Spatial patterns of coral survivorship: The impact of adult proximity versus other drivers of localized mortality for two brooding corals

David A Gibbs, Mark E Hay

Species-specific enemies may promote prey coexistence through negative distance- and density-dependent survival of juveniles near conspecific adults. We tested this mechanism by transplanting juvenile-sized fragments of the brooding corals *Pocillopora damicornis* and Seriatopora hystrix 3, 12, 24 and 182 cm up- and down-current of conspecific adults and monitoring their survival and condition over time. We also characterized the spatial distribution of *P. damicornis* and *S. hystrix* within replicate plots on three Fijian reef flats and measured the distribution of small colonies within 2 m of larger colonies of each species. Juvenile-sized transplants exhibited no differences in survivorship as a function of distance from adult P. damicornis or S. hystrix. Additionally, both P. damicornis and S. hystrix were aggregated rather than overdispersed on natural reefs. However, a pattern of juveniles being aggregated near adults while larger (and probably older) colonies were not suggests that greater mortality near large adults could occur over longer periods of time. While we found minimal evidence of greater mortality of small colonies near adult conspecifics in our transplant experiments, we did document hot-spots of species-specific corallivory. We detected spatially localized and temporally persistent predation on P. damicornis by the territorial triggerfish Balistapus undulatus. This patchy predation did not occur for S. hystrix. This variable selective regime in an otherwise more uniform environment could help maintain the diversity of corals on Indo-Pacific reefs.

1	Spatial patterns of coral survivorship: The impact of adult proximity versus other drivers of
2	localized mortality for two brooding corals
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9	Abstract
10	Species-specific enemies may promote prey coexistence through negative distance- and
11	density-dependent survival of juveniles near conspecific adults. We tested this mechanism by
12	transplanting juvenile-sized fragments of the brooding corals <i>Pocillopora damicornis</i> and

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Introduction

The processes maintaining high numbers of species in tropical rainforests and coral reefs	S
have long been investigated (Connell 1978). One suggested mechanism for maintaining diversit	У
is the Janzen-Connell hypothesis (Janzen 1970; Connell 1971), which proposes that species-	
specific enemies clustered near adults increase the local mortality of conspecific juveniles and	
prevent any single species from monopolizing resources. It has generally been applied to long-	
lived, stationary, terrestrial organisms such as trees (Zhu et al. 2013). Although there are	
examples of species-specific distance- or density-dependent mortality affecting community	
species richness (e.g., Packer and Clay 2000; Petermann et al. 2008; Bagchi et al. 2014), a meta	-
analysis found no general, net effect of distance from parent on offspring mortality across a	
variety of plant types, habitats, or life stages (Hyatt et al. 2003). Thus, some tree species may	
experience Janzen-Connell effects (Johnson et al. 2012) but the generality of the pattern has bee	n
difficult to document (Hyatt et al. 2003). In part, this may be because numerous other processes	
(habitat heterogeneity, spatial patterns of competitors, etc.) could obscure Janzen-Connell	
effects. This makes experimental tests difficult in field settings, especially when spatial scales	
over which they may be relevant are unclear.	
Research addressing Janzen-Connell effects on coral reefs is rare (Marhaver et al. 2013)	
Explanations for maintenance of coral diversity more often invoke disturbance regimes, abiotic	
gradients (e.g., light, sedimentation), and competition hierarchies (Lang 1973; Connell 1978;	
Buss and Jackson 1979; Porter et al. 1981). One reason for the paucity of tests in marine system	ıS
may be that the hypothesis assumes that dispersal decreases monotonically with distagrant from	
parents and that the average dispersal distance is greater than the average predation distance but	
on the same order of magnitude (Nathan and Casagrandi 2004), neither of which necessarily	

51	applies to marine species with pelagic larvae. Coral larvae are competent to settle for ys to
52	months after release (Richmond 1987; Miller and Mundy 2003; Nozawa and Harrison 2008) and
53	may disperse up to hundreds of kilometers (Jones et al. 2009; Torda et al. 2013). Therefore,
54	unlike seeds of many tree species, coral larvae need not be distributed as "seed shadows" with
55	juveniles clustered near parents.
56	Nevertheless, distance- or density-dependent mortality of juveniles could affect both
57	brooding and broadcasting corals because larvae select settlement sites and can be attracted to
58	the chemical cues of conspecifics (Dixson, Abrego, and Hay et al. 2014); this could lead to
59	larvae settling near conspecific adults or in aggregations (e.g., Dunstan and Johnson 1998).
60	Additionally, brooding corals may cast larval shadows akin to terrestrial seed shadows because
61	larvae from brooding corals frequently settle quickly and close to their parents (Carlon and Olson
62	1993; Tioho, Tokeshi, and Nojima 2001; Vermeij 2005; Vermeij and Sandin 2008; Torda et al.
63	2013)
64	Selective mortality of juveniles near conspecific adults is often assumed to be due to
65	specialist enemies that accumulate near adults over their lifetimes. While this may be the case for
66	terrestrial plants where many herbivores and pests are specialists (Bernays 1989), it is unclear to
67	what extent this applies to corals, of which there are relatively few identified species-specific
68	consumers (Copell and Karlson 2000; Rotjan and Lewis 2008) that might be expected to
69	accumulate near adults of specific prey species.
70	However, there is growing evidence from both terrestrial (Packer and Clay 2000; Bagchi
71	et al. 2014; Fricke, Tewksbury, and Rogers 2014) and marine systems (Marhaver et al. 2013)
72	that microbial pathogens may accumulate near adults and suppress the survivorship of
73	conspecific juriles. In the most direct test of the Janzen-Connell hypothesis in corals,

74	Marhaver et al. (2013) used a series of lab and field investigations in the Caribbean to attribute
75	higher mortality of Orbicella (formerly Montastraea) faveolata recruits placed near adult
76	conspecifics to adult-associated microbial enemies. However, they found a complex relationship
77	between distance from adult colonies, current direction, and recruit morta. In less direct tests,
78	Vermeij (2005) and Vermeij and Sandin (2008) observed that survival of juvenile corals
79	decreased with increasing cover of conspecifics; they hypothesized that this was due to species-
80	specific microorganisms rather than to saturation of a limiting resource.
81	We experimentally evaluated distance-dependent mortality of juvenile-sized corals in the
82	field and correlatively surveyed multiple reefs for patterns of spatial distribution suggestive of
83	Janzen-Connell effects. We focused on two brooding coral species (Seriatopora hystrix and
84	Pocillopora damicornis) whose planulae recruit over short distances
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Methods

Study site characteristics

This study was conducted on reef flats within no-take marine protected areas (MPAs) adjacent to Votua, Vatuo-lailai, and Namada villages along the Coral Coast of Viti Levu, Fiji. These reserves are scattered along 11 km of fringing reef and are separated by ~3-8 km. The reserves have high coral cover (38-56%), low macroalgal cover (1-3%), and a high biomass and diversity of herbivorous fishes (Rasher, Hoey, and Hay 2013; Bonaldo and Hay 2014). The reef flats range from ~1-3 m deep at high extend ~500-600 m from shore to the reef crest, and are typical of exposed reef flats occurring throughout Fjii.

