that the proposed mother was rescued and in
460 recovery in the Palić zoo in the year the proposed son was hatched and marked.
461 Discrepancies in the full-sibling dyad may be contributed to the fact that although the
462 full-siblings will share 50% of their alleles on average more variance from dyad to dyad
463 is expected, making related individuals closely related than in reality due to the choice
464 of the marker (Weir et al. 2006). It also may suggest an inbreeding signature on neutral
465 genetic variability and bottleneck effect. Most of the now living birds in the Serbian
466 Griffon vulture population are descendants from 10 breeding pairs in the 1990s and it is
467 expected that some birds may share a higher proportion of the same alleles due to the
468 stochasticity of a bottleneck. This is further corroborated by the mean F which is
469 0.08268, indicating that the parents of the average individual in the Serbian population
470 are more closely related than outbred first cousins whose offspring by definition have F
471 = 0.0625 (Nietlisbach et al. 2017). Although it seems high, the inbreeding level of 8%
472 detected for the Griffon vulture population of Serbia is not unusual compared to other
473 analysed populations in Europe. The same level of inbreeding was detected in the
474 population from Israel, while the population from Cyprus exhibited a lower value, F =
475 0.07 (Arshad et al. 2009b). The lowest level of inbreeding was detected in the Griffon
476 vulture population from Spain (F = 0.01) which is expected since this population is the
477 biggest and the most stable population in Europe and does not have a history of

478   dramatic population decline (Arshad et al. 2009b). Furthermore, when compared with
479   the inbreeding levels of the related species with similar biology and behaviour, like the
480   African Cape vulture (*Gyps coprotheres*), the *F* value is well in the range found in
481   different populations of this species (-0.5 to 0.19) (Kleinhans& Willows-Munro 2019). In
482   the same study, the inbreeding levels were analysed for the South African populations
483   of Bearded vultures (*Gypaetus barbatus*), *F* = 0.17, Hooded vultures
484   (*Necrosyrtesmonachus*), *F* = -0.07, and African White-backed vulture (*Gyps africanus*),
485   *F* = 0.07 (Kleinhans & Willows-Munro 2019). Other related species such as Oriental
486   White-backed Vulture and Long-billed Vulture exhibited inbreeding levels of 0.01 and
487   0.07 respectively (Arshad et al. 2009b). The same inbreeding levels may be attributed
488   to the philopatric nature of the vultures as well as detected historical population declines
489   recorded in most of the analysed species and populations. What clearly distinguishes
490   the Serbian Griffon vulture population from other analysed ones is the fact that it retains
491   a high level of microsatellite diversity comparable to ones in abundant populations
492   (Davidovic et al. 2020).
493

Conclusions

495   Results presented in this study further support the evidence that the Serbian Griffon
496   vulture population is stable, healthy and has a great perspective to be used as a stock
497   population for further conservation efforts in the Balkan Peninsula from the genetic
498   diversity point of view. Even though a relatively high level of *F* was detected among the
499   birds from the Serbian Griffon vulture population, the detected values are in the range of
500   values found in other Griffon vulture populations and related species. In addition,
501   previous research showed comparable values of the genetic diversity on the level of
502   microsatellites and *Cytb* between the Griffon vulture population from Serbia and other
503   populations with the identification of private alleles (Davidovic et al. 2020) and
504   haplotypes (Davidovic et al. 2022) present only in the Balkan Peninsula. In order to
505   preserve the specific genetic diversity of the Griffon vulture population from the Balkan
506   Peninsula, the future restocking efforts in this region should use the birds originating
507   from the autochthonous Balkan population.
508

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516

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518

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The ratio of 1182 hatched and 169 sampled chicks of the Griffon vulture population from Serbia with the sex ratio for the period of 2013-2021

2013*- including the birds captured in 2009. (three female birds) and 2011. (one male bird);

#- total number of hatched chicks for the period 2013-2021 including the birds hatched in 2009. and 2011.; Nh - total number of hatched chicks detected in the nests for a given year; Ns - number of sampled birds; Nm – number of males; Nf – number of females; Rs - a percentage of sampled chicks in relation to the total number of hatched chicks for the given year; Rm - a ratio of male chicks in a sample for a given year; Rf - a ratio of female chicks in a sample for a given year

1 **Table 1. The ratio of 1182 hatched and 169 sampled chicks of the Griffon vulture**
2 **population from Serbia with the sex ratio for the period of 2013-2021**

Year	Nh	Ns	Nm	Nf	Rs (%)	Rm (%)	Rf (%)
2013	112	16	10	6	14.29	62.50	37.50
2014	108	13	7	6	12.04	53.85	46.15
2015	113	16	7	9	14.16	43.75	56.25
2016	119	21	11	10	17.65	52.38	47.62
2017	125	17	9	8	13.60	52.94	47.06
2018	144	20	9	11	13.89	45.00	55.00
2019	145	13	6	7	8.97	46.15	53.85
2020	164	29	11	18	17.68	37.93	62.07
2021	148	20	10	10	13.51	50.00	50.00
2013*-2021	1182#	169	81	88	14.29	47.93	52.07

3 2013*- including the birds captured in 2009. (three female birds) and 2011. (one male bird); #-
4 total number of hatched chicks for the period 2013-2021 including the birds hatched in 2009. and
5 2011.; Nh - total number of hatched chicks detected in the nests for a given year; Ns - number of
6 sampled birds; Nm – number of males; Nf – number of females; Rs - a percentage of sampled
7 chicks in relation to the total number of hatched chicks for the given year; Rm - a   ratio of male
8 chicks in a sample for a given year; Rf - a ratio of female chicks in a sample for a given year

Statistical significance of balance between the sexes for each year in the period of 2013-2021 for the sampled chicks of the Griffon vulture population from Serbia.

