#### Metagenomics analysis of the effects of Agaricus 1

#### microbial mycelia diversity bisporus on and 2

#### **CAZymes in compost** 3

5

4

Wanqiu Chang<sup>1,2</sup>, Weilin Feng<sup>2</sup>, Yang Yang<sup>3</sup>, Yingyue Shen<sup>2</sup>, Tingting Song<sup>2</sup>-, Yu Li<sup>1</sup><sub>2</sub> and

Weiming Cai<sup>2</sup> 6

7 8

- <sup>1</sup> Engineering Research Centre of Chinese Ministry of Education for Edible and Medicinal
- Fungi, Jilin Agricultural University, Changchun, Jilin, China 9
- <sup>2</sup> Institute of Horticulture, Zhejiang Academy of Agricultural Sciences, Hangzhou, Zhejiang, 10
- 11 China
- <sup>3</sup> Environment and Plant Protection Institute, Chinese Academy of Tropical Agricultural 12
- Sciences, Haikou, Hainan, China 13

14

- Corresponding Author: 15
- Yu Li<sup>1</sup>, Jilin Agricultural University, Changchun, Jilin, 130118, China. 16
- Email address: fungi966@126.com 17
- Weiming Cai<sup>2</sup>, Zhejiang Academy of Agricultural Sciences, Hanzghou, Zhejiang, 310021, 18
- China. 19
- Email address: caiwm0527@126.com 20

21

22

23

24

25

26

27

28 29

30

31

32 33

34 35

36

37

38

39

# **Abstract**

Agaricus bisporus growth alters the lignocellulosic composition and structure of compost. However, the use of compost by A. bisporus mycelia during large-scale tunneling production of A. bisporus is unclear owing to the difficulty to-of completely separate separating the mycelia from compost cultures. Besides, it is difficult to differentiate the enzyme activities of A. bisporus mycelia from the wider microbial community. Although few studies have performed comparative investigations of inoculated and uninoculated mushroom composts, research on the use of compost by A. bisporus mycelia is still limited. Therefore, this study employed macro-genomics analysis to examine the fermentation substrate of A. bisporus before and after mycelial growth, and elucidate the molecular mechanism of substrate utilization by A. bisporus mycelia from the perspective of microbial communities and CAZymes in the substrate. The results showed that *Thermobifida* was always the dominant genus in the composting and mycelial growth stages, and the abundance of *Thermobifida*, Thermostaphylospora, and Sphaerobacter significantly increased after mycelial growth, whereas that of Thermopolyspora, Rhodothermus, and Pseudoxanthomonas significantly decreased. Although the presence of A. bisporus mycelia caused dramatic changes in the microbial community, the functional differences were not very pronounced during this period. The relative abundance of A. bisporus mycelia increased by 77.57-fold after the completion of

mycelial colonization. Interestingly, laccase-producing strains associated with lignin degradation were mostly bacteria rather than fungi, and the main laccase-producing bacteria belonged to *Thermobifida* and *Thermostaphylospora*, with *Thermostaphylospora* presenting a significant increase, suggesting that these bacteria may play a synergistic role in lignin decomposition along with *A. bisporus* mycelia. These findings provide preliminary evidence for the molecular mechanism of compost utilization by *A. bisporus* mycelia and offer a reference for the development and utilization of strains related to lignocellulose degradation.

Keywords: *Agaricus bisporus*; Metagenome; Community succession; CAZymes; Lignocellulose; Laccase; Xylanase

## Introduction

Agricultural biomass wastes comprise organic substances generated by humans during agricultural activities (*Malool et al.*, 2021). As these wastes are produced in abundant quantities and pose disposal problems, there has been an increasing interest to develop efficient and safe strategies to utilize agricultural biomass waste. At present, compost is still a primary mode of organic matter degradation(*Wang et al.*, 2021).

Industrial-scale production of Agaricus bisporus, an edible mushroom with a long history of cultivation, has solved part of the problem of agricultural waste reuse to a certain extent. In the process of large-scale production of A. bisporus, agricultural wastes, such as wheat straw and chicken manure, are mainly used as raw materials for fermentation, which is both environmentally-friendly and economical, addressing the issue of reusing of agricultural waste to a certain extent. In recent years, significant improvements in the A. bisporus cultivation process have been achieved, and the application of tunnel inoculation has altered the cultivation pattern and increased the mycelial growth rate. It has been reported that the localized tunnel-growth model achieved a 13.6% increase in A. bisporus growth rate, when compared with the cultivation house growth model (Wang et al., 2021). However, only a few studies have performed comparative investigations of inoculated and uninoculated mushroom compost. Some studies have suggested that A. bisporus mycelial growth produces a range of extracellular enzymes that are involved in the degradation of the lignocellulosic fraction in compost. Lignin has been found to beis mainly degraded during A. bisporus mycelial growth stage (PIII), with an increase in guaiacyl lignin content (G-type lignin)(Wood et al., 1983)-, and the lignin-degradation products have been speculated to be the substrate for subsequent growth of A. bisporus (Jurak et al., 2015, Jurak et al., 2014). The decrease and changes in lignin during the mycelial growth stage can improve the digestibility of carbohydrates in the later growth phases. In a previous study, A. bisporus appeared to be the dominant fungal species based on visual observation of cropping beds; however, phospholipid fatty acid analysis (PLFA) conducted on mushroom compost revealed that A. bisporus mycelia accounted for 6.8% w/w of the mushroom compost after complete colonization, with only less than half of the mycelia being active(McGee et al., 2017).