Except during low tides in calm weather, waves push water over the reef front, and water then flows laterally across the reef flats to discharge through channels bisecting the flats. This creates a relatively predictable current direction at most location.

Approval for our studies was granted by the Fijian Ministry of Education, National Heritage, Culture, and Arts, Youth, and Sports, and by the Korolevu-i-wai district environmental committee.

Survival experiments

To test whether juvenile corals experienced distance-dependent mortality near adult conspecifics, we collected ~5 mm tall fragments of *P. damicornis* and *S. hystrix*, selected suitable adult focal colonies (defined below), and attached conspecific fragments 3, 12, 24 and 182 cm up- and down-current from each focal lit. We deployed fragments around focal colonies in Votua village's MPA, which supports a diverse assemblage of corals covering about 50% of hard substrates (Rasher, Hoey, and Hay 2013). We used fragments from older colonies as proxies for ~6 month old juveniles (Sato 1985) because, despite these species reproducing monthly in some locations (Fan et al. 2002; Kuanui et al. 2008), neither species planulated at our site during the months of this study (August through October 2013).

We used pliers to clip 16 fragments of 30-40 polyps each from the tips of each of 24 large *P. damicornis* and 24 large *S. hystrix* colonies in the Votua village MPA. The fragments from each of four source colonies for a species were collected in six rounds over two days. Each round was taken to shore and four fragments (one from each source colony) were epoxied (Emerkit epoxy) onto the unglazed side of 16 2.54 x 2.54 cm tiles. Thus, each tile had fragments from four different colonies and sets of 16 tiles had fragments from the same for source colony. After epoxying, tiles were held in a tub of seawater for ~1 h, allowing the epoxy to harden. Tiles were then cable-

118	tied onto metal racks at ~1 m deep in the MPA and allowed to acclimate for two weeks before
119	deployment in the experiment. Survivorship during acclimation was 100%, producing 384
120	fragments on 96 tiles for each coral species.
121	Within the MPA, 10 adult P. damicornis and 10 adult S. hystrix colonies served as focal
122	colonies. Focal colonies: i) were >10 cm at their smallest diameter (10 to 35 cm for P .
123	damicornis and 10 to 75 cm for S. hystr ii) had no conspecific colonies within 4 m (so as not
124	to confound effects of the focal colony with effects of nearby conspecifics), iii) were 5-40 cm
125	deep at low tide, and iv) had space for 190 cm PVC pipes to be placed roughly east and west (the
126	predominant current direction was west) without disturbing other corals. Focal colonies were
127	photographed from above and their size determined using ImageJ (Rasband 1997).
128	Twenty mm diameter by 190 cm long PVC pipes served as platforms to which we
129	attached the tiles. Pipes were anchored to the reef by driving steel rebar through pre-drilled holes
130	and cementing the rebar to the pipe. Notches 2.54 cm long allowed us to cable-tie tiles onto the
131	pipes at distances of 3, 12, 24 and 182 cm from focal colonies. This approach secured all pipes
132	and tiles throughout the experiment. These distances and this scale were chosen to match a
133	previous experiment in the Caribbean that had detected distance dependent mortality of newly
134	settled recruits for a broadcast spawning coral (Marhaver et al. 2013).
135	Tiles were randomly assigned to positions on pipes. Thus, fragments at each distance and
136	around each conspecific focal colony were random with respect to source colony. Unassigned
137	tiles were kept on the rack as spares (64 fragments on 16 tiles for each coral species).
138	Every 1-2 d after deployment, we examined all fragments, recording survivorship,
139	consumption, overgrowth by algae, bleaching, or other changes in state

On some *P. damicornis* tiles, three or four of the fragments disappeared within a 24 h period between checks on their condition, appearing to have been bitten off. To determine the agents of this localized mortality, we replaced tiles whose four fragments had been eaten with spare tiles holding four healthy fragments around three of the focal colonies that had experienced localized mortality and videotaped the tiles (GoPro II HD) from about 1 m away during the following high tides. Cameras were retrieved at the next low tide and the videos watched.

We evaluated survival patterns using mixed-effects Cox proportional hazards survival models (coxme package, Therneau 2012) in R (R Core Team 2013). In ance and continuous focal colony were fixed effects and focal colony and tile nested within focal colony were random effects because fragments were blocked by tile and focal colony. The size of the focal colony and the depth of the tiles were included as random effects.

Distribution surveys

We characterized the spatial distribution of P. damicornis and S. hystrix in the reef flat MPAs of Namada, Vatuo-lailai, and Votua villages at two scales (August through October 2013). For our larger-scale survey, we mapped each colony within 8 x 8 m plots (N=5, 5, and 10 for Namada, Vatuo-lailai, and Votua, respectively). Each plot was divided into 256 0.5 x 0.5 m cells and each coral ≥ 1 cm across mapped into a cell. The location of each survey plot was determined by randomly choosing a point on shore, swimming 100, 200, or 300 kicks directly away from shore at that point, and surveying the closest bommie large enough to fill more than three quarters of an 8 x 8 m plot. In four of 10 surveys at Votua and in all five surveys at Vatuo-lailai and Namada, we also measured the largest diameter of each P. damicornis colony. We did not measure S. hystrix colony size because individual colonies were more frequently discontinuous.

To avoid confounding biotically-driven spatial distribution with patterns caused by patchiness of suitable substrate, we also recorded which cells were comprised primarily of unsuitable habitats such as sand-scoured pools or channels and bommie tops covered in rubble.

We analyzed these data using the neighborhood density function O(r) in the point pattern analysis program Programita (Wiegand and Moloney 2004). This analysis identifies distances at which individuals are aggregated, randomly spaced, or overdispersed compared to a specified null model. Unlike the more frequently used Ripley's K(r) statistic, each distance category is not affected by those inside it; expected aggregation at each distance is compared to the observed value independently of nearer distances. Each concentric ring centered on an individual coral is separately placed on the aggregated-overdispersed continuum and displays the spatial pattern within a different distance category. Ring width was 0.5 m extending up to 4 m. The null model for this analysis was complete spatial randomness (CSR). Because the variance in substrate types violated CSR's assumption of uniform likelihood of coral placement, we conducted the below analyses once using the entirety of all 8 x 8 m plots and a second time excluding cells of unsuitable habitat (which should better meet CSR's assumption of uniform likelihood).

To determine whether the observed spatial pattern was random, significantly aggregated, or overdispersed, Programita simulated placement of each plot's colonies 999 times using CSR, calculated O(r) for each simulation, then combined replicate O(r)'s from each reef and from all three reefs. This generated a distribution of simulated O(r)'s from which we established the significance of the observed spatial patterns. The distance(s) at which significant aggregation or overdispersion occurred were determined by the distances at which the observed pattern fell above or below the 95% simulation envelopes, respectively. This analysis does not parse aggregating and overdispering processes; it shows the net resulting pattern.

