

1    **Nickel mine soil is a potential source for soybean plant growth promoting and heavy metal**  
2    **tolerant rhizobia**

3    Han Liu <sup>1, #</sup>, Yongliang Cui <sup>2, #</sup>, Jie Zhou <sup>1</sup>, Petri Penttinen <sup>1</sup>, Jiahao Liu <sup>1</sup>, Lan Zeng <sup>1</sup>, Qiang  
4    Chen <sup>1</sup>, Yunfu Gu <sup>1</sup>, Likou Zou <sup>1</sup>, Ke Zhao <sup>1</sup>, Qianju Xiang <sup>1</sup>, Xiumei Yu <sup>1 \*</sup>

5    <sup>1</sup> College of Resources, Sichuan Agricultural University, Chengdu 611130, China

6    <sup>2</sup> Sichuan Provincial Academy of Natural Resource and Sciences, Chengdu 610015, China

7

8    **Corresponding Author:**

9    Xiumei Yu

10    211th Huimin Street, Chengdu, Sichuan Province, 611130, China

11    Email address: xiumeiyu@sicau.edu.cn

12

13    <sup>#</sup>Han Liu and Yongliang Cui contributed equally to this work.

14   **Abstract**

15       Mine soil is not only barren but also contaminated by some heavy metal. It is unclear  
16 whether some rhizobia survived under extreme conditions in the nickel mine soil. Therefore, this  
17 study tries to isolate some effective soybean plant growth promoting and heavy metal resistant  
18 rhizobia from nickel mine soil, and to analyze their diversity. Soybean plants were used to trap  
19 rhizobia from the nickel mine soil. A total of 21 isolates were preliminarily identified as rhizobia,  
20 which were clustered into eight groups at 87% similarity level using BOXA1R-PCR  
21 fingerprinting technique. Four out of the eight representative isolates formed nodules on soybean  
22 roots with effectively symbiotic nitrogen-fixing and plant growth promoting abilities in the  
23 soybean pot experiment. Phylogenetic analysis of 16S rRNA, four housekeeping genes (*atpD*-  
24 *recA-glnII-rpoB*) and *nifH* genes assigned the symbiotic isolates YN5, YN8 and YN10 into  
25 *Ensifer xinjiangense* and YN11 into *Rhizobium radiobacter*, respectively. They also showed  
26 different tolerance levels to the heavy metals including cadmium, chromium, copper, nickel, and  
27 zinc. It was concluded that there were some plant growth promoting and heavy metal resistant  
28 rhizobia with the potential to facilitate phytoremediation and alleviate the effects of heavy metals  
29 on soybean cultivation in nickel mine soil, indicating a novel evidence for further exploring more  
30 functional microbes from the nickel mine soil.

31   **Keywords:** Rhizobia, Nickel mine soil, Soybean, Plant growth promoting, Diversity

32

Deleted: was

## 34 Introduction

35 Heavy metal contamination in mining related soils affects both the mining site and the  
36 surrounding environment. Heavy metals in soil can originate from natural minerals, yet  
37 anthropogenic activities are the main source (Lebrazi & Fikri-Benbrahim 2018). Heavy metal  
38 contamination is a risk to food security, ecological environment, and even to human health  
39 through bioaccumulation in the food chain (Long et al. 2021; Qin et al. 2021; Zhang et al. 2012;  
40 Zhou et al. 2013). Therefore, areas with severe heavy metal-contaminations need to be  
41 remediated before being used for the cultivation of crops, and the selected crops should not  
42 accumulate contaminants when growing on the slightly-contaminated areas. Remediating the  
43 contaminated soils requires efficient and economical methods such as bioremediation, which is  
44 considered eco-friendly, secondary contamination-free, and suitable for non-point source  
45 contamination (Shao et al. 2020; Yu et al. 2021; Zhang et al. 2020).

46 Phytoremediation, especially *in-situ* enhanced phytoremediation, is widely used for the  
47 bioremediation of heavy metal contaminated soil (Thakare et al. 2021). The growth of plants in  
48 contaminated soil can be facilitated by utilizing the biological nitrogen fixation (BNF) ability of  
49 legume-rhizobia symbionts (Hao et al. 2014; Salmi & Boulila 2021; Wang et al. 2019; Yu et al.  
50 2017). For example, soybean (*Glycine max* L. Merrill) is applicable in remediating heavy metal-  
51 contaminated sites (Li et al. 2019). In the symbiosis, rhizobia induce the formation of nodules on  
52 the roots of the host plant. Inside the nodules, rhizobia fix atmospheric nitrogen into ammonia  
53 which serves as a N source for the legume (Lindstrom & Mousavi 2020; Wang et al. 2020).  
54 Inoculating with effective N fixing rhizobial strains promotes the growth of legumes (Catroux et  
55 al. 2001). It has been proposed that strains suitable for legume-rhizobia phytoremediation can be  
56 isolated from the contaminated sites (Balakrishnan et al. 2017; Dhuldhaj & Pandya 2020; Fan et  
57 al. 2016; Limcharoensuk et al. 2015). Rhizobia include strains with heavy metal resistance and

Deleted: ir

are capable to promote plant growth under heavy metal stress (Fan et al. 2018b; Grandlic et al. 2009). It showed that a copper-resistant *S. meliloti* strain promoted the growth of alfalfa under copper stress (Duan et al. 2019). Rhizobial strains resistant against several heavy metals have the potential to be applied in the *in-situ* bioremediation of soils contaminated with multiple heavy metals (Abd-Alla et al. 2012; Grandlic et al. 2009; Hao et al. 2015; Ke et al. 2021; Yu et al. 2017; Yu et al. 2014). However, indigenous rhizobia resources that could be applied in *in situ* phytoremediation are still scarce in Southwest China.

