

Litter inputs and standing stocks in riparian zones and streams under secondary forest and managed and abandoned cocoa agroforestry systems

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Background. Cocoa is an important tropical tree crop that is mainly cultivated in agroforestry systems (AFS). This system, known as cabruca in northeastern Brazil-, holds promise to reconcile biodiversity conservation and economic development. However, since cocoa AFS alters forest structure composition, it can affect litter dynamics in riparian zones and streams. Thus, our objective was to determine litter inputs and standing stocks in riparian zones and streams under three types of forest: managed cocoa AFS, abandoned cocoa AFS, and secondary forest. **Methods.** We determined terrestrial litter fall (TI), vertical (VI) and lateral (LI) litter inputs to streams, and litter standing stocks on streambeds (BS) in the Atlantic Forest of northeastern Brazil. Litter was collected every 30 days from August 2018 to July 2019 using custom-made traps. The litter was dried, separated into four fractions (leaves, branches, reproductive organs, and miscellaneous material) and weighed. Results. Terrestrial litter fall was similar in all forests, ranging from 57 g m⁻² in secondary forest (SF) to 86 g m⁻² in abandoned cocoa AFS (AC). Vertical input was higher in AC (58 g m⁻²) and MC (39 g m⁻²) than in SF (34g m⁻²), whereas lateral input was higher in MC (57 g m⁻²) than in AC (18 g m⁻²) and SF (24 g m⁻²). Standing stocks followed the order SF>AC>MC, corresponding to 357, 251 and 84 g m⁻². Leaves contributed most to all litter fractions in all forests. Reproductive plant parts accounted for a larger proportion in managed AFS. Branches and miscellaneous litter were also similar in all forests, except for higher benthic standing stocks of miscellaneous litter in the SF. Despite differences in the amounts of litter inputs and standing stocks among the forests,

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seasonal patterns in the abandoned AFS (AC) were more similar to those of the secondary forest (SF) than the managed AFS, suggesting potential of abandoned AFS to restore litter dynamics resembling those of secondary forests.



2	forest and managed and abandoned cocoa agroforestry systems
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Litter inputs and standing stocks in riparian zones and streams under secondary



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27	systems (AFS). This system, known as cabruca in northeastern Brazil, holds promise to
28	reconcile biodiversity conservation and economic development. However, since cocoa AFS
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40	MC (39 g m^{-2}) than in SF (34g m^{-2}), whereas lateral input was higher in MC (57 g m^{-2}) than in
41	AC (18 g m ⁻²) and SF (24 g m ⁻²). Standing stocks followed the order SF>AC>MC,
42	corresponding to 357, 251 and 84 g m ⁻² . Leaves contributed most to all litter fractions in all
43	forests. Reproductive plant parts accounted for a larger proportion in managed AFS. Branches
44	and miscellaneous litter were also similar in all forests, except for higher benthic standing stocks
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46	stocks among the forests, seasonal patterns in the abandoned AFS (AC) were more similar to

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- 47 those of the secondary forest (SF) than the managed AFS, suggesting potential of abandoned
- 48 AFS to restore litter dynamics resembling those of secondary forests.



Introduction

50	Riparian zones are important for the functioning of headwater streams (Vannote et al., 1980;
51	Naiman, Décamps & McClain, 2005), including in tropical zones (Gonçalves Júnior et al., 2014;
52	Bambi et al., 2016; Rezende et al., 2017a; Rezende et al., 2019; Calderón et al., 2019). The
53	riparian canopy limits instream primary production and provides allochthonous organic matter to
54	stream and riparian food webs in the form of litter, which increases heterotrophic metabolism
55	(Gonçalves Júnior et al., 2014; Rezende et al., 2019). Therefore, litter dynamics are a
56	fundamental characteristic of headwater streams (Abelho & Graça, 1996; Neres-Lima et al.,
57	2017). In the tropics, litter is typically supplied throughout the year (Tonin et al., 2017), although
58	this pattern varies among forest types (Lindman et al., 2017; Seena et al., 2017), largely driven
59	by precipitation and temperature regimes (Bambi et al., 2016; Tonin et al., 2017). Changes in the
60	structure and composition of riparian forests can affect the supply of litter to streams and their
61	riparian zones (Delong & Brusven, 1994; Ferreira et al., 2019; Wild, Gücker & Brauns, 2019), as
62	well as in-stream litter dynamics (Stufin, Wohl & Dwire, 2016; Tiegs et al., 2019).
63	Many tropical forests are jeopardized by rapid deforestation and expansion of agriculture (Bawa
64	et al., 2004). This includes the Atlantic Forest of Brazil as one of the most threatened tropical
65	forests worldwide (Winbourne et al., 2018; Taubert et al., 2018). Agroforestry systems (AFSs),
66	however, have potential to partly reconcile the conservation of tropical forest patches with
67	economic development (Cassano et al., 2009; Schroth et al., 2011). One example is the
68	cultivation of cocoa in the Atlantic Forest of northeast Brazil where cocoa trees (Theobroma
69	cacao L.) are grown in AFS that cover a large portion of the remnant Atlantic Forest (Piasentin
70	et al., 2014). The cocoa trees are planted in the shade of native forest trees (dominant and
71	codominant strata) and are surrounded by natural vegetation. Therefore, cocoa AFS are thought



72	to cause less environmental impact than other crop systems (Johns, 1999; Sambuichi, 2002), with
73	benefits for local biodiversity (Faria et al., 2007; Cassano et al., 2009; Schroth et al., 2011).
74	Important processes such as the incorporation of large amounts of organic matter into the forest
75	soil are indeed maintained in cocoa AFS (Beer et al., 1998; Gama-Rodrigues et al., 2010; Barreto
76	et al., 2011; Fontes et al., 2014; Costa et al., 2018). Nevertheless, changes in the vegetation
77	structure of AFS compared to unmanaged forest may affect the amount of litter deposited in
78	riparian zones (Delong & Brusven, 1994; Wild, Gücker & Brauns, 2019) and supplied to streams
79	(Gonçalves Júnior et al., 2014).
80	Studies on litter dynamics in tropical streams and their riparian zones are scarce, especially under
81	cocoa AFS, although some evidence suggests that replacing cocoa AFS changes the cycling of
82	carbon and nitrogen in streams (Costa et al., 2017; Souza et al., 2017; Costa et al., 2018),
83	possibly as a result of altered litter supply by riparian vegetation. Thus, the current study aimed
84	to assess the influence of cocoa AFS on litter dynamics by determining differences in secondary
85	forest and managed and abandoned AFS on litter inputs and benthic standing stocks in streams
86	and riparian zones in these forests. We expected that i) managed and abandoned cocoa AFS
87	produce more litter than secondary forest where forest structure (Curvelo et al., 2009; Dawoe,
88	Isaac & Quashie-Sam, 2010; Fontes et al., 2014) and soil carbon stocks differ (Gama-Rodrigues
89	et al. 2010, Costa et al. 2018) and nutrients are rapid cycled (Nair et al. 1999); ii) streams
90	running through forests with high litter production tend to receive larger amounts of litter
91	(França et al., 2009; Gonçalves Júnior et al., 2014), resulting in greater litter standings stocks in
92	the streambeds (Webster et al., 1994; Lisboa et al., 2015); and iii) seasonal patterns of litter
93	inputs and standing stocks reflect precipitation patterns because water availability controls litter
94	production (Tonin et al., 2017).