Many studies have focused on the crucial role of bacteria and fungi in the degradation of organic compounds, and their diverse modes of action on organic matter decomposition. While bacterial growth becomes restricted owing to their enhanced propagation on the surface of organic matter that acts as the main source of nutrients, the fungal hyphae have <u>a</u> strong

penetrating ability. With regard to About concerning Agaricus growth, both mycelia and fruiting body production are not only dependent on the mushroom itself, but also on bacteria and other fungi in the substrate, and the microbial community dynamics can completely change at the end of the composting process(Song et al., 2021). It has been noted that the final secondary fermentation (PII) compost mainly comprised lignocellulosic components from wheat straw together with microbial biomass(Martínez et al., 2008). Pasteurization of the compost material prior tobefore inoculation has been found to result in the predominance of fungal community in the substrate, with A. bisporus becoming the major fungal strain and its mycelia subsequently colonizing the substrate by degrading the organic material to release nutrients(McGee, 2018). However, little is known about the composition and activity of the wider fungal community in the compost substrate besides A. bisporus throughout the mushroom cultivation process. Therefore, the present study aimed to reveal the utilization of compost substrate before and after A. bisporus mycelial growth and compare the differences in the microbial communities and enzyme families in the compost. Furthermore, the effects of A. bisporus mycelial growth on other microorganisms during large-scale cultivation of A. bisporus were determined to identify novel microorganisms with potential roles in lignin degradation.

#### **Materials & Methods**

83

84

85

86

87

88 89

90

91

92 93

94

95

96

97

98

99 100

101

102

103

104

105 106

107

108

109 110

111

112113

114

115

116

117118

119 120

121

122

123 124

125

## Sample collection and DNA extraction

Commercial strain A15 of A. bisporus from Sylvan (USA) was used in this study and stored in the Engineering Research Center of Chinese Ministry of Education for Edible and Medicinal Fungi (ERCCMEEMF) at Jilin Agricultural University (Changchun, China). The compost fermentation and mycelial culture experiments were performed at Zhejiang Longchen Modern Agricultural Science and Technology Co. Ltd, Jiaxing City, Zhejiang Province, China. The compost comprised wheat straw (90 t), chicken manure (83 t), peanut meal (3 t), and gypsum (9 t). Peanut meal was added as an auxiliary nitrogen source because the components of wheat straw and chicken manure in China were different from those in Europe and America. Before commencing compost fermentation, the initial C/N ratio of the compost was adjusted to 25:1. The straws were completely dampened and piled up for 1-3 days, and then the other materials were mixed and piled again for 2–3 days. The pre-compost was placed into the tunnel for primary fermentation (PI). The pile was turned for three times on days 2, 4, and 7, respectively. The treatment parameters were adjusted based on compost temperature to allow the material temperature to reach 70°C-80°C and remain constant for 6 days. Immediately after that, secondary fermentation (Compost-PII) was conducted for 6–7 days by pasteurization. After secondary fermentation, inoculation was performed when the temperature decreased to 24°C and NH<sub>3</sub> level was ≤10 mg/L. The temperature, humidity, air pressure, air volume, and other environmental factors were adjusted using Dutch Christiaens Group equipment for the intelligent control system. Under the optimal environmental conditions, A. bisporus mycelia could grow all over the compost after 18 days (Mycelium-PIII).

Subsequently, samples were collected from the uninoculated compost and mycelia-filled compost, respectively. Before sample collection, the composts in the top, middle, and bottom

layers of the reactor were fully mixed. All the samples were divided into two parts: one part was stored at 4°C for physicochemical analysis and the other was frozen at -80°C for DNA extraction. The genomic DNA was extracted from the samples using an Omega EZNA soil DNA kit, and the integrity, purity, and concentration of the extracted genomic DNA were examined by 1% agarose gel electrophoresis (100 V, 1.5 h), NanoDrop 2000, and Qubit 3.0, respectively. The extracted DNA was stored in an ultra-low-temperature freezer at -20°C and transported to OE Biomedical Technology (Shanghai, China) for sequencing.

#### **Analysis of compost physicochemical properties**

 The compost samples were dried in an oven at  $105^{\circ}$ C for 5 h to assess the moisture content. The dried samples were crushed and placed in the control box of a resistance furnace at  $600^{\circ}$ C for 2 h to determine the ash content. The total nitrogen (Total-N) and carbon (Total-C) content in the samples were determined using the Kjeldahl method and  $K_2Cr_2O_3$  oxidation. Determination of the pH values with an electronic pH meter (Mettler-Toledo Instruments Co., Ltd., Shanghai, China) using 10% (w/v) sample suspensions were used. Laccase and xylanase activities in the samples were evaluated using a Solarbio kit, and lignin, cellulose, and hemicellulose components were determined according to Van Soes method.