In this study, through the cultivation experiment of soybean plant in nickel mine soil, we aimed to find effective plant growth promoting and heavy metal resistant strains to facilitate phytoremediation of heavy metal contaminated soil. We hypothesized that heavy metal contaminated soil could be a putative source for such strains. Thus, soybean plants were used to trap rhizobia from nickel mine soil in Xichang, Sichuan Province, China. The isolates were identified using molecular methods, and soybean growth-promoting abilities and heavy metal resistance of these strains were tested. The results provide better understanding of the potential of using indigenous microbial resources for the alleviation of heavy metal contamination.

## Materials and Methods

### Sampling and soil analysis

Soil samples were collected from a nickel mine in Xichang, Sichuan Province, China. Three sampling sites with 50 to 100 m apart were randomly selected within the area. Five topsoil (0-20 cm) subsamples from sampling points with at least 5 m in between were collected and mixed to make one composite sample per site. The composite samples were quartered and stored on ice before being taken to the laboratory. The soil samples were ground, passed through a 2 mm nylon sieve, and air-dried. Soil water content in the fresh soil was determined by measuring the

Deleted: ,

Deleted: t

weight difference after soil samples had been dried at 105 °C for eight hours. Soil pH was measured using a PHS-3C pH meter (Shanghai Yoke, China) in a 2.5:1 water-soil slurry which had been left to settle overnight. Soil organic carbon content was determined using the K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub>-H<sub>2</sub>SO<sub>4</sub> method (Schumacher, 2002). Total nitrogen, phosphorus and potassium contents in the soil samples were determined using Kjeldahl method (Kjeltec 8400, FOSS, Sweden), Mo-Sb colorimetric method (WFJ2100, UNICO, China) and flame spectrophotometry (FP6410, Shanghai Precision & Scientific, China), respectively (Murphy & Riley 1962; Page et al. 1982; Yu et al. 2021). Available nitrogen content was determined using the alkali N-proliferation method; soil available phosphorus and potassium were extracted with sodium bicarbonate solution and NH<sub>4</sub>AC solution, respectively, and measured using the previously described method (Wu et al. 2021). The contents of heavy metals were determined after digestion using mixed acid (HNO<sub>3</sub>: HClO<sub>4</sub>= 3: 1) using inductively coupled plasma optical emission spectrometry (ICP-OES; IRIS Intrepid II, Thermo Electron Corporation, USA).

Deleted: ) (

#### Trapping and isolation of rhizobia

The indigenous rhizobial strains in the nickel mine soil were trapped using soybean cultivar Nandou No.12 bred by Nanchong Institution of Agricultural Sciences, Sichuan, China. The seeds were sterilized by dipping into 95% alcohol for 3 min and 1% HgCl<sub>2</sub> for 5 min, followed by rinsing with sterile water (Yu et al. 2017). The soybean seeds were germinated in a pot filled with sterilized damp vermiculite in the dark for 24 hours, and transplanted into pots filled with soil collected from the nickel mine (S<sub>N</sub>). The soybean plants were harvested after 90 d. Three root nodules per soybean plant were selected and sterilized using the above-mentioned methods. The nodules were incubated on beef extract peptone agar at 28 °C, and the surface sterilization was considered successful when no colonies appeared in 24 h. After that, nodules were crushed in plastic tubes and inoculated on yeast-extract mannitol agar (yeast extract 1.5 g L<sup>-1</sup>, mannitol

109 1.0 g L<sup>-1</sup>, K<sub>2</sub>HPO<sub>4</sub> 0.5 g L<sup>-1</sup>, MgSO<sub>4</sub> 7H<sub>2</sub>O 0.2 g L<sup>-1</sup>, NaCl 0.1 g L<sup>-1</sup>, Congo red 0.04 g L<sup>-1</sup>, agar  
110 20 g L<sup>-1</sup>). The plates were maintained at 28 °C for 7 to 10 days. During the incubation period,  
111 single colonies were selected based on colony morphology and purified by repeated streaking  
112 (Yu et al. 2017). Purified isolates with round, plump, milky white, mucilaginous and smooth  
113 margin colonies were examined using light microscopy after Gram staining. Gram-negative  
114 isolates with rod-shaped cells were preserved in 20% (w/v) glycerol at -80 °C.

#### 115 **Phylogenetic diversity analysis**

116 The isolates were grown in YM liquid medium (yeast extract 1.5 g L<sup>-1</sup>, mannitol 1.0 g L<sup>-1</sup>,  
117 K<sub>2</sub>HPO<sub>4</sub> 0.5 g L<sup>-1</sup>, MgSO<sub>4</sub> 7H<sub>2</sub>O 0.2 g L<sup>-1</sup>, NaCl 0.1 g L<sup>-1</sup>, Congo red 0.04 g L<sup>-1</sup>) for 24 h at  
118 28 °C, in a shaking incubator, and DNA was extracted using TIANamp Bacteria DNA Kit  
119 (TIANGEN, China). The genetic diversity of the isolates was assessed using BOX-A1R PCR  
120 fingerprinting with the primer 5'-CCTCGGCAAGGACGCTGACG-3' (Chen et al. 2014).  
121 Amplification was done in a 10 µL volume system containing 5 µL of 2 × PCR mix, 0.2 µL of  
122 the primer (10 µmol L<sup>-1</sup>), 1 µL of template DNA (50 ng mL<sup>-1</sup>), and 3.8 µL double distilled water  
123 (ddH<sub>2</sub>O). The BOX-A1R PCR thermal profile included initial denaturation at 94 °C for 3 min, 30  
124 cycles of denaturation at 94 °C for 1 min, annealing at 52 °C for 1 min and extension at 65 °C for  
125 8 min, and a final extension at 65 °C for 16 min. The amplified fragments in 8 µL of the  
126 amplification mixture and the 200 bp DNA ladder were separated in 2 % (w/v) agarose gel with  
127 ethidium bromide at 80 V for 2.5 h, and were visualized under UV light with the patterns  
128 recorded. Based on the patterns, a BOX-A1R PCR cluster tree diagram was created using the  
129 unweighted pair group method with arithmetic averages (UPGMA) in NTSYSpc 2.1 (Yu et al.  
130 2014).