186	In addition to analyzing all <i>P. damicornis</i> and <i>S. hystrix</i> colonies, we analyzed <i>P</i> .
187	damicornis <5 cm, ≥5 cm, ≥10 cm, and ≥15 cm in diameter to see if spatial patterns changed
188	with colony size. The <5 cm and ≥5 cm categories were mutually exclusive but because there
189	were not enough colonies between 5 and 10 cm and between 10 and 15 cm to analyze as
190	mutually exclusive groups, larger size categories were subsets of smaller ones.
191	The 8 x 8 m quadrat surveys could not resolve spatial patterns below the cell size of 0.5 x
192	0.5 m , meaning that patterns occurring at less than 0.25^2 m could be undocumented. To
193	determine the spatial distribution of <i>P. damicornis</i> and <i>S. hystrix</i> at smaller scales, we conducted
194	2 m radius circular surveys around focal P. damicornis and S. hystrix colonies that i) were the
195	largest colony of that species within 4 m (to reduce the effects of conspecifics), and ii) occurred
196	where >75% of the substrate within 2 m was suitable habitat for <i>P. damicornis</i> and <i>S. hystrix</i> ,
197	again to equalize the likelihood of colonies occurring everywhere in the survey.
198	The distance to each surrounding (radial) P. damicornis and S. hystrix colony was the
199	average of the distance to that colony's near and far sides (N=45 focal colonies for <i>P. damicornis</i>
200	around P. damicornis, 10 for S. hystrix around P. damicornis, and 24 each for P. damicornis and
201	S. hystrix around S. hystrix). We analyzed radial colony counts in 10 cm concentric rings using a
202	generalized linear mixed effects model with Poisson errors and the canonical log link function in
203	R (lme4 package, Bates et al. 2013). Distance was ixed effect and focal colony with distance
204	nested inside was a random effect, with the log ₁₀ of the ring sizes as an offset to control for
205	unequal area sampled at each distance (i.e. ring area increased with distance from the focal
206	colony). We repeated this analysis with just the closest 0.5 m and 1 m of the circles in case radial
207	colonies beyond those distances were masking short-range effects of the focal colonies.

We also analyzed the P. damicornis data from the 8 x 8 m plots in the same manner as we did the circular surveys. To convert the plot data, an R script identified every surveyed P. damicornis colony ≥ 2 m from all edges of its plot and equal to or larger than a specified diameter (either 15 or 20 cm) as a focal colony (N=38 and 19 focal colonies, respectively). In order to have an appreciable sample size, we did not restrict focal colonies to those that were the largest within 4 m. The script then calculated the distances to all P. damicornis colonies less than the specified diameter within 2 m and placed them into 10 cm concentric rings as above. W ped generalized linear mixed effects models as described for the circular surveys.

Results

Survival experiments

In our field experiment, neither distance nor direction from focal colony significantly affected survival of *P. damicornis* or *S. hystrix* fragments (Fig. 1a & b, respection). We observed two main categories of mortality: bleaching preceding death in place (potentially due to microbes [e.g., Ben-Haim, Zicherman-Keren, and Rosenberg 2003]) and partial or complete disappearance, putatively due to predation (akin to provide the disappearance, putatively due to predation (akin to provide the disappearance). Bleaching (47 and 46 fragments out of 320 for *P. damicornis* and *S. hystrix*, respectively) of neither species was affected by distance or direction (Fig. 1c & d). Distance and direction did not affect the number of *P. damicornis* fragments that partially or fully disappeared (putative predation), and direction did not affect this for *S. hystrix* but distance was significant (z=2.23, p=0.03) (Fig. 1e & f), with disappearance increasing with distance from the focal colony. In contrast, 0% of the 64 fragments of each species on the coral rappleached or disappeared despite being on the same reef at the same time (Cox proportional hazards survival analysis, likelihood ratio for *P*.

damicornis=16.5, likelihood ratio for S. hystrix=24.7, p<0.0001 for both species). Fragments on
the coral rack were ~1 m above the benthos and may have experienced more flow or fewer
benthic associated biotic or physical stressors compared to the fragments on PVC pipes, which
were 5-15 cm above the benthos.

The rapid disappearance of *P. damicornis* fragments around some focal colonies suggested spatially localized predation. Therefore, we further divided deaths due to putative predation between isolated predation incidents (disappearance of one or two fragments on a tile in 24 h) and localized predation episodes (disappearance of three or four fragments from a tile in 24 h). We distinguished between these two types of putative predation because their causes were potentially different and therefore either one could have been distance-dependent or masked distance-dependence in the results of 10 *P. damicornis* replicates (23 out of 160 tiles) experienced localized predation on at least one of their eight tiles; three of those experienced localized predation on five or more tiles within 24 h. Two of 10 *S. hystrix* replicates experienced localized predation (on one tile each). We further investigated localized predation only for *P. damicornis* because localized predation on *S. hystrix* was infrequent.

When tiles that had experienced localized predation around three focal colonies were replaced with spare tiles holding healthy fragments, all three sets of replacement tiles again experienced localized predation and their collective survival was significantly lower than that of the replicates that did not experience localized predation in the initial run (mixed effect Cox proportional hazards, z=3.5, p<0.0005). Videos of these tiles showed the territorial triggerfish *Balis pus undulatus* consuming multiple fragments from multiple tiles around two of the three focal colonies. *Balistapus undulatus* feeding resulted in fragments irregularly broken at or above the top of the epoxy, as was seen for most localized predation episodes in the initial outplanting.

We next examined whether localized predation was distance-dependent and whether it masked distance-dependent mortality in replicates that did not experience localized predation. Distance and direction did not significantly affect mortality in replicates that did not experience localized predation (Fig 2a). Considering only replicates that experienced localized predation (both original and replacement tiles), neither distance nor direction significantly affected mortality from all causes (Fig. 2b) or just from localized predation (Fig. 2c).

Pocillopora damicornis fragments were significantly more likely to die of putative predation as opposed to bleach and die in place than were *S. hystrix* fragments (chi-square test, $\chi^2=17.2$, df=1, p<0.0001). More than three times as many *P. damicornis* fragments died from putative predation as bleached prior to death (169 vs. 47 out of 320, respectively), while numbers of *S. hystrix* fragments that died from putative predation versus bleaching did not differ significantly (58 vs. 46 out of 320, respectively). Excluding replicates with localized predation, *P. damicornis* and *S. hystrix* appeared equally susceptible to isolated predation and bleaching ($\chi^2=0.022$, df=1, p=0.88).