### Metagenome sequencing, assembly, and annotation

Metagenome sequencing was accomplished using the Illumina HiSeq platform with a 500-bp sequencing library. The raw data (raw reads) quality was pre-processed using Trimmomatic (*Bolger et al., 2014*) software, and optimized sequences were spliced and assembled using MEGAHIT (*Li et al., 2015*, *Li et al., 2016*) software based on De-Brujin Bruijn graph principle, and contigs with length < 500 bp were filtered out for subsequent analysis. The open reading frame (ORF) of the spliced contigs was predicted with Prodigal software (*Hyatt et al., 2010*). CD-HIT software was adopted to remove redundant- and non-redundant initial unigenes. The clustering parameters included 95% identity and 90% coverage. The clean reads of each sample were aligned to the non-redundant genes set (95% identity) using bowtie 2 software to calculate the gene abundance in the corresponding samples. The representative sequences in the non-redundant unigenes set were annotated to the obtained species information according to the best alignment attained by BLASTP (E value<1e-5) to National Center for Bio-technology Information (NCBI) Non-Redundant Database (Nr). Then, the sum of gene abundances for the corresponding species was used to calculate species abundance-.

#### Identification of carbohydrate-active enzymes

To evaluate the carbon utilization potential of microbial communities during *A. bisporus* mycelial growth, the non-redundant genes were compared with the carbohydrate-active enzymes database (CAZy) using DIAMOND software (e<1e<sup>-5</sup>)(*Buchfink et al.*, 2015). First, all proteins with the highest sequence similarity were screened and subjected to CAZy to search against sequence libraries with the families of glycoside hydrolases (GHs), auxiliary activities (AAs), carbohydrate-binding modules (CBMs), glycosyltransferases (GTs), polysaccharide lyases (PLs), and carbohydrate esterases (CEs). Then, the differences in the CAZyme family between the two samples (uninoculated compost and mycelia-filled compost) were compared and analyzed(*Donhauser et al.*, 2021)-.

## Data and statistical analyses

Raw data were entered and stored in Excel. The differences among the samples were

examined by independent samples t-tests with statistical significance at p < 0.05 and p < 0.01.

171 The data are presented as mean  $\pm$  standard deviation (SD). GraphPad Prism 8.0 software and

172 Origin 2021 were applied for statistical analysis and plotting, and a cloud platform

173 (<a href="http://www.cloudtutu.com/">http://www.cloudtutu.com/</a>) was employed for plotting.

174

175

176

177

178179

180

181

182

183

184

185 186

187 188

189

190

191

192 193

194

195

196

197

198

199 200

201

202

203

204205

206

207

208209

210211

212

## **Results**

#### Physicochemical properties of Compost-PII and Mycelium-PIII

After completion of PII, the material temperature was reduced using fans. Subsequently, A. bisporus was inoculated (4% (w/w)) and the compost substrate was filled with mycelial growth after 18 days of incubation at 22 °C–24 °C. During this period, the water content in the compost decreased from 66.48% to 61.77% with the increase in mycelial growth, whereas the ash content increased by 1.71% (Table 1). It must be noted that the water content can affect microbial activities, which in turn can influence the enzyme activities. During A. bisporus mycelial growth, carbon consumption predominantly increased, whereas nitrogen utilization was relatively less maintained at 2.12-2.13%, and no significant difference(Table 1). After mycelial growth, the pH of the compost decreased from 7.80 to 6.27. Analysis of the cellulose and lignin contents in the compost by Van Soes method revealed that the cellulose content and lignin content significantly decreased(Fig. 1). Furthermore, evaluation of the activities of several known carbon-source-degradation-related enzyme families indicated a moderate increase in xylanase and laccase activities after A. bisporus mycelial growth (Figs. 2B & 2C). Similar findings have also been reported in several previous studies that indicated a sharp increase in laccase and xylanase activities during mycelial growth(Zhang et al., 2019, Hildén et al., 2013, De Groot et al., 1998). It must be noted that xylanase is mainly associated with hemicellulose degradation, whereas laccase is predominantly involved in lignin degradation. The solubility of lignin in aqueous solutions was low, and the decrease in the pH of the culture material— may had have a significant effect on lignin solubility. In addition, the increase in protease(Figs. 2D) also promotes better development of the mycelium(Wang et al., 2021).

#### Diversity of microbial communities in Compost-PII and Mycelium-PIII

The effective data volume of each sample in this experiment was 11.23-17.22 G. The N50 statistics of Contigs were distributed between 1631-2349 bp, and the number of OFR in the set of non-redundant genes was 1029162 after redundancy. The annotation rates were 89.44%, 74.62%, 42.21% and 2.57% for the non-redundant genes compared with NR, eggNOG, KEGG and CAZy databases respectively. Metagenomics analysis indicated the dominance of bacterial community in Compost-PII and Mycelium-PIII samples (93.17% and 94.27%, respectively), followed by fungi (0.25% and 0.29%, respectively), whereas the archaeal abundance remained almost unchanged. However, the abundance of viruses declined with *A. bisporus* mycelial growth (1.18% in Compost-PII and 0.13% in Mycelium-PIII). A total of 181 phyla, 157 classes, 805 families, 3460 genera, and 22,567 species were detected in the samples. The six most prominent bacterial phyla were Proteobacteria, Actinobacteria, Chloroflexi, Planctomycetes, Bacteroidetes, and Firmicutes(*Zhang et al., 2014*), and the abundances of Actinobacteria and Planctomycetes significantly increased after *A. bisporus* mycelial growth(Fig. 3A). Intriguingly, numerous lignocellulose-decomposing bacteria have

213 been reported to belong Proteobacteria, Firmicutes, Actinobacteria, to 214 Bacteroidetes(Lewin et al., 2013, Pankratov et al., 2011). Proteobacteria and Bacteroidetes are 215 known to play a major role in organic matter degradation and C cycling (Wang et al., 2018), and Bacteroidetes can break down lignocellulose into short-chain fatty acids (Dodd et al., 216 2011). Therefore, it was speculated that Actinobacteria and Planctomycetes are the key fungi 217 that degrade lignocellulose at the late stage of mycelium growth. It is noteworthy that 218 219 although the relative abundance of the Basidiomycota phylum was low, it showed a great 220 increase(Fig. 3B).