#### 131 **Plant growth promotion ability test**

Deleted: with

133 Based on the BOX-A1R PCR fingerprints, eight representative isolates were selected for the  
134 plant growth promotion ability test. The test was done in the above-mentioned Leonard jars. The  
135 glass bottle was filled with vermiculite and flushed with RO water until the vermiculite was  
136 thoroughly soaked. The plastic bottle was filled with nutrient solution (KCl 0.5 g L<sup>-1</sup>, CaSO<sub>4</sub>  
137 2H<sub>2</sub>O 0.2 g L<sup>-1</sup>, MgSO<sub>4</sub> 7H<sub>2</sub>O 0.2 g L<sup>-1</sup>, KH<sub>2</sub>PO<sub>4</sub> 0.2 g L<sup>-1</sup>, FeEDTA 1 mg L<sup>-1</sup>, CaCO<sub>3</sub> 2 g L<sup>-1</sup>,  
138 H<sub>3</sub>BO<sub>3</sub> 1 mg L<sup>-1</sup>, ZnSO<sub>4</sub> 7H<sub>2</sub>O 1 mg L<sup>-1</sup>, CuSO<sub>4</sub> 5H<sub>2</sub>O 0.5 mg L<sup>-1</sup>, MnCl<sub>2</sub> 4H<sub>2</sub>O 0.5 mg L<sup>-1</sup>,  
139 Na<sub>2</sub>MoO<sub>4</sub> 2H<sub>2</sub>O 0.1 mg L<sup>-1</sup>). The bottles were connected with a cotton wick which  
140 transferred nutrient solution from the plastic to the glass bottle, whose mouth was plugged with  
141 absorbent cotton (Trung et al. 1982). The Leonard jars were autoclaved at 121°C for 60 min.  
142 Soybean seeds were surface-sterilized and germinated as described above. Three seeds were  
143 transferred into one Leonard jar, which were then covered with a layer of approximately 3 cm  
144 moist vermiculite. When the seedlings were 2-3 cm tall, isolate culture in exponential phase was  
145 inoculated around the roots and the topsoil was covered with a layer of 1 cm autoclaved quartz  
146 sand. The non-inoculated control group included a N<sup>-</sup> treatment with the nutrient solution only  
147 and a N<sup>+</sup> treatment with 1g L<sup>-1</sup> KNO<sub>3</sub> as the nitrogen source in the nutrient solution. Each  
148 treatment included three replicates. When the seedlings were 10-15 cm tall, the weakest seedling  
149 was removed. Growth conditions were as follows: in the day mode, the temperature was 25 °C  
150 and the relative humidity was 80% for 17 h; in the night mode, the temperature was 20 °C and  
151 the relative humidity was 85% for 7 h. The jars were replenished with nutrient solution when  
152 needed. The soybean plants were harvested after 55 days, the number of nodules, root length,  
153 and plant height and weight were measured. The plant samples were dried at 105 °C for 20 min  
154 and at 55 °C for 7 d for analyses of nitrogen, phosphorus, and potassium contents.

## 155 Sequence analysis

156 The isolates that nodulated soybean plants were further characterized using sequence

Commented [PS1]: Please mention the cell density

analyses. The almost full length 16S rRNA gene was amplified using the primer pair 27F/1492R (Yu et al. 2017), the housekeeping genes *atpD*, *recA*, *glnII*, and *rpoB* using primer pairs *atpD*255F/ *atpD*782R, *recA*63F/ *recA*555R, *glnII*12F/ *glnII*tsR, and *rpoB*454F/ *rpoB*1364R, respectively (Tang et al. 2012; Zhao et al. 2014), and the nitrogen fixation gene *nifH* using primer pair *nifHF*/ *nifHI* (Laguerre et al. 2001). Amplification was done in a 30 µL volume system containing 15 µL of 2 × PCR mix, 0.15 µL of each primer, one µL of template DNA (50 ng mL<sup>-1</sup>), and 13.7 µL ddH<sub>2</sub>O. In the amplification of 16S rRNA, *atpD*, *recA*, *glnII* and *rpoB* genes, the thermal cycling conditions included an initial denaturation at 94 °C for 3 min, 30 cycles of denaturation at 94 °C for 1 min, annealing at 55 °C for 1 min, extension at 72 °C for 2 min, and a final extension at 72 °C for 10 min. In the amplification of *nifH*, the thermal cycling conditions included an initial denaturation at 95 °C for 3 min, 30 cycles of denaturation at 94 °C for 1 min, annealing at 59 °C for 1 min, extension at 72 °C for 5 min, and a final extension at 72 °C for 6 min. The amplification products were sequenced by Sangon Biotech (Shanghai, China). The fragments of *atpD* (274 bp), *recA* (245 bp), *glnII* (413bp) and *rpoB* (534 bp) were concatenated for multilocus sequence analysis (MLSA). The sequences were matched against reference sequences in the NCBI database using BLAST. The sequences of the isolates and the reference sequences were subjected to phylogenetic analysis using neighbor joining method in MEGA7.0. The phylogenetic trees were bootstrapped with 1,000 replications for each sequence to evaluate the reliability of the tree topologies (Saitou & Nei 1987 ).

**Commented [PS2]:** Please provide the list of primers for all the genes used in this study in the form of a supplementary table.