Distribution surveys

We analyzed patterns of distribution using both entire 8 x 8 m plots and after excluding habitat deemed unsuitable for *P. damicornis* or *S. hystrix* (e.g., sand-scoured channels and pools, bommie tops covered in rubble). The analyses using only suitable habitat were quantitatively similar to those using the entire plots but were more conservative. Neighborhood density graphs using only suitable habitat are included here; those using entire 8 x 8 m plots can be found in the appendix. Neighborhood density analysis indicated that both *P. damicornis* and *S. hystrix* were significantly aggregated at up to 1 m when all size classes were considered and surveys from all

277	villages were pooled (Fig. 3a & b, respectively). When analyzed by site, the distance below
278	which colonies were aggregated ranged from < 1 m in Votua and Vatuo-lailai to nearly 3 m in
279	Namada (see appendix). At no distance on any reef were colonies significantly overdispersed.
280	Identical analyses with <i>P. damicornis</i> separated into size categories (Fig. 3c-f) indicated
281	that the largest colonies (≥15 cm) were not aggregated at any scale, but all smaller size classes
282	were strongly aggregated at scales of up to 1 Thus, smaller colonies appeared to drive the
283	aggregation at up to \sim 1 m when we analyzed all sizes together. However, the sample size for
284	large colonies was limited (n=187), which may have constrained our ability to detect spatial
285	patterns for large colonies.
286	To resolve the spatial distribution of <i>P. damicornis</i> and <i>S. hystrix</i> more finely, we
287	conducted separate circular surveys (radius = 2 m) around focal colonies that met specific
288	criteria. Across all 2 m, there was a significant negative relationship between distance from focal
289	P. damicornis colonies and radial P. damicornis count (corrected for area surveyed at each
290	distance and henceforth called density focal P. damicornis and radial S. hystrix density, and
291	focal S. hystrix and radial P. damicornis density (GLM: z=-4.4, p<0.0001; z=-3.9, p<0.0005; z=-
292	3.6, p<0.0005, respectively) (Fig. 4a & b). The relationships within the first 0.5 m or 1 m for
293	these focal-radial combinations were not significant (see Table 1 for all values not provided in
294	text).
295	Across all 2 m, there was no significant relationship between distance from focal S.
296	hystrix colony and radial S. hystrix density (GLM, z=-1.9, p=0.06) (Fig. 4b). However, there was
297	a significant positive relationship between distance and density within the first 0.5 m (GLM, z=-
298	12.99, p<0.05) but not within the first 1 m.

When we converted the 8 x 8 m surveys into data analogous to the circular surveys and considered any P. damicornis colony ≥ 15 cm across as a focal colony and any smaller individual as a radial colony, there was a significant negative relationship between distance and radial P. damicornis density (GLM, z=-3.6, p<0.0005) across all 2 m but not across the first 0.5 m or 1 m (Fig. 4c; Table 2). However, when the cutoff for focal colonies was 20 cm, there was no relationship between distance and P. damicornis colony count at 0.5 m, 1 m, or 2 m (Fig. 4c; Table 2).

Discussion

Using small portions of adult *P. damicornis* and *S. hystrix* colonies to represent ~6 month old juveniles, we tested for distance-dependent survivorship as a function of proximity to adult conspecifics. Survival experiments with *P. damicornis* and *S. hystrix* fragments did not generally show distance-dependent mortality around conspecific adults (Fig. 1a & b). The lack of distance-dependent mortality in this study is consistent with a meta-analysis of distance-dependent mortality studies of the seeds and seedlings of terrestrial plants (Hyatt et al. 2003), in which distance from parents did not affect overall survival. However, when separated by life stage, that meta-analysis found that seedling survival increased with distance from parents while seed survival was not affected, suggesting that the strength of distance-dependent mortality may be a function of age. Our experiment using small coral fragments attempted to conduct this test with corals an older stage and thus our procedures would not have detected distance-dependent mortality of larvae occurring just after settlement. We would have preferred to conduct a reciprocal transplant experiment of larvae from both corals at differing distances to both conspecific and heterospecific adults but we were unable to gain permission to use that many

coral colo	nies. Thus,	we could	document	spatial patt	erns of surv	vivorship r	relative to	conspecific
adults but	not relative	e to hetero	specific ad	lults.				

Spatial analyses of the distribution of conspecific colonies might uncover patterns that
our short-term experiment could not detect. Observed spatial patterns represent the balance of
multiple, potentially opposing processes, such as greater recruit density near brooding parents
(similar to terrestrial seed shadows) versus detrimental effects of adult-associated enemies or
intraspecific competition on aggregated, nearby recruits. Rather than overdispersion, we found
significant clumping within 1 m of conspecifics for both P. damicornis and S. hystrix (Fig. 3a &
b). The 8 x 8 m surveys and the 2 m radius surveys both supported this pattern; there was a
significant negative relationship between P. damicornis radial colony density and distance from
focal P. damicornis and a nearly significant negative relationship (with a much more limited
sample size) between S. hystrix radial colony density and distance from focal S. hystrix (Fig. 4a
& b). We also observed a significant negative relationship between S. hystrix density and
distance from P. damicornis and P. damicornis density and distance from S. hystrix (Fig. 4a &
b), suggesting that the cause of declining density with distance need not be species-specific.
Since <i>P. damicornis</i> and <i>S. hystrix</i> are confamilial, it is possible that they aggregate because a
location that is physiologically beneficial for one might also be for the open. We did detect one
pattern consistent with the Janzen-Connell hypothesis: small colonies of <i>P. damicornis</i> were
aggregated at scales of up to 1 m, while colonies ≥ m in diameter were not aggregated at any
scale (Fig. 3). This selective loss of small colonies near adults is consistent with the Janzen-
Connell hypothesis, but is also consistent with self thinning from intraspecific competition
without mortality due to enemies aggregated near adults (Zhu et al. 2013).

There are a few potential causes for the observed clumping of conspecifics that could
counteract Janzen-Connell effects (Carlon and Olson 1993). Aggregated settlement near adults
may occur for P. damicornis and S. hystrix because both brood planulae that can settle quickly
after release (Richmond 1987; Isomura and Nishihira 2001; Underwood et al. 2007; Torda et al.
2013), and even if planulae disperse meters or kilometers, they may still aggregate near
conspecific adults (Babcock 1988; Tioho, Tokeshi, and Nojima 1). Moreover, pocilloporid
recruitment is inherently spatially heterogeneous (Dunstan and Johnson 1998) and occurs in
hotspots that may be partially determined by water flow, density of adult confamilials (Eagle
2006), and substrate suitability (Pri 010). Thus, multiple ecologically important processes and
interactions can generate aggregation of juveniles, and some of these could overwhelm Janzen-
Connell effects and make them seem unimportant in the field (at least in the short term), even if
they were occurring.
The only other direct test of the Janzen-Connell hypothesis in corals was conducted on
planulae and recruits of broadcasting Orbicella (formerly Montastraea) faveolata in the
Caribbean (Marhaver et al. 2013). In that study distance-dependent mortality appeared to be
microbially mediated, with effects differing upstream and downstream of focal O. faveolata. The
results of that study and ours differ in several ways.
First, Marhaver et al. (2013) used planulae and recruits a few days old in their distance-
dependent survival experiments, whereas we used fragments taken from mature colonies. The
physiology and microbiomes of fragments from adult corals may differ from those of recruits
and similarly sized juveniles (Harriott 1983; Christiansen et al. 2009). Most recruit mortality in
the study by Marhaver et al. (2013) appeared to be microbe-related as opposed to predator-
generated. In contrast, predators generated considerable mortality of our juvenile sized

transplants. Liberally assuming that every bleaching death in our study was due to microbes,
only about one quarter of P. damicornis and half of S. hystrix fragments could have died directly
because of microbes; thus, about 50-75% of the mortality we observed appeared to be due to
consumption by fish.