#### Analysis of bacterial and fungal communities

present study, Thermobifida, Thermostaphylospora, Sphaerobacter, Thermopolyspora, Pseudoxanthomonas, and Rhodothermus were the predominant bacterial genera in the Compost-PII and Mycelium-PIII samples(Cao et al., 2019, Durrant et al., 1991). When compared with the Compost-PII samples, the relative abundances of *Thermobifida*, Thermostaphylospora, Sphaerobacter, Thermomonospora, and Chelatococcus were significantly increased in Mycelium-PIII samples; in contrast, the relative abundances of Thermopolyspora, Rhodothermus, and Pseudoxanthomonas presented the opposite trend (Fig. 4A). Analyses of the microbial community composition confirmed significant shifts in the microbial community structure between the two groups, and many of the enriched genera also co-varied with function. Moreover, the variability of these microbial communities may be correlated with nutrients, compost temperature, moisture content, and pH.

At the species level, the relative abundance of A. bisporus presented the highest increase among fungi, exhibiting a 77.57-fold increase after complete mycelial growth (Figs. 4D). Despite the relatively low mass of A. bisporus mycelia in the compost substrate and less than half of the mycelia being active, the A. bisporus mycelia massively proliferated to become dominant during this period. In addition to A. bisporus, the activities of other microorganisms, such as the bacteria (Figs. 4C) Thermostaphylospora\_chromogena, Thermomonospora\_sp.\_CIF\_1, Sandaracinaceae\_bacterium, and Chelatococcus\_composti, (Figs.4D) Spizellomyces\_punctatus, Rozella\_allomycis, and and Basidiobolus\_meristosporus, were also enhanced during A. bisporus mycelial growth.

#### **Analysis of CAZymes**

221

222

223

224

225

226

227228

229

230

231232

233

234

235

236237

238

239

240

241

242

243244

245

246

247

248249

250251

252

253

254255

256

The carbon-utilization potential of the microbial communities in the compost was assessed to evaluate the effects of altered substrate quantity and quality resulting from the shift in microbial activity during *A. bisporus* mycelial growth. Among the genes annotated with CAZy, in total, 431 different CAZyme families (229 GHs, 81 GTs, 48 PLs, 17 AAs, 16 CEs, and 40 CBMs) were detected in the samples. The most abundant enzyme classes at all temperatures were GHs and GTs, whereas those with the lowest abundance were PLs and AAs. At the family level, GTs were especially abundant in all the samples, with GT2 (cellulose/chitin synthase and other functions), GT4 (sucrose synthase and other functions), and GT83 (galacturonosyl transferase and other functions) being the most abundant (*Paixão et al.*, 2021, *Leadbeater et al.*, 2021)-.

Cellulose, hemicellulose, and lignin are major constituents of lignocellulose-containing raw materials (*Stech et al.*, 2014). In the present study, cellulose- and hemicellulose-degrading enzymes exhibited the highest activities in Compost-PII and Mycelium-PIII samples, with GH5, GH8, and GH9 families being the predominant cellulose-degrading enzymes and GH2,

GH10, GH11, GH26, and GH53 families being the major hemicellulose-degrading enzymes(Table 2). The GH2 family includes multiple enzymes, and it has been demonstrated that  $\alpha$ -1,3-L-arabinofuranosidase activity on substituted xylan does not improve compost degradation by A. bisporus. Additionally, the degradation of lignin is an oxidative process mainly attributed to secondary metabolism or to restricted availability of carbon and nitrogen, and lignin is usually not degraded as the sole carbon and energy source (Silva et al., 2010). In nature, it is generally attributed to the metabolism of basidiomycetes white-rot fungi, since they degrade lignin more rapidly and extensively than other microorganisms (Woiciechowski et al., 2013). Although not a wood-rotting fungus, the Agaricus bisporus still plays a key role in the degradation of lignin as a grass-rotting fungus. Lignin consumption is mainly accomplished by laccases, manganese-dependent peroxidases, lignin peroxidases, and versatile peroxidases, which are the major groups of ligninolytic enzymes produced by the white-rot fungi(Kracher et al., 2019). Although AA7 family genes have been reported to play a role in lignin degradation (Andlar et al., 2018) and the content of these genes was relatively high in the lignin-degradation-related enzymes family in the present study, this enzymes family was not significantly different. There is a large content of the AA3 and AA6 enzyme families associated with lignin breakdown(Table 2).