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dy area The study was conducted in the riparian zones of three small watersheds (Fig. 1) representing 98 secondary forest (E 485415, N 8397615), abandoned cocoa AFS (E 481551, N 8364478), and 99 100 managed cocoa AFS (E 448466, N 8363187). All sites are located in the Atlantic Forest of southern Bahia in northeast Brazil. The Climate is wet tropical (hot and humid with no defined 101 102 dry season, Af according to the Köppen classification) with annual rainfall ranging from 1100 to 103 2200 mm. The study streams are second-order according to the Strahler classification. Daily 104 rainfall data were obtained from the website of the Real-Time Climate Monitoring Program of the Northeast Region (PROCLIMA; http://proclima.cptec.inpe.br) for the municipalities of 105 106 Itacaré, Ilhéus, and Barro Preto (Fig. 2). 107 The secondary forest, which covers 9,275 ha, is located in a conservation area (Serra do Conduru 108 State Park - License 2017-013654/TEC/PESQ-0014) (Martini et al., 2007). The vegetation is a 109 mosaic of different developmental stages, including secondary forest and remnants of mature 110 forests with different degrees of selective logging in the past (Winbourne et al., 2018). The 111 uniform canopy of the forest exceeds 25 m in height and includes a few emerging individual 112 trees, epiphytes, large lianas, and a dense understory (Martini et al., 2007; Costa et al., 2018). 113 Tree species density levels in the area were high at all sites, independent of forest successional 114 stage; old growth forest totaled 144 species, old logged forest had 137 species, and recently 115 logged forest 134. Of the species recorded in the Serra do Conduru State Park, 51.4% are 116 endemic to the Atlantic Forest and 26% occur only in the south of Bahia (Martini et al., 2007). 117 The abandoned AFS covers 73.4 ha and is located in an AFS (Santa Cruz) where crop



118	management was abandoned 20 years before the present study. Old cocoa trees and other,
119	irregularly distributed species such as jackfruit, erythrina, embaúba, and jequitibá trees (Argôlo,
120	2009) resulted in a medium level of shading (70%). The managed AFS is located in another AFS
121	(Nova Harmonia) with a total area of 89.8 ha. It comprises areas under cocoa production, a forest
122	patch in the central portion, and two areas undergoing regeneration (Santos et al., 2016).
123	Management consists of pruning cocoa trees every six months, with the biomass left in place
124	(Costa et al., 2018), complemented by vegetation cutting, and some liming for soil amelioration.
125	The cocoa plants were spaced at 3x3m and intercropped with introduced shade trees (erythrina),
126	according to the proposed management for the area.
127	Litter inputs and benthic standing stocks
128	Litter inputs and benthic standing stocks were determined from August 2018 to July 2019.
129	Details of the methodology are described in Gonçalves Júnior et al. (2014) and Bambi et al.
130	(2016). Terrestrial litter fall (TI), vertical (VI) and lateral (LI) litter inputs to streams, and litter
131	deposited on the streambeds (benthic standing stock – BS) were assessed along 100 m stream
132	stretches at each location (Figure 3).
133	TI deposited on the riparian soil represents the amount of litter that can potentially be transported
134	to the stream. It was collected with 10 nets (1 mm mesh, 2.5 m² total area), five on both sides of
135	the streams, installed 1 m above the ground at 20 m distance from one another in the riparian
136	zone. VI represents litter that falls directly into the streams from the riparian canopy. It was
137	collected with 27 buckets (30 cm diameter, 1.9 m² total area) fixed to trees, perpendicular to the
138	stream channel at a height of approximately 2 m. The buckets were arranged in three groups of
139	nine, spaced approximately 30 m apart with a distance of 1 m between the individual buckets.
140	Small holes in the bottom of the buckets allowed any collected water to drain. LI represents the





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indirect input of litter by lateral movement from the forest floor to the stream due to gravity, runoff, wind, or animal action. LI was collected with 10 nets (1 mm mesh, 0.5 m length, 1.5 m² total area) arranged at ground level at the stream margins, five on both sides of the streams. Total litter input to the streams was calculated as the sum of lateral and vertical inputs. Finally, benthic standing stocks represent the litter accumulated on the streambed. It was estimated monthly, by taking Surber samples (0.25 mm mesh, 0.45 m² total area), five in each stream at 20 m distance from one another (Fig. 3). The litter trapped in the nets and buckets was collected at monthly intervals and sorted into four fractions upon return to the laboratory: leaves, branches (i.e. woody pieces less than 25 cm in length), reproductive organs such as flowers and fruits, and miscellaneous material (i.e. unidentified plant matter and animal remains). The sorted litter was dried in an oven at 60 °C for 72 hours and weighed. TI, VI and LI were expressed in g dry mass m⁻² d⁻¹ and BS in g dry mass m⁻² LI per m² was calculated by dividing the collected litter mass by the trap width and multiplying the result by two (to account for inputs from both stream banks) and by the mean channel width (Pozo et al., 2009). The annual litter inputs to the streams and riparian zones corresponds to the sum of the mean monthly litter inputs during the study year (Table 1). Statistical analysis Differences among the forests in litter inputs and benthic standing stocks were assessed by generalized linear mixed-effects models (glmer function in the lme4 package of R) with forest type (= site), time and the interaction of forest type and time as predictive variables (Bates et al., 2015). We considered trap and time as random factors to account for the pseudoreplicated design of the study, since the three forest types were represented only by a single site each. Separate





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models were run for each of the plant organic matter fractions (leaves, branches, reproductive organs, miscellaneous material). P-values were obtained by likelihood ratio tests (chi-square distribution) of the full model against a partial model without the explanatory variable. All models were tested for error distribution by using the hnp package and function in R, and corrected for over- or underdispersion. Differences in litter inputs and standing stocks among forest types (sites) were also assessed by using bootstrapped 95% confidence intervals, which were computed by the bias-corrected and accelerated (BCa) method using the boot package and function in R, based on 1,000 bootstrap replicates (Davison & Hinkley, 1997; Canty & Ripley, 2016). Differences were considered statistically significant when the bootstrapped confidence intervals did not overlap. Given their flexibility, generalized additive mixed models (GAMM) were used as an additional approach to explore the seasonal patterns of litter inputs (i.e. vertical, lateral and terrestrial inputs) and standing stocks (Tonin et al., 2017; 2019). The input of leaves, branches, reproductive matter, or miscellaneous material over the 12-months study period was used as a normally distributed (identity-link function) predictor, nested in-within sites, as a random component of the GAMM models. Trap and time were used as random factors in the GAMM models. The degree of smoothing in an additive model is expressed as effective degrees of freedom (edf). Higher edf values indicate a lower degree of linearity (i.e. here variation over time), with a value of 1 indicating a perfectly linear effect. The additive mixed models were fitted by using the by command in the mgcv package in R. Validation was used to estimate the optimal degree of smoothing (Wood, 2017). The residual spread within models among sampling dates was measured by using the *varIdent* function in *R*.