**Commented [PS3]:** What were the reference sequences used.

## Heavy metal resistance test

For assessing the heavy metal-resistance ability of the isolates, 10 g L<sup>-1</sup> stock solutions of cadmium (CdCl<sub>2</sub> 2.5H<sub>2</sub>O), chromium (K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub>), copper (CuCl<sub>2</sub> 2H<sub>2</sub>O), nickel (NiSO<sub>4</sub>), and zinc (ZnSO<sub>4</sub> 7H<sub>2</sub>O) were prepared by dissolving the salts in ultrapure water. The isolates were grown in 5 mL YM liquid medium with 0, 4, 8, 12, 16, 20, 40, 60, 80, 100 mg L<sup>-1</sup> metal in an orbital



181 shaker at 28 °C for 72h, followed by measuring optical density at 600 nm using a  
182 spectrophotometer (UV-3300, Shanghai MAPADA, China) (Abd-Alla et al. 2012).

183 The minimum inhibitory concentrations (MIC) were determined by comparing the OD<sub>600nm</sub>  
184 value of cultures spiked with metals to those without metals. MIC was defined as the lowest  
185 metal concentration resulting in a visually observable decrease in the growth curve of the isolates.  
186 Minimum lethal concentration (MLC) was defined as the lowest metal concentration resulting in  
187 an OD<sub>600nm</sub> value lower than 0.1.

## 188 Statistical analysis

189 Differences in soil properties and soybean growth parameters were tested using one-way  
190 ANOVA and Fisher's protected LSD (least significant difference) test at  $P \leq 0.05$  test in IBM  
191 SPSS statistics 22.0. The differences in MIC and MLC values were not tested due to their zero  
192 variance.

## 193 Results

### 194 Isolation of rhizobia from nickel mine soil

195 The soil in the nickel mining area was slightly acidic (pH 6.35), and the organic matter  
196 content was low (1.07 %). The contents of total N, P and K were 261.35, 479.85, and 5869.73  
197 mg kg<sup>-1</sup>, respectively, and 15.08 %, 2.93 %, and 0.38 % of them were available fragments (Table  
198 1). Both nickel and lead contents were approximately 20 mg kg<sup>-1</sup>, and cadmium, chromium,  
199 copper, iron, manganese and zinc were also detected (Table 1).

200 As a result of trapping rhizobial strains from the nickel mine soil using soybean as the host  
201 plant, 21 isolates were preliminarily identified as rhizobia based on colony morphology and  
202 microscopic examination of cell shape and Gram staining. The BOXA1R-PCR fingerprinting

Commented [PS4]: Please include the Table title and footnote with all the abbreviations

(Fig. 1) of the isolates showed that sixteen distinct fingerprint patterns were found, indicating that these isolates from nickel mine soil were genetically diverse. The isolates were clustered into 8 groups at 87% similarity level.

**Commented [PS5]:** Please include the legends for all the figures

#### **Plant growth promotion ability of selected isolates**

Based on the fingerprinting, we selected 8 isolates (YN1, YN5, YN8, YN10, YN11, YN13, YN27, YN30) for the plant growth promotion ability test. Only four isolates (YN5, YN8, YN10, YN11) nodulated soybean plants, with nodule numbers ranging from 50 to 54 per plant (mean value of six plants). The biomass of soybeans inoculated with the four symbiotic isolates (YN5, YN8, YN10, YN11) and isolate YN13 was significantly higher than that in the  $N^-$  treatment ( $P = 0.00$ ) (Fig. 2a). The biomass of soybeans inoculated with isolate YN8 was on the same level as in the  $N^+$  treatment. The shoots of soybeans inoculated with the 4 symbiotic isolates (YN5, YN8, YN10, YN11) and isolate YN1 were significantly longer than those in the  $N^-$  treatment ( $P=0.00, 0.01, 0.00, 0.00, 0.00$ ). The roots of soybeans inoculated with the symbiotic isolates YN8, YN10, and YN11 were significantly longer than those in the  $N^-$  treatment ( $P=0.00$ ) (Fig. 2b). The shoot N content of soybean plants inoculated with the symbiotic strains was at the same level as in the  $N^+$  treatment, and the shoot N content of soybeans inoculated with the symbiotic strains of YN5, YN8, YN10, and YN11 as well as the isolate YN1 was significantly higher than that in the  $N^-$  treatment ( $P=0.00$ ) (Fig. 3a). The root N content of inoculated soybean plants was higher than that in the  $N^-$  treatment (Fig. 3a). The root P content of soybeans inoculated with the symbiotic isolates and isolate YN30 was lower than that in the  $N^-$  treatment (Fig. 3b). The shoot K content of soybeans inoculated with the isolate YN10 was higher than that in the  $N^-$  treatment (Fig. 3c). The root K content of soybeans inoculated with the isolates YN5, YN8, YN10, YN13, YN27 and YN30 was higher than that in the  $N^-$  treatment (Fig. 3c).

#### **Sequence analyses of symbiotic isolates**

As a result of the sequence analysis of the almost full length 16S rRNA gene, the symbiotic isolates were assigned to the genera *Ensifer* (formerly designated as *Sinorhizobium*) and *Rhizobium* (Fig. 4). The isolates YN5, YN8, and YN10 were respectively grouped together with the type strains *E. fredii* USDA205, *E. americanum* CFNEI156 and *E. xinjiangense* CCBAU110, and isolate YN11 with *R. radiobacter* ICMP 5856 (formerly *Agrobacterium tumefaciens* ICMP 5856). Based on the multilocus sequence analysis (MLSA) of the concatenated fragments of *atpD* (274 bp), *recA* (245 bp), *glnII* (413bp) and *rpoB* (534 bp), the isolates YN5, YN8, and YN10 were identified as *E. xinjiangense* strains, and YN11 as *Rhizobium radiobacter* strain (Fig. 5). The *nifH* sequences from YN5, YN8, and YN10 were 100% similar with those from *E. fredii* CCBAU23314 and *E. xinjiangense* CCBAU110, and the *nifH* sequence from YN11 with that from *R. radiobacter* (Fig. 6).