#### Variations in enzyme families between Compost-PII and Mycelium-PIII

The effect of *A. bisporus* mycelium on the compost substrate was mainly reflected in the enzyme families with low relative abundance at the population level. For instance, the abundances of GT32 and GH24 significantly decreased, whereas those of GH42, AA10, and CBM13 significantly increased. Furthermore, the abundance of GH8 presented a slight decrease(Figs. 5A). GH24 is known to act in association with lysozyme, GH8 is the main family of enzymes involved in cellulose degradation, and GH42 plays an important role in hemicellulose degradation. *Thermostaphylospora* belonging to Actinobacteria mainly causes an increase in the activities of  $\beta$ -galactosidase and  $\alpha$ -L-arabyranosidase of the GH42 family in the galactose metabolism pathway during the mycelial growth stage, and CBM67 has been reported to exhibit  $\alpha$ -L-rhamnose-binding activity(*Fujimoto et al.*, 2013).

In the present study, the increase in the laccase content during the mycelial growth stage was mainly related to the AA10 family(Table 2). AA10 (formerly, CBM33) proteins are copper-dependent lytic polysaccharide monooxygenases (LPMOs), and the functionality of C1/C4 oxidizing, cellulose-active, and chitin-active AA10 LPMOs has been documented in bacteria and fungi(*Bissaro et al., 2020*). The major strains that cause differential laccase production are *Thermobifida, Thermostaphylospora*, and *Cellulomonas*, and in the present study, the relative abundances of *Thermobifida* and *Cellulomonas* decreased, while that of *Thermostaphylospora* increased(Figs. 5B&5C). Furthermore, the increase in the xylanase content during *A. bisporus* mycelial growth stage was mainly related to the CBM13 family, and *Thermostaphylospora* was also the dominant genus that caused this increase(Figs. 5D&5E).

#### **Discussion**

Degradation of lignin during the composting process has been reported to be caused by certain fungi and several species of bacteria and Actinomycetes (Kabel et al., 2017, Fermor et

al., 1981). In the present study, the artificial introduction of A. bisporus at the end of the composting process and the resultant dominance of A. bisporus in the fungal community played a major role in lignin degradation. Similar finding has also been reported by Jean-Michel Savoie et al. (Savoie, 1998). During the mycelial growth of A. bisporus, the oxidative phosphorylation pathway became dominant, which subsequently affected the bacterial and fungal communities' composition, and a part of lignin degradation originated from bacterial action. Jurak et al. showed that xylan solubility increased by 20% during mycelial growth, indicating partial degradation of the xylan skeleton. In the present study, xylan degradation was mainly associated with the action of Thermostaphylospora. However, despite xylan degradation, the carbohydrate composition and degree of substitution of xylan in the compost at the beginning and end of the mycelial growth stage were rather similar (Jurak et al., 2014).

In a previous study, some researchers (*Zhang et al.*, 2014) determined the rRNA gene copy number of Actinomycetes and fungi during the composting process by quantitative PCR, and found that the fungal genus *Agaricus* and unknown fungal community accounted for 45% and 55% of the microbial community, respectively, while the bacterial genus Streptomyces accounted for 60% of the total bacterial community during the mycelial growth phase. In contrast, in the present study, although the abundance of Streptomyces varied before and after the mycelial growth stage, the difference was not significant. Moreover, besides Agaricus, Basidiobolus and Spizellomyces were also the predominant fungi, thus providing further insights into the composition of the fungal community in the compost.

During the A. bisporus mycelial growth stage, the relative abundances of gene sequences still remained high at high-temperature composting, and although the abundance of fungal communities increased, the number of bacterial genes was much higher than that of fungal genes. However, as it was not possible to determine the number of active bacteria during this stage, microbial communities with a higher relative abundance of gene sequences were compared, and A. bisporus was found to be dominant in the fungal community. While enzymes related to lignocellulose breakdown were not detected in the macro-genome of A. bisporus, analysis of whole-genome data of A. bisporus substrate confirmed the presence of a large number of genes encoding lignocellulose-degrading enzymes(Morin et al., 2012), because macrogenome sequencing results did not assemble genes encoding lignocellulosedegrading enzymes in A. bisporus. Moreover, variations were also noted in the expression of genes encoding CAZymes between compost-grown mycelia and fruiting body, with genes encoding plant cell wall degrading enzymes detected in compost-grown mycelia, but largely undetected in the fruiting body. Similarly, Patyshakuliyeva et al. also confirmed that compostgrown mycelia could express a large variety of CAZyme genes related to the degradation of plant biomass components(Patyshakuliyeva et al., 2013). In addition, transcriptomics and proteomics investigations performed in a previous study also demonstrated that genes related to lignin degradation were only highly expressed on day 16 of mycelial growth, indicating that lignin was degraded at the initial stage of mycelial growth and was no longer altered after complete growth of mycelia (Patyshakuliyeva et al., 2015). Moreover, compost-grown mycelia were found to express a large number of CAZymes-encoding genes associated with the degradation of plant biomass components. In summary, the present study uncovered lignocellulose-degrading microorganisms and enzyme expressions in bacteria during A.

bisporus mycelial growth stage in the composting process, providing further insights into lignin degradation in compost. The results obtained can further strengthen our understanding of the specificity of *A. bisporus* mycelial growth.