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Results

188	Leaf material was the single largest litter fraction, with percentages >60% for TI and VI and
189	>70% for LI in the secondary forest and abandoned AFS (Fig. 4, Table 1). In the managed AFS,
190	the percentages of TI, VI, and LI were 56, 41, and 62%, respectively. Leaves also represented
191	large portions of the instream standing stock of litter, 61% in the abandoned AFS, 57% in the
192	managed AFS, and 38% in the secondary forest. Miscellaneous types of organic matter was the
193	second most abundant litter fraction of TI and VI observed in secondary forest and abandoned
194	AFS (23% and 16% for TI, 23% and 15% for VI). Miscellaneous litter accounted for 43% of the
195	standing stock in the secondary forest (Table 1). Branches were also a large portion of the litter
196	standing stock in secondary forest (19%) and abandoned AFS (16%) (Table 1). Because of their
197	sporadic and transient occurrence, reproductive parts were a low proportion of the total litter,
198	except in managed AFS (Table 1), where they accounted for a higher proportion in all types of
199	litter inputs and standing stocks (Fig. 4-c,h,m,r). Branches and miscellaneous litter were
200	generally similar in all forests, except for higher benthic standing stocks of miscellaneous litter
201	in the secondary forest (Fig. 4).
202	The greatest seasonal variation in litter inputs and standing stocks was found for leaves, with the
203	highest contributions during the transition between dry and rainy months and two seasonal peaks
204	in some cases (Fig. 5a,c). Seasonal variation was less pronounced and generally not significant
205	for branches, reproductive plant parts, and miscellaneous litter (Figs S1 to S3). Vertical litter
206	inputs showed seasonal patterns for leaves in the secondary forest (effective degrees of freedom
207	- edf = 5.5; Fig. 5d and abandoned AFS (edf = 5.1; Fig. 5e), with higher contributions during the
208	rainy months, from November to February, in both SF and AC. In managed AFS, the
209	contribution of leaves decreased linearly over time, as reflected by an edf value close to 1 (Fig.





5f). Terrestrial leaf inputs showed a sinusoidal pattern is the secondary forest (Fig. 5a) and
managed AFS (Fig. 5c), which was reflected by high edf values of 7.0 and 6.7, respectively. A
peak in the rainiest months was observed in all three forests but the second peak was missing in
the abandoned AFS (Fig. 5b). The largest lateral inputs of leaves were observed from January to
March in managed AFS (edf = 1.9; Fig. 5f) the period of least rainfall (Fig. 2). Standing stocks
of leaf litter showed different trends than leaf litter inputs. The sine wave shifted to the right in
SF, indicating that leaf input occurred just after the rainiest periods (peak in April; edf = 3.5; Fig
5j) and abandoned AFS (minimum in November and maximum in February; edf = 5.6; Fig. 5k).
No seasonal patterns were observed for leaf litter of managed AFS (edf = 1; Fig. 5f,i,l).
Total annual litter fall in the riparian zone of the abandoned AFS, managed AFS, and secondary
forest was 181, 122, and 118 g dry mass m ⁻² , respectively. In the abandoned and managed AFS,
56% of the terrestrial litter input (TI) was deposited on the forest floor and 44% directly entered
the streams through vertical litter input (VI). In the secondary forest, the percentages of TI and
VI were 63% and 37%, respectively. In the managed AFS, 66% of the total litter fall in the
riparian zone entered the stream by lateral movement and 34% remained in the riparian zone. In
the abandoned AFS, only 16% entered the stream and 84% remained in the riparian zone. The
corresponding percentages in the secondary forest were 37 and 63%. The average annual total
litter standing stock in the managed AFS was more than two times lower than in the abandoned
AFS and more than three times lower than in the secondary forest (Fig. 6).
Discussion

Riparian zones with natural vegetation are important for the structural and functional integrity of streams (Gonçalves Junior et al., 2014, Rezende et al., 2017b). Our results on the contribution of



233	leaf litter in forests subject to different management practices are important information to assess
234	the role of cocoa agroforestry in efforts to restore impacted forests in northeastern Brazil.
235	Previous studies in the area have reported that the cocoa agroforestry system alters the
236	biogeochemistry of C and N in streams and soils (Costa et al., 2017, 2018, Souza et al., 2017).
237	However, these studies could not determine which forest attributes determined the observed
238	changes in C and N cycling. Although differences between forests could be associated with
239	different management regimes, it is necessary also to consider the potential influence of other
240	factors that could not be evaluated in those study.
241	High litter production in the abandoned AFS may be due to the riparian vegetation structure
242	(Gonçalves Junior et al., 2014; Rezende et al., 2017a) and successional stage during forest
243	recovery (Sambuichi & Haridasan, 2007; Rolim et al., 2017). Factors such as abundant deposits
244	of crop biomass (Beer et al., 1998), which are related to high carbon stocks in the soil (Gama-
245	Rodrigues et al., 2010; Costa et al., 2018), and rapid nutrient cycling in these systems (Nair et al.,
246	1999) are likely to play a role. The pattern of litter inputs in the abandoned AFS was more
247	similar to that in the secondary forest than the managed AFS. Streams in abandoned AFS are
248	usually lined by riparian vegetation (Ferreira et al., 2019) similar to that of streams in secondary
249	forest, whereas in managed AFS, native shade trees are replaced by species with high
250	commercial value (Cassano et al., 2009, Piasentin Saito & Sambuichi, 2014). Additional factors
251	linked to higher litter production in abandoned AFS (Gama-Rodrigues et al., 2010) include
252	greater stand age of the vegetation. Moreover, leaf surface area may be greater in the shaded
253	stands where low solar radiation limits rates of photosynthesis (Beer et al., 1998). These factors
254	tend to increase litter production, to which leaves contribute the largest fraction (Gonçalves
255	Junior et al., 2014; Bambi et al., 2016; Tonin et al., 2017).



256	Higher litter production in the abandoned AFS is reflected in the observed spatial and seasonal
257	patterns of litter inputs and standing stocks, which were also closer to those in the secondary
258	forest than in the managed AFS. This indicates that abandoned AFS with little human
259	intervention may provide favorable conditions for forest regeneration (Rolim et al., 2017) and
260	could thus be valuable strategic sites for the conservation of riparian forest remnants in these
261	agroforestry systems of Brazil (Faria et al., 2007; Cassano et al., 2009; Scroth et al., 2011;
262	Sambuich et al., 2012). The presence of pioneer species in abandoned AFS could be a critical
263	factor promoting this regeneration by facilitating ecological succession (Rolim & Chiarello 2004)
264	Sambuichi & Haridasan 2007, Sambuichi et al. 2012).
265	Crop management, which involves hoeing and soil cleaning (Sambuichi et al., 2012; Mello &
266	Gross, 2013), may be a critical factor accounting for the observed tendency of greater lateral
267	litter inputs to streams in managed AFS. Possible mechanisms include facilitation of litter
268	leaching and movement along streams by runoff (Afonso, Henry & Rodella, 2000; Wantzen et
269	al., 2008) as well as effects on the structure and composition of the tree vegetation (Deheuvels et
270	al., 2014). Furthermore, some management practices introduce exotic species necessary for
271	cultivation (Sambuichi, 2002; Piasentin, Saito & Sambuichi, 2014; Rolim et al., 2017) and
272	involve thinning of the vegetation to obtain the desired shade levels for crop production (Johns,
273	1999).
274	The management practices in cocoa cultivation and phenology of the shade species may also
275	have determined the greater contribution of reproductive plant parts in the managed AFS. This
276	observation is related to the high proportion of exotic species (e.g. Artocarpus heterophyllus,
277	Spondias mombin and, particularly in this study, Clitoria fairchildiana), which show different
278	phenological patterns than native riparian forests (Sambuichi & Haridasan, 2007; Sambuichi et