#### Heavy metal resistance ability of symbiotic isolates

In general, *E. xinjiangense* YN5 and *R. radiobacter* YN11 tolerated higher levels of heavy metals compared to *E. xinjiangense* YN8 and YN10 (Fig. 7, Supplementary Table S1). For Cd<sup>2+</sup>, the MIC and MLC values of *E. xinjiangense* YN5 and *R. radiobacter* YN11 were the highest. For Cr<sup>2+</sup>, the MLC of all the strains was 16 mg L<sup>-1</sup>. For Cu<sup>2+</sup>, the MIC and MLC values of *E. xinjiangense* YN8 were the lowest. For Ni<sup>2+</sup>, the MLC values of YN5 and YN11 were higher than those of YN8 and YN10. For Zn<sup>2+</sup>, the MLC value of *E. xinjiangense* YN5 was 300 mg L<sup>-1</sup>, i.e., over three times higher than that of *R. radiobacter* YN11 and over 18 to 75 times higher than those of YN11 and YN8, respectively.

#### Discussion

We trapped rhizobia from nickel mine soil using soybean plants in Xichang, Sichuan Province, China, to find effectively plant growth promoting and heavy metal resistant strains for

**Commented [PS6]:** Please use a brighter color for the isolate names. Please check Cr<sup>2+</sup> in the figure

the enhancement of phytoremediation of heavy metal contaminated soil and for the promotion of soybean growth on slightly contaminated farmland. The low organic matter, N, P and K implied that the soil was barren (Wu et al. 2021; Zhang et al. 2012), yet the trap plants were nodulated, and the isolates from nodules were diverse based on the BOXA1R-PCR fingerprints. However, when inoculated on soybeans, only four out of the eight representative isolates formed nodules on the roots. Similar to rhizobia isolated from *Glycyrrhiza* spp. (Li et al. 2012), the four non-nodulating isolates may have been sporadic symbionts or other endophytes that had entered the trap plant nodules together with a genuine symbiont. Similar to the model inoculant of soybean, *Bradyrhizobium diazoefficiens* USDA110 (Sibponkrung et al. 2020), inoculation with the symbiotic isolates resulted in over two times higher biomass than in the uninoculated control; the higher biomass was accompanied with higher shoot nitrogen content. In addition, even the non-nodulating isolates showed some plant growth promoting abilities. Especially, inoculation with all the representative isolates resulted in higher root N content than in the nitrogen free control. In most of the inoculated plants, the increase in root N content was accompanied with lower P content.

As a host plant, soybean is promiscuous and may be nodulated with both fast and slow growing rhizobia (Chen et al. 2021). Based on the 16S rRNA gene analysis, three of our isolates were identified as *Ensifer americanum*, *E. fredii* or *E. xinjiangense*, i.e., as species with closely related 16S rRNA genes (Peng et al. 2002; Wang et al. 2013). Further analysis using MLSA of four housekeeping genes assigned the symbiotic isolates into the fast-growing rhizobial species *Ensifer xinjiangense* and *Rhizobium radiobacter*, strains of which have been identified as plant growth promoting symbionts of soybean plants (Iturralde et al. 2019; Peng et al. 2002). To our knowledge, neither *E. xinjiangense* (formerly *Sinorhizobium xinjiangense*) nor *R. radiobacter* (formerly *Agrobacterium tumefaciens*) strains have been applied as a single-inoculant plant

274 growth promoter in bioremediation, yet co-inoculation of an *A. tumefaciens* strain with *S.*  
275 *meliloti* promoted the growth of *Medicago lupulina* under Cu and Zn stresses (Jian et al. 2019).  
276 *R. radiobacter* is traditionally considered as a plant pathogen and is a free-living nitrogen fixer  
277 (Kanvinde et al. 1990). In our study, the amplification and sequencing of the *nifH* gene, which  
278 encodes nitrogenase iron protein, showed that both the *Ensifer* strains and *R. radiobacter* YN11  
279 had the genetic potential for nitrogen fixation. The nodulation and plant growth promotion  
280 capacity of *R. radiobacter* YN11 added to the growing body of evidence that when carrying the  
281 nodulation genes, *Rhizobium* (*Agrobacterium*) clade strains can be legume-nodulating symbionts  
282 (Cummings et al. 2009).

283 In the soil from mining areas, the concentrations of heavy metals vary considerably from  
284 below the background values for, to hundreds of times higher than average values in all soils (Li  
285 et al. 2014). The bacteria inhabiting the heavy metal-contaminated sites include legume  
286 nodulating strains with high tolerance against the metals (Mohamad et al. 2017). In our study, the  
287 symbiotic strains showed varied heavy metal resistance; *E. xinjiangense* YN5 outperformed the  
288 other *E. xinjiangense* isolates and the resistance levels of *R. radiobacter* YN11 fell in-between.  
289 Compared to the rhizobia isolated directly from V-Ti magnetite mine tailing soil and those from  
290 the nodules *Robinia pseudoacacia* in a Pb-Zn mining area (Fan et al. 2018a; Yu et al. 2014), our  
291 isolates tolerated lower levels of Cd<sup>2+</sup> and Cu<sup>2+</sup>. One possible explanation is the level of  
292 contamination on the sites where the strains were isolated; the V-Ti magnetite and Pb-Zn mining  
293 areas were seriously contaminated (Fan et al. 2018a; Yu et al. 2014), but only Zn content in the  
294 nickel mine soil was higher than in the background value for soils in China (Li et al. 2014). The  
295 levels of heavy metals tolerated are approximately 10 to 100 times lower in liquid medium than  
296 on solid medium (Hassen et al. 1998). It is also important to take into account the different  
297 testing methods for the Zn tolerance of *E. xinjiangense* YN5. The V-Ti magnetite and Pb-Zn

298 mining area isolates were tested on solid media (Fan et al. 2018a; Yu et al. 2014) but our isolates  
299 in liquid medium, yet *E. xinjiangense* YN5 tolerated a higher level of  $Zn^{2+}$ .