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Second, Marhaver et al. (2013) studied the broadcast spawning species O. faveolata, while we studied two brooding species whose planulae may be more likely to settle near their parents, and whose larvae may receive critical components of their microbiome via vertical transmission from adults. Data on the make-up and function of juvenile coral microbiomes are limited but at present there is some evidence that larvae from brooding species may be more consistently endowed with parental components of the microbiome than are the larvae of broadcast spawners (Littman, Willis, and Bourne 2009; Apprill et al. 2012; Sharp, Distel and Paul 2012; Lema, Willis, and Bourne 2014). In some acroporid corals, juveniles do not develop microbiomes typical of adult colonies until greater than 9 months of age (Littman, Willis, and Bourne 2009) but a core component of the microbiome appears in all the early stages, despite additions of other microbial species from the environment later in development (Lema, Willis, and Bourne 2014). However, in brooding species such as *Porites* and *Pocillopora*, critical microbes are transmitted from adults to larvae, or very quickly acquired from the environment, and even very young juvenile stages resemble adults in their composition of key microbes comprising the symbiotic microbiome (Apprill et al. 2012; Sharp, Distel, and Paul 2012). We do not know these relationships for the species we investigated but if their microbiomes take months to develop and are important defenses against microbial enemies, then our use of small adult portions may not mimic juvenile susceptibility to adult-associated pathogens. In contrast, if the critical components of the microbiome are present in even the earliest stages, then our adult

390	fragments should be more representative. We would have preferred to use recently recruited
391	larvae but neither P. damicornis nor S. hystrix planulated at our study site during our experiment.
392	Finally, the focal adult colonies of Orbicella investigated by Marhaver et al. (2013) form
393	larger, longer-lived colonies than the colonies of <i>Pocillopora</i> and <i>Seriatopora</i> that we
394	investigated. It is possible that larger, longer-lived colonies accumulate more species-specific
395	enemies over their lifetimes; if so, this could more strongly suppress juvenile survivorship near
396	these longer-lived adults.
397	Although we did not detect distance-dependent mortality, we did document spatially
398	heterogeneous corallivory on P. damicornis. This may promote species coexistence by producing
399	a mosaic of favorable and unfavorable patches for P. damicornis across the reef (Levin and Paine
400	1974; Holt 1984). Corallivore activity can structure coral distribution on reefs in both the Pacific
401	and Caribbean (Neudecker 1979; Littler, Taylor, and Littler 1989) and parrotfish and
402	butterflyfish density can impact coral recruit and juvenile mortality, respectively (Penin et al.
403	2010). Localized predation by the triggerfish <i>Balistapus undulatus</i> on small <i>P. damicornis</i> could
404	have a similar effect here. <i>Balistapus undulatus</i> is a generalist with territories of 100-200 m ²
405	(McClanahan 2000) and eats the tips of branching corals, including P. damicornis (Hiatt and
406	Strasburg 1960; Neudecker 1979). This triggerfish species' territoriality may delineate certain
407	patches on reefs in which some species (e.g., P. damicornis) have high mortality while other
408	species (e.g., S. hystrix) are not directly affected, akin to what is seen with seaweed in territories
409	of the steephead parrotfish on the Great Barrier Reef (Welsh and Bellwood 2012) or <i>Pocillopora</i>
410	and Pavona in the interaction between damselfish territories and roving corallivores in the
411	Eastern Pacific (Wellington 1982). Additional experiments are necessary to determine how

patchy corallivory contributes to the coexistence of P. damicornis, S. hystrix, and coral	s in
general.	

Overall we found little evidence for distance-dependent mortality relative to focal conspecific adults and for the pattern of over-dispersion that distance-dependent mortality would be expected to produce. Instead, both *P. damicornis* and *S. hystrix* aggregated at the scale of 1 m or less, with a tendency for small colonies to be clumped around larger ones. These findings suggest that local dispersal shadows or areas of physiological benefit near prospering adult conspecifics equal or exceed Janzen-Connell effects for the brooding corals we studied on these Fijian reef flats. Our experiments using small coral fragments did not detect distance-dependent mortality by species-specific enemies; we did, however, observe spatially heterogeneous corallivory on *P. damicornis*, which could facilitate species coexistence by delineating reef patches that are more or less favorable to different corals.

Acknowledgments

We thank the Fijian government and the Korolevu-i-wai district environmental committee for permitting this research, Victor Bonito for local support, and Kristen Marhaver and anonymous reviewers for helpful comments.

References

- Apprill A, Marlow HQ, Marntidale MQ, Rappe MS. 2012. Specificity of associates between bacteria and the coral *Pocillopora meandrina* during early development. *Applied Environmental Microbiology* 70: 7467-7475.
- Babcock R. 1988. Fine-scale spatial and temporal patterns in coral settlement. *Proceeding of the* 6th International Coral Reef Symposium 2:635-639.
- Bagchi R, Gallery RE, Gripenberg S, Gurr SJ, Narayan L, Addis CE, Freckleton RP, Lewis OT.
 2014. Pathogens and insect herbivores drive rainforest plant diversity and composition.
 Nature 506:85-88.
- Bates D, Maechler M, Bolker B, Walker S. 2013. lme4: Linear mixed-effects models using Eigen and S4. R package version 1.0-5. http://CRAN.R-project.org/package=lme4.
- Ben-Haim Y, Zicherman-Keren M, Rosenberg E. 2003. Temperature-regulated bleaching and lysis of the coral *Pocillopora damicornis* by the novel pathogen *Vibrio corallilyticus*. *Applied Environmental Microbiology* 69:4236-4242.
- Bernays, EA. 1989. Host range in phytophagous insects: the potential role of generalist predators. *Evolutionary Ecology* 3:299-311.
- Bonaldo RM, Hay ME. 2014. Seaweed-coral interactions: variance in seaweed allelopathy, coral susceptibility, and potential effects on coral resilience. *PLoS One* 9:e85786
- Buss LW, Jackson JBC. 1979. Competitive networks: nontransitive competitive relationships in cryptic coral reef environments. *American Naturalist* 113:223-234.
- Carlon DB, Olson RR. 1993. Larval dispersal distance as an explanation for adult spatial pattern in two Caribbean reef corals. *Journal of Experimental Marine Biology and Ecology* 173:247-263.
- Christiansen N, Ward S, Harii S, Tibbetts I. 2009. Grazing by a small fish affects the early stages of a post-settlement stony coral. *Coral Reefs* 28:47-51.
- Connell JH. 1971. On the role of natural enemies in preventing competitive exclusion in some marine animals and in rain forest trees. In: Den Boer PJ and Gradwell GR (eds):