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al., 2012). Furthermore, variation in the size, shape, texture and anatomy of reproductive plant parts of different tree species in different forest types could account for the higher contribution of this litter fraction in managed AFS. This applies particularly to fruits of the legume *Clitoria* fairchildiana, the species contributing most to the reproductive plant parts in litter vertical inputs and standing stocks in this forest. The large dry and dehiscent fruits of the tree are 25 to 30 cm long and 2.6 to 2.9 cm wide (Silva & Môro, 2008), Rezende et al. (2017). A potentially higher nutritional quality of such reproductive plant parts compared to leaves may favor rapid decomposition of the litter supplied to tropical streams. However, the fruits of *Clitoria* fairchildiana and other species in the managed AFS we investigated may not provide a highquality nutritional resource (Rezende et al., 2017), and we do not currently have data to evaluate the consequences for litter decomposition. Litter standing stocks in ecosystems are the net result of litter inputs and losses by movement and decomposition (Elosegi & Pozo, 2005; Tank et al., 2010). The high standing stocks we observed in streams of secondary forests could imply high inputs (França et al., 2009; Lisboa et al., 2015; Bambi et al., 2016) combined with slow decomposition and downstream transport (Gonçalves Júnior et al., 2014; Rezende et al., 2017a). In particular, the high proportion of miscellaneous matter in the secondary forest stream could be related to rapid decomposition. Effective litter retention favors long residence times in streams (Bilby & Likens, 1980) and the generation of small organic particles released during decomposition ((Gessner et al. 1999, Boyero et al., 2011)) instead of coarse litter being transported downstream-. This contrasts with the situation in the managed AFS where cocoa leaves are the predominant litter type, the high lignin and cellulose concentrations of which slow litter decomposition in the soil of cocoa AFS (Dawoe, Isaac & Quashie-Sam, 2010) and likely also in streams (Tank et al., 2010; Lemes da Silva et al., 2017).



In the managed AFS, in contrast, it may have been the high proportion of recalcitrant reproductive plant parts (see above) that favored high benthic standing stock in this forest. This type of litter was rare in the secondary forest and abandoned AFS.

Our finding that litter inputs and standing stocks in the abandoned AFS were more similar to those in the secondary forest than in the managed cocoa AFS suggests that abandoning management measures, such as litter removal and soil cleaning, could create favorable conditions for reestablishing natural litter dynamics in riparian zones and streams of former AFSs. The capacity of tree species richness to regenerate is high in those forests (Sambuichi & Haridasan, 2007), if surrounding vegetation remains intact to provide a seed source for forest recovery (Rolim et al., 2017). It appears that the absence of management in the riparian zone of abandoned AFS sufficiently reduces pressure on species during regeneration (Rolim et al., 2017) for the phenology of species to become the main factor determining litter dynamics. Intensive management, in contrast, overrides the importance of phenology by altering the structure of riparian plant communities (Delong & Brusven, 1994; Ferreira et al. 2019).

Conclusion

Although our study was restricted to one location for each of the three investigated forest types (SF, AC and MC), our findings are a starting point to evaluate differences in litter inputs and standing stocks between those forests. The observed similarity with litter inputs and standing stocks in the secondary forest suggests potential of the abandoned AFS to provide favorable conditions for restoring natural litter dynamics in streams and riparian zones. However, future investigations are needed to elucidate the nutritional quality of litter and its variation depending on the composition and structure of the riparian vegetation.



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331	References
332	Abelho M, Graça, AS. 1996. Effects of eucalyptus afforestation on leaf litter dynamics and
333	macroinvertebrate community structure of streams in Central Portugal. Hydrobiologia 324
334	195-204.
335	Afonso AAO, Henry R, Rodella RCSM. 2000. Allochthonous matter input in two different
336	stretches of a headstream (Itatinga, São Paulo, Brazil). Brazilian Archives of Biology and
337	Technology 43: 335-43.
338	Argôlo LMH. 2009. Avaliação de genótipos de Heliconia spp. sob cultivo a pleno sol e cabruca.
339	Master's Thesis. Universidade Estadual de Santa Cruz.
340	Bambi P, Rezende RS, Feio MJ, Leite GFM, Alvin E, Quintão JMB, Araújo F, Gonçalves Júnior
341	JF. 2016. Temporal and spatial patterns in inputs and stock of organic matter in savannah
342	streams of central Brazil. Ecosystems 20: 757-768.
343	Barreto PA, Gama-Rodrigues EF, Gama-Rodrigues AC, Fontes AG, Polidoro JC, Moço MKS,
344	Machado RC, Baligar VC. 2011. Distribution of oxidizable organic C fractions in soils
345	under cacao agroforestry systems in Southern Bahia, Brazil. Agroforestry systems 81:213-
346	220.



347	Bawa KS, Kress WJ, Nadkarni NM, Lele SR, Janzen DH, Lugo AE, Ashton PS, Lovejoy TE.
348	2004. Tropical ecosystems into the 21st century. Science 306: 227-228.
349	Bates D, Mächler M, Bolker BM, Walker SC. 2015. Fitting linear mixed-effects models using
350	lme4. Journal of Statistical Software 67: 1-48.
351	Beer J, Muschler R, Somarriba E, Kass D. 1998. Shade management in coffee and cacao
352	plantations. Agroforestry Systems 38: 139-164.
353	Bilby RE, Likens GE. 1980. Importance of organic debris dams in the structure and function of
354	stream ecosystems. <i>Ecology</i> 61: 1107-1113.
355	Boyero L, Pearson RG, Gessner MO, Barmuta LA, Ferreira V, Graça MAS, Dudgeon D, Boulton
356	AJ, Callisto M, Chauvet E, Helson JE, Bruder A, Albariño RJ, Yule CM, Arunachalam M,
357	Davies JN, Figueroa R, Flecker AS, Ramírez A, Death RG, Mathooko JM, Mathuriau C,
358	Gonçalves JF, Moretti MC, Jinggut T, Lamothe ST, M'Erimba C, Ratnarajah L, Schindler
359	MH, Castela J, Buria LM, Cornejo A, Villanueva VD, West DC, 2011. A global experimen
360	suggests climate warming will not accelerate litter decomposition in streams but might
361	reduce carbon sequestration Ecology Letters 14: 289–294
362	Calderón CC, Rezende RS, Calor A, Dahora JAS, Aragao LN, Guedes L, Caiafa AN, Medeiros
363	AO. 2019. Temporal dynamics of organic matter, hyphomycetes and invertebrate
364	communities in a Brazilian savanna stream. Community Ecology 20: 301-313.
365	Canty AJ, Ripley B. 2016. Boot: Bootstrap R (S-Plus) functions, R-package version 1.3-10.
366	Cassano, CR, Schroth, G, Faria D, Delabie JHC, Bede L. 2009. Landscape and farm scale
367	management to enhance biodiversity conservation in the cocoa producing region of
368	southern Bahia, Brazil. Biodiversity Conservation 18: 577-603.