## 300 **Conclusions**

301 Our study used soybean pot experiment to trap 21 rhizobia strains from nickel mine soil. As  
302 a result, we selected three *Ensifer xinjiangense* strains (YN5, YN8, and YN10) and one  
303 *Rhizobium radiobacter* (YN11) with good nitrogen fixing ability, which can significantly  
304 improve the soybean plant height, root length, and biomass yield. Moreover, these 4 strains  
305 carried the symbiotic gene *nifH* that can encode dinitrogenase reductase enzyme, which further  
306 confirmed their root nodule formation and nitrogen fixation abilities. *E. xinjiangense* YN5 and *R.*  
307 *radiobacter* YN11 tolerated higher levels of heavy metals than *E. xinjiangense* YN8 and YN10.  
308 Taken together, the results showed that the nickel mine soil is a potential source for plant growth  
309 promoting rhizobia strains to be applied as indigenous inoculants in phytoremediation on slightly  
310 contaminated farmland to alleviate the adverse effects of heavy metals on soybean cultivation.

## 311 **Acknowledgments**

312 This study was supported by the National Natural Science Foundation of China [grant number  
313 31872696] and the Key Research Project of Sichuan Province [NO. 2021YFS0293]. The funders  
314 had no role in study design, data collection and analysis, decision to publish, or preparation of  
315 the manuscript. We thank Dr Xia Kang (School of Life Sciences, University of Dundee, United  
316 Kingdom) for English language editing.

317 **Reference**

- 318 Abd-Alla MH, Morsy FM, El-Enany A-WE, and Ohyama T. 2012. Isolation and characterization  
319 of a heavy-metal-resistant isolate of *Rhizobium leguminosarum* bv. *viciae* potentially  
320 applicable for biosorption of Cd<sup>2+</sup> and Co<sup>2+</sup>. *International Biodeterioration &*  
321 *Biodegradation* 67:48-55.
- 322 Balakrishnan B, Sahu BK, Kothilmozhian Ranishree J, Lourduraj AV, Nithyanandam M,  
323 Packiriswamy N, and Panchatcharam P. 2017. Assessment of heavy metal concentrations  
324 and associated resistant bacterial communities in bulk and rhizosphere soil of *Avicennia*  
325 *marina* of Pichavaram mangrove, India. *Environmental Earth Sciences* 76.
- 326 Catroux G, Hartmann A, and Revellin C. 2001. Trends in rhizobial inoculant production and use.  
327 *Plant and Soil* 230.
- 328 Chen JY, Gu J, Wang ET, Ma XX, Kang ST, Huang LZ, Cao XP, Li LB, and Wu YL. 2014. Wild  
329 peanut *Arachis duranensis* are nodulated by diverse and novel *Bradyrhizobium* species in  
330 acid soils. *Systematic Applied Microbiology* 37:525-532.
- 331 Chen WF, Wang ET, Ji ZJ, and Zhang JJ. 2021. Recent development and new insight of  
332 diversification and symbiosis specificity of legume rhizobia: mechanism and application.  
333 *Journal of Applied Microbiology* 131:553-563.
- 334 Cummings SP, Gyaneshwar P, Vinuesa P, Farruggia FT, Andrews M, Humphry D, Elliott GN,  
335 Nelson A, Orr C, Pettitt D, Shah GR, Santos SR, Krishnan HB, Odee D, Moreira FM,  
336 Sprent JI, Young JP, and James EK. 2009. Nodulation of Sesbania species by *Rhizobium*  
337 (*Agrobacterium*) strain IRBG74 and other rhizobia. *Environmental Microbiology*  
338 11:2510-2525.
- 339 Dhuldhaj U, and Pandya U. 2020. Combinatorial study of heavy metal and microbe interactiona  
340 and resistance mechanism consort to microbial system. *Geomicrobiology Journal* 38:

181-189.

Duan C, Razavi BS, Shen G, Cui Y, Ju W, Li S, and Fang L. 2019. Deciphering the rhizobium inoculation effect on spatial distribution of phosphatase activity in the rhizosphere of alfalfa under copper stress. *Soil Biology and Biochemistry* 137.

Fan M, Lin Y, Huo H, Liu Y, Zhao L, Wang E, Chen W, and Wei G. 2016. Microbial communities in riparian soils of a settling pond for mine drainage treatment. *Water Research* 96:198-207.

Fan M, Liu Z, Nan L, Wang E, Chen W, Lin Y, and Wei G. 2018a. Isolation, characterization, and selection of heavy metal-resistant and plant growth-promoting endophytic bacteria from root nodules of *Robinia pseudoacacia* in a Pb/Zn mining area. *Microbiological Research* 217:51-59.

Fan M, Xiao X, Guo Y, Zhang J, Wang E, Chen W, Lin Y, and Wei G. 2018b. Enhanced phytoremediation of *Robinia pseudoacacia* in heavy metal-contaminated soils with rhizobia and the associated bacterial community structure and function. *Chemosphere* 197:729-740.

Grandlic CJ, Palmer MW, and Maier RM. 2009. Optimization of Plant Growth-Promoting Bacteria-Assisted Phytostabilization of Mine Tailings. *Soil Biology and Biochemistry* 41:1734-1740.