Parameters used for the analysis: χ^2=3.841, df=1.

2013*- including the birds captured in 2009 and 2011; Nh - total number of hatched chicks detected in the nests for a given year; p - statistical significance p-value

1 **Table 2. Statistical significance of balance between the sexes for each year in the period of**
2 **2013-2021 for the sampled chicks of the Griffon vulture population from Serbia.**

3 Parameters used for the analysis: χ^2=3.841, df=1.

Year	Nh	p
2013	112	0.80259
2014	108	0.93868
2015	113	0.90052
2016	119	0.96202
2017	125	0.95309
2018	144	0.92034
2019	145	0.93868
2020	164	0.80926
2021	148	1.00000
2013*-2021	1182#	0.96696

4 2013*- including the birds captured in 2009 and 2011; Nh - total number of hatched chicks
5 detected in the nests for a given year; p - statistical significance p-value

Relatedness estimates based on the TrioML estimator for individual pairs of diploid individuals and average numbers of birds that have full-siblings, second-order and third-order relationships for each bird in the sample of the Griffon vulture population

r_{xy} - mean relatedness for individual pairs of diploid individuals; Afs - average number of full-siblings per individual; Asor - average number of second order relatives per individual; Ator - average number of third order relatives per individual; Ar - average number of relatives (full-sibling/parent-offspring relationships, second-order relationships and third-order relationships)

1 **Table 3. Relatedness estimates based on the TrioML estimator for individual pairs of**
2 **diploid individuals and average numbers of birds that have full-sibling, second-order and**
3 **third-order relationships for each bird in the sample of the Griffon vulture population of**
4 **Serbia.**

	Relatedness (Estimate)	Relatedness (CI95)
r_{xy}	0.09669	0.01334
Variance	0.01701	0.00234
Afs	2.98 ± 2.21	0.17 ± 0.38
Asor	7.93 ± 3.14	0.95 ± 1.02
Ator	9.47 ± 2.74	1.86 ± 1.39
Ar	20.39 ± 4.62	2.98 ± 1.82

5 r_{xy} - mean relatedness for individual pairs of diploid individuals; Afs - average number of full-
6 siblings per individual; Asor - average number of second order relatives per individual; Ator -
7 average number of third order relatives per individual; Ar - average number of relatives (full-
8 sibling/parent-offspring relationships, second-order relationships and third-order relationships)

MtDNA haplotypes of the birds from the sample of the Griffon vulture population from Serbia that according to the COANCESTRY analysis exhibited the full-sib/parent-offspring relationship.

MtDNA haplotypes are determined by the analysis of *Cytb* and *D-loop I* regions and comparison to the reference mitogenome NC_036050.1.

1 **Table 4. MtDNA haplotypes of the birds from the sample of the Griffon vulture population**
2 **from Serbia that according to the COANCESTRY analysis exhibited the full-sib/parent-**
3 **offspring relationship.**

4 MtDNA haplotypes are determined by the analysis of *Cytb* and *D-loop I* regions and comparison
5 to the reference mitogenome NC_036050.1.

Sample ID	Dyad	*Cytb* haplotype	*D-loop I* haplotype
S31	S31/SF1	13884C 13887C 13896T 14650C	15188G 15752T 15791A 15830T
SF1	S31/SF1	13884C 13887C 13896T 14650C	15188G 15752T 15791A 15830T
S95	S95/S53	13884C 13887C 13896T 14650C	15188G 15752T
S53	S95/S53	13884C 13887C 13896T 14650C	15188G 15752T
SA3	SA3/S01	14650C	15188G 15618G 15752T 15830T
S01	SA3/S01	13884C 13887C 13896T 14650C	15188G 15752T
S34	S34/BE423	13884C 13887C 13896T 14650C	15188G 15752T 15791A 15830T
BE423	S34/BE423	13884C 13887C 13896T 14650C	15188G 15343G 15346T
S033	S033/S43	13884C 13887C 13896T 14650C 14682G	15188G 15343G 15346T 15451A 15752T 15791A 15830T
S43	S033/S43	14650C	15188G 15752T 15830T

Figure 1

Dyads representing full-sibling or parent-offspring relationships.

Dyads are in different colors representing the relationship between two different birds. Individual birds are represented with the ring markings that they have received during the annual monitoring activity (S31, SF1, S95, S53, SA3, S01, S34, BE423, S033 and S43); numbers in the brackets represent the year of hatching/marking; sex is represented using the following symbols: ♂ for males and ♀ for females. For better presentation the dyads were extracted from Figure S2.