369	Costa END, De Souza MFL, Marrocos PCL, Lobão D, Silva DML. 2018. Soil organic matter
370	and CO2 fluxes in small tropical watersheds under forest and cacao agroforestry. PloS one
371	13 (7). DOI 10.1371/journal.pone.0200550.
372	Costa END, Souza JC, Pereira MA, Souza WFL, Souza MFL, Silva DML. 2017. Influence of
373	hydrological pathways on dissolved organic carbon fluxes in tropical streams. Ecology and
374	Evolution 1: 1-12.
375	Curvelo K, Calasans NA, Lobão, DE, Sodré, GA, Pereira, JM, Marrocos, PCL, Barbosa, JW,
376	Valle, RR. 2009. Aporte de nutrientes na serapilheira e na água do solo em cacau-cabruca,
377	floresta secundária e pastagem. Agrotrópica 21: 55 – 64.
378	Davison AC, Hinkley DV. 1997. Bootstrap methods and their application. Cambridge
379	University Press, Cambridge, England.
380	Dawoe EK, Isaac ME, Quashie ASJ. 2010. Litterfall and litternutrient dynamics under cocoa
381	ecosystems in lowland humid Ghana. Plant and Soil 330: 55-64.
382	Deheuvels O, Rousseau GX, Quiroga GS, Franco MD, Cerda R, Mendoza SJV, Somarriba E.
383	2014. Biodiversity is affected by changes in management intensity of cocoa-based
384	agroforests. Agroforestry Systems 88: 1081-1099.
385	Delong MD, Brusven MA. 1994. Allochthonous input of organic matter from different riparian
386	habitat of an agricultural impacted stream. Environmental Management: 18: 59-71.
387	Elosegi A, Pozo J. 2005. Litter input. In Graça, MAS, Bärlocher F, Gessner MO, ed. Methods to
388	study litter decomposition: a practical guide. Springer, New York, USA, 3-11.
389	Faria D, Laps RR, Baumgarten J, Cetra M. 2007. Bat and bird assemblages from forests and
390	shade cacao plantations in two contrasting landscapes in the Atlantic Forest of southern
391	Bahia, Brazil. Biodiversity and Conservation 15: 587-612.



392	reffella v, Boyelo L, Calvo C, Collea r, Figueloa R, Goliçaives Jr, Goyellola G, Glaça MA,
393	Hepp LU, Kariuki S, López-Rodríguez A. 2019. A Global assessment of the effects of
394	eucalyptus plantations on stream ecosystem functioning. Ecosystems 22: 629-642.
395	Fontes AG, Gama-Rodrigues AC, Gama-Rodrigues EF, Sales MVS, Costa MG, Machado RCR.
396	2014. Nutrient stocks in litterfall and litter in cocoa agroforests in Brazil. Plant and Soil
397	383: 313-335.
398	França JS, Gregório RS, D'Arc de Paula J, Gonçalves Júnior JF, Ferreira FA, Callisto M. 2009.
399	Composition and dynamics of allochthonous organic matter inputs and benthic stock in a
400	Brazilian stream. Marine and Freshwater Research 60: 990-998.
401	Gama-Rodrigues EF, Nair PKR, Nair VD, Gama-Rodrigues AC, Baligar VC, Machado RCR.
402	2010. Carbon storage in soil size fractions under two cacao agroforestry systems in Bahia,
403	Brazil. Environmental Management 45: 274-283.
404	Gessner MO, Chauvet E, Dobson M. 1999. A Perspective on Leaf Litter Breakdown in Streams.
405	Oikos 85: 377-384.
406	Gonçalves Júnior JF, Rezende RS, Gregorio RS, Valentin GC. 2014. Relationship between
407	dynamics of litterfall and riparian plant species in a tropical stream. Limnologica 44: 40-8.
408	Johns ND. 1999. Conservation in Brazil's chocolate forest: the unlikely persistence of the
409	traditional cocoa agroecosystem. Environmental Management 23: 31-47.
410	Lemes da Silva AL, Lisboa LK, Siegloch AE, Petrucio MM, Gonçalves Júnior JF. 2017.
411	Connecting the litterfall temporal dynamics and processing of coarse particulate organic
412	matter in a tropical stream. Marine and Freshwater Research 68: 1260.
413	Lindman J, Jonsson M, Burrows RM, Bundschuh M, Sponseller RA. 2017. Composition of
414	riparian litter input regulates organic matter decomposition: Implications for headwater



415	stream functioning in a managed forest landscape. Ecology and Evolution 7. 1008-1077
416	1271.
417	Lisboa LK, Da Silva ALL, Siegloch AE, Gonçalves Júnior JF, Petrucio MM. 2015. Temporal
418	dynamics of allochthonous coarse particulate organic matter in a subtropical Atlantic
419	rainforest Brazilian stream. Marine and Freshwater Research 66: 1-7.
420	Martini AMZ, Fiaschi P, Amorim AM, Paixão JL. 2007. A hot-point within a hot-spot: a high
421	diversity site in Brazil's Atlantic Forest. Biodiversity and Conservation 16: 3111-31288.
422	Mello DLN, Gross E. 2013. Guia de manejo do agroecossistema cacau cabruca. Editora
423	Instituto Cabruca, Ilhéus, Brazil.
424	Naiman RJ, Décamps H, Mcclain ME. 2005. Riparian ecology, conservation, and management
425	of stream side communities. Elsevier Academic Press, Burlington, USA.
426	Nair PKR, Buresh RJ, Mugendi DN, Latt CR. 1999. Nutrient cycling in tropical agroforestry
427	systems: Myths and science. In Buck LE, Lassoie JP, Fernandes ECM, ed. Agroforestry in
428	sustainable agricultural systems: Advances in agroecology. CRC Press, Boca Raton, FL,
429	USA, pp. 1-31.
430	Neres-Lima V, Machado Silva F, Baptista DF, Oliveira RBS, Andrade PM, Oliveira A F, Sasada
431	Sato CY, Silva Junior EF, Feijó Lima R, Angelini R, Camargo PB, Moulton TP. 2017.
432	Allochthonous and autochthonous carbon flows in food webs of tropical forest streams.
433	Freshwater Biology 62: 1012-1023.
434	Piasentin FB, Saito CH, Sambuichi RHR. 2014. Local tree preferences in the cacao-cabruca
435	system in the southeast of Bahia, Brazil. Ambiente & Sociedade 17: 55-78.



436	Pozo J, Elosegi A, Díez J, Molinero J. 2009. Dinámica y relevancia de la materia organia. In
437	Elosegi A, Sabater S, ed. Conceptos y técnicas en ecología fluvial. Fundación BBVA, País
438	Vasco, Espanha,141-168.
439	Rezende RS, Sales MA, Hurbath F, Roque N, Gonçalves Júnior JF, Medeiros AO. 2017a. Effect
440	of plant richness on the dynamics of coarse particulate organic matter in a Brazilian
441	Savannah stream. Limnologica 63: 57-64.
442	Rezende RS, Santos AM, Medeiros AO, Gonçalves Júnior JF. 2017b. Temporal leaf litter
443	breakdown in a tropical riparian forest with an open canopy. Limnetica 36: 445-459.
444	Rezende RS, Medeiros AO, Gonçalves Júnior JF, Feio MJ, Gusmão EP, de Andrade Gomes VÂ,
445	Calor A, Almeida JDSD. 2019. Patterns of litter inputs, hyphomycetes and invertebrates in
446	a Brazilian savanna stream: a process of degradative succession. Journal of Tropical
447	Ecology 35: 297-307.
448	Rolim SG, Chiarello AG. 2004. Slow death of Atlantic forest trees in cocoa agroforestry in
449	southeastern Brazil. Biodiversity and Conservation 13: 2679-2694.
450	Rolim SG, Sambuichi RHR, Schroth G, Nascimento MT, Gomes JML. 2017. Recovery of forest
451	and phylogenetic structure in abandoned cocoa agroforestry in the Atlantic Forest of
452	Brazil. Environmental Management 59: 410-418.
453	Sambuichi RHR. 2002. Fitossociologia e diversidade de espécies arbóreas em cabruca (Mata
454	Atlântica raleada sobre plantação de cacau) na região sul da Bahia, Brasil. Acta Botanica
455	Brasiliensis 16: 89-101.
456	Sambuichi RHR, Haridasan M. 2007. Recovery of species richness and conservation of native
457	Atlantic Forest trees in the cacao plantations of southern Bahia in Brazil. Biodiversity and
458	Conservation 16: 3681-3701.