346 347

348

349

350

351 352

353

354 355

356

357

358

359

344

345

#### Conclusions

The present study found the dominance of *A. bisporus* during the mycelial growth stage. Although pasteurization during composting effectively inhibited lignin-degrading fungi in the compost substrate, these fungi competed with *A. bisporus* during the mycelial growth stage. As the majority of sequences belonging to the AA family were of fungal origin and the fungal community abundance increased after the mycelial growth stage, the potential for lignin-degrading enzymes of bacterial origin may be grossly underestimated. Furthermore, *Agaricus* proliferation may require an interacting consortium of both bacteria and fungi for effective lignocellulose degradation. The results obtained offer insights into the difference in enzyme activities between *A. bisporus* mycelia and other microbial communities, and enhance our understanding of the changes in microbial communities and enzyme families during *A. bisporus* mycelial growth phase in the composting process.

360

361

362

366

371

373

378

#### ADDITIONAL INFORMATION AND DECLARATIONS

# **Acknowledgements**

- 363 Sincere thanks to Zhejiang Longchen Modern Agricultural Science and Technology Co. Ltd, (Jiaxing
- 364 City, Zhejiang Province) for help in this work. We thank International Science Editing (
- http://www.internationalscienceediting.com ) for editing this manuscript.

## Funding

- This work was supported by the China Agriculture Research System (CARS-20), the Zhejiang Science
- and Technology Major program on Agriculture (2021C02073), Central Public-interest Scientific
- 369 Institution Basal Research Fund (NO.1630042022003), and New Variety Breeding and National Natural
- 370 Science Foundation of China (U20A2046).

#### Conflict of interest statement

The authors declare that they have no conflict of interest.

#### **Author Contributions**

- •Wanqiu Chang designed the experiments, analyzed the data, and approved the final draft.
- •Yang Yang and Tingting Song reviewed and modified the original draft.
- Weilin Feng and Yingyue Shen provide study materials, and reagents.
- •Yu Li and Weiming Cai conceived and designed the experiments.

## **Data Availability**

379 | Sequence data were deposited in the NCBI Sequence Read Archive under the BioProject Accession

References

380

381

382	Malool ME, Keshavarz Moraveji M, Shayegan J. 2021. Optimized production, Pb(II) adsorption and
383	characterization of alkali modified hydrochar from sugarcane bagasse. Scientific Reports
384	<b>11</b> :22328-22328. DOI 10.1038/s41598-021-01825-y.
385	Wang WK, Liang CM. 2021. Enhancing the compost maturation of swine manure and rice straw by
386	applying bioaugmentation. Scientific Reports 11:6103. DOI 10.1038/s41598-021-85615-6.
387	Wang Q, Xiao TT, Song XX, Shen XF, Jun JX, Zhang JX, Chen H, Huang JC. 2021. Effects of
388	Running Phase III in a Tunnel on Agaricus bisporus mycelial growth and compost
389	degradation. Acta Edulis Fungi 28:86-92. DOI 10.16488/j.cnki.1005-9873.2021.03.011.
390	Wood DA, Leatham GF. 1983. Lignocellulose degradation during the life cycle of Agaricus bisporus.
391	FEMS Microbiology Letters 20:421-424. DOI 10.1111/j.1574-6968.1983.tb00160.x.
392	Jurak E, Punt AM, Arts W, Kabel MA, Gruppen H. 2015. Fate of Carbohydrates and Lignin during
393	Composting and Mycelium Growth of Agaricus bisporus on Wheat Straw Based Compost.
394	PLoS One 10:1-16. DOI 10.1371/journal.pone.0138909.
395	Jurak E, Kabel MA, Gruppen H. 2014. Carbohydrate composition of compost during composting and
396	mycelium growth of Agaricus bisporus. Carbohydrate Polymers 101:281-288. DOI
397	10.1016/j.carbpol.2013.09.050.
398	McGee, Conor F. 2017. Microbial ecology of the Agaricus bisporus mushroom cropping process.
399	Applied Microbiology and Biotechnology 102:1075-1083. DOI 10.1007/s00253-017-8683-9.
400	Song TT, Shen YY, Jin QL, Feng WL, Fan LJ, Cao GT, Cai WM. 2021. Bacterial community diversity,
401	lignocellulose components, and histological changes in composting using agricultural straws
402	for Agaricus bisporus production. PeerJ 9:e10452. DOI 10.7717/peerj.10452.