 Dynamics of Population. Wageningen, Pudoc 298-312.
- Connell JH. 1978. Diversity in tropical rain forests and coral reefs. *Science* 199:1302-1310
- Cornell JH, Karlson R. 2000. Coral species richness: ecological versus biogeographical
 influences. *Coral Reefs* 19:37-49.
- Dixson DL, Abrego D, Hay ME. 2014. Chemically-mediated behavior of recruiting corals and fishes: a tipping point that may limit reef recovery. *Science* 345:892-897.
- Dunstan P, Johnson C. 1998. Spatio-temporal variation in coral recruitment at different scales on Heron Reef, southern Great Barrier Reef. *Coral Reefs* 17:71-81.
- Eagle JV. 2006. Recruitment hotspots around a coral reef: the roles of hydrodynamics and habitats. Ph.D theis, James Cook University
- Fan TY, Li JJ, Ie SX, Fang LS. 2002. Lunar periodicity of larval release by pocilloporid corals in southern Taiwan. *Zoological Studies* 41:288-294.
- Fricke EC, Tewksbury JJ, Rogers HS. 2014. Multiple natural enemies cause density-dependent mortality at the seed-to-seeding transition. *Ecology Letters* 17: 593-598.
- Harriott VJ. 1983. Reproductive seasonality, settlement, and post-settlement mortality of *Pocillopora damicornis* (Linnaeus), at Lizard Island, Great Barrier Reef. *Coral Reefs* 2:151-157.
- Hiatt RW, Strasburg DW. 1960. Ecological relationships of the fish fauna on coral reefs of the

- 476 Marshall Islands. *Ecological Monographs* 30:65-127.
- Holt RD. 1984. Spatial heterogeneity, indirect interactions, and the coexistence of prey species. *American Naturalist* 124:377-406.
- Hyatt LA, Rosenberg MS, Howard TG, Bole G, Fang W, Anastasia J, Brown K, Grella R,
 Hinman K, Kurdziel JP. 2003. The distance dependence prediction of the Janzen-Connell hypothesis: a meta-analysis. *Oikos* 103:590-602.
- Isomura N, Nishihira M. 2001. Size variation of planulae and its effect on the lifetime of planulae in three pocilloporid corals. *Coral Reefs* 20:309-315.
- Janzen DH. 1970. Herbivores and the number of tree species in tropical forests. *American Naturalist* 104:501-528.
- Johnson DJ, Beaulieu WT, Bever JD, Clay K. 2012. Conspecific negative density dependence and forest diversity. *Science* 336:904-907.
- Jones G, Almany G, Russ G, Sale P, Steneck R, Van Oppen M, Willis B. 2009. Larval retention and connectivity among populations of corals and reef fishes: history, advances and challenges. *Coral Reefs* 28:307-325.
- Kuanui P, Chavanich S, Raksasab C, Viyakarn V. 2008. Lunar periodicity of larval release and larval development of *Pocillopora damicornis* in Thailand. Mar Freshw Res 11:375-377.
- Lang J. 1973. Interspecific aggression by scleractinian corals. 2. Why the race is not only to the swift. *Bulletin of Marine Science* 23:260-279.
- Lema KA, Willis BL, Bourne DG. 2012. Corals form characteristic associations with symbiotic nitrogen-fixing bacteria. *Applied Environmental Microbiology* 78:3136-3144.
- Levin SA, Paine RT. 1974. Disturbance, patch formation, and community structure. *Proceeding* of the National Academy of Sciences 71:2744-2747.
- Littler MM, Taylor PR, Littler DS. 1989. Complex interactions in the control of coral zonation on a Caribbean reef flat. *Oecologia* 80:331-340.
- Littman RA, Willis BL, Bourne DG. 2009. Bacterial communities of juvenile corals infected with different Symbiodinium (dinoflagellate) clades. *Marine Ecology Progress Series* 389:45-59.
- 504 Marhaver K, Vermeij M, Rohwer F, Sandin S. 2013. Janzen-Connell effects in a broadcast-505 spawning Caribbean coral: distance-dependent survival of larvae and settlers. *Ecology* 506 94:146-160.
- McClanahan T. 2000. Recovery of a coral reef keystone predator, *Balistapus undulatus*, in East African marine parks. *Biological Conservation* 94:191-198.
- Miller K, Mundy C. 2003. Rapid settlement in broadcast spawning corals: implications for larval dispersal. *Coral Reefs* 22:99-106.
- Nathan N, Casagrandi R. 2004. A simple mechanistic model of seed dispersal, predation and plant establishment: Janzen-Connell and beyond. *Journal of Ecology* 92: 733-746.
- Neudecker S. 1979. Effects of grazing and browsing fishes on the zonation of corals in Guam. *Ecology* 60:666-672.
- Nozawa Y, Harrison PL. 2008. Temporal patterns of larval settlement and survivorship of two broadcast-spawning acroporid corals. *Marine Biology* 155:347-351.
- Packer A, Clay K. 2000. Soil pathogens and spatial patterns of seedling mortality in a temperate tree. *Nature* 404:278-281.
- Penin L, Michonneau F, Carroll A, Adjeroud M. 2011. Effects of predators and grazers exclusion on early post-settlement coral mortality. *Hydrobiologia* 663:259-264.
- Penin L, Michonneau F, Baird AH, Connolly SR, Pratchett MS, Kayal M, Adjeroud M. 2010.