459	Sambuichi RH, Vidal DB, Piasentin FB, Jardim, JG, Viana, TG, Menezes AA, Baligar VC.
460	2012. Cabruca agroforests in southern Bahia, Brazil: tree component, management
461	practices, and tree species conservation. Biodiversity and Conservation 21: 1055-1077.
462	Santos ERM, Araujo QR, Faria Filho AF, Vieira RB, Assunção Neto JF, Cabral LCC. 2016.
463	Aspectos geoambientais dos recursos hídricos das propriedades rurais do Projeto Barro
464	Preto, Bahia, Brasil. In: Bruck MME, Lorandi R. ed. Métodos e técnicas de pesquisa em
465	bacias hidrográficas. Editus, Ilhéus, Brazil, 121-138.
466	Schroth G, Faria D, Araujo M, Bede L, Van Bael SA, Cassano CR, Oliveira LC, Delabie JH.
467	2011. Conservation in tropical landscape mosaics: the case of the cacao landscape of
468	southern Bahia, Brazil. Biodiversity and Conservation 20: 1635-1654.
469	Seena S, Carvalho F, Cássio F, Pascoal C. 2017. Does the developmental stage and composition
470	of riparian forest stand affect ecosystem functioning in streams? Science of the Total
471	Environment 609: 1500-1511.
472	Silva BM, Moro FV. 2008. Aspectos morfológicos do fruto, da semente e desenvolvimento pós-
473	seminal de faveira (Clitoria fairchildiana R. A. Howard Fabaceae). Revista Brasileira de
474	Sementes 30: 195-201.
475	Souza JCS, Pereira MA, Costa END, Silva DML. 2017. Nitrogen dynamics in soil solution under
476	different land uses: Atlantic forest and cacao-cabruca system. Agroforestry Systems 92:
477	425-435.
478	Sutfin NA, Wohl EE, Dwire KA. 2016. Banking carbon: a review of organic carbon storage and
479	physical factors influencing retention in floodplains and riparian ecosystems. Earth Surface
480	Processes and Landforms 41: 38-60.



481	Tank JL, Rosi-Marshall EJ, Griffiths NA, Entrekin SA, Stephen ML. 2010. A review of
482	allochthonous organic matter dynamics and metabolism in streams. Journal of the North
483	American Benthological Society 29: 118-146.
484	Taubert F, Fischer R, Groeneveld J, Lehmann S, Müller MS, Rödig E, Wiegand T, Huth A.
485	2018. Global patterns of tropical forest fragmentation. <i>Nature</i> , 554(7693): 519-522.
486	Tiegs SD, Costello DM, Isken MW, Woodward G, McIntyre PB, Gessner MO, Chauvet E,
487	Griffiths NA, Flecker AS, Acuña V, Albariño R, Allen DC, Alonso C, Andino P, Arango
488	C, Aroviita J, Barbosa MVM, Barmuta LA, Baxter CV, Bell TDC, Bellinger B, Boyero L,
489	Brown LE, Bruder A, Bruesewitz DA, Burdon FJ, Callisto M, Canhoto C, Capps KA,
490	Castillo MM, Clapcott J, Colas F, Colón-Gaud C, Cornut J, Crespo-Pérez V, Cross WF,
491	Culp JM, Danger M, Dangles O, Eyto E, Derry AM, Villanueva VD, Douglas MM,
492	Elosegi A, Encalada AC, Entrekin S, Espinosa R, Ethaiya D, Ferreira V, Ferriol C,
493	Flanagan KM, Fleituch T, Follstad Shah JJ, Barbosa AF, Friberg N, Frost PC, Garcia EA,
494	Lago LG, Soto PEG, Ghate S, Giling DP, Gilmer A, Gonçalves Jr. JF, Gonzales RK, Graça
495	MAS, Grace M, Grossart HP, Guérold F, Gulis V, Hepp LU, Higgins S, Hishi T, Huddart
496	J, Hudson J, Imberger S, Iñiguez-Armijos C, Iwata T, Janetski DJ, Jennings E, Kirkwood
497	AE, Koning AA, Kosten S, Kuehn KA, Laudon H, Leavitt PR, da Silva ALL, Leroux SJ,
498	LeRoy CJ, Lisi PJ, MacKenzie R, Marcarelli AM, Masese FO, McKie BG, Medeiros AO,
499	Meissner K, Miliša M, Mishra S, Miyake Y, Moerke A, Mombrikotb S, Mooney R,
500	Moulton T, Muotka T, Negishi JN, Neres-Lima V, Nieminen ML, Nimptsch J, Ondruch J,
501	Paavola R, Pardo I, Patrick CJ, Peeters ETHM, Pozo J, Pringle C, Prussian A, Quenta E,
502	Quesada A, Reid B, Richardson JS, Rigosi A, Rincón J, Rîşnoveanu G, Robinson CT,
503	Rodríguez-Gallego L, Royer TV, Rusak J A, Santamans AC, Selmeczy GB, Simiyu G,



504	Skuja A, Smykla J, Sridhar KR, Sponseller R, Stoler A, Swan CM, Szlag D, Mello FT,
505	Tonkin JD, Uusheimo S, Veach AM, Vilbaste S, Vought LBM, Wang CP, Webster JR,
506	Wilson PB, Woelfl S, Xenopoulos MA, Yates AG, Yoshimura C, Yule CM, Zhang YX,
507	Zwart JA. Global patterns and drivers of ecosystem functioning in rivers and riparian
508	zones. Science Advances 518, p.eaav0486.
509	Tonin AM, Boyero L, Bambi P, Pearson RG, Correa-Araneda F, Gonçalves Júnior JF. 2019.
510	High within-stream replication is needed to predict litter fluxes in wet-dry tropical streams.
511	Freshwater Biology 65: 688-697. DOI:10.1111/fwb.13459
512	Tonin AM, Gonçalves JF, Bambi P, Couceiro SR, Feitoza LA, Fontana LE, Hamada N, Hepp
513	LU, Lezan-Kowalczuk VG, Leite GF, Lemes-Silva AL. 2017. Plant litter dynamics in the
514	forest-stream interface: precipitation is a major control across tropical biomes. Scientific
515	Reports 7: 10799. DOI: 10.1038/s41598-017-10576-8
516	Vannote RL, Minshall GW, Cummins KW, Sedell, JR, Cushing CE. 1980. The river continuum
517	concept. Canadian Journal of Fisheries and Aquatic Sciences 37: 130-137.
518	Wantzen KM, Yule CM, Mattooko J.M, Pringle CM. 2008. Organic matter processing in tropical
519	streams. In: Dudgeon D, ed. Tropical stream ecology. Elsevier, Amsterdam, Netherlands,
520	43-64
521	Webster J, Covich A, Tank J, Crockett, T. 1994. Retention of coarse organic particles in streams
522	in the southern Appalachian Mountains. Journal of the North American Benthological
523	Society 13: 140-150.
524	Wild R, Gücker B, Brauns M. 2019. Agricultural land use alters temporal dynamics and the
525	composition of organic matter in temperate headwater streams. Freshwater Science, 38(3):
526	566-581.