403	Martínez áT, Rencoret J, Marques G, Gutiérrez A, Ibarra D, Jiménez-Barbero J, Río J. 2008.
404	Monolignol acylation and lignin structure in some nonwoody plants: A 2D NMR study.
405	Phytochemistry 69:2831-2843. DOI 10.1016/j.phytochem.2008.09.005.
406	McGee CF. 2018. Microbial ecology of the Agaricus bisporus mushroom cropping process. Applied
407	Microbiology and Biotechnology 102:1075-1083. DOI 10.1007/s00253-017-8683-9.
408	Bolger AM, Lohse M, Usadel B. 2014. Trimmomatic: a flexible trimmer for Illumina sequence data.
409	Bioinformatics 30:2114-2120. DOI 10.1093/bioinformatics/btu170.
410	Li D, Liu CM, Luo R, Sadakane K, Lam TW. 2015. MEGAHIT: an ultra-fast single-node solution for
411	large and complex metagenomics assembly via succinct de Bruijn graph. Bioinformatics
412	<b>31</b> :1674-1676. 10.1093/bioinformatics/btv033.
413	Li D, Luo R, Liu CM, Leung CM, Ting HF, Sadakane K, Yamashita H, Lam TW. 2016. MEGAHIT v1.0:
414	A fast and scalable metagenome assembler driven by advanced methodologies and
415	community practices. <i>Methods</i> <b>102</b> :3-11. DOI 10.1016/j.ymeth.2016.02.020.
416	Hyatt D, Chen GL, Locascio PF, Land ML, Larimer FW, Hauser LJ. 2010. Prodigal: prokaryotic gene
417	recognition and translation initiation site identification. BMC Bioinformatics 11:119. DOI
418	10.1186/1471-2105-11-119.
419	Buchfink B, Xie C, Huson DH. 2015. Fast and sensitive protein alignment using DIAMOND. Nat
420	Methods 12:59-60. DOI 10.1038/nmeth.3176.
421	Donhauser J, Qi W, Bergk-Pinto B, Frey B. 2021. High temperatures enhance the microbial genetic
422	potential to recycle C and N from necromass in high-mountain soils. Global Change Biology
423	<b>27</b> :1365-1386. DOI 10.1111/gcb.15492.
424	Zhang HL. Wei JK. Wang QH. Yang R. Gao XJ. Sang YX. Cai PP. Zhang GQ. Chen QJ. 2019.

425	Lignocellulose utilization and bacterial communities of millet straw based mushroom (Agaricus
426	bisporus) production. Scientific Reports 9:1151. DOI 10.1038/s41598-018-37681-6.
427	Hildén K, Mäkelä MR, Lankinen P, Lundell T. 2013. Agaricus bisporus and related Agaricus species
428	on lignocellulose: Production of manganese peroxidase and multicopper oxidases. Fungal
429	Genetics and Biology 55:32-41. DOI 10.1016/j.fgb.2013.02.002.
430	De Groot PW, Basten DE, Sonnenberg A, Van Griensven LJ, Visser J, Schaap PJ. 1998. An endo-
431	1,4-beta-xylanase-encoding gene from Agaricus bisporus is regulated by compost-specific
432	factors. Journal of biological chemistry 277:273-284. DOI 10.1006/jmbi.1997.1605.
433	Wang Q, Juan JX, Xiao TT, Zhang JJ, Chen H, Song XX, Chen MJ, Huang JC. 2021. The physical
434	structure of compost and C and N utilization during composting and mushroom growth in
435	Agaricus bisporus cultivation with rice, wheat, and reed straw-based composts. Applied
436	Microbiology Biotechnology 105:3811-3823. DOI 10.1007/s00253-021-11284-0.
437	Zhang X, Zhong YH, Yang SD, Zhang WX, Liu WF. 2014. Diversity and dynamics of the microbial
438	community on decomposing wheat straw during mushroom compost production. Bioresource
439	Technology 170:183-195. DOI 10.1016/j.biortech.2014.07.093.
440	Lewin A, Wentzel A, Valla S. 2013. Metagenomics of microbial life in extreme temperature
441	environments. Current Opinion in Biotechnology 24:516-525. DOI
442	10.1016/j.copbio.2012.10.012.
443	Pankratov TA, Ivanova AO, Dedysh SN, Liesack W. 2011. Bacterial populations and environmental
444	factors controlling cellulose degradation in an acidic Sphagnum peat. Environmental
445	Microbiology 13:1800-1814. DOI 10.1111/j.1462-2920.2011.02491.x.
446	Wang K, Mao H, Li X. 2018. Functional characteristics and influence factors of microbial community in

447	sewage sludge composting with inorganic bulking agent. Bioresource Technology 249:527-
448	535. DOI 10.1016/j.biortech.2017.10.034.
449	Dodd D, Mackie RI, Cann IK. 2011. Xylan degradation, a metabolic property shared by rumen and
450	human colonic Bacteroidetes. Molecular Microbiology 79:292-304. DOI 10.1111/j.1365-
451	2958.2010.07473.x.
452	Cao GT, Song TT, Shen YY, Jin QL, Cai WM. 2019. Diversity of Bacterial and Fungal Communities in
453	Wheat Straw Compost for Agaricus bisporus Cultivation. HortScience: a publication of the
454	American Society for Horticultural Science 54:100-109. DOI 10.21273/HORTSCI13598-18.
455	Durrant AJ, Wood DA, Cain RB. 1991. Lignocellulose biodegradation by Agaricus bisporus during
456	solid substrate fermentation. <i>Microbiology</i> <b>137</b> :751-755. DOI 10.1099/00221287-137-4-751.
457	Paixão DAA, Tomazetto G, Sodré VR, Gonçalves TA, Uchima CA, Büchli F, Alvarez TM, Persinoti GF,
458	Da Silva MJ, Bragatto J, Liberato MV, Franco Cairo JPL, Leme AFP, Squina FM. 2021.
459	Microbial enrichment and meta-omics analysis identify CAZymes from mangrove sediments
460	with unique properties. Enzyme and Microbial Technology 148:109820. DOI
461	10.1016/j.enzmictec.2021.109820.
462	Leadbeater DR, Oates NC, Bennett JP, Li Y, Dowle AA, Taylor JD, Alponti JS, Setchfield AT, Alessi
463	AM, Helgason T, McQueen-Mason SJ, Bruce NC. 2021. Mechanistic strategies of microbial
464	communities regulating lignocellulose deconstruction in a UK salt marsh. Microbiome 9:48.
465	DOI 10.1186/s40168-020-00964-0.
466	Stech M, Hust M, Schulze C, Dübel S, Kubick S. 2014. Cell-free eukaryotic systems for the production,
467	engineering, and modification of scFv antibody fragments. ENGINEERING IN LIFE
468	SCIENCES 14:387-398. DOI 10.1002/elsc.201400036.