Hao X, Taghavi S, Xie P, Orbach MJ, Alwathnani HA, Rensing C, and Wei G. 2014. Phytoremediation of heavy and transition metals aided by legume-rhizobia symbiosis. *International Journal of Phytoremediation* 16:179-202.

Hao X, Xie P, Zhu YG, Taghavi S, Wei G, and Rensing C. 2015. Copper tolerance mechanisms of *Mesorhizobium amorphae* and its role in aiding phytostabilization by *Robinia pseudoacacia* in copper contaminated soil. *Environmental Science and Technology* 49:2328-2340.



366 Hassen A, Saidi N, Cherif M, and Boudabous AJBT. 1998. Resistance of environmental bacteria  
367 to heavy metals. *Bioresource Technology* 64:7-15.

368 Iturralde ET, Covelli JM, Alvarez F, Pérez-Giménez J, Arrese-Igor C, and Lodeiro AR. 2019.  
369 Soybean-Nodulating Strains With Low Intrinsic Competitiveness for Nodulation, Good  
370 Symbiotic Performance, and Stress-Tolerance Isolated From Soybean-Cropped Soils in  
371 Argentina. *Frontiers in Microbiology* 10:1061.

372 Jian L, Bai X, Zhang H, Song X, and Li Z. 2019. Promotion of growth and metal accumulation  
373 of alfalfa by coinoculation with *Sinorhizobium* and *Agrobacterium* under copper and zinc  
374 stress. *PeerJ* 7:e6875.

375 Kanvinde L, Sastry GJA, and Microbiology E. 1990. *Agrobacterium tumefaciens* Is a  
376 Diazotrophic Bacterium. *Applied & Environmental Microbiology* 56:2087-2092.

377 Ke T, Guo G, Liu J, Zhang C, Tao Y, Wang P, Xu Y, and Chen L. 2021. Improvement of the Cu  
378 and Cd phytostabilization efficiency of perennial ryegrass through the inoculation of  
379 three metal-resistant PGPR strains. *Environmental Pollution* 271:116314.

380 Laguerre G, Nour SM, Macheret V, Sanjuan J, Drouin P, and Amarger N. 2001. Classification of  
381 rhizobia based on nodC and *nifH* gene analysis reveals a close phylogenetic relationship  
382 among *Phaseolus vulgaris* symbionts. *Microbiology* 147(Pt 4):981-993.

383 Lebrazi S, and Fikri-Benbrahim K. 2018. Rhizobium-Legume Symbioses: Heavy Metal Effects  
384 and Principal Approaches for Bioremediation of Contaminated Soil. *Legumes for Soil*  
385 *Health and Sustainable Management*, 205-233.

386 Li L, Hanna Sinkko, Leone Montonen, Gehong Wei, Kristina Lindström, and Räsänen LA. 2012.  
387 Biogeography of symbiotic and other endophytic bacteria isolated from medicinal  
388 *Glycyrrhiza* species in China. *FEMS Microbiology Ecology* 79:46–68.

389 Li X, Wang X, Chen Y, Yang X, and Cui Z. 2019. Optimization of combined phytoremediation  
390 for heavy metal contaminated mine tailings by a field-scale orthogonal experiment.

391 *Ecotoxicology and Environmental Safety* 168:1-8.

392 Li Z, Ma Z, van der Kuijp TJ, Yuan Z, and Huang L. 2014. A review of soil heavy metal  
393 pollution from mines in China: pollution and health risk assessment. *Science of Total*  
394 *Environmental* 468-469:843-853.

395 Limcharoensuk T, Sooksawat N, Sumarnrote A, Awutpet T, Kruatrachue M, Pokethitiyook P, and  
396 Auesukaree C. 2015. Bioaccumulation and biosorption of Cd<sup>2+</sup> and Zn<sup>2+</sup> by bacteria  
397 isolated from a zinc mine in Thailand. *Ecotoxicology and Environmental Safety* 122:322-  
398 330.

399 Lindstrom K, and Mousavi SA. 2020. Effectiveness of nitrogen fixation in rhizobia. *Microbial*  
400 *Biotechnology* 13:1314-1335.

401 Long Z, Huang Y, Zhang W, Shi Z, Yu D, Chen Y, Liu C, and Wang R. 2021. Effect of different  
402 industrial activities on soil heavy metal pollution, ecological risk, and health risk.  
403 *Environmental Monitoring and Assessment* 193: 1-12.

404 Mohamad R, Maynaud G, Le Quere A, Vidal C, Klonowska A, Yashiro E, Cleyet-Marel JC, and  
405 Brunel B. 2017. Ancient heavy metal contamination in soils as a driver of tolerant  
406 *Anthyllis vulneraria* rhizobial communities. *Applied & Environmental Microbiology*  
407 83(2):e01735-16.

408 Murphy J, and Riley J. 1962. A modified single solution method for the determination of  
409 phosphate in natural waters. *Anal. Chim. Acta* 27:31-36.

410 Page A, Miller R, and Keeney DJASoA, Madison, WI. 1982. *Methods of Soil Analysis*, part II  
411 15(1):99-100.

412 Peng G, Tan Z, Wang E, Reinhold-Hurek B, Chen WF, and Chen WX. 2002. Identification of  
413 isolates from soybean nodules in Xinjiang Region as *Sinorhizobium xinjiangense* and  
414 genetic differentiation of *S. xinjiangense* from *Sinorhizobium fredii*. *International*  
415 *Journal of Systematic and Evolutionary Microbiology* 52:457-462.

416 Qin G, Niu Z, Yu J, Li Z, Ma J, and Xiang P. 2021. Soil heavy metal pollution and food safety in  
417 China: Effects, sources and removing technology. *Chemosphere* 267 :129205.