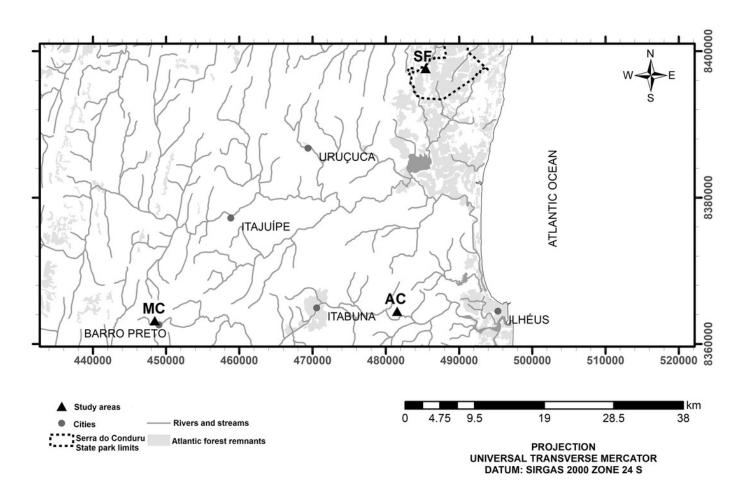
PeerJ

527	Winbourne JB, Feng A, Reynolds L, Piotto D, Hastings MG, Porder S. 2018. Nitrogen cycling
528	during secondary succession in Atlantic Forest of Bahia, Brazil. Scientific Reports 8: 1-9
529	Wood SN. 2017. Generalized Additive Models: An Introduction with R. Chapman & Hall/CRC
530	Boca Raton, EUA.
531	Zuur AF, Ieno EN, Walker N, Saveliev AA, Smith GM. 2009. Mixed Effects Models and
532	Extensions in Ecology with R. Springer-Verlag, New York, USA.
533 534	



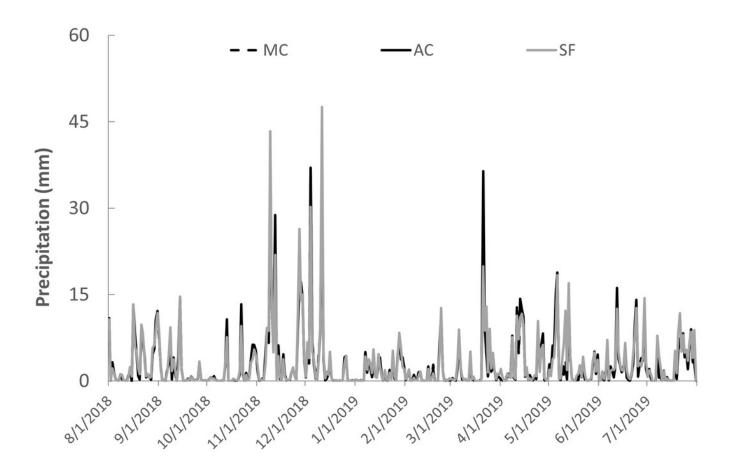
Location of study sites in secondary forest (SF), a managed AFS (MC) and an abandoned AFS (AC) in northeastern Brazil

Map database: Instituto Brasileiro de Geografia e Estatística (Brazilian territory, cities and hydrography, 2017); Ministério do Meio Ambiente (Serra do Conduru State Park limits, 2012), Fundação SOS Mata Atlântica and Instituto Nacional de Pesquisas Espaciais (Atlantic Forest remnants, 2016).

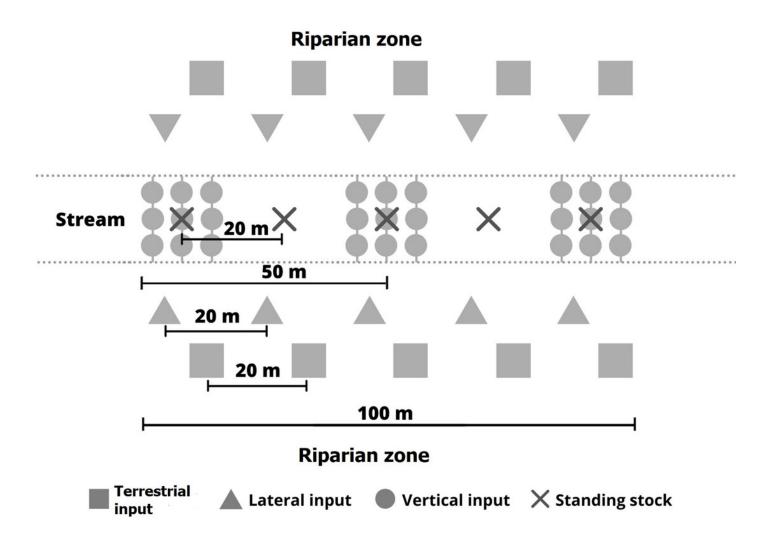




Daily precipitation at the study sites in secondary forest (SF), a managed AFS (MC) and an abandoned AFS (AC) in northeastern Brazil



Sampling design to determine litter inputs and standing stocks in riparian zone and streams



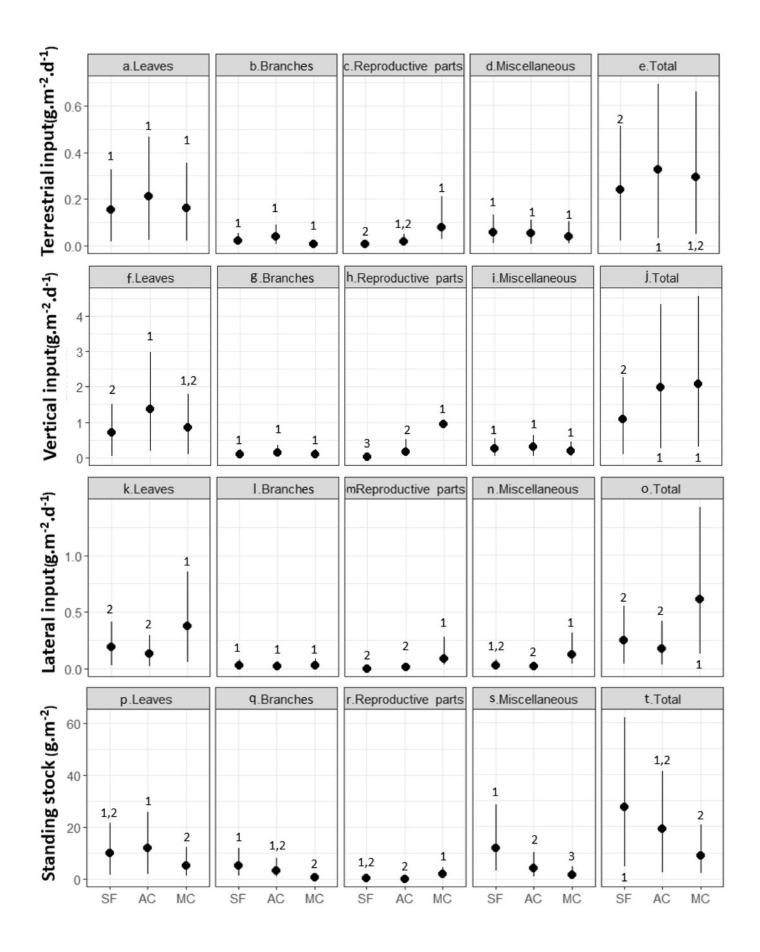


Inputs and standing stocks of various litter fractions in streams and riparian zones of secondary forest (SF), a managed cabruca (MC) and an abandoned cabruca (AC).