469	Silva, Santos, Menezes. 2010. Degradation of lignosulfonic and tannic acids by ligninolytic soil fungi
470	cultivated under microaerobic conditions. Brazilian Archives of Biology and Technology
471	<b>53</b> :693-699. DOI 10.1111/j.1574-6941.1996.tb00348.x.
472	Woiciechowski AL, Vandenberghe LPDS, Karp SG, Letti L, Soccol CR. 2013. The Pretreatment Step
473	in Lignocellulosic Biomass Conversion: Current Systems and New Biological Systems.
474	Lignocellulose Conversion. DOI 10.1007/978-3-642-37861-4_3.
475	Kracher D, Ludwig R. 2019. Cellobiose dehydrogenase: an essential enzyme for lignocellulose
476	degradation in nature -a review. Die Bodenkultur: Journal of Land Management, Food and
477	Environment 67. DOI 10.1515/boku-2016-0013.
478	Andlar M, Rezić T, Marđetko N, Kracher D, Ludwig R, Šantek B. 2018. Lignocellulose degradation: An
479	overview of fungi and fungal enzymes involved in lignocellulose degradation. Engineering in
480	life science 18:768-778. DOI 10.1002/elsc.201800039.
481	Fujimoto Z, Jackson A, Michikawa M, Maehara T, Momma M, Henrissat B, Gilbert HJ, Kaneko S. 2013.
482	The structure of a Streptomyces avermitilis $\alpha\text{-L-rhamnosidase}$ reveals a novel carbohydrate-
483	binding module CBM67 within the six-domain arrangement. Journal of biological chemistry
484	<b>288</b> :12376-12385. DOI 10.1074/jbc.M113.460097.
485	Bissaro B, Kommedal E, Røhr ÅK, Eijsink VGH. 2020. Controlled depolymerization of cellulose by
486	light-driven lytic polysaccharide oxygenases. Nature Communications 11:890-890. DOI
487	10.1038/s41467-020-14744-9.
488	Kabel MA, Jurak E, Mäkelä MR, De Vries RP. 2017. Occurrence and function of enzymes for
489	lignocellulose degradation in commercial Agaricus bisporus cultivation. Applied Microbiology
490	and Biotechnology volume 101:4363-4369. DOI 10.1007/s00253-017-8294-5.

491 Fermor TR, Wood DA. 1981. Degradation of Bacteria by Agaricus bisporus and Other Fungi. 492 Microbiology 126:377-387. DOI 10.1099/00221287-126-2-377. 493 Savoie J-M. 1998. Changes in enzyme activities during early growth of the edible mushroom, Agaricus 494 Mycological Research **102**:1113-1118. DOI bisporus, in compost. 495 10.1017/S0953756298006121. Morin E, Kohler A, Baker AR, Foulongne-Oriol M, Lombard V, Nagy LG, Ohm RA, Patyshakuliyeva A, 496 497 Brun A, Aerts AL, Bailey AM, Billette C, Coutinho PM, Deakin G, Doddapaneni H, Floudas D, 498 Grimwood J, Hildén K, Kües U, Labutti KM, Lapidus A, Lindquist EA, Lucas SM, Murat C, 499 Riley RW, Salamov AA, Schmutz J, Subramanian V, Wösten HA, Xu J, Eastwood DC, Foster 500 GD, Sonnenberg AS, Cullen DD, De Vries RP, Lundell T, Hibbett DS, Henrissat B, Burton KS, 501 Kerrigan RW, Challen MP, Grigoriev IV, Martin F. 2012. Genome sequence of the button 502 mushroom Agaricus bisporus reveals mechanisms governing adaptation to a humic-rich 503 ecological niche. Proc Natl Acad Sci USA 109:17501-17506. DOI 10.1073/pnas.1206847109. Patyshakuliyeva A, Jurak E, Kohler A, Baker A, Battaglia E, De Bruijn W, Burton KS, Challen MP, 504 505 Coutinho PM, Eastwood DC, Gruben BS, Mäkelä MR, Martin F, Nadal M, Van Den Brink J, Wiebenga A, Zhou M, Henrissat B, Kabel M, Gruppen H, De Vries RP. 2013. Carbohydrate 506 507 utilization and metabolism is highly differentiated in Agaricus bisporus. BMC Genomics 14:663. 508 DOI 10.1186/1471-2164-14-663. 509

Patyshakuliyeva A, Post H, Zhou M, Jurak E, Heck AJ, Hildén KS, Kabel MA, Mäkelä MR, Altelaar MA,

De Vries RP. 2015. Uncovering the abilities of *Agaricus bisporus* to degrade plant biomass throughout its life cycle. *Environmental microbiology* 17:3098-3109. DOI 10.1111/1462-2920.12967.

510

511

512