- Early post-settlement mortality and the structure of coral assemblages. *Marine Ecology Progress Series* 408:55-64.
- Petermann JS, Fergus AJ, Turnbull LA, Schmid B. 2008. Janzen-Connell effects are widespread and strong enough to maintain diversity in grasslands. *Ecology* 89:2399-2406.
- Porter JW, Woodley JD, Jason Smith G, Neigel JE, Battey JF, Dallmeyer DG. 1981. Population trends among Jamaican reef corals. *Nature* 294:249-250.
- Pratchett MS. 2007. Dietary selection by coral-feeding butterflyfishes (Chaetodontidae) on the Great Barrier Reef, Australia. *Rafflesia Bulletin of Zoology* 14:155-160.
- Price N. 2010. Habitat selection, facilitation, and biotic settlement cues affect distribution and performance of coral recruits in French Polynesia. *Oecologia* 163:747-758.
- R Core Team. 2014). R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL http://www.R-project.org/
- Rasband WS. 1997. ImageJ, US National Institutes of Health, Bethesda, Maryland, USA
- Rasher DB, Hoey AS, Hay ME. 2013. Consumer diversity interacts with prey defenses to drive ecosystem function. *Ecology* 94:1347-1358.
- Richmond R. 1987. Energetics, competency, and long-distance dispersal of planula larvae of the coral *Pocillopora damicornis*. *Marine Biology* 93:527-533.
- Roff G, Mumby PJ. 2012. Global disparity in the resilience of coral reefs. *Trends in Ecology and Evolution* 27:404-413.
- Rotjan RD, Lewis SM. 2008. Impact of coral predators on tropical reefs. *Marine Ecology Progress Series* 367:73-91.
- Sato M. 1985. Mortality and growth of juvenile coral *Pocillopora damicornis* (Linnaeus). *Coral Reefs* 4:27-33.
- Sharp KH, Distel D, Paul VJ. 2012. Diversity and dynamics of bacterial communities in early life stages of the Caribbean coral *Porites astreoides*. *ISME Journal* 6:790–801.
- Smith JE, Shaw M, Edwards RA, Obura D, Pantos O, Sala E, Sandin SA, Smriga S, Hatay M,
 Rohwer FL. 2006. Indirect effects of algae on coral: algae-mediated, microbe-induced
 coral mortality. *Ecology Letters* 9:835-845.
- Therneau T. 2012). coxme: Mixed Effects Cox Models.. R package version 2.2-3. http://CRAN.R-project.org/package=coxme
- Tioho H, Tokeshi M, Nojima S. 2001. Experimental analysis of recruitment in a scleractinian coral at high latitude. *Marine Ecology Progress Series* 213:79-86.
- Torda G, Lundgren P, Willis B, Oppen M. 2013. Revisiting the connectivity puzzle of the common coral *Pocillopora damicornis*. *Molecular Ecology* 22:5805-5820.
- Underwood J, Smith L, Van Oppen M, Gilmour J. 2007. Multiple scales of genetic connectivity in a brooding coral on isolated reefs following catastrophic bleaching. *Molecular Ecology* 16:771-784.
- Vermeij MJ. 2005. Substrate composition and adult distribution determine recruitment patterns in a Caribbean brooding coral. *Marine Ecology Progress Series* 295:123-133.
- Vermeij MJ, Sandin SA. 2008. Density-dependent settlement and mortality structure the earliest life phases of a coral population. *Ecology* 89:1994-2004.
- Wellington GM. 1982. Depth zonation of corals in the Gulf of Panama: control and facilitation by resident reef fishes. *Ecological Monographs* 52:223-241.
- Welsh J, Bellwood D. 2012. Spatial ecology of the steephead parrotfish (*Chlorurus microrhinos*): an evaluation using acoustic telemetry. Coral Reefs 31:55-65.
- Wiegand T, Moloney AK. 2004. Rings, circles, and null-models for point pattern analysis in

568	ecology. Oikos 104:209-229.
569	Zhu Y, Getzin S, Wiegand T, Ren H, Ma K. 2013. The relative importance of Janzen-Connell
570	effects in influencing the spatial patterns at the Gutianshan subtropical forest. PLoS One
571	8:e74560
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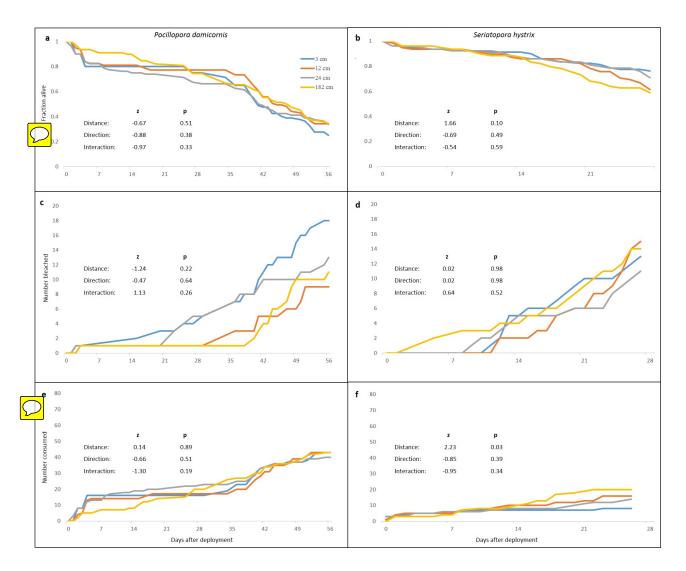


Figure 1- Left column is *Pocillopora damicornis* and right column is *Seriatopora hystrix*.

Statistical values are from mixed-effects Cox proportional hazards survival analyses.

n=80 fragments at each distance across 10 focal colonies and pooled between both directions. a & b) Survivorship through time for *Pocillopora damicornis* and *Seriatopora hystrix* fragments. c & d) Cumulative number of fragments that bleached over time. e & f) Cumulative number of fragments that partially or fully disappeared over time (putative predation).

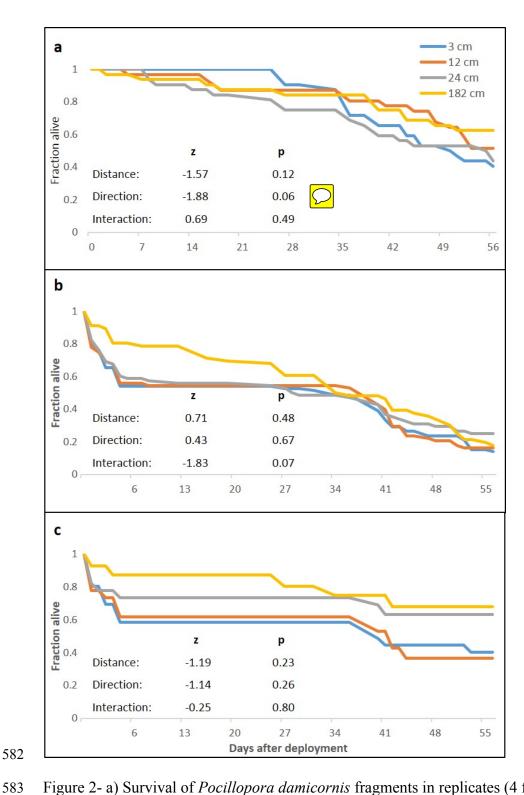


Figure 2- a) Survival of *Pocillopora damicornis* fragments in replicates (4 focal colonies, 32 fragments at each distance) that did not experience localized predation. b) Survival of *P. damicornis* fragments in the six focal colony replicates that did experience localized

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586	predation and in the replacement replicates. Deaths are from all causes. c) Fraction of <i>P</i> .
587	damicornis fragments not killed by localized predation episodes in original replicates that
588	experienced localized predation and in the replacement replicates. Direction not shown.
589	Analyses as in Fig. 1.
590	