418 Saitou N, and Nei M. 1987 The neighbor-joining method a new method for reconstructing  
419 phylogenetic trees. *Molecular Biology & Evolution* 4:406.

420 Salmi A, and Boulila F. 2021. Heavy metals multi-tolerant *Bradyrhizobium* isolated from  
421 mercury mining region in Algeria. *Journal of Environmental Management* 289:112547.

422 Schumacher BA. 2002. Methods for the Determination of Total Organic Carbon (TOC) in Soils  
423 and Sediments. *Ecological risk assessment support center*.

424 Shao Y, Yan T, Wang K, Huang S, Yuan W, and Qin FGF. 2020. Soil heavy metal lead pollution  
425 and its stabilization remediation technology. *Energy Reports* 6:122-127.

426 Sibponkrung S, Kondo T, Tanaka K, Tittabutr P, Boonkerd N, Yoshida KI, and Teaumroong N.  
427 2020. Co-Inoculation of *Bacillus velezensis* Strain S141 and *Bradyrhizobium* Strains  
428 Promotes Nodule Growth and Nitrogen Fixation. *Microorganisms* 8(5):678.

429 Tang J, Bromfield ES, Rodrigue N, Cloutier S, and Tambong JT. 2012. Microevolution of  
430 symbiotic *Bradyrhizobium* populations associated with soybeans in east North America.  
431 *Ecology & Evolution* 2:2943-2961.

432 Thakare M, Sarma H, Datar S, Roy A, Pawar P, Gupta K, Pandit S, and Prasad R. 2021.  
433 Understanding the holistic approach to plant-microbe remediation technologies for  
434 removing heavy metals and radionuclides from soil. *Current Research in Biotechnology*  
435 3:84-98.

436 Trung BC, Yoshida SJSS, and Nutrition P. 1982. Improvement of Leonard jar assembly for  
437 screening of effective rhizobium. *Soil Science and Plant Nutrition* 29:97-100.

438 Wang H, Gu C, Liu X, Yang C, Li W, and Wang S. 2020. Impact of Soybean Nodulation  
439 Phenotypes and Nitrogen Fertilizer Levels on the Rhizosphere Bacterial Community.  
440 *Frontiers in Microbiology* 11:750.

441 Wang Q, Ma L, Zhou Q, Chen B, Zhang X, Wu Y, Pan F, Huang L, Yang X, and Feng Y. 2019.  
 442 Inoculation of plant growth promoting bacteria from hyperaccumulator facilitated non-  
 443 host root development and provided promising agents for elevated phytoremediation  
 444 efficiency. *Chemosphere* 234:769-776.

445 Wang Y, Wang F, Hou B, Wang E, Chen W, Sui X, Chen W, Li Y, and Zhang Y. 2013. Proposal of  
 446 *Ensifer psoraleae* sp. nov., *Ensifer sesbaniae* sp. nov., *Ensifer morelense* comb. nov. and  
 447 *Ensifer americanum* comb. nov. *Systematic and Applied Microbiology* 36:467-473.

448 Wu B, Peng H, Sheng M, Luo H, Wang X, Zhang R, Xu F, and Xu H. 2021. Evaluation of  
 449 phytoremediation potential of native dominant plants and spatial distribution of heavy  
 450 metals in abandoned mining area in Southwest China. *Ecotoxicology and Environmental*  
 451 *Safety* 220:112368.

452 Yu X, Li Y, Li Y, Xu C, Cui Y, Xiang Q, Gu Y, Zhao K, Zhang X, and Penttinen P. 2017.  
 453 *Pongamia pinnata* inoculated with *Bradyrhizobium liaoningense* PZHK1 shows potential  
 454 for phytoremediation of mine tailings. *Applied Microbiology and Biotechnology*  
 455 101:1739-1751.

456 Yu X, Li Y, Zhang C, Liu H, Liu J, Zheng W, Kang X, Leng X, Zhao K, Gu Y, Zhang X, Xiang Q,  
 457 and Chen Q. 2014. Culturable heavy metal-resistant and plant growth promoting bacteria  
 458 in V-Ti magnetite mine tailing soil from Panzhihua, China. *PLoS One* 9:e106618.

459 Yu X, Shen T, Kang X, Cui Y, Chen Q, Shoaib M, Liu H, Zhang F, Hussain S, Xiang Q, Zhao K,  
 460 Gu Y, Ma M, Li S, Zou L, and Liang Y. 2021. Long-term phytoremediation using the  
 461 symbiotic *Pongamia pinnata* reshaped soil micro-ecological environment. *Science of the*  
 462 *Total Environment* 774: 145112.

463 Zhang H, Yuan X, Xiong T, Wang H, and Jiang L. 2020. Bioremediation of co-contaminated soil  
 464 with heavy metals and pesticides: Influence factors, mechanisms and evaluation methods.  
 465 *Chemical Engineering Journal* 398:125659.

466 Zhang X, Yang L, Li Y, Li H, Wang W, Ye BJEM, and Assessment. 2012. Impacts of lead/zinc  
467 mining and smelting on the environment and human health in China. *Environmental*  
468 *Monitoring & Assessment* 184:2261-2273.

469 Zhao L, Fan M, Zhang D, Yang R, Zhang F, Xu L, Wei X, Shen Y, and Wei G. 2014. Distribution  
470 and diversity of rhizobia associated with wild soybean (*Glycine soja* Sieb. & Zucc.) in  
471 Northwest China. *Systematic & Applied Microbiology* 37:449-456.

472 Zhou H, Zeng M, Zhou X, Liao BH, Liu J, Lei M, Zhong QY, and Zeng H. 2013. Assessment of  
473 heavy metal contamination and bioaccumulation in soybean plants from mining and  
474 smelting areas of southern Hunan Province, China. *Environmental Toxicology &*  
475 *Chemistry* 32:2719-2727.

476