Black circles and vertical lines represent means and bootstrapped 95% confidence intervals.

Different numbers indicate significant differences of means as judged based on nonoverlapping confidence intervals.



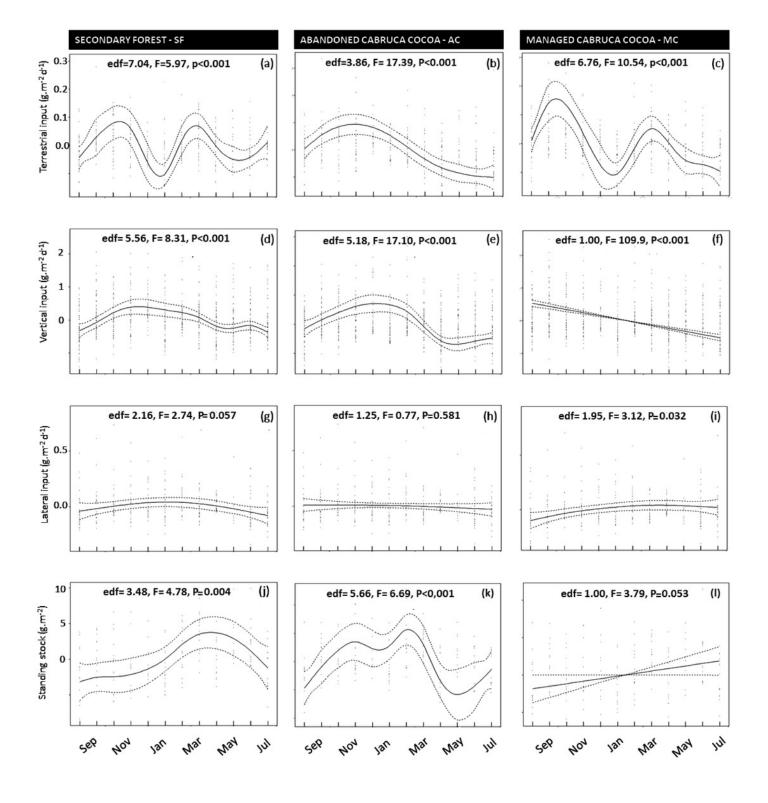




Temporal changes of litter inputs (g dry mass m^{-2} d^{-1}) and standing stocks (g dry mass m^{-2}) in secondary forest (SF), an abandoned cabruca (AC) and a managed cabruca (MC).

Also shown are F and P values as well as the effective degrees of freedom (edf) of GAMM analyses. Continuous lines are the GAMM smoothers and dotted lines indicate 95% confidence limits.







Summary of annual litter inputs and average standing stocks (g dry mass m⁻² year⁻¹) in streams and riparian zones

SF - secondary forest, AC- abandoned cabruca and MC - managed cabruca

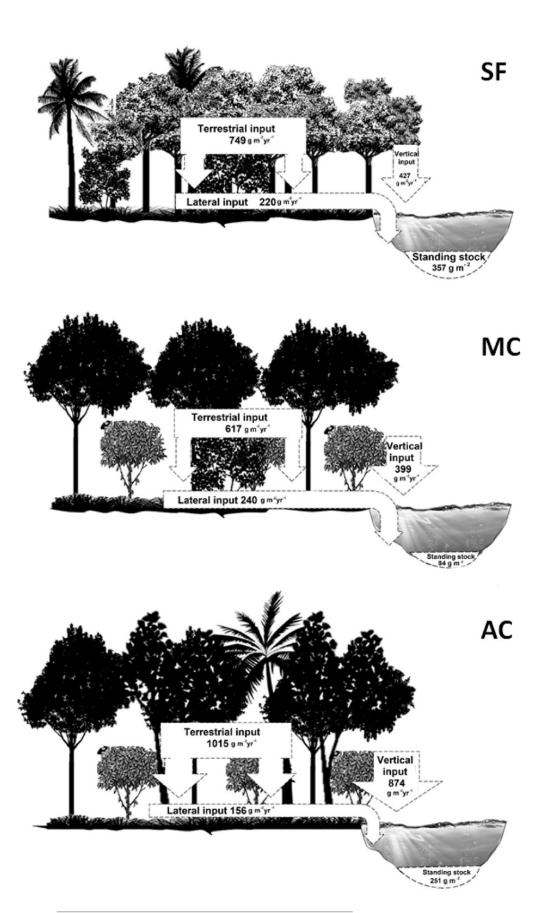




Table 1(on next page)

Absolute inputs and standing stocks (g m⁻²) as well as relative contributions (%) of various litter fractions to litter inputs and standings stocks in streams and riparian zones

Secondary Forest (SF), Managed Cabruca cocoa (MC) and Abandoned Cabruca cocoa (AC)

I :ttou	Farant		Terresti	ial		Vertical in	nput		Lateral in	put		Standing st	tock
Litter	Forest	Mean	Range	Contri-									
fraction	type	(g m ⁻²)	(g m ⁻²)	bution (%)	(g m ⁻²)	(g m ⁻²)	bution (%)	(g m ⁻²)	(g m ⁻²)	bution (%)	(g m ⁻²)	(g m ⁻²)	bution (%)
	SF	41	0 - 76	64	25	0 - 122	67	21	0 - 164	72	132	6 - 483	36
Leaves	MC	42	0 - 193	56	26	0 - 194	41	42	0 - 441	62	41	0 - 422	57
	AC	53	4 - 322	65	35	0 - 287	69	12	0 - 5160	74	150	30 - 424	61
	SF	2	0 - 92	9	0	0 - 201	8	0	0 - 83	11	52	0 - 431	19
Branches	MC	0	0 - 48	3	0	0 - 137	4	0	0 - 103	5	0	0 - 95	6
	AC	8	0 - 75	12	0	0 - 86	7	0	0 - 99	10	33	0 - 386	16
Reproductive	SF	0	0 - 32	4	0	0-33	2	0	0 - 15	1	0	0 - 76	2
•	MC	0	0 - 398	28	0	0 - 493	46	0	0 - 389	14	0	0 - 289	19
parts	AC	0	0 - 73	6	0	0 - 306	9	0	0 - 62	7	0	0 - 57	1
Miscellaneous	SF	7	0 - 137	23	4	0 - 108	23	0	0 - 63	12	116	0 - 1432	43
litter	MC	5	0 - 217	13	2	0 - 137	9	3	0 - 293	19	9	0 - 395	17
ntter	AC	13	0 - 46	16	9	0 - 56	15	0	0 - 47	9	50	0 - 382	22
	SF	61	8 - 221	-	34	0 - 230	-	24	0 - 215	-	357	28 - 2050	-
Total	MC	57	0 - 511	-	39	0 - 638	-	57	0 - 816	=	84	0 - 559	-
	AC	86	4 - 339	-	58	0 - 521	-	18	0 - 303	-	251	78 - 726	-