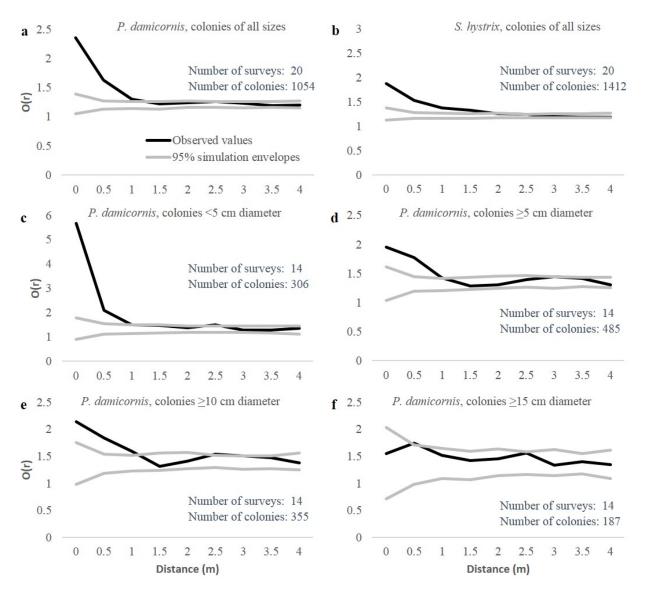
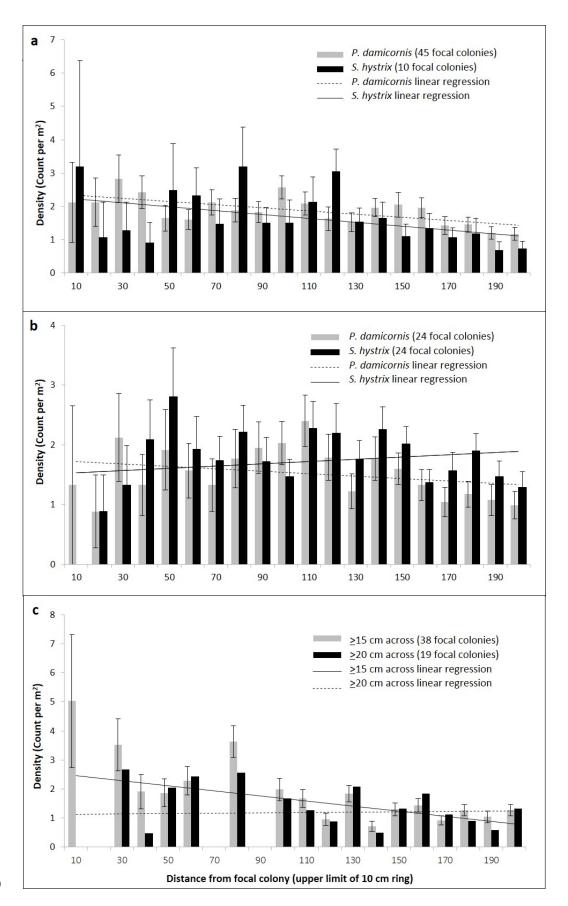


Figure 3- Neighborhood density analysis of *Pocillopora damicornis* and *Seriatopora hystrix* in 8 x 8 m quadrats with replicates from all three reefs combined. Black lines are observed patterns; grey lines are the 95% simulation envelopes from 999 simulations. Where black lines are above the upper grey line colonies are significantly aggregated, where they are between the grey lines colonies are randomly spaced, and where they are below the lower grey line colonies are significantly overdispersed. These analyses used only areas of suitable substrate (see text for definition).



601	Figure 4- Density (± SE) of <i>Pocillopora damicornis</i> and <i>Seriatopora hystrix</i> at 10 cm intervals
602	from focal a) P. damicornis and b) S. hystrix colonies. The linear regressions shown are
603	to indicate the slope of the relationship found in the generalized linear mixed effects
604	models but do not represent the models' outputs. Radial colony count significantly
605	declined with distance from focal colony over 2 m for three of the four focal-radial
606	combinations (focal <i>P. damicornis</i> -radial <i>P. damicornis</i> — z=-4.4, p<0.001; focal <i>P.</i>
607	damicornis-radial S. hystrix— z=-3.9, p<0.001; focal S. hystrix-radial P. damicornis—
608	z=-3.6, p<0.001; focal <i>S. hystrix</i> -radial <i>S. hystrix</i> — z=-1.9, p=0.06). c) Density (mean ±
609	SE) of <i>Pocillopora damicornis</i> within 2 m of focal <i>P. damicornis</i> based on the 8 x 8 m
610	surveys. Focal colonies are <i>P. damicornis</i> that are \ge 15 cm across or \ge 20 cm across.
611	Radial colonies are any colonies below that size. Radial colony count significantly
612	declined with distance over 2 m when colonies \geq 15 cm were considered focal (focal
613	colonies \ge 15 cm— z=-3.6, p<0.0005; focal colonies \ge 20 cm— z=-1.09, p<0.27).
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Focal species- radial species	Maximum distance	Slope	z value	p-value
P. damicornis-	0.50 m	-0.0098	-1.0	0.32
	1.0 m	0.00055	0.22	0.83
P. damicornis	2.0 m	-0.0032	-4.4	<0.0001
P. damicornis-	0.50 m	0.016	0.59	0.56
	1.0 m	0.0012	0.19	0.85
S. hystrix	2.0 m	-0.0057	-3.9	<0.0005
C hystriv	0.50 m	0.0099	0.57	0.57
S. hystrix-	1.0 m	0.0042	1.1	0.28
P. damicornis	2.0 m	-0.0036	-3.6	<0.0005
C hystriv	0.50 m	0.042	2.3	<0.05
S. hystrix-	1.0 m	-0.00065	-0.18	0.86
S. hystrix	2.0 m	-0.0017	-1.9	0.06

Table 1- Relationship between count of radial *Pocillopora damicornis* and *Seriatopora hystrix* colonies and distance from focal *P. damicornis* and *S. hystrix* colonies using generalized linear mixed effects models. "Maximum distance" is the distance up to which radial colonies were considered.

Threshold size	Maximum	Slope	z value	p-value
for focal colony	distance			
	0.50 m	-0.0064	-0.46	0.65
15 cm	1.0 m	-0.0063	-1.5	0.14
	2.0 m	-0.0034	-3.6	<0.0005
	0.50 m	0.047	1.2	0.23
20 cm	1.0 m	0.00032	0.045	0.96
	2.0 m	-0.0016	-1.09	0.27

Table 2- Relationship between count of radial *Pocillopora damicornis* colonies and distance from focal *Pocillopora damicornis* colonies using the data from the 8 x 8 m surveys. "Threshold size for focal colony" is the size above which surveyed colonies were designated "focal" and below which colonies were designated "radial." "Maximum distance" is the distance up to which radial colonies were considered.